


Trade-offs in carbon-degrading enzyme activities limit long-term soil carbon sequestration with biochar addition

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ABSTRACT

Biochar amendment is one of the most promising agricultural approaches to tackle climate change by enhancing soil carbon (C) sequestration. Microbial-mediated decomposition processes are fundamental for the fate and persistence of sequestered C in soil, but the underlying mechanisms are uncertain. Here, we synthesise 923 observations regarding the effects of biochar addition (over periods ranging from several weeks to several years) on soil C-degrading enzyme activities from 130 articles across five continents worldwide. Our results showed that biochar addition increased soil ligninase activity targeting complex phenolic macromolecules by 7.1%, but suppressed cellulase activity degrading simpler polysaccharides by 8.3%. These shifts in enzyme activities explained the most variation of changes in soil C sequestration across a wide range of climatic, edaphic and experimental conditions, with biochar-induced shift in ligninase:cellulase ratio correlating negatively with soil C sequestration. Specifically, short-term (<1 year) biochar addition significantly reduced cellulase activity by 4.6% and enhanced soil organic C sequestration by 87.5%, whereas no significant responses were observed for ligninase activity and ligninase:cellulase ratio. However, long-term (≥1 year) biochar addition significantly enhanced ligninase activity by 5.2% and ligninase:cellulase ratio by 36.1%, leading to a smaller increase in soil organic C sequestration (25.1%). These results suggest that shifts in enzyme activities increased ligninase:cellulase ratio with time after biochar addition, limiting long-term soil C sequestration with biochar addition. Our work provides novel evidence to explain the diminished soil C sequestration with long-term biochar addition and suggests that earlier studies may have overestimated soil C sequestration with biochar addition by failing to consider the physiological acclimation of soil microorganisms over time.

Key words: biochar addition, enzyme activity, soil carbon sequestration, experimental duration, soil microorganism, meta-analysis.

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I. INTRODUCTION

Biochar amendment of carbon (C)-rich products from biomass pyrolysis has increasingly been regarded as a cost-effective and environmentally friendly method of increasing soil C sequestration (Lehmann *et al.*, 2021; Molina *et al.*, 2009; Moore, Jevrejeva & Grinsted, 2010; Yang *et al.*, 2021). The amount of biochar amendment has increased substantially in recent decades, with predicted sequestration of 0.3–2.0 Pg CO₂ annually by 2050 (Fawzy *et al.*, 2021; Sohi *et al.*, 2010; Woolf *et al.*, 2010). Despite increasing evidence demonstrating the advantages of biochar in enhancing soil C sequestration (Han *et al.*, 2022; Hernandez-Soriano *et al.*, 2016; Zhang, Voroney & Price, 2015a), whether and how biochar addition affects soil C dynamics over time remains unclear. Indeed, several recent studies have shown that effects of biochar addition on soil C sequestration can be positive (Ameloot *et al.*, 2014; Azlan Halmi *et al.*, 2018), negative (Peng *et al.*, 2019; Tian *et al.*, 2016) or neutral (Elzobair *et al.*, 2016; Rafael *et al.*, 2019). Such large discrepancies illustrate a poor understanding of the underlying mechanisms. Positive effects of biochar addition on soil C sequestration can be explained by the stimulation of plant growth (Lehmann *et al.*, 2021; Liu *et al.*, 2016a) with a subsequent increase in inputs of plant (e.g. litter and root) residues into soil (Hagemann *et al.*, 2017). On the other hand, biochar addition can accelerate decomposition of pre-existing soil organic C (SOC) by changing microbial community composition and activities, leading to negative effects on SOC (Pei *et al.*, 2021; Tian *et al.*, 2016). However, a mechanistic understanding of the composite effects of biochar addition on SOC decomposition process is lacking, hampering the prediction of the long-term effects of biochar addition on soil C dynamics.

Soil extracellular enzymes catalyse the degradation of soil organic matter, deconstructing plant and microbial residues by breaking down large macromolecules into simpler molecules (Margida, Lashermes & Moorhead, 2020; Sinsabaugh, 2010).

Various extracellular enzymes target different pools of SOC, for example, ligninases target structurally complex polyphenolic macromolecules, and cellulases degrade ordered polysaccharides with a simpler structure (Chen *et al.*, 2018a; Margida *et al.*, 2020; Ren *et al.*, 2017; Yang *et al.*, 2022b). Biochar addition may have different impacts on ligninase and cellulase activity, partly due to changes in the chemical composition of soil organic matter and also due to shifts in microbial community composition (Gul *et al.*, 2015; Jing *et al.*, 2022; Singh & Cowie, 2014). For instance, the condensation of cellulose and hemicellulose into humic-like macromolecules on the surface of biochar (Lehmann *et al.*, 2021; Quilliam *et al.*, 2013) could lead to induction of microbial secretion of ligninase relative to cellulase, as enzyme production is commonly induced by the presence of suitable substrates (German *et al.*, 2012; Sinsabaugh *et al.*, 2008). The observed increases in the proportion of structurally complex macromolecules and fungal abundance with time after biochar addition may also lead to increased induction of ligninase relative to cellulase activity over time (Pei *et al.*, 2021; Yi *et al.*, 2020). Although some recent studies indicated that soil C sequestration varied significantly with time after biochar addition, biochar production technologies (e.g. feedstock type and thermal pyrolysis temperature of biochar), and site-specific conditions (e.g. climate and soil properties) (Gronwald *et al.*, 2015; Mitchell *et al.*, 2016), a comprehensive understanding of the underlying mechanisms remains unexplored. In particular, there is no direct evidence for how biochar addition affects key enzyme activities (e.g. cellulase and ligninase) that are likely to influence long-term impacts on soil C sequestration across various environmental conditions.

To address these knowledge gaps, we conducted a global meta-analysis to evaluate the responses of soil cellulase and ligninase activities to biochar addition, and how these responses may affect long-term soil C sequestration. We compiled a database of 923 soil C-degrading enzyme activity observations from 130 biochar addition studies (with biochar

addition periods ranging from several weeks to several years) across five continents worldwide (see online Supporting Information, Fig. S1, Table S1). We combined the advantages of classic meta-analysis with advanced model selection analysis to quantify the relative importance of potential predictors in explaining the effects of biochar addition on enzyme activities and soil C sequestration. This approach allows us to assess the role of enzyme activities in determining changes in soil C sequestration with biochar addition across a wide range of climatic, edaphic and experimental factors. Specifically, we test the following hypotheses: (i) biochar addition induces shifts in C-degrading enzyme activity, enhancing ligninase activity while suppressing cellulase activity; and (ii) ligninase:cellulase ratio may increase with time after biochar addition, leading to weakened soil C sequestration by stimulating the decomposition of structurally complex macromolecules with long-term biochar addition.

II. MATERIALS AND METHODS

(1) Enzymes included in this study

Seven kinds of extracellular enzymes involved in SOC decomposition were included in this meta-analysis (Table S2) following previous studies (Chen *et al.*, 2018a; Margida *et al.*, 2020; Ren *et al.*, 2017; Yang *et al.*, 2022b); four cellulases [β -1,4-glucosidase (BG), α -1,4-glucosidase (AG), β -1,4-xylosidase (BX) and β -D-cellobiosidase (CBH)] and three ligninases [phenol oxidase (PO), polyphenol oxidase (PPO) and peroxidase (PER)].

(2) Data collection

We searched for articles on the effects of biochar addition on cellulase and ligninase activities using *Web of Science* (<http://apps.webofknowledge.com/>), *Google Scholar* (<http://scholar.google.com/>), and *China National Knowledge Infrastructure* (<http://www.cnki.net/>). Specifically, we searched for peer-reviewed articles, academic theses, and book chapters published in English or Chinese before May 2022. We used the following search string [(‘biochar addition’ OR ‘biochar amendment’) AND (‘cellulase’ OR ‘ligninase’ OR ‘glucosidase’ OR ‘xylosidase’ OR ‘cellobiosidase’ OR ‘peroxidase’ OR ‘phenol oxidase’) AND (‘soil’ OR ‘terrestrial’ OR ‘land’)] and their equivalents in Chinese. Additionally, we searched for articles through other sources, including manual searches of reference cited by or citing the articles identified by our search string.

Articles included in this study had to meet the following criteria: (i) climatic, vegetation and soil attributes were similar for the control and biochar addition treatments; (ii) biochar properties (biochar materials, biochar pH, biochar temperature, biochar C% and biochar N%) and application protocols (biochar application method, biochar application rate and duration) were clearly described; (iii) ecosystem types were reported; and (iv) standard deviation (SD) and sample size were reported or could be calculated from the data presented

in the publication. Measurements with different durations of biochar addition at the same site were considered as independent observations because one of our primary purposes was to explore the impacts of duration on enzyme activities and soil C dynamics. Measurements with insufficient information on study sites and from contaminated soil were excluded to eliminate any confounding effects of pollutants on soil enzyme activities (Campos *et al.*, 2020; Li *et al.*, 2020c). The PRISMA flowchart illustrating the processes for selection of the included articles is shown in Fig. S1. All data were selected and collected following PRISMA-EcoEvo guidelines (PRISMA-EcoEvo_WordChecklist) (O’Dea *et al.*, 2021). Based on these criteria, we obtained 923 observations of soil C-degrading enzyme activities from 130 independent publications worldwide (Fig. S2, Table S1; references included in the meta-analysis are identified with asterisks in the reference list). Data were extracted directly from the tables, main text, or appendices of the articles and theses, or digitised from figures using Getdata Graph Digitizer (version 2.26) (<http://www.getdata-graph-digitizer.com/download.php>).

We first extracted information on cellulase and ligninase activities. If one paper reported two or more kinds of cellulase or ligninase, the sum of these enzyme activities was calculated as the overall responses of cellulase and ligninase activities, respectively (Chen *et al.*, 2018a; Wu *et al.*, 2022). If multiple measurements over time were conducted, values from the last time were selected. Multiple methods have been used to measure enzyme activities based on assessments of substrate concentrations or products over time at certain temperatures (Deforest, 2009; Marx, Wood & Jarvis, 2001). Most studies measured soil cellulase activity using fluorimetric methods with fluorescent 4-methylumbelliferone substrates, and assessed soil ligninase activity by colorimetric methods using L-3,4-dihydroxy-phenylalanine as the substrate (Elzobair *et al.*, 2016; Li *et al.*, 2020c). Methods and incubation conditions often varied among studies. Moreover, soil enzyme activities may also vary significantly depending on sampling season and soil water content (i.e. dry or moist periods). However, in our analyses we only consider the logarithmic response ratio of enzymes in each individual study, in which experimental conditions such as the type and concentration of substrates, buffer pH, incubation temperature and sampling season, etc., were the same for each paired observation. Comparisons between different sites were carried out using the logarithmic response ratio rather than absolute values of soil enzymes. Therefore, differences in measurement methods and sampling seasons should only have a minor influence on our assessment of the effects of biochar addition on enzyme activities in this meta-analysis (Chen *et al.*, 2020, 2018a; Hedges, Gurevitch & Curtis, 1999).

To quantify drivers of biochar effects on enzyme activities and SOC sequestration, we further collected information on a wide range of environmental variables, including elevation (0–1746 m), latitude (–42.95° S to 55.37° N), longitude (–119.74° W to 147.10° E), mean annual precipitation (MAP, 27–2500 mm), mean annual temperature (MAT, –1.0–32.3 °C), and vegetation type (cropland, grassland,

forest, open area or wetland) for each site. Edaphic properties, including SOC, soil total nitrogen (N), soil C:N ratio, soil pH and soil texture (classified as sandy, loamy or clay following USDA Textural Soil Classification) were also recorded. Table S1 and our Supplementary Data set (available at <https://doi.org/10.6084/m9.figshare.21769979>) provides detailed values or ranges for these variables obtained from the 130 publications. For missing environmental and edaphic variables, we searched for relevant publications by the same research group at the same study sites or contacted the corresponding authors. Alternatively, missing data for climatic (MAT, MAP) and soil attributes (SOC, total N and soil texture) were obtained from the WorldClimate Database (<http://www.worldclim.org/>) and Soil grids database (<https://www.isric.org/explore/soilgrids>), respectively.

For biochar properties and application protocols, we recorded feedstock, pH, pyrolysis temperature, C and N content (%) as well as method (Field, Pot or Laboratory incubation), rate (%) and duration (year) of biochar application. The feedstock types for biochar production were classified into five groups: herb, manure, residue, wood and urban waste. The pyrolysis temperatures used to produce biochar were classified into three groups: low ($\leq 350^\circ\text{C}$), medium ($350\text{--}550^\circ\text{C}$) and high temperature ($> 550^\circ\text{C}$). We recorded microbial biomass, the abundance of fungi, bacteria, the fungi:bacteria ratio, and plant biomass for both ambient and biochar addition treatments, when these variables were reported.

In total, this data set included 12,194 records of the above environmental, edaphic, and experimental factors, or the responses of soil organic C, soil nutrient contents, microbial or plant attributes to biochar addition.

(3) Data analysis

We used meta-analysis to investigate the effects of biochar addition on soil cellulase and ligninase activities, ligninase:cellulase ratio and other edaphic and microbial variables. Specifically, we calculated the logarithmic response ratio (LnR) of each variable using the following equation (Chen *et al.*, 2017b; Hedges *et al.*, 1999):

$$\text{LnR} = \text{Ln} \left(\frac{\overline{X^B}}{\overline{X^C}} \right) = \text{Ln}(\overline{X^B}) - \text{Ln}(\overline{X^C}) \quad (1)$$

where $\overline{X^B}$ and $\overline{X^C}$ are the arithmetic average values in the biochar addition and control treatments, respectively. The variance (V_i) of LnR was calculated as:

$$V_i = \frac{\text{SD}_B^2}{n_B \overline{X}_B^2} + \frac{\text{SD}_C^2}{n_C \overline{X}_C^2} \dots \quad (2)$$

where SD_B and SD_C are the standard deviations, and n_B and n_C are the number of replicates, respectively.

The overall effects of biochar addition on different variables and 95% confidence intervals (CI) were evaluated using *rma.mv*

function in the *metafor* package of R project (version 4.0.2) (Viechtbauer, 2010). We included 'Publication' and 'Observation' as random factors in the mixed-effect models, because some studies contributed more than one paired observation (Chen *et al.*, 2018a; van Groenigen *et al.*, 2017). To facilitate the interpretation of data, the effect size was back-transformed to percentage change using the equation $(\text{LnR}-1) \times 100$ (Chen *et al.*, 2018a). The effect of biochar addition on each variable was considered significant if the 95% CI did not overlap with zero. The normality of data for each kind of enzyme activity was tested using the Kolmogorov–Smirnov and Shapiro–Wilk tests, except for PPO due to its small sample size. A funnel plot and Egger's test were used to detect potential publication bias on soil enzyme activities, ligninase:cellulase ratio and soil organic C using the *metafor* package of R project (version 4.0.2). A sensitivity analysis was conducted to investigate the stability of the results by excluding one study a time using the *leave1out* function in *metafor* (Copas & Shi, 2000).

We conducted a meta-analysis to analyse the combined effects of environmental, edaphic and experimental factors on the responses of soil cellulase activities, ligninase activities, ligninase:cellulase ratio and soil C sequestration to biochar addition. In brief, we used a mixed-effects meta-regression model using the *gmulti* package in R (Calcagno & de Mazancourt, 2010; Chen *et al.*, 2020, 2018b). The importance of different factors was evaluated using the sum of Akaike weights. The weight was considered as the overall support for each variable in all potential models. A cutoff of 0.8 was set to identify the significant predictors for each model (Chen *et al.*, 2018a, 2017b; Terrer *et al.*, 2016). We used Spearman's rank correlation analysis to evaluate the relationships of cellulase activities, ligninase activities, and ligninase:cellulase ratio with environmental, edaphic and experimental factors. To explore further the effect of experiment duration on soil C sequestration, we conducted both linear regression and piece-wise regression models to fit the relationship between the LnR of SOC and time after biochar addition. Specifically, a piece-wise regression model was carried out using the *segmented* R package (Muggeo, 2003). The optimal regression model was selected by comparing regression coefficients (R) and the model was statistically significant if $P < 0.05$. Furthermore, studies were separated into short-term and long-term according to the slope of the curves for the relationship between LnR-SOC and time after biochar addition. Spearman's rank correlation analysis was conducted to investigate factors associated with enzyme activities and soil C sequestration in short-term and long-term studies, respectively.

III. RESULTS

(1) Responses of cellulase and ligninase activity to biochar addition

Averaged across all studies, biochar addition significantly suppressed cellulase activity by 8.3% ($P < 0.001$; Fig. 1A).

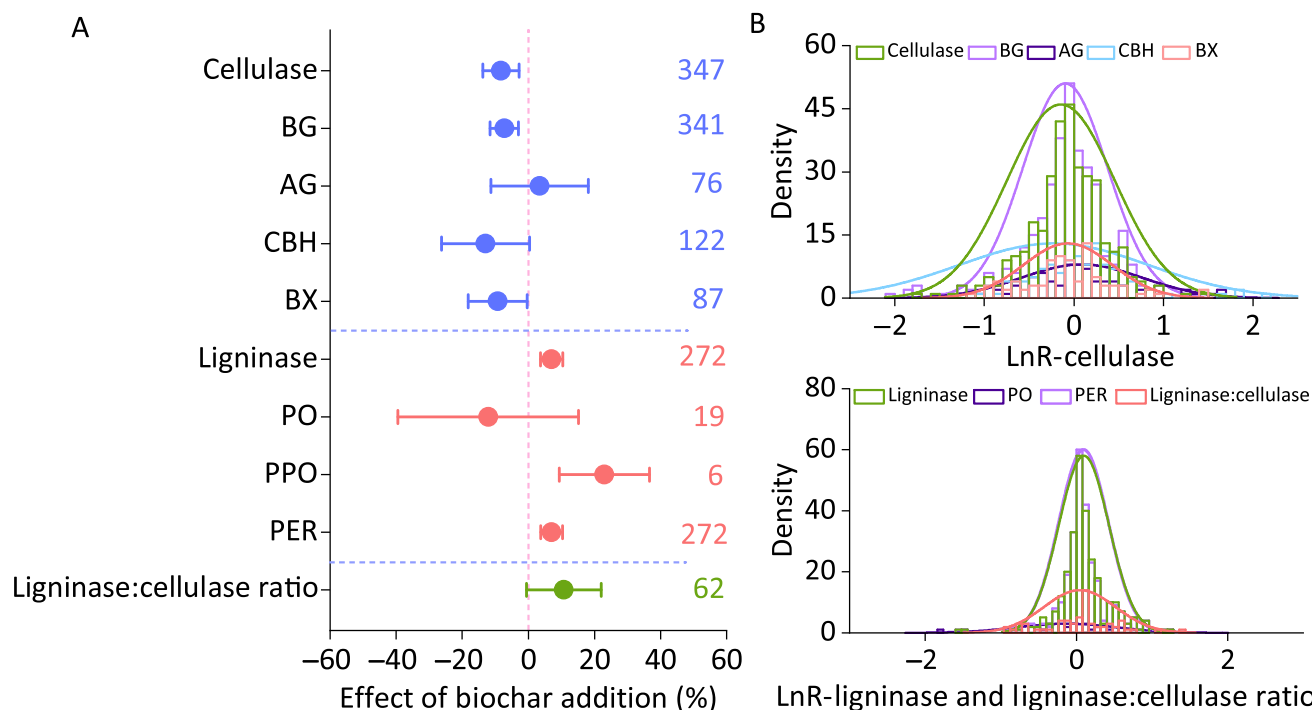


Fig. 1. Carbon-degrading enzyme activities. (A) Effects of biochar addition on activity of cellulases, ligninases, and on the ligninase:cellulase ratio. Values represent the mean percentage change in each variable with biochar addition *versus* control; error bars indicate 95% confidence intervals. Sample sizes for each variable are shown on the right. (B) Distribution of the log response ratios (LnR) of cellulase activity (top), and ligninase activity and the ligninase:cellulase ratio (bottom) to biochar addition. BG, β -1,4-glucosidase; AG, α -1,4-glucosidase; CBH, β -D-cellobiohydrolase; BX, β -1,4-xylosidase; PO, phenol oxidase; PPO, polyphenol oxidase; PER, peroxidase.

Specifically, biochar addition decreased the activities of BG and BX by 7.3% and 9.3% ($P < 0.05$), respectively. By contrast, biochar addition increased ligninase activity by 7.1% on average ($P < 0.001$), with an increase of PER activity by 7.0% ($P < 0.001$) and of PPO by 23%. The differential responses of cellulase and ligninase activities to biochar addition led to a marginally increased ratio of ligninase:cellulase activities by 10.7% ($P = 0.052$). In addition, the responses of cellulase, ligninase and ligninase:cellulase ratio were normally distributed according to both Kolmogorov–Smirnov and Shapiro–Wilk tests ($P > 0.05$; Fig. 1B). Moreover, Egger’s funnel plots appeared symmetrical ($P > 0.05$), and the results showed no significant publication bias in this meta-analysis (Fig. S3).

The effects of biochar addition on soil enzyme activities depended on the feedstock type and pyrolysis temperature for biochar production and variation in soil texture (Fig. S4). Herb, wood and urban waste biochar significantly reduced cellulase activity by 4.4–22.0% ($P < 0.05$), while manure and residue biochar enhanced ligninase activity by 2.5–12.7% ($P < 0.05$). Biochar produced at medium (350–550 °C) and high (>550 °C) temperatures significantly decreased the activity of cellulase by 10.0% and 6.7%, respectively ($P < 0.05$). By contrast, biochar produced at low (<350 °C) and medium temperature increased ligninase activity by 16.0% and 4.6%, respectively. There were significant reductions in cellulase activity in sandy (by 16.6%, $P < 0.05$) and clay (by 7.0%, $P < 0.05$) soils, and

significant increases in ligninase activity in sandy (by 13.2%, $P < 0.05$) and loamy (by 15.3%, $P < 0.05$) soils. These differential responses of cellulase and ligninase activities resulted in significant increases in ligninase:cellulase ratio by 7.8–14.5% ($P < 0.05$) with biochar produced using feedstock from manure or wood materials, at medium temperature and in sandy and loamy soils (Fig. S4).

Model selection analysis showed that the main factors influencing the response of cellulase and ligninase activities to biochar addition were different (Fig. 2). The effects of biochar addition on cellulase activity were best explained by biochar application rate, MAP, longitude and soil clay content (Fig. 2A). By contrast, the responses of ligninase activity to biochar addition were best explained by soil N content, biochar temperature, site location (i.e. longitude) and biochar pH. Linear regression analysis confirmed that LnR of cellulase activity was negatively correlated with biochar application rate, whereas a positive correlation was found with MAP ($P < 0.05$). Moreover, LnR of ligninase had negative relationships with soil N content and biochar pyrolysis temperature, but a positive correlation with biochar pH ($P < 0.05$; Fig. 2B). For the ligninase:cellulase ratio, the most important predictors were the time after biochar addition, soil C:N ratio, biochar C content, and biochar C:N ratio. Specifically, the LnR-ligninase:cellulase ratio was positively correlated with time after biochar addition but negatively correlated with soil C:N ratio, biochar C content and biochar C:N ratio ($P < 0.05$).

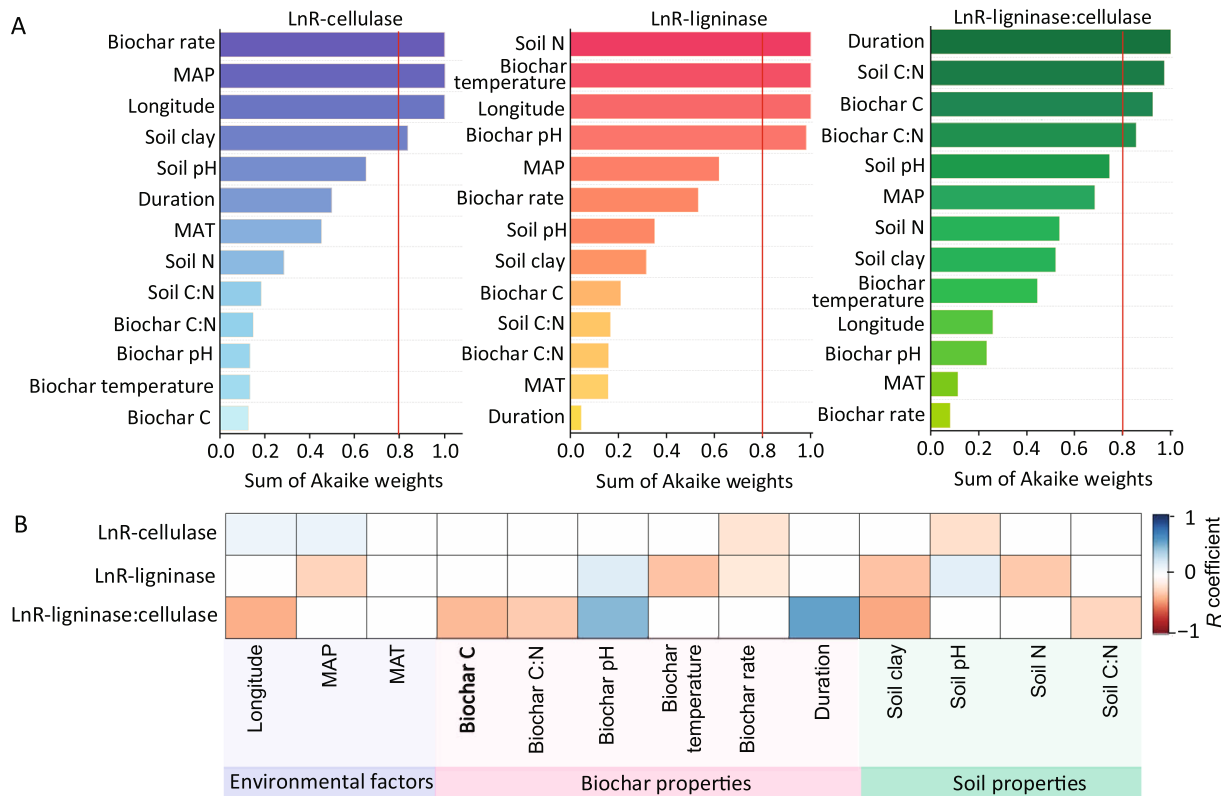


Fig. 2. Factors influencing the responses of soil enzymes. (A) Relative importance of different variables regulating the effects of biochar addition on cellulase activity, ligninase activity, and ligninase:cellulase ratio. Relative importance is shown according to the sum of Akaike weights of model selection. A cutoff of 0.8 was set to differentiate the important *versus* non-essential predictors. (B) Correlations between studied variables and the responses (log response ratio, LnR) of cellulase activity, ligninase activity, and ligninase:cellulase ratio to biochar addition. Biochar temperature, temperature of biochar production; Biochar rate, application rate of biochar addition; Duration, time after biochar addition; MAP, mean annual precipitation, MAT, mean annual temperature.

(2) Linking shifts in soil enzyme activities to changes in SOC with biochar addition

For studies that have reported SOC, biochar addition enhanced SOC by an average of 52.8% ($P < 0.001$) (Fig. 3A). Biochar-induced increases in SOC were found regardless of differences in feedstock, pyrolysis temperature of biochar addition and soil types (Fig. S4). Specifically, wood- and herb-derived biochar significantly increased SOC by over 60% ($P < 0.001$), and biochar produced from manure and residue enhanced SOC by 24.9 and 43.5%, respectively. Biochar produced at high pyrolysis temperature had a more positive effect on SOC (with SOC increased by 119.7%) compared with those produced at low and medium temperature (enhanced SOC by 47.6 and 54.8%, respectively). Moreover, SOC in loamy and clay soils (increased by 68.7 and 62.6%, respectively) showed a more positive response to biochar addition than in sandy soils (increased by 24.5%) (Fig. S4). Our piece-wise regression analysis showed that biochar-induced increases in SOC overall diminished with time after biochar addition (Fig. 3B; $P < 0.001$). The relationship between LnR-SOC and time of biochar addition could be divided into two periods (i.e. < 1 year and ≥ 1 year) according to the slope of the curves

identified using piece-wise regression. The model selection analysis showed that the response of SOC to biochar addition was best explained by LnR-ligninase:cellulase ratio (Fig. 3C). Specifically, changes in SOC with biochar application were negatively related to the ligninase:cellulase ratio (Fig. 3D, $P < 0.001$).

(3) The effect of time after biochar addition on soil enzyme activities and SOC sequestration

Studies were separated into short term (< 1 year) and long term (≥ 1 year) according to the results of piece-wise regression between LnR-SOC and time after biochar addition (Fig. 3B). Short-term and long-term biochar addition reduced the activity of cellulase by 4.6 and 12.7%, respectively ($P < 0.05$; Fig. 4A). By contrast, there were no significant effects of biochar addition on ligninase activity and ligninase:cellulase ratio in short-term studies ($P > 0.05$), while long-term biochar addition significantly enhanced soil ligninase activity by 5.2% ($P < 0.05$) and ligninase:cellulase ratio by 36.1% ($P < 0.001$). Short-term biochar addition had more positive effects on SOC compared with long-term studies, with significant increase in SOC by 87.5 and

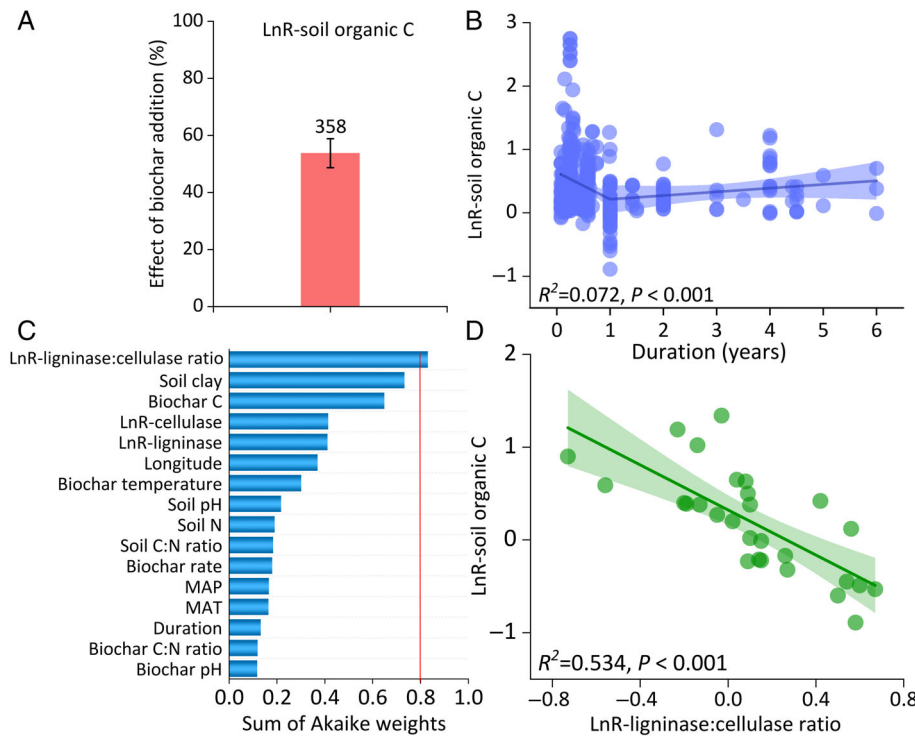


Fig. 3. The response of soil organic C to biochar addition and associated driving factors. (A) Effects of biochar addition on soil organic C sequestration. Value represents the mean percentage change with biochar addition *versus* control; error bar indicates 95% confidence intervals. Sample size is shown above the column. (B) Relationship between duration of biochar addition and the log response ratio (LnR) of soil organic C to biochar addition. The relationship between duration of biochar addition and LnR-soil organic C was analysed using a piece-wise regression model. (C) Relative importance of different variables regulating the effects of biochar addition on soil organic C sequestration. A cutoff of 0.8 is set to differentiate the important *versus* non-essential predictors. (D) Relationship between the response of ligninase:cellulase ratio and the response of soil organic C to biochar addition. Biochar temperature, temperature of biochar production; Biochar rate, application rate of biochar addition; Duration, time after biochar addition; MAP, mean annual precipitation; MAT, mean annual temperature.

25.1%, respectively. In short-term studies, the response of soil enzyme activities was significantly associated with changes in soil properties. Specifically, LnR-cellulase activity had negative relationship with soil clay content, but positive associations with soil pH and soil N content, whereas LnR-ligninase activity showed opposite relationships with these soil properties (Fig. 4B). LnR-SOC correlated negatively with soil pH and positively with biochar pyrolysis temperature. In long-term studies, the response of both soil enzyme activities and SOC were significantly related to soil clay content. LnR-cellulase activity and LnR-SOC showed positive relationships with soil clay content, whereas LnR-ligninase and LnR-ligninase:cellulase ratio correlated negatively with soil clay content.

IV. DISCUSSION

Our results indicate that shifts in C-degrading enzyme activities are key drivers of soil C sequestration with time (ranging from several weeks to several years) after biochar addition.

Most importantly, our model selection analysis underscores that changes in ligninase:cellulase ratio explains the most variation in the response of SOC to biochar addition, with biochar-induced shift in ligninase:cellulase ratio correlating negatively with SOC content (Fig. 3D). Specifically, the significantly increased ligninase:cellulase ratio under long-term (≥ 1 year) biochar addition leads to progressively diminished soil C sequestration. These results provide strong support for our hypothesis that the increasing ligninase:cellulase ratio with time contributes to a declining capacity for soil C sequestration with long-term biochar addition. To the best of our knowledge, this is the first comprehensive study linking enzyme activity and soil C with biochar addition, providing novel evidence to unravel the mechanisms controlling soil C sequestration with prolonged biochar exposure.

We propose several possible underlying mechanisms to explain differential responses of cellulase and ligninase activity with biochar addition (Fig. 5). First, biochar-induced reductions in soil N availability could stimulate ligninase rather than cellulase activity. Several lines of evidence have demonstrated reductions in soil N availability after biochar addition, which might be driven by (i) higher plant biomass

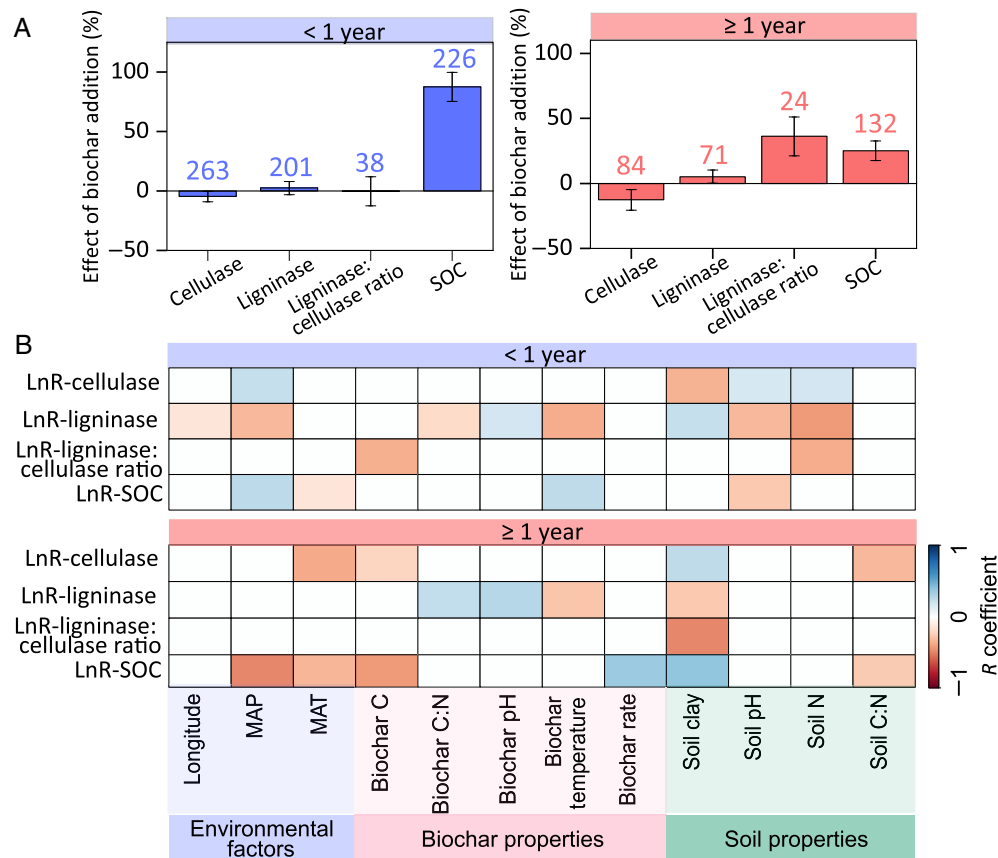


Fig. 4. Responses of soil enzymes and soil organic C to short-term and long-term biochar additions. (A) Effects of biochar addition on soil enzyme activities, ligninase:cellulase ratio, and soil organic C (SOC) sequestration in short-term (<1 year) and long-term (≥1 year) studies. Values represent the mean percentage change in each variable with biochar addition *versus* control; error bars indicate 95% confidence intervals. Sample sizes for each variable are shown above the column. (B) Relationships between different variables and the log response ratio (LnR) of enzyme activities, ligninase:cellulase ratio, and SOC sequestration. Biochar temperature, temperature of biochar production; Biochar rate, application rate of biochar addition; MAP, mean annual precipitation; MAT, mean annual temperature.

production (Fig. S5) and associated translocation of N from soil to vegetation, (ii) the additional C inputs increasing bulk soil stoichiometric C:N ratio (Fig. S6A), and (iii) occlusion of soil NH₄⁺ by phenolic- and lignin-like compounds through complex organo-mineral interactions on biochar surfaces (Fig. S6B). In response, soil microorganisms may increase ligninase production to stimulate the breakdown of complex phenolic- and lignin-like compounds to acquire bound N. In support of this explanation, we found that soil N content was the most important predictor (negative relationship) of the effects of biochar addition on soil ligninase activity (Fig. 4). This explanation supports the ‘microbial N mining theory’, which assumes that soil microorganisms will likely invest resources to decompose complex structural macromolecules to acquire N under N limitation (Craine, Morrow & Fierer, 2007; Meyer *et al.*, 2017; Moorhead & Sinsabaugh, 2006).

Second, shifts in microbial community composition could contribute to the opposite effects of biochar addition on cellulase and ligninase activity. Our results show positive

associations between ligninase activity and microbial biomass with biochar addition, but not for cellulase activity (Fig. S7A,B). These results suggest that soil microbial community composition or enzyme production efficiency may change under biochar addition. Indeed, previous studies have reported that biochar addition stimulates fungal growth including the two most commonly occurring types of mycorrhizal (arbuscular mycorrhizal and ectomycorrhizal) fungi (Lehmann *et al.*, 2011; Yang *et al.*, 2022a). This increase in fungal abundance with biochar addition could arise because fungal hyphae can grow into biochar pores and thereby access complex macromolecules adsorbed on biochar (Gul *et al.*, 2015; Lehmann *et al.*, 2011). Moreover, the protection of soil fungi from grazers or competitors on biochar pores may also contribute to increased soil fungal abundance (Li *et al.*, 2020b). Consistently, we found that increased fungal abundance associated with biochar addition was positively correlated with changes in ligninase activity (Figs S7 and S8). These results indicate that biochar-induced shifts towards a fungi-dominant microbial community could

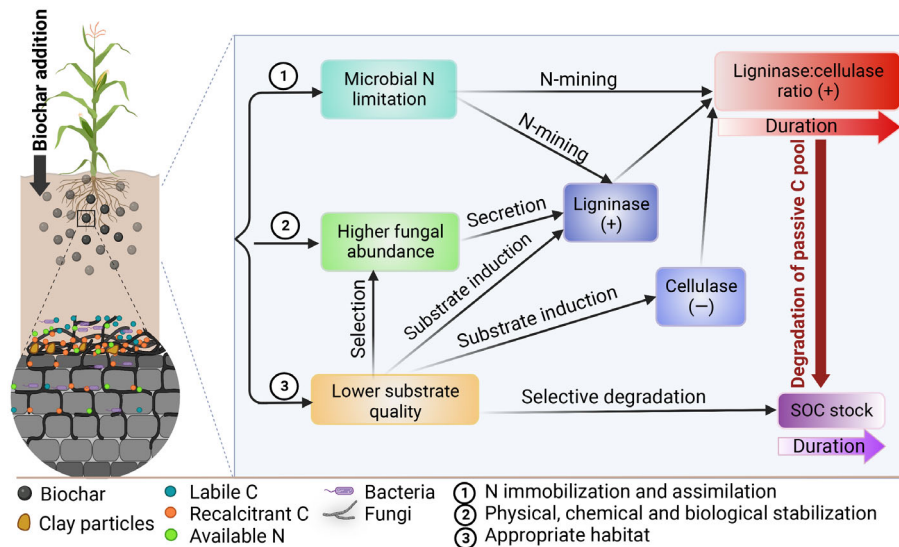


Fig. 5. Schematic of enzyme-mediated soil organic C sequestration. A conceptual paradigm illustrating the mechanisms of biochar addition on soil carbon-degrading enzyme activities and their impacts on soil organic C (SOC) dynamics. Biochar addition has differential effects on soil cellulase and ligninase activities *via* direct and indirect effects on substrate quality, microbial community composition and soil N availability. The differential responses of cellulase and ligninase activities to biochar addition results in an increasing ligninase:cellulase ratio with duration of biochar addition, which may reduce SOC sequestration over time.

promote ligninase activity, possibly because fungi are more efficient at mineralizing structural complex macromolecules than bacteria and are the primary producers of phenol oxidase (Burke & Cairney, 2002). In addition, we observed a positive relationship between LnR-ligninase:cellulase ratio and LnR-fungi:bacteria ratio (Fig. S7D), further suggesting that the observed shifts in extracellular enzyme activities were related to changes in microbial community composition.

Third, biochar addition significantly altered the chemical composition of soil organic matter (Mitchell *et al.*, 2016; Tian *et al.*, 2016), which likely contributed to the differential responses of cellulase and ligninase activities. By introducing additional phenolic- and lignin-like compounds, biochar could reduce the availability of readily decomposable C compounds because they could be occluded within macromolecule assemblages through complex organo-mineral interactions (e.g. H-bonding, cation bridging, and hydrophobic interactions) on biochar surfaces (Jing *et al.*, 2022; Kleber, Sollins & Sutton, 2007; Singh & Cowie, 2014). In particular, the resulting non-polar and hydrophobic compounds are composed predominantly of alkyl and aromatic functional groups (Hernandez-Soriano *et al.*, 2016; Kleber *et al.*, 2007). Microbial utilisation of readily decomposable C compounds thus could be suppressed due to limitation in substrate availability (Jing *et al.*, 2022; Singh & Cowie, 2014). By contrast, the increase in lignin-like soil organic C would possibly induce expressional and/or translational upregulation of ligninase activity. Therefore, our findings indicate that after the initial depletion and stabilisation of readily decomposable C by biochar, soil microbes likely stimulate ligninase production to access more chemically recalcitrant soil C pools (Li *et al.*, 2020b).

Biochar production technologies (e.g. feedstock type and thermal pyrolysis temperature of biochar) and soil properties (e.g. soil texture) may also be potential reasons underlying shifts in enzyme activities with biochar addition. For instance, biochar made from wood fibres commonly had higher degree of aromaticity and recalcitrance (Liu *et al.*, 2016a; Wang, Xiong & Kuzyakov, 2016), which could reduce cellulase activity and enhance ligninase:cellulase ratio by affecting the chemical composition of soil substrates (Fig. S4). By contrast, the addition of carbohydrate-rich biochar (e.g. biochar derived from crop residues) may alleviate substrate restriction on cellulase activity and contribute to a reduced ligninase:cellulase ratio (Fig. S4). Additionally, our model selection and correlation analysis indicates that the pyrolysis temperature of biochar had differential effects on the response of cellulase and ligninase activity to biochar addition. Biochar produced at high (>550°C) pyrolysis temperatures commonly has more C present as aromatic compounds (Mukherjee, Zimmerman & Harris, 2011; Wang *et al.*, 2015), which may explain its reduced cellulase activity, but lack of a significant effect on ligninase activity (Fig. S4). Moreover, the response of soil ligninase activities to biochar addition correlated negatively with soil clay content, because humics and complex aromatic compounds could be preferentially or competitively adsorbed by soil clay particles (Balcke *et al.*, 2002). Therefore, loamy and clay soils with higher soil clay content had higher retention capacity for SOC compared with sandy soils (Fig. S4).

Shifts in responses of cellulase and ligninase activity to biochar addition could exert inverse effects on soil C sequestration. Specifically, suppressed cellulase activity may promote soil C sequestration with biochar by reducing the

decomposition of ordered polysaccharides with simpler structure (Margida *et al.*, 2020). By contrast, enhanced ligninase activity may cause increased decomposition of complex phenolic macromolecules, which is commonly considered to be a rate-limiting step of SOC decomposition (Fontaine *et al.*, 2007; Schmidt *et al.*, 2011). For instance, the decomposition of phenolics such as tannins may result in reduced stabilisation of fungal necromass (Adamczyk *et al.*, 2019) and alleviate the toxicity and binding effects of phenols on hydrolase activities (Freeman *et al.*, 2004; Sinsabaugh, 2010), consequently limiting the effect of biochar addition on soil C sequestration. Therefore, biochar-induced sequestration of SOC could reflect the differential responses of these key extracellular enzymes. In support of this idea, our model selection analysis results indicate that shifts in soil enzyme activity from cellulase to ligninase explained the most variation in soil C sequestration with biochar addition (Fig. 3). Our regression analysis further showed a significant negative relationship between ligninase:cellulase ratio and SOC with biochar addition (Fig. 3D), suggesting that the increased decomposition of lignin-like substrates relative to cellulose-like substrates may limit soil C sequestration with biochar addition.

Moreover, the response of ligninase activity and ligninase:cellulase ratio to biochar addition varied between short-term (<1 year) and long-term (≥ 1 year) studies, contributing to the observed changes in the soil C sequestration capacity with time after biochar addition. In short-term studies, biochar addition significantly reduced cellulase activity, but had no significant effect on ligninase activity and ligninase:cellulase ratio, contributing to a significant 87.5% increase in SOC. However, the increased ligninase activity and ligninase:cellulase ratio observed in long-term studies may counteract the effects of cellulase activity on SOC sequestration and lead to weakened soil C sequestration (increased by only 25.1%) under biochar addition. These results indicate that some studies may have overestimated the long-term effects of biochar addition on soil C sequestration by failing to consider dynamics in activities of different enzymes with time after biochar addition (Woolf & Lehmann, 2012). Furthermore, the ligninase:cellulase ratio increases with time after biochar addition, which could exacerbate the decline in soil C over time. Previous studies have reported that complex-macromolecular C adsorbed on the surface of biochar could be used by the microbial community when polysaccharides with simple ordered structure are depleted over time (Gul *et al.*, 2015; Yi *et al.*, 2020). Therefore, this gradual increase in ligninase:cellulase ratio over time may also reflect changes in the chemical composition of soil organic matter and associated shifts in microbial community composition with time after biochar addition (Acosta-Martínez & Harmel, 2006; Pei *et al.*, 2021). Similar declines in soil C sequestration with time after biochar addition have also been observed in long-term case studies, and are mainly considered the result of declining adsorption capacity of biochar over time (Lefebvre *et al.*, 2020; Quilliam *et al.*, 2013). These results provide new evidence from enzyme activities that a functional acclimation of soil microorganisms to the chemical composition of organic substrates

affects the response of soil C sequestration to biochar addition over time.

However, the moderate correlations between soil C sequestration and ligninase:cellulase ratio suggest that shifts in soil enzyme activity alone cannot fully explain the variations in soil C sequestration with biochar addition (Fig. 3D). Indeed, overall soil C sequestration is determined by interactions among at least three C pools, namely biochar, the pre-existing SOC and plant litter/root exudates (Lehmann *et al.*, 2021). Therefore, other soil processes, such as the decomposition of labile components of biochar and priming of pre-existing C in soil could also affect soil C sequestration with biochar addition (Singh *et al.*, 2014; Zhang *et al.*, 2022). For example, a previous meta-analysis indicated that a small part of biochar ($\sim 3\%$) is bioavailable with a mean residue time of 108 days (Wang *et al.*, 2016), which may partly contribute to the initial decline in the response of soil C sequestration during the first year of biochar addition in this study. Further studies deciphering these processes (e.g. using isotopic tracers) are needed to predict accurately the long-term consequences of biochar addition to soil C sequestration. In addition, heterogeneity in experimental design (e.g. sampling season, properties of the original soil, type and concentration of substrate used for enzyme analysis, etc.) may affect our results and inferences. To investigate the sensitivity of our results to these heterogeneities, the influences of individual studies on the overall results were estimated by excluding one study a time. Our results showed that the response ratio of enzyme activities and SOC to biochar addition were relatively constant and without marked fluctuations (Table S3). This stability of results suggests that our findings do not merely reflect biases in the data set and provides support for our conclusions.

Models used to predict soil C sequestration with biochar addition vary significantly, with annual increases in SOC ranging from 0.07 to 10% per unit of biochar C added (Lehmann *et al.*, 2021; Woolf & Lehmann, 2012). These large uncertainties may stem from the timescales and soil C mineralisation processes simulated in different models. Existing biochar models commonly consider soil C mineralisation as simple first-order reactions (Woolf & Lehmann, 2012). However, C mineralisation is a complex process that combines enzyme-mediated catalysis of both fast- and slow-mineralised organic fractions (Chen *et al.*, 2020; Wu *et al.*, 2022). Our results suggest that retaining inflexible microbial functional traits over the duration of biochar addition may lead to inaccurate predictions of soil C sequestration. Therefore, it is necessary to include the temporal shifts in microbial C-degrading enzyme activity to improve model predictions of soil C sequestration over time with added biochar.

Our study provides evidence for the contribution of biochar addition to enzyme-catalysed microbial decomposition processes and soil C sequestration over wide temporal and environmental scales. Our results show that shifts in cellulase and ligninase activities drive long-term impacts of biochar addition on soil C sequestration. This physiological acclimation in microbial metabolic activity has been overlooked to

date (Hernandez-Soriano *et al.*, 2016; Jing *et al.*, 2022). In addition, our analyses offer insights to options for regulating soil C sequestration with biochar addition across a broad range of environmental and experimental conditions. Specifically, the responses of cellulase and ligninase depend on different environmental, edaphic and experimental factors (Fig. 2). Therefore, it should be possible to promote soil C accrual with biochar addition by regulating factors controlling different soil enzyme activities. For instance, biochar can be applied with N fertiliser to promote C sequestration by reducing N-mining and associated ligninase activity. Moreover, reduced response of soil ligninase activity and associated increase in soil C sequestration could also be achieved by selecting the appropriate pyrolysis temperature of biochar production. Therefore, there are promising potentials for innovative biochar management techniques for long-lasting climate mitigation, which will require the collective actions of policy makers, farmers and industry at both local and regional scales.

V. CONCLUSIONS

- (1) Our synthesis identifies differential responses of cellulase and ligninase to biochar addition, with important implications for long-term soil C sequestration.
- (2) Biochar addition increased ligninase activity but reduced cellulase activity, with an increasing ligninase:cellulase ratio with time after biochar addition.
- (3) Biochar-induced changes in ligninase:cellulase ratio were negatively related to SOC pool size, suggesting a progressive reduction in soil C sequestration with long-term biochar addition.
- (4) Various factors influenced the responses of cellulase and ligninase activities to biochar addition, providing insights into options for increasing soil C sequestration under prolonged biochar exposure.
- (5) We provide new evidence to explain the diminished soil C sequestration with long-term biochar addition, and highlight that the C sequestration potential of biochar may be overestimated without considering temporal changes in the physiological acclimation of soil microorganisms.

VI. ACKNOWLEDGEMENTS

We would like to thank all the authors whose data and work are included in this meta-analysis. This study was supported by the National Natural Science Foundation of China (32071595, 41830756 and 42177022). We also thank the Fundamental Research Funds for the Central Universities (Program no. 2662019QD055). We acknowledge Cunbin Gao, Qianqian Zhao and Qin Liu for their assistance in data collection. J.C. received funding from Aarhus Universitets Forskningsfond (AUFF-E-2019-7-1), EU H2020 Marie

Skłodowska-Curie Actions (839806), Danish Independent Research Foundation (1127-00015B), and Nordic Committee of Agriculture and Food Research (<https://nordicagriresearch.org/2020-5/>). The authors declare no competing interests.

VII. AUTHOR CONTRIBUTIONS

J. C. designed the meta-analysis. D. Y., J. F., Y. S. and X. L. conducted independent parallel screening of studies and collected data, J. F. analysed the data with help from J. C. and Y.-R. L. The manuscript was written by J. F., edited by J. C., Y.-R. L., R. L. S., D. L. M., M. N. A., P. S., and Q. H. All authors reviewed the paper and approved the final version.

VIII. DATA AVAILABILITY STATEMENT

The raw data used in the meta-analysis are available in the online digital repository figshare at <https://doi.org/10.6084/m9.figshare.21769979>.

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- References identified with an asterisk (*) are the primary studies included in the meta-analysis. Additional references cited only within the supporting information are identified with a dagger symbol (†).
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X. SUPPORTING INFORMATION

Additional supporting information may be found online in the Supporting Information section at the end of the article.

Fig. S1. PRISMA flowchart illustrating the processes for the selection of articles included in the meta-analysis.

Fig. S2. Global distribution of the biochar addition experiments selected for this meta-analysis.

Fig. S3. Funnel plots for soil enzyme activities, ligninase:cellulase ratio and soil organic carbon in this meta-analysis.

Fig. S4. Effects of biochar addition on soil enzyme activities, ligninase:cellulase ratio and soil organic carbon sequestration as categorised by feedstock type, pyrolysis temperature of biochar and soil type.

Fig. S5. Effects of biochar addition on plant biomass and relationship between the responses of soil organic carbon and plant biomass to biochar addition.

Fig. S6. Effects of biochar addition on soil carbon:nitrogen ratio and NH_4^+ availability, and relationship between biochar-induced changes in carbon:nitrogen ratio and the application rate of biochar addition.

Fig. S7. Relationships between the responses of cellulase, ligninase and ligninase:cellulase ratio with various microbial attributes after biochar addition.

Fig. S8. Effects of biochar addition on fungal abundance, and relationship between the log response ratio of fungal abundance and biochar carbon:nitrogen ratio after biochar addition.

Table S1. Overview of studies included in our meta-analysis.

Table S2. Overview of cellulases and ligninases included in this meta-analysis.

Table S3. Results of sensitivity analysis for soil enzyme activities and soil organic carbon with biochar addition.

(Received 18 August 2022; revised 27 February 2023; accepted 1 March 2023; published online 13 March 2023)