

1 **Biocultural aspects of species extinctions**

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29 **Impact Statement**

30 Human responses to impending extinctions are complex, highly dependent on cultural and
31 socioeconomic context, and have typically been far less studied than the ecological and genetic
32 aspects of extinction. Specifically, the way in which science and societies respond to population
33 decline, extirpation, and species extinction can also have a profound influence on whether a species
34 goes extinct, either positively or negatively. For example, while some rare species suffer higher
35 extinction risk the rarer they become, some charismatic species benefit from significantly higher
36 conservation effort and elevated levels of scientific research. A more comprehensive and nuanced
37 understanding of which species will go extinct, and which will be 'rescued' by conservation and
38 stewardship efforts, requires an explicit interdisciplinary, biocultural approach to extinction that
39 draws on expertise from the natural and social sciences, and dialogue with holders of different
40 knowledge systems, and in particular with Indigenous Peoples and local communities. Ultimately,
41 many currently threatened species will only go extinct if society allows it to occur, either through a
42 lack of motivation, knowledge, resources, or local conservation capacity.

43 **Abstract**

44 Predicting whether a species is likely to go extinct (or not) is one of the fundamental objectives of
45 conservation biology, and extinction risk classifications have become an essential tool for conservation
46 policy, planning and research. This sort of prediction is feasible because the extinction processes
47 follow a familiar pattern of population decline, range collapse and fragmentation, and, finally,
48 extirpation of sub-populations through a combination of genetic, demographic and environmental
49 stochasticity. Though less well understood and rarely quantified, the way in which science and society
50 respond to population decline, extirpation, and species extinction can also have a profound influence,
51 either negative or positive, on whether a species goes extinct. For example, species that are highly
52 sought after by collectors and hobbyists can become more desirable and valuable as they become
53 rarer, leading to increased demand and greater incentives for illegal trade – known as the
54 anthropogenic Allee effect. Conversely, species that are strongly linked to cultural identity are more
55 likely to benefit from sustainable management, high public support for conservation actions and fund-
56 raising and, by extension, may be partially safeguarded from extinction. More generally, human
57 responses to impending extinctions are extremely complex, highly dependent on cultural and
58 socioeconomic context, and have typically been far less studied than the ecological and genetic
59 aspects of extinction. Here, we identify and discuss biocultural aspects of extinction and outline how
60 recent advances in our ability to measure and monitor cultural trends with big data are, despite their
61 intrinsic limitations and biases, providing new opportunities for incorporating biocultural factors into
62 extinction risk assessment.

63

64 **Keywords:** biocultural, culturally important species, anthropogenic Allee effect, culturomics

65 Introduction

66 Extinction is typically viewed as a logical end point of the process of population decline—the point on
67 the graph where the population size curve meets the x-axis and terminates abruptly and finally (Ladle,
68 Jepson, 2008). Accordingly, the International Union for the Conservation of Nature (IUCN) defines a
69 species as extinct if there is no reasonable doubt that the last individual has died (Hughes et al., 2021).
70 Reasonable doubt in this context is the lack of evidence in the face of exhaustive surveys or
71 extrapolation from historical observations (Solow, 2005). The intrinsic uncertainty about extinction
72 led E. O. Wilson to characterize it as “the most obscure and local of all biological processes” (Wilson,
73 1992, p.255). If scientists were to restrict their analysis to only well documented extinctions, there
74 would be a huge risk of underestimating current extinctions (Pimm et al., 2014). Thus, extinctions are
75 often extrapolated onto unknown species, whose existence is inferred from species discovery curves,
76 from biodiversity ratios, or from species area relationships (Bebber et al., 2007, García-Robledo et al.,
77 2020, Kunin et al., 2018, Chisholm et al., 2016).

78 Early conceptualizations of the extinction process strongly emphasized the role of population decline
79 and the effects of small population size on population viability (Caughley, 1994). The former process
80 is the result of deterministic factors such as habitat loss, degradation, and overexploitation, ultimately
81 leading to small fragmented populations that are highly susceptible to stochastic factors (Lande,
82 1998). Accordingly, extinction risk assessment schemes such as the IUCN Red List emphasize the rate
83 of population decline, distributional range, and the size and fragmentation of extant populations as
84 important factors that influence long term prospects of a species’ survival
85 (<https://www.iucnredlist.org/assessment/process>). Likewise, these important insights led directly to
86 the development of the concept of minimum viable populations and sophisticated tools performing
87 population viability analysis (Ladle, 2009, Traill et al., 2010, Flather et al., 2011).

88 In summary, extinction is typically understood to mean the death of the last individual of a species
89 and is a consequence of the process of population decline and the negative consequences of small
90 population size. However, contemporary extinction is almost never a purely biological process (Ladle,
91 Jepson, 2008). Rather, contemporary extinction is almost always a consequence of the interaction of
92 cultural and biological phenomena, i.e. a *biocultural* process. Indeed, cultural practices now influence
93 almost every aspect of the extinction process, how we perceive it, measure it, and act upon it, and,
94 critically, whether a threatened species actually goes extinct (or stays extinct). Echoing calls for the
95 adoption of biocultural approaches to conservation (Garibaldi, Turner, 2004, Gavin et al., 2015,
96 Bridgewater, Rotherham, 2019). We contend that there is a need for a broader conceptualization and
97 exploration of species extinctions as a biocultural process. Here, we outline some of the key biocultural
98 aspects of extinction, including taxonomic change, human impacts on population decline, and how
99 the behaviour of the global conservation movement plays a key role in contemporary extinction
100 dynamics.

101 Taxonomy and extinction

102 Although extinction is sometimes used to describe the extirpation of populations, geographic variants,
103 or subspecies (often called “local extinction”), the term is more commonly applied to the loss of all
104 populations of an officially recognized species (Ladle, Jepson, 2008). Extinction statistics are therefore
105 highly sensitive to changes in taxonomic practice, especially changes in normative use of species
106 concepts and species delimitation criteria (Zachos, 2016). Such changes have become increasingly
107 prevalent, partly as a consequence of advances in molecular taxonomy, leading to significant recent
108 increases (and sometime decreases) in the number of recognized species in many taxa (Garnett,
109 Christidis, 2017). Every taxonomic decision to ‘split’ a species into two or more species or to ‘lump’

110 two or more species into a single species necessarily has consequences for the extinction risk of each
111 newly defined species (Mace, 2004). Splitting means that each newly recognized species will have a
112 smaller population size and/or geographic range, potentially increasing threat level. Such increasing
113 threat may potentially be compensated by the increased conservation attention afforded to a fully
114 recognized species rather than a sub-species or regional variant. It could also reduce the average
115 societal attention given to each newly designated species (Ladle et al., 2019). Moreover, estimates of
116 the numbers of unknown species (sometimes referred to as the Linnean Shortfall; Hortal et al. 2015)
117 are highly sensitive to the number of documented species, with knock-on effects for estimates of
118 global extinction rates (Stropp et al., 2022).

119 **Cultural push factors**

120 The first scientists to study extinction at the end of the 19th century had enormous difficulties
121 attributing the decline and eventual loss of a species to human actions (Ladle, Jepson, 2008). For
122 example, despite clear evidence of overhunting (Bengtson, 1984), James Orton described his
123 confusion about the causes of the extinction of the Great Auk (*Pinguinus impennis*) as follows: “The
124 upheaval or subsidence of strata, the encroachments of other animals, and climatal revolutions—by
125 which of these great causes of extinction now slowly but incessantly at work in the organic world, the
126 Great Auk departed this life, we cannot say” (1869; p. 540). This reluctance to attribute human causes
127 to species extinction continued well into the 20th century (Ladle, Jepson, 2011). In contrast, modern
128 current conceptualizations of the factors driving population declines foreground the indirect and
129 direct role of human actions and how they are shaped by socio-cultural practices and beliefs (Lande,
130 1998, Díaz et al., 2019).

131 Although extinction can occur in the absence of human influence (De Vos et al., 2015), the vast
132 majority of contemporary extinctions are ultimately or proximately connected to human action
133 (Ceballos et al., 2015, Díaz et al., 2019, 2015) and are underpinned by societal values and behaviours.
134 Ultimate causes include human population growth (McKee et al., 2004; McKee, 2003), the seemingly
135 universal desire to accumulate surplus capital (McBrien, 2016), the political need for economic growth
136 (Spash, Smith, 2019), and the grinding hardship of rural poverty that forces individuals into a reliance
137 on natural resource exploitation (Adams et al., 2004). These factors, in turn, drive the proximate
138 causes of population decline, the most significant of which are habitat loss, fragmentation and
139 transformation (Powers, Jetz, 2019, Maxwell et al., 2016), climate change (Cahill et al., 2013, Thomas
140 et al., 2004), biological invasions (Clavero, García-Berthou, 2005) and over-exploitation (Bennett et al.,
141 2002, Maxwell et al., 2016).

142 While habitat loss, climate change, and biological invasions can be considered as by-products of other
143 human activities such as agriculture, trade, transport and recreation; population decline due to over-
144 exploitation of species is a direct consequence of human cultural practices and values. This is clearly
145 illustrated by the anthropogenic Allee effect, the idea that human predisposition to place exaggerated
146 economic value on rarity drives the disproportionate exploitation of rare species, causing them to
147 become rarer and therefore even more desirable (Courchamp et al., 2006, Tournant et al., 2012,
148 Palazy et al., 2012a). Even when a species becomes so rare that it cannot, alone, support livelihoods,
149 opportunistic exploitation, while targeting more common species, will ensure that population decline
150 continues (Branch et al., 2013). For example, Chinese bahaba (*Bahaba taipingensis*) is a highly sought
151 after fish for use in traditional Chinese medicine, but fishermen seeking this species must make their
152 living off other species because only a few are caught each year (Sadovy, Cheung, 2003). The
153 anthropogenic Allee effect, an explicitly biocultural model of extinction risk, is particularly applicable
154 to ‘collectable’ exotic species (Siriwat et al., 2019) or their products, as in the example of traditional
155 Chinese medicine highlighted above (but see Mateo-Martín et al., 2023). More generally, it illustrates

156 the complex ways in which cultural practices and beliefs can become entangled in the process of
157 population decline and extinction. In this case, human perceptions of rarity and economic dynamics
158 interact with the population trends of the exploited species to create an ‘extinction vortex’
159 (Courchamp et al., 2006).

160 The ultimate and proximate drivers of population declines of species are typically considered to be
161 insufficient to cause the actual extinction of a species (Lande, 1998). Instead, populations become so
162 small and fragmented that they become subject to a range of stochastic processes (genetic drift,
163 demographic and environmental stochasticity, natural catastrophes) that ultimately lead to the death
164 of the last individual (extinction) (Lande, 1998). As described above, in a few cases the increasing rarity
165 value of these last individuals vastly inflates their economic value and therefore incentivizes their
166 capture or elimination (Courchamp et al., 2006, Hall et al., 2008). In species without economic value
167 and without sufficient human intervention (see below), remnant populations eventually succumb to
168 one of the many risk factors associated with small populations as highlighted by individual animals
169 that become famous for being the ‘last’ of their species (Nicholls, 2012, Jarić et al., 2023).

170 **Cultural push-back factors**

171 While cultural ‘push-factors’ for extinction are generally well known and quantified, far less attention
172 has been given to the role of humans in delaying or preventing extinction (‘push-back’ factors). Since
173 the emergence of the global conservation movement in the late 19th century (Soulé, 1985),
174 conservationists have increasingly monitored threatened populations and, when deemed necessary,
175 intervened in aiming to halt or slow the extinction process (Hoffmann et al., 2015, Bolam et al., 2021).

176 The increasing capacity of the global conservation community to identify species at risk of extinction
177 and to take action to mitigate this risk highlights the key role that humans now play in the extinction
178 process. In other words, species become extinct (or avoid impending extinction) due to the interplay
179 between human-mediated biological processes (range collapse, population decline, small
180 populations) and human capacity to monitor and successfully intervene in this process (Ladle, Jepson,
181 2008). There is a global safety net provided by the conservation movement as represented by
182 government bodies and various international and national non-governmental conservation
183 organizations (NGOs). In a similar way that extinction threats vary geographically, the motivation,
184 capacity, resources and effectiveness of conservation also vary immensely by country and region
185 (Waldron et al., 2013). Sometimes extinction threats and capacity to deal with those threats align, but
186 frequently there is a mismatch with geographic areas hosting a high frequency of threatened species
187 mainly located in the “Global South” (Schipper et al., 2008) while the capacity of the global
188 conservation movement to act is often more concentrated in the “Global North” (Balmford et al.,
189 2003). While this is generally true, there are many exceptions, and further research into this area is
190 needed.

191 Although conservation capacity is clearly central to the probability and time-scale of extinction for
192 many species, it is a very poorly defined concept hampering measurement and mapping efforts. The
193 broadest definition of conservation capacity in relation to species extinctions includes at least three
194 dimensions (Table 1): i) willingness and motivation to act; ii) knowledge to design effective
195 interventions; and iii) institutional, technical, and economic resources to implement effective
196 interventions. The interactions of these three dimensions will largely determine whether a species is
197 identified as being at risk of extinction, whether efforts are made to reduce the risk of extinction, and
198 whether those efforts are successful in the short and long term.

199 **Table 1:** Main dimensions of conservation capacity and some of the methods of measurement (see
 200 text for details).

Dimension	Definition	Example Methods	Example
Willingness and Motivation to act	Societies vary in their desire to prevent different species from going extinct, largely depending on the societies' cultural values and on the species' cultural characteristics (public awareness, interest, sentiment, etc.).	- Culturomic analysis	- (Millard et al., 2020)
		- Social surveys	- (Samojlik et al., 2023)
Knowledge	The amount of scientific and/or local knowledge of species that is relevant to their conservation, management, and stewardship varies due to a wide range of cultural and historic factors.	- Bibliometrics	- (dos Santos et al., 2020)
		- Expert assessment	- (Pearce-Higgins et al., 2017)
		- Reports from Indigenous and local knowledge systems	- (Ziembicki et al., 2013)
Institutional, technical and economic resources	The capacity of local actors and conservation organizations (NGOs, governmental bodies, international institutions and private organizations) to fund and implement successful conservation interventions varies geographically in relation to complex socio-economic, political and historic factors.	- Desk-based analysis of institutional capacity	- (Malhado et al., 2020)
		- Social surveys	- (Fu, Shumate, 2020)

201

202 *Cultural willingness and motivation to prevent extinction*

203 It is self-evident that societies vary enormously in their willingness to allocate resources to conserve
 204 different species depending upon their values and valuation of nature (Díaz et al., 2015). Among the
 205 conservation community, charismatic and culturally iconic species of vertebrates (mainly mammals
 206 and birds) are prioritized for conservation funding and action over equally threatened but culturally
 207 less visible species (Mammola et al., 2020, Davies et al., 2018). A recent analysis of the internet
 208 salience of 36,873 vertebrate taxa revealed that search interest was higher for more threatened
 209 mammal and bird species than it was for fish, reptiles and amphibians (Davies et al., 2018). Similarly,
 210 Kim et al (2014) examined web search data for 246 threatened species in Korea and found that the
 211 interest for mammals, birds, amphibians and reptiles were ten times higher than those for other taxa.
 212 This bias towards vertebrates also has a geographical component with temperate species receiving
 213 more conservation attention than those in the tropics (Titley et al., 2017). Although plants generally
 214 receive less conservation attention than vertebrates, they are also strongly influenced by cultural
 215 perceptions (Adamo et al., 2021) with a recent study indicating that species with attractive flowers
 216 received more funding, irrespective of extinction risk (Adamo et al., 2022). Similarly, fungi
 217 conservation is significantly biased towards macrofungi since these are most easily observed and
 218 include many edible taxa (Gonçalves et al., 2021). It should be noted that even charismatic vertebrate
 219 species may still be lacking adequate resources to prevent continued population decline (Courchamp
 220 et al., 2018, Di Minin et al., 2015).

221 Culturally prominent species that generate high public interest and positive sentiment are more likely
 222 to be the target of conservation actions for two main reasons. Firstly, it is easier to mobilize support
 223 and resources through campaigns and other fund-raising actions when a species already has a high
 224 public profile (Thomas-Walters & Raihani, 2017). Secondly, societal preferences also extend into

225 scientific research, with researchers across the world preferentially collecting data on larger, more
226 charismatic taxa independent of the threat status (Troutet et al., 2017, Caro, 2007). Conversely, many
227 species receive little to no attention and are likely to suffer from a process of societal extinction – the
228 decline of collective attention and memory of an extinct or threatened, extant species (Jarić et al.,
229 2022). The process of societal extinction of species is linked to that of biological extinctions, as it is
230 likely to result in decreased support for conservation action, ultimately affecting negatively the
231 outcome of such efforts.

232 Until recently it was challenging to quantify the level of public awareness, interest and sentiment
233 about threatened species at national, regional and global scale because this required the use of time-
234 consuming and costly social surveys. However, the recent development of ‘conservation culturomics’
235 (Correia et al., 2021, Ladle et al., 2016, Di Minin et al., 2015), the analysis of digital data generated by
236 people to provide novel insights on human–nature interactions allows the evaluation of multiple
237 aspects of societal preferences for species and higher taxa. For example, Ladle et al. (2019) found that
238 the salience of bird species on the global internet was strongly correlated with species that have wide
239 geographic ranges that overlap with technologically advanced societies, that are phenotypically
240 conspicuous and, critically, that have direct interactions with humans (e.g., hunting, pet keeping, etc.).
241 A related study based on Wikipedia page views for all extant species of birds found that farmed species
242 and species in the pet bird trade were particularly prominent over multiple language editions
243 (Mittermeier et al., 2021). These examples demonstrate how culturomic metrics can be used to
244 capture and quantify different aspects of human interest in nature at scales that are beyond the reach
245 of standard social surveys (see also Fink et al., 2020, Johnson et al., 2023, Falk, Hagsten, 2022).

246 Big data approaches such as culturomics have enormous potential but also many limitations related
247 to scale and coverage (reviewed in Correia et al., 2021, Di Minin et al., 2021). For example, many
248 Indigenous Peoples (and many other socially and economically marginalized groups) have limited
249 access to the global internet and understanding their interactions, attitudes and sentiment towards
250 local species are also critical for effective conservation, but frequently ignored in conservation
251 management (Zanotti, Knowles, 2020). Recognizing Indigenous Peoples and local communities’ rights
252 and agency in conservation management (Reyes-García et al., 2022) is of critical importance, both
253 because much of the world’s biodiversity now exists in landscapes and seascapes traditionally owned,
254 managed, used and or occupied by Indigenous Peoples (Garnett et al., 2018) or by local communities
255 (Brondizio, Tourneau, 2016) and because such a strategy might ultimately improve conservation
256 outcomes (Büscher, Fletcher, 2019). Rates of biodiversity decline are slower in such areas than
257 elsewhere, including protected areas (Garnett et al., 2018, Fa et al., 2020, O’Byrne et al., 2021). Reyes-
258 García et al. (2023) recently developed a framework around the concept of culturally important
259 species that could be used to integrate different nature values into the management of threatened
260 species. Such species predominated among areas where Indigenous Peoples live and, critically, include
261 a high proportion of species that the IUCN classify as Data Deficient. Species in their study were more
262 likely to be culturally than biologically threatened, especially those associated with Indigenous Peoples
263 due to the high levels of cultural loss they have experienced.

264 *Conservation-relevant knowledge of species*

265 It has long been recognized that there are large and persistent taxonomic biases in which species have
266 been researched and, consequently, the volume and quality of scientific knowledge about different
267 species (Clark, May, 2002, Fleming, Bateman, 2016). Such variations potentially have a significant
268 impact on the capacity of societies to prevent species from going extinct. For many species, to be
269 ‘saved’ from extinction there should be sufficient biological, ecological, and cultural knowledge of the
270 species and its habitat to support the design and implementation of appropriate conservation

271 interventions (Murray et al., 2015, Cooke et al., 2017). For other species, habitat and site-based
272 conservation may be sufficient to prevent extinction and population decline. Moreover, more
273 knowledge of a species or higher taxon does not always lead to better conservation interventions or
274 swifter responses when a species is threatened but, all things being equal, a well-studied species or
275 group is more likely to be the subject of effective conservation actions than a poorly known
276 counterpart. It should be noted that it is not only published scientific knowledge that is potentially
277 important, but also the practical and contextual knowledge of researchers (and other stakeholders)
278 about the species in question. The greater the research effort, the greater number of people with such
279 knowledge that can be mobilized to facilitate conservation efforts.

280 The causes of these biases are relatively well understood (Jarić et al., 2019). For example, scientists
281 tend to study species within the country where they work due to a combination of funding priorities,
282 cost, and convenience. It follows that countries with low scientific capacity typically have fewer
283 qualified conservation scientists and ecologists and less resources available for research leading to
284 geographical biases in conservation research effort (Meyer et al., 2015). Species also vary in their
285 ‘researchability’ – any characteristic of the species that potentially increases the costs of data
286 collection or which impedes or reduces the feasibility of a research project (dos Santos et al., 2020, da
287 Silva et al., 2020). For field-based conservation research, this could include characteristics that make
288 it harder to observe a species, such as small body size, habitat characteristics and accessibility,
289 nocturnal activity patterns, elusiveness, or cryptic coloration. Researchability could also be correlated
290 with ‘conservability’, if the traits that make a species more challenging to study overlap with those
291 that make it more difficult to implement conservation interventions. Moreover, when species are
292 challenging to study they become less desirable targets for researchers whose chances of career
293 advancement may depend on their publication record or the completion of a high level research
294 dissertation (Caro, 2007).

295 As with human interest in species (see above), the last decade has seen great advances in our capacity
296 to quantify taxonomic biases in research at scale through the analysis of bibliometric databases such
297 as Scopus, Web of Knowledge, or Google Scholar. For example, a recent regional-scale bibliometric
298 analysis of Australian birds demonstrated significantly more publications on species with larger body
299 sizes, larger ranges, higher relative abundance, and which are present in urban environments
300 (Yarwood et al., 2019). A similar study on all extant species of mammals found that research volume
301 was strongly associated with the scientific capacity within the range of species, high body mass and
302 whether the species was non-native, with a very weak effect of conservation threat status (dos Santos
303 et al., 2020).

304 Additionally, it should be noted that there can be a mismatch between the “researchability” of a
305 species, as determined by scientists, and its cultural relevance, as defined by local criteria (Crane et
306 al., 2016). Reyes-García et al. (2023) found that culturally important species had a much higher
307 proportion of Data-Deficient species than the full set of IUCN species, most likely resulting in an
308 underestimation of their biological threat, as species categorized as Data-Deficient by the IUCN seem
309 to be more threatened than data-sufficient species (Borgelt et al., 2022). The data gap underscores
310 that cultural considerations remain disregarded in much current biological research (Bridgewater,
311 Rotherham, 2019) despite the fact that Indigenous and local knowledge has long been deemed as
312 essential to setting realistic and effective biodiversity targets (Reyes-García et al., 2022, Brondízio et
313 al., 2021, Berkes et al., 2000).

314 *Institutional, technical and economic capacity to intervene*

315 Even when there is good scientific knowledge about a threatened species and strong public support
 316 for conservation action, weak institutional capacity means that interventions may be poorly planned
 317 and executed or not even be implemented. In this context, institutional capacity normally refers to
 318 governmental departments, conservation NGOs and other civil society groups, and occasionally
 319 private sector organizations. Measuring such capacity is highly challenging and there have been very
 320 few systematic analyses of conservation organizations at national, regional or global scales
 321 (Brockington et al., 2018). To our knowledge, there have not yet been attempts to evaluate
 322 institutional conservation capacity at the level of species or geographic regions (e.g., countries),
 323 though such quantifications could play a major role in determining the number, type and quality of
 324 interventions in the face of endangerment (Ladle, Jepson, 2008). Such a lack is partly attributable to
 325 the difficulties of collecting data on diverse conservation actors (Malhado et al., 2020), and partly due
 326 to the complexity of factors that contribute to institutional capacity, severely limiting the potential to
 327 develop robust metrics.

328 Over the last decades, there has been an institutionalization of co-management and bottom-up
 329 approaches to conservation (e.g., Indigenous and Community Conserved Areas, community
 330 monitoring). For example, community-based monitoring is increasingly proposed to improve scientific
 331 understanding of biodiversity status and trends, or local uses of plants and animals, among other
 332 processes (Danielsen et al., 2021). Understanding the role of such initiatives in curbing extinction
 333 processes also requires monitoring.

334 *How many species have been 'saved' from extinction?*

335 Estimating the number of species that would have gone extinct if conservationists had not intervened
 336 is, by definition, exceedingly challenging and the fact remains that the number of species currently
 337 threatened with extinction is unprecedented in human history (Ceballos et al., 2010, Díaz et al., 2019).
 338 Well known examples of 'near extinction' events, such as the black footed ferret (Dobson, Lyles, 2000)
 339 or the Chatham Island Black Robin (von Seth et al., 2022) are indisputable. However, the imminent
 340 demise of the rescued population is often less clear cut and there have been few large-scale
 341 estimations, mainly restricted to birds and mammals (Table 2). For example, Bolam et al. (2021)
 342 estimated the number of species 'saved' by canvassing the opinion of experts; their estimate that bird
 343 and mammal extinction rates would have been 2.9–4.2 times greater without conservation action is
 344 almost certainly an underestimate given that only clear cases were considered. There are many more
 345 situations where, had conservation not intervened earlier in the extinction process (i.e., before a
 346 species becomes Critically Endangered), a species would arguably have gone extinct. Moreover, many
 347 species have avoided extinction due to actions aimed at conserving sites, habitats and ecosystems.
 348 This category of 'saved' species is even more difficult to quantify since they include many lesser known
 349 taxa, some of which may be undescribed (Hortal et al., 2015).

350 **Table 2:** Estimated number of bird and mammal species that would have gone extinct without direct
 351 human intervention.

Taxon	Number of species	Timeframe	Reference
Birds	16	1994-2004	(Butchart et al., 2006)
	21 to 32	1993-2020	(Bolam et al., 2021)
	9 to 18	2010-2020	(Bolam et al., 2021)

	7 to 16	1993-2020	(Bolam et al., 2021)
Mammals (all)	2 to 7	2010-2020	(Bolam et al., 2021)
Mammals (Ungulates)	6	1996-2008	(Hoffmann et al., 2015)

352

353 A closely related issue to species saved from extinction is the number of species that would eventually
 354 go extinct without continued conservation investment. Such ‘conservation dependent’ species could
 355 justifiably include, among others, species whose: i) populations are periodically augmented with
 356 captive bred individuals; ii) populations are not declining due to continued management efforts such
 357 as anti-poaching measures and surveillance, genetic management, control of invasive species,
 358 supplemental feeding, etc.; and iii) last remaining individuals exist only in captivity. Taking the latter
 359 group of species as an example, there are 85 species currently classified as Extinct in the Wild (i.e.,
 360 only *ex-situ* populations remain) and some of these species have persisted in captivity for over 70
 361 years (Smith et al., 2023). Other forms of conservation dependence are poorly quantified but
 362 potentially represent a significant limitation to future conservation actions given the limited resources
 363 available for new initiatives.

364

365 **Extinction risk forecasting using push and push-back factors**

366 As quantitative assessments of extinction risk, Red Lists are a crucial knowledge product and underpin
 367 much conservation law and policy (Hoffmann et al., 2008; Rodrigues et al., 2006). Red List
 368 categorizations are used, among other things, to: i) support conservation decisions at, and across,
 369 multiple governance levels; ii) guide strategy and investments in species conservation; and iii) inform
 370 progress towards targets of international agreements. Perhaps more importantly, Red Lists have
 371 translated the key conservation value of avoiding anthropogenic extinctions into a governance tool
 372 that has helped produce global norms governing relations between society, economy, and the non-
 373 human world (Jepson et al., 2011).

374 IUCN Red Lists assign species to extinction threat categories based on five quantitative criteria: i)
 375 population vulnerability; ii) population size reduction; iii) geographic range; iv) population size; and v)
 376 population viability analysis. These categories have a population ecology/life history focus, yet - as
 377 argued above - extinction (and its avoidance) is a biocultural phenomenon. Before and after a taxon
 378 is assigned to a Red List category it is subject to cultural forces that determine the success (or
 379 otherwise) of conservation actions (Ladle, Jepson, 2008). We would argue that a species well-known
 380 to science is, *ceteris paribus*, less likely to be at risk of extinction compared with a lesser-known species
 381 in the same threat category because publics and institutions will mobilise more effectively to save it.
 382 Exceptions may include species that are highly sought after as pets, trophies, food, or fashion
 383 accessories (Leclerc et al., 2015, Palazy et al., 2012b, Gault et al., 2008) that may suffer more intense
 384 exploitation than less desirable species (Courchamp et al., 2006). Moreover, IUCN species lists do not
 385 explicitly include the importance of species for local cultures (Reyes-García et al., 2023), a factor that
 386 could also play a vital role in the success of any proposed conservation intervention. In short, IUCN
 387 Red List categories currently omit a range of non-biological factors that may be critical in determining
 388 whether a species will be ‘saved’ from extinction (or not).

389 Creating a systematic and comprehensive system of assessing public interest and local cultural
390 importance of species could, along with information of scientific knowledge of species, provide
391 interesting complementary information to support and add nuance to extinction categorizations
392 (Figure 1). Specifically, information from Indigenous and local knowledge systems or from macroscale
393 cultural analysis (e.g., culturomics) could potentially be used to: i) identify threatened species
394 assemblages and geographic regions (or parts of regions) where potential for rallying support may be
395 weaker and where greater investment may be required to improve the conservation status of a taxon
396 (Ladle et al., 2016); ii) further support the use of Red Lists in business and investment decisions
397 (Bennun et al., 2018), making material the reputational risks associated with activities that impact
398 publicly visible globally threatened species; iii) enhance the quality of conservation actions by
399 increasing the recognition of Indigenous Peoples and local communities' knowledge, values, and rights
400 (Reyes-García et al., 2022); and iv) provide complementary information to support and prompt
401 innovative actions to reduce extinction risk. For example, "digital interventions" can raise public
402 profile of threatened species and more effectively link communities of interest with specific taxa.

403 The above suggestions come with the caveat that cultural evaluation based on big data approaches
404 such as culturomics have many limitations and biases (Correia et al., 2021, Troumbis, Iosifidis, 2020),
405 and much research is still needed to develop robust, well validated metrics that can be used with
406 confidence for conservation planning and assessment. Furthermore, as cultural metrics are eventually
407 integrated into extinction risk assessment it is inevitable that different threatened species might
408 benefit or lose out depending on how conservation organizations choose to use this information. For
409 example, deciding on the balance between funding the conservation of well-known species versus
410 promoting the conservation of species deemed to have little or no cultural importance.

411 All things being equal, in areas where conservation capacity is low: i) species are less likely to be 'saved'
412 from extinction due to a lack of scientific knowledge, resources, and effective conservation
413 interventions; ii) threatened species may be less effectively monitored (Fisher et al., 2011) leading to
414 incomplete knowledge of species distributions/population status and slow or absent conservation
415 responses; and iii) technological interventions such as captive breeding, reintroductions and
416 translocations are less likely to be implemented or successful. However, willingness and capacity of
417 institutions to act to prevent a species from imminent extinction is not straightforward to evaluate
418 and partially depends on the cultural characteristics of the threatened species, with far less effort
419 expended on the conservation of non-charismatic species. For example, Bellon (2019) found that
420 species popularity (higher internet salience) had a greater effect than federal priority ranking for the
421 funding of threatened species by various US federal agencies under the Endangered Species Act.

422

423 **Conclusions**

424 Species extinction is a complex phenomenon that involves both biological and social factors.
425 Understanding and addressing these factors is crucial for the conservation of biodiversity. Most, if not
426 all, species currently in danger of imminent extinction are in that state due to the direct and/or indirect
427 impacts of humans on the environment. Moreover, whether these species actually go extinct will
428 largely depend on the willingness to act and the technical capacity of local, national, and international
429 conservation organizations, along with the support of local communities and other stakeholders. In
430 contrast to the assessment of biological and ecological risk factors, our understanding and
431 quantification of the cultural and political vulnerability of species is at an early stage of development.
432 A more comprehensive understanding of which species will go extinct and which will be 'rescued' by
433 conservation and stewardship efforts will require an explicit interdisciplinary, biocultural approach to

434 extinction that draws on expertise from the social sciences, and dialogue with holders of other
435 knowledge systems, and in particular with Indigenous Peoples and local communities. For many
436 currently threatened species, extinction will only occur if society allows it to occur, either through a
437 lack of motivation, knowledge, resources, or local conservation capacity.

438

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451 There is no conflict interest.

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