



Soil greenhouse gas emissions and crop production with implementation of alley cropping in a Mediterranean citrus orchard

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ABSTRACT

The implementation of alley cropping in orchards has been suggested as a sustainable strategy to increase farmer revenues by crop diversification, enhance soil organic matter (SOM) and fertility, water retention, overall biodiversity, and contribute to climate change mitigation. Thus, the objective of this study was to assess if alley cropping with annual crops can contribute to i) mitigate soil greenhouse gas (GHG) emissions, ii) enhance C sequestration in a semiarid Mediterranean irrigated citrus orchard, and iii) increase land productivity. For this, two different treatments were established: i) conventional mandarin monoculture (MC) with no alley cropping; and ii) mandarin diversified with alley cropping of barley/vetch and fava bean (DIV). Measurements of soil CO₂ and N₂O emissions were periodically performed (every 7–20 days) during two years. Soil CO₂ emission rates followed the soil moisture trend, and showed no significant differences between treatments. As an average, soil CO₂ emission rates were 147 mg m⁻² h⁻¹ in MC and 196 mg m⁻² h⁻¹ in DIV. Soil N₂O emission rates were not correlated to soil moisture nor temperature, and showed average values of 0.026 mg m⁻² h⁻¹ in MC and – 0.002 mg m⁻² h⁻¹ in DIV. Alley cropping did not contribute to significantly increase soil organic C and total nitrogen in two years' time. With regard to production, mandarin yield showed no significant differences between treatments, but alley crops contributed to complementary commodities to the main cash crop, increasing overall land productivity. Thus, alley cropping in irrigated Mediterranean orchards has no significant effect on soil C sequestration and GHG emissions at short-term, with increased land productivity owing to new commodities grown in the alleys. These results confirm that under semiarid Mediterranean climate, long periods are needed to efficiently assess soil C sequestration potential of sustainable practices in orchards.

1. Introduction

The green revolution was a milestone in the last century since it allowed the increase of crop production worldwide, thanks to the adoption of high-yielded crop varieties, chemical fertilizers, pesticides and mechanization (Pingali, 2012). This new paradigm was fostered for decades leading to the current agri-business models based on monocultures and intense mechanization and dependence on external inputs (Morugán-Coronado et al., 2020). However, it is of international consensus that this production model has also led to environmental degradation, with soil and water pollution, decrease of biodiversity, loss

of soil health, loss of soil by erosion, greenhouse gas (GHG) emissions, higher incidence of pests/diseases with higher resistance to pesticides, and, in the end, low resilience of the agro-ecosystems (Maleki et al., 2021; Martínez-Núñez et al., 2020; Morugán-Coronado et al., 2020; Zeng et al., 2022). This fact is even aggravated in most tree orchards under Mediterranean conditions, where trees are irrigated and alleys remain bare to avoid competition for water and nutrients with weeds or other crops. In addition, there has been a traditional social belief that maintaining vegetation cover in the alleys is associated to abandonment and dirt, and so weeds are controlled by tillage or herbicides application (Cerdà et al., 2018; Tsanis et al., 2021). This is usually associated to the

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lack of consideration of soil as a non-renewable resource, but as a mere physical substrate (Cerdà et al., 2018). As a consequence, soil erosion rates in Mediterranean orchards are high, with high mineralization of soil organic matter (SOM) and GHG emissions, leading to a soil with poor structure and low content of SOM and nutrients, with low capacity to hold water and favor infiltration (Cerdà et al., 2021; Martínez-Mena et al., 2019; Tsanis et al., 2021).

The use of cover crops is being fostered in orchards worldwide to counteract the negative effects of keeping the soil bare, with positive effects such as enhancement of SOM, increased soil fertility, higher water retention capacity and attraction of auxiliary fauna, increasing overall biodiversity (Almagro et al., 2017; Martínez-Mena et al., 2019; Martínez-Núñez et al., 2020). However, their use is still not generalized in the Mediterranean region, and more incentives are needed to encourage farmers to adopt this practice. In this sense, alley cropping could be a suitable strategy to achieve environmental benefits but also increasing land profitability by production of complementary commodities (Swieter et al., 2021; Yang et al., 2021). Alley cropping is the agronomic strategy of growing annual or perennial crops in the alleys of tree orchards to increase farmer revenues, while increasing the land productivity by producing several commodities in the same space during the same year (Tsonkova et al., 2012; Wolz and DeLucia, 2018). If properly selected and managed, alley crops can increase soil organic matter content in the soil, contributing to soil C sequestration (Reyes et al., 2021), can increase soil fertility by stimulation of soil microorganisms that can fix atmospheric N or solubilize soil nutrients (Beule and Karlovsky, 2021) and can attract pollinators and auxiliary fauna to fight against pests (Morugán-Coronado et al., 2020; Staton et al., 2019). Under this strategy, farmers could reduce the use of external fertilizers and pesticides while increase C sequestration in soil and biomass, reducing the C footprint of the agroecosystems (Pavlidis et al., 2020). In fact, this strategy is aligned with the European Green Deal (European Commission, 2019) and the European Climate Law (European Commission, 2020), which aim to make a fair transition in the EU's economy to achieve climate neutral farms by 2050. In this sense, agricultural soils present a unique opportunity for C sequestration and compensating emissions by sustainable cropping systems (Chabbi et al., 2017).

It is essential, so, to define what strategies can lead to climate neutrality in farms by reducing GHG emissions and increasing C sequestration in soil and biomass. For this, a proper selection of species as alley cropping is vital. Legumes have been proposed as perfect candidates in crop diversification, since they can improve soil fertility by biological N fixation with their symbiosis with rhizobia, with higher efficiency in the use of N (Sánchez-Navarro et al., 2019). Besides, legumes have a very active rhizodeposition, stimulating microbial communities that can stabilize organic matter in the soil and solubilize essential nutrients to the plant (Sánchez-Navarro et al., 2020), attracting beneficiary insects into the farm (Cole et al., 2022).

Hence, there is a need to provide scientific data about agronomic strategies than can foster climate neutrality in farms, by decreasing GHG emissions and increasing C storage. In previous research, we focused on GHG emissions on the orchard tree rows and how irrigation affects their trends (Zornoza et al., 2018, 2016); so, new information is needed on how alley management can contribute to GHG in tree orchards. Accordingly, the objectives of this study were to: i) assess if alley cropping with the multiple cropping of barley/vetch and fava bean can contribute to mitigate alley soil GHG emissions and increase soil C sequestration in a Mediterranean irrigated citrus orchard; and ii) evaluate if alley cropping can contribute to increase land productivity. Thus, we hypothesized that the growth of alley crops among the tree rows may increase soil CO₂ and N₂O emissions by active rhizodeposition and stimulation of soil microbial communities. However, higher organic outputs by root exudates and plant residues would foster an increase in SOM, which may counteract increased GHG emissions; so, the net balance would be C sequestration in the system. The growth of barley/vetch and fava bean in the same space than the citrus trees would increase land

productivity by the harvest of new products, with no negative effect on the main tree crop yield.

2. Materials and methods

2.1. Study site and experimental design

This experiment was performed from February 2019 to March 2021 in a commercial mandarin orchard (*Citrus reticulata* Blanco var. Clemenvilla), located in Murcia, SE Spain (37° 57' 31''N; 0° 56' 17'' W). The *C. reticulata* orchard had an extension of 2.3 ha, with 970 trees, at a spacing of 6 m between rows and 4 m between trees within the same row, planted in the year 2000. The climate is semiarid Mediterranean with mean annual temperature of 18 °C and mean annual rainfall of 280 mm. The potential evapotranspiration rate is 1300 mm year⁻¹. The soil is a Calcaric Regosol (IUSS Working Group WRB, 2014), with silt loam texture (7 %, 70 % and 23 % of sand, silt and clay, respectively), 55 % of CaCO₃ content, pH of 7.56, bulk density of 1.20 g cm⁻³, total organic carbon of 7.6 g kg⁻¹, total nitrogen of 1.02 g kg⁻¹ and a cation exchange capacity of 14.5 cmol₊ kg⁻¹. A drip irrigation system was installed in all tree rows, with one line per tree row and 3 pressure-compensated emitters (4 L h⁻¹) per tree. Irrigation was scheduled weekly and applied at night or early morning. The frequency of irrigation varied according to the evaporative demand which was 1–2 times per week in winter, 2–7 times per week in spring and autumn, and 7–14 times per week in summer. The EC of the irrigation water varied between 1.0 and 1.5 dS m⁻¹.

Two different treatments were established as split-plot design with three replicates in February 2018. Plots of 12 m × 28 m were established, with the long side of each one following the direction of the maximum slope, including three rows of 6 trees. The average plot slope was 12 %. Treatments were: i) mandarin monoculture with no alley cropping (MC); and ii) mandarin diversified (DIV) with fava bean (*Vicia faba* L.) grown from September to December for human consumption and barley/vetch (*Hordeum vulgare* L./*Vicia sativa* L.) grown from February to June for feed as alley cropping (see photos of the plots in the Figs. S1–3 of the Supplementary material). Fava bean seeds were manually sown under drip irrigation in three rows in each alley in early September 2018, 2019 and 2020, with a spacing of 100 cm between rows and 40 cm between plants (2.5 plants m⁻²). Crop residues were incorporated in the soil after harvest of pods. Barley/vetch seeds (1:3 ratio) were manually sown at 150 kg ha⁻¹ covering the entire alley surface in early February 2018, 2019, 2020 and 2021. Three drip irrigation lines were established for irrigation of alley crops in DIV with pressure-compensated emitters (4 L h⁻¹) every 40 cm. Tillage was performed (0–30 cm) immediately before sowing the alley crops to prepare the soil and after the harvest of aerial parts for animal feed to incorporate crop residues into the soil. For DIV, the alley crop water needs were estimated as the product of precipitation, reference crop evapotranspiration (ET₀) and the crop coefficients (between 0.25 and 0.55) proposed by the Agricultural Information System of Murcia (<http://siam.imida.es>) for this area. Consumption of water in the alley crops in DIV was 1770 m³ ha⁻¹ for fava bean and 1280 m³ ha⁻¹ for barley/vetch (a total of 3050 m³ ha⁻¹ for the two years of study).

Usual cultural practices (e.g. fertilization, pruning, fruit thinning and banding) were carried out by the technical department of the commercial orchard. Weed control was carried out in MC by tillage (chisel plowing 4–5 times yr⁻¹ at 20 cm depth), while weeds were not controlled in DIV. Pruning residues were chopped *in situ* and left as much on the soil surface in the monoculture. Pruning residues were not applied as mulch in DIV because at that time alley was cover with barley/vetch crop. So, pruning residues were removed from the system. Table S1 shows the C and N inputs applied to soil as pruning residues in MC and alley crop residues in DIV. Fertilizers were applied as fertigation by use of the commercial products Neptuno PK 28, Neptuno Triton and Neptuno Pandora (Medifer, Constantino Gutiérrez, SA), as a mixture of soluble N,

P, K, Ca, Mg and chelated micronutrients. No irrigation was performed in the allays of MC. Table S2 shows the quantity of N, P and K added with fertilizers per treatment and crop each month.

2.2. Soil greenhouse gas measurements

Measurements of soil CO₂ and N₂O were made every 7–20 days, depending on climate conditions, in all replicated treatments from 05/04/2019 to 04/03/2021, between 9:00 and 11:00. The basic experimental procedure used in this study was the dynamic gas chamber technique. The chamber was made of non-oxidizable steel, with a diameter of 7.5 cm and a height of 20 cm, with one inlet and one outlet connected to a photoacoustic infrared spectroscopy multi-gas analyser with ultra-sensitive cantilever pressure sensor (Gasera One, Gasera Ltd). The dynamic system with inlet and outlet in the chamber permits a continuous flow and avoids pressure fluctuations. The chambers were adjusted over a non-oxidizable steel base which was randomly inserted into the bare soil to a depth of 15 cm in the middle of the alleys in MC and in DIV when no alley crop was present and with the barley/vetch crop. Chambers were randomly inserted between two fava bean plants in the central row. Bases were kept inserted in the soil and without vegetation, except for tillage events, when they were removed and replaced. Inserts were also kept out of vegetation in a surrounding ring of 20 cm wide to reduce the effect of autotrophic respiration (Smith et al., 2010). CO₂ and N₂O were quantified every 1 min for a period of 5 min to assess the linear trend. CO₂ and N₂O emissions rates were expressed in mg m⁻² h⁻¹ as the difference between the quantification at the end and the beginning of the measure period divided by the time. CO₂ and N₂O cumulative emissions for each treatment were estimated by numerical integration (Chen et al., 2013). GHG emissions were converted into CO₂equivalent (CO₂e) for cumulative emission data (g m⁻²) for the experimental period. For this, N₂O emissions were converted into CO₂e according to their global warming potential, which is 265 (Vasconcelos et al., 2022). Overall CO₂e emission was the sum of CO₂ emissions and N₂O emissions converted into CO₂e.

2.3. Crop production, soil sampling and analytical methods

Meteorological data were measured using an automatic weather station located in a nearby orchard (4 km). Soil temperature (T) and soil moisture (M) were measured using a ProCheck and 5TM sensors (Decagon Devices, USA) introduced at 15 cm depth adjacent to the place where GHG measures were done.

Mandarin yield was calculated by weighing all the mandarins harvested directly from the trees in each plot on 24/01/2019, 07/01/2020 and 03/02/2021. Barley/vetch yield was determined by weighing the aerial biomass cut on 13/06/2019 and 22/06/2020 (no commercial production was obtained in 2018). Fava bean was harvested by collecting all the pods in each plot when the seeds were fresh, on 09/01/2019, 23/12/2020 and 12/02/2021, and weighed as crop yield.

Two soil sampling campaigns were performed: 21/02/2019 and 17/02/2021 at two different depths (0–10 cm and 10–30 cm) with an auger. Three composite samples derived from 5 random subsamples were collected in each plot (9 composite soil samples per treatment). Soil cores using steel cylinders were taken to determine soil bulk density (BD). Soil was air-dried for one week and sieved at < 2 mm.

Particle size distribution was measured using a Mastersizer analyser 2000LF (Malvern Instruments). Soil pH and electrical conductivity (EC) were measured in deionized water (1:2.5 and 1:5 w/v, respectively). Total organic carbon (TOC), total inorganic C and total nitrogen (Nt) were determined by an elemental CHNS-O analyser (EA-1108, Carlo Erba). Soil NH₄⁺ was extracted with 2 M KCl in a 1:10 soil:extractant ratio and calorimetrically measured (Kandeler and Gerber, 1988; Keeny and Nelson, 1982). Soil NO₃⁻ was extracted with deionized water in a 1:10 soil:extractant ratio and measured by ion chromatography (Metrohm 861).

2.4. Statistical analysis

Data were checked to ensure normal distribution using the Kolmogorov-Smirnov test at P < 0.05. Homoscedasticity was checked by the Levene test. GHG data were submitted to two-way repeated measures ANOVA, with measurement date as within-subject factor, and treatment (MC and DIV) as between-subject factor. Crop yield and GHG data were submitted, independently for each date, to a t-test to compare significant differences between MC and DIV. Soil data were submitted to two-way repeated measures ANOVA, with sampling date (2019 and 2021) as within-subject factor, and treatment (MC and DIV) as between-subject factor. Relationships among properties were studied using Pearson correlations. Statistical analyses were performed with the software IBM SPSS for Windows, Version 20.

3. Results

3.1. Crop yields

Mandarin crop yield showed no significant differences between treatments in the experimental period of study (Table 1). However, during the last crop cycle, mandarin yield decreased ten times the values obtained in the first year, owing to intense affection of a *Alternaria* sp. that hindered fructification of most flowers. Fava bean and barley/vetch contributed to complementary commodities to the main cash crop, which supposed 10–15 % of overall land production during the first two years but reached 50 % of overall land production in the second year owing to the low mandarin yields. In fact, overall land productivity in DIV (8565 kg ha⁻¹) was significantly higher than in MC (3020 kg ha⁻¹) during the last year, with no significant differences the previous years.

3.2. Greenhouse gas emission rates

Soil CO₂ emission rates followed the soil moisture trend, as shown in Fig. 1, with a positive significant correlation between both properties (R = 0.41; P < 0.01). Highest CO₂ emissions were associated to highest soil moisture and high temperature values. Alley cropping contributed to significantly increase (P < 0.05) soil moisture compared to monoculture owing to irrigation (Fig. 1). As an average, soil moisture was 11.6 % in MC and 14.9 % in DIV for the entire experimental period (24 months).

Table 1

Crop yield for mandarin trees and alley crops in the two crop cycles of study (kg ha⁻¹). Values are mean ± standard error (n = 3). Overall land productivity represents the sum of all crop yields during that cycle.

Treatment	Mandarin	Barley/vetch	Fava bean	Overall land productivity
Cycle 2018/2019				
Diversified system	18,676 ± 869	–	1019 ± 254	19,695 ± 2520
Monoculture	20,922 ± 2509	–	–	20,922 ± 2509
t-Student	1.587 ns			0.278 ns
Cycle 2019/2020				
Diversified system	33,107 ± 4721	4728 ± 258	1589 ± 164	39,421 ± 1686
Monoculture	45,436 ± 1860	–	–	45,436 ± 4950
t-Student	1.281 ns			-2.25 ns
Cycle 2020/2021				
Diversified system	4456 ± 1853	1417 ± 141	2691 ± 271	8565 ± 1881
Monoculture	3020 ± 247	–	–	3020 ± 247
t-Student	-0.728 ns			-2.92*

ns: not significant (P > 0.05).

* Significant at P < 0.05.

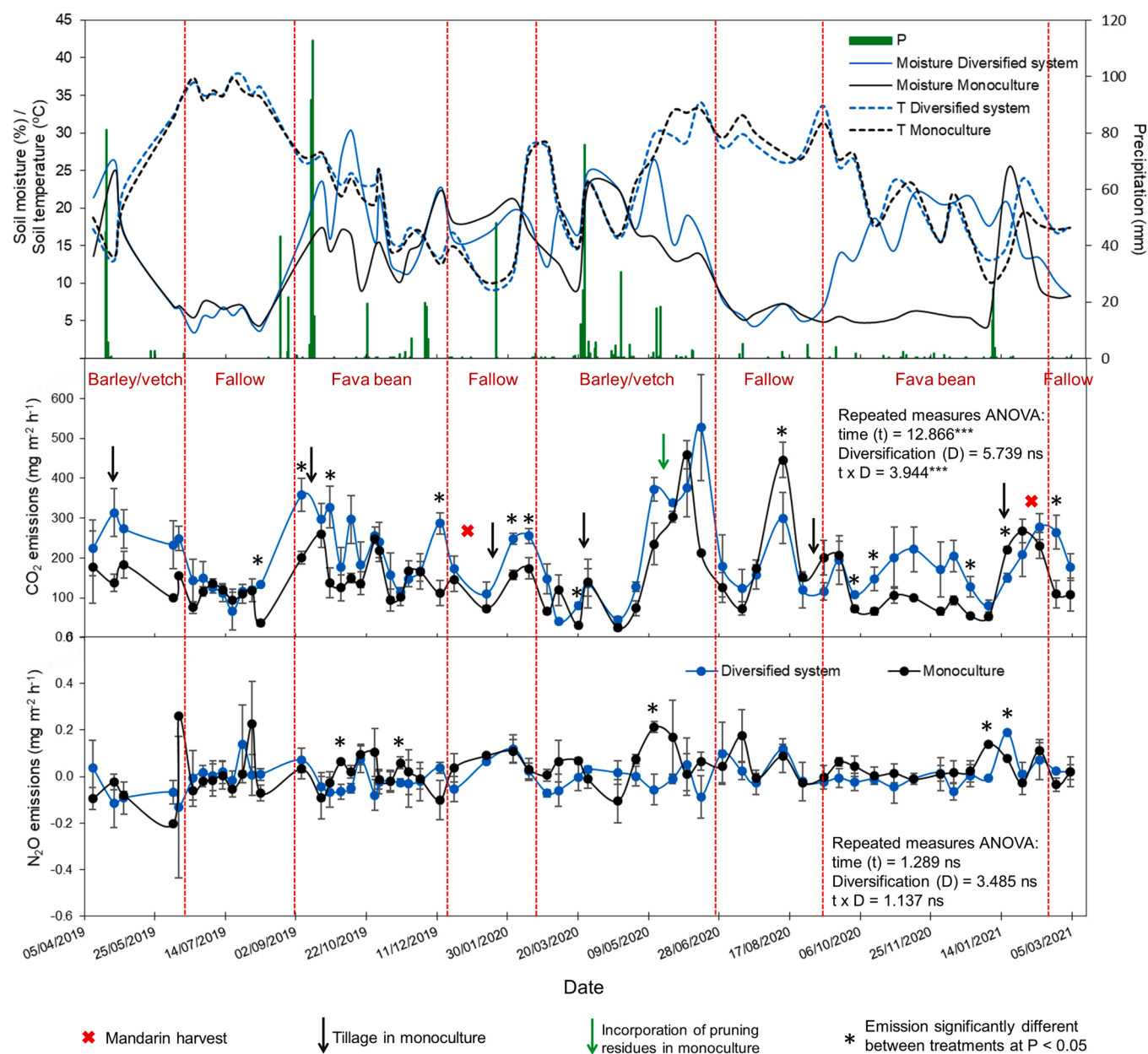


Fig. 1. Environmental conditions during the duration of the experiment (top), soil CO₂ emission rates (center) and soil N₂O emission rates (bottom) from the mandarin monoculture and diversified orchard. Values are mean \pm standard error ($n = 3$). P: precipitation; T: soil temperature. For repeated measures ANOVA data: significant at *** $P < 0.001$; ns: not significant ($P > 0.05$).

Soil CO₂ emission rates were, in general, not significantly different between MC and DIV (diversification factor) for the complete experimental period (Fig. 1). However, 11 out of the 60 CO₂ emission rate measures were significantly higher in DIV than in MC, mostly associated to the fava bean growth; these significant increases also appeared in the fallow periods and in one occasion in the 2020 barley/vetch cycle (Fig. 1). There were two occasions when MC had significantly higher emissions than DIV: i) in the fallow period after barley/vetch 2020, and ii) in the fava bean cycle of 2020, both associated to an increase in soil moisture owing to rainfall events. As an average, CO₂ emission rates were 152 ± 114 mg m⁻² h⁻¹ in MC and 196 ± 109 mg m⁻² h⁻¹ in DIV for the entire experimental period.

Soil N₂O emission rates were not correlated to soil moisture nor temperature, with a flat trend with small oscillations up and down of 0 mg m⁻² h⁻¹ (Fig. 1). Soil N₂O emission rates were not significantly different between MC and DIV for the complete experimental period.

However, four out of the 60 N₂O emission rate measures were significantly higher in MC, with two episodes in autumn 2019, one episode in late spring 2020 and the last episode in winter 2021. There was only once, on 14/01/2021, when N₂O emission was higher in DIV, when fava bean was growing. As an average, N₂O emission rates were 0.026 ± 0.114 mg m⁻² h⁻¹ in MC and -0.002 ± 0.118 mg m⁻² h⁻¹ in DIV for the entire experimental period.

There was no clear evidence about the direct effect of tillage on changes in soil CO₂ or N₂O emissions (Fig. 1). The incorporation of pruning residues in the alleys of MC on 25/05/2020 was led by a peak in soil CO₂ emissions. However, this peak was also observed in DIV, and so it may be likely due to high soil temperatures during those days, associated to high soil moisture in all plots.

3.3. Overall cumulative emissions

The estimation of cumulative CO₂, N₂O and CO₂e released during the experimental period confirmed the lack of significant differences between treatments owing to the high temporal variability of these fluxes (Table 2). Nonetheless, DIV tended to decrease the cumulative N₂O emission and to increase the CO₂ emission. Cumulative CO₂ emission was not correlated with any soil property measured (clay, sand, silt, pH, EC, TOC, soil inorganic carbon, Nt, NH₄⁺, NO₃⁻), while cumulative N₂O emission was significantly correlated to soil NO₃⁻ (R = 0.48; P < 0.01), but not with Nt neither NH₄⁺.

3.4. Soil organic carbon and total nitrogen contents

TOC and Nt was initially lower in DIV than in MC at 0–10 cm depth, but alley cropping did not contribute to significantly increase these properties in any of the soil depths in two years' time (Fig. 2).

4. Discussion

Alley cropping was successfully established in the study site, with increased land productivity, achieving several crops. Nonetheless, to ensure alley crop production, irrigation was needed, and thus, soil from the alleys in DIV received more water and nutrients than the alley soil in MC. Root exudates by alley crops may have stimulated and activated soil microbial communities, as previously reported (D'Hervilly et al., 2021; Wachendorf et al., 2020). An evidence of that is that fava bean showed no nodules during the first cycle but developed a lot during the last crop cycle (Supplementary Table S3). This is a sign of improvement in soil health by increases in the abundance of beneficial microorganisms (Bertola et al., 2021). Thus, factors introduced in the alleys of DIV and not present in those of MC such as: i) irrigation; ii) fertilization and iii) vegetation growth and inputs of organic compounds by rhizodeposition and crop residues, have likely stimulated microbial communities and so microbial activity should be higher, associated with the metabolic activity of plant roots (root respiration). In this line, D'Hervilly et al. (2021) observed that alley cropping with herbaceous strips showed higher microbial biomass and a higher density of earthworms than the uncultivated alleys. Guillot et al. (2021) also reported that compared to tree monocultures, microbial basal respiration and glucose-induced respiration increased with alley cropping. However, we did not find significant differences in terms of soil CO₂ emissions between the bare alley soil in MC and the plant covered alley soil in DIV in the entire experimental period of 24 months. Thus, despite some isolated episodes with highest CO₂ emissions in DIV, the overall system is not significantly contributing to increase GHG emissions compared to a bare alley, typical of orchards in the Mediterranean regions. Similarly, Almagro et al. (2016) showed no significant differences in soil CO₂ emissions between rainfed almond orchards with bare soil and covered with cover crops. However, it was expected, as initial hypothesis, that higher soil moisture would contribute to even higher CO₂ emissions from alley soil in DIV, as reported by Zornoza et al. (2018). These authors concluded that increases in soil moisture were associated to significant increases in CO₂ emissions in irrigated orchards. On the contrary, the lack of differences

Table 2

Cumulative values of soil CO₂, N₂O and CO₂ equivalent emissions released from alley soil in the monoculture and diversified treatments during the entire experimental period (11/04/2019 to 04-03-2021).

	Cumulative CO ₂ g m ⁻²	Cumulative N ₂ O	Cumulative CO ₂ e
Diversified system	3330 ± 318	-0.044 ± 0.199	3316 ± 369
Monoculture	2575 ± 64	0.329 ± 0.176	2673 ± 115
t-Student	2.292 ns	-1.446 ns	1.649 ns

ns: not significant (P > 0.05).

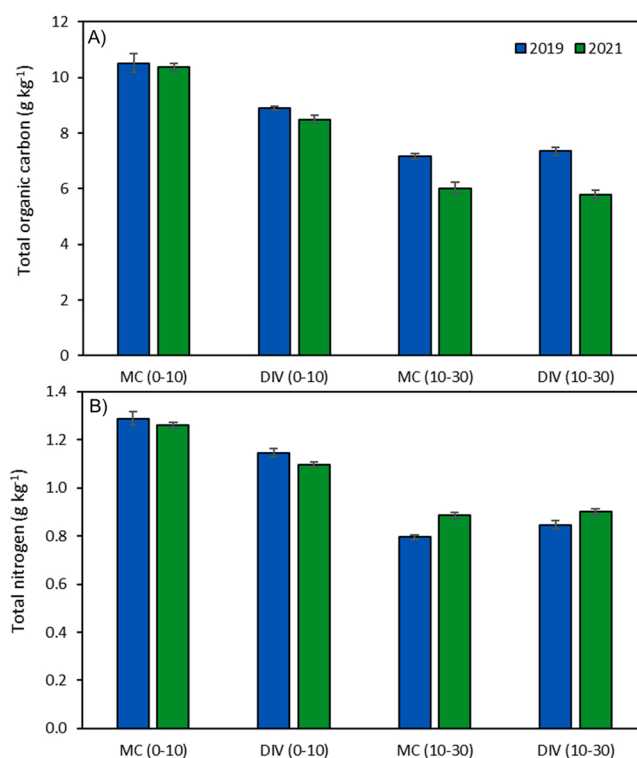


Fig. 2. Evolution of soil total organic carbon (A) and total nitrogen (B) from 2019 to 2021 in the two different systems at the two sampling depths (0–10 cm and 10–30 cm). Vertical bars indicate standard error. MC: monoculture; DIV: diversified orchard. (0–10) and (10–30) in the X file denotes soil depth. No significant differences between sampling dates for each treatment and depth.

in emissions between DIV and MC may indicate that current soil management in MC with frequent tillage and incorporation of pruning residues (Table S1) also contributes to high soil GHG emissions, as reported in literature (García-Orenes et al., 2010; Morell et al., 2010; Pramanik and Phukan, 2020). The lack of correlations between CO₂ emission and soil properties such as texture, pH, EC, TOC, Nt, nitrates and ammonium may suggest that climatic conditions and management are controlling CO₂ emissions by soil microbial activation through the presence of alley crops, tillage or pruning residues additions rather than inherent soil characteristics.

Contrarily to our initial hypothesis, alley cropping, incorporation of fava bean crop residues (with barley/vetch only roots remain in soil) and the likely activation of microbial communities did not lead to increases in TOC neither Nt in the DIV soil at the short term (2 years). This confirms the high organic matter mineralization rates under Mediterranean climate and the long time needed to achieve C storage in these soils, despite the high content of clay. In fact, alley soil in MC also received pruning residues, but this strategy has not contributed in two years' term to evidence an increase in C storage in the soil. This is contrary to a meta-analysis performed on Mediterranean orchards highlighting the positive effect of the implementation of alley cropping or cover crops, minimum or no tillage and on-farm organic amendments to increase TOC, with an average increase of 3.8 t C ha⁻¹ yr⁻¹ (Vicente-Vicente et al., 2016). However, these authors also indicated that Mediterranean sub-climates of low annual precipitation and high temperatures had very low values of soil C sequestration rate (< 1 t C ha⁻¹ yr⁻¹). This is the situation of our study site, with semiarid Mediterranean conditions and 18 °C of mean annual temperature. These results are in agreement with Martínez-Mena et al. (2013), who reported that TOC did not change after three years of implementation of reduced tillage and cover crops in an almond orchard from SE Spain. However, an increase in TOC was observed after 6 years of implementation of these practices compared to

the intensive tilled almond orchard in the same study site (Martínez-Mena et al., 2021), confirming that under semiarid conditions, C sequestration is a slow and long-term process.

With regard to soil N₂O emissions, it has been reported that the use of legumes can contribute to reduce N₂O emissions by higher efficiency in the use of N (Rochette and Janzen, 2005; Zhong et al., 2009). In fact, soil nitrate content was higher in the alley soil from MC compared to DIV (Supplementary Table S4). This may be due to its uptake by alley crops in DIV, which seems to contribute to decrease N₂O emissions owing to lower nitrate content in the soil, confirmed with the positive correlation between cumulative N₂O emissions and soil nitrate content. This fact may also contribute to sequester mobile N forms by biomass that otherwise can leach and pollute groundwater. In this line, alley cropping and cover crops have been proposed to foster nutrient cycling and avoid nutrient losses by leaching (Bergeron et al., 2011; Wolz et al., 2018).

Hence, results confirm that most variation in GHG emission rates was provided by climatic conditions, being soil moisture the main factor controlling emissions, followed by soil temperature. With this regard, Zornoza et al. (2018, 2016) reported that CO₂ emissions were more influenced by temperature than by moisture when soil water was not limited in Mediterranean irrigated orchards; however, when water content became a limiting factor, CO₂ emissions were more related to soil moisture than to temperature. Thus, since soil heterotrophic activity is regulated by soil water content, CO₂ emissions followed the soil water trend, confirming the water deficit in this system. Thus, the higher the soil water content, the higher soil microbial activity and so, the CO₂ emissions (Diao et al., 2022; Song et al., 2018). Higher differences among MC and DIV with regard to soil moisture were found with fava bean crop, mostly during the last cycle 2020, associated with the highest differences in CO₂ emissions between MC and DIV. This should be associated to higher microbial activity with high soil moisture. Thereby, although as a general pattern, the DIV system did not significantly contribute to increase CO₂ emissions compared to MC, when fava bean grew, CO₂ emissions increased likely due to high soil moisture that enhanced rhizodeposition that stimulated microbial activity.

Alley cropping, when citrus production was high, counted only 15 % of the 2019 annual production, confirming that main revenues will come from the cash crop (mandarin). However, alley cropping reached 50 % of the 2020 annual production owing to a high incidence of a fungal disease that affected mandarin fructification. This confirms the benefits of alley cropping not only from an environmental point of view, but also as an economic benefit since diversified production can make farmers more resilient to negative effects of pests/diseases on one crop. Accordingly, Kurdyś-Kujawska et al. (2021) proved higher economic efficiency in diversified farms compared to monocultural farms in Poland. John et al. (2021) demonstrated that diversifying maize with legumes increased yield stability, nutritional gains, and profitability in central Africa. Di Falco and Zoupanidou (2016) also evidenced the important role of agrobiodiversity in the resilience of agroecosystems, proposing crop diversification as a strategy to support productivity and farmers revenues when soils are less fertile and monocultures cannot provide high yields.

5. Conclusion

Alley cropping with barley/vetch and fava bean contributed to increase land productivity in a mandarin orchard from the Mediterranean region. Alley cropping did not contribute to higher overall GHG emissions, with no differences between bare and plant covered alleys. Soil CO₂ was regulated by soil moisture in the area, with higher emissions during those days with higher soil moisture owing to rainfall events or irrigation in the diversified system. Furthermore, management (alley crops, tillage, incorporation of pruning residues) had higher influence on CO₂ emissions than inherent soil characteristics. The growth of fava bean and barley/vetch seems to contribute to decrease N₂O emissions. After two years, the addition of crop residues in the alley soil was not

enough to increase soil organic C content in the soil, confirming the trend to organic matter mineralization under these Mediterranean climatic conditions. More time is so needed to demonstrate if this practice is actually efficient for soil C sequestration and storage in irrigated orchards from the Mediterranean basin. These findings can encourage farmers, land managers and decision-makers to implement and foster the adoption of alley cropping as a sustainable practice to enhance land production with no negative effect on the environment, with potential positive impact long-term. However, long-term monitoring programs are needed to provide robust data about the impact of alley cropping on the delivery of ecosystem services, since changes in soil are a long process.

CRedit authorship contribution statement

Virginia Sánchez-Navarro: Conceptualization, Investigation, Writing – original draft. **Silvia Martínez-Martínez:** Conceptualization, Investigation, Writing – review & editing. **Jose A. Acosta:** Conceptualization, Investigation, Writing – review & editing. **María Almagro:** Investigation, Writing – review & editing. **María Martínez-Mena:** Conceptualization, Investigation, Writing – review & editing. **Carolina Boix-Fayos:** Conceptualization, Investigation, Writing – review & editing. **Elvira Díaz-Pereira:** Conceptualization, Investigation, Writing – review & editing. **Abdelmalek Temnani:** Investigation. **Pablo Berrios:** Investigation. **Alejandro Pérez-Pastor:** Conceptualization, Writing – review & editing. **Raúl Zornoza:** Conceptualization, Investigation, Writing – original draft, Data curation, Project administration, Funding acquisition.

Declaration of Competing Interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

Data Availability

Data will be made available on request.

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Appendix A. Supporting information

Supplementary data associated with this article can be found in the online version at doi:10.1016/j.eja.2022.126684.

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