



Airborne remote sensing and in situ measurements of atmospheric CO₂ to quantify point source emissions

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Abstract. Reliable techniques to infer greenhouse gas emission rates from localised sources require accurate measurement and inversion approaches. In this study airborne remote sensing observations of CO₂ by the MAMAP instrument and airborne in situ measurements are used to infer emission estimates of carbon dioxide released from a cluster of coal-fired power plants. The study area is complex due to sources being located in close proximity and overlapping associated carbon dioxide plumes. For the analysis of in situ data, a mass balance approach is described and applied, whereas for the remote sensing observations an inverse Gaussian plume model is used in addition to a mass balance technique. A comparison between methods shows that results for all methods agree within 10 % or better with uncertainties of 10 to 30 % for cases in which in situ measurements were made for the complete vertical plume extent. The computed emissions for individual power plants are in agreement with results derived from emission factors and energy production data for the time of the overflight.

plausible methods. For example, Ackerman and Sundquist (2008) found differences of more than 20 % between emissions from US power plants. Differences in inventories of about a factor of 2 were found for CO₂ emissions from flaring in oil and gas production (Ciais et al., 2015).

Top-down estimates of localised sources are generally obtained using airborne or ground-based in situ measurements. Recently the use of airborne remote sensing has also demonstrated the ability to accurately estimate emissions. All methods have their distinctive advantages and disadvantages. Ground-based in situ measurements are fairly low cost. However, they generally do not sample the complete atmospheric boundary layer, which is necessary for an accurate emission estimate. Airborne in situ data allow accurate concentration and wind speed measurements from which emissions can be derived (e.g. Karion et al., 2013; Cambaliza et al., 2014; Caulton et al., 2014; Gordon et al., 2015; Lavoie et al., 2015). However, they require a dense flight pattern and assumptions, for example about the layer below the lowest flight track, have to be made. Furthermore, airspace restrictions can interfere with required flight patterns. Airborne remote sensing of atmospheric column concentrations of CO₂ allows sounding of the complete boundary layer and offers the opportunity to survey large areas in short time spans (Krings et al., 2011, 2013; Thompson et al., 2015; Frankenberg et al., 2016; Thorpe et al., 2014; Tratt et al., 2014). In contrast to in situ measurements, however, clear-sky conditions are mostly required as they measure backscattered solar

1 Introduction

Knowledge of emissions of the greenhouse gas carbon dioxide (CO₂) originating from localised sources is often inadequate (Ciais et al., 2015; NRC, 2010). Even for well-monitored localised CO₂ emitters, there are significant differences between inventories calculated using different but

electromagnetic radiation, except for those instruments operating in the thermal infrared (e.g. Tratt et al., 2014). However, interpretation of thermal infrared measurements depends on the thermal contrast (e.g. Young, 2002). Wind information has to be additionally gathered, for example from model data (see, for example, Krings et al., 2011) or additional in situ wind measurements (see, for example, Krings et al., 2013).

Using the example of CO₂ emissions from a cluster of power plants in western Germany, top-down results from airborne in situ and remote sensing data are evaluated and compared to each other, as well as to an independent bottom-up estimate computed from energy production and emission factors. Additional complexity is added to the top-down inverse problem since sources are partly located in close proximity to each other with overlapping CO₂ plumes. Generally, it is necessary to achieve a correct source attribution in the presence of multiple neighbouring sources for validation purposes.

2 Measurement campaign and target area

An airborne measurement campaign combining remote sensing measurements of column-averaged concentrations of CO₂ and CH₄, denoted XCO₂ and XCH₄ with in situ measurements of atmospheric CO₂ and CH₄ concentration (mass per volume) as well as wind speed and direction in the boundary layer, took place in Germany in August 2012. This campaign was carried out in the framework of ESA's Earth Explorer 8 activities for the candidate mission CarbonSat (Bovensmann et al., 2010; Buchwitz et al., 2013). The CarbonSat concept laid the foundation for the CO₂ monitoring mission now under preparation within the European Copernicus programme.

For the remote sensing of the greenhouse gases CO₂ and CH₄, MAMAP (Gerilowski et al., 2011) was flown above the boundary layer on the Cessna T207A aircraft belonging to the Free University of Berlin. MAMAP is an airborne 2 channel NIR-SWIR grating spectrometer system for accurate measurements of gradients in column-averaged methane and carbon dioxide concentrations. It was jointly developed by the Institute of Environmental Physics/Remote Sensing (IUP/IFE), University of Bremen (Germany) and the Helmholtz Centre Potsdam, German Research Centre for Geosciences (GFZ). It was demonstrated that the instrument is able to detect and retrieve the total dry column of the greenhouse gases CH₄ and CO₂ with a precision of 0.3–0.4 % (1 σ) at local scales (\approx 10 km), and that MAMAP is an appropriate tool for detection and inversion of localised greenhouse gas emissions from aircraft (Gerilowski et al., 2011; Krings et al., 2011, 2013; Krautwurst et al., 2017).

For the in situ airborne measurements, the small research aircraft METAIR-DIMO was flown in the boundary layer to perform measurements in the plumes emitted from the targets, to gather background concentration measurements and to perform wind measurements needed for the interpretation

of the total column measurements of MAMAP. The aircraft is equipped with underwing pods, carrying up to 50 kg of scientific payload each. The standard equipment measures the meteorological parameters wind (three-dimensional components in 10 Hz resolution), fast temperature and fast, redundant humidity. A two-channel aerosol counter (MetOne for > 0.3 and > 0.5 μm) can characterise the structure of the boundary layer. The chemical measurements are for CO₂ (redundant) and CH₄ corrected for H₂O interference (dilution and spectroscopic; for details see Hiller et al., 2014) as well as CO, O₃, NO₂, NO_x, NO_y and O_x. The CO₂ is measured with three different time resolutions and accuracies, resulting in an overall accuracy and precision of 0.4 ppm. The individual contributions are (i) a fast (10 Hz) measurement with a short-term precision (e.g. while crossing a plume) of about 0.2 ppm, with a limited absolute accuracy of about 5 ppm, using a modified LiCor LI-7500. (ii) A more accurate continuous but slower (0.3 Hz) reference with a modified LI-6262, with an accuracy of better than 0.5 ppm. The highest accuracy is from flask samples, analysed at MPI Jena. This method is described in Hiller et al. (2014).

The surveying strategy was to simultaneously probe the atmospheric boundary layer with in situ and remote sensing measurements where the MAMAP remote sensing measurements were performed via the second aircraft above the boundary layer.

While the complete campaign also covered other CO₂ and CH₄ emitting targets (Bovensmann et al., 2014), this study focuses on measurements obtained in an area with several lignite-fired power plants in western Germany (see Fig. 1) on 15 August 2012. The power plants are Niederaußem, Neurath (old and new blocks) and Frimmersdorf. The remote sensing flights were performed at about 11:50–13:40 UTC (that is, 13:50–15:40 local time, CEST). The in situ survey over the same area was conducted between about 12:15 and 14:20 UTC. Wind was blowing from the south-east (145–148°), so the CO₂ plumes of individual power plants were overlapping.

The in situ measurements concentrate on transects at several altitude layers around the two Neurath power plants and around the extended area, including the Frimmersdorf power plant.

3 Methodology

3.1 Flux estimates from in situ measurements

Calculating fluxes of trace gases through an imaginary vertical plane is simple when the concentration field and the wind field are known for a sufficient time during quasi-stationary conditions. However, in reality, such perfect measurements are not possible. Neither the accuracy nor precision of the measurements are a prime concern, but there are unknown parts in the fields (inter- and extrapolations), and most im-

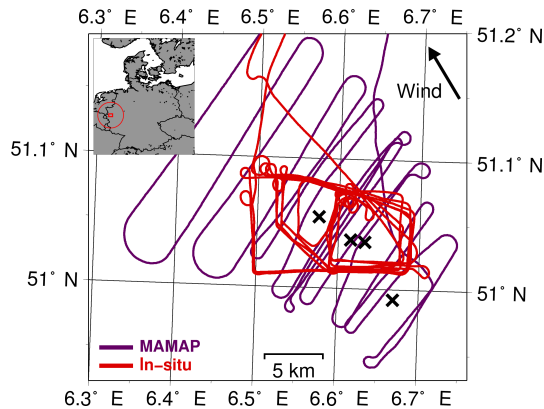


Figure 1. Map of the target area in western Germany. The crosses denote the four lignite-fired power plants in the area. From upwind to downwind, i.e. from south-east to north-west, the power plants are Niederaußem, Neurath (new), Neurath (old) and Frimmersdorf.

portantly, remaining instationarities. These could be caused by short-term fluctuations (hitting a part of a plume or not) and by varying source strengths and changing meteorological conditions. Cambaliza et al. (2014), Gordon et al. (2015), Lavoie et al. (2015) and Caulton et al. (2014) discuss comparable airborne in situ observations downwind of the regions and individual sources on different scales. The basic method used to derive emissions from atmospheric CO₂ or CH₄ measurements is described in Mays et al. (2009). The instrumentation and methods that were used in the work presented here are quite similar and use the following characteristics with reference to Gordon et al. (2015): (i) a “single-screen” approach was chosen (as opposed to a box method); (ii) due to the short distance to the sources, terms like chemical conversion or storage in the volume are irrelevant; (iii) the CO₂ emissions were from hot stacks placed at short distances which reduces the problem of extrapolation to the surface; (iv) the plume concentrations were large (see Fig. 2); (v) the time resolution of the instrumentation and the relatively low speed of the aircraft (typically 40 m s⁻¹) allowed us to include the turbulent horizontal fluxes (at 5 Hz for wind, concentrations and density, the spacial resolution was better than 10 m).

The single-screen approach was chosen for practical reasons, because flying around a source means spending most of the time in background concentrations, with a higher risk of missing quasi-stationary conditions. Some circumferential tracks (Fig. 2) confirmed the background concentrations found on the edges of the single screens. In contrast to these benefits we suffered from a non-ideal flight pattern caused by the following: (a) the airspace restrictions both horizontally and vertically were complex, and (b) this was the first day in a yet unknown region; i.e. we were still in the process of optimizing and preparing for other days with other wind directions (south to south-west). However, this was one of only a few days of the campaign when the remote sens-

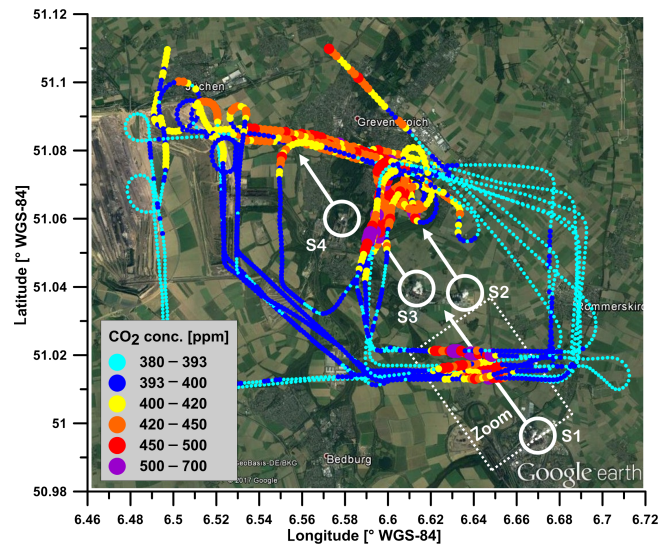


Figure 2. The in situ flight pattern around and between the four sources (S1: Niederaußem; S2: Neurath (new); S3: Neurath (old); S4: Frimmersdorf), with the wind direction (white darts, from 146°). The CO₂ concentrations (1 dot per 3 s-average, i.e. about every 120 m) are colour coded. In the north, the flight pattern was along an airspace that could not be entered. Also, on top, the cross sections were limited to 3500 ft AMSL (1050 m a.m.s.l.).

ing method could be directly compared with in situ measurements, because on many other days the cloud conditions were less ideal for remote sensing.

As in other work mentioned above, the measured data had to be inter- and extrapolated to a grid covering the whole of the cross sections. Because the data density on most cross sections was relatively high (see Fig. 3), and the plumes were not at all regularly shaped (rather an ensemble of puffs than a Gaussian plume), a method for the inter- and extrapolation was developed that mainly uses the grid cells of the cross sections in which measurements can be found. This structure can be seen in Fig. 4.

Before describing the interpolation process, the main difference between the method and other approaches is emphasised. In all the work referenced above, the fluxes were calculated by multiplying an interpolated wind field with an interpolated field of the mass distribution. This was done as well (flux 2, shown in Fig. 5). However, when the pointwise measurements of wind and concentration (or mass, when including density as a function of locally measured pressure, temperature and humidity) are known, the mass flux can be calculated locally for each point of the measurements. With the mentioned common sampling rate of 5 Hz (original data 10 Hz or faster; all sensors good enough for 5 Hz), this means that every 10 m there is a measurement of CO₂ flux perpendicular to the chosen cross section, including the turbulent contribution up to this frequency or length scale. When averaging several such measurements, this results in a direct

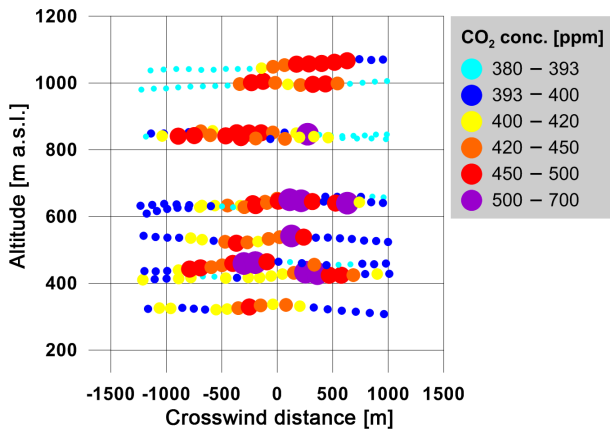


Figure 3. Details of the cross section through the plume of Niederaußem (S1). The projected CO₂ concentrations (1 dot per 3 s-average, i.e. about every 120 m) are colour coded. It is important to average the fluxes along the wind, i.e. all values have to be projected to an imaginary cross section perpendicular to the wind and not parallel to the flight pattern (see Fig. 2 and main text). The background concentration was determined to be at about 392 to 393 ppm, which is a small uncertainty compared to the maximum plume concentrations of about +100 ppm).

average mass flux for a chosen grid cell in the vicinity of the flight track. The most extreme coarse approach is to average over the whole flight near the cross section. These results are listed as the “bulk average” in the Supplement. It is astonishing that these results, which make no assumptions about the shape of the plumes whatsoever and do not need any method for interpolation, are mostly near or even within the range of the more sophisticated results based on reconstructed fields. Even more extreme is bulk average 2, for which the averages of all the mass measurements (above background) were multiplied with the averaged winds perpendicular to the whole of the cross sections. This finding to some degree relativizes the discussion about subtleties in interpolation methods. Nevertheless, in the following, we explain our linear inter- and extrapolation methods and compare them with kriging.

The cells shown in Figs. 4 and 7 with bold, larger numbers for the fluxes result from several measurements in these cells. The inter- and extrapolation was done following three rules: (i) single gaps in columns or layers were filled by linear interpolation (first horizontal, then vertical, and again horizontal); (ii) the lowest layer with measurements was completed by taking the value of the second lowest cell above if there were valid measurements in it; (iii) larger gaps were filled differently for two types of parameters: (a) wind and air mass (density) were filled with the averages in the layer; (b) concentrations, CO₂ masses and fluxes (all above background) were filled with the background. These rules conserve the values in the cells with measurements and tend to underestimate the fluxes through cells without measurements, resulting in con-

servative estimates. These rules apply for cells in the range of the highest and the lowest levels with measurements.

In all cases, the lowest flight track was about 150 m a.g.l. (legal limit without special permission), and the highest track was limited by the airspace, which was still in the plume on this day. This means that the layers between 1000 m a.m.s.l. (or 1100 m a.m.s.l. for the plume Niederaußem) and the top of the stable layer at 1300 m a.m.s.l. (centred altitude of the cell), and the layers below 250 m a.m.s.l. (lower boundary of the cells at nominal 300 m) had to be extrapolated. Below the lowest layer of measurements, the fluxes were interpolated linearly to zero at the surface (using the SRTM topography for the flat terrain with an average elevation of 48 m a.m.s.l.). Since we are primarily dealing with direct flux measurements, this is the only field we have to extrapolate. However, for calculating flux 2 as the product of wind and mass, we extrapolated the concentrations to background at the surface and kept the wind constant. This was due to the observation that the wind on flight altitude was usually the same as reported at the nearest airport on 10 m a.g.l. The areas of the cells for mass and fluxes were adapted accordingly in cells with topography. On top, the same procedure was applied: fluxes and concentrations (above background) were extrapolated to zero, with constant wind. In principal, more sophisticated extrapolation schemes could be applied, such as those discussed in Gordon et al. (2015). Ultimately, however, these profiles are not known, and their contribution here is low (less than other uncertainties).

Except for the very small plume of Frimmersdorf, the contribution of the extrapolated fluxes was less than 15 %. Dealing with individual local fluxes instead of mass and wind separately has an additional advantage for the extrapolation, since only one field has to be extrapolated. In cases for which the changes in wind and concentration are expected to compensate each other (e.g. increasing concentrations for sources near the surface, while the wind is decreasing), the range of plausible values for the extrapolation also has a smaller uncertainty. Whenever highly resolved data of good quality are available, we see no need to treat wind and concentrations (mass) separately, neither for the cells with measurements nor for those that need to be inter- or extrapolated.

Because the standard method for the inter- and extrapolation of measurements to a cross section in the literature referenced above is kriging, this was applied as well (graphics programme Surfer from Golden Software, including gridding with several options for kriging and other interpolation methods). An example of such a result is shown in Fig. 6, and all results are listed in the Supplement. By qualitative reasoning it seems that the results are less consistent than with the rules for limited linear interpolation described above. Gordon et al. (2015) has shown that kriging is best for simulated smooth plumes (Gaussian or similar). However, here highly irregular shapes had to be dealt with. Independently of which settings were chosen for the kriging (the variogram for the parameter was calculated in any case and both block

z [m a.s.l.]	Niederaussem: CO ₂ flux [kg s ⁻¹] above background in the grid cells (sums and linear interpolations & extrapolations)																						
1300	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0			
1200	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.2	2.4	1.9	2.7	2.3	1.9	3.4	3.2	5.7	7.1	6.6	0.6	0.1	0.1	0.1
1100	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.5	4.7	3.8	5.3	4.6	3.9	6.8	6.5	11.4	14.2	13.1	1.2	0.3	0.3	0.1
1000	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.5	4.7	3.8	5.3	3.9	0.5	4.0	11.3	9.7	2.9	2.4	0.1	0.0	0.0	0.0
900	0.3	0.1	2.0	0.3	0.3	5.0	3.6	1.3	6.3	11.2	15.9	15.4	5.1	0.1	0.0	7.8	5.8	1.5	1.2	0.1	0.0	0.0	0.0
800	0.6	0.2	4.1	10.2	9.4	7.2	5.0	4.6	19.2	5.6	7.7	2.0	1.2	8.1	9.8	4.3	1.8	0.1	0.0	0.0	0.0	0.0	0.0
700	0.5	0.4	2.4	5.4	4.9	4.2	3.6	3.3	10.6	5.5	9.5	0.3	6.4	16.1	7.7	0.2	0.2	0.2	0.1	0.2	0.2	0.1	0.1
600	0.4	0.6	0.8	0.5	0.3	1.2	2.2	2.0	2.1	5.4	11.3	5.2	8.9	23.7	14.0	7.7	1.3	7.0	26.5	3.0	1.1	0.1	0.2
500	0.5	0.2	0.5	1.0	0.9	1.8	0.9	2.6	8.4	9.7	8.2	7.9	1.4	7.7	5.8	2.8	0.9	0.3	0.3	0.3	0.4	0.4	0.4
400	1.0	0.7	0.6	1.0	5.8	4.8	2.5	0.8	1.3	1.0	1.9	2.6	1.7	3.0	18.6	31.3	18.3	16.5	11.2	6.2	0.8	1.5	0.3
300	0.5	1.1	1.3	1.0	0.3	0.4	1.3	1.0	3.0	7.2	12.5	2.6	1.7	3.0	1.6	0.7	0.3	0.4	0.4	0.4	0.3	0.3	0.3
200	0.4	0.7	0.9	0.7	0.2	0.3	0.9	0.7	2.0	4.8	8.3	1.7	1.2	2.0	1.0	0.5	0.2	0.3	0.2	0.3	0.2	0.2	0.2
100	0.1	0.2	0.2	0.1	0.0	0.0	0.1	0.1	0.3	0.7	1.4	0.3	0.2	0.4	0.2	0.1	0.0	0.0	0.0	0.0	0.0	0.0	0.0
0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
x [m] >	-1200	-1100	-1000	-900	-800	-700	-600	-500	-400	-300	-200	-100	0	100	200	300	400	500	600	700	800	900	1000

Figure 4. Example of results for the cells on a cross section. Shown is the direct CO₂ flux in the plume of Niederaußem (S1, same plume as in Fig. 3). Since the grid cells are not points, but areas filling 100 × 100 m² (or more precisely, voxels of 100 × 100 × 1 m³), the values are not displayed as contour plots, but as pixels. The x and z coordinates are centred in the cells. The standard inter- and extrapolation were done with the methods described in the text. Cells with bold numbers [kg s⁻¹] contain measurements (20 points in the average), while the inter- and extrapolated cells are shown with smaller italic numbers.

z [m a.s.l.]	Niederaussem: CO ₂ flux type 2 (wind x mass) [kg s ⁻¹] above background in the grid cells																						
1300	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
1200	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.2	2.7	2.4	3.9	2.3	2.0	3.4	3.2	5.6	7.0	6.7	0.6	0.1	0.1	0.1
1100	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.5	5.3	4.7	7.8	4.6	4.0	6.8	6.4	11.1	14.1	13.3	1.2	0.3	0.3	0.1
1000	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.5	5.3	4.7	7.8	2.6	0.6	4.0	11.3	9.7	3.0	2.5	0.1	0.0	0.0	0.0
900	0.3	0.1	1.8	0.3	0.3	4.9	3.6	1.3	6.7	11.2	16.0	15.3	4.9	0.1	0.0	8.0	6.2	1.7	1.4	0.1	0.0	0.0	0.0
800	0.7	0.2	3.7	10.2	9.5	7.2	4.0	4.5	19.2	9.6	7.4	2.0	2.8	7.8	8.4	3.6	2.0	0.1	0.0	0.0	0.0	0.0	0.0
700	0.6	0.5	2.1	5.0	5.1	4.6	3.5	3.5	10.8	7.2	10.3	1.1	6.1	16.4	7.6	0.2	0.2	0.2	0.1	0.2	0.2	0.1	0.1
600	0.4	0.6	0.8	0.5	0.4	1.2	2.1	1.9	2.2	5.2	12.0	4.9	8.8	23.7	13.5	8.3	1.3	6.6	26.4	2.9	1.0	0.1	0.2
500	0.5	0.2	0.5	1.0	0.9	1.8	1.1	2.9	8.3	9.0	6.1	6.7	2.0	9.4	6.0	3.3	1.0	0.3	0.3	0.3	0.4	0.4	0.4
400	1.0	0.7	0.6	1.0	5.6	5.1	2.9	0.8	1.3	1.0	2.0	2.6	1.7	2.8	18.3	31.4	20.2	14.3	11.3	6.2	0.8	1.5	0.3
300	0.5	1.1	1.3	1.0	0.3	0.4	1.3	1.0	2.9	7.4	13.2	1.6	4.0	4.8	1.6	0.7	0.3	0.4	0.4	0.4	0.3	0.3	0.3
200	0.4	0.7	0.9	0.7	0.2	0.3	0.8	0.7	2.0	4.9	8.8	1.1	2.7	3.2	1.0	0.5	0.2	0.3	0.2	0.3	0.2	0.2	0.2
100	0.1	0.2	0.2	0.1	0.0	0.0	0.1	0.1	0.3	0.8	1.5	0.2	0.5	0.6	0.2	0.1	0.0	0.0	0.0	0.0	0.0	0.0	0.0
0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
x [m] >	-1200	-1100	-1000	-900	-800	-700	-600	-500	-400	-300	-200	-100	0	100	200	300	400	500	600	700	800	900	1000

Figure 5. As Fig. 4 but for CO₂ flux 2, i.e. the product of averaged wind and mass in the individual cells. The difference between these two methods is discussed in the text and is one of the sensitivity cases as summarized in Table 3 and detailed in the Supplement.

fits and point fits were tried) it seemed that the fields became too smooth. Kriging was applied both to the directly measured fluxes and the flux 2 resulting from mass and wind fields after kriging. We do not claim to have found the optimum method, but we think that the difference between the methods is smaller than other uncertainties and that basically, we do not know which concentrations, winds and fluxes are present between the cells with measurements. However, with

the rules described above, we make sure that the values in the cells with enough measurements are not affected.

Referring to the importance of turbulent fluxes, Foken et al. (2009) gives a concise overview of the difficulties of complete closures of fluxes. The method with the direct local calculation of fluxes at each measuring point includes the turbulent fluxes in the mean wind direction. In some convective situations, the vertical turbulent flux above the source

z [m a.s.l.]	Total CO ₂ flux [kg s ⁻¹] above background in the grid cells (by Kriging, without extrapolations)																						
1300																							
1200																							
1100	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	1.5	1.4	3.0	3.5	4.0	5.6	7.9	10.3	12.8	9.6	2.5	0.3	0.2	0.1
1000	0.1	0.0	0.0	0.0	0.0	0.0	0.0	0.0	1.7	5.0	6.8	5.1	3.4	2.0	4.7	8.2	9.1	8.3	5.0	1.3	0.1	0.1	0.1
900	0.3	0.1	1.7	2.5	1.6	3.3	2.3	1.8	7.3	8.0	9.7	7.7	2.5	1.7	7.3	7.2	4.2	2.3	0.8	0.1	0.0	0.0	0.0
800	0.5	0.3	2.9	3.9	2.8	4.6	3.1	3.3	8.7	7.8	10.1	7.1	4.1	6.3	8.3	4.0	1.1	0.6	1.9	0.3	0.1	0.0	0.1
700	0.5	0.4	2.0	2.5	2.0	3.0	2.6	3.2	5.4	6.7	9.9	5.5	5.9	11.7	8.9	3.5	0.9	2.0	7.2	0.9	0.4	0.1	0.1
600	0.5	0.5	0.7	0.6	0.7	1.2	1.5	2.4	3.4	5.2	7.3	3.5	5.6	13.4	11.2	4.5	0.7	2.9	7.7	1.5	0.6	0.3	0.3
500	0.6	0.5	0.7	1.0	3.6	3.9	1.8	3.0	6.3	7.0	6.6	5.9	2.7	5.2	8.6	8.5	2.9	1.6	1.7	1.1	0.5	0.6	0.4
400	0.8	0.8	0.8	1.0	3.2	2.7	1.7	1.9	3.8	5.8	5.2	3.5	0.7	1.7	7.3	12.2	6.5	4.8	4.9	2.3	0.7	0.7	0.4
300	0.7	0.9	1.2	0.9	0.5	0.5	1.1	1.6	3.6	6.4	3.8	0.0	0.1	0.4	0.9	1.2	0.9	0.4	0.7	0.7	0.3	0.4	0.3
200																							
100																							
0																							
x [m]	-1200	-1100	-1000	-900	-800	-700	-600	-500	-400	-300	-200	-100	0	100	200	300	400	500	600	700	800	900	1000

Figure 6. The same as Fig. 4 but for the CO₂ flux calculated by kriging of the original data instead of averaging in cells and applying our method for linear inter- and extrapolations. The kriging was only done for the part of the cross section with layers containing data. The comparison with our standard method (see Supplement) was done by adding the percentage of the extrapolated flux (up to the stable layer and down to the surface) of the standard method (+10 % in this case).

z [m a.s.l.]	Cluster: CO ₂ flux [kg s ⁻¹] above background in the grid cells (sums and linear interpolations & extrapolations)																										
1300	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0			
1200	0.0	0.0	0.0	0.0	0.0	0.0	2.4	2.0	3.7	1.1	1.5	1.1	2.7	4.9	4.1	8.3	5.6	1.8	1.0	0.1	6.1	3.3	2.5	2.7	1.2	1.0	0.4
1100	0.0	0.0	0.0	0.0	0.0	0.0	4.8	3.9	7.4	2.2	3.0	2.3	5.5	9.7	8.3	16.7	11.1	3.7	2.0	0.2	12.2	6.6	4.9	5.4	2.5	2.0	0.7
1000	0.0	0.0	0.1	0.0	0.0	0.0	7.2	5.9	11.2	3.3	4.4	3.4	8.2	14.6	12.4	25.0	16.7	5.5	3.0	0.3	18.3	9.8	7.4	8.0	3.7	3.0	1.1
900	0.0	0.0	4.1	8.0	5.9	12.2	12.2	8.1	13.3	9.1	9.6	8.5	12.4	17.1	13.1	18.5	11.7	8.9	11.8	12.2	24.7	18.8	0.0	0.0	8.0	3.4	0.0
800	0.0	0.0	8.2	16.0	13.3	9.6	17.3	10.4	15.4	14.9	14.8	13.5	16.5	19.6	13.7	12.0	6.7	12.2	20.6	24.0	31.2	2.9	4.0	5.2	12.4	3.8	0.0
700	0.0	0.0	0.0	11.1	9.8	14.5	14.0	10.0	14.2	16.3	13.0	11.4	10.9	18.8	15.9	6.9	4.7	24.7	10.4	12.0	15.6	1.4	2.1	2.6	2.9	3.3	4.1
600	0.0	0.0	0.0	6.1	6.3	19.5	10.8	9.5	13.0	17.8	11.1	9.3	5.2	17.9	18.2	1.9	2.8	1.6	0.2	0.0	0.0	0.0	0.2	0.0	0.0	0.0	0.0
500	1.1	0.0	0.0	0.0	8.9	13.1	11.0	2.4	0.6	3.3	2.6	3.4	10.2	8.6	7.0	4.2	12.5	16.1	10.0	9.0	1.8	0.7	0.2	2.8	0.0	0.0	0.0
400	1.0	1.3	0.0	0.0	0.5	0.9	1.0	1.8	8.3	14.1	5.3	6.5	7.6	9.7	6.1	10.0	8.8	1.9	0.7	1.4	4.7	2.5	0.2	0.3	0.1	0.0	0.0
300	1.0	1.1	1.1	0.9	1.2	3.9	15.2	1.8	8.3	14.1	5.3	4.3	9.9	3.9	0.8	0.3	0.3	0.3	0.3	0.2	0.2	0.1	0.3	0.4	0.1	0.0	0.0
200	0.7	0.7	0.7	0.6	0.8	2.6	7.0	1.2	5.6	9.4	3.5	2.9	6.6	2.6	0.6	0.2	0.2	0.2	0.2	0.1	0.1	0.0	0.2	0.3	0.0	0.0	0.0
100	0.2	0.2	0.2	0.2	0.3	0.7	0.0	0.3	1.9	3.2	1.2	1.0	2.2	0.9	0.2	0.1	0.1	0.1	0.1	0.0	0.0	0.0	0.1	0.1	0.0	0.0	0.0
0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
x [m]	-2800	-2600	-2400	-2200	-2000	-1800	-1600	-1400	-1200	-1000	-800	-600	-400	-200	0	200	400	600	800	1000	1200	1400	1600	1800	2000	2200	2400

↑

Frimmersdorf
 1.23 Mt a⁻¹
 39.10 kg s⁻¹

Figure 7. The same as Fig. 4 but for the mixed plume of the whole cluster on the northern transect. The weak plume from Frimmersdorf, which is only 4 km away from this cross section could be isolated below 450 m a.m.s.l. The confirmation was done with an analysis at higher resolution. The layer between 350 and 450 m (centred at 400 m) has enough measurement points allowing the small lower plume to be separated from the main plume from Niederaußem (probably the whole width of about 4.5 km) and the two plumes from Neurath old and new. See Fig. 8 for the separation on a cross section upwind of Frimmersdorf, which is closer to the three other plumes.

can also be considered, which was not a necessity in this case. In principle, the crosswind turbulent diffusion could also be calculated. However, this does not contribute to the flux from the source and would only deliver an estimate for the plume broadening (lateral entrainment and detrainment).

The results in the detailed table in the Supplement (difference between 200 × 100 fb and flux 2) show that the contribution of the turbulent fluxes is not positive, and in all the cases only a few percent. Two reasons are possible in combination: (i) uncertainties in the calculations; (ii) when the turbulent

flux is mainly responsible for dilution (entrainment), the turbulence reduces the net flux.

The positions of the cross sections were selected (filtered from the whole flight) with minimum and maximum distance to the source, the mean wind direction and the lateral distance from the centerline. The angle of the cross section was adjusted for a crosswind component of less than 0.1 m s^{-1} , and the width of the cross section should include enough of the background concentrations.

The background concentration for this case study was relatively easy to find, and the results are not sensitive to it because the peak plume concentrations were high above the background (100 ppm and more; see Fig. 2). Originally, the standard method was to find the background on each layer by finding the minimum concentration. However, a constant background concentration of 392 ppm was selected in order to avoid artefacts in widely contaminated layers. There is another reason against taking the minimum for each layer, typically resulting in slightly decreasing background concentrations with altitude in this region: the emissions were injected at low altitudes into the background concentrations that were present there. When such a plume is rising into lower (or higher) background concentrations, taking the local background would lead to an over- or underestimation of the emission. Therefore it is better to take the background concentration at the altitude of the sources for the whole cross section.

Figure 2 can also be used to explain the steps of the processing. Based on the flight track on the map, the minimum and maximum distances from the source were defined. Within these distances, the measurements are projected onto a vertical plane perpendicular to the mean wind vector. The fluxes from Niederaußem were calculated using measurements between 2 and 5 km, on a cross section 3.5 km away (this distance is not relevant, because the projection is parallel). The orientation of the cross section was adjusted until the average crosswind component was less than 0.1 m s^{-1} . Larger crosswind components would cause artefacts in the same way as if the cross section was aligned with the flight track, applying wind components perpendicular to this plane. If all the flight tracks perfectly overlapped, the orientation of the cross section would not be important, and both options would be possible (exactly perpendicular to the wind or along the flight track). However, when different tracks at different distances are involved, the maximum concentrations on the different traverses would not be aligned and would not be averaged in the same grid cell of the cross section. Instead, they would contribute to neighbouring cells, increasing these fluxes. Since the flight track in other sectors was quite complex, observing this rule was very important. Another precaution was applied: even if the wind measurement should also be accurate in steep turns, parts with more than five degrees bank angle (roll) were excluded. Since most parts of the plumes were crossed on straight flight legs, this did not reduce the available data considerably.

For the separation of the individual contributions of sources that were emitting in the same cross section, two methods were applied: in the case of Frimmersdorf, the small plume was identified on the cross section in the lee of all the sources. The height and width parameters of the cross section could then be adjusted to cover this plume only. This part of the larger cross section is marked in Fig. 7. However, on the cross section cluster 3 (see Fig. 8), the three plumes (Niederaußem, Neurath new and old) were overlapping in a way that a direct separation was not possible. Therefore, two regions on the cross section were defined where the sources Neurath old and new most likely dominated. The source Niederaußem, 9 km upwind of this cross section, was most likely contaminating this whole cross section, mainly above 450 m, laterally within the limits marked with dashed lines. The percentage of contributions in the overlapping parts was varied in a wide range between 20 and 80 % (from Neurath old or new vs. Niederaußem).

Using the difference between the whole cluster (measured directly on the most downwind cross section) and cluster 3 (all power plants except Frimmersdorf) measured directly with the small plume of Frimmersdorf, it was possible to estimate the flux that was not captured on the cross section cluster 3. The underestimation is due to a rather cautious extrapolation above the highest flight track. Of course the direct measurement of the whole cluster could also be underestimated in the same way, but by adjusting the two, the budget is coherent as a lower limit of the real emissions. Of course all these operations are not exact. Therefore the error estimates were done quite conservatively by taking the minima and maxima of the results of all the assumptions (e.g. the percentage mentioned above), resulting in error estimates of 22 % for Neurath old and 50 % for Neurath new. The difference is due to the fact that Neurath old is closer to the cross section flown than Neurath new and was therefore more easily distinguishable on the cross section.

3.2 Discussion about main uncertainties in situ

3.2.1 Measurement errors

The wind components (u , v , w) have an accuracy of 0.5 m s^{-1} each, while the CO₂ concentrations have an accuracy of better than 1 ppm. However, as the background concentrations are subtracted, the absolute concentrations are not important, resulting in uncertainties in terms of precision (stability of the sensitivity within an hour) instead of absolute accuracies. This results in maximum uncertainties of 0.5 ppm for CO₂ (using the two CO₂ instruments in combination with flask samples). The uncertainties of the wind measurements remain in the order of 0.5 m s^{-1} per component (as for the perpendicular component on the cross section). The main uncertainty is the crosswind component relative to the aircraft. However, when flying back and forth through any plume on a similar altitude, this error is averaged out. Since this was

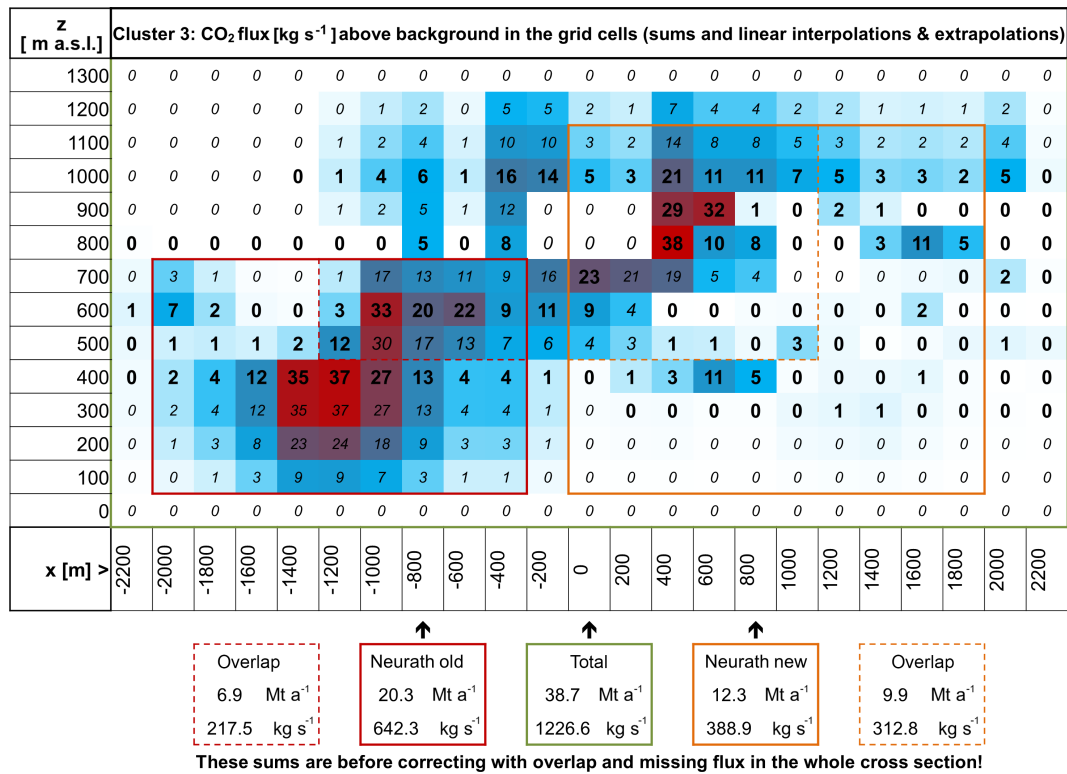


Figure 8. The same as Figs. 4 and 7, on the cross section downwind of the total cluster without Frimmersdorf (S4). The two domains and the overlap with the plume from Niederaußem (S1) in the background were used to separate the almost collinear plumes of Niederaußem and Neurath old (S2) and new (S3). The uncertainty was quantified by varying the parameters in the overlapping cells between 20 and 80 % (background from Niederaußem versus Neurath old and new).

not always possible, we take the worst case 0.5 m s^{-1} . Then the relative error based on the wind measurement is 10 % at wind speeds in the order of 5 m s^{-1} , increasing with weaker winds, and decreasing with higher wind speeds. For CO₂, a plume with moderate 50 ppm above background only adds another percent. In the case discussed here, wind speeds were around 8 m s^{-1} , with excessive CO₂ concentrations of more than 100 ppm (see Figs. 2 and 3), leading to the conclusion that the error for horizontal fluxes due to the uncertainties of the primary measurements is clearly below 10 %. This is in agreement with the assessment found in Gordon et al. (2015). In summary, the measurement itself – if state of the art – is not the main source of errors.

3.2.2 Other sources of errors

The accuracy and precision discussed above are important but are not the main criterion for reliable flux estimates. Only under unrealistically ideal conditions could winds and concentrations over the whole of the cross sections be captured completely and instantaneously, and repeated many times, and the remaining uncertainties would arise from the (systematic) measurement errors. These are, as discussed above, quite small. However, as already mentioned in Sect. 3.1 and

discussed in similar work referenced there, other reasons for uncertainties can dominate, which are harder to quantify. The approach used here was to vary the assumptions and parameters used in the calculations in a wide range. This sensitivity analysis allowed the methodological variations to be compared with the average results.

The main uncertainties are dependent on the meteorological conditions and the flight patterns. An ideal flight pattern covers the plume as completely as possible. For the CO₂ plumes, the minimum height of 50 or 150 m above ground is generally sufficient when measuring close to the elevated hot sources. The main source for uncertainties in the flux estimates are the interpolations and extrapolations as discussed in Sect. 3.1.

Another source of uncertainty already mentioned is instationarity, i.e. varying source strengths from day to day or even by the hour, while the measurements account only for the emissions for the time of the overflight. For this case study, the source variation based on energy production was less than 0.5 % for Niederaußem and Neurath new and old blocks. For Frimmersdorf the variability was about 4 % but with considerably lower total fluxes (see Sect. 5). Another type of instationarity is caused by the atmospheric turbulence on the scale of a few hundred metres, due to which a

maximum concentration (puff instead of continuous plume) can be missed (underestimation) or captured by coincidence (overestimation). The only way to minimize this is by repeating the pattern as often as possible and is another reason to spend as much time as possible on the single screens (cross sections). Generally spoken, the 4-D inhomogeneities cannot be captured in a snapshot.

Another issue is the definition of the boundary layer. In an ideal case, cross sections are flown to an altitude at which no concentrations in excess of the background are detected anymore; i.e. where the plume is confined below. This plume was captured under ideal conditions with no extrapolation needed at all on 23 August near a single source. Therefore, the source Weisweiler is listed in the results of Table 3 in Sect. 4.1 as a reference. The astonishing conclusion is that the variation due to different interpolations is comparable to the non-ideal cases in this study in which extrapolation was necessary. However, the uncertainty of the extrapolations has to be added, as in Table 3.

Since a deductive error estimate as is possible for the basic measurements is not possible for the overall flux results, a sensitivity analysis was applied to all cases. In this case study, the five individual cross sections with fluxes were calculated using six sets of parameters for our approach and for kriging, and with three extreme (unrealistic) assumptions. The percentages of the contributions from the extrapolations are listed in Table 3 as well. They ranged between 10 and 20 % for the directly measured plumes, with one exception (Frimmersdorf), where the weak plume increased the extrapolated contribution to 45 %. A conservative approach for the total error is to add half of these contributions (assuming a 50 % uncertainty in the extrapolations) to the differences found by the sensitivity analysis. The overall uncertainty for the total emissions of the cluster would then be 18 %, 14 % for Niederaußem and between 32 and 63 % for other individual sources (Table 3), which were separated indirectly.

When combining sources by adding or subtracting (e.g. subtracting Frimmersdorf and Niederaußem in order to get an estimate for old and new Neurath), these uncertainties are added in an unknown way (some systematic errors do not add but compensate; others are added as components via a root of sums of squares) or have to be added for a possible maximum error. Therefore, when separating sources as described in Sect. 3.1, the extreme values within the parameter space were taken.

3.3 Fluxes from remote sensing greenhouse gas information

3.3.1 Measurement data

The processing of the MAMAP remote sensing data is based on the methods described by Krings et al. (2011, 2013). A modified version of the weighting function modified differential optical absorption spectroscopy (WFM-DOAS) algo-

rithm (Buchwitz et al., 2000) is used to obtain vertical column information of CH₄ or CO₂. It relies on a least squares fit of the logarithmic simulated radiance spectrum to the measurements after correction for dark signal and pixel-to-pixel gain. Additionally a look-up table approach has been implemented, accounting for varying solar zenith angle (SZA) and surface elevation. The conversion from total columns to column-averaged dry-air mole fractions (XCH₄, XCO₂) is performed using the proxy method, assuming that locally CH₄ is sufficiently constant to compute XCO₂ (or vice versa for XCH₄). See Frankenberg et al. (2005) and Schepers et al. (2012) for more information on the proxy method and Krings et al. (2011) for its application to MAMAP measurements. This method is suitable for point sources, as is the case in this study.

The corresponding background profiles for the linearization points use the US standard atmosphere (U.S. Committee on Extension to the Standard Atmosphere, 1976) scaled to actual values. In this case, background XCO₂ was determined to 390 ppm based on in situ data upwind of the power plants in the boundary layer and results from the SECM model (Reuter et al., 2012) above. The methane background XCH₄ was estimated to about 1.805 ppm, also based on in situ measurements in the upwind area of the boundary layer scaling the standard atmosphere. An assumed uncertainty of the ratio of the background columns of 1 % accounts for possible deviations from these values. The spectroscopic database used for the computation of absorption cross sections was HITRAN 2012 (Rothman et al., 2013).

Aircraft altitude during the measurements was almost constant at about 1590 m a.m.s.l. (± 25 m), which was also selected for the reference radiative transfer. Assuming a constant aircraft velocity of 200 km h⁻¹, the ground scene size is about 22 m \times 54 m (cross track \times along track) for the installed optical front telescope. Thereby the along-track ground scene size describes the full width at half maximum for the sensitivity along the flight track. During the measurement, the solar zenith angle varied from about 37.5 to 45.3°.

The radiative transfer model was interpolated using a two dimensional look-up table (LUT) based on solar zenith angle and surface elevation. For that, the SRTM digital elevation model was used (Shuttle Radar Topography Mission version 2.1, http://dds.cr.usgs.gov/srtm/version2_1/, a collaborative effort from NASA, NGA as well as the German and Italian Space Agencies). Due to the changing measurement geometry, the conversion factor to correct for the altitude sensitivity effect (Krings et al., 2011) has to be determined for each measurement independently, also using a LUT. This correction takes into account that light passes twice below the aircraft where the observed plumes are located. On average, the conversion factor for the present measurements is about 0.49.

3.3.2 Quality filtering

Filtering, based on the spectroscopic fit quality, has been applied by rejecting measurements with a root mean square (RMS) value of the differences between measurement and model after a fit larger than 0.9 % relative to the model affecting about 0.1 % of the total measurements. The threshold was empirically determined from the distribution of RMS values ordered by size (compare Krings et al., 2011, 2013).

An additional filter has been applied that is dependent on the signal strength to avoid measurements close to saturation (more than $\approx 90\%$ detector filling) and in the lower signal to noise range, e.g. over water which has a lower surface spectral reflectance in the short-wave infrared. Filtering of the data accounts for not only SNR but also whether linear full well capacity is achieved. For the full well ADC (analog-to-digital converter) range selected by the manufacturer a non-linear behaviour could be observed for very high detector fillings. Therefore data with very high filling factors are excluded from further processing. However, out of all measurements, the selected maximum threshold value affects only four single measurements during the whole measurement period.

Measurements with a detector filling of less than about 20 % (13 000 counts) appear to have a slight signal dependency and were neglected for the inversion process. Furthermore to ensure nadir viewing geometry the deviation from the vertical was not allowed to exceed 5°.

The XCO₂(CH₄) precision after filtering is approximately 0.29 % determined from the standard deviation of the data outside the plume area.

Figure 9 shows the XCO₂(CH₄) data acquired over the coal-fired power plants without and with the filtering applied. Clearly visible are the overlapping CO₂ plumes originating at the individual power plant locations and advected in a downwind direction towards the north-west in agreement with the wind field as computed by the routine analysis of the numerical weather prediction model COSMO-DE (Doms, 2011).

3.3.3 Atmospheric conditions and wind information

A fundamental parameter for the inversion is wind speed. To compute an average wind speed throughout the CO₂ plume from the model and in situ data, the boundary layer depth is also important.

The aerosol and virtual potential temperature profiles give no indication that the transition to the free troposphere is located in the lower 1100 m. This is furthermore confirmed by the fact that there are enhanced CO₂ amounts throughout the probed altitude layers. Consequently it can only be concluded that the boundary layer depth is likely larger than 1100 m. For the analysis a boundary layer depth of 1500 m was assumed (with uncertainty estimates for cases of 1200 and 1800 m; see Sect. 3.4).

Figure 10 shows measured wind speed and direction for the day analysed in this study compared to results from the COSMO-DE model. The measured in situ wind speed averaged over 60 s ranges from about 7 to 11 ms⁻¹ for the time and altitude range of the overflight and is rather constant with altitude. Altitudes covered by in situ measurements range from about 300 m to 1100 m a.m.s.l.

Modelled and measured wind direction are similar in lower altitudes but agree less well at higher altitudes and later times (Fig. 10). However, the wind direction is derived from the remote sensing data directly without using the model information.

Since in situ wind information is not always available in time and space where remote sensing measurements are taken, the in situ wind data are used to calibrate the COSMO-DE model result in the measurement area during the time of the remote sensing overflights.

The precision of the wind model was estimated to about 0.9 ms⁻¹ (1σ) with negligible bias for the case described in Krings et al. (2011). Assuming the same error holds in the present study, this leads to a wind-based relative error (1σ) on the inversion of about 10 %. This error can be reduced when on site wind information is available, for example from airborne turbulence measurements that were used in Krings et al. (2013) and also performed during the present campaign.

To quantify the difference between measured and modelled wind speed, the probed altitude range has been divided into five equally thick layers in which the deviation between the in situ measurements on the one hand and the associated model data interpolated in time and space on the other hand were computed and subsequently averaged over the altitude layers.

For the available model and measurement data from the target area and time, this yields an average overestimation of about 0.58 ms⁻¹, i.e. about 6.1 %. This is well within the approximate error of about 0.9 ms⁻¹. Similarly to Krings et al. (2013) the in situ wind error of 0.5 ms⁻¹ is assigned to the calibrated wind.

The complete results are shown in Table 1. Note that the results do not directly relate to Fig. 10, which only shows the model wind speed at one specific COSMO-DE grid point, while model wind data from the whole measurement area enter the computations in Table 1. The standard deviation of the model wind speed over the measurement area for the model layer shown in Fig. 10 is about 5.8 %.

3.3.4 Inversion for emission data

Emission rate estimates are then obtained using an inverse Gaussian plume model fitting flux and atmospheric stability. In a second approach mass balance estimates are computed leading to two independent inversion methodologies with the exception of wind information, which is taken from COSMO-DE model, and the in situ turbulence probe of the DIMO aircraft and that is used for both methods.

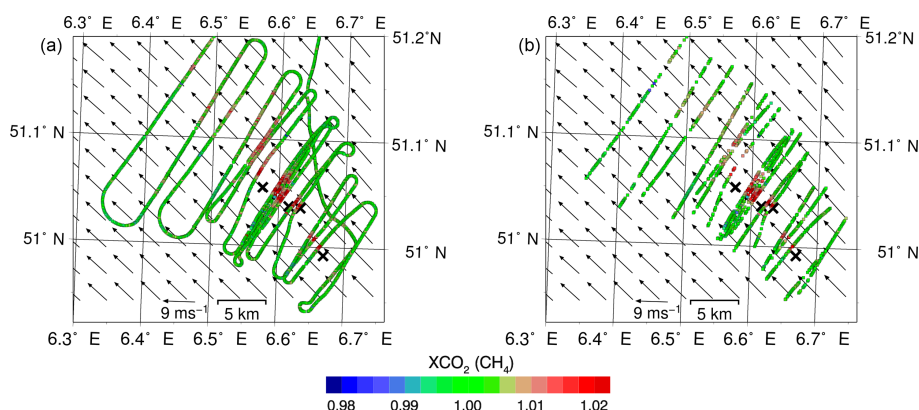


Figure 9. Qualitative MAMAP remote sensing XCO₂ data unfiltered (a) and filtered as described in the main text (b). The crosses denote the power plant locations (see Fig. 1). Arrows denote wind vectors from the COSMO-DE model at an altitude of about 350 m a.m.s.l. (model layer 45) at 13:00 UTC.

Table 1. Comparison of modelled and measured wind speed for several altitude layers.

Altitude range (m a.m.s.l.)	Model wind speed (m s ⁻¹)	In situ wind speed (m s ⁻¹)	Wind speed difference (in situ – model) (m s ⁻¹)	Relative wind speed difference (in situ – model)/model (%)
291–440	8.99	8.53	–0.46	–5.1
440–588	9.33	8.47	–0.86	–9.2
588–737	9.48	8.53	–0.95	–10.0
737–885	9.84	9.11	–0.74	–7.5
885–1034	9.27	9.40	+0.13	+1.4
			Average: –0.58 m s ⁻¹	Average: –6.1 %

Preparation and performance of the Gaussian plume inversion and the integral inversion method is very much in line with Krings et al. (2011, 2013). Since the inversion proved to be extremely stable no a priori information on emission rate or stability was required. This simplified the cost function to be minimised in the iterative inversion process.

The data were gridded to pixels of 35 m × 35 m with approximately the same area as the MAMAP ground scene size. The impact of different pixel sizes for the gridding is assessed in Sect. 3.4. No additional smoothing was applied. Note that the gridding was not used for the mass balance method.

Prior to the inversion the data were normalised by dividing them by the regional background. Since the measured area and time are somewhat larger than in the previous studies, no constant normalisation was selected, but a track-by-track procedure that is also able to account, for example, for linear gradients that are unrelated to the source. Thereby data from each cross section (see Fig. 9) are normalised by a linear function that is determined by the flanks of the track excluding the plume area. If the track does not measure sufficient data outside the plume, then this method results in an underestimation of emissions. Because of that and because gaps

due to filtering are rather large from track 10 onwards, only the first 9 downwind tracks with respect to the Niederaußem power plant could be analysed using the standard filtering. To investigate the measurements from further downwind, the signal threshold was relaxed for the first three tracks downwind of the power plant at Frimmersdorf to a minimum of 3000 counts and the inclination filter to 15°. In this way it is ensured that a sufficient set of measurements, even if of lower quality, are available for interpretation. The Gaussian plume method was not applied to these data as that would require data to be mixed which were subject to different filter criteria.

Obtaining an adequate estimate of the mean wind speed with which the emitted gas is transported is generally challenging when there are several sources which are separated in downwind direction. In the present case study the low variability of wind speed with altitude (see Fig. 10), however, makes the estimation less sensitive to errors in this regard. Here, the mean wind speed was estimated assuming Niederaußem power plant, the power plant that is located most upwind of all emitters, as an only source. The emitted CO₂ was then distributed using a vertical Gaussian dispersion with the stability parameter resulting from a 2-D horizontal Gaussian

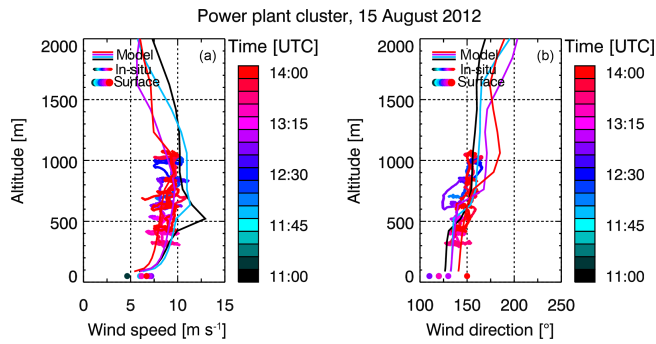


Figure 10. Wind speed (a) and direction (b) as measured from the in situ turbulence probe as well as COSMO-DE model data and surface in situ information from Düsseldorf airport about 30 km north-north-east of Neurath power plant (obtained from weather underground, <http://www.wunderground.com>).

plume inversion model. This information could be used to obtain an altitude-weighted mean wind speed for the remote sensing flight legs based on relative concentrations per altitude layer. The resulting wind speeds for each remote sensing flight leg were used individually for the mass balance approach and averaged for the inverse plume model over the relevant area. At the first remote sensing leg 700 m downwind of the power plant at Niederaußem, the plume reaches about 1 km in height, and at 2 km downwind the CO₂ is already well mixed according to the plume model, which represents a temporal average.

Boundary layer top height (represented as a reflective layer as in Turner, 1970) and emission height needed for the vertical dispersion were varied around the baseline parameters to estimate the range of errors resulting from these assumptions (see Sect. 3.4).

3.4 Error assessment

The influence of various parameters on the inversion results was investigated. This was mainly accomplished by evaluating errors introduced by uncertainties in the input parameters for the inversion methods, except for the statistical error in the plume inversion, which yields about 3 % for the total emission estimate and about 6 % maximum error for a single power plant emission and for the statistical error in the mass balance approach derived from the variability in the inversion results (see Sect. 4.2).

The wind direction of 147.5° for the inversion was determined by visually matching the best wind direction and minimizing the retrieved stability parameter a of the Gaussian plume inversion and hence minimizing the plume width relative to the centre axis (see Fig. 11). This also minimizes the total emission rate. However this minimum does not exactly coincide with the best fit (lowest RMS of differences between model and measurements) which is accomplished for about 145.5°. To account for the fact that the minimum in

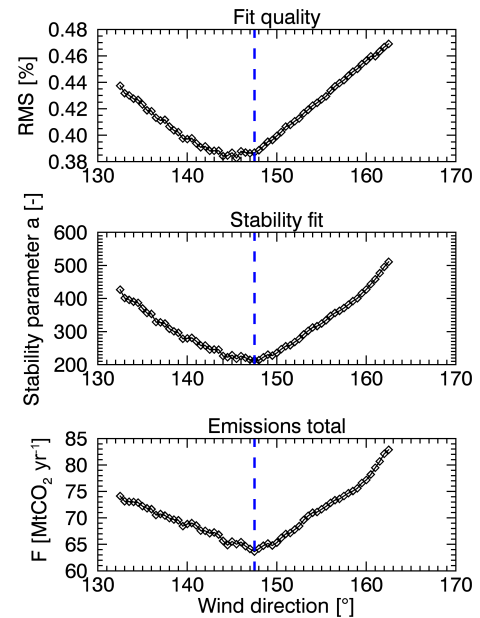


Figure 11. Sensitivity of the inverse Gaussian plume model to the wind direction for flux F , stability parameter a and the RMS between measurement and fit.

stability and fit quality is rather wide, an error of about $\pm 4^\circ$ was assumed. Importantly, Fig. 11 furthermore shows that, in principal, wind direction can be fitted directly to the data.

Several quality filters were applied to the data. The filter for fit quality (see Sect. 3.3.2) has been set relatively broad to reject only the data of poorest quality. It does not improve the results to relax the filter further or apply a stricter criteria to reject data. The signal filter was set to reject data with a signal below 13 000 counts (see Sect. 3.3.2). When relaxing the filter criteria by -2000 counts the total emission estimate is affected by about 6 % for the plume inversion and about 1 % for the mass balance approach. More strict filtering will significantly reduce the data which particularly impact the mass balance and the defined plume area and was therefore not applied.

The error due to uncertainties in the surface elevation model was not investigated because it largely cancels out for the proxy approach.

The overall errors (see Table 2) have been computed as the root sum square, assuming no correlation between errors and yielding about 10 % for the mass balance method (not yet including uncertainties based on variability between results from different flight legs as mentioned above) and 15 % for the plume inversion approach.

Table 2. Overview of maximum absolute values of the different error components of the estimated emission rates for the remote sensing results. The uncertainties are derived from a sensitivity analysis and the total error is the root sum square of the individual error components. In the case of the mass balance approach an additional uncertainty can be derived from the scatter of inversions for individual tracks (see Fig. 15 and Sect. 4.2).

Error source	Mass balance approach	Inverse plume model
Wind speed uncertainty (0.5 m s ⁻¹)	6 %	6 %
Statistical error (maximum)	*	6 %
Emission height (200–300 m a.g.l.)	1 %	0.5 %
Boundary layer depth (1200–1800 m a.m.s.l.)	6 %	6 %
Wind direction (143.5–151.5°)	4 %	6 %
Source width (100–500 m)	–	3 %
Grid size (15–55 m)	–	2 %
Signal filter (11 000–13 000 counts)	1 %	6 %
Inclination filter (5–10°)	1 %	5 %
CO ₂ background profile (390 ppm ± 1 %)	1 %	1 %
Total error	10 %	15 %

* Statistical errors for the mass balance approach are derived from the resulting emission rates by track in Sect. 4.2.

4 Results

4.1 In situ

The results of the in situ flux analysis are shown in Table 3. The average of the six methods of calculation (see Sect. 3.1 and 3.2.2) as the best estimate for the total emission flux is 51.6 MtCO₂yr⁻¹, of which 86 % was measured directly.

Emissions for all individual power plants could be derived. For directly measured fluxes (Niederaußem and total cluster in Table 3), the estimates have an uncertainty in the order of 10–20 %. Fluxes derived from differences and sums of the primary fluxes have a much higher uncertainty, especially when the plumes are overlapping (Neurath new and old). Frimmersdorf has a high relative uncertainty due to the fact that 45 % of the flux had to be derived from extrapolation.

4.2 Remote sensing

Wind direction was determined from the measured remote sensing data to about 147.5° both by visual inspection and by minimizing the stability parameter (see also Sect. 3.4). When fitting the stability parameter a to the retrieved XCO₂ this yields $a = 214.0$ (±8.8 % statistical error), i.e. stability class A (very unstable atmospheric conditions, Martin, 1976; Masters, 1998), independent of wind speed. This is in agreement with the observed convective mixing. However, the stability parameter obtained within the inversion is an effective parameter that also subsumes other effects such as increased flue gas temperature and variation in wind direction.

Using this stability as input for a vertical Gaussian dispersion model as a function of distance to the source to compute a weighted mean of the wind profile, an average wind speed

of 8.2 m s⁻¹ was obtained for the plume area covered by the first 9 downwind tracks of the MAMAP data. For this the in situ calibrated wind model data were used.

After applying this wind speed to the Gaussian plume inversion, the result for the average emission rate for the time of the overflight is in total about 63.6 MtCO₂yr⁻¹ (±3.0 % statistical error) split into 24.0 MtCO₂yr⁻¹ (±4.6 %), 14.2 (±6.3 %) and 25.4 MtCO₂yr⁻¹ (±5.2 %) for the power plants Niederaußem, Neurath new and Neurath old, respectively. The overall uncertainty including other components is about 15 % (see Sect. 3.4). The contour lines based on the multiple sources as an overlay on the retrieved XCO₂ can be seen in Fig. 12. Additionally, data and model result per flight leg are shown in Fig. 13. As mentioned before the evaluated tracks are all located upwind of the power plant Frimmersdorf. Therefore no emission rate for Frimmersdorf was derived with this approach.

For the mass balance approach a wind speed was computed for each individual track ranging from about 8 to 9 m s⁻¹. Similarly to the inverse plume model, the first 9 downwind tracks were analysed and the associated data are shown in Fig. 13. For three tracks further downwind (see Fig. 14) where the usual quality filtering could not be applied, the results have to be interpreted with more caution. At this distance, the plume is considerably wider than on the more upwind tracks so that there are fewer data available for the normalization.

The data were normalised for each flight track individually using a linear fit based on the data outside the plume (see Sect. 3.3.4). The definitions of the outside plume area are listed in Table 4.

The results are shown in Fig. 15 for the individual tracks and the average emission between power plants. Figure 15

Table 3. Summary of all standard methods applied to the different plumes measured in situ (for more details see the extended table in the Supplement). For the discussion see text in Sect. 3.1. The best estimate is the average of the sensitivity analyses. The minima and maxima were calculated combining the minima and maxima of either the primary measurements (e.g. for the total cluster, Frimmersdorf, and Niederaußem), or by using the worst-case combinations for sums and differences. The split into Neurath old and new was done as shown in Fig. 8.

Source	Best estimate (MtCO ₂ yr ⁻¹)	Min (MtCO ₂ yr ⁻¹)	Max (MtCO ₂ yr ⁻¹)	Error relative to best estimate	Fraction of extrapolation	Overall uncertainty
Total cluster	51.6	50.0	61.3	11 %	14 %	18 %
Frimmersdorf	1.7	1.3	2.7	41 %	45 %	63 %
Neurath old	16.8	13.3	20.8	22 %	20 %	32 %
Neurath new	7.3	3.9	11.3	50 %	20 %	60 %
Niederaußem	25.5	22.0	26.6	9 %	10 %	14 %
Weisweiler (ideal reference case)	18.4	16.3	20.9	13 %	0 %	13 %

Table 4. Normalisation distances to the end of the measurement track for each individual remote sensing transect.

Track	Distance to end of track	Comment
Upwind, downwind 3–5	2000 m	Baseline normalisation length for shorter tracks
Downwind 1	3000 m	Avoid measurements with increase in CH ₄ next to plume
Downwind 2	1500 m	Plume not centred on track
Downwind 6–9	3000 m	Track lengths increased
Downwind 10–12	2000 m	Plume is widened due to distance from the source

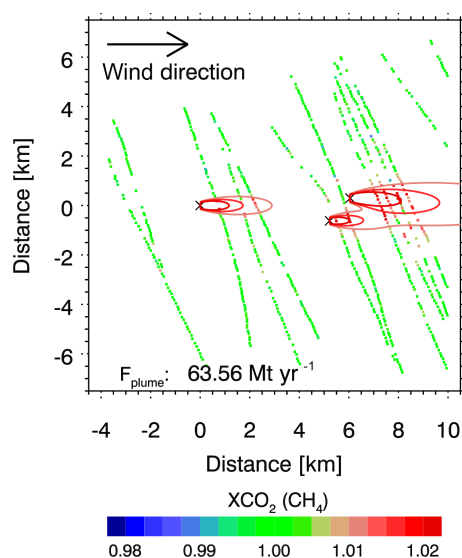


Figure 12. Gridded MAMAP XCO₂ results rotated so that wind direction points into positive x direction and contour lines of the inferred plume models for the individual power plants. Total emission rate is 63.6 MtCO₂yr⁻¹ for the time of the overflight. Ground scenes are shown slightly enlarged for better visibility.

also shows that there is basically no CO₂ influx from upwind of the measurement area. Shown in light grey are the tracks further downwind with decreased data quality. The average

absolute deviation from the mean was used as an indicator for the precision. The precision is worse for cases in which emissions are derived as differences which are subject to error propagation from emission estimates from upwind as well as downwind.

5 Discussion of results

The CO₂ emission rate estimates calculated using the different methods for the different power plants are shown in Fig. 16 and comprise the following: MAMAP remote sensing data analysed with inverse plume model and mass balance approach (Sect. 3.3), in situ data analysed using the presented mass balance method (Sect. 3.1) and emission rate estimates based on emission factors and energy production data for the time of the overflight. Error bars for the emissions derived from energy production are not shown. The error in power generation itself is generally about 1 % (compare Krings et al., 2011) and the annual error of derived emissions is required to be within 2.5 % (European Commission, 2007). The error for the time of the overflight is most likely not much larger. However, a study by Ackerman and Sundquist (2008) shows for individual US power plants that inventories based on monitoring of stack gases and inventories based on emission factors can in principal differ more than 20 %.

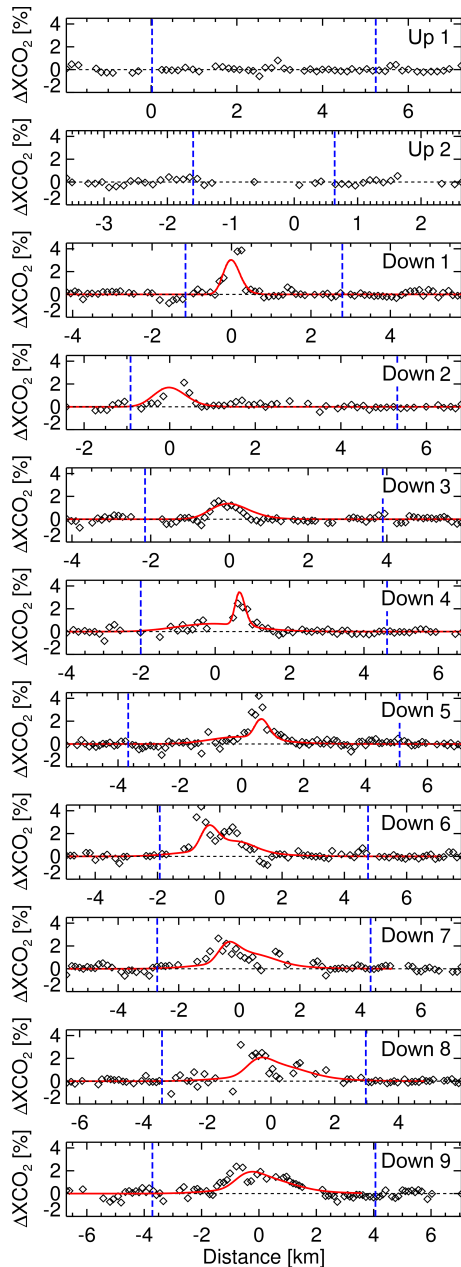


Figure 13. MAMAP XCO₂ transects for the tracks used for the emission rate estimates (black) and modelled Gaussian plume (red). The areas outside the dashed vertical lines denote the data that were used for the normalisation.

Generally the two inversion approaches for MAMAP agree very well within their uncertainties for the three individual power plants Niederaußem, Neurath old and Neurath new. However, for the mass balance approach, the uncertainties determined from the variability of emission estimates for individual power plants (see Fig. 15) are larger when differences were computed due to error propagation. This track-by-track variability is likely due to in-stationarity of the atmosphere and shows that repeated measurements are

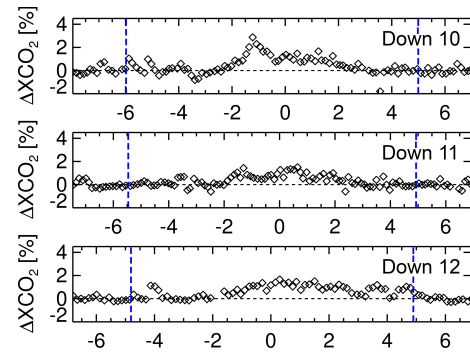


Figure 14. MAMAP XCO₂ measurement transects for the three flight legs downwind of the power plant Frimmersdorf. The area outside the dashed vertical lines denote the data that were used for the normalisation.

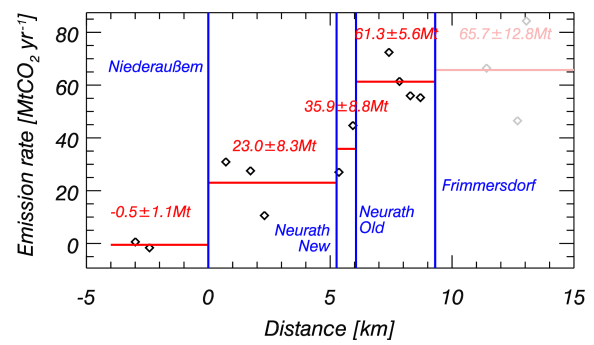


Figure 15. Mass balance results based on MAMAP remote sensing data. Vertical lines denote the location of power plants in a downwind direction from Niederaußem. Diamonds show emission rates derived from individual aircraft legs, where grey diamonds indicate reduced data quality (see Sect. 3.3.4). Horizontal lines and emission values show average total emissions of the upwind sources.

vital for obtaining accurate emission estimates. The inverse plume model, which inverts for all power plants simultaneously, is less affected by the in-stationarity since all available data are considered for all power plants, reducing the overall uncertainty.

When comparing with the CO₂ release computed from energy production the agreement is very good for all methods for the emissions from the power plant Niederaußem. For the two Neurath power plants, the remote sensing results indicate less emission from the new units and more from the old units, while the overall result is approximately the same. This is then also reflected in the emissions of Niederaußem, Neurath old and new, which is very similar for remote sensing methods and the computed emissions. The combined emissions of the three power plants are $63.6 \pm 15 \%$, $61.3 \pm 13 \%$ and $63.8 \text{ MtCO}_2\text{yr}^{-1}$ for the MAMAP plume inversion, MAMAP mass balance and the computed emissions. The relative difference to the computed emissions is therefore -0.3 (plume inversion) and -3.9% (mass balance). If

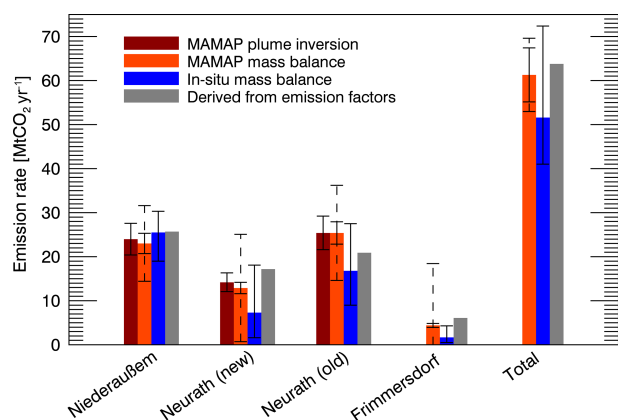


Figure 16. Inversion results compared to results obtained from emission factors and energy production for the time of the overflights. For the remote sensing mass balance approach the smaller error bars denote the uncertainties derived from the sensitivity analysis, while the larger also include the precision (see Fig. 15) by applying the root sum square. The error bars for in situ are worst-case limits based on the sensitivity analysis and half of the extrapolated emissions below and above of the captured plumes. Emissions from Frimmersdorf were not evaluated with the remote sensing plume inversion.

no in situ data had been available, that is, if the wind had been derived only from the COSMO-DE model, the errors would have been -6.4 and -10.0% respectively, reflecting the importance of additional wind measurements.

The total emissions derived from in situ data are in agreement with the emission computed from emission factors but are somewhat lower. The causes and uncertainties were explained in Sects. 3.2 and 4.1. However, the selection of the measurement day for detailed analysis was largely driven by the clear-sky requirement for remote sensing and there was no ideal overlap between the optimal measurements from the in situ instruments and suitable remote sensing measurement days. The case is similar for Neurath and Frimmersdorf, while for Niederaußern, where almost the complete plume could be captured, the agreement is exceptionally good.

The remote sensing mass balance results for power plant Frimmersdorf are based on data with less strict filters to obtain a sufficiently large data set to compute an emission rate. While the scatter and hence the uncertainty is quite large, the mean value indicates an emission rate of $4.4 \text{ MtCO}_2\text{yr}^{-1}$ which is close to the emission rate of $6.1 \text{ MtCO}_2\text{yr}^{-1}$ based on emission factors. However, the associated error based on the sensitivity analysis and precision is about $8.3 \text{ MtCO}_2\text{yr}^{-1}$.

6 Conclusions

This work enhances the comparison between measurement and inversion approaches using in situ and remote sensing

data to obtain emission rates for flue gases from a cluster of point sources with known locations. These sources were partly in close proximity to each other and the plumes of – in this case – CO₂ from coal-fired power plants overlapped, adding complexity to the inverse problem.

In contrast to the in situ method, the remote sensing measurements required clear-sky conditions at the time of the measurement. MAMAP measures solar backscattered electromagnetic radiation in the short-wave infrared. To simplify the radiative transfer calculations, cloud-free atmospheres are generally selected to avoid the radiative transfer issues associated with solar electromagnetic radiation passing through clouds. The selection of the measurement day for this study was largely driven by this requirement. This restriction impacted the selection of the measurement day for the analysis in this work. This resulted in some days with potentially more favourable conditions for the in situ method (coverage, flight restrictions, etc.) being disregarded. Nevertheless, the in situ measurements for the selected day allowed a good estimate of the emission rate when the extrapolation up to the limiting stable layer was applied.

Both remote sensing point source inversion methods are able to quantify the emissions within the error bars – to about 15% for the plume inversion approach and to about 13% for the mass balance method referring to the combination of the three power plants. The mass balance approach requires fewer parameters. It is, for example, not essential to know the exact source location and dimensions, which is an advantage for surveying unknown sources with a non-imaging instrument like MAMAP. However, the mass balance approach showed a lower precision when only a few flight legs per source are available, in particular close to the source and when differences between inversion results are interpreted. To mitigate this effect which is likely based on atmospheric in stationarity it is an advantage to gather measurements on multiple flight legs. This results in a higher precision and an improved error estimate.

One critical external input parameter for the analysis of the remote sensing data is wind information, which in this work was derived using model and in situ data. While the wind direction can be fitted to the data directly, this is not possible for the wind speed, which scales linearly with the emission rate.

The in situ inversion proves to be accurate for the power plant Niederaußern, where an almost complete sampling of the plume was possible. Further away from the sources, capturing the complete vertical plume extent in the higher reaching convective boundary layer was not possible due to airspace restrictions.

While the individual results for remote sensing and in situ measurements yield very similar results, provided sufficient sampling, a joint inversion approach may also complement the individual methods when there is no complete plume coverage. The uncertainty of in situ emission estimates for cases

with better coverage in less restricted places is better than 15 % (reference case in Table 3).

The methods presented here are demonstrated for CO₂ emissions from point sources. However, they are directly applicable in the same way to other gaseous compounds that disperse in the atmosphere and that have a lifetime longer than between emission and measurement, such as CH₄, which can also be derived from MAMAP remote sensing observations.

This case study illustrates the advantages and disadvantages of the used methods. The remote sensing approach needs clear-sky conditions but offers the possibility to perform many flight legs in a short period of time. This is necessary to reduce the uncertainty as can also be seen in Fig. 15. The multiple transects allow for the application of the Gaussian plume model to a multi-source set-up which simultaneously retrieves the emission rates from several sources.

While for MAMAP the plume model usually utilises a priori information on the source location, an imaging instrument with sufficient spatial resolution and sensitivity (similar to, for example, AVIRIS-NG Thompson et al., 2015; Frankenberg et al., 2016 but having a lower sensitivity compared to MAMAP) is able to determine the source location from the data directly and can acquire more data on shorter timescales, potentially reducing uncertainties in derived emission estimates. Furthermore imaging instruments offer the possibility of mapping large areas in a survey for unknown sources.

However, there is generally a need for external wind information which originates from models and/or in situ measurements. The analysis in this study shows an overestimated model wind speed of about 6 % (or about 0.4 m s⁻¹), which is smaller than the uncertainty in wind speed. So in this case, relying on the model alone may be sufficient. In a former study with a similar set-up (Krings et al., 2013) the error was about 10 % (or 0.7 m s⁻¹). A wider and more systematic analysis on the accuracy of the model wind is needed to assess to what extent additional wind (profile) measurements are dispensable. This will also become more relevant with regard to observations of localized sources by current and upcoming satellite missions with increased accuracy and spatial resolution. In these cases additional wind measurements will generally not be available.

In contrast to the remote sensing measurements of the entire vertical column, in situ measurements need to sample the plume with flight legs at different altitude levels. As a result of the time needed to complete a representative vertical cross section of measurements, only a limited number of repeated measurements are typically feasible. Interpolations within the cross sections and extrapolations to the surface and sometimes to the top of the plume have to be applied. This also applies to this study, in which the boundary layer reached into restricted airspace. However, the in situ method has the advantage of delivering vertically and horizontally resolved information in conjunction with co-located wind information, which can be readily used to infer a flux estimate.

The high intrinsic sensitivity also enables the detection of elevated trace gas levels at great distances from the source. Errors in the inversion results from in situ and remote sensing data are rather similar.

Data availability. All airborne remote sensing and in situ data acquired within the C-MAPEX campaign are available from the European Space Agency (ESA). A comprehensive description of the complete data set is given in the final report (Bovensmann et al., 2014) and the data can be requested from ESA on <https://earth.esa.int/web/guest/campaigns>.

Data from the COSMO-DE model can be obtained from the German Weather Service DWD via the Pamore database.

The reference values for CO₂ emission rates of the power plants were derived from power generation values of each power plant location kindly provided by RWE AG, Essen, Germany.

Surface in situ information from Düsseldorf airport were obtained from weather underground (<http://www.wunderground.com>).

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Competing interests. The authors declare that they have no conflict of interest.

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References

- Ackerman, K. V. and Sundquist, E. T.: Comparison of Two U.S. Power-Plant Carbon Dioxide Emissions Data Sets, *Environ. Sci. Technol.*, 42, 5688–5693, <https://doi.org/10.1021/es800221q>, 2008.

- Bovensmann, H., Buchwitz, M., Burrows, J. P., Reuter, M., Krings, T., Gerilowski, K., Schneising, O., Heymann, J., Tretner, A., and Erzinger, J.: A remote sensing technique for global monitoring of power plant CO₂ emissions from space and related applications, *Atmos. Meas. Tech.*, 3, 781–811, <https://doi.org/10.5194/amt-3-781-2010>, 2010.
- Bovensmann, H., Krings, T., Gerilowski, K., Neininger, B., Ruhtz, T., and Lindemann, C.: C-MAPEX Final Report – “Scientific and Technical Assistance for the Deployment of a flexible airborne spectrometer system during C-MAPEX”, ESA Study, 2014.
- Buchwitz, M., Rozanov, V. V., and Burrows, J. P.: A near-infrared optimized DOAS method for the fast global retrieval of atmospheric CH₄, CO, CO₂, H₂O, and N₂O total column amounts from SCIAMACHY Envisat-1 nadir radiances, *J. Geophys. Res.*, 105, 15231–15245, 2000.
- Buchwitz, M., Reuter, M., Bovensmann, H., Pillai, D., Heymann, J., Schneising, O., Rozanov, V., Krings, T., Burrows, J. P., Boesch, H., Gerbig, C., Meijer, Y., and Löscher, A.: Carbon Monitoring Satellite (CarbonSat): assessment of atmospheric CO₂ and CH₄ retrieval errors by error parameterization, *Atmos. Meas. Tech.*, 6, 3477–3500, <https://doi.org/10.5194/amt-6-3477-2013>, 2013.
- Cambaliza, M. O. L., Shepson, P. B., Caulton, D. R., Stirm, B., Samarov, D., Gurney, K. R., Turnbull, J., Davis, K. J., Possolo, A., Karion, A., Sweeney, C., Moser, B., Hendricks, A., Lauvaux, T., Mays, K., Whetstone, J., Huang, J., Razlivanov, I., Miles, N. L., and Richardson, S. J.: Assessment of uncertainties of an aircraft-based mass balance approach for quantifying urban greenhouse gas emissions, *Atmos. Chem. Phys.*, 14, 9029–9050, <https://doi.org/10.5194/acp-14-9029-2014>, 2014.
- Caulton, D. R., Shepson, P. B., Santoro, R. L., Sparks, J. P., Howarth, R. W., Ingraffea, A. R., Cambaliza, M. O. L., Sweeney, C., Karion, A., Davis, K. J., Stirm, B. H., Montzka, S. A., and Miller, B. R.: Toward a better understanding and quantification of methane emissions from shale gas development, *P. Natl. Acad. Sci. USA*, 111, 6237–6242, <https://doi.org/10.1073/pnas.1316546111>, 2014.
- Ciais, P., Crisp, D., van der Gron, H. D., Engelen, R., Janssens-Maenhout, G., Heimann, M., Rayner, P., and Scholze, M.: Towards a European Operational Observing System to Monitor Fossil CO₂ emissions, Final Report from the expert group, European Commission, 2015.
- Doms, G.: A Description of the Nonhydrostatic Regional COSMO-Model, Deutscher Wetterdienst, Technical Report, available at: <http://www.cosmo-model.org/> (last access: 31 January 2018), 2011.
- European Commission: Decision 2007/589/EC, establishing guidelines for the monitoring and reporting of greenhouse gas emissions pursuant to Directive 2003/87/EC of the European Parliament and of the Council, Official Journal of the European Union, L229, 1–85, 2007.
- Foken, T., Aubinet, M., Finnigan, J. J., Leclerc, M. Y., Mauder, M., and Paw, U. K. T.: Results Of A Panel Discussion About The Energy Balance Closure Correction For Trace Gases, *B. Am. Meteorol. Soc.*, 92, ES13–ES18, <https://doi.org/10.1175/2011BAMS3130.1>, 2009.
- Frankenberg, C., Meirink, J. F., van Weele, M., Platt, U., and Wagner, T.: Assessing Methane Emissions from Global Space-Borne Observations, *Science*, 308, 1010–1014, <https://doi.org/10.1126/science.1106644>, 2005.
- Frankenberg, C., Thorpe, A. K., Thompson, D. R., Hulley, G., Kort, E. A., Vance, N., Borchardt, J., Krings, T., Gerilowski, K., Sweeney, C., Conley, S., Bue, B. D., Aubrey, A. D., Hook, S., and Green, R. O.: Airborne methane remote measurements reveal heavy-tail flux distribution in Four Corners region, *P. Natl. Acad. Sci. USA*, 113, 9734–9739, <https://doi.org/10.1073/pnas.1605617113>, 2016.
- Gerilowski, K., Tretner, A., Krings, T., Buchwitz, M., Bertagnolio, P. P., Belemezov, F., Erzinger, J., Burrows, J. P., and Bovensmann, H.: MAMAP – a new spectrometer system for column-averaged methane and carbon dioxide observations from aircraft: Instrument description and performance analysis, *Atmos. Meas. Tech.*, 4, 215–243, <https://doi.org/10.5194/amt-4-215-2011>, 2011.
- Gordon, M., Li, S.-M., Staebler, R., Darlington, A., Hayden, K., O’Brien, J., and Wolde, M.: Determining air pollutant emission rates based on mass balance using airborne measurement data over the Alberta oil sands operations, *Atmos. Meas. Tech.*, 8, 3745–3765, <https://doi.org/10.5194/amt-8-3745-2015>, 2015.
- Hiller, R. V., Neininger, B., Brunner, D., Gerbig, C., Bretscher, D., Künzle, T., Buchmann, N., and Eugster, W.: Aircraft-based CH₄ flux estimates for validation of emissions from an agriculturally dominated area in Switzerland, *J. Geophys. Res.*, 119, 4874–4887, <https://doi.org/10.1002/2013JD020918>, 2014.
- Karion, A., Sweeney, C., Pétron, G., Frost, G., Michael Hardesty, R., Kofler, J., Miller, B. R., Newberger, T., Wolter, S., Banta, R., Brewer, A., Dlugokencky, E., Lang, P., Montzka, S. A., Schnell, R., Tans, P., Trainer, M., Zamora, R., and Conley, S.: Methane emissions estimate from airborne measurements over a western United States natural gas field, *Geophys. Res. Lett.*, 40, 4393–4397, <https://doi.org/10.1002/grl.50811>, 2013.
- Krautwurst, S., Gerilowski, K., Jonsson, H. H., Thompson, D. R., Kolyer, R. W., Iraci, L. T., Thorpe, A. K., Horstjann, M., Eastwood, M., Leifer, I., Vigil, S. A., Krings, T., Borchardt, J., Buchwitz, M., Fladland, M. M., Burrows, J. P., and Bovensmann, H.: Methane emissions from a Californian landfill, determined from airborne remote sensing and in situ measurements, *Atmos. Meas. Tech.*, 10, 3429–3452, <https://doi.org/10.5194/amt-10-3429-2017>, 2017.
- Krings, T., Gerilowski, K., Buchwitz, M., Reuter, M., Tretner, A., Erzinger, J., Heinze, D., Pflüger, U., Burrows, J. P., and Bovensmann, H.: MAMAP – A new spectrometer system for column-averaged methane and carbon dioxide observations from aircraft: retrieval algorithm and first inversions for point source emission rates, *Atmos. Meas. Tech.*, 4, 1735–1758, <https://doi.org/10.5194/amt-4-1735-2011>, 2011.
- Krings, T., Gerilowski, K., Buchwitz, M., Hartmann, J., Sachs, T., Erzinger, J., Burrows, J. P., and Bovensmann, H.: Quantification of methane emission rates from coal mine ventilation shafts using airborne remote sensing data, *Atmos. Meas. Tech.*, 6, 151–166, <https://doi.org/10.5194/amt-6-151-2013>, 2013.
- Lavoie, T. N., Shepson, P. B., Cambaliza, M. O. L., Stirm, B. H., Karion, A., Sweeney, C., Yacovitch, T. I., Herdon, S. C., Lan, X., and Lyon, D.: Aircraft-Based Measurements of Point Source Methane Emissions in the Barnett Shale Basin, *Environ. Sci. Technol.*, 49, 7904–7913, <https://doi.org/10.1021/acs.est.5b00410>, 2015.

- Martin, D. O.: The Change of Concentration Standard Deviations with Distance, *J. Air Poll. Control Assoc.*, 26, 145–147, <https://doi.org/10.1080/00022470.1976.10470238>, 1976.
- Masters, G. M.: Introduction to Environmental Engineering and Science, Prentice-Hall, Inc., 2nd Edn., 1998.
- Mays, K. L., Shepson, P. B., Stirm, B. H., Karion, A., Sweeney, C., and Gurney, K. R.: Aircraft-Based Measurements of the Carbon Footprint of Indianapolis, *Environ. Sci. Technol.*, 43, 7816–7823, <https://doi.org/10.1021/es901326b>, 2009.
- NRC: National Research Council (NRC) – Committee on Methods for Estimating Greenhouse Gas Emissions, Verifying Greenhouse Gas Emissions: Methods to Support International Climate Agreements, ISBN 0-309-15212-7, available at: <http://www.nap.edu/catalog/12883.html> (last access: 31 January 2018), 2010.
- Reuter, M., Buchwitz, M., Schneising, O., Hase, F., Heymann, J., Guerlet, S., Cogan, A. J., Bovensmann, H., and Burrows, J. P.: A simple empirical model estimating atmospheric CO₂ background concentrations, *Atmos. Meas. Tech.*, 5, 1349–1357, <https://doi.org/10.5194/amt-5-1349-2012>, 2012.
- Rothman, L., Gordon, I., Babikov, Y., Barbe, A., Benner, D. C., Bernath, P., Birk, M., Bizzocchi, L., Boudon, V., Brown, L., Campargue, A., Chance, K., Cohen, E., Coudert, L., Devi, V., Drouin, B., Fayt, A., Flaud, J.-M., Gamache, R., Harrison, J., Hartmann, J.-M., Hill, C., Hodges, J., Jacquemart, D., Jolly, A., Lamouroux, J., Roy, R. L., Li, G., Long, D., Lyulin, O., Mackie, C., Massie, S., Mikhailenko, S., Müller, H., Naumenko, O., Nikitin, A., Orphal, J., Perevalov, V., Perrin, A., Polovtseva, E., Richard, C., Smith, M., Starikova, E., Sung, K., Tashkun, S., Tennyson, J., Toon, G., Tyuterev, V., and Wagner, G.: The {HITRAN2012} molecular spectroscopic database, *J. Quant. Spectrosc. Ra.*, 130, 4–50, <https://doi.org/10.1016/j.jqsrt.2013.07.002>, 2013.
- Schepers, D., Guerlet, S., and Butz, A.: Methane retrievals from Greenhouse Gases Observing Satellite (GOSAT) shortwave infrared measurements: Performance comparison of proxy and physics retrieval algorithms, *J. Geophys. Res.*, 117, D10307, <https://doi.org/10.1029/2012JD017549>, 2012.
- Thompson, D. R., Leifer, I., Bovensmann, H., Eastwood, M., Fladland, M., Frankenberg, C., Gerilowski, K., Green, R. O., Kratwurst, S., Krings, T., Luna, B., and Thorpe, A. K.: Real-time remote detection and measurement for airborne imaging spectroscopy: a case study with methane, *Atmos. Meas. Tech.*, 8, 4383–4397, <https://doi.org/10.5194/amt-8-4383-2015>, 2015.
- Thorpe, A. K., Frankenberg, C., and Roberts, D. A.: Retrieval techniques for airborne imaging of methane concentrations using high spatial and moderate spectral resolution: application to AVIRIS, *Atmos. Meas. Tech.*, 7, 491–506, <https://doi.org/10.5194/amt-7-491-2014>, 2014.
- Tratt, D. M., Buckland, K. N., Hall, J. L., Johnson, P. D., Keim, E. R., Leifer, I., Westberg, K., and Young, S. J.: Airborne visualization and quantification of discrete methane sources in the environment, *Remote Sens. Environ.*, 154, 74–88, <https://doi.org/10.1016/j.rse.2014.08.011>, 2014.
- Turner, D. B.: Workbook of Atmospheric Dispersion Estimates, U.S. Department of Health, Education, and Welfare, 95 pp., 1970.
- U.S. Committee on Extension to the Standard Atmosphere: U.S. Standard Atmosphere, 1976, National Oceanic and Atmospheric Administration, National Aeronautics and Space Administration, United States Air Force, Washington D.C., 241 pp., 1976.
- Young, S.: Detection and quantification of gases in industrial-stack plumes using thermal infrared hyperspectral imaging, Aerospace Report No. ATR-2002(8407)-1. El Segundo, Calif: The Aerospace Corporation, 2002.