

1 **Intraspecific trade-offs between facilitation and competition in the non-native mollusc**

2 *Crepidula fornicata*

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6 Running page head: Facilitation in invasive gastropods

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17 species . Population dynamics

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23 **ABSTRACT**

24 We tested the hypothesis that high density populations of the non-indigenous gastropod
25 *Crepidula fornicata* increase settlement of conspecific recruits (facilitation process), and that
26 this facilitation is balanced by competition processes. To verify our hypothesis, we sampled
27 *Crepidula* at two drastically different densities for 10 years. We found that at high densities,
28 the number of one-year old individuals per m² colonizing the habitat and individual growth
29 performances were higher in comparison with the low density condition (Allee effect). On the
30 contrary, the production/biomass ratio (P/B), a good indicator of population fitness, was lower
31 at higher densities. We relate this lower P/B ratio to the deficit of young individuals compared
32 to adult biomass. We conclude that the net effect of high density on the conspecific
33 colonisation processes of the population is positive, thanks to the higher available hard
34 substrate for larvae (facilitation). However, intraspecific competition also occurs and
35 mitigates this positive effect. Therefore, we suggest that it is particularly important to take
36 into account the “net” balance between costs and benefits (i.e. what we observe) when
37 analyzing population growth, in order to better understand its dynamics.

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40 INTRODUCTION

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42 Population demography within communities is thought to be strongly shaped by physical
43 disturbances and negative interactions like competition, predation and parasitism (Hixon et al.
44 2002). However, other studies have highlighted the importance of positive interactions in
45 population and community dynamics (Bruno et al. 2003). In particular, one of these
46 interactions is facilitation, generally defined as all non-consumer interactions among two or
47 more species (named ‘foundation species’) that positively affect at least one of the species
48 involved (Bertness & Callaway 1994), often with major cascade effects on recipient
49 ecosystems (Gribben et al. 2009, Gouhier et al. 2011, Yakovis & Artemieva 2017). Therefore,
50 facilitation occurs when one organism makes the local environment more favourable for
51 another species. This process is mostly described for sessile organisms, which cannot actively
52 escape stress (by locomotion) and consequently profit from the presence of other foundation
53 species mitigating adverse conditions (Bruno & Kennedy 2000, Wonham et al. 2005, Bulleri
54 2009, Branch et al. 2010).

55 Facilitation may consist of providing a suitable substrate for other species, as is the case for
56 the gastropod *Batillaria attramentaris* in its introduced range in Washington, USA. Its
57 proliferation, in fact, increases hard substrate surface (shells) for other species, including
58 *Crepidula convexa* (Wonham et al. 2005). In many other cases, facilitation consists of
59 mitigating physical stress, like reducing wave related physical disturbance either in close
60 vicinity of the foundation species or at a larger scale. Cordgrass (*Spartina alterniflora*), for
61 example, modifies cobble beaches nearby by reducing flow speed, stabilizing the sediments
62 and allowing seedlings of other plants to establish (Bruno & Kennedy 2000). *Mytilus edulis*
63 beds, instead, break the wave energy and facilitate cockle, *Cerastoderma edule*, settlement

64 50-100 m towards the coast (Donadi et al. 2013). Even though facilitation is often defined as a
65 positive interaction among two or more species (Bertness & Callaway 1994), this biotic factor
66 can also act within a single population, in the same way negative interactions such as
67 competition or predation do (i.e. intraspecific competition or cannibalism, respectively). In
68 this regard, Allee effects describe situations where density of conspecifics is positively
69 correlated with population fitness (Courchamp et al. 1999, Stephens & Sutherland 1999), in
70 contradiction with Verhulst's logistic equation, which states mathematically that population
71 size has a negative effect on its own per capita growth rate (Hixon et al. 2002). Besides, Allee
72 effect is often considered as a major mechanism facilitating non-indigenous species
73 expansion, at least when those have already overpass the first settlement and dispersal stages
74 (Simberloff & Von Holle 1999, Wonham et al. 2005) or other types of invasions (Shaw et al.
75 2018). One important output concerns the 'left part' of the theoretical curve of Allee effects,
76 which supports the principle that low conspecific density may lead to population extinction or
77 decreased fitness by genetic inbreeding and loss of heterozygosity or demographic
78 stochasticity (Courchamp et al. 1999). However, few studies focus on the "central part" of the
79 curve that shows the positive effect on population growth at high conspecific densities (i.e.
80 inverse density dependence), beyond the threshold where negative effects would be expected.
81 Authors who assessed Allee effects generally concentrate on life history traits of populations
82 like density/mortality rate (Leslie 2005, Svanfeldt et al. 2017), growth (Tsai et al. 2010) or
83 reproductive output (Leslie 2005). In the present work, based on a 10-yr monitoring of a
84 marine benthic population, we propose to estimate a more comprehensive parameter
85 describing population fitness. – the Production/Biomass (P/B) ratio – in conditions of low
86 and high conspecific densities. The P/B ratio is considered a good indicator of energetic flow
87 in an ecosystem (Brey 1990). It typically estimates the percentage of annually renewed
88 biomass, and consequently integrates recruitment success, individual growth and mortality

89 rate. Brey (1990) reviewed different correlations (or dependences) between P/B ratios and
90 biological features of macrozoobenthic populations, such as age at sexual maturity, lifespan
91 and number of generations. Moreover, the P/B ratio is an indicator of population health. For
92 example, the P/B ratio of a population of cockles, *Cerastoderma edule*, was 27% decreased by
93 the presence of parasites (Gam et al. 2009). Our study also aimed to highlight the need of
94 clarifying the definitions of competition and facilitation. In fact, one definition of facilitation
95 states that facilitative or positive interactions are encountered between organisms when at
96 least one of the participants benefits without “causing harm to neither” (Bruno et al. 2003).
97 This definition seems antagonistic with the one stating that facilitation appears when the
98 “benefits exceed the costs” (Stachowicz 2001).

99 Our biological model is the slipper limpet, *Crepidula fornicata*, a gastropod which was
100 accidentally introduced in Europe in 1872 from the East coast of the United States (Blanchard
101 1997), and is introduced to many coastal areas (Stiger-Pouvreau & Thouzeau 2015). *C.*
102 *fornicata* appeared in France in the 1940s (Blanchard 1995) and was first recorded in 1969 in
103 Arcachon Bay (Bachelet et al. 1980). These organisms form stacks where up to 12 individuals
104 are piled up (Coum 1979). They are protandric hermaphrodites and exhibit planktotrophic
105 development. Larvae need to reach a specific size in order to attain metamorphic competence.
106 Then, they settle on hard substratum, e.g. conspecific live or dead shells (Henry & Lyons
107 2016). Paradoxically, although its planktonic larvae necessitate a hard substrate to settle, *C.*
108 *fornicata* is particularly prolific in soft-substrate habitats. Thus, our hypothesis is that we
109 could observe facilitative processes related to the increase of hard substrates (represented by
110 *C. fornicata*'s shells) with increasing *C. fornicata* density in a soft-sediment environment.
111 Conversely, a high density of conspecifics would beget competition leading to limited growth,
112 higher mortality, and less successful recruitment. Our objective is to assess the net effect of
113 density on population fitness through the P/B ratio and the settlement of one-year individuals.

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116 **METHODS**

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119 **Study location and sampling**

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122 All samples were collected from an intertidal mudflat of Arcachon Bay, France
123 (44°39'18.00"N - 1°08'24.62"W). *Crepidula fornicata* are aggregately distributed within a
124 narrow band of 450 × 15 m² along the lowest intertidal levels which can be reached by foot at
125 ebb-tide, when tidal coefficient > 100. Salinity and water temperature seasonally fluctuate in
126 the range 22–32, and 1–25°C, respectively. In this area, patches with high density of *C.*
127 *fornicata* (ca. 8-m-diameters) are separated by areas where it is rare. Once a year from 2001
128 to 2010, during the same period (February-March), eight quadrats were sampled and washed
129 on a 5-mm mesh sieve. The size of the quadrats was selected according to *C. fornicata*
130 density: four 0.04-m² quadrats were situated in the middle of a patch (“HIGH” density), while
131 four 0.25-m² quadrats were sampled at least 20 m from the closest high density patch, where
132 *C. fornicata* are rare (“LOW” density). In order to obtain a clear quantitative difference
133 between “LOW” and “HIGH” treatments, we defined “LOW” those samples characterized by
134 less than 2 kg Fresh Weight (FW) of live *C. fornicata*.m⁻² and “HIGH” those having more
135 than 7 kg FW.m⁻². The accuracy of our GPS did not allow to detect whether high density
136 patches were at the exact same location each year, but the whole sampling area (see above)
137 was always within the same mudflat. This period of sampling is just before *C. fornicata*

138 begins seasonal reproduction (de Montaudouin et al. 2002, Richard et al. 2006, Bohn et al.
139 2012) and was selected to ensure that our youngest individuals were ca. 1-yr old.

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142 **Cohort determination**

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145 Every *C. fornicata* (n = 10,071) was measured with a calliper (1-mm precision) at the lower
146 edge (straight shell length) and the position in the stack was noted. Former studies have noted
147 the peculiar fact that *C. fornicata* forms stacks with an average of one individual per year
148 settling (Coum 1979). Thus the position of an individual in a stack could provide its age
149 (Deslous-Paoli 1985). An isolated individual (i.e. an individual that is not settled on the top of
150 another) is considered as a stack of one specimen. However, a recent study, performed in the
151 same area as this one, showed that more than one individual per year could settle in a single
152 stack and proposed a model to transform a stack/position matrix into a stack/age matrix (de
153 Montaudouin & Accolla 2017). The R 3.3.1 script of this model (Team 2014) is available in
154 de Montaudouin & Accolla (2017). Consequently, we used this model to identify the different
155 cohorts and calculate population dynamics parameters per year (10 years from 2001 to 2010)
156 and per *C. fornicata* abundance (LOW and HIGH).

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159 **Population dynamics**

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Density, Mortality

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163 We calculated the mortality rate of every *C. fornicata* cohort (Z , yr^{-1}) for each year Y (2001 to
164 2010) and for each of the density conditions (LOW and HIGH *C. fornicata* density) by
165 comparing the density N at age t with density N at age $t-1$. The theoretical density of recruits
166 N_0 (ind m^{-2}) was deduced from the exponential mortality model:

$$167 \quad N_t = N_0 e^{-Zt} \quad (1)$$

168 where N_t is the number of individuals of a cohort at age t (yr).

169 Then, we calculated the percentage of 1-year individuals as the ratio of the number of 1-year
170 individuals over the total number of individuals, $\times 100$. This value is a proxy of the success of
171 recruitment related to the density of adults, ca. one year after settlement.

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173 Individual growth

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175 Von Bertalanffy growth function (VBGF) predicts length as a function of age and is
176 commonly used in gastropod growth analysis with or without seasonal correction (Richardson
177 et al. 2005, Chatzinikolaou & Richardson 2008), including examples for *C. fornicata* (Coun
178 1979, Deslous-Paoli 1985, de Montaudouin et al. 2001, de Montaudouin & Accolla 2017). Its
179 formulation is:

$$180 \quad L_t = L_\infty (1 - e^{-K(t)}) \quad (2)$$

181 where L_∞ is the asymptotic shell length (mm) and K the growth coefficient (yr^{-1}).

182 For each year Y (2001 to 2010) and each of the density conditions (LOW and HIGH *C.*
183 *fornicata* density), we measured growth increment by comparing straight shell length at age t
184 and straight shell length at age $t-1$. The VBGF parameters, L_∞ and K , were assessed using
185 FISAT II software (version 1.2.2, FAO-ICLARM).

186 A growth performance index ϕ' was also calculated using K and L_∞ (Pauly & Munro 1984),
187 as:

188 $\phi' = 2\log(L_\infty) + \log K$ (3)

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Biomass, Production, P/B

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192 Biomass was calculated from individual shell length with the following equation based on 45
193 specimens ($r^2 = 0.92$) (de Montaudouin & Sauriau 1999):

194 $\log DW = 2.15 \log L - 4.17$ (4)

195 where DW is shell-free dry weight (g) and L the straight shell length (mm). Dry weight was
196 obtained after dissecting the flesh and drying it at 60°C for 48 h.

197 Total annual production P was calculated for *C. fornicata* according to incremental
198 summation method for populations with non-continuous recruitment and distinguishable year-
199 classes (Crisp 1984):

200 $P = \sum_{t=0}^{t=n} [(N_t + N_{t+1})/2 \times (W_{t+1} - W_t)]$ (5)

201 where N_t is *C. fornicata* abundance at age t (yr) and W_t is individual dry weight at age t (g).

202 P/B (yr^{-1}) was calculated, dividing production P by mean biomass B , where B is defined as the
203 average *C. fornicata* biomass at year Y .

204

205 **Statistical analysis**

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207 All parameters (density N , mortality Z , % of 1-year individuals, asymptotic shell length L_∞ ,
208 growth coefficient K , growth performance index ϕ' , biomass B , production P , P/B ratio) were
209 compared according to *C. fornicata* sampling conditions (LOW vs. HIGH), using the non-
210 parametric Wilcoxon paired test, with years as paired replicates ($n = 10$ years). Wilcoxon
211 paired test was chosen because we sampled the same population of *C. fornicata* each year,
212 with LOW density patches being geographically close to HIGH density patches.

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215 **RESULTS**

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217 Length-frequency histograms were obtained for each of the ten investigated years,
218 discriminating high *Crepidula fornicata* density patches from low *C. fornicata* density areas
219 (Fig. 1). Mean density was 4.5 times higher in the HIGH density condition than in the LOW
220 density condition ($W = 2.80$, $p = 0.005$) (Table 1). The relative density of 1-year individuals
221 was 2.6 lower in the HIGH density condition than in the LOW density condition ($W = 2.80$, p
222 $= 0.005$) (Fig. 1 & 2, Table 1). However, the density of 1-year individuals is 2.5 higher in the
223 HIGH density condition than in the LOW density condition (Table 1). Mortality rate Z was
224 similar for both conditions ($W = 0.10$, $p = 0.919$). Asymptotic length L_{∞} was similar in both
225 conditions ($W = 1.38$, $p = 0.169$), but growth coefficient K was 15% higher in the HIGH
226 density condition ($W = 2.80$, $p = 0.005$) (Fig. 3). As a consequence, the growth performance
227 index Φ' was 1.4% higher in the HIGH density condition ($W = 2019$, $p = 0.028$) (Table 1).
228 Mean biomass and production were 5.4 and 4.2 times higher in the HIGH density condition
229 than in LOW density condition respectively ($W = 2.80$, $p = 0.005$) (Table 1). Conversely, P/B
230 ratio was 23% lower in the HIGH density condition than in the LOW density condition ($W =$
231 2.50 , $p = 0.013$) (Fig. 4, Table 1).

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233

234 **DISCUSSION**

235

236 The LOW and HIGH density conditions discriminated two drastically different situations of
237 *Crepidula fornicata* biomass and density. The HIGH density condition displayed density

238 values that are currently found in areas with longer histories of *C. fornicata* invasion (Hily
239 1991, de Montaudouin & Sauriau 1999, de Montaudouin et al. 2001, Guérin 2004, de
240 Montaudouin & Accolla 2017), while the LOW density condition rather described a process
241 of colonization (de Montaudouin et al. 2001). Likewise, all parameter values (L_{∞} , K , ϕ' , P and
242 B) but the P/B ratio are within the range of what is already assessed for this species. The P/B
243 ratio, instead, displayed particularly low values in both HIGH and LOW density conditions.
244 Indeed, we obtained values between 0.21 and 0.45 yr⁻¹, while the expected P/B values for
245 invertebrates are usually not smaller than 1.00 yr⁻¹: P/B = 1.77 yr⁻¹ for molluscs, P/B = 1.82
246 yr⁻¹ for filter feeders and P/B = 1.00 for mollusc filter feeders (Cusson & Bourget 2005).
247 Moreover, computing models, which provide an estimation of the P/B ratio according to the
248 lifespan or to the maximum individual weight (Robertson 1979, Tumbiolo & Downing 1994),
249 gives for *C. fornicata* values between 1.18 and 1.28 yr⁻¹. *C. fornicata* is usually not <0.50 yr⁻¹
250 (de Montaudouin & Accolla 2017). This discrepancy could be attributed to the rather
251 uncommon position of the *C. fornicata* population: our samples were at the intertidal level,
252 while this species usually lives at the subtidal level (Loomis & VanNieuwenhuize 1985).
253 The comparison of population dynamics parameter values between LOW and HIGH density
254 conditions provided a new insight in the facilitation/competition processes of species with
255 particularly high performances, such as invasive molluscs. The well-established density-
256 dependent intraspecific competition concept is not fully valid in the present case. Firstly and
257 unexpectedly, growth performance was not affected by high density, but on the contrary was
258 slightly stimulated. This result contradicts many studies demonstrating that high suspension-
259 feeders density is usually related to stunted growth (Peterson & Black 1987, Peterson & Beal
260 1989, Jensen 1992, Kamermans et al. 1992). Moulin et al. (2007) observed a higher particle
261 resuspension (i.e. higher available food) during spring tides when fewer stacks of *C. fornicata*
262 were on the bottom. In our sheltered study site, instead, current speed is low and resuspension

263 of benthic diatoms, a large part of *C. fornicata* trophic regime (Riera et al. 2002), could rather
264 be related to bioturbation activity due to the grazing activity of the younger stage (Chaparro et
265 al. 1998) and to the lifting behaviour of the older individuals (Diederich et al. 2015). Thus, the
266 higher the density, the higher the bioturbation and the microphytobenthic availability. *C.*
267 *fornicata* growth can also be stimulated by the contact of conspecifics (Le Gall 1980), such
268 that high density consequently promotes these processes. Finally, we were not able to measure
269 competitive effects on shell growth, although there was certainly trophic competition at these
270 high levels of biomass, as observed with another suspension-feeder, the cockle *Cerastoderma*
271 *edule* (de Montaudouin & Bachelet 1996). In conclusion, facilitation was the dominating
272 process influencing growth.

273 The density values of 1-year old individuals clearly illustrate the antagonistic effects of
274 competition and facilitation. In the LOW density condition, 1-year old individuals were on
275 average 38% of total population (i.e. 219 ind.m⁻² out of 575), which corresponded to a [1-yr
276 old]/[>1-yr old] ratio of 0.61. If we extrapolate this ratio to the HIGH density condition, we
277 should obtain an average density of 1-yr old individuals of 1347 ind.m⁻². This would be the
278 pure consequence of a gross facilitation process: the higher the availability of suitable
279 substrate (here conspecific shells density), the higher the larval recruitment. However, the
280 observed density of 1-yr old individuals was only 375 ind.m⁻² in the HIGH density condition
281 (14.5% of total density). This could be caused by intraspecific competition and by a
282 consequent alteration of recruitment by suspension-feeding activity of previously settled
283 conspecific adults, as observed for many other marine species (Woodin 1976, Bachelet et al.
284 1992, de Montaudouin & Bachelet 1996). Moreover, cannibalism in suspension-feeders has
285 been commonly observed in the laboratory (Young & Chia 1987), and in particular for *C.*
286 *fornicata* (Pechenik et al. 2004). This relative deficit of younger individuals in the HIGH

287 density condition explained why the P/B ratio was 23% lower than in the LOW density
288 condition.

289 In conclusion, this 10-yr study showed that a 5.4 fold increase of biomass facilitated
290 settlement of 1-year individuals ($\times 6.1$), but this facilitation was modulated by strong
291 intraspecific competition ($\div 3.6$). The net result was positive ($\times 1.7$): facilitation (Allee effect)
292 prevails over competition in terms of colonisation success in high densities. This self-
293 sustained growth of a population, exacerbated by this peculiarity to live piled up in stacks, is
294 certainly a major reason explaining the success of *C. fornicata* in becoming invasive
295 (Blanchard 2009, Stiger-Pouvreau & Thouzeau 2015), but also the success of other sessile
296 species needing hard substrates to settle in environments dominated by soft bottoms
297 (Diederich 2005, Leslie 2005, Lang & Buschbaum 2010, Svanfeldt et al. 2017). Conversely,
298 we also demonstrated that, even though greater density of *C. fornicata* increases population
299 growth (Allee effect), the P/B ratio at high density is diminished, suggesting that a density
300 threshold beyond which fitness is altered has been exceeded in the HIGH density condition.
301 Our study provides evidence for the importance of examining “net” balance between costs
302 and benefits (i.e. what we observe) when analysing population growth. This means that we
303 should stress what is the net competition when costs are high, albeit gross facilitation is
304 present too, or what is the net facilitation when benefits are high in presence of gross
305 competition (Bruno et al. 2003, Leslie 2005).

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308 All applicable institutional and/or national guidelines for the care and use of animals were followed.

309

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461 Fig. 1. Shell length frequency histogram of *Crepidula fornicata* from 2001 to 2010 at the LOW and
462 HIGH density conditions, and by age. Black bars concern the 1-yr individuals.

463 Fig. 2. Mean relative proportion of 1-yr ind.m⁻² (+ 1 standard error) of *Crepidula fornicata* at the LOW
464 and HIGH densities condition, from 2001 to 2010 (N = 4 replicates).

465 Fig. 3. Modelled growth curve of *Crepidula fornicata* with average Von Bertalanffy growth function
466 parameters (2001-2010) at the LOW and HIGH densities conditions (Table 1).

467 Fig. 4. Mean Production/Biomass ratio of *Crepidula fornicata* at the LOW and HIGH densities
468 condition, from 2001 to 2010.

469

Figure 1

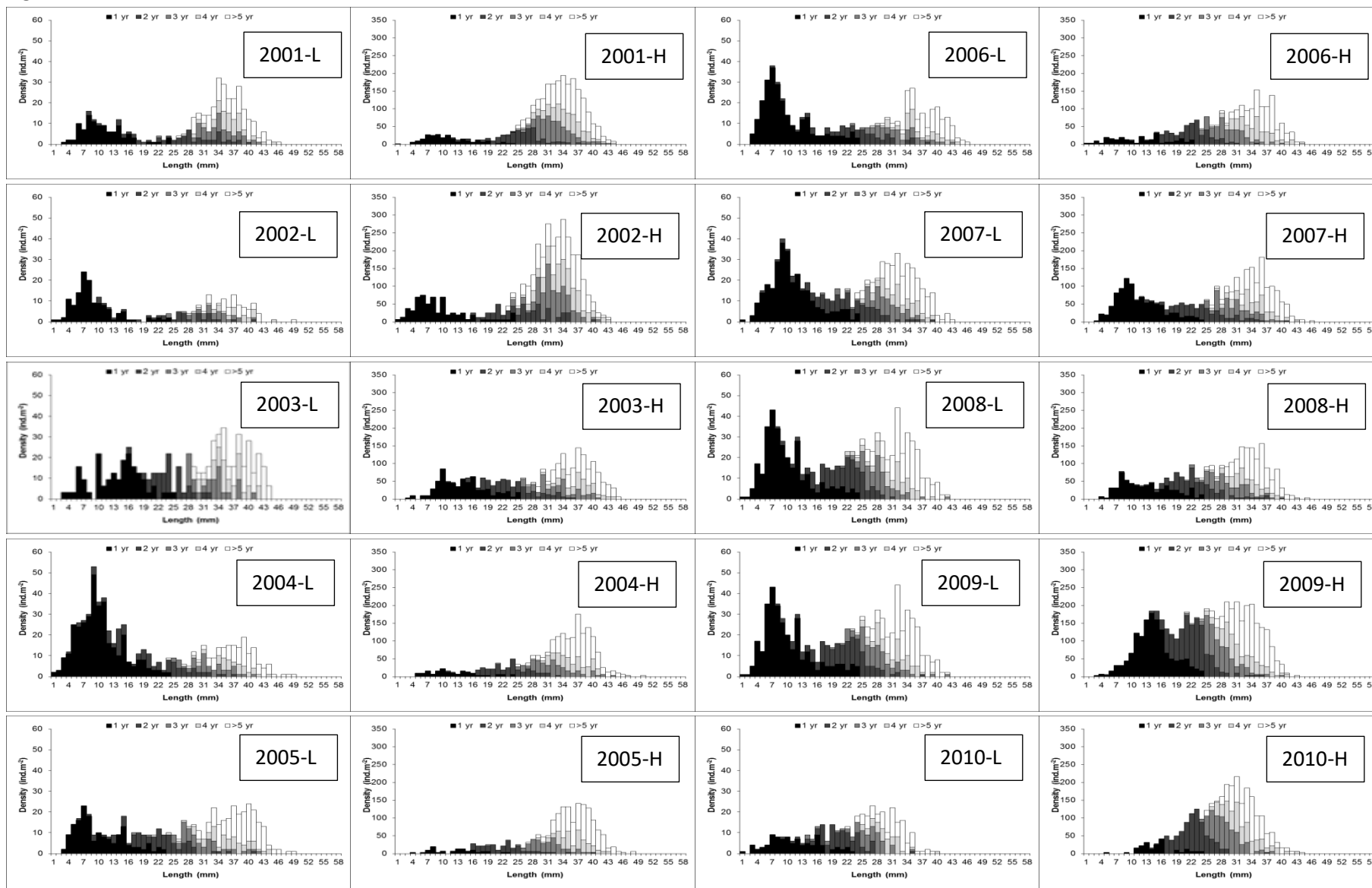


Figure 2

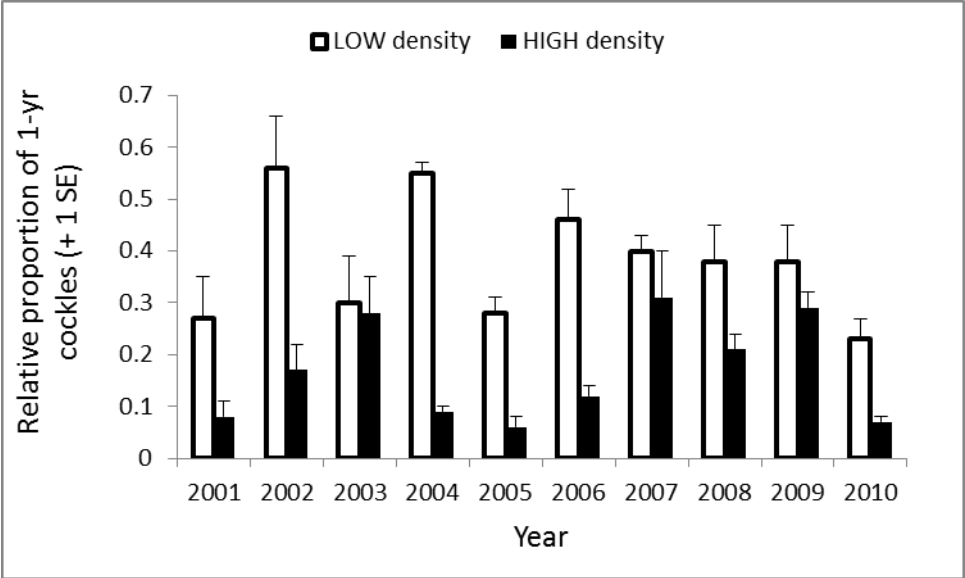


Figure 3

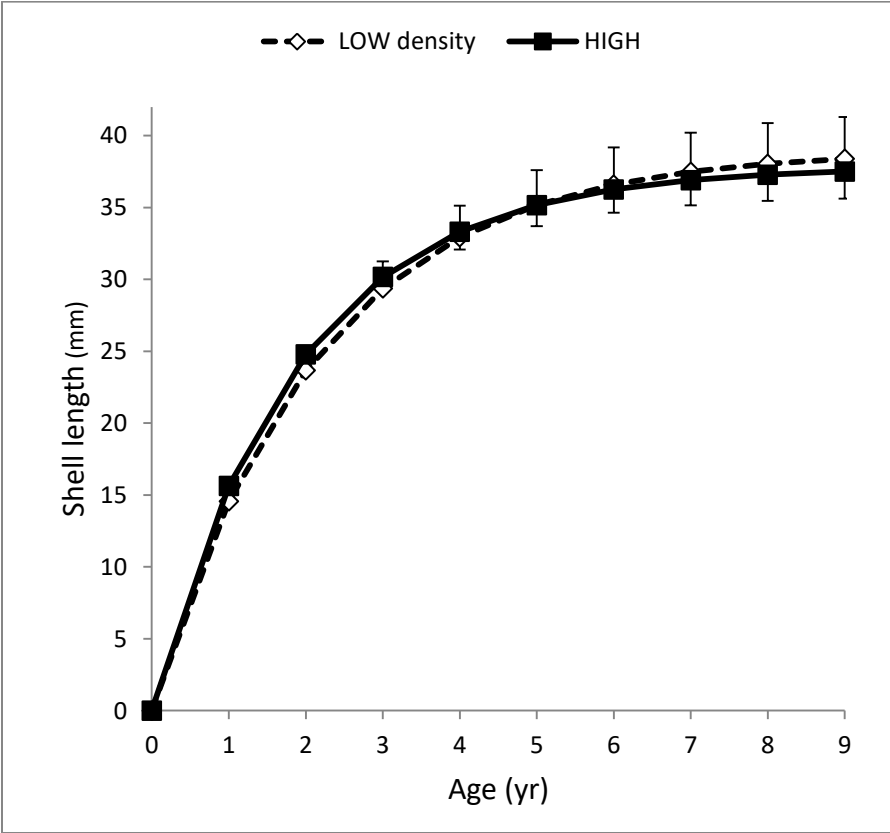


Figure 4

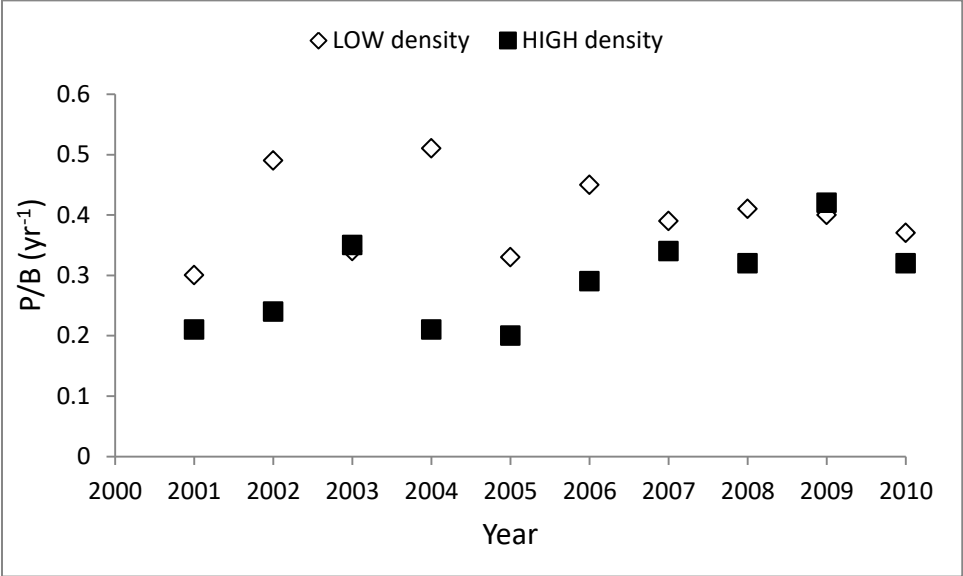


Table 1. Summary of results of Wilcoxon paired test examining the effect of *Crepidula fornicata* density (LOW and HIGH) on different parameters of their population dynamics.

	LOW	HIGH	W	P
Density N (ind.m ⁻²)	575 (403-712)	2583 (2084-2891)	2.80	0.005
Mortality rate Z (yr ⁻¹)	0.44 (0.40-0.47)	0.45 (0.38-0.49)	0.10	0.919
% 1-yr old	38.0 (30.0-47.0)	14.5 (9.0-28.0)	2.80	0.005
Asymptotic length L_{∞} (mm)	39.0 (33.8-41.7)	37.9 (36.1-41.3)	1.38	0.169
Growth coefficient K (yr ⁻¹)	0.47 (0.42-0.54)	0.54 (0.49-0.59)	2.80	0.005
Growth performance index Φ'	2.85 (2.76-2.93)	2.89 (2.84-2.92)	2.19	0.028
Biomass B (gDW.m ⁻²)	47 (39-53)	256 (243-304)	2.80	0.005
Production P (gDW.m ⁻² .yr ⁻¹)	19 (13-20)	79 (64-89)	2.80	0.005
P/B (yr ⁻¹)	0.40 (0.34-0.45)	0.31 (0.21-0.34)	2.50	0.013