



UNIVERSITY OF LEEDS

This is a repository copy of *Unveiling the Potential for Artificial Upwelling in Algae Derived Carbon Sink and Nutrient Mitigation*.

White Rose Research Online URL for this paper:

<https://eprints.whiterose.ac.uk/202057/>

Version: Preprint

Preprint:

Shen, C., Hao, X., An, D. et al. (3 more authors) (2023) Unveiling the Potential for Artificial Upwelling in Algae Derived Carbon Sink and Nutrient Mitigation. [Preprint]

<https://doi.org/10.2139/ssrn.4467862>

Reuse

Items deposited in White Rose Research Online are protected by copyright, with all rights reserved unless indicated otherwise. They may be downloaded and/or printed for private study, or other acts as permitted by national copyright laws. The publisher or other rights holders may allow further reproduction and re-use of the full text version. This is indicated by the licence information on the White Rose Research Online record for the item.

Takedown

If you consider content in White Rose Research Online to be in breach of UK law, please notify us by emailing eprints@whiterose.ac.uk including the URL of the record and the reason for the withdrawal request.



eprints@whiterose.ac.uk
<https://eprints.whiterose.ac.uk/>

Title:

Unveiling the potential for artificial upwelling in algae derived carbon sink and nutrient mitigation

Names of authors:

Chunlei Shen^a, Xinya Hao^b, Dong An^{a,c}, Martin R. Tillotson^d, Lin Yang^a, Xu Zhao^e

Affiliations and addresses:

^a School of Business, Shandong University, Weihai 264209, China

^b School of Energy and Environment, City University of Hong Kong, Hong Kong 999077, Hong Kong Special Administrative Region of China

^c School of Bohai, Hebei Agricultural University, Baoding 071000, China

^d School of Civil Engineering, University of Leeds, Leeds LS2 9JT, UK

^e Institute of Blue and Green Development, Shandong University, Weihai 264209, China

Corresponding authors:

Lin Yang (handle correspondence at all stages)

School of business

Shandong University

Weihai 264209, China

Tel: +86-199-6319-3934

E-mail: yanglin2128@126.com

Xu Zhao

Institute of Blue and Green Development

Shandong University

Weihai 264209, China

Tel: +86-136-2131-0973

Email: xuzhao@sdu.edu.cn

Unveiling the potential for artificial upwelling in algae derived carbon sink and nutrient mitigation

Chunlei Shen^a, Xinya Hao^b, Dong An^{a, c}, Martin R. Tillotson^d, Lin Yang^{a, *}, Xu Zhao^{e, *}

^a School of Business, Shandong University, Weihai 264209, China

^b School of Energy and Environment, City University of Hong Kong, Hong Kong 999077, Hong Kong Special Administrative Region of China

^c School of Bohai, Hebei Agricultural University, Baoding 071000, China

^d School of Civil Engineering, University of Leeds, Leeds LS2 9JT, UK

^e Institute of Blue and Green Development, Shandong University, Weihai 264209, China

Abstract: Mariculture algae may present a crucial part of ocean-based solutions for climate change, with the ability to sequester carbon and remove nutrients. However, the expansion of mariculture algae faces multiple challenges. Here, we measure the changes in algae derived carbon sinks and nitrogen (N) and phosphorus (P) removal between 2010 and 2020 in Shandong Province, China, and identify the key driving factors affecting change. The results show that algae derived carbon sinks and nutrient removal growth rates in Shandong Province have slowed significantly since 2014, mainly due to area limitations, laver-oriented species change, and unstable yields. Artificial upwelling (AU) has the potential to offset these adverse factors through yield enhancement. The results from scenario analysis indicate that a complete deployment of AU by 2030 will offset up to a 44.52% decrease in the mariculture algae area, or a 72.57% increase in the laver share of the algal species combination compared to 2020. Similar conclusions are reached regarding the role of AU in N and P removal. This study also identifies ancillary challenges such as low energy efficiency and high costs faced by applying AU.

Keywords: Artificial upwelling; Carbon sink; Mariculture algae; Nutrient removal; Scenario analysis

Highlights

- Artificial upwelling shows potential for algae carbon sink and nutrient removal.
- Algae carbon sink and nutrient removal are limited by area and algal species.
- Artificial upwelling offsets adverse factors by boosting yield.
- Artificial upwelling has limitations in offsetting loss.

* Corresponding authors.

E-mail addresses: yanglin0631@sdu.edu.cn (L. Yang), xuzhao@sdu.edu.cn (X. Zhao).

27 1. Introduction

28 Mariculture algae is an important component of marine ecosystems, and may deliver both economic and environmental
29 benefits. It has been recognized as having capacity to act as a carbon sink which, under the right conditions, can also
30 control eutrophication (Alvera-Azcarate et al., 2003; Bolton and Stoll, 2013; Ahmed et al., 2017; Tsai et al., 2017; Xiao
31 et al., 2017). Carbon dioxide (CO₂) and dissolved inorganic carbon (DIC) are converted into organic carbon as algae
32 grow, ultimately creating carbon sinks (Smith, 1981; Gao and McKinley, 1994). In addition, a portion of the particulate
33 organic carbon (POC) and dissolved organic carbon (DOC) released by the algae (Tyler and McGlathery, 2006; Tang et
34 al., 2011; Watanabe et al., 2020; Weigel and Pfister, 2021) may be recalcitrant dissolved organic carbon (RDOC) thus
35 forming a stable carbon sink (Jiao et al., 2010; Krause-Jensen and Duarte, 2016; Jiao et al., 2018; Gao et al., 2021).
36 Therefore, mariculture algae has been proposed as the fourth species of blue carbon (IPCC, 2019), and one of the most
37 effective CO₂ removal approaches identified thus far (NASEM, 2021). Harvesting of algae can also remove nitrogen (N)
38 and phosphorus (P) from coastal waters (Fei, 2004; He et al., 2008; Sinha et al., 2022), and has been proposed as an
39 effective ecological restoration tool to control eutrophication (Yang et al., 2015; Buschmann et al., 2017; Jiang et al.,
40 2020). Further research is required on how to fully exploit the function of mariculture algae in addressing climate change
41 and marine pollution.

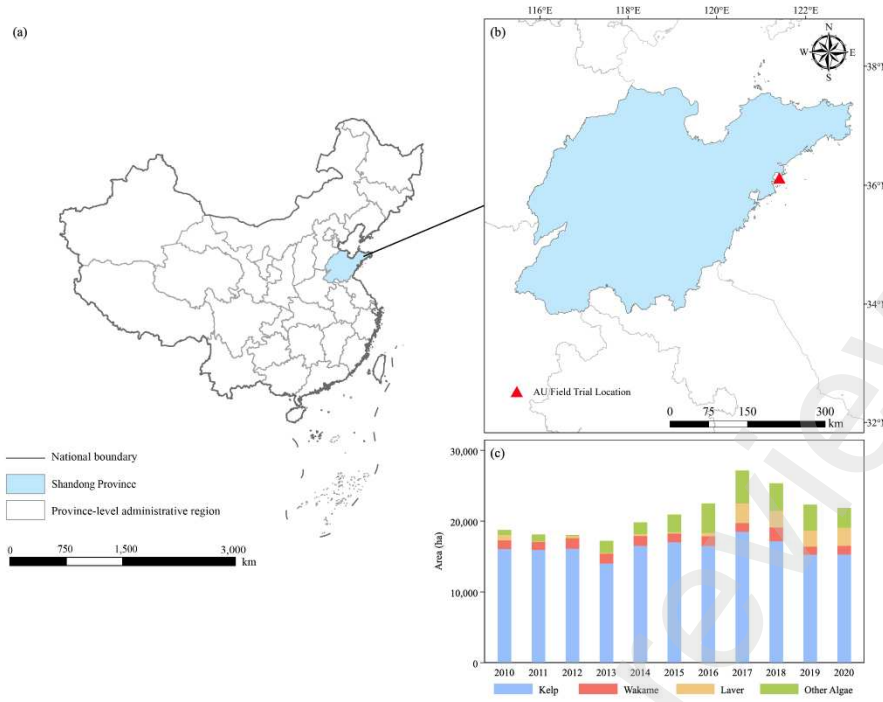
42 China leads the world in the production of mariculture algae (FAO, 2022), and has implemented a number of initiatives
43 to promote the development of algae derived carbon sinks (Jiao et al., 2018; Yang et al., 2021). Many scholars have found
44 increase in the carbon sink of algae between 2010 and 2015. Shao et al. (2019) noted significant growth of 19.63% in the
45 total mariculture derived carbon sink in China between 2010 to 2014, with growth in algae production and algal structure
46 contributing 15.16% and 4.47% to this growth, respectively. Here, algal structure means the proportion of different algal
47 species in total algal production. Yang et al. (2022) found that growth in algae production and algal structure species
48 contributed 18.36% and 1.45% respectively to carbon sink increase in China between 2011 and 2015, but the species-
49 related carbon sink coefficient (carbon sink per production unit) slightly offset the increase in carbon sinks. These studies
50 recognize the major contribution of increased algae production to carbon sinks compared to the more limited effects of
51 algal species (Ren, 2020). Similar conclusions may be drawn for nitrogen and phosphorus removal by algae (Xiao et al.,
52 2017).

53 Despite the growth of carbon sinks between 2010 and 2015, algae derived carbon sink development has slowed in
54 several of China's coastal provinces (Gu and Yin, 2022; Wu and Li, 2022). According to Yang et al. (2022), China's
55 mariculture carbon sinks grew by just 8.77% between 2016 and 2020, of which the contribution rate of algal production
56 growth dropped to 6.39%, while algae structure and carbon sink coefficients contributed only 2.17% and 0.21%,
57 respectively. This slowdown in production growth has led to a significant slowdown in the growth of carbon sinks, which
58 also affects the function of algae in nutrient removal (Wu et al., 2017). Recent ocean warming, coastal pollution,

59 competition for space, and ecological policies to control eutrophication have limited the expansion of mariculture algae
60 development (Filbee-Dexter and Wernberg, 2018; Jouffray et al., 2020; Hu et al., 2021; Wang et al., 2023). The reduction
61 in mariculture area limits production growth, which inevitably affects the amount of algae derived carbon sinks and
62 nutrient removal. However, no studies have yet been carried out to examine area as a driving factor to changes of carbon
63 sinks and nutrient removal. Therefore, we further decompose mariculture production that predominantly affects algae
64 derived carbon sinks and nutrient removal into two components i.e., yield and area. From this we hope to explore how
65 algal yield may be enhanced under limited expansion of mariculture area.

66 Artificial upwelling (AU) is a system of mechanical equipment deployed in the mariculture area, which breaks the
67 nutrient limitations of aquaculture by continuously upwelling the lower temperature, higher nutrient loaded seawater to
68 the surface (Aure et al., 2007; Lovelock and Rapley, 2007; McClimans et al., 2010; Zollmann et al., 2019; Ortiz et al.,
69 2022). AU has been shown to increase mariculture algal yield in small-scale trials (Fan et al., 2019; Lin et al., 2019; Fan
70 et al., 2020). Furthermore, researchers have found application of AU enhances downward flux of POC and carbon
71 sequestration in the deep ocean (Oschlies et al., 2010; Pan et al., 2015; Baumann et al., 2021; Gómez-Letona et al., 2022).
72 AU has been recognized by the United Nations Intergovernmental Panel on Climate Change (IPCC) as a global ocean
73 carbon sink solution (IPCC, 2019). However, the potential of AU to offset the limitations on algae derived carbon sinks
74 and nutrient removal remains unknown.

75 In this paper, we analyse the driving forces that constrain carbon sink growth and investigate the potential of AU in
76 offsetting these factors using Shandong Province, China (Fig. 1) as a case study. Shandong Province, which is bordered
77 by the Bohai Sea and the Yellow Sea, has a long coastline accounting for 1/6 of the total coastline of China (Jiao et al.,
78 2021). As China's most important mariculture location (Zhao et al., 2022), Shandong Province accounts for 27.28% (2020
79 base) of the country's mariculture algae production (SFSY, 2021). Moreover, multiple AU field experiments conducted
80 in Shandong Province analysed the specific enhancement effect of AU application, which provided the necessary
81 technical parameters for predicting the potential of AU (Fan et al., 2019; Lin et al., 2019). Our study is therefore distinct
82 from previous studies by (a) decomposing production, a factor that leads to a decline in the annual growth rates of carbon
83 sinks and nutrient removal in recent years, into yield and area; (b) estimating the potential of AU to offset the loss of
84 carbon sinks and nutrient removal caused by negative drivers; and (c) exploring the upper limits of AU potential. This
85 study can thus inform a new technology pathway for increasing carbon sinks and mitigating seawater eutrophication,
86 which can contribute to policy and management decisions to address climate change and marine pollution.



87

88 **Fig. 1.** General information of the study area. (a) The location of Shandong Province in China; (b) AU field trial location in Shandong
 89 Province (36°22' N, 120°50' E); and (c) mariculture algae area and structure in Shandong Province between 2010 and 2020.

90 2. Methodology and data

91 2.1. Measurement of carbon sink and nutrient removal of mariculture algae

92 We calculated the carbon sink of mariculture algae (TC , assuming a total of i species) by adding three components
 93 (Yang et al., 2022) i.e., the carbon sink of the algal body (C_i), the carbon sink formed by releasing POC (C_i^{POC}), and the
 94 carbon sink formed by releasing DOC (C_i^{DOC}):

$$95 TC = \sum_{i=1}^n (C_i + C_i^{POC} + C_i^{DOC}) \quad (1)$$

96 When measuring the carbon sink of mariculture algae:

$$97 C_i = DW_i \times w_i^C \quad (2)$$

$$98 C_i^{POC} = C_i \times \frac{\alpha}{1-\alpha-\beta} \times r^{POC} \quad (3)$$

$$99 C_i^{DOC} = C_i \times \frac{\beta}{1-\alpha-\beta} \times r^{DOC} \quad (4)$$

100 The carbon sink of the mariculture algal body (C_i) can be estimated from algal production (dry weight) (DW_i) and the
 101 carbon content of algae (w_i^C). α and β represent the proportion of POC and DOC released during algal growth to algal
 102 photosynthetic productivity (Yan et al., 2011). r^{POC} and r^{DOC} are the proportion of POC and DOC released by algae
 103 that are eventually converted into carbon sinks. Most of the POC and DOC released by algae do not form carbon sinks;
 104 Nilsson et al. (2018) found that 96% of the POC broke away from deposition and returned to the water column, re-entering
 105 the ocean carbon cycle. Chen et al. (2020) found that only 1.6% of algal biogenic carbon was deposited in seawater as

106 RDOC. Therefore, we assigned values of 0.04 and 0.016 to the parameters r^{POC} and r^{DOC} respectively, denoting the
 107 proportion of carbon in POC and DOC that eventually forms carbon sinks.

108 N and P removal by algae was determined by algal production (dry weight) and the N and P content of algae. The
 109 specific calculation formula is as follows:

$$110 \quad N_i = DW_i \times w_i^N \quad (5)$$

$$111 \quad P_i = DW_i \times w_i^P \quad (6)$$

112 Here, N_i and P_i represent N and P removal, w_i^N and w_i^P are the N and P content of the algae.

113 2.2. Driving force analysis using the Logarithmic Mean Divisia Index approach

114 We used the Logarithmic Mean Divisia Index (LMDI) approach (Ang et al., 2005) to decompose the changes in the
 115 carbon sink and nutrient removal of mariculture algae, and identified four driving factors i.e., intensity, yield, structure,
 116 and area, as shown in Eq. 7:

$$117 \quad M = \sum_{i=1}^n \frac{M_i}{DW_i} \times \frac{DW_i}{A_i} \times \frac{A_i}{A} \times A = \sum_{i=1}^n I_i \times Y_i \times S_i \times A \quad (7)$$

118 Here, M represents the carbon sink or nutrient removal of mariculture algae; subscript i represents algal species i , and
 119 n represents the total number of algal species (for this study $n = 4$); DW_i represents the production of algal species i ;
 120 A_i is the mariculture area used for growth of algal species i ; A refers to the total mariculture area of algae. I , Y , S , and
 121 A represent intensity, yield, structure, and area, respectively. Intensity is the amount of carbon sink or nutrient removal
 122 per unit of algal species i 's production. Yield describes the amount of production per unit of algal species i 's area.
 123 Structure is the ratio of algal species i 's area to the total area of all algae, representing the effect of algal species changes.
 124 Area reflects how the total area of mariculture algae can impact the carbon sink or nutrient removal of algal species i .

125 The total changes in the carbon sink or nutrient removal of mariculture algae can thus be formulated as:

$$126 \quad \Delta M = M^t - M^0 = \Delta I + \Delta Y + \Delta S + \Delta A \quad (8)$$

127 where ΔI (intensity effect), ΔY (yield effect), ΔS (structure effect), and ΔA (area effect) are changing driving factors
 128 of ΔM .

129 Notably, the value of I_i depends on the carbon, N, and P content of algal species i . Changes in these parameters over
 130 time were not considered in this study, therefore I_i remains unchanged. In the subsequent analysis, the contribution from
 131 the intensity effect (ΔI) to the increase in algae derived carbon sinks and nutrient removal amounts to 0.

132 According to the LMDI approach, the equations to decompose the changes to mariculture algae derived carbon sinks
 133 or nutrient removal are as follows:

$$134 \quad \Delta M_I = \sum_{i=1}^n [L(M_i^t, M_i^0) \times \ln(\frac{I_i^t}{I_i^0})] \quad (9)$$

$$135 \quad \Delta M_Y = \sum_{i=1}^n [L(M_i^t, M_i^0) \times \ln(\frac{Y_i^t}{Y_i^0})] \quad (10)$$

$$136 \quad \Delta M_S = \sum_{i=1}^n [L(M_i^t, M_i^0) \times \ln(\frac{S_i^t}{S_i^0})] \quad (11)$$

$$137 \quad \Delta M_A = \sum_{i=1}^n [L(M_i^t, M_i^0) \times \ln(\frac{A_i^t}{A_i^0})] \quad (12)$$

138 Where t and 0 represent the latter and former year during the change, respectively. L is the log-average function,
139 which satisfies:

$$140 \quad L(M_i^t, M_i^0) = \frac{M_i^t - M_i^0}{\ln(M_i^t) - \ln(M_i^0)}, M_i^t \neq M_i^0 \quad (13)$$

$$141 \quad L(M_i^t, M_i^0) = M_i^t, M_i^t = M_i^0 \quad (14)$$

142 2.3. Scenario setting

143 To estimate the potential for AU to offset the limiting effects on algae derived carbon sinks and nutrient removal by
144 2030, we set a No-AU scenario based on the development characteristics of previous mariculture area growth, as well as
145 four scenarios that consider the application of AU. The LMDI analysis was intended to reveal the driving factors that
146 slow down the carbon sink and nutrient removal growth between 2014 and 2020. Thus, we were interested in establishing
147 whether applying AU can effectively mitigate these negative factors. In the AU application scenarios, we intended to
148 calculate the minimum percentage of areas where AU application can compensate for reducing carbon sinks (nutrient
149 removal). We assumed that the yield of mariculture algae can increase by a factor of μ when applying AU. Our study
150 aimed to determine the minimum AU application proportions required to achieve a comparable scale of carbon sink
151 (nutrient removal) as in the No-AU scenario by 2030 in the four AU application scenarios, namely λ_1 (scenario S1), λ_2
152 (scenario S2), λ_3 (scenario S3), and λ_4 (scenario S4). Between 2021 and 2030, AU would be applied annually in $\lambda/10$
153 of the mariculture area. The yield of mariculture algae in Shandong Province in 2030 would be $Y_{2020}(1 + \lambda \cdot \mu)$. The
154 details of the scenarios were as follows:

155 **No-AU scenario (N1).** In the No-AU scenario, the average annual change rates of the algal area between 2021 and
156 2030 remained consistent with the average change rates of the area between 2010 and 2020. The algal structure and yield
157 remain unchanged at 2020 levels.

158 **Area constant scenario (S1).** We assumed the mariculture area of algae remained at 2020 levels. The structure of algal
159 species would be the same as for the No-AU scenario. By 2030, the algal yield would be $Y_{2020}(1 + \lambda_1 \cdot \mu)$.

160 **Area reduction scenario (S2).** There has been a noticeable decline in the mariculture area in Shandong Province since
161 2017. Hence, this scenario assumed that future changes in the mariculture area would maintain this trend. Specifically,
162 the mariculture area continued to decrease between 2021 and 2030 at an average change rate to that observed between

163 2017 and 2020. while the algal structure would remain unchanged based on 2020 levels. By 2030, the algal yield would
164 be $Y_{2020}(1 + \lambda_2 \cdot \mu)$.

165 **Laver increase scenario (S3).** The contribution of algae to carbon sinks and mitigation of seawater eutrophication
166 varies with algal species (Zheng et al., 2019). The increase in the area proportion of laver will have a negative impact on
167 the growth of carbon sinks (nutrient removal). We therefore assumed that the area proportion of laver would continue to
168 increase by 2030, at a mean growth rate to that observed between 2010 and 2020, while the mariculture area was the same
169 as in the No-AU scenario. By 2030, algal yield would therefore be $Y_{2020}(1 + \lambda_3 \cdot \mu)$.

170 **Area reduction and laver increase scenario (S4).** The area given over to mariculture algal growth would be consistent
171 with scenario S2, and the algal structure would be consistent with scenario S3. We would also calculate the minimum
172 application ratio λ_4 of AU in order to achieve a comparable scale of carbon sink (nutrient removal) as in the No-AU
173 scenario.

174 2.4. Uncertainty and sensitivity test

175 In this study, we utilized a Monte Carlo simulation to estimate the uncertainties in carbon sink and nutrient removal of
176 mariculture algae. The overall uncertainty is calculated under the 95% confidence interval around the arithmetic mean.
177 The distribution characteristics of specific model parameters are shown in Table A1. Additionally, we performed a
178 sensitivity test for the carbon sink and nutrient removal of mariculture algae to analyse the impact of different input
179 parameters on the model outputs.

180 2.5. Data collection

181 We obtained data on the production and area of mariculture algae from the “Shandong Fishery Statistical Yearbook”
182 (SFSY, 2011-2021). The specific biological parameters are shown in Table 1. The main mariculture algal species in
183 Shandong Province were kelp, laver, and wakame, which together contributed approximately 90% of total production.
184 Therefore, in the following study, the mariculture algae in Shandong Province were divided into four categories i.e., kelp,
185 laver, wakame, and others.

Table 1 Biological parameters of mariculture algae (%).			
Species	Carbon content of algae (w_i^C)	Nitrogen content of algae (w_i^N)	Phosphorus content of algae (w_i^P)
Kelp	24.99	3.71	0.52
Laver	29.09	6.30	1.00
Wakame	30.48	5.01	0.76
Other algae	28.19	5.01	0.76

Notes: The carbon content ratio of kelp, laver, and wakame refer to Zhang et al. (2020). The carbon content of other algae species were taken as the mean values of kelp, laver, and wakame. The N and P content of kelp refer to Xiao et al. (2017). The N and P content of laver refer to He et al.

(2008). Other algal species' N and P contents were taken as the mean values of kelp and laver.

186 Other parameters are shown in Table 2. The value of the average promotion rate on yield of AU (μ) was based on
187 previous field experiments. Fan et al. (2019) compared 60 strains of algae from the distribution area of the artificial
188 upwelling system and an area remote from the artificial upwelling system. They found that AU increased the average
189 weight per algae by approximately 109.9%. Lin et al. (2019) found that the average weight of algae in the experimental
190 group grown around the AU area was 33.1g, while the average weight of algae in the control group grown in the natural
191 environment was 10.1g. Based on the above findings, we took a μ of 1.1 to ensure the reliability of the prediction results.

Mechanism parameters	Values	References
r^{POC}	0.04	Nilsson et al. (2018)
r^{DOC}	0.016	Jiao et al. (2010)
α	0.19	Yan et al. (2011)
β	0.05	Yan et al. (2011)

192 3. Results

193 3.1. Carbon sink and nutrient removal of mariculture algae between 2010 and 2020

194 Between 2010 and 2020, the average annual carbon sink of mariculture algae in Shandong Province was 162.20 kt,
195 representing 23.14% of the carbon emissions of marine fisheries in 2014 (Yue et al., 2016). The carbon sink in the algal
196 body accounted for 98.91% of the total mariculture algae derived carbon sinks, while the carbon sink formed via releasing
197 POC and DOC contributed only 1.09%. The proportion of carbon sinks formed by POC and DOC measured in this study
198 was lower than in other studies due to the lower r^{POC} and r^{DOC} values utilised (Yan et al., 2011; Yang et al., 2022).

199 The carbon sink of mariculture algae in Shandong Province showed an increasing trend between 2010 and 2020 (Table
200 3), with an overall rate of 28.76%. The changes in carbon sinks may be divided into two distinct periods: from 2010 to
201 2014, the average annual growth rate of mariculture algae derived carbon sink was 5.98%. While the average annual
202 growth rate between 2014 and 2020 was only 0.34%.

203 The N and P removal trends are similar to those observed for carbon sinks. Specifically, between 2010 and 2014, there
204 was a significant increase in N and P removal, with a rise of 26.78% and 27.49%, respectively. In contrast, the nutrient
205 removal by mariculture algae was relatively stable between 2014 and 2020, with a modest increase of only 4.12% and
206 4.85%, respectively.

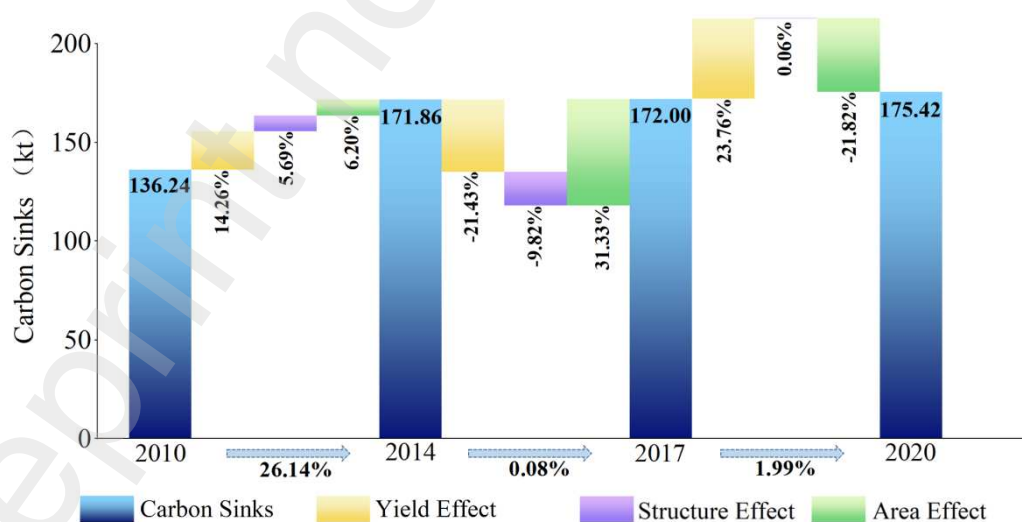
Year	2010	2011	2012	2013	2014	2015	2016	2017	2018	2019	2020
Carbon sinks of algae body	134.75	129.58	145.69	151.04	169.98	170.75	173.67	170.12	172.81	172.76	173.51

Carbon sinks through POC	1.35	1.30	1.46	1.51	1.70	1.71	1.74	1.70	1.73	1.73	1.74
Carbon sinks through DOC	0.14	0.14	0.15	0.16	0.18	0.18	0.18	0.18	0.18	0.18	0.18
Carbon sinks	136.24	131.01	147.30	152.71	171.86	172.64	175.59	172.00	174.72	174.67	175.42
Nitrogen Removal	20.50	19.76	22.28	23.09	25.99	26.12	26.79	26.31	26.94	27.18	27.06
Phosphorus Removal	2.91	2.81	3.18	3.29	3.71	3.73	3.84	3.77	3.88	3.93	3.89

207 3.2. Driving force analysis for carbon sink and nutrient removal of mariculture algae

208 We explored the driving factors (yield, structure, and area) leading to changes in carbon sink and nutrient removal of
 209 mariculture algae during the study period (Fig. 2). The analysis was divided into three periods: 2010-2014, 2014-2017,
 210 and 2017-2020. This division was based on the differences observed in the growth rates of carbon sinks and nutrient
 211 removal around 2014, as well as the clear downward trend in mariculture area used for algal growth since 2017.

212 Between 2010 and 2014 all three factors, i.e., yield, structure, and area, contributed to a rise in carbon sinks, resulting
 213 in a 26.14% increase in the carbon sink of algae relative to 2010. The yield effect stood out as the primary cause for
 214 increased carbon sinks (contributing 14.26% of the increase). Between 2014 and 2017, carbon sinks only increased by
 215 0.08% based on the 2014 level, and the effect of area became the major contributor to increased carbon sinks (53.84 kt,
 216 31.33%). In contrast, yield and structure showed inhibitory effects, resulting in a 21.43% and 9.82% reduction in carbon
 217 sinks, respectively. Between 2017 and 2020, the yield effect (40.87 kt, 23.76%) contributed positively to carbon sink
 218 growth, which was mostly offset by the negative effects of area (37.53 kt, 21.82%), resulting in only a slight increase in
 219 algae derived carbon sinks (1.99%). Meanwhile, the structure effect had little impact on carbon sinks (0.10 kt, 0.06%).
 220 The driving factors for N and P removal from mariculture algae in Shandong Province were similar to those found for
 221 carbon sinks (Fig. S1).



222 **Fig. 2.** Contribution of different driving factors to carbon sink changes in Shandong Province (2010-2020) (kt). The intensity effect
 223 (ΔI) is set to 0 and not shown in the figure.
 224

225 We found driving force effects coincided with changes to the marine environment and policy adjustments. Prior to
 226 2014, production, area, and yield of mariculture algae in Shandong Province grew rapidly, encouraged by policies such

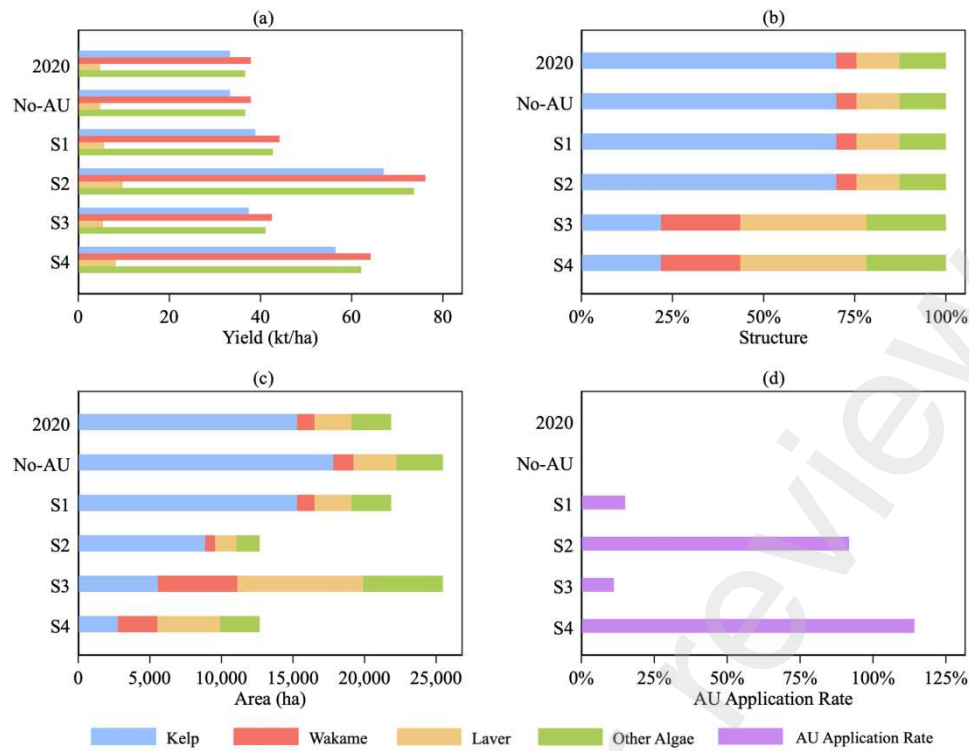
227 as increased investment in marine fishery fixed assets, subsidising of fisheries diesel, and supporting fisheries resources
228 protection (Liang et al., 2018; Han and Jiang, 2019). At the end of 2016, China released the 13th Five-Year Plan of
229 National Fishery Development, which emphasized the implementation of coastal ecological protection and promoted
230 structural reform on the supply side of fisheries (Cao et al., 2017; Su et al., 2021). As a result, many policies began to
231 restrict the expansion of mariculture areas. For example, the Blue Bay Remediation Project (BBRP) was one of the major
232 marine projects in China's 13th Five-Year Plan for ecological environmental protection, with Rizhao, Yantai, Weihai,
233 and Qingdao in Shandong Province being selected as participating cities in early 2017. The project restricted or banned
234 certain aquaculture activities in near-shore waters and targeted algal rafts for cleanup (Liu et al., 2019; Wang et al., 2020).
235 In addition, several ecological policies, such as the "returning ponds to natural wetlands", have been implemented in some
236 coastal aquaculture regions, leading to a significant decline in the mariculture algae area (Wang et al., 2023).

237 The yield effect showed a fluctuant trend between 2010 and 2020. This might be because artificial inputs and immature
238 mariculture techniques dominated algae farming, which makes algal yield susceptible to extreme natural disasters,
239 environmental conditions, water quality, and diseases (Zhang and Han, 2017).

240 The negative structural effect was primarily attributed to the increased share of laver in the mariculture area, as the
241 carbon sink and nutrient removal per unit area of laver were less than 1/5 that of kelp and wakame. The share of laver
242 increased from 1.34% to 11.80% during 2014-2020. The growing market demand for laver, a nutritious and healthy food
243 (Brown et al., 2014), is causing the area of laver to expand. Meanwhile, rising seawater temperatures due to global
244 warming have led to disease outbreaks in Jiangsu Province, China's primary laver producing area, which led to many
245 mariculture companies turning to promote the cultivation and demonstration of the laver in Shandong Province (Lu et al.,
246 2022).

247 ***3.3. Scenario analysis of the potential of AU for algae derived carbon sink and eutrophication mitigation***

248 We conducted a scenario analysis to evaluate the extent to which AU can offset the effects of two negative factors i.e.,
249 area reduction and a more laver-oriented mariculture algal system. Fig. 3 shows the required application ratio of AU and
250 the algal yield, structure, and area in 2030 to achieve the same carbon sink level as the No-AU scenario under different
251 scenarios.



252

253
254
255
256

Fig. 3. Yield, structure, area, and AU application rates in 2020, and the five scenarios in 2030. The No-AU Scenario (No-AU) represents the case in which the mariculture algal area will grow at an average growth rate between 2010 and 2020, with structure and yield remaining unchanged from 2020 levels. Scenarios 1-4 (S1-S4) represent constant area scenarios, area reduction scenarios, increased laver scenario, and area reduction and laver increase scenarios, respectively.

257

258

259

260

In the No-AU scenario, the algal area will continue to increase at an average annual growth rate between 2010 and 2020, with the structure remaining consistent with the 2020 level. When no AU technology is applied, the carbon sink of mariculture algae in Shandong Province will reach 204.41 kt by 2030, with corresponding N and P removal of 31.53 kt and 4.54 kt, respectively.

261

262

263

264

265

266

267

Applying AU may compensate for the loss of carbon sink due to diminishing mariculture area and laver-oriented structural change. In scenario S1, where the mariculture area and structure of algae remain unchanged at 2020 levels, applying AU to 15.02% of the mariculture area was sufficient to achieve the same carbon sink level as in the No-AU scenario by 2030. However, when the mariculture area decreases at the same rate as observed between 2017 and 2020 (scenario S2), AU would need to be applied to 91.81% of the area. In scenario S3, we assumed that the mariculture algal area would maintain the same growth as for the No-AU scenario, while the proportion of laver would grow to 34.53% by 2030. In this case, applying an AU to 11.14% of the mariculture algal area would be necessary.

268

269

270

271

272

273

It is worth noting there is also a limit to the potential of AU to increase carbon sinks. AU will not fully compensate for the negative effects of continuous mariculture area decline and the increase in the proportion of laver area (scenario S4). We found that when AU was implemented across the entire mariculture area by 2030, it would compensate at most for a carbon sink reduction of 44.52% in mariculture algal area compared to 2020, assuming algal structure remained constant. Similarly, supposing the mariculture area was maintained at 2020 levels with 100% application of AU, the loss of carbon sinks would not be compensated for when the share of laver exceeded 72.57%.

274 Applying AU can also compensate for the reduction in N and P removal due to mitigation in algal area and an increase
275 in the amount of laver (see Table A2). In the area reduction scenario (scenario S2), 96.23% and 95.46% of the area would
276 require AU application to secure identical N and P removal, respectively, as for the No-AU scenario by 2030. However,
277 the potential of AU would reach its limit when the area declined by more than 44.52% of the 2020 level. In the increased
278 laver scenario (scenario S3), where the laver area share increased to 34.53%, AU application rates would be 3.82% and
279 0.08% for N and P removal, respectively. If the share of laver exceeded 78.89% and 81.58%, achieving the same N and
280 P removal, respectively, as in the No-AU scenario then applying AU alone would no longer be feasible.

281 **4. Discussion**

282 *4.1. Improving key factors that influence carbon sinks and nutrient removal*

283 China has acknowledged the importance of ocean carbon sinks, particularly algae derived carbon sinks, in mitigating
284 climate change (Yang et al., 2021). The country has laid out a policy system to support the development of ocean carbon
285 sinks around the goal of carbon peak and carbon neutrality. Despite the importance of algae for increasing carbon sinks
286 and achieving carbon neutrality, the incremental carbon sinks of algae have been limited in recent years (Gu and Yin,
287 2022; Wu and Li, 2022; Yang et al., 2022). In this study, we identified the main limiting factors of algae derived carbon
288 sinks and their contributions by proposing driving factors such as yield, structure, and area. Unlike the results of previous
289 studies (Shao et al., 2019; Ren, 2020; Yang et al., 2022), we demonstrated the importance of taking area into account as
290 a driving force. The results showed that area was the most critical factor driving the growth of algae derived carbon sinks
291 until 2017. However, between 2017 and 2020, decreasing area had a significant inhibitory effect on carbon sinks. Our
292 study also revealed the negative impacts of laver expansion and unstable yields on carbon sinks. The biased culture
293 structure of laver hindered the growth of carbon sinks, and yields that fluctuate significantly over time are less conducive
294 to the stable enhancement of carbon sinks. We found similar conclusions regarding influencing factors for N and P
295 removal. The findings have contributed to adjusting mariculture industry policies regarding improved area, structure, and
296 yield to support the growth of mariculture algae derived carbon sink and nutrient removal.

297 To guarantee a steady increase in algae derived carbon sinks and nutrient removal, we propose the application of AU
298 in mariculture areas. AU provides a new impetus to the growth of algae derived carbon sinks and nutrient removal by
299 increasing yield against the negative impacts of area constraint and changes in structure changes. Our research
300 investigated the potential for AU to offset these negative effects. The results showed that enhancing carbon sink and
301 nutrient removal through AU is feasible. However, the promotion of AU also faces challenges, including its low energy
302 efficiency and high installation costs (Fan et al., 2013; Viudez et al., 2016; Qiang et al., 2018). These challenges need to
303 be considered in successful implementation of AU technology and achievement of better results in Shandong Province
304 and other coastal areas. Using clean energy to achieve self-powered AU is crucial in application of AU (Pan et al., 2018),

305 and can effectively reduce energy consumption and greenhouse gas emissions. Specifically, offshore wind, solar and tidal
306 energy can be harnessed for in-situ power generation, while wave or ocean current energy can be utilized to drive
307 upwelling and further optimize energy efficiency. Meanwhile, AU may benefit from special subsidies, tax breaks, and
308 technology research support for blue carbon. Government and market instruments can be used to provide technical and
309 financial support for AU application and promotion.

310 The yield effect was unstable between 2010 and 2020, partially due to the dominance of immature mariculture
311 techniques that make algal yield susceptible to natural disasters, environmental conditions, and disease (Zhang and Han,
312 2017). Whether AU can solve or mitigate yield fluctuation problem remains unknown. To achieve an increased and steady
313 yield, AU could combine with other farming techniques, for example: (a) use of remote sensing technology and marine
314 monitoring technology to plan cultivation sites according to required environmental conditions for the growth of different
315 algal species (Yang et al., 2022; Ai et al., 2023); (b) developing integrated multi-trophic aquaculture (IMTA) and using
316 interactions between aquatic plants and animals at different trophic levels to improve mariculture efficiency (Cutajar et
317 al., 2022; Hargrave et al., 2022); and (c) genetic improvements, such as developing adaptable and disease-resistant algal
318 cultivars (Hu et al., 2021).

319 Notably, there is an upper limit to the benefits achieved through AU. Where mariculture area declines, or the proportion
320 of laver increases, applying AU may not achieve the desired carbon sink and nutrient removal levels. Currently,
321 mariculture grown algae in China is mainly associated with nearshore waters, and some mariculture areas have been
322 reduced or removed due to global climate change, seawater pollution, and policy requirements (Liu et al., 2019; Wang et
323 al., 2020). To solve this dilemma, focusing on pollution control and ecological restoration in the original nearshore
324 mariculture areas will help improve existing farming areas. In addition, offshore mariculture may be developed by
325 cultivating new species suitable for deep-water mariculture and developing new facilities to expand mariculture space.
326 We've also noticed farmers tend to prioritize economic value of algae over environmental function when selecting species
327 for cultivation (Zheng et al., 2019). Laver is more economically valuable and preferred by farmers, while kelp and wakame
328 have a higher carbon sink and nutrient removal rates per unit of farmed area (Ou et al., 2017). By establishing marine
329 carbon sink trading platforms, farmers can be encouraged and guided to grow more species with high carbon sinks to
330 convert algae with high carbon sink functions from resources to assets. As a result, market players who protect and restore
331 the ecological environment can receive reasonable returns.

332 **4.2. Limitations**

333 As with all studies of this nature there are some limitations to our work: (a) we have simplified the complexities of
334 market demand on mariculture algal production. Total algae production may not increase even with productivity-
335 enhancing techniques because the total demand may remain relatively constant; (b) AU works better for areas where

336 surface seawater is nutrient-poor (Fan et al., 2020). The percentage increase in acreage from AU (μ) may vary depending
337 on nutrient salt levels in different waters; (c) as a geo-environmental project, the potential long-term effects of sustained
338 AU implementation on the marine and climate systems remain unknown. These include the potential for increased
339 acidification (Keller et al., 2014), harmful algal blooms (Ryan et al., 2009), and changes in ocean thermocline
340 (Kwiatkowski et al., 2015). Potential risks must be closely monitored when applying AU in a sea trial; (d) AU can increase
341 carbon sink conversion efficiency by enhancing the downward fluxes of POC (Baumann et al., 2021). We have not
342 considered this effect in our projections of AU potential due to a lack of robust and relevant parameters. The effect of AU
343 may potentially increase the carbon sink formed by both POC and DOC, providing an even more significant
344 environmental benefit.

345 It should also be cautioned that our measurements of the carbon sink of mariculture algae were based on numerical
346 models and parameters. The choice of parameters has had an influence on our results. We analysed the sensitivity of our
347 results to the parameters r^{POC} , r^{DOC} , α , and β . The detailed results of the sensitivity test are shown in Table A3. The
348 results showed that the carbon sink of mariculture algae will increase by 0.018% to 0.129% in 2020 if the mechanism
349 parameters were increased by 10%. We also estimated the uncertainties of model parameters using Monte Carlo
350 simulation methods. The uncertainty ranges of the carbon sink of mariculture algae between 2010 and 2020 are presented
351 in Fig. A.2. The uncertainty of carbon sinks (expressed as relative standard deviation (RSD) that equals the standard
352 deviation divided by the mean) ranged from 7.64% to 9.63%, indicating that the results were reliable. However, the N
353 and P removal uncertainties were relatively high, ranging from 21.31% to 26.56%, and 27.90% to 35.13%, respectively,
354 which was due to the lack of precision and relatively large standard deviation in the results of existing studies regarding
355 the measurement of algal N and P content.

356 5. Conclusions

357 This study focused on exploring the potential of AU to enhance algae derived carbon sink and mitigate eutrophication
358 in the face of continued mariculture area degradation and undesired structural change. The limited growth of the
359 mariculture algae area in Shandong Province, China, and the more intensive cultivation of laver in the limited area has
360 resulted in minimal improvements in carbon sinks and nutrient removal levels since 2014. Our findings indicated that
361 applying AU could effectively compensate for the loss of carbon sink and nutrient removal caused by the decrease of
362 mariculture area or the increase of the laver share. Meanwhile, we observed that the potential for AU to achieve these
363 benefits has upper limits. It is worth mentioning that scenario analysis cannot calculate future carbon sinks and nutrient
364 removal accurately, but rather reflects a promising technical pathway for improving algae derived carbon sinks and
365 nutrient removal in the face of shrinking mariculture areas and suboptimal species selection. Further research could

366 investigate the implication of other potential variables, such as the intensity effect changes over time and AU energy
367 efficiency on the carbon sink and nutrient removal potential.

368 **CRedit authorship contribution statement**

369 **Chunlei Shen:** Conceptualization, Writing – original draft, Investigation, Formal analysis. **Xinya Hao:**
370 Conceptualization, Methodology, Software, Writing – review & editing. **Dong An:** Investigation, Data curation. **Martin**
371 **R. Tillotson:** Writing – original draft, Writing – review & editing. **Lin Yang:** Conceptualization, Supervision,
372 Investigation, Funding acquisition. **Xu Zhao:** Conceptualization, Writing – review & editing, Methodology, Funding
373 acquisition.

374 **Declaration of competing interest**

375 The authors declare that they have no known competing financial interests or personal relationships that could have
376 appeared to influence the work reported in this paper.

377 **Acknowledgements**

378 This work was supported by the Major Project of National Social Science Foundation of China (No. 20&ZD100), the
379 National Natural Science Foundation of China (No. 72074136).

380 **References**

- 381 Ahmed, N., Bunting, S.W., Glaser, M., Flaherty, M.S., Diana, J.S., 2017. Can greening of aquaculture sequester blue carbon? *Ambio*
382 46, 468-477. <https://doi.org/10.1007/s13280-016-0849-7>.
- 383 Ai, B., Wang, P.P., Yang, Z.Y., Tian, Y.X., Liu, D.D., 2023. Spatiotemporal dynamics analysis of aquaculture zones and its impact on
384 green tide disaster in Haizhou Bay, China. *Mar. Environ. Res.* 183, 105825. <https://doi.org/10.1016/j.marenvres.2022.105825>.
- 385 Alvera-Azcárate, A., Ferreira, J.G., Nunes, J.P., 2003. Modelling eutrophication in mesotidal and macrotidal estuaries. The role of
386 intertidal seaweeds. *Estuar. Coast. Shelf Sci.* 57, 715-724. [https://doi.org/10.1016/S0272-7714\(02\)00413-4](https://doi.org/10.1016/S0272-7714(02)00413-4).
- 387 Ang, B.W., 2005. The LMDI approach to decomposition analysis: a practical guide. *Energy Pol.* 33, 867e871.
388 <https://doi.org/10.1016/j.enpol.2003.10.010>.
- 389 Aure, J., Strand, O., Erga, S.R., Strohmeier, T., 2007. Primary production enhancement by artificial upwelling in a western Norwegian
390 fjord. *Mar. Ecol. Prog. Ser.* 352, 39-52. <https://doi.org/10.3354/meps07139>.
- 391 Baumann, M., Taucher, J., Paul, A.J., Heinemann, M., Vanharanta, M., Bach, L.T., Spilling, K., Ortiz, J., Arístegui, J., Hernández-
392 Hernández, N., Baños, I., Riebesell, U., 2021. Effect of intensity and mode of artificial upwelling on particle flux and carbon export.
393 *Front. Mar. Sci.* 8, <https://doi.org/10.3389/fmars.2021.742142>.
- 394 Bolton, C.T., Stoll, H.M., 2013. Late Miocene threshold response of marine algae to carbon dioxide limitation. *Nature* 500, 558-562.
395 <https://doi.org/10.1038/nature12448>.
- 396 Brown, E.M., Allsopp, P.J., Magee, P.J., Gill, C.I.R., Nitecki, S., Strain, C.R., McSorley, E.M., 2014. Seaweed and human health.
397 *Nutr. Rev.* 72, 205-216. <https://doi.org/10.1111/nure.12091>.
- 398 Buschmann, A.H., Camus, C., Infante, J., Neori, A., Israel, Á., Hernández-González, M.C., Pereda, S.V., Luis Gomez-Pinchetti, J.,
399 Golberg, A., Tadmor-Shalev, N., Critchley, A.T., 2017. Seaweed production: overview of the global state of exploitation, farming
400 and emerging research activity. *Eur. J. Phycol.* 52, 391-406. <https://doi.org/10.1080/09670262.2017.1365175>.

401 Cao, L., Chen, Y., Dong, S., Hanson, A., Huang, B., Leadbitter, D., Little, D.C., Pikitch, E.K., Qiu, Y., de Mitcheson, Y.S., Sumaila,
402 U.R., Williams, M., Xue, G., Ye, Y., Zhang, W., Zhou, Y., Zhuang, P., Naylor, R.L., 2017. Opportunity for marine fisheries
403 reform in China. *Proc. Natl. Acad. Sci. U. S. A.* 114, 435-442. <https://doi.org/10.1073/pnas.1616583114>.

404 Chen, J., Li, H.M., Zhang, Z.H., He, C., Shi, Q., Jiao, N.Z., Zhang, Y.Y., 2020. DOC dynamics and bacterial community succession
405 during long-term degradation of *Ulva prolifera* and their implications for the legacy effect of green tides on refractory DOC pool in
406 seawater. *Water Res.* 185, 116268. <https://doi.org/10.1016/j.watres.2020.116268>.

407 Cutajar, K., Falconer, L., Massa-Gallucci, A., Cox, R.E., Schenke, L., Bardócz, T., Sharman, A., Deguara, S., Telfer, T.C., 2022.
408 Culturing the sea cucumber *Holothuria poli* in open-water integrated multi-trophic aquaculture at a coastal Mediterranean fish farm.
409 *Aquaculture* 550, 737881. <https://doi.org/10.1016/j.aquaculture.2021.737881>.

410 Fan, W., Chen, J.W., Pan, Y.W., Huang, H.C., Chen, C.-T.A., Chen, Y., 2013. Experimental study on the performance of an air-lift
411 pump for artificial upwelling. *Ocean Eng.* 59, 47-57. <https://doi.org/10.1016/j.oceaneng.2012.11.014>.

412 Fan, W., Zhang, Z.J., Yao, Z.Z., Xiao, C.B., Zhang, Y., Zhang, Y.Y., Liu, J.H., Di, Y.N., Chen, Y., Pan, Y.W., 2020. A sea trial of
413 enhancing carbon removal from Chinese coastal waters by stimulating seaweed cultivation through artificial upwelling. *Appl. Ocean*
414 *Res.* 101, 102260. <https://doi.org/10.1016/j.apor.2020.102260>.

415 Fan, W., Zhao, R.L., Z., Y.Z., B., X.C., Pan, Y.W., Chen, Y., Jiao, N.Z., Zhang, Y., 2019. Nutrient removal from Chinese coastal
416 waters by large-scale seaweed aquaculture using artificial upwelling. *Water* 11, 1754. <https://doi.org/10.3390/w11091754>.

417 FAO. 2022. The State of World fisheries and aquaculture 2022. Towards Blue Transformation. Rome.
418 <https://doi.org/10.4060/cc0461en>.

419 Fei, X.G., 2004. Solving the coastal eutrophication problem by large scale seaweed cultivation. *Hydrobiologia* 512, 145-151.
420 <https://doi.org/10.1023/B:HYDR.0000020320.68331.ce>.

421 Filbee-Dexter, K., Wernberg, T., 2018. Rise of turfs: a new battlefield for globally declining kelp forests. *Bioscience* 68, 64-76.
422 <https://doi.org/10.1093/biosci/bix147>.

423 Gao, K., McKinley, K.R., 1994. Use of macroalgae for marine biomass production and CO₂ remediation: a review. *J. Appl. Phycol.* 6,
424 45-60. <https://doi.org/10.1007/BF02185904>.

425 Gao, Y.P., Zhang, Y.T., Du, M.R., Lin, F., Jiang, W.W., Li, W.H., Li, F.X., Lv, X.N., Fang, J.H., Jiang, Z.J., 2021. Dissolved organic
426 carbon from cultured kelp *Saccharina japonica*: production, bioavailability, and bacterial degradation rates. *Aquac. Environ.*
427 *Interact.* 13, 101-110. <https://doi.org/10.3354/aei00393>.

428 Gómez-Letona, M., Sebastián, M., Baños, I., Fernanda Montero, M., Pérez Barrancos, C., Baumann, M., Riebesell, U., Arístegui, J.,
429 2022. The importance of the dissolved organic matter pool for the carbon sequestration potential of artificial upwelling. *Front. Mar.*
430 *Sci.* 9, 969714. <https://doi.org/10.3389/fmars.2022.969714>.

431 Gu, H.L., Yin, K.D., 2022. Forecasting algae and shellfish carbon sink capability on fractional order accumulation grey model. *Math.*
432 *Biosci. Eng.* 19, 5409-5427. <https://doi.org/10.3934/mbe.2022254>.

433 Han, H., Jiang, Y., 2019. The evolution of mariculture structures and environmental effects in China. *J. Coastal Res.* 83, 155-166.
434 <https://doi.org/10.2112/si83-024.1>.

435 Hargrave, M.S.S., Nylund, G.M.M., Enge, S., Pavia, H., 2022. Co-cultivation with blue mussels increases yield and biomass quality
436 of kelp. *Aquaculture* 550, 737832. <https://doi.org/10.1016/j.aquaculture.2021.737832>.

437 He, P.M., Xu, S.N., Zhang, H.Y., Wen, S.S., Dai, Y.J., Lin, S.J., Yarish, C., 2008. Bioremediation efficiency in the removal of dissolved
438 inorganic nutrients by the red seaweed, *Porphyra yezoensis*, cultivated in the open sea. *Water Res.* 42, 1281-1289.
439 <https://doi.org/10.1016/j.watres.2007.09.023>.

440 Hu, Z.M., Shan, T.F., Zhang, J., Zhang, Q.S., Critchley, A.T., Choi, H.G., Yotsukura, N., Liu, F.L., Duan, D.L., 2021. Kelp aquaculture
441 in China: a retrospective and future prospects. *Rev. Aquacult.* 13, 1324-1351. <https://doi.org/10.1111/raq.12524>.

442 IPCC (Intergovernmental Panel on Climate Change), 2019. Special report on the ocean and cryosphere in a changing climate.
443 Technical Report. <https://www.ipcc.ch/srocc/>.

444 Jiang, Z.B., Liu, J.J., Li, S.L., Chen, Y.Y., Du, P., Zhu, Y.L., Liao, Y.B., Chen, Q.Z., Shou, L., Yan, X.J., Zeng, J.N., Chen, J.F., 2020.
445 Kelp cultivation effectively improves water quality and regulates phytoplankton community in a turbid, highly eutrophic bay. *Sci.*
446 *Total Environ.* 707, 135561. <https://doi.org/10.1016/j.scitotenv.2019.135561>.

447 Jiao, N.Z., Herndl, G.J., Hansell, D.A., Benner, R., Kattner, G., Wilhelm, S.W., Kirchman, D.L., Weinbauer, M.G., Luo, T.W., Chen,
448 F., Azam, F., 2010. Microbial production of recalcitrant dissolved organic matter: long-term carbon storage in the global ocean. *Nat.*
449 *Rev. Microbiol.* 8, 593-599. <https://doi.org/10.1038/nrmicro2386>.

450 Jiao, N.Z., Wang, H., Xu, G., Aricò, S., 2018. Blue carbon on the rise: challenges and opportunities. *Natl. Sci. Rev.* 5, 464-468.
451 <https://doi.org/10.1093/nsr/nwy030>.

452 Jiao, Y.N., Yang, L.P., Kong, Z.Q., Shao, L.J., Wang, G.L., Ren, X.F., Liu, Y.J., 2021. Evaluation of trace metals and rare earth
453 elements in mantis shrimp *Oratosquilla oratoria* collected from Shandong Province, China, and its potential risks to human health.
454 *Mar. Pollut. Bull.* 162, 111815. <https://doi.org/10.1016/j.marpolbul.2020.111815>.

455 Jouffray, J.B., Blasiak, R., Norstrom, A.V., Österblom, H., Nyström, M., 2020. The blue acceleration: the trajectory of human
456 expansion into the ocean. *One Earth* 2, 43-54. <https://doi.org/10.1016/j.oneear.2019.12.016>.

457 Keller, D.P., Feng, E.Y., Oschlies, A., 2014. Potential climate engineering effectiveness and side effects during a high carbon dioxide-
458 emission scenario. *Nat. Commun.* 5, 3303. <https://doi.org/10.1038/ncomms4304>.

459 Krause-Jensen, D., Duarte, C.M., 2016. Substantial role of macroalgae in marine carbon sequestration. *Nat. Geosci.* 9, 737-742.
460 <https://doi.org/10.1038/ngeo2790>.

461 Kwiatkowski, L., Ricke, K.L., Caldeira, K., 2015. Atmospheric consequences of disruption of the ocean thermocline. *Environ. Res.*
462 *Lett.* 10, 034016. <https://doi.org/10.1088/1748-9326/10/3/034016>.

463 Liang, Y.X., Cheng, X.W., Zhu, H., Shutes, B., Yan, B.X., Zhou, Q.W., Yu, X.F., 2018. Historical evolution of mariculture in China
464 during past 40 years and its impacts on eco-environment. *Chin. Geogra. Sci.* 28, 363-373. <https://doi.org/10.1007/s11769-018-0940-z>.

465 Lin, T.C., Fan, W., Xiao, C.B., Yao, Z.Z., Zhang, Z.J., Zhao, R.L., Pan, Y.W., Chen, Y., 2019. Energy management and operational
466 planning of an ecological engineering for carbon sequestration in coastal mariculture environments in China. *Sustainability* 11, 3162.
467 <https://doi.org/10.3390/su11113162>.

468 Liu, F.L., Liang, Z.R., Zhang, P.Y., Wang, W.J., Sun, X.T., Wang, F.J., Yuan, Y.M., 2019. Preliminary discussion on the development
469 of *Saccharina japonica* offshore aquaculture in China. *Prog. Fish. Sci.* 40 (in Chinese), 161-166.

470 Lovelock, J.E., Rapley, C.G., 2007. Ocean pipes could help the Earth to cure itself. *Nature* 449, 403. <https://doi.org/10.1038/449403a>.

471 Lu, F., Zhan, D.M., Ding, G., Liu, W., Tang, L.Q., Wu, H.Y., 2022. Effects of Nitrogen and Phosphorus Enrichment on Growth and
472 Nutritional Components of *Pyropia haitanensis* in Changdao, Shandong Province. *Guangxi Sci.* 29 (in Chinese), 168-175.

473 McClimans, T.A., Handá, A., Fredheim, A., Lien, E., Reitan, K.I., 2010. Controlled artificial upwelling in a fjord to stimulate non-
474 toxic algae. *Aquacult. Eng.* 42, 140-147. <https://doi.org/10.1016/j.aquaeng.2010.02.002>.

475 NASEM (National Academies of Sciences, Engineering, and Medicine), 2022. A research strategy for ocean-based carbon dioxide
476 removal and sequestration. Washington, DC: The National Academies Press. <https://doi.org/10.17226/26278>.

477 Nilsson, M.M., Kononets, M., Ekeröth, N., Viktorsson, L., Hylén, A., Sommer, S., Pfannkuche, O., Almroth-Rosell, E., Atamanchuk,
478 D., Andersson, J.H., Roos, P., Tengberg, A., Hall, P.O.J., 2018. Organic carbon recycling in Baltic Sea sediments - An integrated
479 estimate on the system scale based on in situ measurements. *Mar. Chem.* 209, 81-93. <https://doi.org/10.1016/j.marchem.2018.11.004>.

480 Ortiz, J., Aristegui, J., Hernandez-Hernandez, N., Fernandez-Mendez, M., Riebesell, U., 2022. Oligotrophic phytoplankton community
481 effectively adjusts to artificial upwelling regardless of intensity, but differently among upwelling modes. *Front. Mar. Sci.* 9,
482 <https://doi.org/10.3389/fmars.2022.880550>.

483 Oschlies, A., Pahlow, M., Yool, A., Matear, R.J., 2010. Climate engineering by artificial ocean upwelling: Channeling the sorcerer's
484 apprentice. *Geophys. Res. Lett.* 37, L04701. <https://doi.org/10.1029/2009gl041961>.

485 Ou, G., Wang, X., Yang, A., Ke, A., Guan, W., 2017. Interspecific differences in the carbon sink capacity of macroalgae. *Journal of*
486 *Zhejiang Agricultural Sciences* 58 (in Chinese), 1436-1439+1443. <https://doi.org/10.16178/j.issn.0528-9017.20170843>.

487 Pan, Y.W., Fan, W., Huang, T.-H., Wang, S.-L., Chen, C.-T.A., 2015. Evaluation of the sinks and sources of atmospheric CO₂ by
488 artificial upwelling. *Sci. Total Environ.* 511, 692-702. <https://doi.org/10.1016/j.scitotenv.2014.11.060>.

489 Pan, Y.W., You, L., Li, Y., Fan, W., Chen, C.-T.A., Wang, B.-J., Chen, Y., 2018. Achieving highly efficient atmospheric CO₂ uptake
490 by artificial upwelling. *Sustainability* 10, 664. <https://doi.org/10.3390/su10030664>.

491 Qiang, Y.F., Fan, W., Xiao, C.B., Pan, Y.W., Chen, Y., 2018. Effects of operating parameters and injection method on the performance
492 of an artificial upwelling by using airlift pump. *Appl. Ocean Res.* 78, 212-222. <https://doi.org/10.1016/j.apor.2018.06.006>.

493 Ren, W.H., 2021. Study on the removable carbon sink estimation and decomposition of influencing factors of mariculture shellfish and
 494 algae in China - a two-dimensional perspective based on scale and structure. *Environ. Sci. Pollut. Res.* 28, 21528-21539.
 495 <https://doi.org/10.1007/s11356-020-11997-1>.

496 Ryan, J.P., Fischer, A.M., Kudela, R.M., Gower, J.F.R., King, S.A., Marin, R., III, Chavez, F.P., 2009. Influences of upwelling and
 497 downwelling winds on red tide bloom dynamics in Monterey Bay, California. *Cont. Shelf Res.* 29, 785-795.
 498 <https://doi.org/10.1016/j.csr.2008.11.006>.

499 SFSY (Shandong Fishery Statistical Yearbook), 2011-2021. Shandong Provincial Department of Ocean and Fisheries, Jinan 2011-
 500 2021.

501 Shao, G.L., Liu, B., Li, C., 2019. Evaluation of carbon dioxide capacity and the effects of decomposition and spatio-temporal
 502 differentiation of seawater in China's main sea area based on panel data from 9 coastal provinces in China. *Acta Ecol. Sin.* 39 (in
 503 Chinese with English abstract), 2614-2625.

504 Sinha, R., Thomas, J.B.E., Strand, A., Soderqvist, T., Stadmark, J., Franzen, F., Ingmansson, I., Grondahl, F., Hasselstrom, L., 2022.
 505 Quantifying nutrient recovery by element flow analysis: Harvest and use of seven marine biomasses to close N and P loops. *Resour.*
 506 *Conserv. Recycl.* 178, 1106031. <https://doi.org/10.1016/j.resconrec.2021.106031>.

507 Smith, S.V., 1981. Marine macrophytes as a global carbon sink. *Science* 211, 838-840. <https://doi.org/10.1126/science.211.4484.838>.

508 Su, M., Wang, L.L., Xiang, J.H., Ma, Y.X., 2021. Adjustment trend of China's marine fishery policy since 2011. *Mar. Pol.* 124, 104322.
 509 <https://doi.org/https://doi.org/10.1016/j.marpol.2020.104322>.

510 Tang, Q.S., Zhang, J.H., Fang, J.G., 2011. Shellfish and seaweed mariculture increase atmospheric CO₂ absorption by coastal
 511 ecosystems. *Mar. Ecol. Prog. Ser.* 424, 97-104. <https://doi.org/10.3354/meps08979>.

512 Tsai, D.D.-W., Chen, P.H., Ramaraj, R., 2017. The potential of carbon dioxide capture and sequestration with algae. *Ecol. Eng.* 98, 17-
 513 23. <https://doi.org/10.1016/j.ecoleng.2016.10.049>.

514 Tyler, A.C., McGlathery, K.J., 2006. Uptake and release of nitrogen by the macroalgae *Gracilaria vermiculophylla* (Rhodophyta). *J.*
 515 *Phycol.* 42, 515-525. <https://doi.org/10.1111/j.1529-8817.2006.00224.x>.

516 Viudez, A., Balsells, M.F.P., Rodriguez-Marroyo, R., 2016. Artificial upwelling using offshore wind energy for mariculture
 517 applications. *Sci. Mar.* 80, 235-248. <https://doi.org/10.3989/scimar.04297.06B>.

518 Wang, G.G., Shuai, L., Li, Y., Lin, W., Zhao, X.W., Duan, D.L., 2008. Phylogenetic analysis of epiphytic marine bacteria on Hole-
 519 Rotten diseased sporophytes of *Laminaria japonica*. *J. Appl. Phycol.* 20, 403-409. <https://doi.org/10.1007/s10811-007-9274-4>.

520 Wang, M., Mao, D., Xiao, X., Song, K., Jia, M., Ren, C., Wang, Z., 2023. Interannual changes of coastal aquaculture ponds in China
 521 at 10-m spatial resolution during 2016–2021. *Remote Sens. Environ.* 284, 113347.
 522 <https://doi.org/https://doi.org/10.1016/j.rse.2022.113347>.

523 Wang, M.Q., Wang, X.H., Zhou, R., Zhang, Z.P., 2020. An indicator framework to evaluate the Blue Bay Remediation Project in
 524 China. *Reg. Stud. Mar. Sci.* 38, 101349. <https://doi.org/10.1016/j.rsma.2020.101349>.

525 Watanabe, K., Yoshida, G., Hori, M., Umezawa, Y., Moki, H., Kuwae, T., 2020. Macroalgal metabolism and lateral carbon flows can
 526 create significant carbon sinks. *Biogeosciences* 17, 2425-2440. <https://doi.org/10.5194/bg-17-2425-2020>.

527 Weigel, B.L., Pfister, C.A., 2021. The dynamics and stoichiometry of dissolved organic carbon release by kelp. *Ecology* 102, e03221.
 528 <https://doi.org/https://doi.org/10.1002/ecy.3221>.

529 Wu, H.L., Kim, J.K., Huo, Y.Z., Zhang, J.H., He, P.M., 2017. Nutrient removal ability of seaweeds on *Pyropia yezoensis* aquaculture
 530 rafts in China's radial sandbanks. *Aquat. Bot.* 137, 72-79. <https://doi.org/10.1016/j.aquabot.2016.11.011>.

531 Wu, J.H., Li, B., 2022. Spatio-temporal evolutionary characteristics of carbon emissions and carbon sinks of marine industry in China
 532 and their time-dependent models. *Mar. Policy* 135, 104879. <https://doi.org/10.1016/j.marpol.2021.104879>.

533 Xiao, X., Agusti, S., Lin, F., Li, K., Pan, Y.R., Yu, Y., Zheng, Y.H., Wu, J.P., Duarte, C.M., 2017. Nutrient removal from Chinese
 534 coastal waters by large-scale seaweed aquaculture. *Sci. Rep.* 7, 46613. <https://doi.org/10.1038/srep46613>.

535 Yan, L.W., Hang, H.J., Chen, J.T., Yan Yang, X.G., 2011. Estimation of carbon sink capacity of algal mariculture in the coastal areas
 536 of China. *Adv. Mar. Sci.* 29 (in Chinese with English abstract), 537-545.

537 Yang, Y., Chai, Z., Wang, Q., Chen, W., He, Z., Jiang, S., 2015. Cultivation of seaweed *Gracilaria* in Chinese coastal waters and its
 538 contribution to environmental improvements. *Algal Res.* 9, 236-244. <https://doi.org/10.1016/j.algal.2015.03.017>.

- 539 Yang, Y.F., Luo, H.T., Wang, Q., He, Z.L., Long, A.M., 2021. Large-scale cultivation of seaweed is effective approach to increase
540 marine carbon sequestration and solve coastal environmental problems. *Bull Chin. Acad. Sci.* 36 (in Chinese with English abstract),
541 259-269. <https://doi.org/10.16418/j.issn.1000-3045.20210217103>.
- 542 Yang, L., Hao, X.Y., Shen, C.L., An, D., 2022. Assessment of carbon sink capacity and potential of marine fisheries in China under
543 the carbon neutrality target. *Resour. Sci.* 44 (in Chinese with English abstract), 716-729. <https://doi.org/10.18402/resci.2022.04.06>.
- 544 Yang, Z.Y., Yu, X.Y., Dedman, S., Rosso, M., Zhu, J.M., Yang, J.Q., Xia, Y.X., Tian, Y.C., Zhang, G.P., Wang, J.Z., 2022. UAV
545 remote sensing applications in marine monitoring: Knowledge visualization and review. *Sci. Total Environ.* 838, 155939.
546 <https://doi.org/10.1016/j.scitotenv.2022.155939>.
- 547 Yue, D.D., Wang, L.M., Fang, H., Geng, R., Zhao, P.F., Xiong, M.S., Wang, Q., Zhou, Y.S., Xiao, L., 2016. Development Strategy of
548 Marine Fisheries in China Based on the Carbon Balance. *J. Agric. Sci. Technol.* 18 (in Chinese with English abstract), 1-8.
549 <https://doi.org/10.13304/j.nykjdb.2015.695>.
- 550 Zhang, J.F., Cai, H.J., Zhao, Y.R., Chen, W.H., Hu, S.Q., Liu, Y., Liu, C.F., 2020. Seasonal variation in the total organic carbon
551 contents and the $\delta^{13}\text{C}$ values of macroalgae in the rocky intertidal zone of the Zhangzi island. *Mar. Sci.* 44 (in Chinese), 56-65.
- 552 Zhang, L.T., Han, L.M., 2017. The problems and policy recommendations on the development of Chinese seaweed industry. *Chin.*
553 *Fish. Econ.* 35 (in Chinese), 89-95.
- 554 Zhang, W.J., Dong, Z.J., Zhang, C., Sun, X.Y., Hou, C.W., Liu, Y.L., Wang, L., Ma, Y.Q., Zhao, J.M., 2020. Effects of physical-
555 biochemical coupling processes on the *Noctiluca scintillans* and *Mesodinium* red tides in October 2019 in the Yantai nearshore,
556 China. *Mar. Pollut. Bull.* 160, <https://doi.org/10.1016/j.marpolbul.2020.111609>.
- 557 Zhao, Y.Z., Li, Y.F., Wang, X.W., 2022. The land-sea system dynamics model with shared socioeconomic pathways can identify the
558 gaps in achieving Sustainable Development Goal 14. *Resour. Conserv. Recycl.* 181, 106257.
559 <https://doi.org/10.1016/j.resconrec.2022.106257>.
- 560 Zheng, Y.H., Jin, R.J., Zhang, X.J., Wang, Q.X., Wu, J.P., 2019. The considerable environmental benefits of seaweed aquaculture in
561 China. *Stochastic Environ. Res. Risk Assess.* 33, 1203-1221. <https://doi.org/10.1007/s00477-019-01685-z>.
- 562 Zollmann, M., Traugott, H., Chemodanov, A., Liberzon, A., Golberg, A., 2019. Deep water nutrient supply for an offshore *Ulva sp.*
563 cultivation project in the Eastern Mediterranean Sea: Experimental simulation and modeling. *Bioenergy Res.* 12, 1113-1126.
564 <https://doi.org/10.1007/s12155-019-10036-3>.