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1 **Simultaneous removal of malachite green and hexavalent chromium by**
2 ***Cunninghamella elegans* biofilm in a semi-continuous system**

3

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14 **Abstract**

15 The present study was conducted to evaluate the potential of the fungus *Cunninghamella*
16 *elegans* for simultaneous decolourisation of a triphenylmethane dye malachite green (MG)
17 and hexavalent chromium [Cr(VI)] in the same media. This fungus can degrade MG through
18 its reduction into leucomalachite green and then demethylation followed by oxidative
19 cleavage. Along with MG degradation, *C. elegans* biofilm could effectively and repeatedly
20 remove Cr(VI) from the liquid cultures even in the presence of high concentrations (40 g L⁻¹)
21 of NaCl and various other metal ions. *C. elegans* biofilm was also found to adsorb different
22 dyes (reactive black-5, acid orange 7, direct red 81 and brilliant blue G) concurrently with
23 Cr(VI). Based on its potential for simultaneous removal of dyes and Cr(VI) as well as
24 reusability, *C. elegans* biofilm is envisaged as an efficient bioresource to devise strategies for
25 treatment of wastewaters loaded with multiple pollutants.

26

27 **Keywords:** Immobilization; fungus; dye decolorisation; textile wastewater; adsorption

28

29 **1. Introduction**

30 Dyes are a common constituent of wastewaters originating from various industrial processes.
31 Malachite green (MG) is a triphenylmethane cationic dye which is used in textile, leather,
32 medical, food and paper industries in addition to its use as a biocide to control protozoan and
33 fungal infections in fish farming (Culp and Beland, 1996; Srivastava et al., 2004). However,
34 discharge of MG-loaded wastewaters into the environment reduces light penetration in the
35 water bodies and affects the living organisms present owing to the carcinogenic, mutagenic
36 and teratogenic properties of MG and its metabolites (Culp and Beland, 1996; Srivastava et
37 al., 2004; Donya et al., 2012). For example, MG is toxic to mammalian cells and has been
38 shown to cause cancer in different organs including liver and thyroid of experimental animals
39 (Rao, 1995; Srivastava et al., 2004; Donya et al., 2012). Leucomalachite green, which is a
40 major metabolite arising from the reduction of malachite green, is also of particular concern
41 owing to its toxicity, mutagenicity and its relatively higher lipophilicity, which result in it
42 being retained in fish muscle and fat (Bilandzic et al., 2012). Despite the fact that MG has
43 been banned in some countries it is still being used in others owing to its low cost, ready
44 availability and high efficacy. In addition to dyes, wastewaters originating from different
45 industries, including textile and leather, have also been found to contain considerable
46 amounts of different salts and metal ions (Tuzen et al., 2008; Ngah and Hanafiah, 2008). The
47 latter are present either from the use of metal complex dyes or metal-containing salts as
48 mordant for better fixation of dyes. Among the metal ions, hexavalent chromium [Cr(VI)] is a
49 common pollutant which co-exists with dyes in the wastewaters originating from textile and
50 leather industries (Desai et al., 2009). It is not only the second most common inorganic
51 contaminant of ground water and hazardous waste sites but also listed by the United States
52 Environmental Protection Agency among the 17 chemicals for posing the greatest threat to
53 human health (Horton et al., 2006; Cheung and Gu, 2007; Quintelas et al., 2008). In addition
54 to disruption of biochemical and physiological functions in bio-systems owing to its strong
55 oxidizing nature, high solubility in water and rapid permeability, it has also been reported to
56 harbor mutagenic, carcinogenic and teratogenic properties (McLean and Beveridge, 2001;
57 Ilias et al., 2011). Hence, the co-existence of Cr(VI) and synthetic dyes, including malachite
58 green, in wastewaters is a matter of serious concern and there is a need to find effective,
59 innovative and economic treatment technologies to eliminate them or minimize their quantity
60 in the environment.

61 Exploitation of microorganisms for bioremediation of contaminated environments has
62 attracted attention as a cost-effective and environmentally friendly approach. Several

63 researchers have isolated and characterized various bacterial and fungal strains for removal
64 and detoxification of chromium in soil and water resources (Prigione et al., 2008; Dhal et al.,
65 2010; Ilias et al., 2011; Essahale et al., 2012; Maqbool et al., 2015). Similarly, a number of
66 bacterial strains belonging to different genera have been isolated and characterized for
67 decolourisation of MG (Li et al., 2009; Kalyani et al., 2012). The potential for
68 decolourisation and degradation of this dye has also been reported in various fungi including
69 *Phanerochaete chrysosporium*, *Cyathus bulleri*, *Cyathus stercoreus*, *Cyathus striatus*, and
70 *Penicillium ochrochloron* (Vasdev et al., 1995; Jadhav and Govindwar, 2006; Shedbalkar and
71 Jadhav, 2011; Jasinska et al., 2012). The non-lignolytic fungus *Cunninghamella elegans* is
72 well known for its ability to transform a broad range of xenobiotics (Murphy, 2015) and the
73 inactivated biomass of the fungus is an effective biosorbent (Tigini et al., 2010). Cha et al.
74 (2001) observed the formation of leucomalachite green, *N*-demethylated and *N*-oxidized
75 metabolites upon incubation of *C. elegans* with the MG. Microsomal fractions also catalysed
76 the production of leucomalachite green and *N*-demethylated metabolites, and the
77 biotransformation was inhibited by 1-aminobenzotriazole, metyrapone and SKF 525-A, thus
78 it was reasoned that the reduction and *N*-demethylation reactions were catalysed by
79 cytochrome P450. Kim et al. (2010) purified a cytochrome c, CeCyt, from the mitochondria
80 of *C. elegans*, that catalysed the decolourisation of malachite green and suggested that the
81 protein functions to reduce malachite green under conditions of oxidative stress.

82 Whilst there are some reports on the simultaneous removal of different dyes and
83 Cr(VI) from synthetic textile wastewaters by using some multifunctional bacterial strains
84 (Desai et al., 2009; Mahmood et al., 2013; Anwar et al., 2014; Maqbool et al., 2016), to the
85 best of our knowledge, simultaneous microbial removal of MG and Cr(VI) has not yet been
86 the focus of any study. Moreover, there is no report regarding the application of fungal strains
87 for such simultaneous removal of dyes and metal ions. In this context, the present study has
88 been conducted for simultaneous removal of MG and Cr(VI) by using *C. elegans*. Biofilms
89 of this fungus have already been reported to demonstrate improved biotransformation of
90 drugs and xenobiotics compared with suspended cells (Amadio et al., 2013; Mitra et al.,
91 2013; Quinn et al., 2015). The aim of this study is to extend the possible application of the
92 fungal biofilm to the bioremediation of dye/metal contaminated wastewater.

93

94 **2. Materials and Methods**

95 **2.1 Dyes**

96 Malachite green (technical grade) was acquired from BDH (Poole, UK), reactive black-5 and
97 direct red-81 were purchased from Santa Cruz Biotechnology (Dallas, TX, USA), and acid
98 orange-7 ($\geq 85\%$) and brilliant blue G were obtained from Sigma Aldrich (Arklow, Ireland).
99

100 **2.2. Cultivation of *C. elegans* biofilm and planktonic cells**

101 *Cunninghamella elegans* DSM 1908 was grown on sabouraud glucose agar for 120 h at 28
102 °C. Inoculum was prepared by homogenizing one plate of agar and mycelia in 100 mL of
103 0.8% autoclaved saline. The planktonic cell cultures were grown in 250 mL Erlenmeyer
104 flasks containing 45 mL of sterilised sabouraud dextrose broth and 5 mL of *C. elegans*
105 homogenate. For cultivating biofilms the method described by Amadio et al. (2013) was
106 followed. For biofilm cultivation, stainless steel compression springs (1.2 mm, T316 wire,
107 Shannon Coiled Springs, Ireland) were placed at the bottom of 250 mL Erlenmeyer flasks
108 containing sterilized sabouraud dextrose broth (49 mL). The springs were kept completely in
109 contact with the inner walls of the flasks for optimum biofilm growth. Each flask was
110 inoculated with *C. elegans* homogenate (1 mL) and incubated for 72 h with rotary agitation
111 (150 rpm) at 28 °C.
112

113 **2.3. Decolourisation of MG by *C. elegans* biofilm and planktonic cells**

114 After 72 h of biofilm growth, the medium in the flasks was replaced with 50 mL sterile MG
115 aqueous solution (80 μM or 29 mg L^{-1} , unless stated) and incubated with shaking (150 rpm)
116 at 28 °C, alongside an un-inoculated control. When the MG was degraded and both the
117 supernatant and biofilm had been decolourized, fresh MG was added to the flasks. For
118 estimation of decolourisation by planktonic cultures, the cells were harvested by centrifuging
119 (3500 rpm for 15 min) and the biomass was re-suspended in 50 mL of the aqueous MG
120 solution and incubated as before; for the biofilm cultures, the supernatant was decanted and
121 replaced. The supernatants (1.5 mL) and biomass (200 mg) of biofilm and planktonic cells
122 were collected aseptically at regular intervals. Malachite green decolourisation in the
123 supernatant was determined spectrophotometrically as previously described (Jasinska et al.,
124 2012) by measuring the change in absorbance at 617 nm (λ_{max}). The biomass was immersed
125 in 1 mL of methanol and shaken vigorously for 30 s. The methanol extract was then used to
126 determine malachite green decolourisation (Nanodrop 1000). To recycle biofilms, the
127 supernatants were decanted directly and fresh aqueous dye solution was added. Planktonic
128 cells were harvested by centrifuging and the biomass re-suspended in 50 mL of fresh dye.

129 The biofilms and the planktonic cells were both rejuvenated by replacing the supernatants
130 with 50 mL of fresh sabouraud dextrose broth and incubated for up to 16 h.

131

132 **2.3.1. Effect of pH on MG decolourisation by *C. elegans* biofilm**

133 To determine the effect of pH on MG decolourisation, biofilm was cultivated as described
134 and incubated for 48 h with dye dissolved in water (50 mL). The supernatant was decanted
135 and replaced by 50 mL of 20 mM phosphate buffer (pH 3, 5-7) or 2-(*N*-morpholino)
136 ethanesulfonic acid (pH 4) containing malachite green; the pH experiments were conducted
137 with the same biofilm, starting with pH 7 and ending at pH 3.

138

139 **2.3.2. Metabolite identification**

140 The degradation products of malachite green were extracted from the biomass by incubating
141 it with 50 mL of ethyl acetate for 3 hours. The organic layer was evaporated to dryness, the
142 residue redissolved in 1 mL ethyl acetate and analysed by gas chromatography-mass
143 spectrometry (GC-MS) using a method similar to that described by Du et al. (2011). Samples
144 (1 μ L) of the extract were injected in the splitless mode onto a HP5MS column (30 m \times 0.25
145 mm \times 0.25 μ m). The oven temperature held at 120 $^{\circ}$ C for 2 min and then increased to 300 $^{\circ}$ C
146 at 10 $^{\circ}$ C min⁻¹. The metabolites were identified by retention time and mass spectra.

147

148 **2.4. Simultaneous removal of dye and Cr(VI) by *C. elegans* biofilm and planktonic cells**

149 *C. elegans* biofilm and planktonic cells were prepared as described in section 2.1 and
150 incubated under the standard conditions with either MG (80 μ M) or Cr(VI) (20 mg L⁻¹) only,
151 or a combination of both dye and metal. Decolourisation was monitored
152 spectrophotometrically and Cr(VI) removal was assessed following the diphenyl carbazide
153 (DPC) method described by Maqbool et al. (2016). In order to test the reusability of the
154 biofilms for MG decolourisation and/or Cr(VI) removal, the aqueous solutions of dye and/or
155 metal were continuously replaced with the fresh solutions after >95% of the pollutants were
156 eliminated from the supernatant. The biofilms were rejuvenated after every three cycles of
157 decolourisation by replacing the supernatants with 50 mL of fresh sabouraud dextrose broth
158 and incubating for up to 16 h under shaking (150 rpm) at 28 $^{\circ}$ C.

159 To evaluate the ability of *C. elegans* biofilms to decolourise other dyes (reactive
160 black-5, acid orange-7, direct red-81 & brilliant blue G) concurrently with Cr(VI) removal,
161 triplicate biofilms were separately incubated with aqueous solutions containing 20 mg L⁻¹ of
162 Cr(VI) and 50 mg L⁻¹ one of the selected dyes. Triplicate un-inoculated controls were also

163 incubated for each treatment. Decolourization of the dyes in the supernatant was monitored
164 spectrophotometrically at 597 nm (reactive black-5), 485 nm (acid orange-7), 540 nm (direct
165 red-81) and 595 nm (brilliant blue G).

166

167 **2.3.1. Impact of initial Cr(VI) concentration on removal efficiency**

168 Triplicate biofilms were incubated with aqueous solutions of MG (80 μ M) and varying
169 concentrations (20 mg L⁻¹, 40 mg L⁻¹, 60 mg L⁻¹, 80 mg L⁻¹, 100 mg L⁻¹, 150 mg L⁻¹) of
170 Cr(VI). Triplicate un-inoculated flasks for each treatment were also incubated as controls.
171 The decrease of both pollutants in the supernatants was measured as described previously.
172 The re-usability of biofilms following rejuvenations at varying initial Cr(VI) concentrations
173 was also evaluated.

174

175 **2.4.2. Impact of NaCl and metal ions on simultaneous removal of MG and Cr(VI) by *C.*** 176 ***elegans* biofilm**

177 Biofilms were incubated with aqueous solutions containing MG (80 μ M) and Cr(VI) (20 mg
178 L⁻¹) plus varying concentrations (up to 100 g L⁻¹) of NaCl. The presence of 20 mg L⁻¹ various
179 metal ions (Ag⁺, Cu²⁺, Zn²⁺, Mn²⁺, Ni²⁺, Ba²⁺, Fe³⁺) on simultaneous removal of both
180 pollutants was similarly investigated. Triplicate un-inoculated flasks for each treatment were
181 also incubated as controls.

182

183 **2.5 Statistical analysis**

184 The results are presented as means \pm standard deviation. The means were compared using
185 Least Significance Difference (LSD) test after the analysis of variance (ANOVA) at $p \leq 0.01$
186 using the software R (3.4.1).

187

188 **3. Results**

189 **3.1. MG decolourisation by *C. elegans***

190 MG was decolourised by *C. elegans* biofilm and planktonic cultures. Over the first 6 h
191 incubation in the biofilm culture, the colour in the supernatant had decreased by
192 approximately 60 % and, after 24 h incubation, almost a complete (> 95%) removal of colour
193 was observed in the supernatant (data not shown). Upon the second addition of dye to the
194 flasks the supernatant and biomass were monitored spectrophotometrically at different time
195 points (Fig 1). The dye was removed from the supernatant within 15 min by absorption to the
196 biomass (Fig 1A); the colour in the biomass dissipated more slowly (Fig 1 B and C). The

197 biomass of both the biofilm and planktonic cultures gave almost a similar pattern of colour
198 removal from the supernatant and absorbance over the incubation period. Thus, rapid initial
199 decolourisation of the supernatant through biosorption was followed by a slower
200 biodegradation of the MG dye.

201 In order to study the impact of decreasing pH on decolourisation of MG by *C. elegans*
202 biofilm, the decolourisation experiments were carried out at pH from 7 to 3. The cultures
203 incubated at pH values from 4 to 7 were found to decolourise more than 95% of the initially
204 added MG in the supernatants within the first 24 hours. However, only 80% decolourisation
205 of the supernatant was observed in the same period with cultures at pH 3 (Supplemental
206 Information). Furthermore, the time for complete decolourisation (i.e. supernatant and
207 biomass) of MG by the cultures increased as pH was lowered.

208

209 **3.2. Assessment of biodegradation of MG by *C. elegans***

210 The biomass from biofilm cultures incubated with 80 μ M malachite green was extracted with
211 ethyl acetate and the extractable metabolites were analyzed by GC-MS. The GC-MS analysis
212 revealed the presence of leucomalachite green, *N*-demethylated metabolites, 4-
213 (dimethylamino) benzophenone and aminobenzophenone (Table 1). The presence of these
214 metabolites suggests a stepwise demethylation followed by oxidative cleavage as previously
215 suggested by Cha et al. (2001). Interestingly, upon subsequent dye addition to the biofilm, no
216 metabolites were detectable by GC-MS after 24 h incubation, indicating complete
217 biodegradation.

218

219 **3.3. Semi-continuous biofilm-catalyzed simultaneous removal of MG and Cr(VI)**

220 **3.3.1. Simultaneous removal of MG and Cr(VI) by *C. elegans* biofilm and planktonic 221 cells**

222 *C. elegans* biofilm and planktonic cultures were tested for their potential not only to remove
223 MG and Cr(VI) individually but also for simultaneous removal of MG and Cr(VI) in the
224 same solution. The data are summarized in Table 2 and show that MG removal was
225 comparable in planktonic and biofilm cultures whether in the absence or presence of Cr (VI),
226 with approx. 80 % decolourisation within 16 h and complete degradation within 22-26 h.
227 Planktonic cultures were more effective at Cr(VI) removal than biofilm, with 83 % removed
228 in 16 h compared to 71 %, and a shorter time required for complete removal (22 h compared
229 with 24 h). However, whereas the efficiency of biofilm was not significantly impacted with

230 the combination of dye and metal, the removal of Cr (VI) in planktonic cultures after 16 h
231 decreased noticeably compared with the cultures incubated with the metal only.

232 One of the main potential advantages of employing the biofilm is the ease of re-
233 usability, which was demonstrated in these experiments, showing that complete (> 95 %)
234 removal of dye and metal was possible for at least 19 repeated additions (Table 2).
235 Planktonic cultures are more difficult to recycle, as a centrifugation (or filtration) step is
236 necessary, and the suspended cells have previously shown to cease functioning after approx.
237 three cycles (Amadio et al., 2013).

238

239 **3.3.2. Impact of initial Cr(VI) concentration on biofilm efficiency**

240 Varying initial concentrations of Cr(VI) had an impact on the simultaneous removal of MG
241 and Cr(VI) by the *C. elegans* biofilm (Figure 2). After 16 h incubation, over 90% of the
242 initially added MG was decolourized in the solutions containing Cr (VI) concentrations up to
243 60 mg L⁻¹; however, higher concentrations of the metal resulted in a decrease in
244 decolourisation ability. The total amount of Cr (VI) removed within the same period in these
245 experiments increased from 0.91 mg, when an initial concentration of 20 mg L⁻¹ was used, up
246 to 1.95 mg when the initial Cr (VI) concentration was 60 mg L⁻¹. At higher concentrations
247 the removal progressively declines. Notably, increasing the initial concentrations of Cr(VI)
248 also resulted in an increase in the time required for complete (>95%) simultaneous removal
249 of Cr(VI) and MG, and a decrease in number of cycles of complete simultaneous removal of
250 both the pollutants (Table 3).

251

252 **3.3.3. Effect of NaCl and other metals on biofilm efficiency**

253 Simultaneous removal of MG and Cr(VI) by the *C. elegans* biofilm was not substantially
254 affected by NaCl concentrations of 20 g L⁻¹ (Fig 3 and Table 3); however, at higher
255 concentrations the removal efficiency after 16 h, the time required for complete removal and
256 the number of cycles of complete dye/metal removal were all affected. Fig 4 shows the effect
257 of a selection of metal ions (2 mM) on the simultaneous removal of dye and Cr(VI). Most of
258 the metals tested inhibited the removal of both pollutants to some degree, although complete
259 (>95 %) removal was still achieved within 40 h in all experiments.

260

261 **3.4. Simultaneous removal of Cr(VI) and other dyes by *C. elegans* biofilm**

262 The ability of *C. elegans* biofilm for simultaneous removal of Cr(VI) and other dyes
263 (Reactive Black 5, Acid Orange 7, Direct Red 81 and Brilliant Blue G) was also examined.

264 This biofilm showed a good potential for parallel removal of Cr(VI) and different dyes from
265 the culture supernatant (Table 4). After 40 h incubation, a complete (>95%) removal of
266 Cr(VI) was observed along with at least 85 % simultaneous removal of the initially added
267 dye. The dyes were biosorbed by the biofilms, but, unlike MG, the biomass was not
268 completely decolourized, even after 120 hours incubation.

269

270 **4. Discussion**

271 Environmental pollution due to synthetic textile dyes is one of the leading contributors in
272 degradation of natural resource. This negative impact of synthetic dyes is intensified when
273 these dyes loaded effluents are also accompanied by the presence of different pollutants.
274 Hexavalent chromium [Cr(VI)] is one of such pollutants which has often been found to co-
275 exist as a contaminant with synthetic dyes in textile and tanneries effluents. Hence, there is
276 need to devise the strategies for concurrent removal of such co-existing pollutants and the
277 present study was conducted to evaluate the potential of *C. elegans* biofilm for simultaneous
278 removal of a synthetic dye, malachite green (MG), and Cr(VI) in a semi-continuous system.

279 The decolourisation of MG in planktonic and biofilm cultures occurred following a
280 similar pattern, with the dye rapidly adsorbed by the biomass, followed by a slower
281 biodegradation step, resulting in complete removal of 80 µM dye in 24 h. This pattern of
282 decolourisation with initial biosorption followed by degradation has also been observed in
283 other studies that focused on fungal biodegradation of MG (Jadhav and Govindwar, 2006;
284 Jasinska et al., 2012). It was observed here that the time taken for dye decolourisation by *C.*
285 *elegans* biofilm upon initial addition of MG was longer compared to the subsequent rounds,
286 in contrast to planktonic cultures. One possible reason for this difference is that in the
287 biofilm there are specific genes required for decolourisation that are induced upon dye
288 addition, but in planktonic cells the genes are already expressed. Transcriptomic and
289 proteomic analyses of other fungi demonstrate that expression of genes can vary between
290 planktonic and biofilm cultures (Gutierrez-Correa et al., 2012).

291 In general biofilms are stable and active over long periods (Halan et al., 2012) and *C.*
292 *elegans* biofilms have been shown to be conveniently reused for biotransformations (Amadio
293 et al., 2013; Quinn et al., 2015), thus they have potential for application in continuous or
294 semi-continuous processes. The *C. elegans* biofilm can decolourize MG over a range of
295 acidic pH, which is an added advantage; however, at pH 3 decolourisation ability is
296 compromised. This is comparable with the decolourisation activity of some other fungal

297 strain including *Penicillium ochrochloron*, which is completely inhibited at pH 3 (Shedbalkar
298 and Jadhav, 2011).

299 Leucomalachite green, demethylated leucomalachite green, 4-(dimethylamino)
300 benzophenone and aminobenzophenone were observed as intermediate metabolites during
301 initial decolourisation of MG by *C. elegans* biofilm. Cha et al. (2001) identified mono-, di-,
302 and tri-demethylated derivatives of malachite green and leucomalachite green after
303 decolourisation by suspended *C. elegans* ATCC 36112. The other metabolites detected in the
304 present study had not previously been identified from the fungus, but are known
305 intermediates in the biodegradation of malachite green in other microorganisms. For
306 example, 4-(dimethylamino) benzophenone and 4-aminobenzophenone were observed during
307 degradation of MG by *Micrococcus* sp. strain BD15 (Du et al., 2013). 4-(Dimethylamino)
308 benzophenone has also been detected during decolourisation of MG by *Shewanella*
309 *decolourationis* NTOU1 under anaerobic conditions (Chen et al., 2010).

310 *C. elegans* biofilm as well as planktonic cultures were shown here for the first time to
311 simultaneously remove MG and Cr(VI) from contaminated water. Although bacterial
312 cultures have been reported to concurrently remove Cr(VI) and different azo-dyes (Maqbool
313 et al., 2016; Anwar et al., 2014; Mahmood et al., 2013), to the best of our knowledge, there is
314 no report of the simultaneous removal of MG and Cr(VI) by any single microbial strain.
315 Furthermore, the biofilm can tolerate the presence of Cr(VI) up to 60 mg L⁻¹, and up to 40 g
316 L⁻¹ NaCl, but is sensitive to higher concentrations of both, resulting in longer times for
317 complete removal of dye/metal and a reduction in the number of times the biofilm can be re-
318 used. Simultaneous removal of dye/metal by the biofilm is possible even in the presence of
319 metal ions, such as silver and copper, albeit at a slower rate.

320 There are numerous reports of immobilized fungi applied to the decolourisation of
321 dye-contaminated water (Couto, 2009), but only a handful of these concern MG, and none
322 that also involve Cr(VI) removal. Barapatre et al. (2017) reported MG decolourisation in
323 *Aspergillus flavus* and demonstrated that immobilization on a number of inert materials, such
324 as polyurethane foam and clay brick, resulted in improved decolourisation compared with
325 suspended culture. However, no experiments to investigate the recycling of the immobilized
326 fungus were done. In the present study, *C. elegans* was immobilized as a biofilm, which
327 enabled repeated use (at least 19 cycles of dye/metal removal), which is attractive for
328 bioremediation applications. Furthermore, a screen of other dyes demonstrated that the
329 biofilm could biosorb these also, thus expanding its potential for remediation of dye-
330 contaminated water.

331

332 **Conclusion**

333 Based on the findings of this study, it can be concluded *C. elegans* biofilm might serve as
334 potential bioresource to devise the strategies for simultaneous removal of Cr(VI) and MG
335 even in the presence of NaCl and metal ions that are characteristically present in real textile
336 and tanneries effluents. Confirmation of the re-usability of this biofilm is an important feature
337 for its potential use in wastewater treatment processes, which require continuous operation.

338

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341 Protection Agency (LQ).

342

343 **Conflict of interest**

344 None

345

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467 **Figure legends**

468 Fig 1. The decolourisation malachite green in *C. elegans*. **(A)** Absorbance (617 nm) of
469 supernatants (S/N) and biomass (BM) after 15 min incubation with the fungus. Error bars
470 represent standard deviation n=2. **(B)** Absorbance spectra of methanolic extracts of
471 planktonic biomass. **(C)** Absorbance spectra of methanolic extracts of biofilm biomass. The
472 slightly lower absorbance in biofilm reflects the effectiveness of the extraction method using
473 methanol

474 Fig 2. The effect of initial Cr (VI) concentration on the simultaneous removal of MG and Cr
475 (VI) after 16 h incubation with the fungus. The means of MG removal compared using Least
476 Significance Difference (LSD) test after the analysis of variance (ANOVA) at $p \leq 0.01$ (LSD
477 value=17.43). The mean values labelled by the same letter(s) are not significantly different.

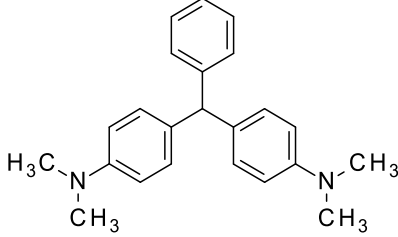
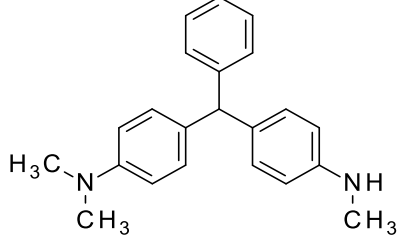
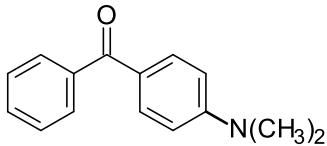
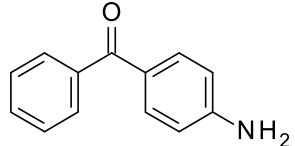
478 Fig 3. The effect of NaCl concentration on simultaneous removal of MG and Cr (VI). The
479 means of MG removal and Cr(VI) removal compared using Least Significance Difference
480 (LSD) test after the analysis of variance (ANOVA) at $p \leq 0.01$ (LSD value for MG
481 removal=13.04, LSD value for Cr(VI) removal=17.42). The mean values within either
482 response (MG removal or Cr(VI) removal) labelled by the same letter(s) are not significantly
483 different.

484 Fig 4. The effect of metal ions on the removal of MG **(A)** and Cr (VI) **(B)** by *C. elegans*. The
485 means of MG removal and Cr(VI) removal at varying time intervals compared using Least
486 Significance Difference (LSD) test after the analysis of variance (ANOVA) at $p \leq 0.01$ (LSD
487 value for MG removal after 8 h= 4.57, LSD value for MG removal after 24 h= 5.73, LSD
488 value for Cr(VI) removal after 8 h= 5.01). The mean values within either response (MG
489 removal or Cr(VI) removal) at a specific time labelled by the same letter(s) are not
490 significantly different. The unlabelled mean values for MG removal over 40 h and Cr(VI)
491 removal over 24 hours were found statistically non-significantly different among themselves.

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494 Table 1. GC-MS data for the metabolites of malachite green incubated with *Cunninghamella*
 495 *elegans* biofilm

| Intermediate products | Molecular structure | T _R (min) | m/z of M ⁺ |
|-----------------------------------|---|----------------------|-----------------------|
| Leucomalachite green |  | 14.65 | 330 |
| Desmethyl Leucomalachite green |  | 14.29 | 316 |
| 4-(Dimethylamino) benzophenone |  | 15.39 | 225 |
| 4-Aminobenzophenone |  | 14.16 | 197 |

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498 Table 2. Removal of Cr(VI) and malachite green (MG) by the suspended cells (planktonic)
 499 and biofilm of *Cunninghamella elegans*. > 95 % Decrease in dye/metal is considered
 500 complete removal.

| Culture condition | | Cr (VI) Removal | MG Removal | Simultaneous removal | |
|-------------------|-----------------------------------|-----------------|------------|----------------------|----------|
| | | | | Cr (VI) | MG |
| Planktonic | % Removal after 16 hours | 82.9±5.9 | 81.6±4.4 | 74.3±3.4 | 79.5±6.1 |
| | Time for complete removal (h) | 22 | 22 | 24 | 24 |
| Biofilm | % Removal after 16 hours | 71.4±3.5 | 82.5±6.2 | 69.2±2.9 | 80.5±4.6 |
| | Time for complete removal (h) | 24 | 22 | 26 | 26 |
| | No. of cycles of complete removal | 24 | 19 | 23 | 20 |

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Table 3. Effect of NaCl and initial Cr(VI) concentrations on simultaneous removal of Cr(VI) and malachite green (MG) by *Cunninghamella elegans* biofilm. > 95 % Decrease in dye/metal is considered complete removal.

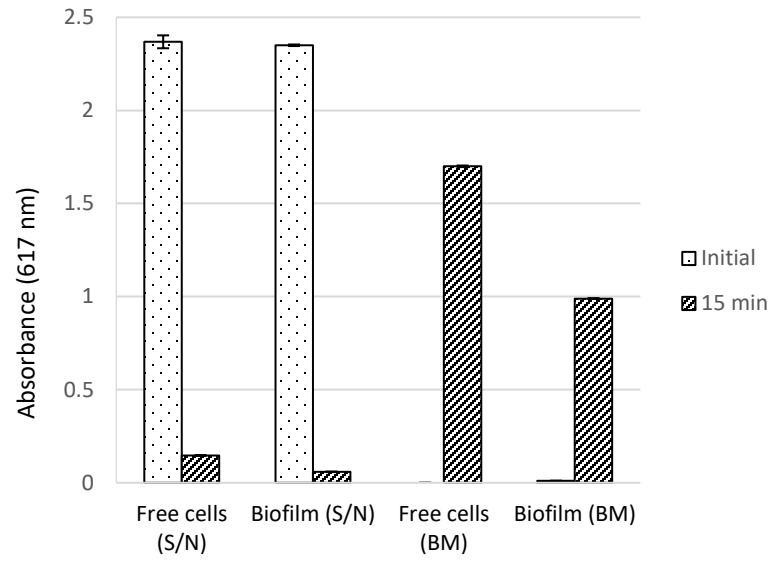
| NaCl concentration | No NaCl | | 20 g L⁻¹ | | 40 g L⁻¹ | | 60 g L⁻¹ | | 80 g L⁻¹ | | 100 g L⁻¹ | |
|--|-----------------------------|---------------|-----------------------------|---------------|-----------------------------|---------------|-----------------------------|---------------|------------------------------|---------------|------------------------------|---------------|
| | MG | Cr(VI) | MG | Cr(VI) | MG | Cr(VI) | MG | Cr(VI) | MG | Cr(VI) | MG | Cr(VI) |
| Time for complete removal (h) | 24 | 24 | 24 | 24 | 28 | 32 | 32 | 54 | 96 | 144 | >144 | >144 |
| No. of cycles of complete removal | >10 | >10 | >10 | >10 | 6 | 4 | 3 | 2 | 1 | 1 | 0 | 0 |
| | | | | | | | | | | | | |
| Initial Cr(VI) concentration | 20 mg L⁻¹ | | 40 mg L⁻¹ | | 60 mg L⁻¹ | | 80 mg L⁻¹ | | 100 mg L⁻¹ | | 150 mg L⁻¹ | |
| | MG | Cr(VI) | MG | Cr(VI) | MG | Cr(VI) | MG | Cr(VI) | MG | Cr(VI) | MG | Cr(VI) |
| Time for complete removal (h) | 24 | 24 | 24 | 32 | 28 | 52 | 48 | 72 | 76 | >144 | >144 | >144 |
| No. of cycles of complete removal | >10 | >10 | 6 | 5 | 2 | 2 | 1 | 1 | 1 | 0 | 0 | 0 |

Table 4. Simultaneous removal of Cr(VI) and various dyes by *Cunninghamella elegans* biofilm.

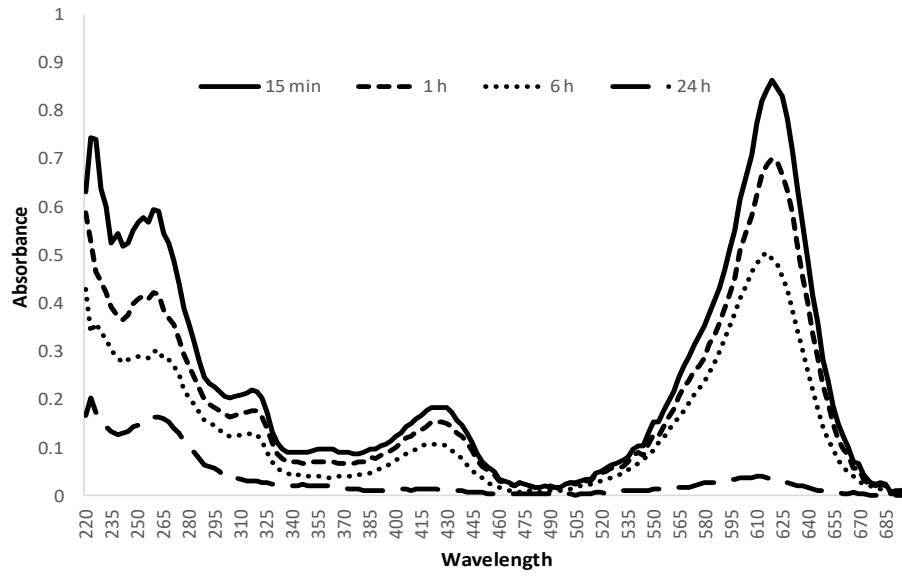
| | Colour Removal (%) | | Cr(VI) Removal (%) | |
|-------------------------|--------------------|------------|--------------------|------------|
| | 24 h | 40 h | 24 h | 40 h |
| Reactive Black 5 | 38.6 ± 3.6 | 90.7 ± 3.1 | 85.3 ± 4.5 | 98.6 ± 2.1 |
| Acid Orange 7 | 51.9 ± 3.1 | 92.6 ± 3.9 | 82.1 ± 3.4 | 96.6 ± 2.6 |
| Direct Red 81 | 72.7 ± 1.4 | 96.3 ± 2.6 | 89.1 ± 2.4 | 96.7 ± 3.1 |
| Brilliant Blue G | 58.9 ± 2.3 | 86.6 ± 3.9 | 90.1 ± 1.9 | 96.6 ± 2.8 |

Fig 1.

A



B



C

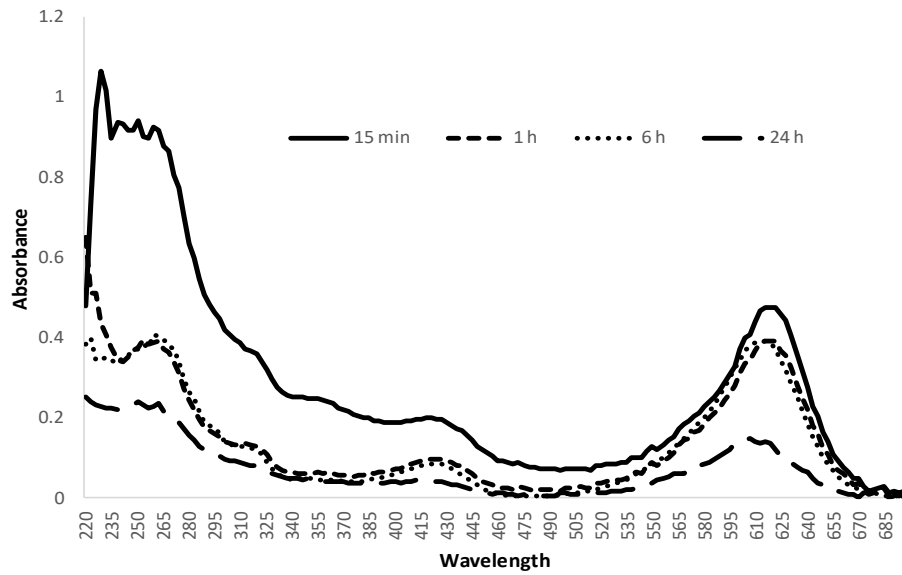


Fig 2.

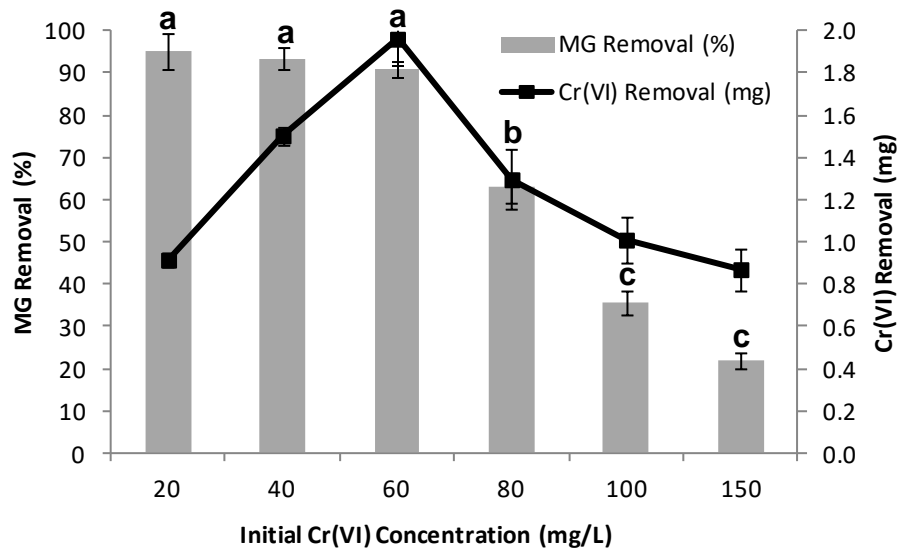


Fig 3

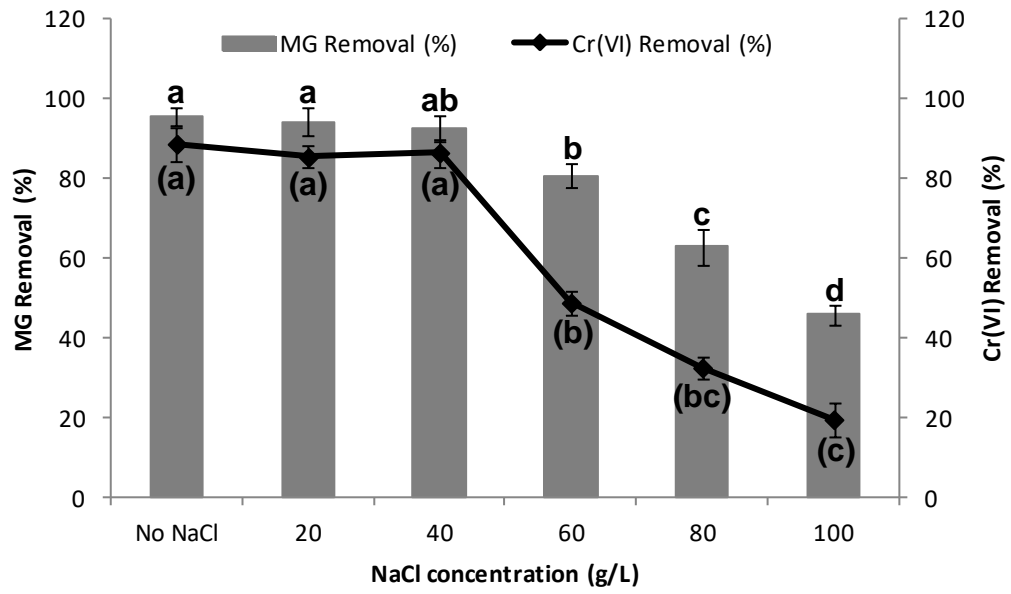
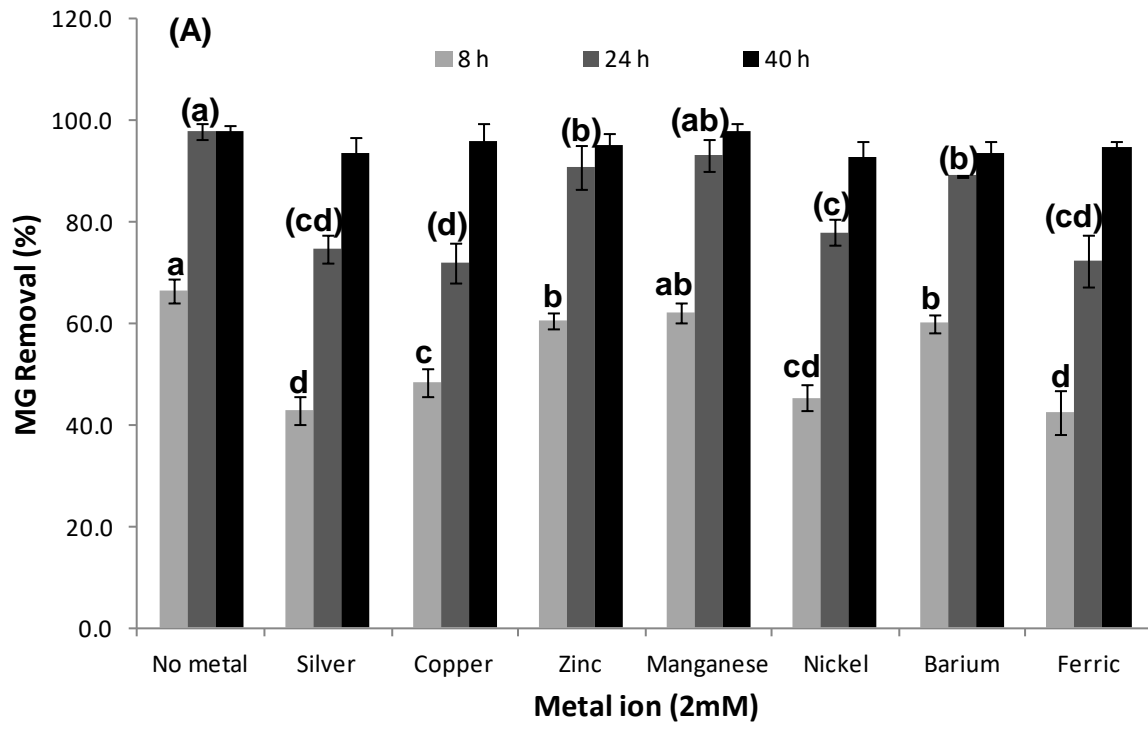


Fig 4



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