




Catchment and climatic influences on spatio-temporal variations in suspended sediment transport dynamics in rivers

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ABSTRACT

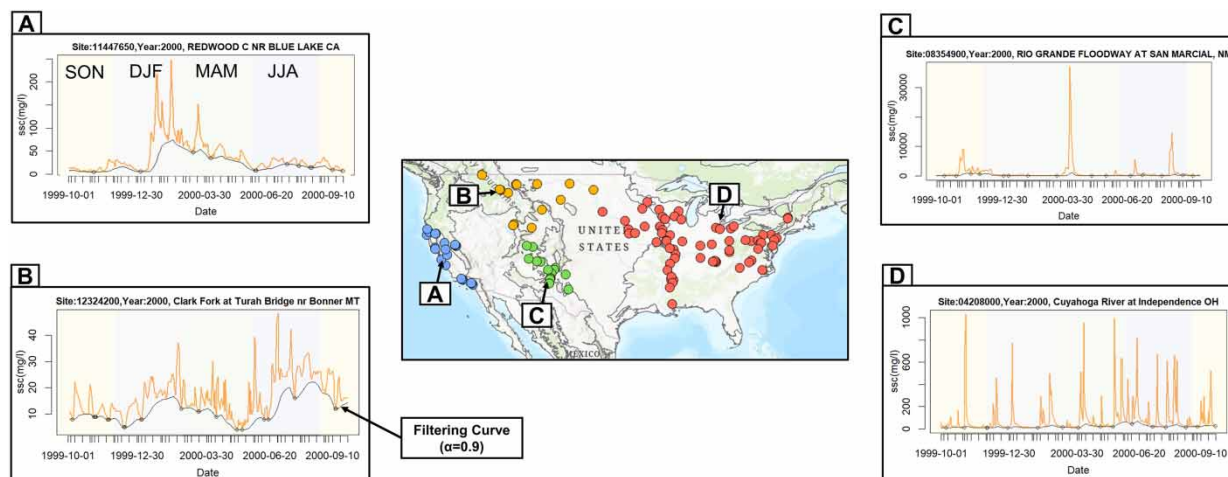
Suspended sediment concentration (SSC) is an important attribute for water resources management. However, the interactions between climate and catchment characteristics that control the temporal variability of SSC in rivers are not fully resolved. The study aim is to evaluate how these variables influence spatial and seasonal variations in SSC dynamics at a continental scale. Daily SSC (mg/l) and site attribute data from 120 sites (USA) with minimum 10 years of record (1971–2000) were analysed. New indicators of SSC dynamics (magnitude and frequency) were developed and applied annually and seasonally. Geographically weighted regression (GWR) models were created for each ordinary least squares (OLS) regression model, and GWR coefficients were analysed by ecoregion. Land cover, rainfall and erosivity, baseflow index and soil texture were the most common variables in the OLS models. GWR coefficients displayed significant variation across the continent. Agricultural cover was positively associated with low frequency SSC events, while urban and forest cover predicted higher frequency events, except in the desert areas. PPT30 was generally a negative predictor for SSC magnitude, except the marine west coasts forests. These findings on catchment and climate controls on SSC will support future predictive models of SS transport dynamics.

Key words: catchment, climate, land cover, land use, sediment yield, suspended sediment concentrations

HIGHLIGHTS

- Influence of catchment and climate variables on short-term (day-month) suspended sediment transport dynamics were investigated using new indicators.
- Regression analysis identified land cover/land use, precipitation, baseflow index and soil clay content as significant predictor variables.
- Spatial modelling identified significant regional and seasonal differences in the influence of catchment and climate variables.

GRAPHICAL ABSTRACT



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INTRODUCTION

Climate change is predicted to affect the flux of suspended sediment in river systems. Changes in rainfall, temperature and CO₂ concentrations are expected to alter the forces (i.e. rainfall intensity, river discharge and wind) that drive erosion and transport but also the factors that moderate these forces, such as vegetation growth (Naylor *et al.* 2017; Lowe *et al.* 2018; IPCC 2019). These moderating factors operate naturally and via anthropogenic activity, including changes to land use and land cover (LULC) in response to or as part of mitigating actions against climate change, e.g. shifts in crop types or conversion of agricultural land to forest (Murphy 2020). Considerable research has been conducted on the impact of LULC and climate change on annual sediment load (i.e. sediment yield) at catchment scales (Dai *et al.* 2009; Hung *et al.* 2014; Camilo & Restrepo 2017; Kemper *et al.* 2019) and at continental scales (Syvitski 2003; Walling & Fang 2003; Wang *et al.* 2011; Syvitski *et al.* 2014; Moragoda & Cohen 2020). However, there has been less focus on how these factors and their interactions affect suspended sediment transport dynamics, i.e. short-term changes in suspended sediment concentration (SSC) at a location over days to months (Mount & Abrahart 2011; Roman *et al.* 2012; Murphy 2020).

Uncertain future trajectories in SSC dynamics exist because numerous factors affect SS transport. Climate, LULC and human activities all affect the magnitude, frequency and timing of SS events (Vercruyse *et al.* 2017). Climate change is shifting temperatures and rainfall, but is also altering the frequency and magnitude of extreme weather events (van de Pol *et al.* 2010). These climate perturbations can influence SS generation and transport directly, through rainfall-induced detachment of soil particles and resuspension of stored sediment on land and in river channels. They can also alter SS flux at catchment scale through indirect effects on vegetation growth, both positive and negative. The combined effect can lead to multiple possible future trajectories in SSC dynamics in rivers with climate change. Some studies suggest that increased vegetation growth triggered by higher temperatures and carbon dioxide concentrations in the air will reduce soil erosion, despite a higher frequency and magnitude of precipitation events and runoff (Tietjen *et al.* 2017). Other studies conclude that rainfall has less effect on sediment generation than the catchment hydrological regime, controlled by lithology and geological characteristics, which determines the partitioning of rainfall on the land surface (Fortesa *et al.* 2021). Natural responses to climate change could be compounded or mitigated by human activity, such as LULC changes, engineering infrastructure and erosion control methods implemented across the landscape. However, suspended sediment transport is not linearly responsive to natural processes or anthropogenic activities (Bussi *et al.* 2016). While rainfall and discharge are often used to predict suspended sediment transport, relationships between these drivers and SSC are variable in space and time due to impacts by different processes (Vercruyse *et al.* 2017). Thus, to better predict future changes to SSC transport dynamics (i.e. sediment regime), further research is needed on the climate and catchment factors that interact to affect the magnitude, frequency and timing of high magnitude SSC events in rivers.

Suspended sediment transport dynamics are often investigated explicitly with discharge, either in the form of ratings curves or sediment loads (Vercruyse *et al.* 2017), but an approach that considers the magnitude, frequency and timing of high SSC events is important for sediment management (Reckendorfer 2019; Shin *et al.* In Press). A magnitude, frequency, timing approach can also be analysed using regression approaches to determine the contribution of climate and catchment factors, as done in previous research to predict long term mean annual SS loads (Roman *et al.* 2012; Grauso *et al.* 2021) and total SS loads (Ayadi *et al.* 2010). A deeper understanding of climate and LULC controls on SSC and their interactions with catchment characteristics is crucial to predicting changes in the magnitude, frequency and timing of high SSC in rivers in the future.

The aim of this study is to determine the influence of climate and catchment characteristics on suspended sediment transport dynamics in rivers. The objectives are to (i) develop statistical and spatial models of indicators of SSC dynamics (magnitude, frequency and timing) at annual and seasonal scales in order to (ii) assess the relative contribution of climate and catchment characteristics. The aims and objectives are achieved through analysis of long time series of SSC data from river gauging stations distributed across the contiguous United States of America (USA). This statistical approach can generate empirical models capable of differentiating the relative contributions of climate (Roman *et al.* 2012; Eurich *et al.* 2021) that have the potential to be more transferable than process-based modelling (Girolamo *et al.* 2015).

STUDY AREA AND DATA

SSC data were obtained from the United States Geological Survey (USGS), which has a large database of SS monitoring data, collected over a long time period, and encompassing a range of climatic regions and catchment characteristics (U.S. Geological Survey 2016). The study started with daily SSC (mg/l) data from 1,667 sites and daily discharge (m³/s) from 1,605 sites from USGS distributed across the continent and overseas territories of the US. Text files containing each SSC (Parameter code: 80154) and discharge

(Parameter code: 00060) were downloaded from USGS Surface-Water Daily Data for the Nation. More information on the collection and processing of the SSC data can be found in *Shin et al. (In Press)*. The spatial extent of the study was reduced to the contiguous U.S., due to limitation of the site characteristics datasets in Geospatial Attributes of Gages for Evaluating Streamflow (GAGES) (e.g. environmental features and anthropogenic influences) in the remaining areas of the USA (*Falcone et al. 2010*). The number of sites was then reduced further due to SS data availability and the application of site selection criteria.

Suspended sediment data and pre-processing

The SSC data were processed, and final sites were selected, according to five steps. First, sediment monitoring sites were filtered to include only those co-located with a river discharge gauging station. Secondly, extreme outlier sites, which could be linked to geological events, were identified and removed. For example, the SSC monitoring site on the Yakima River (Sunnyside, Washington) was removed from the dataset because SSC records were affected by the 1980 Mount Saint Helens volcanic eruption. Thirdly, sites were selected that had a good completeness of SSC data, defined as 80% SSC data/year. To ensure that ephemeral and intermittent rivers were included in the study, this threshold was applied to days on which discharge was recorded at the gauging station (*Goodrich et al. 2018*). Fourthly, sites were filtered further by selecting those that had at least 10 years of data (non-consecutive) within the 30-year period, 1971–2000. This time period was selected because it corresponded with the period used for estimation of catchment characteristics (*Falcone et al. 2010*) and is consistent with climate change impact studies (*Yue et al. 2020*). Finally, gap filling was done using Kalman’s smoothing for short gaps in data (≤ 3 days), because the approach to calculating the frequency of high SSC events is sensitive to gaps (*Moritz & Bartz-beielstein 2017*). A total of 120 sites passed the selection process and were included in this study (*Figure 1*).

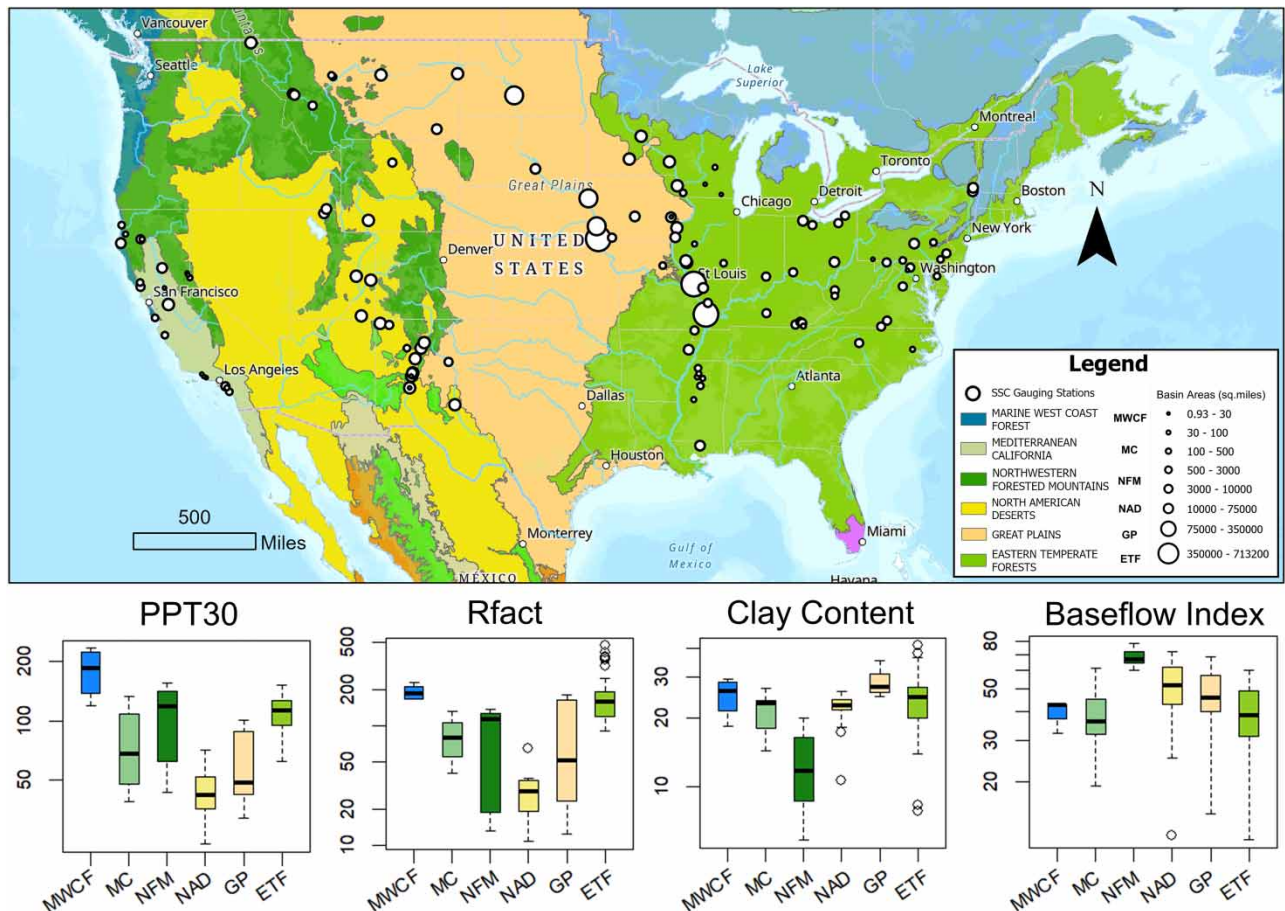


Figure 1 | Top: Data were analysed from 120 sites across the contiguous USA, which encompasses six ecoregions (level-I; *Falcone et al. 2010*). Bottom: Summary of climate and catchment characteristics by ecoregion, including average annual precipitation (PPT30; cm/year), 30-year rainfall and runoff factor (Rfact; 100’s foot tonforce inch/acre/h/year)10), clay content (%), and baseflow index (%). Please refer to the online version of this paper to see this figure in colour: <http://dx.doi.org/10.2166/nh.2023.127>.

The daily SSC dataset was used, rather than turbidity, despite concerns raised by previous research regarding its use for calculating annual sediment yield, in which a lack of strong correlation between SSC and turbidity or suspended solids was noted (Sommerfield 2016). However, analyses for this study found generally high R^2 and, importantly, coherence in the timings of rises and falls between SSC and FNU (Formazin Nephelometric Unit) turbidity datasets (Supplementary material, Figure A1). The observed correlations for these sites provide confidence in the use of the SSC daily data for this study to investigate catchment and climate influences on suspended sediment dynamics.

Site attribution data

The study used summary data on climatic and catchment characteristics from USGS for all 120 sites (Table 1). This dataset aligns with the SSC data pre-processing period; climate data was compiled from the 30 year time period (1971–2000) (Daly *et al.* 2008) and LULC calculated for the year 2001 (GAGES) (Falcone *et al.* 2010). The site attribute data are broadly grouped into climate, land use, land cover, soils, and river hydrology data. Ecoregion distinguishes areas with internally consistent LULC and climate in consideration of natural vegetation community (Dodds & Whiles 2004; Omernik & Griffith 2014). The research used Ecoregion level-I (7.0, 11.0, 6.0, 10.0, 9.0, 8.0), which for this study included Marine West Coast Forest (MWCF), Mediterranean California (MC), Northwestern Forested Mountains (NFM), North American Deserts (NAD), Great Plains (GP) and Eastern Temperate Forest (ETF).

METHODS

The research used a multivariate statistical analysis approach to determine the contribution of catchment and climate explanatory factors to spatial and temporal variations of suspended sediment dynamics. The following steps were used: (i) calculation of SSC dynamics indicators, (ii) data exploration and a principal component analysis (PCA) to inform the reduction in the number of SSC dynamics indicators to model, (iii) regression analysis using ordinary least squares regression and the removal of spatial autocorrelation with geographically weighted regression (GWR), and (iv) cluster analysis and statistical analysis of GWR coefficients by ecoregions.

SSC dynamics indicators

Indicators to describe SSC dynamics in magnitude (e.g. percentiles and rates of changes in mg/l), frequency (durations and counts) were calculated at annual and seasonal timescales (Shin *et al.* In Press). The indicators were calculated from the daily SSC time series, separately for each year of record, and then averaged annually and seasonally. A total of 80 SS dynamics indicators were initially calculated (Table 2).

Table 1 | Climatic and catchment characteristics (Falcone *et al.* 2010)

Category	Abbreviation	Content
Climate	PPT30	Mean annual precipitation (1971–2000, cm)
	Rfact	30-year rainfall and runoff factor ($100's \text{ ft-tonf in h}^{-1} \text{ ac}^{-1} \text{ year}^{-1}$)
Land Use	Agperc	Percentage of agricultural land (sum of NLCD classes 81 and 82, %)
	Urbanperc	Percentage of urban land (sum of NLCD classes 21, 22, 23, and 24, %)
Land Cover	Perm	Avg. soil permeability (inches/h)
	K-fact	Avg. K-factor for the uppermost soil horizon (unitless)
	Forestperc	Percentage of forest land (sum of NLCD classes 41, 42, and 43, %)
	Silt	Avg. percentage of soil silt content (%)
	Sand	Avg. percentage of soil sand content (%)
	Clay	Avg. percentage of soil clay content (%)
	Eco L1 Code	Level-I Ecoregion Code
River Environment	Catch.A	Upstream drainage area (sq. miles)
	BFI	Avg. baseflow index (%)
	Flow.50p	50th percentile of daily flows
	Flow.95p	95th percentile of daily flows
	Major_Dams	Number of dams (≥ 50 feet in height or having storage $\geq 5,000$ acre-feet)
	R.storage	Dam storage (acre-feet)

NLCD, National Land Cover Database.

Table 2 | The indicators calculated to quantify SS dynamics annually and seasonally

SS indicators	Indicators calculated for statistical analyses (average annual)	Abbreviation	Number of indicators	Seasons (4 ×)
Magnitude	(1) Magnitudes in percentiles (95th, 75th, 50th, 25th)	M	4	4
	(2) Rising/falling rates	R.r/F.r	2	2
	(3) Magnitude on a peak day value	Mpkday	1	1
	(4) Standard deviation of monthly max/min	StdevMax/Min	2	×
	(5) Annual average change rate	Aacr	1	1
Frequency	(1) Event durations ($\alpha = 0.975, 0.9, 0.6, 0.2$)	D	4	3
	(2) Counts of events ($\alpha = 0.975, 0.9, 0.6, 0.2$)	C.of.e	4	3
	(3) Peaks/event ($\alpha = 0.975, 0.9, 0.6, 0.2$)	P.p.e.	4	×
	(4) Duration between peaks	D.b.p.	1	×

Seasonal subset values of standard deviation of monthly max/min as well as counting peak-related indicators were not calculated due to less variance observed in the annual data

Intra-annual variation in SSC dynamics was investigated seasonally so that the potential impacts of seasonally varying land management practices for the different LULC units (%) could be explored: winter (December, January and February; DJF), spring (March, April and May; MAM), summer (June, July and August; JJA), and autumn (September, October and November; SON) (Table 2). All pre-processed time series were divided into seasons, and magnitude and frequency indicators were recalculated. Visual checks of time series were done at each step to verify whether there were impacts from missing months as well as yearly gaps in time series.

Preliminary statistical analysis – principal component analysis

A PCA was conducted, using the ‘factoextra’ package in R (R Core Team 2023), to explore the multivariate dataset and reduce the number of response variables in the subsequent statistical models (i.e. the SSC dynamics indicators) (Haddadchi & Hicks 2021). The analyses used all continuous and discrete (e.g. C.of.e. and P.p.e.) response variables to identify SSC indicators that correspond to different dimensions in the PCA plots, aided by K-means clustering (Fang *et al.* 2012). The redundant variables with low contribution were removed.

Regression analysis – OLS and GWR

OLS regression

OLS linear regression analysis was used to determine the relationship between climatic and catchment characteristics (Predictor variables) and the SSC dynamics indicators (Response variables) for the selected 120 sites, annual and seasonal. Data transformation was conducted to produce more symmetric residuals with linearity and homoscedasticity. A log 10 transformation was conducted on the response variables to achieve normality (Cho & Lee 2018). The OLS linear regression analysis was conducted using a best subset selection approach with the reduced number of response variables identified in the PCA and all predictor variables. In addition to the identified significant predictor variables, the OLS analysis was deemed appropriate to bypass detrending of the input variables spatial prior to exploring the spatial patterns in the GWR. Due to the number of predictor variables (<30) (Nakariyakul & Casasent 2007), the ‘leaps’ package with Branch and Bound method (Fouad *et al.* 2020; Lumley 2022) was used to develop parsimonious models in R (R Core Team 2023). The subset selection process started with five independent variables for each model, which was one more than the number of categories of independent variables (Table 1). Variables were added or removed to maximise Adjusted R² and minimise Mallows’ Cp (Cp, a variant of AIC – Akaike Information Criterion) and the Bayesian information criteria (BIC).

Cross validation of the models (K-fold) was done to determine predictor variables with the least prediction error using the ‘caret’ package in R (R Core Team 2023). A five-fold model was used (Max *et al.* 2023) as informed by recent research in environmental sciences and earth surface processes (Coker *et al.* 2021; Su *et al.* 2021), with the data divided into test (20%) and training (80%) subsets. The number of independent variables in each model was determined based on the least cross validation errors. Then, the variance inflation factor (VIF) was checked for multicollinearity, removing independent variables with VIF greater than 10 (Arabameri *et al.* 2020; Wang & Wang 2022), followed by checking interaction terms between independent variables. Finally, predictor variables with statistical significance greater than *p*-value of 0.05 were removed.

Geographically weighted regression

GWR was conducted, using the 'spgwr' package in R, to reduce the spatial autocorrelation in the linear regression models as a means to explore spatial patterns in the predictor variables for the SSC indicators (Ha & Tu 2018). In addition, GWR incorporates non-linear elements between sites allowing independent quantification of regression coefficients between sites by estimation of weighted least squares (Middya & Roy 2021) (Equation (1)).

$$y_i = \beta_{i0} + \sum_{k=1}^n \beta_k(\mu_i, \nu_i) x_{ki} + \delta_i \quad (1)$$

where y_i is the response variable at location i . β_{i0} is the intercept at location i . $\beta_k(\mu_i, \nu_i)$ is the local coefficient with coordinates of a location denoted by μ_i and ν_i . x_{ki} is the k th predictor variable. n is the number of predictor variable. δ_i is the random error at location i .

The Weighted Least Square (WLS) method estimates the local parameters, $\beta_k(\mu_i, \nu_i)$, allocating different weight to each site. The estimation is done in a form of matrix, which is,

$$\hat{\beta}(\mu, \nu) = (X^T W(\mu, \nu) X)^{-1} X^T W(\mu, \nu) Y \quad (2)$$

where X is a matrix of predictor variables, Y is a vector of response variables, $\hat{\beta}(\mu, \nu)$ represents a vector of local regression parameters, $W(\mu, \nu)$ represents geographical weighting from regression location as a diagonal matrix, which puts more weights to closer points in the regression.

An important step in GWR is setting the bandwidth for the spatial weighting ($w(g)$), based on a Gaussian scheme (Equation (3)). The bandwidth type (e.g. fixed or adjusted) was selected considering the spatial variability in the density of the gauging stations and modelling performance (e.g. higher R^2 and less AIC).

$$w(g) = \exp\left\{-\left(\frac{1}{2}\right)\left(\frac{d}{h}\right)^2\right\} \quad (3)$$

where d is the distance between the observations, and h is the bandwidth.

Spatial autocorrelation was calculated using Moran's I for OLS and GWR using the coordinates available for each site and distance of derived bandwidths.

Spatial patterns and differences in GWR coefficients

The spatial distribution of GWR coefficients was investigated directly and indirectly using K-means clustering (Kopczewska & Ćwiakowski 2021). Patterns in GWR coefficients were related to continental scale variations in precipitation. Those were summarised by Ecoregion Level-I. To aid visual comparison of results, only the two predictor variables with the first and second highest relevant weights are presented in figures in the results (Johnson 2010). The significant differences in mean values of the GWR coefficients by Ecoregion were evaluated using one-way ANOVA and Tukey's test (Oh *et al.* 2019).

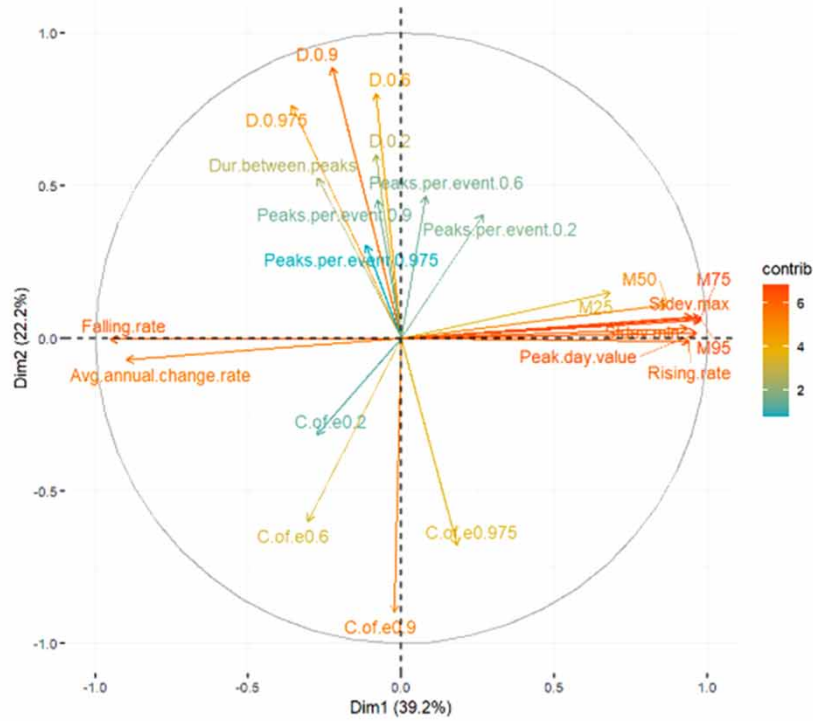
Finally, given the large number of SSC indicators investigated, K-means clustering was conducted to identify regions of the U.S. with distinct combinations of GWR coefficients (ArcGIS Pro 2.8). All SS dynamics indicators and statistically significant predictor variables were included in the cluster analysis with cluster size determined by the spatial coverage of Ecoregion level-I. Side-by-side boxplots were used to check variability between predictor variables. Maps of standard error (s.e.) of each GWR model were produced to check locally different reliability of the model and possible local multicollinearity due to implicit constraints tied to the local coefficients, i.e. larger deviations in SSC magnitude (Wheeler & Tiefelsdorf 2005).

RESULTS

Preliminary statistical analysis – PCA

The PCA results using the pre-processed SSC data identified covariation in the response variables. The first principal component from the PCA corresponded with the magnitude indicators and rising/falling rates, whereas the second component corresponded with duration and count of events (Figure 2(a)). Magnitude indicators were positively correlated with each

(a) Result of 24 annual indicators



(b) Result of selected indicators

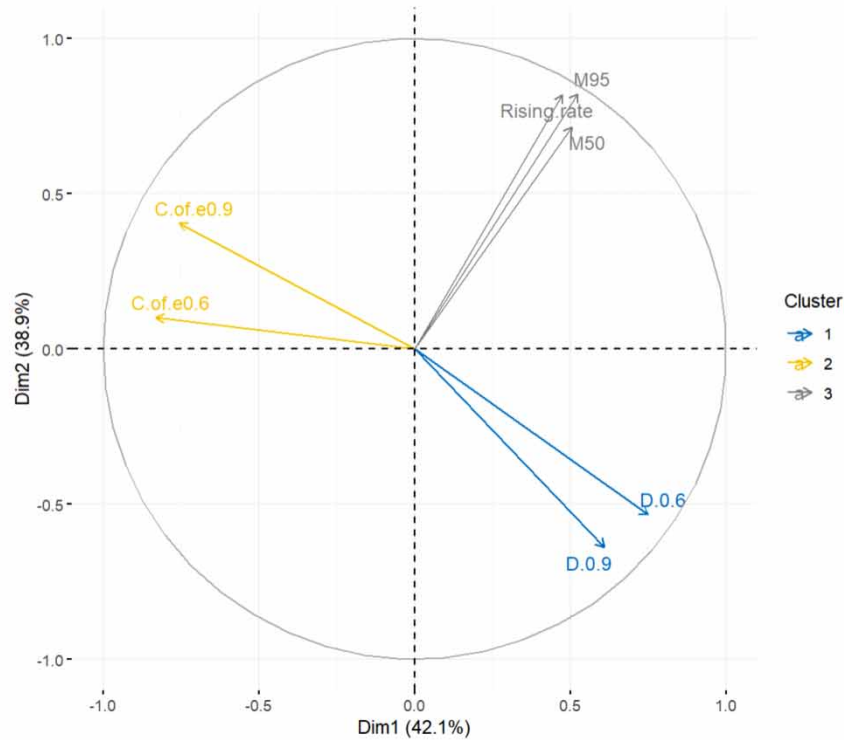


Figure 2 | PCA results of (a) all annual SSC dynamics indicators and (b) indicators selected with K-means clustering, in which redundant indicators with less significant meanings (Low contribution) were removed. Please refer to the online version of this paper to see this figure in colour: <http://dx.doi.org/10.2166/nh.2023.127>.

other, except for falling rates and annual average change rates. Similarly, for frequency, the duration of events indicators were positively correlated with each other, but negatively correlated with count of events. The K-means clustering analysis applied to check the PCA results identified clear distinction between the magnitude indicators and two clusters of frequency indicators: (i) three magnitude indicators – M95, M50 and R.r, (ii) two frequency indicators related to the duration of events ($\alpha = 0.9$ and 0.6), and (iii) two frequency indicators of count of events ($\alpha = 0.9$ and 0.6 ; Figure 2(b)). The seven indicators were taken forward for further analyses. Similar results were obtained for the PCA with indicators at seasonal resolution (Supplementary material, Figure A2), therefore, the same sub selection of indicators was used for the seasonal models. The spatial distribution of indicators is presented in Supplementary material, Figure A3, and the trend analysis is in Supplementary material, Figure A4.

Regression analyses

For the seven SSC dynamics indicators identified in the PCA analysis, a total of 35 OLS regression models (annual and seasonal) were initially developed for regression analyses. The regression models included different predictor variables and interactions for the SSC indicators, but the most common predictor variables were LULC (Urbanperc, Agperc, Forestperc), BFI, clay content and PPT30 (Table 3). The GWR model with fixed bandwidths accounted for spatial autocorrelation in the regression models, thus yielded models with better R^2 and AIC (Supplementary material, Table A1). Following tests for significant spatial autocorrelation in the residuals of the GWR model, 23 GWR models had successfully removed autocorrelation and were studied further (Table 3).

M95 had the greatest number of models, with statistically significant models created for annual and all four seasons (Table 3). For M50, models were generated for winter (DJF) and spring (MAM). For rising rate, models were created for annual and all seasons except summer (JJA). For these magnitude indicators, there were similar predictor variables and directions of effects (positive, +, or negative, -), generally PPT30 (-), BFI (-), Urbanperc (-), Agperc (-), Clay(+/-), and BFI \times Clay (+), though Urbanperc was not included in the model for rising rate. There is some suggestion of seasonal variations in the impacts of predictor variables; Forestperc appears in two seasonal models, the direction of the impact of clay content varies considerably between seasons, and BFI \times Clay appears in models for winter (DJF).

For the duration of longer events ($\alpha = 0.9$), statistically significant models were created for all four seasons, but few predictor variables were common amongst them, except BFI (+). For the duration of shorter events ($\alpha = 0.6$), models were significant for summer (JJA) and autumn (SON) with Forestperc exhibiting a negative effect in both seasons. An annual model and two seasonal models were created for count of events for both longer and short events showing generally positive associations with LULC and, for $\alpha = 0.9$, a positive association with K-fact.

Spatial patterns and differences in GWR coefficients

The GWR model adjusts the regression coefficients to account for spatial autocorrelation. The spatial distribution of these coefficients identifies geographical areas in which the influence of a variable is higher or lower than the original regression model. For many of the models, particularly for the magnitude indicators, PPT30 was the variable that explained the most variation in the linear regression, i.e. the highest relative weight based on the heuristic method (Johnson 2010). PPT30 showed a general spatial pattern of higher GWR coefficients in the coastal regions (MWCF, MC, and ETF) and a lower coefficient in the driest region (NAD) (Figures 3–5), which is similar to the spatial pattern observed for some of the climate and catchment characteristics (Figure 1). However, models for the frequency indicators showed distinctly different regional patterns in the GWR coefficients for LULC (Figures 6 and 7). The spatial patterns by ecoregion for all GWR coefficients are shown in Supplementary material, Figures A10–A16; only a subset are included here for presentation and later discussion.

GWR coefficients varied spatially and seasonally, suggesting differences and shifts in dominant factors affecting sediment erosion and transport. For example, PPT30 had a negative effect on M95 in all ecoregions, except MWCF, at annual resolution (Figure 3). However, the GWR coefficient for PPT30 was positive only in spring (MAM) and autumn (SON) (Figure 3). Similarly, GWR coefficients for PPT30 were higher in some of the ecoregions, e.g. MC and NFM, in the autumn (SON), with some shifting to positive, and were much lower in other ecoregions in summer (JJA), e.g. NAD. In general, clay content had the 2nd highest weighting in the linear regressions for M95. At annual resolution, clay content was negatively associated with M95 in all ecoregions, but there were substantial differences seasonally. GWR coefficients for clay content were strongly positive in spring (MAM) and autumn (SON) (Figure 3), but negative in winter (DJF) (Supplementary material, Figure A10). In spring (MAM) and autumn (SON), clay content has the greatest positive effect on M95 in the

Table 3 | Final statistical models without significant autocorrelation in the residuals ($n = 120$, p -value > 0.05)

Response Variable	OLS Model																	GWR (Fixed bandwidths)		
	Season	Intercept	PPT30	Rfact	Kfact	BFI	Urbanperc	Agperc	Forestperc	Clay	Sand	BFI x Clay	Rfact x Clay	R.storage	Catch.A	Catch.A x BFI	R ²	R ²	Observed Moran's I	
M95	Annual	5.80	-1.00×10^{-2}			-6.00×10^{-2}	-1.10×10^{-2}	-1.00×10^{-2}		-4.20×10^{-2}		2.00×10^{-3}						6.30×10^{-1}	8.30×10^{-1}	-7.20×10^{-2}
	DJF	5.44	-5.00×10^{-3}			-6.30×10^{-2}		-1.20×10^{-2}		-4.30×10^{-2}		2.00×10^{-3}						5.90×10^{-1}	6.80×10^{-1}	-2.90×10^{-2}
	MAM	3.73	-1.00×10^{-2}			-1.70×10^{-2}	-9.00×10^{-3}	-8.00×10^{-3}		4.00×10^{-2}								5.50×10^{-1}	8.20×10^{-1}	-7.40×10^{-2}
	JJA	2.29	-2.80×10^{-2}	1.10×10^{-2}					1.00×10^{-2}	7.60×10^{-2}				-2.00×10^{-4}				5.90×10^{-1}	8.40×10^{-1}	-5.50×10^{-2}
	SON	3.71	-1.10×10^{-2}			-2.20×10^{-2}	-1.20×10^{-2}	-1.10×10^{-2}		4.60×10^{-2}									5.10×10^{-1}	8.70×10^{-1}
M50	DJF	4.25	-8.00×10^{-3}			-4.90×10^{-2}	-1.10×10^{-2}			-4.40×10^{-2}		2.00×10^{-3}						5.10×10^{-1}	7.60×10^{-1}	-4.60×10^{-2}
	MAM	3.06	-1.30×10^{-2}			-1.30×10^{-2}	-1.20×10^{-2}	-6.00×10^{-3}		3.80×10^{-2}								6.30×10^{-1}	7.90×10^{-1}	-3.80×10^{-2}
R.r	Annual	3.25	-9.00×10^{-3}			-2.30×10^{-2}		-1.30×10^{-2}		3.50×10^{-2}								4.60×10^{-1}	7.70×10^{-1}	-6.60×10^{-2}
	DJF	4.60				-5.60×10^{-2}		-1.60×10^{-2}	-7.00×10^{-3}	-4.20×10^{-2}		1.00×10^{-3}						5.40×10^{-1}	5.90×10^{-1}	-1.30×10^{-2}
	MAM	3.00	-8.00×10^{-3}			-2.10×10^{-2}		-1.00×10^{-2}		3.90×10^{-2}								4.60×10^{-1}	7.50×10^{-1}	-6.80×10^{-2}
	SON	4.96	-7.00×10^{-3}			-6.10×10^{-2}		-1.20×10^{-2}		-4.60×10^{-2}		2.00×10^{-3}						5.20×10^{-1}	7.80×10^{-1}	-6.10×10^{-2}
D0.9	DJF	1.12				2.00×10^{-3}			-1.00×10^{-3}	6.00×10^{-3}				8.70×10^{-10}				2.70×10^{-1}	6.90×10^{-1}	-7.90×10^{-2}
	MAM	1.22				2.00×10^{-3}		1.00×10^{-3}						-3.90×10^{-6}	9.50×10^{-8}			2.70×10^{-1}	4.20×10^{-1}	-3.10×10^{-2}
	JJA	1.20				2.00×10^{-3}		2.00×10^{-3}										1.60×10^{-1}	2.70×10^{-1}	-3.30×10^{-2}
	SON	1.36				3.00×10^{-3}			-2.00×10^{-3}	-4.00×10^{-3}								2.70×10^{-1}	5.10×10^{-1}	-5.20×10^{-2}
D0.6	JJA	1.20	1.00×10^{-3}		-8.09×10^{-1}				-3.00×10^{-3}									2.00×10^{-1}	5.80×10^{-1}	-6.20×10^{-2}
	SON	1.01							-2.00×10^{-3}									1.40×10^{-1}	7.00×10^{-1}	-1.58×10^{-1}
C.of.e0.9	Annual	2.81×10^{-1}								3.00×10^{-3}	-4.00×10^{-3}				-7.40×10^{-10}			3.80×10^{-1}	7.10×10^{-1}	-7.80×10^{-2}
	MAM	9.60×10^{-2}				6.88×10^{-1}	-2.00×10^{-3}	2.00×10^{-3}		3.00×10^{-3}		4.00×10^{-3}						3.40×10^{-1}	5.60×10^{-1}	-6.80×10^{-2}
	JJA	1.35×10^{-1}				6.15×10^{-1}		2.00×10^{-3}		2.00×10^{-3}								2.10×10^{-1}	4.80×10^{-1}	-4.20×10^{-2}
C.of.e0.6	Annual	1.35	-2.00×10^{-3}		1.35			4.00×10^{-3}	8.00×10^{-3}									1.90×10^{-1}	4.50×10^{-1}	-2.70×10^{-2}
	DJF	7.59×10^{-1}							2.00×10^{-3}									6.00×10^{-2}	4.30×10^{-1}	-6.90×10^{-2}
	SON	8.84×10^{-1}							2.00×10^{-3}	-3.00×10^{-3}								1.50×10^{-1}	4.80×10^{-1}	-7.30×10^{-2}

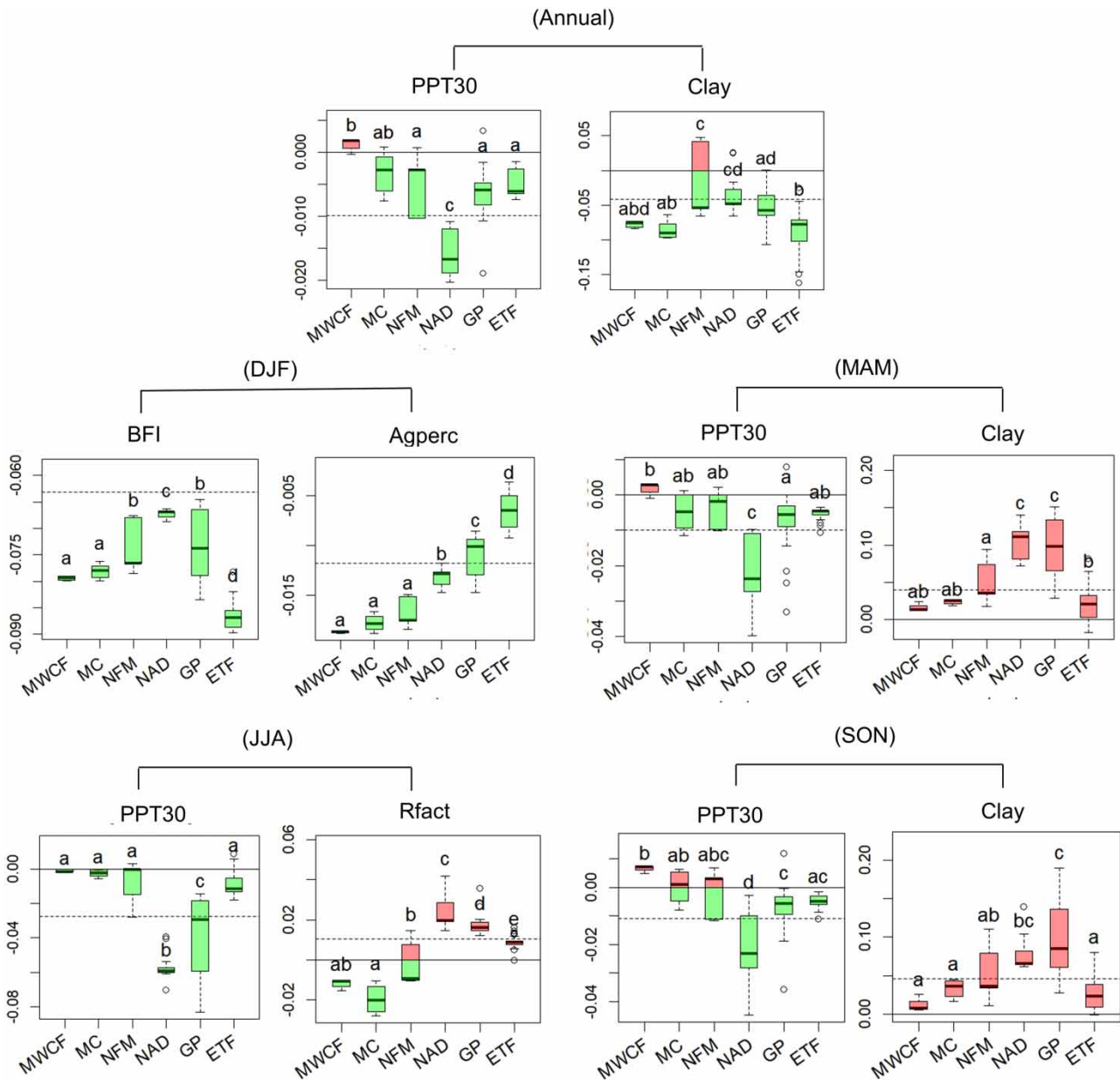


Figure 3 | GWR coefficients of M95 in different ecoregions (Log10 transformed): Horizontal lines in black represent zero, and the dashed horizontal lines are OLS coefficients. Similar mean values in each plot from Tukey's tests are in the same initials above each box plots. Please refer to the online version of this paper to see this figure in colour: <http://dx.doi.org/10.2166/nh.2023.127>.

NFM, NAD and GP ecoregions. Finally, the predictor variables with the highest weighting differed for winter (DJF). BFI had a negative relationship with M95, with the lowest GWR coefficient value in ETF and highest in NAD. Agperc had very different relationships with M95 by ecoregions, with a low negative GWR coefficient in the western MWCF ecoregion rising to a high positive in the eastern ETF ecoregion.

Urbanperc, BFI, Agperc, clay cover and PPT30 were the catchment variables affecting M50, mainly in winter (DJF) and spring (MAM) (Table 3 and Figure 4 and Supplementary material, Figure A11). PPT30 had a consistently negative association with M50 for winter (DJF) and spring (MAM), though there were significant differences across the ecoregions. The general west-east negative unimodal pattern was observed for both seasons, with a greater range observed in winter (DJF). Clay cover was positively related to M50 but significantly lower in MWCF and ETF than in other ecoregions. BFI was negatively associated with M50 overall (Table 3), especially in the winter (DJF) when GWR coefficients are significantly lower for GP and ETF

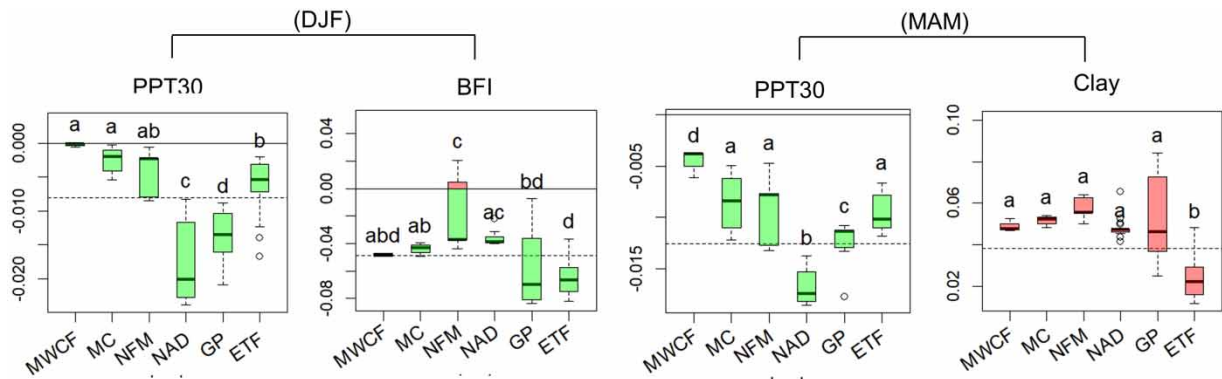


Figure 4 | GWR coefficients of M50 in different ecoregion (Log10 transformed): PPT30 is mean annual precipitation (1971–2000, cm). BFI is the average baseflow index (%). Clay is average percentage of soil clay content. Please refer to the online version of this paper to see this figure in colour: <http://dx.doi.org/10.2166/nh.2023.127>.

(Figure 4). However, BFI was positively associated with M50 in MWCF and MC in spring (MAM) (Supplementary material, Figure A11).

The models predicting rising rates were composed of PPT30 and LULC components, e.g., Agperc, Forestperc, clay content and BFI × Clay (Table 3). However, the predictor variables with the highest weighting were PPT30 and BFI (Figure 5). GWR coefficients for PPT30 displayed the negative unimodal distribution across the ecoregions, identified earlier, with the lowest values (Negative) in NAD and highest (Positive) in MWCF and ETF. BFI was generally negatively associated with rising rates

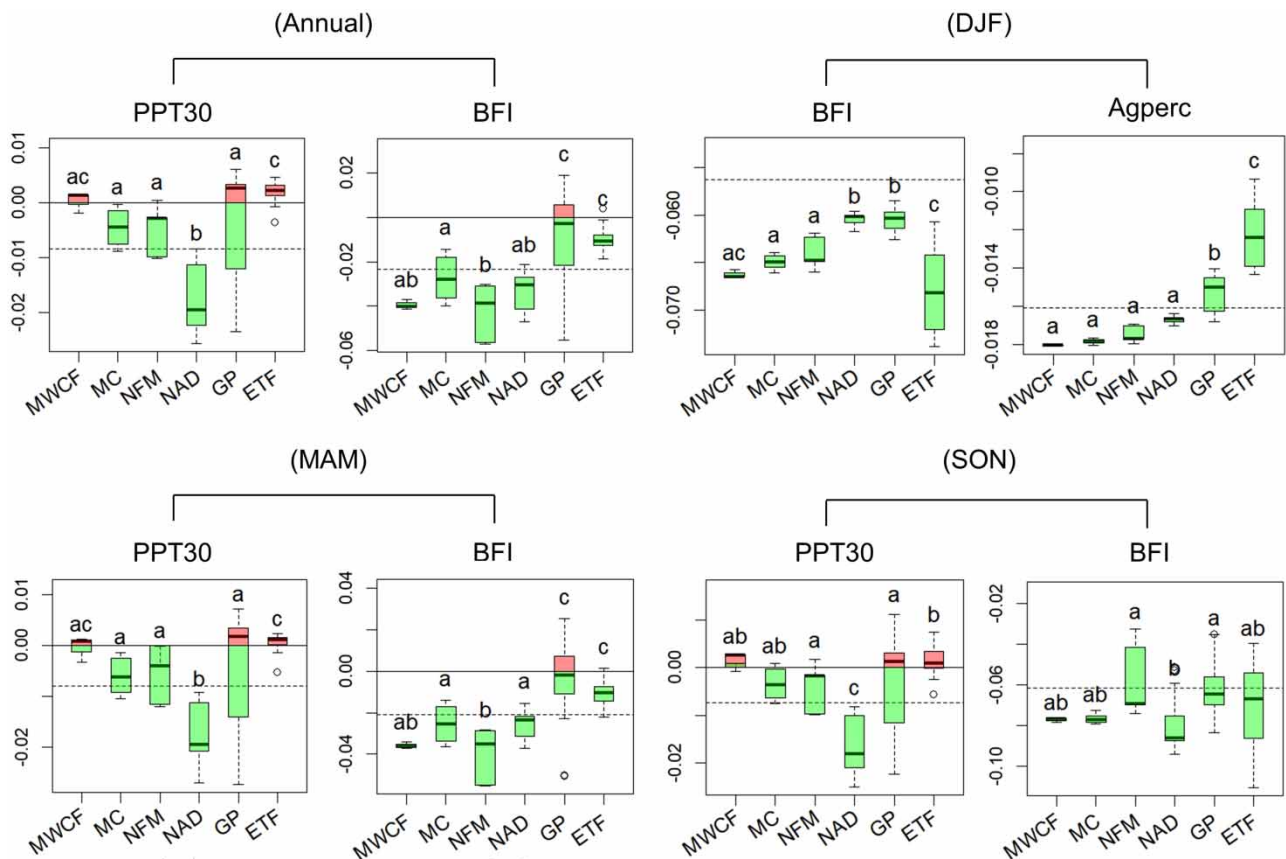
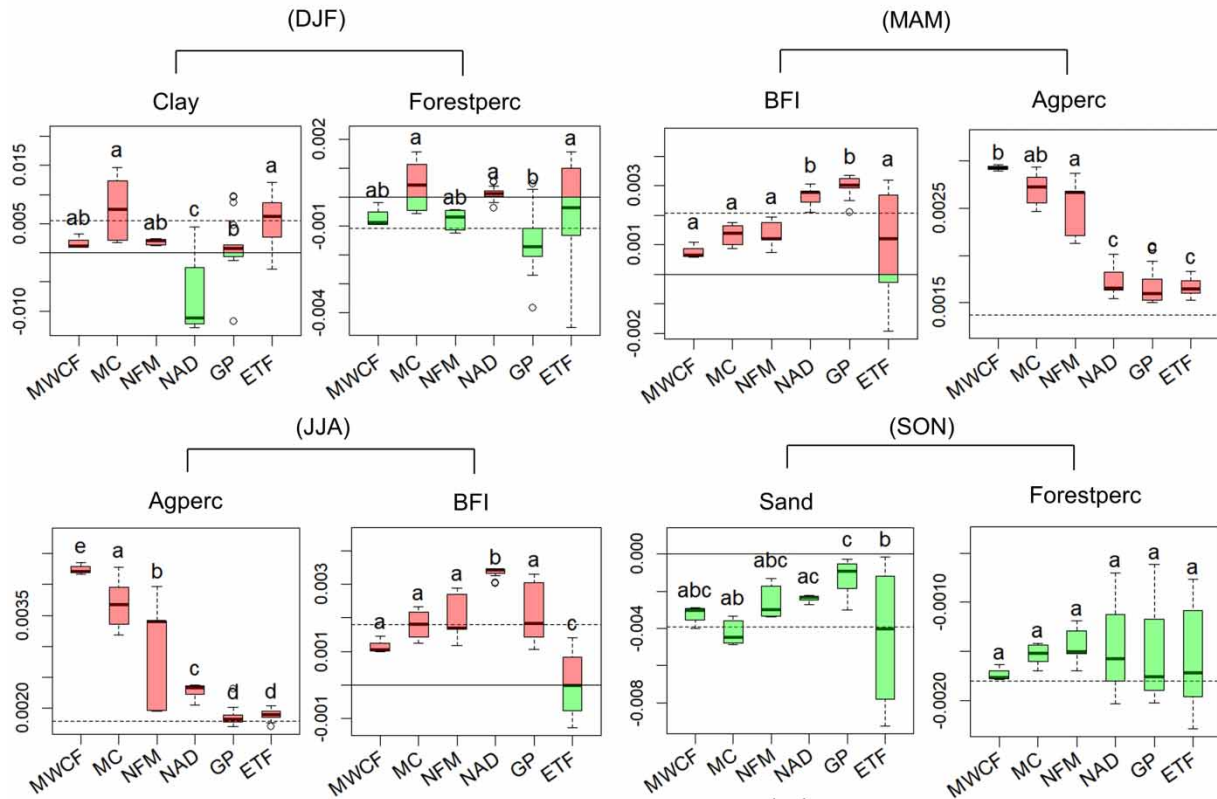


Figure 5 | GWR coefficients of R.r in different ecoregion (Log10 transformed): PPT30 is mean annual precipitation (1971–2000, cm). BFI is the average baseflow index (%). Agperc is average percentage of agricultural area (%). Please refer to the online version of this paper to see this figure in colour: <http://dx.doi.org/10.2166/nh.2023.127>.

(a) $\alpha = 0.9$



(b) $\alpha = 0.6$

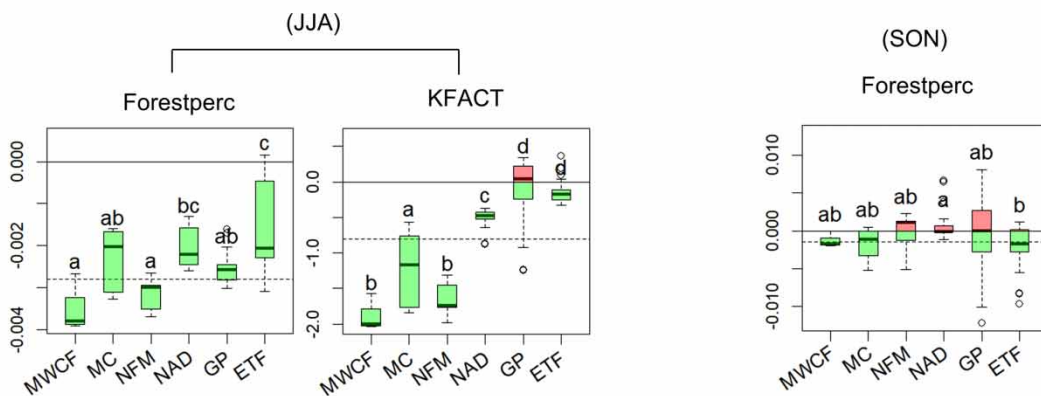
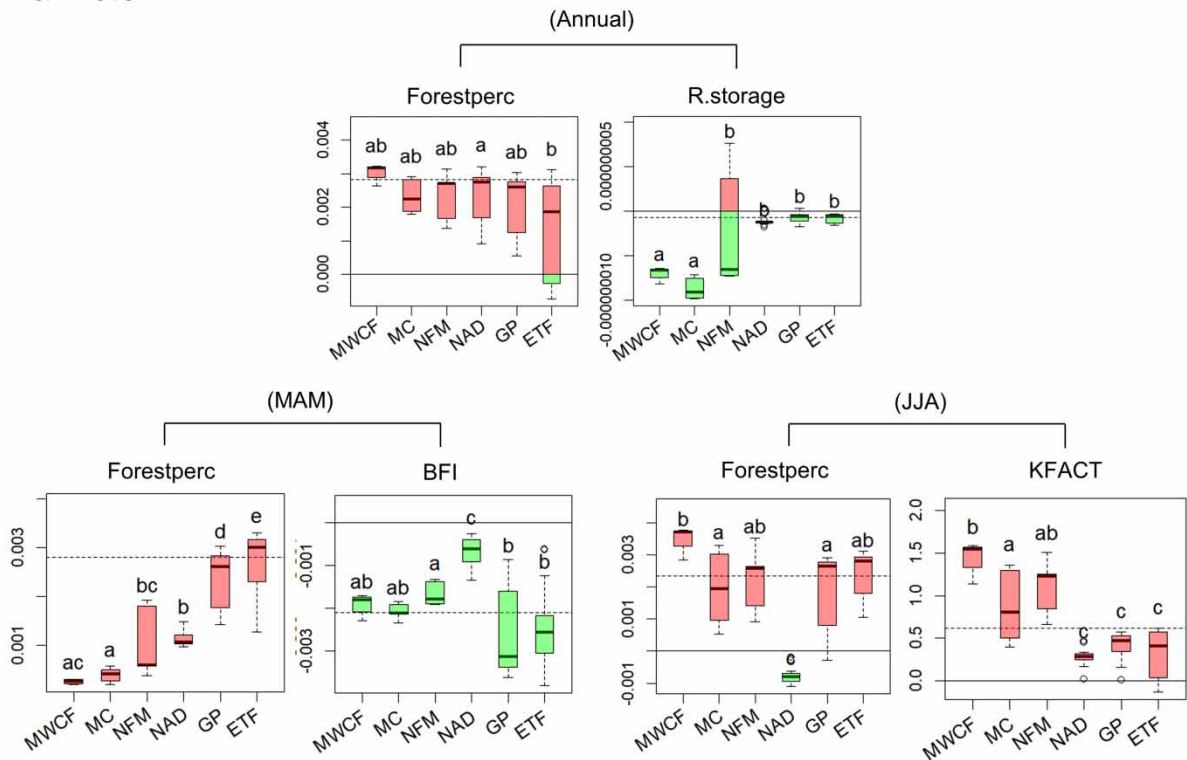


Figure 6 | GWR coefficients of duration of events (a, $\alpha = 0.9$ and b, $\alpha = 0.6$) (Log10 transformed): α from Lyne and Hollick filter that larger value yielded longer duration of event. Forestperc is average percentage of forest cover. Sand is average percentage of sand. K-factor is average K-factor for upper most soil horizon. Please refer to the online version of this paper to see this figure in colour: <http://dx.doi.org/10.2166/nh.2023.127>.

with greater differences by ecoregion in winter (DJF) and spring (MAM). Agperc identified solely with negative coefficients without impact from PPT30.

For duration of events, a much wider range of predictor variables were identified as having high weighting across the seasonal models than for the other magnitude indicators, including Clay, Forestperc, Agperc, BFI, Sand, and K-factor (Figure 6). For the duration of events ($\alpha = 0.9$), Forestperc and soil types (Clay or Sand) had the highest weightings in the linear

(a) $\alpha = 0.9$



(b) $\alpha = 0.6$

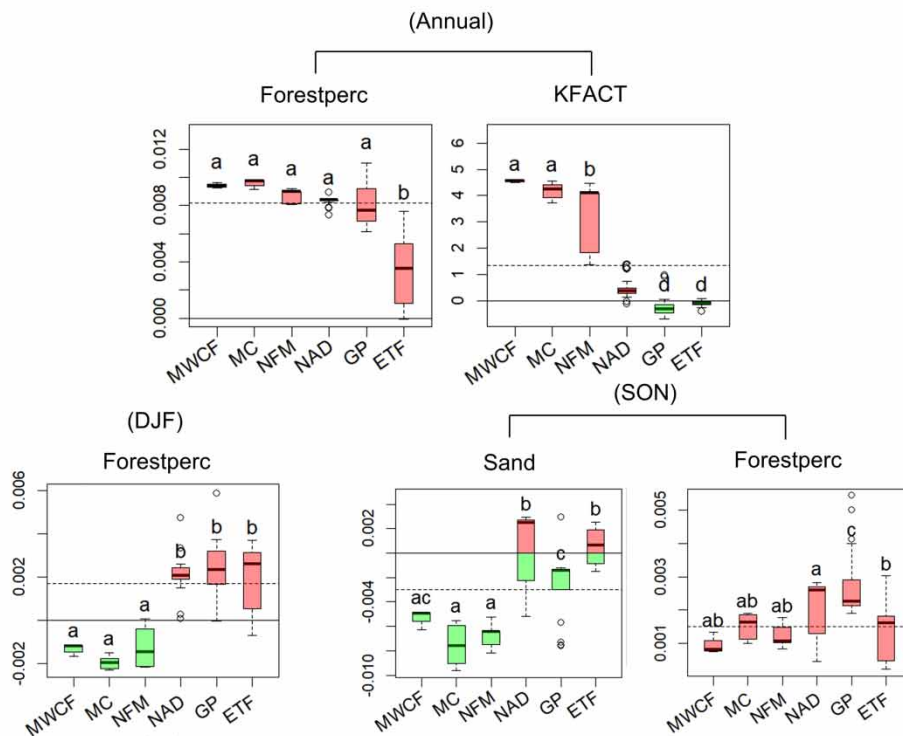


Figure 7 | GWR coefficients of count of events in different ecoregion (Log10 transformed): Urbanperc is percentage of urban land (%). Forestperc is percentage of forest land (%). PPT30 is mean annual precipitation (1971–2000, cm). Please refer to the online version of this paper to see this figure in colour: <http://dx.doi.org/10.2166/nh.2023.127>.

regression variables in autumn (SON) and winter (DJF), while Agperc and BFI had the highest weightings in spring (MAM) and summer (JJA). Forestperc was generally negatively associated with duration of longer events in autumn (SON), though there were mixed effects (positive and negative) in winter (DJF). Clay cover was generally positively associated with duration, while sand content was negative. Agperc showed a pattern by ecoregion that was opposite to M95 (DJF), with the highest GWR coefficients in MMCF and lowest in GP and ETF. BFI had a positive unimodal pattern across the ecoregions, with the highest positive contribution to the duration of long events in NAD and GP. Duration of events ($\alpha = 0.6$) were more related with Forestperc and K-factor. Significant differences by ecoregion were only noted for summer (JJA), when Forestperc and K-factor were negatively associated with duration of shorter events, with significantly lower GWR coefficients in MWCF and NFM than ETF, and GP for K-factor.

For the count of events ($\alpha = 0.9$ and 0.6), Forestperc was identified as a high weighted predictor variable in all models, annual and seasonal, along with K-factor, BFI, Sand and R.storage (Table 3 and Figure 7). It was generally positively associated with count of events with strong spatial variations in some seasons, i.e. lower in the west and higher in the east in the spring (MAM) for longer duration events ($\alpha = 0.9$). The association became negative for count of shorter events ($\alpha = 0.6$) in the western MWCF, MC and NFM ecoregions in winter (DJF). K-fact had a strongly declining impact from west to east, with GWR coefficients highest in MWCF and lowest in NAD, GP and ETF, turning negative for the short duration events ($\alpha = 0.6$). Sand content in the soil was generally negatively associated with the duration of shorter events ($\alpha = 0.6$, SON), except in NAD and ETF ecoregions. Finally, both BFI and R.storage were negatively associated with the duration of longer SSC events ($\alpha = 0.9$).

The results of the K-means clustering analysis identify clear groups of sites with similar GWR coefficient assemblages, suggesting regional characteristics underlying SSC dynamics (Figure 8). K-means and side-by-side boxplots allow further identification of spatially and temporally varying relationships between the indicators and climatic and LULC factors.

Sites located in the western U.S. (e.g. California) grouped together consistently, where M95 and M50 were more strongly associated with PPT30, duration of event ($\alpha = 0.6$) was less strongly associated with Forestperc and K-fact, and count of event was more associated with Agperc and K-fact, relative to the coefficient mean. Sites in the Rocky Mountain region grouped together, except for sites in Montana (Figure 8(c)–8(f)). BFI and Clay had a stronger relationship with M95 in this region in winter (DJF), relative to the coefficient mean. Forestperc with Rfact had a stronger relationship with M95 in summer (JJA). Count of event ($\alpha = 0.6$) was also affected by Forestperc and Agperc in this region. Sites in the Rocky Mountain and the NAD region grouped together for many of the indicators and seasons, especially M95 and the frequency indicators (Figure 8(a)–8(g)). Southern sites group separately in spring, when M50 is associated with Agperc (Figure 8(d)), and again in the autumn when Urbanperc has a greater influence on M95 (Supplementary material, Figure A17). Sites in the Great Plains, Midwest and Mississippi River regions group together for most of the indicators, especially duration of event ($\alpha = 0.6$) and count of event ($\alpha = 0.9$), which were closely associated with K-fact. The regions grouped separated for some of the magnitude indicators. For example, clay content was more closely associated with M95 (JJA) in the Midwest than lower Mississippi. Model coefficients are generally closer to the average than for sites in the western US. Sites (M95) on the Mississippi cluster together all year, except in summer (JJA) (Figure 8(b)). Similarly, the Mid Atlantic and Midwest sites group together, except duration and count of events. For example, in the GWR models for count of events ($\alpha = 0.6$), the Mid Atlantic has lower coefficient values for Agperc and Forestperc, in comparison to sites located further west (Figure 8(g)). The sites in New England and Mid Atlantic always grouped together for SSC indicators without relatively significant influences from catchment characteristics, except M95 (Agperc and $BFI \times Clay$), event duration ($\alpha = 0.6$, Forestperc) and event count ($\alpha = 0.6$, PPT30).

DISCUSSION

In this section, we discuss the results of the OLS regression models and how the geographical variations in model coefficients generated by the GWR models provide insight into the interplay of catchment and climate factors that affect sediment transport processes (i.e. fine sediment generation, transport and storage overland, and transport in river systems) (Table 4). Indicators of SS dynamics were affected by climate and catchment characteristics with additional spatial and seasonal variations that are indicative of further regional controls that were not captured in the predictor variables. These additional controls could relate to climate, physiographic setting, and anthropogenic activities, which influence the annual and seasonal models.

Table 4 | Summary of the key findings of the study, showing the dominant relationships (▲ – positive, ▼ – negative) between catchment and climate factors and indicators of SS transport dynamics

Factors		Key Findings
LULC	Agperc	<ul style="list-style-type: none"> ▼ M95 overall, but ▲ in NAD in spring and in MWCF, MC and NFM in autumn (Figures 3 and 5, Supplementary material, Figure A10) ▲ M50 in MWCF in spring (Supplementary material, Figure A11) ▲ Duration of long events overall, especially in MWCF, MC and NAD in spring and summer (Figure 6) ▲ Count of short events in all ecoregions, except ETF (Supplementary material, Figure A16)
	Forestperc	<ul style="list-style-type: none"> ▼ Duration of events (long and short), except in MC in winter (Figure 6) ▲ Count of long events (Figure 7(a)) and in short events, except in MWCF, MC and NFM in winter (Figure 7(b)) ▲ M95 in summer (Supplementary material, Figure A10)
	Urbanperc	<ul style="list-style-type: none"> ▲ M95 for MWCF, MC and NFM (Supplementary material, Figure A10) ▲ M50 in MWCF in winter and spring (Supplementary material, Figure A11) ▲ Count of long events in spring and summer (Supplementary material, Figure A15)
Soil and hydrogeology	K-factor	<ul style="list-style-type: none"> ▼ Duration of short events, especially MWCF and NFM, but not GP (Figure 6) ▲ Count of long events for all ecoregions in summer, but only NAD, GP and ETF in spring (Supplementary material, Figure A15) ▲ Count of short events in MWCF, MC and NFM (Figure 6)
	Clay	<ul style="list-style-type: none"> ▲ M95 for all ecoregions in spring, summer and autumn, but ▼ in winter (Figures 3 and Supplementary material, Figure A10) ▲ M50 in all ecoregions in spring (Figure 4) ▲ Rising rate in spring, especially in NAP and GP, but ▼ in winter (Supplementary material, Figure A10) ▲ Duration of long events in all ecoregions except NAD in winter (Figure 6(a))
	BFI	<ul style="list-style-type: none"> ▼ M95 and M50, especially in winter (Figures 3 and 4) ▲ M50 in MWCF and MC in spring (Supplementary material, Figure A11) ▼ Rising rate, except in some sites in GP (Figure 5) ▲ Duration of long events in all ecoregions and seasons (Figures 6 and Supplementary material, Figure A13)
Climate	PPT30	<ul style="list-style-type: none"> ▼ M95 and M50 in all ecoregions and seasons, except M95 in MWCF, especially in spring and autumn (Figures 3 and 4, Supplementary material, Figure A10) ▼ Rising rate in most ecoregions, except ETF and, in autumn, MWCF and ETF (Figure 5)
	Rfact	<ul style="list-style-type: none"> ▲ M95 in NAD, GP and ETF in summer (Supplementary material, Figure A10)

MWCF, Marine West Coast Forest; MC, Mediterranean California; NFM, Northwestern Forested Mountains; NAD, North American Deserts; GP, Great Plains; ETF, Eastern Temperate Forest.

LULC impacts on SS transport

LULC was the most common type of predictor variable affecting the SSC dynamics indicators. In the OLS models, at least one of the three LULC classes was included in each model for every indicator. However, the direction and strength of the association varied between indicators and seasons (Table 3 and Figures 3–8).

The magnitude indicators (M95 and M50) were negatively associated with Urbanperc, but the frequency indicators (C.of.e.0.9) were dominated by positive coefficients (Table 3). Therefore, in general, higher cover of urban development was associated with lower extreme and average SSC, but a higher frequency of events (Figure 8(f)). From the GWR model, though, the negative association between Urbanperc and SSC magnitude is largely confined to the driest ecoregion (Supplementary material, Figures A10 and 11). For the rest of the USA, there is either no association (GP and ETF) or it

is positive (MWCF, MC, and NFM). SS is a major contaminant in stormwater runoff from urban areas (Yin & Li 2008; Borris *et al.* 2016), which tends to be more problematic in winter-spring than summer (Westerlund *et al.* 2003; Ciupa *et al.* 2021; Gong *et al.* 2021). Increased connectivity due to storm water drains, which carry material washed from impervious surfaces in urban areas (Gellis *et al.* 2020), and channel incision and scouring due to urban storm water could possibly explain the positive association with M95 and M50 in western regions (Zeiger & Hubbart 2016). The increased flashiness of urban environment explains the increasing number of events. However, the reduction in the magnitude can relate with street cleaning and storm water management. Seasonal differences by ecoregion are likely affected by intra-annual variations in rainfall and discharge (Supplementary material, Figures A10, A11, A15).

Agperc was associated negatively with magnitude indicators (M95, M50 and rising rate), but positively with the duration of long SSC events (MAM and JJA; $\alpha = 0.9$) and the count of short duration events (DJF and SON; $\alpha = 0.6$) (Tables 3 and 4). However, there are some important variations by ecoregion, including a positive association with duration of events (Figure 6(a), MAM), and seasonally, such as a negative association with M95 in winter (DJF; Figure 3). In addition, an increasing pattern from west to east (i.e. fewer negative values) was observed for rising rate. This pattern was most pronounced in DJF (Figure 5) but was also observed in MAM and SON (Supplementary material, Figure A12). A similar west–east pattern was also identified for M95, but only for DJF, the pattern was opposite in SON (Supplementary material, Figure A10). Similarly, this opposite pattern (decreasing positive values, west–east) was also identified in the duration of event ($\alpha = 0.9$; Supplementary material, Figure A13) and count of events ($\alpha = 0.6$; Supplementary material, Figure A16). These results suggest that Agperc affects seasonal variations in the supply of sediment to rivers, which may be related to the timing of agricultural practices or bare soil, which would vary regionally due to the presence of different crops, irrigation, and soil erodibility that affect soil erosion and delivery, often with time lags (Cooper 2009; Darama *et al.* 2021; Haddadchi & Hicks 2021). The agricultural lands consist of NLCD classes 81 (Pasture/Hay) and 82 (Cultivated crops) (Falcone *et al.* 2010), which are LCLU types known to contribute suspended sediment to rivers (Tiecher *et al.* 2017). However, the generally negative GWR coefficients from agricultural lands are not supported by previous research on soil erosion and delivery to channels but could be explained by confounding factors, such as the spatial distribution of LULC (e.g. farms not in the close proximity to river networks) and erosion mitigation measures such as buffer strips and local policies for maintaining soil resources retention (Tiecher *et al.* 2017; Shi *et al.* 2021).

Forestperc was positively associated with the count of long SSC events and negatively associated with the duration of events (long and short) (Table 3). Spatially, the positive association with count varies depending on the seasons but was highest for the count of longer duration events in JJA in MWCF and lowest in the driest region (NAD; Figure 7(a)). The GWR clusters suggest that Forestperc in the Rocky Mountains area experience higher pulses of sediment (M95, Figures 8(b) & Supplementary material, Figure A10) and count of events in summer ($\alpha = 0.9$ in Supplementary material, Figure A15) and autumn ($\alpha = 0.6$ in Supplementary material, Figure A16). While Forestperc was mostly negatively associated with duration, which could relate to higher runoff in managed forests, some sites in MC and ETG had a positive association for longer duration ($\alpha = 0.9$; i.e. seasonal) events. These variable results are related to the impacts of forests on hydrological pathways and soil erosion processes on forest rich areas (Riitters *et al.* 2004; Zimmermann *et al.* 2012), complicated by potential covariations between forest cover, slope steepness, management practices, and climate.

Soil and hydrogeology impacts on SS transport indicators

Soil and hydrogeological properties (K-factor, clay content, and BFI) have significant relationships with SSC indicators. Soil clay content and BFI were identified in models for magnitude (M95, M50, and rising rate) and duration, while K-factor, and to a lesser degree sand content, featured in models for frequency (count of events).

In general, K-factor is positively associated with the number of SSC events (Figures 7, Supplementary material, Figures A15 and 16), thus areas with more erodible soils have a greater frequency of high SSC events. The spatial pattern is very pronounced for shorter duration events, showing a decreasing pattern west to east, with high coefficients in MWCF, MC, and NFM (Figure 7). For longer duration events, the lack of a clear west–east pattern in the annual models is caused by opposing patterns in different seasons; there is an increasing pattern west to east in spring (MAM) but a decreasing one in summer (JJA) (Figure 6 & Supplementary material, Figure A15). These spatial patterns may be due to the timing of high intensity regional rainfall and availability of erodible materials, especially in California and Great Plains (Figure 8(f)). K-factor is associated with soil structure (e.g. particle size distribution and organic matter) and highly variable depending on LULC. The different distribution of K-factor is likely the result of the impacts of both climate and catchment controls (Martínez-Murillo *et al.* 2020).

Soil clay content had a variable association with SSC dynamics indicator, varying between magnitude indicators and seasonally within an indicator. While it was negatively associated with the annual model for M95 and winter (DJF) model for M50, it was positively associated with these magnitude indicators for other seasons (Table 3). The GWR results suggest that clay content may have a greater influence on SSC in some ecoregions, especially NFM, NAD, and GP (Figures 3 and 4). However, duration of events ($\alpha = 0.9$) from the clay cover (Figure 6) did not follow the pattern of Figure 1 in DJF, which may be due to seasonal differences in flashiness of the sedimentograph. In cold regions, snowmelt runoff would cause soil loss in spring (Aygün & Kinnard 2021), which would be exacerbated where there is a high proportion of precipitation as snow and LCLU with bare soil or grass (Tramblay *et al.* 2010; Huang *et al.* 2020), and may contribute to sediment availability and delivery to the channel. In addition, ephemeral rivers are correlated with fine grained channels with silt or clay (Li *et al.* 2021).

BFI was negatively associated with magnitude indicators (M95, M50, and rising rate) and positively associated with duration of longer SSC events ($\alpha = 0.9$) (Table 3). Significant variation was observed in the GWR coefficients spatially and seasonally, with stronger negative associations in winter (DJF) for M95 and M50. For duration and count of events, GWR coefficients of BFI by ecoregion broadly mirror the spatial pattern for BFI (Figure 1), being highest in NFM and NAD, with some evidence of seasonal variations (Figures 5 and 6, Supplementary material, Figure A10–A13). These spatial patterns are likely driven by a combination of climate, geology, soil matrix and groundwater flows (Price 2011). Baseflow is supported by aquifers fed from ephemeral rivers (Goodrich *et al.* 2018). As baseflow is affected by ground water and soil structure, the negative coefficients from magnitude indicators could be due to dilution impacts from precipitation, and the positive effects to the duration of events might be caused more by additional environmental factors that also correlate with BFI and topography (e.g. the Rocky Mountains, NFM) (Thomas *et al.* 2004; Price 2011).

These additional environmental factors may also underly the interaction terms included in the models. A positive association between BFI \times Clay and magnitude indicators was identified in the OLS models (Table 3). The GWR model coefficients displayed a west–east variation that mirrored many of the other variables, i.e. highest along the coasts (MWCF and ETF) for M95, M50 and rising rates (Supplementary material, Figures A10–A12). The K-means clustering identified the northern Rocky Mountains, NAD and Great Plains regions grouped together (Figure 8(a) and 8(c)), and those are areas where BFI \times Clay was strongly positively associated with M95 and M50 (Supplementary material, Figures A10 and A11). These interactions are likely driven by covariations in baseflow, topography and climate, which affect sediment generation, delivery overland to channel, and transport within the river network. For example, mountainous areas often have greater forest cover and a colder climate with a greater proportion of precipitation as snow, which affect rainfall partitioning, hydrological flow pathways, and baseflow generation through snowmelt (Santhi 2008; Huang *et al.* 2021).

Climate impacts on SS transport indicators

GWR coefficients were significantly different by ecoregion (Level-I) and varied seasonally (Figures 3–7), resulting in large-scale spatial patterns in model coefficients (Figure 8), which could be related to climate factors. Many of the GWR model coefficients were higher in the west coast, lower in the central U.S, and higher in the eastern temperate region, similar to the west-to-east pattern in annual precipitation (PPT30) across the contiguous USA (Figure 1). For example, PPT30 is associated positively with M95 in the moist coastal region (MWCF) but switches from positive to negative seasonally in the semi-arid ecoregions and northwest forest (MC and NFM; Figure 3). PPT30 is a positive driver of event frequency in ETF (Supplementary material, Figure A16).

The annual precipitation indices used in this study were generally negatively associated with magnitude indicators of SSC transport dynamics. However, there is evidence of differences by ecoregion that could relate to the timing and intensity of the predictor variable (PPT30). For example, while GWR coefficients were generally lowest for M95 in NAD over most of the seasons, they were significantly lower in GP and ETF in DJF (Supplementary material, Figure A10). In general, the negative relationship between magnitude indicators and PPT30 is likely caused by dilution effects (Kozak *et al.* 2019), which is moderated regionally by different frequencies and intensities of rainfall that affect the generation and transport of sediment (Hancock 2012). However, PPT30 is a static representation of 30-year rainfall (1971–2000), which may not capture well the seasonality inherent in the SS delivery system. Rfact is positively associated with M95 in JJA (Table 3), with the highest coefficients found in the Rocky Mountains, desert and Northern Great Plains (Figure 8(b)). The Rfact values may represent well the intensity and frequency of rainfall runoff generation cycles at these locations in these months (Supplementary material, Figure A10). Although larger standard errors (s.e.) in GWR may indicate the effects of local collinearity (Fadliana

et al. 2019), climate controls reacted to the catchment characteristics more in wetter seasons, e.g. positive impacts of Rfact (M95, Figure 3) and PPT30 (D0.6, Supplementary material, Figure A14) in summer.

Spatial patterns in SSC transport dynamics and predictor variables

This research identified possible climate and catchment controls on SSC dynamics at continental scale using daily SSC and annual average descriptors of potential predictor variables using a regression and spatial analysis approach. Further modelling (e.g. improvement of gap filling techniques for catchment characteristics with finer time intervals, applying AI or machine learning) (Tao *et al.* 2021) can help to validate these findings to support the prediction of changes in SSC dynamics in the future with climate change. While this study used a part of GAGES dataset because of the site coverage and suitability of data types, future studies using SSC data from US gauging stations could consider using both GAGES and GAGES-II. These datasets have a greater number of potential explanatory variables, including seasonal precipitation data, which would align more closely with seasonal summaries of SSC dynamics indicators.

CONCLUSIONS

The research applied recently developed suspended sediment dynamics indicators to investigate how catchment and climate factors influence spatial and seasonal variations in SSC. Regression models were developed for seven SS transport dynamic indicators for 120 SSC monitoring sites on rivers across the US at annual and seasonal timescales. The models included a range of statistically significant predictor variables, of which LULC, PPT30, baseflow index, and soil properties, like clay content, were the most common across the SS transport dynamics indicators. When additional spatial modelling was conducted to remove spatial autocorrelation (GWR), significant differences in coefficients were identified by ecoregion, suggesting regional influences from catchment and climatic factors. Key results are as follows:

- Agricultural percent cover is positively associated with longer duration SSC events and more frequent short events, and seasonally with SSC magnitude, likely due to effects on soil hydrology and sediment supply.
- Forest percent cover is positively associated with shorter and more frequent SSC events, and extreme high magnitude SSC in summer months, which might be related to the impacts of forest management on hydrological processes or covariations with topography and climate.
- Urban percent cover was the least common LULC variable in the models. It is positively associated with the frequency of long SSC events, especially in spring and summer, and with extreme high SSC magnitude in the western US.
- Soil erodibility (K-factor) is positively associated with the frequency of SSC events, but negatively with the duration of short events, especially in the west coast and forested mountain regions. Thus, more erodible soils yield more frequent and short-lasting high SSC events, though climate and catchment variables likely have strong influences.
- Soil clay content is positively associated with average SSC in all ecoregions over most seasons (spring, summer and autumn) and with the duration of long SSC events, especially in winter. Clay content may relate to the availability of fine sediment, but there are spatial covariations with topography and climate that suggest an influence of snowmelt in mountainous regions and ephemeral rivers in dry regions.
- Baseflow index is positively associated with the duration of long SSC events and slow rates of SSC increase, but negatively associated with average and extremely high SSC, which likely relates to reduced overland flow and more stable flow regimes that reduce sediment erosion and transport.
- Annual precipitation is negatively associated with SSC magnitude and rates of SSC increase overall, likely due to dilution effects. However there are important spatial variations with positive associations in moist forested regions (MWCF and ETF).
- The rainfall and runoff factor is positively associated with extreme high SSC in desert, Great Plains and ETF regions in summer, possibly related to intense seasonal storms.

This study is one of the first to investigate the short-term dynamics (days to months) of suspended sediment transport in rivers at continental scale and the underlying catchment and climatic variables that control spatial and seasonal variations. It provides new understanding of the varying influence of LULC on the magnitude and frequency of high suspended sediment concentrations caused by climate and catchment factors. Further research should be conducted on the impact of climate on suspended sediment transport dynamics to inform future monitoring, modelling and mitigation related to the climate change.

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DATA AVAILABILITY STATEMENT

The original SSC data were obtained freely from USGS water data portal (<https://waterdata.usgs.gov/nwis>). The processed data used in the modelling are available at CORD data repository (<https://dx.doi.org/10.17862/cranfield.rd.22233721>).

CONFLICT OF INTEREST

The authors declare there is no conflict.

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