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A Bimodal Extension of the ARYA&PARIS Approach for Predicting Hydraulic Properties of Structured Soils

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Abstract:	<p>The main purpose of this paper is to develop a bimodal pedotransfer function to obtain soil water retention (WRC) and hydraulic conductivity (HCC) curves. The proposed pedo-transfer function (PTF) extends the Arya and Paris (AP) approach, which is based on particle size distribution (PSD), by incorporating aggregate-size distribution (ASD) into the PTF to obtain the bimodal WRC. A bimodal porosity approach was developed to quantify the fraction of each of the porous systems (matrix and macropores) in overall soil porosity. Saturated hydraulic conductivity, K_0, was obtained from WRC using the Kozeny-Carman equation, whose parameters were inferred from the behaviour of the bimodal WRC close to saturation. Finally, the Mualem model was applied to obtain the HCC. In order to calibrate the PTF, measured soil physical and hydraulic properties data were used, coming from field infiltration experiments from an irrigation sector of 140 ha area in the "Sinistra Ofanto" irrigation system in Apulia, southern Italy. The infiltration data were fitted by using both bimodal and unimodal hydraulic properties by an inverse solution of the Richards equation. The bimodal "measured" hydraulic properties were then used to calibrate the scaling parameter (α_{AP}) of the proposed bimodal AP (bimAP) PTF. Similarly, for the sake of comparison with the bimodal results, the unimodal hydraulic properties were used to calibrate the α_{AP} of the classical unimodal AP (unimAP) PTF. Compared to the unimAP PTF, the proposed bimAP significantly improves the predictions of the mean WRC parameters and K_0, as well as the prediction of the shape of the whole HCC. Moreover, compared to the unimodal approach, it also allows keeping the hydraulic parameters' spatial variability observed in the calibration dataset. Multiple linear regression (MLR) was also applied to analyse the sensitivity of the bimodal α_{AP} parameter to textural and structural features, confirming significant predictive effects of soil structure.</p>
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1 **A BIMODAL EXTENSION OF THE *ARYA&PARIS* APPROACH FOR PREDICTING HYDRAULIC PROPERTIES OF**
2 **STRUCTURED SOILS**

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9
10 **Abstract:** The main purpose of this paper is to develop a bimodal pedotransfer function to obtain
11 soil water retention (WRC) and hydraulic conductivity (HCC) curves. The proposed pedo-transfer
12 function (PTF) extends the Arya and Paris (AP) approach, which is based on particle size
13 distribution (PSD), by incorporating aggregate-size distribution (ASD) into the PTF to obtain the
14 bimodal WRC. A bimodal porosity approach was developed to quantify the fraction of each of the
15 porous systems (matrix and macropores) in overall soil porosity. Saturated hydraulic conductivity,
16 K_0 , was obtained from WRC using the Kozeny-Carman equation, whose parameters were inferred
17 from the behaviour of the bimodal WRC close to saturation. Finally, the Mualem model was applied
18 to obtain the HCC. In order to calibrate the PTF, measured soil physical and hydraulic properties
19 data were used, coming from field infiltration experiments from an irrigation sector of 140 ha area
20 in the "Sinistra Ofanto" irrigation system in Apulia, southern Italy. The infiltration data were fitted
21 by using both bimodal and unimodal hydraulic properties by an inverse solution of the Richards
22 equation. The bimodal "measured" hydraulic properties were then used to calibrate the scaling
23 parameter (α_{AP}) of the proposed bimodal AP (*bimAP*) PTF. Similarly, for the sake of comparison
24 with the bimodal results, the unimodal hydraulic properties were used to calibrate the α_{AP} of the
25 classical unimodal AP (*unimAP*) PTF. Compared to the *unimAP* PTF, the proposed *bimAP*

Abbreviations: AP, Arya and Paris model; *bimAP*, bimodal Arya and Paris model; *unimAP*, unimodal Arya and Paris model

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26 significantly improves the predictions of the mean WRC parameters and K_0 , as well as the
27 prediction of the shape of the whole HCC. Moreover, compared to the unimodal approach, it also
28 allows keeping the hydraulic parameters' spatial variability observed in the calibration dataset.
29 Multiple linear regression (MLR) was also applied to analyse the sensitivity of the bimodal α_{AP}
30 parameter to textural and structural features, confirming significant predictive effects of soil
31 structure.

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33 properties variability, soil water retention, soil hydraulic conductivity

34

35 **1. Introduction**

36 The basis for understanding and solving agro-environmental problems increasingly lies in the use
37 of agro-hydrological models. Such models frequently rely on mechanistic descriptions of
38 fundamental processes involved in water and solute transport in soils (Abrahamsen and Hansen,
39 2000; Coppola et al., 2019; Šimůnek et al., 2008; Van Dam et al., 1997). Richards' equation (RE) and
40 the Advection-Dispersion equation (ADE) are generally used for water flow and solute transport,
41 respectively. Solving RE requires that soil water-pressure head, $\theta(h)$, and hydraulic conductivity-
42 water content, $K(\theta)$, functions be specified at the space scale of concern. For large-scale
43 applications, large hydraulic properties datasets are required to characterize the high spatial (and
44 temporal) variability of soil hydraulic properties naturally found in extensive areas (Coppola et al.,
45 2009a; Sposito, 1998). This is one of the more frustrating problems for soil scientists and
46 hydrologists, because direct measurements are cumbersome and expensive, and may represent the
47 main limit to using mechanistic models for large scale applications. This is also the chief
48 justification for the use of simpler approaches (bucket approach, for example) than the RE.
49 In attempts to overcome this problem, in recent decades great efforts have been made to develop
50 methods to estimate soil hydraulic properties from simpler data in the case of extensive direct

51 characterizations. Since hydraulic properties are affected by other physical, chemical and biological
52 properties, which are considered easier and cheaper to measure, empirical relations to predict
53 them have been proposed. Most of these expressions can be classified as pedotransfer functions
54 (PTFs, after Bouma, 1987) as they translate “readily” available information into the properties
55 needed to solve RE. Continuous PTFs (Rawls and Brakensiek, 1985; Minasny et al., 1999) that
56 calculate hydraulic properties from particle size distribution and additional soil variables such as
57 bulk density via a mathematical relationship are being constantly improved. Neural network
58 analysis has also been used to generate empirical PTFs (Schaap and Bouten, 1996). Leij et al. (2004)
59 extended the use of neural networks by introducing terrain attributes. Overviews of the current
60 status of PTF approaches are given by Basile et al. (2019) and Pachepsky et al. (2004).

61 Particle size distribution (PSD) data have also been used as a basis for estimating soil water
62 retention using semi-physical PTFs (Haverkamp and Parlange, 1986). Arya and Paris (1981)
63 significantly contributed to the expansion and spread of the approach. Their physico-empirical
64 approach is mainly based on the similarity between shapes of the cumulative PSD and $\theta(h)$ curves.
65 The model originally developed was refined (Arya et al., 1999a), also after later investigations (e.g.,
66 Basile and D’Urso, 1997), suggesting improvements pertaining to the limited flexibility of the
67 formulation. Arya et al. (1999b) also derived an expression to compute $K(\theta)$ directly from PSD,
68 based on the same soil structure model leading to the $\theta(h)$ relationship (Arya et al., 1999a; Arya and
69 Paris, 1981). Hereafter, such an approach will be referred to as an AP approach.

70 Although the performance of PTFs for the retention curve has continuously improved, due also to
71 increasing database size, they have still to be much improved on at least two interrelated issues: 1)
72 ability to accurately predict saturated/unsaturated hydraulic conductivity; 2) ability to predict the
73 spatial variability naturally found in measured soil hydraulic properties.

74 1) As for the issue of saturated hydraulic conductivity, K_0 , Loague (1992) used textural-based K_0
75 estimates in a rainfall-runoff model to be applied in a small catchment and concluded that texture

76 was not a substitute for actual K_0 field data. Sobieraj et al. (2001) compared the performance of
77 nine PTFs for estimating K_0 in modelling the storm flow generated in a rainforest catchment. They
78 concluded that the PTFs used generally underestimated measured K_0 , thus inadequately predicting
79 hydrograph attributes, and grossly overestimating total runoff and peak runoff for almost all the
80 events they examined. Tietje and Hennings (1996) tested six different PTFs and found different
81 accuracy in the results between soils from the US and soils from Germany. Their study also showed
82 that the prediction of K_0 using PTFs is inaccurate including the mean values of K_0 and their
83 geometric standard deviation, especially for clay and silt soils. Vereecken et al. (2010) studied the
84 use of PTFs to estimate van Genuchten and Mualem parameters. Their results showed inaccuracy in
85 estimating hydraulic conductivity parameters using texture-based PTFs. A likely reason for this
86 failure is that saturated hydraulic conductivity is largely dependent on soil structure and that
87 currently used PTFs do not adequately (or at all) account for macroporosity in soils. The
88 characteristics of macropores (mostly interaggregate pores) are not related to soil texture, such
89 that soils with similar texture may have completely different saturated hydraulic conductivity
90 (Coppola et al., 2009b; Pachepsky et al., 2004; Vereecken et al., 2010).

91 Besides on K_0 , excluding the macropore information in a PTF may have an impact even on the shape
92 of the whole HCC, especially when models based on the Hagen-Poiseuille, such as Mualem's
93 conductivity model (Mualem, 1976), are used to predict hydraulic conductivity starting from the
94 WRC. Actually, almost all of the existing PTFs assume pore systems with unimodal pore size
95 distributions. This is justified by the fact that these PTFs have been calibrated by using datasets
96 with either limited or no measurements at all close to saturation, which hold the information on the
97 soil structure. The van Genuchten (1980) model is widely adopted to parametrize the unimodal
98 WRC. Using the van Genuchten parameters in the Mualem model, estimation of the hydraulic
99 conductivity is obtained by using the measured K_0 as matching factor. By contrast, when data close
100 to saturation are available, a macropore portion of the water retention curve becomes frequently

101 evident, which may require a bimodal model to be correctly described (Coppola, 2000; Coppola et
102 al., 2009a, 2009b; Durner, 1994; Othmer et al., 1991; Ross and Smettem, 1993; Wilson et al., 1992).
103 In all of the above papers, it has been shown that using either a unimodal or a bimodal model to
104 describe WRC may induce significant changes in the prediction of the whole HCC. This is because, as
105 discussed by Durner (1994) and shown experimentally by Coppola (2000), the formulation itself of
106 Mualem's conductivity model makes the model particularly sensitive to the slope of the retention
107 curve near saturation. This subject has received considerable theoretical and experimental
108 treatment which shows that relatively small variations in water content close to saturation may be
109 amplified by the algorithm for determining hydraulic conductivity (Coppola, 2000; Van Genuchten
110 and Nielsen, 1985; Vogel and Cislerova, 1988).

111 2) As for the issue of spatial variability, which is strictly related to the issue described above, most
112 of the studies around PTFs have focused more on the predictive capability of the mean values of
113 hydraulic parameters than on their spatial variability. Much rarer are the attempts to evaluate the
114 ability of PTFs to describe the spatial variability of soil hydraulic properties (Espino et al., 1996;
115 Romano and Santini, 1997; Leij et al., 2004). Coppola et al. (2013) found that Rosetta PTF-based
116 hydraulic parameters resulted in very low variability compared to the measured hydraulic
117 parameters (see graphs 5 and 6 and tables 1 and 2 in their paper). This was especially true for the α
118 and K_0 parameters of the van Genuchten-Mualem model (van Genuchten, 1980), which are known
119 to be the parameters mainly related to the soil structure. The authors ascribed this behaviour to the
120 fact that even in a quite homogeneous soil from a textural perspective, the structure may induce a
121 variability in the soil hydraulic parameters which cannot be reproduced by PTFs not including
122 explicitly structural information. Additionally, for the reasons already discussed above, the use of a
123 unimodal model to describe bimodal porous media may also contribute to flatten the variability
124 observed in the measurements.

125 Earlier efforts to properly account for macroporosity were mainly oriented to introducing

126 additional information in PTFs explicitly considering macropore sizes and counts (McKeague et al.,
127 1982; Nimmo et al., 2007). As also argued by Pachepsky and Rawls (2003), these studies showed
128 that there is a potential usefulness in using aggregate size distribution in PTFs to improve water
129 retention and hydraulic conductivity estimates.

130 Based on these premises, the aim of this study was to propose a bimodal extension of the AP
131 approach (hereafter *bimAP*), which incorporates information on aggregate size distribution, to
132 improve the prediction of the hydraulic properties and their spatial variability for structured soils.
133 A large dataset of in situ infiltration measurements was used to establish the bimodal nature of
134 hydraulic properties. Compared to the original AP model (hereafter *unimAP*), the *bimAP* model
135 requires additional measurements of aggregate size distribution (ASD) and single-aggregate bulk
136 density. Also, the *bimAP* water retention estimates require fitting by a bimodal water retention
137 model to obtain the *bimAP* scaling parameter. A multiple linear regression was applied to analyse
138 the degree of dependence of this scaling parameter on the textural and structural information. All
139 the estimates from the *bimAP* were compared to those from the *unimAP*, to show the effects of not
140 considering the effects of the structure on the predictions of the hydraulic properties and their
141 spatial variability.

142 **2. Materials and Methods**

143 2.1. Hydraulic property models

144 In this paper, we use water retention models assuming pore systems with either unimodal or
145 bimodal pore-size distributions. The van Genuchten (van Genuchten, 1980) model for unimodal
146 porous systems is as follows:

$$S_e = \frac{\theta - \theta_r}{\theta_0 - \theta_r} = [1 + |\alpha_{VG} h|^n]^{-m} \quad h < 0$$
$$\theta = \theta_s \quad h \geq 0 \quad (1)$$

147 where h is the pressure head ($h \leq 0$), S_e is effective saturation and θ is the water content (θ_s and θ_r
 148 are the water content at $h=0$ and for $h \rightarrow \infty$, respectively). α_{VG} (cm^{-1}), n and $m=1-1/n$ are shape
 149 parameters. The effective saturation, S_e , may be considered a cumulative distribution function of
 150 pore size with a density function $f(h)$ which may be expressed by:

$$f(h) = \frac{dS_e}{dh} \quad (2)$$

151 The presence of aggregates frequently results in a retention function curve with at least two
 152 inflection points. To represent such behaviour, a double porosity approach can be used which
 153 assumes that the pore space from θ_r to θ_s consists of two pore-size distributions, each occupying a
 154 fraction W_i of that pore space (Coppola, 2000; Durner, 1994). The model proposed by Durner
 155 (1994) is as follows:

$$S_e = \sum \beta_i \left[\frac{1}{1+(\alpha_{VG,i}h)^{n_i}} \right]^{m_i} \quad 0 < \beta_i < 1 \quad \text{and} \quad \sum \beta_i = 1 \quad i = 1,2 \quad (3)$$

156 in which β_1 and β_2 are the weighting of the total pore space fraction to be attributed respectively to
 157 inter-aggregate pores (the macropores) and intra-aggregate pores (the micropores or matrix
 158 pores), and $\alpha_{VG,i}$, n_i and m_i still represent the fitting parameters for each of the partial curves.
 159 The unsaturated hydraulic conductivity function is described by using the Mualem model (Mualem,
 160 1976). It is based on the capillary bundle theory and relates relative hydraulic conductivity, K_r , to
 161 the pore-size distribution function $f(h)$ with the equation:

$$K_r(h) = \frac{K(h)}{K_0} = S_e^\tau [\eta(h)/\eta(0)]^2 \quad (4)$$

$$\eta(h) = \int_{-\infty}^{\infty} h^{-1} f(h) dh$$

162 in which τ is a parameter accounting for the dependence of the tortuosity and the correlation
 163 factors on the water content. τ was fixed at a value of 0.5. The relative hydraulic conductivity is thus
 164 scaled using the saturated hydraulic conductivity, K_0 (hydraulic conductivity at $h=0$), as matching

165 factor.

166 In the case of the unimodal van Genuchten model and assuming $m=1-1/n$, K_r becomes

$$K_r(S_e) = \frac{K(S_e)}{K_0} = S_e^\tau \left[1 - \left(1 - S_e^{1/m} \right)^m \right]^2 \quad (5)$$

167 In the case of bimodal Durner water retention, equation (5) becomes equation (6) (Priesack and
168 Durner, 2006).

$$K_r(S_e) = \frac{K(S_e)}{K_0} = \left(\sum_{i=1}^2 \beta_i S_{e,i} \right)^\tau \left\{ \frac{\sum_{i=1}^2 \beta_i \alpha_i \left[1 - \left(1 - S_{e,i}^{1/m_i} \right)^{m_i} \right]}{\sum_{i=1}^2 \beta_i \alpha_i} \right\}^2 \quad (6)$$

169 2.2. Developing the Bimodal Arya and Paris approach for soil water retention

170 This section is devoted to describing in detail the *bimAP* approach. For the *unimAP* approach, the
171 sequence of steps to obtain water retention curves by applying its classical concepts may be found
172 in Arya and Paris (1981). We only recall here that it assumes cylindrical pores which are built by
173 overlying the single particles (coming from the PSD) in a given size range.

174 In developing the *bimAP* approach, the main assumption is that the *unimAP* approach may be
175 extended also to the ASD. Specifically, in the *bimAP* case, the total porous system is assumed to be
176 partitioned into structural pore space (macropores, arising from particle aggregation) and matrix
177 pore space (micropores, the only pores classically considered by the *unimAP*). Similar to matrix,
178 which in the *unimAP* are ideally cylindrical and come from the overlying of the single particles of
179 the PSD, the pores of the structural part of the porous system are also assumed to be cylindrical and
180 simply built by overlying the aggregates in a given size range.

181 We will consider two cases: 1) the whole sample only consists of an ensemble of aggregates,
182 without particle inclusions in the interspace among aggregates (see Figure 1a); 2) the whole sample
183 consists of an ensemble of aggregates with particle inclusions in the interspace among aggregates
184 (see Figure 1b).

185 Figure 1

186

187 In the *bimAP* approach, the porosity of the soil sample was divided into two porosities: macropore
188 porosity (or structural porosity) which occupies the space between the aggregates, and matrix
189 porosity which occupies the space between soil particles.

190 Below we give definitions and calculations, first for the single aggregate and then for the whole soil
191 sample, which will be used in the *bimAP* approach applied to both the PSD and ASD.

192

193 *i) Single aggregate*

194 Let us start from the calculations for a single aggregate (one of the red clods in figure 1a and 1b).

195 The aggregate will consist of solid particles and matrix pores (intra-aggregate pores). Below, the
196 label *ag* is used for single aggregates.

197 The volume of solid particles in a single aggregate, assuming that the particle density $\rho_s=2.65$
198 g/cm^3 , is:

$$v_{ag,prt} = \frac{w_{ag,prt}}{2.65} \quad (7)$$

199 where $w_{ag,prt}$ = dry weight of the aggregate = dry weight of the particles in the aggregate

200 The total volume of the aggregate is:

$$v_{ag,t} = v_{ag,prt} + v_{ag,p} \quad (8)$$

201

202 where $v_{ag,p}$ = volume of pores in the aggregate (determined by the ethyl alcohol method, see
203 section 2.4).

204 The bulk density of the aggregate, which will be used in the calculations for the whole sample, is:

$$\frac{w_{ag,prt}}{v_{ag,t}} = \rho_{b,ag} \quad (9)$$

205

206 Finally, the porosity of the aggregate is:

$$\frac{v_{ag,p}}{v_{ag,t}} = \varphi_{ag} \quad (10)$$

207

208 *ii) Sample (ensemble of aggregates, without particle inclusions in the interspace among aggregates)*

209 *(figure 1a)*

210 In this case, it is assumed that all the sample weight consists of aggregates (no single particles

211 leaving the finest sieve for aggregates). In the following, label S stands for sample.

212 The bulk density, $\rho_{b,S}$ and total porosity, $\Phi_{S,T}$, of the sample are:

$$\rho_{b,S} = \frac{W_{S,T}}{V_{S,T}} \quad (11)$$

$$\Phi_{S,T} = 1 - \frac{\rho_{b,S}}{2.65}$$

213

214 where $W_{S,T}$ is the total dry weight of particles in the sample and $V_{S,T}$ the total volume of the soil

215 sample.

216 The volume occupied by the aggregates, $V_{S,AG}$, and by the inter-aggregate pores, in the soil sample

217 (respectively the red and blue parts in figure 1) is:

$$V_{S,AG} = \frac{W_{S,T}}{\rho_{b,ag}} \quad (12)$$

$$V_{S,MCP} = V_{S,T} - V_{S,AG}$$

218 In equation (12), note that we use the bulk density of the aggregates determined on the single

219 aggregate. We assume that $W_{S,T}$ is also the weight of the aggregates in the soil sample (no particle

220 inclusion).

221 The inter-aggregate porosity, $\Phi_{S,MCp}$, and the intra-aggregate or matrix porosity, $\Phi_{S,MXp}$, are:

$$\Phi_{S,MCp} = \frac{V_{S,MCp}}{V_{S,T}} \quad (13)$$

$$\Phi_{S,MXp} = \Phi_{S,T} - \Phi_{S,MCp}$$

222

223 $\Phi_{S,MCp}$ and $\Phi_{S,MXp}$ will be used, respectively, as saturated water contents for the structural part and
 224 for the matrix part of the water retention curve obtained by the AP method applied to both the
 225 aggregates and the particles (see section *iv.* below *bimAP approach for calculating the pore-spaces,*
 226 *capillary radius, pressure head and water contents for each class of particles or aggregates*).

227 The volume of the intra-aggregate pores (matrix pores), $V_{S,MXp}$, and that of the solid particles, $V_{S,prt}$,
 228 in the soil sample are:

$$V_{S,MXp} = \Phi_{S,T}V_{S,T} - V_{S,MCp} \quad (14)$$

$$V_{S,prt} = V_{S,T} - V_{S,MCp} - V_{S,MXp}$$

229

230 The matrix particle density, $\rho_{S,prt}$, and the aggregate density, $\rho_{S,AG}$, are respectively:

$$\rho_{S,prt} = \frac{W_{S,T}}{V_{S,prt}} \quad (15)$$

231

232 Note that in equation (15), $\rho_{S,AG}$ corresponds to the density $\rho_{b,ag}$ determined on the single
 233 aggregates.

234 Finally, from equations (14), one can define the void index of the whole sample, e_s , the void index

235 coming from the pores in the aggregates (intra-aggregate pores = micro-pores), e_{sMX} , and the void
 236 index from the pores among the aggregates (inter-aggregate pores = macro-pores), e_{MC} :

$$e_s = \frac{V_{S,MCp} + V_{S,MXp}}{V_{S,prt}}$$

$$e_{MX} = \frac{V_{S,MXp}}{V_{S,prt}} \quad (16)$$

$$e_{MC} = \frac{V_{S,MCp}}{V_{S,prt}}$$

237

238 *iii. Sample (ensemble of aggregates with particle inclusions in the interspace between aggregates)*
 239 *(figure 1b)*

240 In this case, we assume that the inclusions (the orange particles) in the inter-aggregate space
 241 consist of single particles just occupying a part of this but without an inner porosity. The total
 242 weight of the sample does not correspond to the total weight of the aggregates. The latter
 243 corresponds to a fraction ε of the total weight. The inclusions occupy a space, $V_{S,MCprt}$, which is the
 244 volume of solid particles included in the inter-aggregate space.

245 In this case, the part of the total volume occupied by the aggregates in the soil sample (the red part
 246 in figure 1b) is:

$$V_{S,AG} = \frac{\varepsilon W_{S,T}}{\rho_{b,ag}} \quad (17)$$

247

248 Note that, as in equation (12), we use the bulk density of the aggregates determined on the single
 249 aggregate. We assume that $\varepsilon W_{S,T}$ is the weight of the aggregates in the soil sample.

250 Now, the total volume of inter-aggregate pores is:

$$V_{S,MCp} = V_{S,T} - V_{S,AG} - V_{S,MCprt} \quad (18)$$

251

252 where $V_{S,MCprt} = \frac{W_{S,T} - \varepsilon W_{S,T}}{2.65}$ = volume of the inclusions in inter-aggregate space. The remaining

253 calculations for $\Phi_{S,MCp}$ and $\Phi_{S,MXp}$ are as for the no-inclusions case.

254 *iv. bimAP approach for calculating pore spaces, capillary radius, pressure head and water contents for*
255 *each class of particles or aggregates*

256 As for the matrix pores, also the pores of the structural part of the porous system are assumed to be
257 cylindrical and simply built by overlying the aggregates in a given size range.

258 All the calculations require that PSD and ASD (both expressed as percentages) be divided into a
259 number of N radius classes.

260 Should only a negligible fraction of the air-dried soil sample be left in the smaller sieve for the
261 aggregates, the case *ii. Sample (ensemble of aggregates, without particle inclusions in the interspace*
262 *among aggregates)* will apply (this is our case in this paper). Of course, if this fraction were to be
263 more significant, the approach would be simply extended by including its weight $(1-\varepsilon)$ in the
264 calculations (see case *iii. Sample (ensemble of aggregates with particle inclusions in the interspace*
265 *among aggregates)*).

266 In case *ii.*, the dry weight of the total number of single particles in the sample coincides with that of
267 the total number of aggregates:

268

$$W_{S,T} = \sum_{i=1}^{N_{PSD}} W_{S,T,i_{R<1}} = \sum_{i=1}^{N_{ASD}} W_{S,T,i_{R>1}} \quad (19)$$

269

270 where $W_{S,T,i_{R<1}}$ is the dry weight of the particles in the i^{th} class of particles of radius $R<1\text{mm}$,

271 $W_{S,T,i_{R>1}}$ is the dry weight of the aggregates in the i^{th} class of aggregates of radius $R>1\text{mm}$, and

272 N_{PSD} and N_{ASD} are the number of classes into which the PSD and ASD distribution are divided.

273 The volume of pores from the i^{th} class of particles of radius $R<1\text{mm}$ may be obtained as

274

$$V_{S,P,i_{R<1}} = \frac{W_{S,T,i_{R<1}} e_{MX}}{\rho_{S,prt}} = \frac{W_{S,T,i_{R<1}} e_{MX} V_{S,prt}}{W_{S,T}} \quad (20)$$

275

276 where $\frac{W_{S,T,i_{R<1}}}{W_{S,T}}$ is the solid mass per unit sample mass in the i^{th} particle-size range. It is obtained by
277 taking the differences in cumulative percentages corresponding to successive particle sizes divided
278 by 100, such that the sum of $\frac{W_{S,T,i_{R<1}}}{W_{S,T}}$ for all the n classes is unity.

279 Similarly, the volume of pores from the i^{th} class of aggregates of radius $R>1\text{mm}$ is:

$$V_{S,P,i_{R>1}} = \frac{W_{S,T,i_{R>1}} e_{AG}}{\rho_{S,AG}} = \frac{W_{S,T,i_{R>1}} e_{AG} V_{S,AG}}{W_{S,T}} \quad (21)$$

280

281 where $\frac{W_{S,T,i_{R>1}}}{W_{S,T}}$ is the solid mass per unit sample mass in the i^{th} aggregate-size range. Again, it is
282 obtained by taking the differences in cumulative percentages corresponding to successive
283 aggregate sizes divided by 100, such that the sum of the $\frac{W_{S,T,i_{R>1}}}{W_{S,T}}$ for all the n classes is unity.

284 By using equations (20) and (21), it is possible to calculate the water content in the matrix pores,
285 $\theta_{S,i_{R<1}}$, and that in the macropores, $\theta_{S,i_{R>1}}$, as follows:

$$\theta_{S,i_{R<1}} = \frac{\sum_{j=1}^i V_{S,P,i_{R<1}}}{V_{S,T}} \quad (22)$$

$$\theta_{S,i_{R>1}} = \frac{\sum_{j=1}^i V_{S,P,i_{R>1}}}{V_{S,T}}$$

286

287 which are obtained respectively by progressively filling matrix (and macropore) volumes with
288 water up to the selected i^{th} $V_{S,P,i_{R<1}}$ volume (i^{th} $V_{S,P,i_{R>1}}$ volume).

289 Now, the number of particles in the i^{th} class of particles of radius $R_i < 1\text{mm}$ (all the particles in the
 290 range are assumed to form a single cylindrical pore of volume $V_{S,P,iR < 1}$) may be obtained as follows:

291

$$n_{i,prt} = \frac{3V_{S,P,iR < 1}}{4\pi R_i^3} \quad (23)$$

292

293 Similarly, the number of aggregates in the i^{th} class of aggregates of radius $R_i > 1\text{mm}$ (all the
 294 aggregates in the range are assumed to form a single cylindrical pore of volume $V_{S,P,iR > 1}$) are
 295 calculated as:

296

$$n_{i,AG} = \frac{3V_{S,P,iR > 1}}{4\pi R_i^3} \quad (24)$$

297

298

299 The radius of the pores in the i^{th} class of particles of radius $R_i < 1\text{mm}$, $r_{i,MX}$, and the radius of the
 300 pores in the i^{th} class of aggregates of radius $R_i > 1\text{mm}$, $r_{i,AG}$, are:

$$r_{i,MX} = R_i \frac{[4e_{MX}n_{i,prt}^{(1-\alpha_{MX})}]^{0.5}}{6} \quad (25)$$

301

302 from which the pressure head for the radius of the pores in the i^{th} class of particles of radius
 303 $R_i < 1\text{mm}$, $h_{i,MX}$, and that corresponding to the radius of the pores in the i^{th} class of aggregates of
 304 radius $R_i > 1\text{mm}$, $h_{i,MC}$, may be calculated as follows:

$$h_{i,MX} = \frac{2\sigma\cos\vartheta}{\rho_w g r_{i,MX}} \quad (26)$$

$$h_{i,MC} = \frac{2\sigma\cos\vartheta}{\rho_w g r_{i,MC}}$$

305

306 where σ is water-air surface tension, ϑ is contact angle, ρ_w is density of water, and g is gravity
 307 acceleration.

308 α_{MX} and α_{MC} are scaling parameters for pore length, accounting for the fact that the actual soil
 309 particles and aggregates are not spherical. In their classical *unimAP*, Arya and Paris (1981) assumed
 310 the parameter α_{MX} to be >1 under the hypothesis that each particle contributes a length greater
 311 than the diameter of an equivalent sphere. As a first approximation, in our *bimAP* approach, we will
 312 assume that this hypothesis extends to α_{MC} and that the two parameters have the same value, such
 313 that $\alpha_{MX} = \alpha_{MC} = \alpha_{AP}$. Parameter α_{AP} has to be estimated by fitting the AP estimates to measured
 314 water retention curves for both the *unimAP* and *bimAP* cases (see section 2.7 below. *Fitting the AP*
 315 *estimates to the measured hydraulic properties to calibrate the α_{AP} parameter*).

316 The matrix and the macropore parts of the water retention are simply obtained by combining pairs
 317 of $\theta_{S,i_{R<1}} - h_{i,MX}$ and $\theta_{S,i_{R>1}} - h_{i,MC}$. Total water retention is obtained by summing up the two partial
 318 contributions (see the symbols in Figure 2).

319

320 Figure 2

321

322 *v. Using the information from bimAP to estimate K_0*

323 Saturated hydraulic conductivity, K_0 , was obtained by using the following Kozeny-Carman equation

324 (Kozeny, 1927; Carman, 1937):

325

$$K_0 = \eta \phi_e^\gamma \quad (27)$$

326

327 where ϕ_e is the effective porosity, which is the difference between the saturated water content and
328 the water content at field capacity (330 cm matric suction), and η and γ are constants. All these
329 parameters were estimated from the curve obtained by fitting the water retention model (whether
330 unimodal or bimodal) to the AP estimates. Obviously, all the parameters in the Kozeny-Carman
331 equation will change depending on the approach used (*unimAP* or *bimAP*).

332 In order to estimate the other parameters in equation (27), in our approach we used the version of
333 the Kozeny-Carman equation proposed by Timlin et al. (1999):

334

$$K_0 = 0.0131 \left(\frac{F}{l} \right)^{0.5} \phi_e^{2.5} \quad (28)$$
$$F = 0.148/h_b$$
$$l = 1.86(2 - \lambda)^{5.34}$$

335

336 where F and l are parameters related to the fractal dimensions of porosity (Rawls et al., 1993), λ is
337 the pore size distribution index, and h_b is the air-entry potential in the Brooks and Corey water
338 retention model (BC model) (Brooks and Corey, 1964):

$$S_e = (\theta - \theta_r)/(\theta_s - \theta_r) = \left(\frac{h}{h_b} \right)^\lambda \quad (29)$$

339

340 In the case of *bimAP*, since K_0 is related to the water retention characteristics near saturation, λ and
341 h_b were obtained by fitting the BC model to the upper part of the *bimAP* WRC. In the case of the
342 *unimAP* approach, λ and h_b were estimated by fitting the BC model to the whole water retention
343 curve.

344

345 2.3. The study area

346 The study area is sector 6 of irrigation district 10 in the “Sinistra Ofanto” irrigation system, which is
347 located on the left bank of the Ofanto river in Puglia, Southeast Italy (see figure 3). The study sector
348 covers an area of 140 hectares of agricultural land and is irrigated with an on-demand pressurized
349 network. The whole district covers 22,500 hectares of agricultural land.

350

351 Figure 3

352

353 2.4. Measurement of soil physical parameters

354 The approach proposed in this paper assumes the bimodality of soil porosity. It requires
355 measurement of the size distribution of the single particles (PSD), as well as the size distribution of
356 the aggregates (ASD) present in an undisturbed soil sample.

357 Undisturbed soil samples were collected using steel cylinders 7 cm in diameter and 7 cm high in 90
358 sites selected in the study area (about 140 hectares). Additionally, some disturbed soil was also
359 sampled to determine the average bulk density of single aggregates.

360 After measuring the sample volume, the undisturbed soil was removed from the sampler and air-
361 dried for at least one week. Sieve analysis was carried out on each sample to obtain the ASD curve
362 according to the dry-sieving method proposed by Nimmo & Perkins (2002). The sieve sizes used in
363 this analysis were: 40, 31.5, 25, 20, 16, 10, 8, 5, 2 and 1 mm. All the soil remaining in a sieve was
364 considered to consist of aggregates of a radius exceeding the sieve size. All the soil passing the
365 narrowest sieve was kept. All the soil initially contained in a sampler was then oven-dried at 105 °C
366 for 24 hours to determine bulk density, ρ_b , and thus PSD. Total porosity was calculated from the
367 measured bulk density assuming that particle density was 2.65 g/cm³. PSD curves were obtained by
368 using the hydrometer method combined with sieve analysis to characterize the range of particle

369 diameter from 2 up to 2000 μm (Gee and Or, 2002). The dry soil was lightly crushed on a tray using
370 a rolling pin to break up clods until the soil passed through a 2 mm sieve. Fifty grams of the sieved
371 soil were pre-treated with 30% (w/v) hydrogen peroxide until no reaction was revealed to remove
372 organic matter. After washing and air-drying of residual soil, chemical dispersion of soil particles
373 was achieved by mixing the soil sample with a 5 g L⁻¹ sodium-hexametaphosphate (HMP) solution
374 adjusted to pH 8.5, allowing the soil to soak overnight. Physical dispersion was obtained with
375 mechanical mixing with an electric stirrer working at 10000 rpm. Then soil samples were
376 transferred to 1000-mL sedimentation cylinders. After thorough mixing of soil suspension, the
377 suspension density was measured and recorded after 3, 10, 30, 60, 210, 1440 minutes with an
378 ASTM 152H. The hydrometer readings were also made at the same times on a blank solution to
379 correct for the density of HMP solution. At the end of readings, the contents of the cylinder were
380 poured out through a 45- μm sieve to retain coarser particles. The retained material was oven-dried
381 for 24 h at 105 °C and sieved with a nest of sieves of 1000, 500, 250, 106, 53 μm . The portion of
382 sand retained on each sieve was weighed and annotated. Following the above procedure, we
383 determined a particle size distribution curve composed by 11 experimental points for all of the soil
384 samples. Sand, silt and clay contents were expressed as percentages by mass of the fine-earth
385 fraction (<2 mm). According to the USDA soil classification, the texture of the soil samples in the
386 examined dataset ranged from silty-clay-loam to sandy-loam (see figure 4). Overall, the above
387 methods allowed 20 points to be obtained for PSDs and 10 for ASDs.

388

389 Figure 4

390

391 Both PSD and ASD curves were described by a parametric (van Genuchten-type) equation:

$$P = 100 + (T_s - 100)(1 + (T_\alpha D)^{T_n})^{-T_m} \quad (30)$$

392 where P is the percentage of the particle or aggregate passing from a sieve size; D is the sieve size;

393 T_s , T_α , T_n and T_m are parameters similar to those of van Genuchten (1980) model, and $T_m=1-1/T_n$.

394 The parameters were obtained by fitting equation (1) to the measured PSDs and ASDs (Figure 5).

395

396 Figure 5

397

398 The average bulk density of the single aggregates was determined using the disturbed soil samples.

399 As with undisturbed soil samples, the soil was left to air-dry for a week. Aggregates of different

400 sizes were then selected, whose porosity was determined by using the ethyl alcohol method

401 proposed by Moret-Fernandez and Lopez (2019). First, the dry aggregate was weighed. Then it was

402 immersed in ethanol in a beaker, which was covered well with biofilm to avoid ethanol evaporation.

403 The bubbling was observed for at least 20 minutes until it stopped, indicating aggregate saturation

404 with ethanol. After saturation, the aggregate was carefully taken from ethanol and placed on a

405 paper filter for less than 10 seconds before measuring its new weight. At the same time, the

406 temperature of the alcohol was measured using a mercury thermometer in order to determine the

407 alcohol density. This process was done for several aggregates from different locations and with

408 different sizes at room temperature set to be less than 25 °C. Thus, the volume of the pores, V_p , was

409 calculated as:

$$V_p = \frac{W_{agg-al} - W_{agg}}{\rho_{al}} \quad (31)$$

410 where W_{agg-al} is the weight of the aggregate after saturation with alcohol, W_{agg} is the dry weight of

411 the aggregate, and ρ_{al} is the alcohol density. The volume of the solid phase, V_s , in the aggregate was

412 then calculated as:

$$V_s = \frac{W_{agg}}{\rho_s} \quad (32)$$

413 where ρ_s is the solid particle density, which can be assumed to be 2.65 g/cm³. Finally, the

414 aggregate bulk density was calculated as follows:

$$\rho_{b,ag} = W_{agg}/(V_s + V_p) \quad (33)$$

415

416 2.5. Direct measurement of soil hydraulic parameters

417 Soil hydraulic parameters at each of the 90 studied sites were obtained using tension infiltrometers
 418 (Ankeny et al., 1988; Coppola et al., 2011). First, the soil surface was levelled. Then a ring was
 419 placed on the surface and a thin layer of homogeneous fossil sand was added to the soil surface to
 420 ensure good contact with the infiltrometer disc. At each site, infiltration experiments were carried
 421 out at four sequential water pressure head values (-15, -10, -5 and -1 cm). Water pressure was
 422 controlled by raising or lowering the tube in the bubble tower. A soil sample was taken before and
 423 after the infiltration process to measure the initial and final water content.

424 The cumulative infiltration data were used as input in an inverse solution of the 3D Richards
 425 equation to obtain both the unimodal and bimodal hydraulic property parameters by a parameter
 426 estimation procedure. The van Genuchten-Mualem and Durner-Mualem models were used to
 427 describe unimodal and bimodal hydraulic properties, respectively. As for the unimodal properties,
 428 they were estimated by using DISC software (Šimůnek and van Genuchten, 1996). Inverse solution
 429 using bimodal properties was carried out by using the software HYDRUS 2D/3D (Rassam *et al.*,
 430 2003; Šimůnek *et al.*, 2008). In both cases, the saturated water content for each site was fixed at the
 431 total porosity, residual water content was fixed as zero, τ was fixed as 0.5, in order to minimize the
 432 number of variables to be optimized. Parameter m was assumed to be $m=1-1/n$ for both the
 433 unimodal and the bimodal descriptions. The weight β_1 in the Durner model was assumed to be
 434 equal to the fraction of macroporosity to total porosity $\Phi_{S,MCp}/\Phi_{S,T}$. Eventually, the inverse
 435 solution would estimate the three parameters involved for the unimodal scenario (namely, α , n , and
 436 K_0) and five parameters for the bimodal scenario (namely: α_1 , n_1 , α_2 , n_2 , and K_0).

437

438 2.6. Akaike Information Criterion (AIC) to test the bimodality of the porous medium

439 The goodness of fit from the inverse solution for both the unimodal and bimodal scenarios was
 440 compared in order to test the bimodality of soil pores. The root-mean-square error (*RMSE*) was
 441 used as a measure of the distance between the predicted and the measured infiltrated depths
 442 (equation 35). Unimodal and bimodal scenarios involve a different number of parameters. The
 443 Akaike Information Criterion (AIC) was used to balance the goodness of fit and the number of
 444 parameters involved:

445

$$AIC = N_o \ln \left(\frac{RMSE}{N_o} \right) + 2k \quad (34)$$

$$RMSE = \sqrt{\frac{\sum_{i=1}^{N_o} (x_i - \hat{x})^2}{N_o}} \quad (35)$$

446

447 where N_o is the number of observations, $k =$ the number of parameters + 1, x_i is the variable
 448 obtained from field measurements and \hat{x}_i is the variable estimated from *unimAP* or *bimAP*. The
 449 lower the AIC, the better the fit.

450

451 2.7. Fitting AP estimates to the measured hydraulic properties to calibrate parameter α_{AP}

452 The proposed model is physically-based with one unknown parameter (both for *unimAP* and
 453 *bimAP*), which is the scaling parameter (α_{AP}). This parameter was estimated by fitting the *unimAP*
 454 and *bimAP* estimates to respectively the unimodal and bimodal measured water retention curves
 455 (see again the graph in figure 2).

456

457 2.8. Evaluating the dependence of parameter α_{AP} on textural and structural physical properties

458 Multiple linear regression (MLR) was applied to relate the scaling parameter α_{AP} to texture and
 459 aggregate properties with a view to predicting α_{AP} with only physical properties available and with

460 no prior knowledge of the hydraulic parameters. The regression analysis included: 1) the
461 parameters of the equations used to interpolate PSD and ASD (namely, T_s , T_n and T_α); 2) soil bulk
462 density (ρ_b); 3) the fraction of the macropores from total porosity (β_I). The regression model was
463 developed by using the hydraulic parameters obtained from the field measurements.

464

465 2.9. Schematic view of the approach used in the paper

466 For easier interpretation of the results of the *bimAP* application, Figure 6 summarizes the steps
467 followed in this paper to test the *bimAP* PTF and to compare it to the *unimAP* approach. On the one
468 hand, inverse solution of tension infiltration experiments was used twice to obtain the unimodal
469 van Genuchten (van Genuchten, 1980) and the bimodal Durner (Durner, 1994) hydraulic
470 properties. Hereafter, they will be called the measured hydraulic properties. The Akaike
471 Information Criterion (AIC) was used to establish the bimodality of the hydraulic properties. On the
472 other, both the unimodal and bimodal AP approaches were used to obtain, respectively, *unimAP*
473 and *bimAP* estimates of the WRC. These were fitted respectively to the unimodal and bimodal
474 measured WRC to obtain the scaling parameter (α_{AP}) for both the *unimAP* and *bimAP* approaches.
475 Saturated hydraulic conductivity (K_0) was then estimated from both *unimAP* and *bimAP* WRCs using
476 the Kozeny-Carman equation (Carman, 1937; Kozeny, 1927), and was subsequently used to obtain
477 $K(h)$ curves (HCCs) using the Mualem model (Mualem, 1976; Priesack and Durner, 2006). Finally,
478 multiple linear regression (MLR) was used to analyse the dependence of the scaling parameter, α_{AP} ,
479 on soil physical parameters.

480

481 Figure 6

482

483 3. Results and Discussion

484 3.1. Testing the bimodality of the measured hydraulic property dataset

485 Some results of the inverse solutions for three locations in the study area are shown in figure 7. The
486 symbols represent the observed infiltration depths, while the solid lines represent the infiltration
487 depths obtained from the inverse solution. Each plot on the left represents a unimodal inverse
488 solution, whereas the corresponding right-hand plot represents the bimodal inverse solution for
489 the same location. Figure 8 shows the results of AIC values for unimodal and bimodal inverse
490 solutions for all 90 sites investigated.

491

492 Figure 7

493

494

495 Figure 8

496

497 Looking at both the graphs in figure 7 and the AIC values in figure 8, fitting infiltrated depths by
498 using the bimodal model frequently gives better results than the unimodal model. AIC analysis
499 shows that the bimodal model provides better estimates in almost 73% of the locations in the study
500 area. These results allowed us to conclude that the hydraulic property dataset is mostly bimodal
501 and is thus appropriate for calibrating the proposed *bimAP* PTF.

502 The first two columns in Table 1 show the average and the standard deviation of the parameters of
503 the unimodal and bimodal hydraulic properties obtained from the measurements (subscript *meas*).

504 It is worth noting that in the unimodal case the parameter K_0 is on average higher than the bimodal
505 K_0 , with also a much higher standard deviation. Moreover, the relatively high unimodal water
506 retention parameter α would also indicate a quite low (in absolute value) air-entry pressure head,
507 which is typical of sandy soils (which is not the case of the investigated soils). This behaviour may
508 be ascribed to a fast infiltration rate observed during the infiltration experiments, which the
509 unimodal model tries to capture by using relatively high values of α and K_0 . The variability of the

510 infiltration rate thus induces a high variability of the unimodal K_0 . By contrast, the bimodal model
511 explicitly includes additional parameters (α_1 and n_1), which allow rapid infiltration to be described
512 as a swift emptying of the structural pores without the need to increase the saturated hydraulic
513 conductivity excessively. Actually, in the bimodal case the variability observed in the infiltration
514 rate is now fulfilled by the relatively high standard deviation of α_1 and n_1 , whereas the variability of
515 the bimodal K_0 remains quite limited. This should open a discussion on the real meaning of the high
516 coefficient of variations generally found in the saturated hydraulic conductivity, especially when
517 arising from inversion procedures, which could come partly from the inadequate model used for
518 describing hydraulic properties in the presence of soil structure.

519

520 3.2. Comparing measured hydraulic properties and *unimAP* and *bimAP* estimates

521 The graphs in Figure 9 compare the measured WRC (solid lines - coming from the inversion of
522 infiltration experiments) to those obtained by both the *unimAP* and *bimAP* (dashed lines) for three
523 of the sites investigated. Figure 10 compares the corresponding HCCs. In both figures, the graphs on
524 the left side show the comparison of unimodal measured and estimated curves, whereas those on
525 the right compare the bimodal measured and estimated curves. All comparisons were carried out in
526 terms of root-mean square error (RMSE).

527 We recall that the AP estimates are obtained from the optimization of a single parameter, namely
528 the α_{AP} scaling parameter. Graphical results show that introducing the aggregate information in the
529 *bimAP* significantly improves the ability of the approach to estimate soil water retention (with
530 average RMSE values of 0.43 and 0.11 for *unimAP* and *bimAP*, respectively). Even more importantly,
531 the WRC parameters obtained under the *bimAP* (the Durner parameters) and K_0 from the Kozeny-
532 Carman model significantly improve hydraulic conductivity (with average RMSE values of 0.315
533 and 0.28 cm/min for *unimAP* and *bimAP*, respectively) predicted by applying the Mualem model.

534 The substantial enhancement of the K_0 estimates is apparent in Figure 11, showing a comparison of

535 saturated hydraulic conductivity as measured and obtained by *unimAP* (empty triangles) and
536 *bimAP* (solid triangles), with much smaller RMSE and scattering around the 1:1 line for the *bimAP*
537 case (RMSE = 0.747) compared to *unimAP* (RMSE = 12.580).
538 The third and fourth columns in Table 1 show the average and the standard deviation of the
539 hydraulic property parameters obtained from the AP PTF estimates (subscript *PTF*). It is worth
540 noting in the table that the *unimAP* is unable to obtain either the average value of the saturated
541 hydraulic conductivity, or its variability. As discussed by Coppola et al. (2009a, 2009b), PTFs tend
542 to flatten the variability generally found in measured hydraulic properties. This is mostly due to the
543 information on the structure being overlooked, as well as using unimodal models to describe the
544 hydraulic properties of structured soils, when developing PTFs. This is clearly demonstrated by the
545 parameter values for the *bimAP* approach given in the table, showing that accounting explicitly for
546 the structure in developing PTFs allows much better estimates of K_0 and, importantly, its variability.
547 The same may be said for all the parameters describing both the textural and structural parts of the
548 water retention curve (see the average and standard deviations for n_1 , α_2 and n_2 in table 1).

549

550 Figure 9

551

552

553 Figure 10

554

555

556 Figure 11

557

558 3.3. Relationship between the scaling parameter, α_{AP} , and soil physical properties

559 Multiple linear regression was used to evaluate the degree of dependence of the scaling factor α_{AP}

560 on soil physical parameters. This is especially important in view of using the *bimAP* PTF in soils
561 where no references to measured hydraulic properties are available. Table 2 summarizes the
562 coefficients and the intercepts of soil physical parameters to predict α_{AP} using MLR. The physical
563 parameters used in the regression are: 1) the parameters of the PSD and ASD curves (see section
564 *2.4. Measurements of soil physical parameters*), 2) soil bulk density, and 3) macropore fraction in the
565 sample's overall porosity (β_1).

566 The values in the table show a relatively strong correlation of the scaling parameter with bulk
567 density and the slope of the PSD curve (both in *unimAP* and *bimAP*). However, when the structure is
568 explicitly taken into account (the *bimAP* case), a clear correlation emerges between α_{AP} and the
569 aggregate parameters, namely $T_{\alpha_{ASD}}$ and the fraction of aggregate porosity to total porosity, β_1 ,
570 which cannot be detected when a unimodal approach without structure is considered.

571 Improvement in the correlation with soil structural properties of *bimAP* is also apparent when
572 plotting the α_{AP} values obtained by the MLR against the original values of α_{AP} for both the *unimAP*
573 (white symbols) and *bimAP* (black symbols) approaches (see Figure 12). The RMSE is 0.418 and
574 0.227 for *unimAP* and *bimAP*, respectively. That said, regardless of the better overall description of
575 the curves, *bimAP* always appropriately captures the behaviour close to saturation, which is crucial
576 for predicting hydraulic conductivity and hydraulic property variability.

577

578 Figure 12

579

580 **Conclusions**

581 The main purpose of this paper was to develop a bimodal physically-based PTF to estimate soil
582 hydraulic parameters. The proposed PTF (*bimAP*) is based on the principles of the Arya and Paris
583 (1981) PTF, incorporating aggregate-size distribution to obtain bimodal soil hydraulic parameters.
584 The proposed approach provides bimodal WRCs and HCCs starting from soil physical parameters:

585 PSD, ASD, sample bulk density, and single-aggregate bulk density.

586 Overall, *bimAP* provides better estimates of soil hydraulic parameters and their variability
587 compared to the *unimAP* PTF. K_0 , the whole shape of the HCC, as well as their variability, are better
588 predicted by accounting for soil structure and bimodal porosity in the development of the PTF. In
589 general, the *bimAP* approach produces hydraulic parameter estimates remaining within a more
590 physically plausible region than in the *unimAP* approach. It also enhances the ability of MLR to
591 predict the scaling parameter, α_{AP} .

592 Our results confirm that, in the perspective of PTF calibration for estimating K_0 and, more generally,
593 the hydraulic conductivity function, the relevant information on the bimodal character of the
594 porous medium included in the soil water retention near saturation must be described in detail.
595 Unfortunately, unimodal hydraulic functions are unable to describe the transition between pore
596 systems frequently indicated by the retention data in aggregated soils. Consequently, if a unimodal
597 water retention function is used to fit measured retention data with a bimodal behaviour and then
598 to calibrate PTF parameters, a poor performance of the PTFs is expected when used to estimate
599 hydraulic conductivities.

600 From our data set, it may be observed that by introducing bimodality excellent AP estimates can be
601 obtained for aggregated soils. Owing to the flexibility arising from the structural-matrix partition
602 specifically built into the modified AP retention model, the *bimAP* estimates keep the fundamental
603 information on soil aggregation in the measured soil water retention within the range of soil water
604 potential near saturation, thus providing accurate predictions of pore size distribution and hence of
605 the hydraulic conductivity curve.

606 Of course, to be effectively and reliably applied the bimodal approaches always require that the
607 predominant effects of the soil hydrological behaviour near saturation be supported by accurate
608 and detailed experimental descriptions of the retention curve and hydraulic conductivity for high
609 water contents, which would allow less uncertain identification of the processes and related

610 parameters involved.

611

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615

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748

749 **Tables**

750 Table 1. Mean (μ) and standard deviations (σ) of measured (subscript *meas*) and AP PTF (subscript
 751 *PTF*) hydraulic parameters obtained from unimodal and bimodal inverse models. The scaling
 752 parameter, α_{AP} , is also reported only for the PTF case

Parameter	μ_{meas}	σ_{meas}	μ_{PTF}	σ_{PTF}	Scenario
α	0.137	0.129	0.192	0.439	Unimodal
n	1.476	0.431	1.273	0.187	
K_0 (cm/min)	2.367	15.062	0.171	0.193	
α_{AP}	-	-	1.269	0.264	
α_1	0.590	1.055	2.249	2.204	Bimodal
n_1	2.569	2.160	3.127	2.522	
α_2	0.049	0.042	0.042	0.126	
n_2	1.496	0.423	1.539	0.126	
K_0 (cm/min)	0.266	0.915	0.206	0.434	
α_{AP}	-	-	1.156	0.311	

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754

755 Table 2: Results of MLR application to predict the α_{AP} parameter from soil physical parameters.

756 Subscript PSD and ASD stand for particle and aggregate size distribution, respectively

MLR Parameter	Coefficients	Lower 95%	Upper 95%	Scenario
Intercept	9.939	3.031	16.847	
ρ_b (g/cm ³)	0.480	-0.104	1.063	
$T_{s,PSD}$	0.007	-0.043	0.057	Unimodal
$T_{\alpha,PSD}$	0.002	-0.003	0.007	
$T_{n,PSD}$	-11.835	-20.893	-2.777	
Intercept	-5.398	-15.139	3.105	
ρ_b (g/cm ³)	0.684	-0.056	1.424	
$T_{s,PSD}$	-0.028	-0.107	0.051	
$T_{\alpha,PSD}$	0.000	-0.006	0.007	
$T_{n,PSD}$	9.902	-0.751	20.554	Bimodal
$T_{s,ASD}$	-0.009	-0.034	0.016	
$T_{\alpha,ASD}$	0.934	-1.430	3.298	
$T_{n,ASD}$	0.006	-0.025	0.037	
β_1	-0.619	-1.552	2.790	

757

Abstract: The main purpose of this paper is to develop a bimodal pedotransfer function to obtain soil water retention (WRC) and hydraulic conductivity (HCC) curves. The proposed pedo-transfer function (PTF) extends the Arya and Paris (AP) approach, which is based on particle size distribution (PSD), by incorporating aggregate-size distribution (ASD) into the PTF to obtain the bimodal WRC. A bimodal porosity approach was developed to quantify the fraction of each of the porous systems (matrix and macropores) in overall soil porosity. Saturated hydraulic conductivity, K_0 , was obtained from WRC using the Kozeny-Carman equation, whose parameters were inferred from the behaviour of the bimodal WRC close to saturation. Finally, the Mualem model was applied to obtain the HCC. In order to calibrate the PTF, measured soil physical and hydraulic properties data were used, coming from field infiltration experiments from an irrigation sector of 140 ha area in the “Sinistra Ofanto” irrigation system in Apulia, southern Italy. The infiltration data were fitted by using both bimodal and unimodal hydraulic properties by an inverse solution of the Richards equation. The bimodal “measured” hydraulic properties were then used to calibrate the scaling parameter (α_{AP}) of the proposed bimodal AP (*bimAP*) PTF. Similarly, for the sake of comparison with the bimodal results, the unimodal hydraulic properties were used to calibrate the α_{AP} of the classical unimodal AP (*unimAP*) PTF. Compared to the *unimAP* PTF, the proposed *bimAP* significantly improves the predictions of the mean WRC parameters and K_0 , as well as the prediction of the shape of the whole HCC. Moreover, compared to the unimodal approach, it also allows keeping the hydraulic parameters’ spatial variability observed in the calibration dataset. Multiple linear regression (MLR) was also applied to analyse the sensitivity of the bimodal α_{AP} parameter to textural and structural features, confirming significant predictive effects of soil structure.

Highlights

- Bimodal Arya&Paris hydraulic properties were developed based on soil structure
- The bimodal PTF provides better prediction of un/saturated hydraulic conductivity
- The bimodal PTF keeps the spatial variability of the original hydraulic properties

Declaration of interests

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:

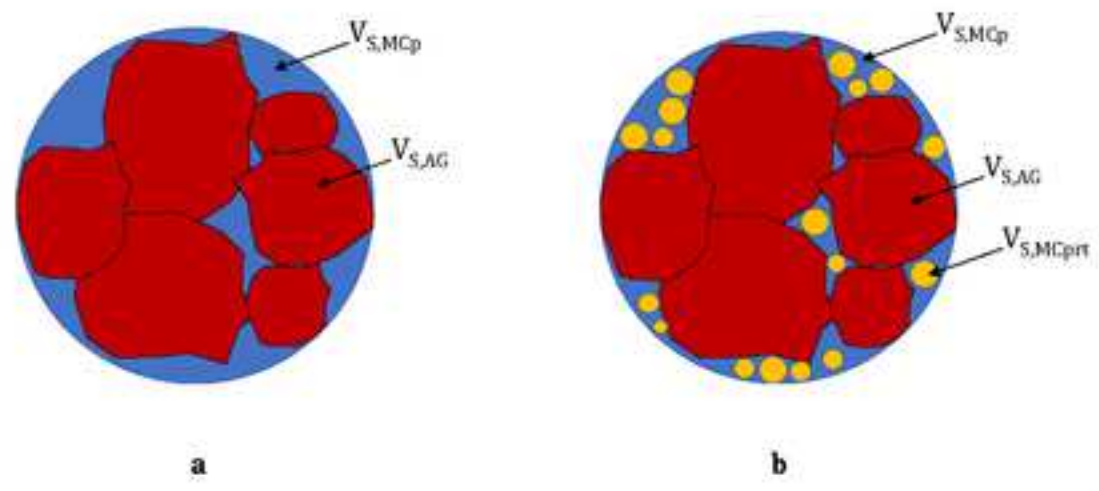
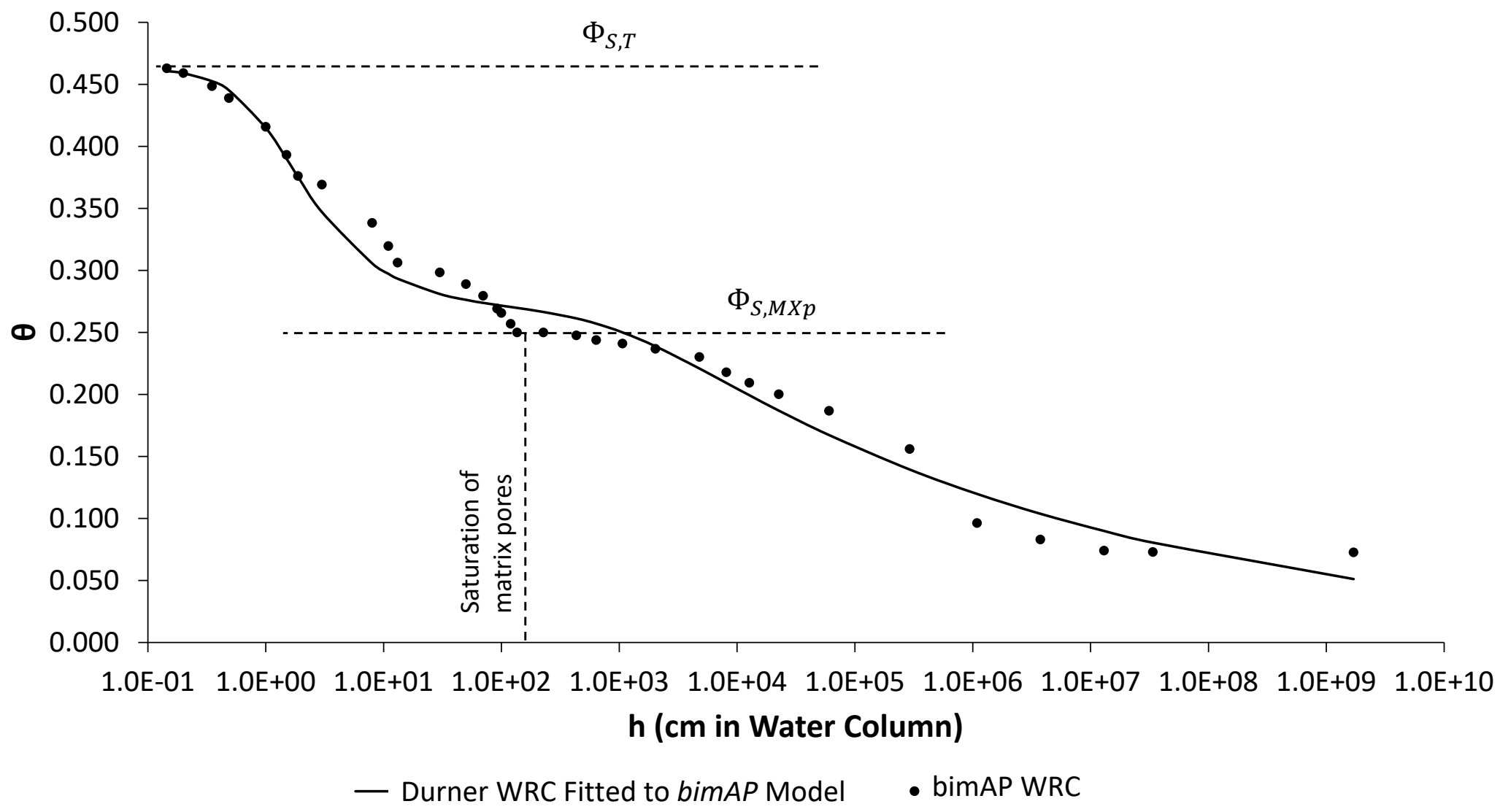
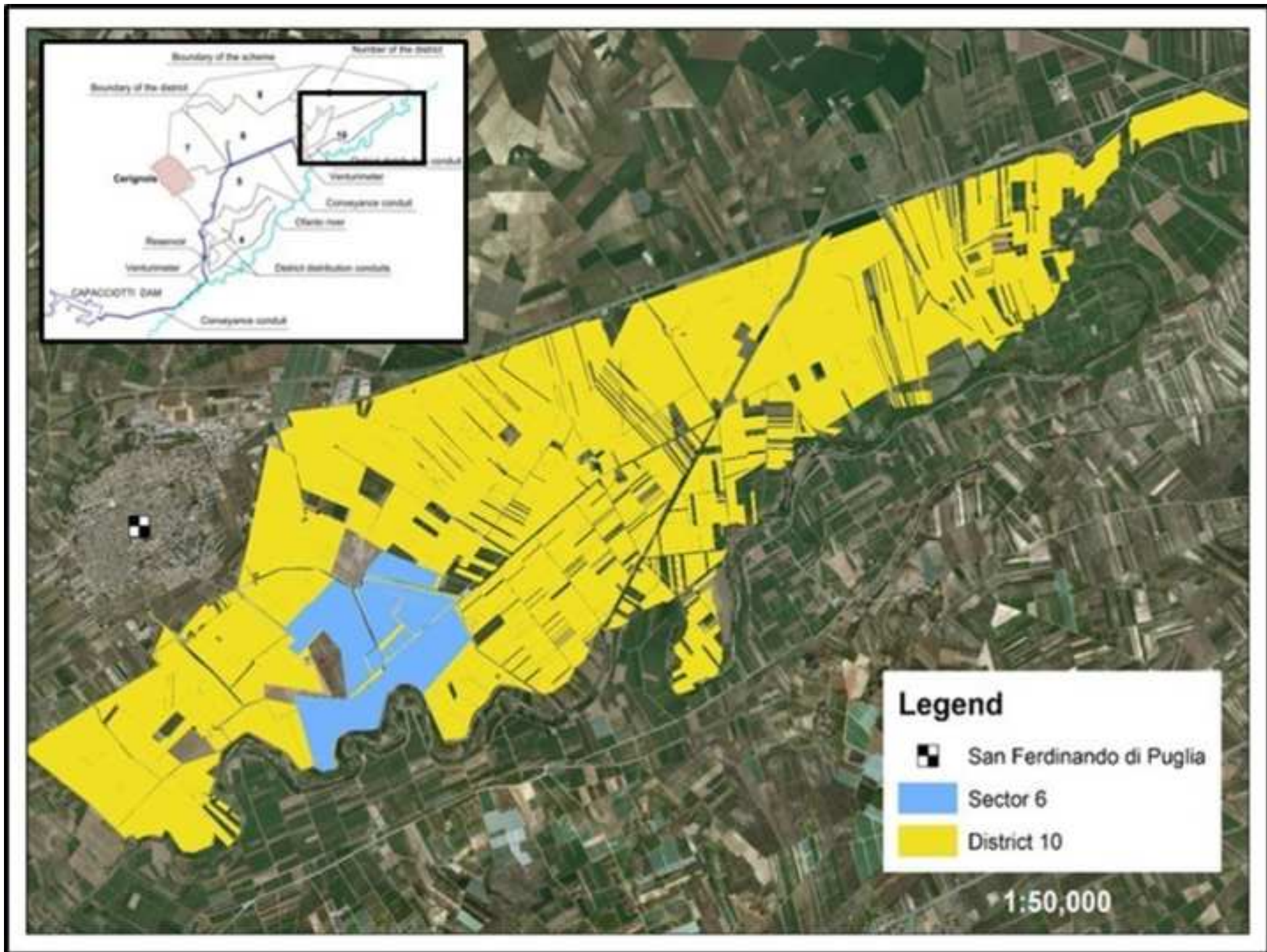


Figure 2





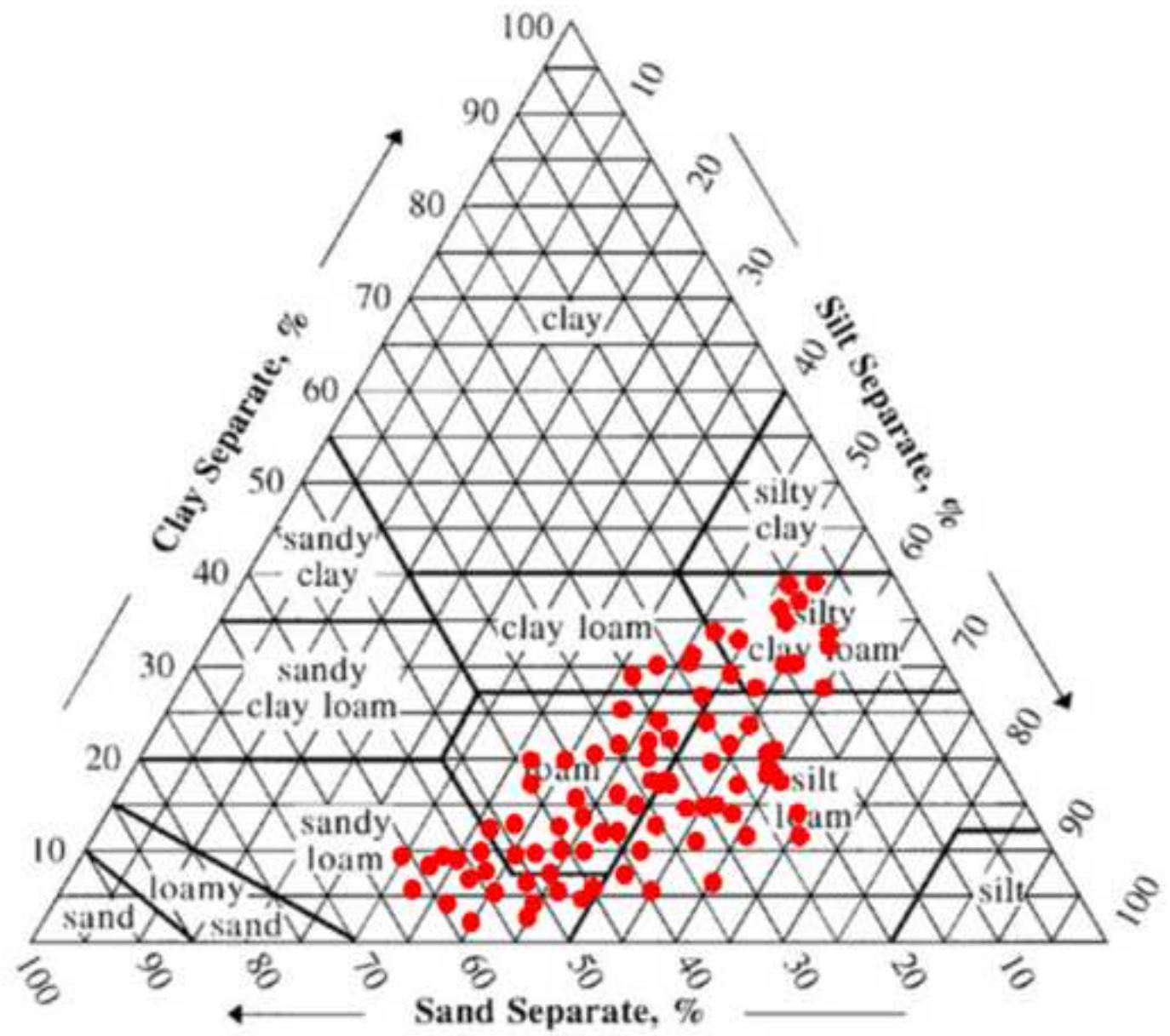


Figure 5

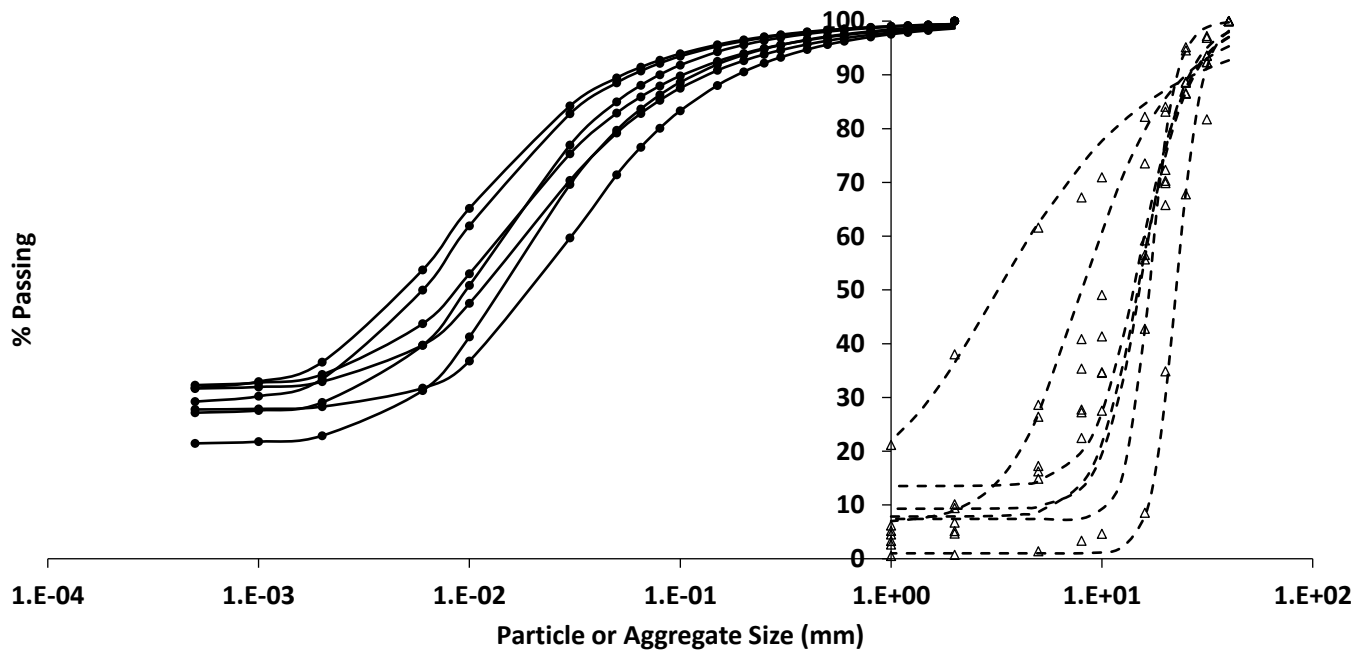


Figure 7

Unimodal Inverse Solution

Bimodal Inverse Solution

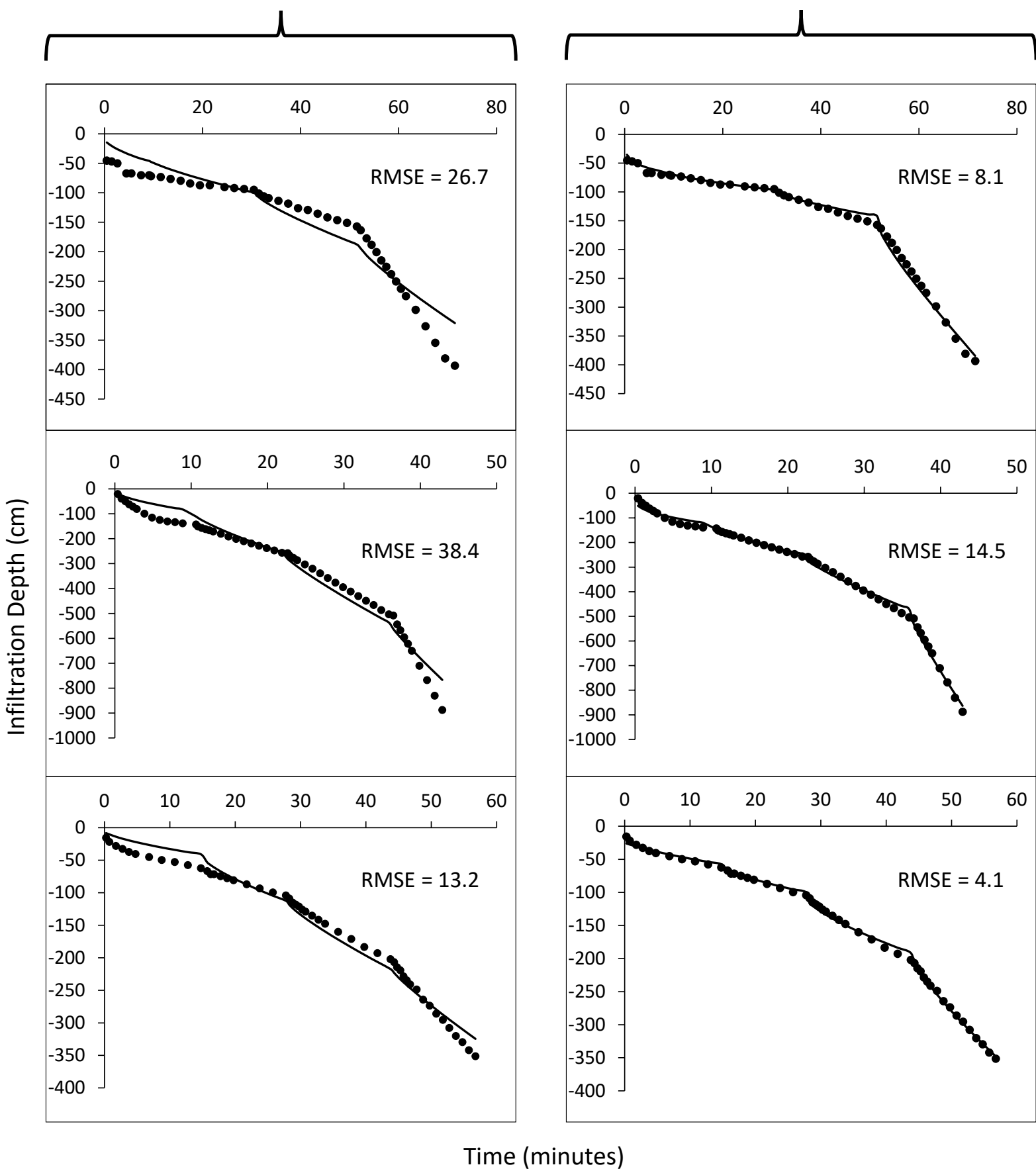


Figure 8

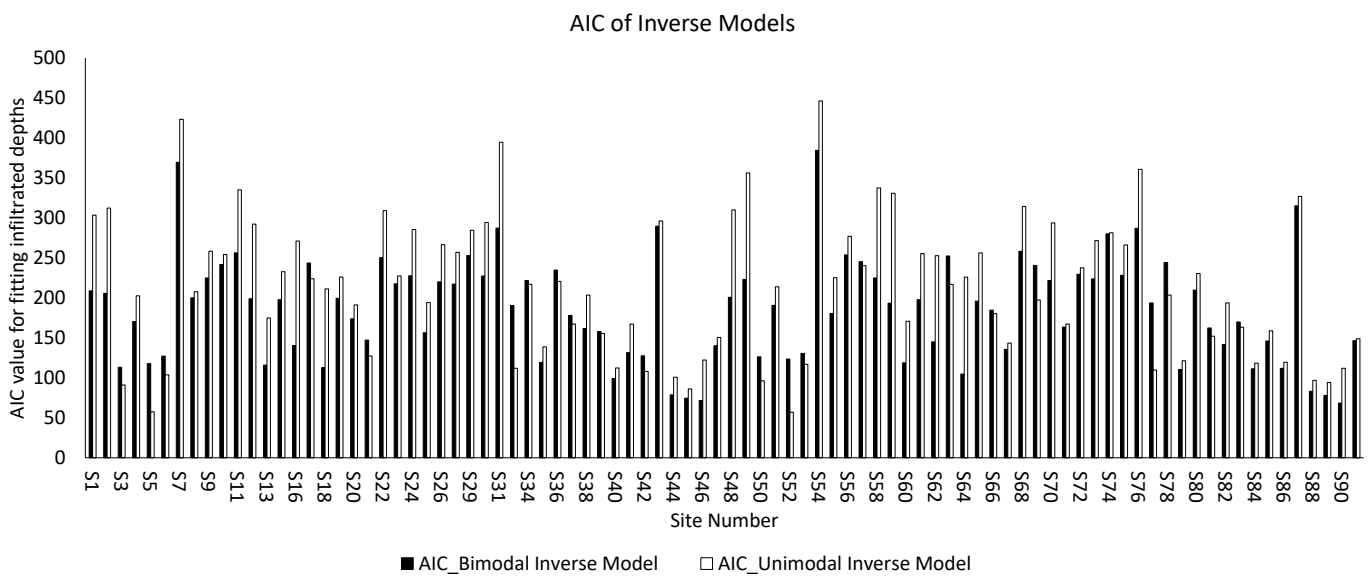


Figure 9

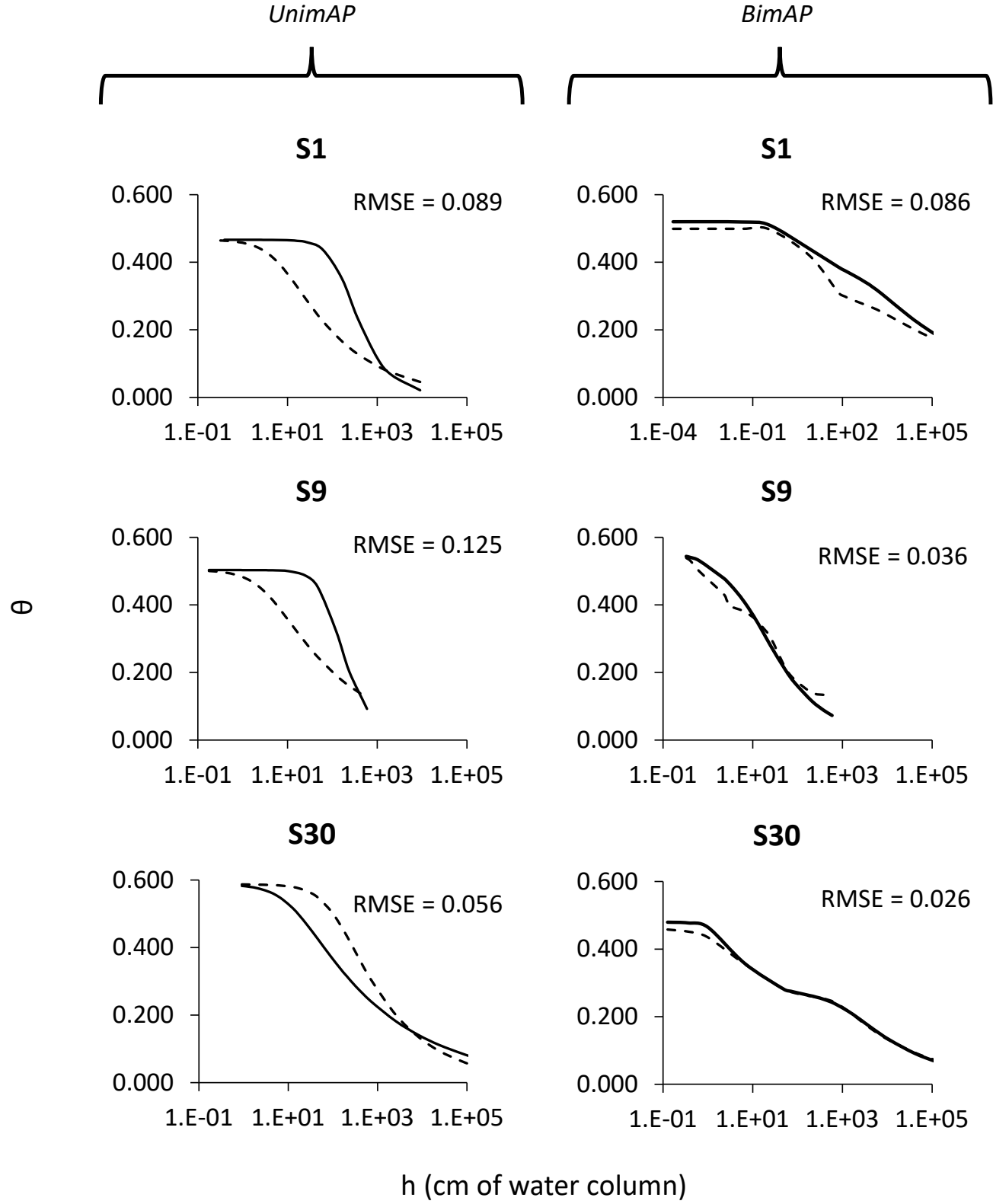


Figure 10

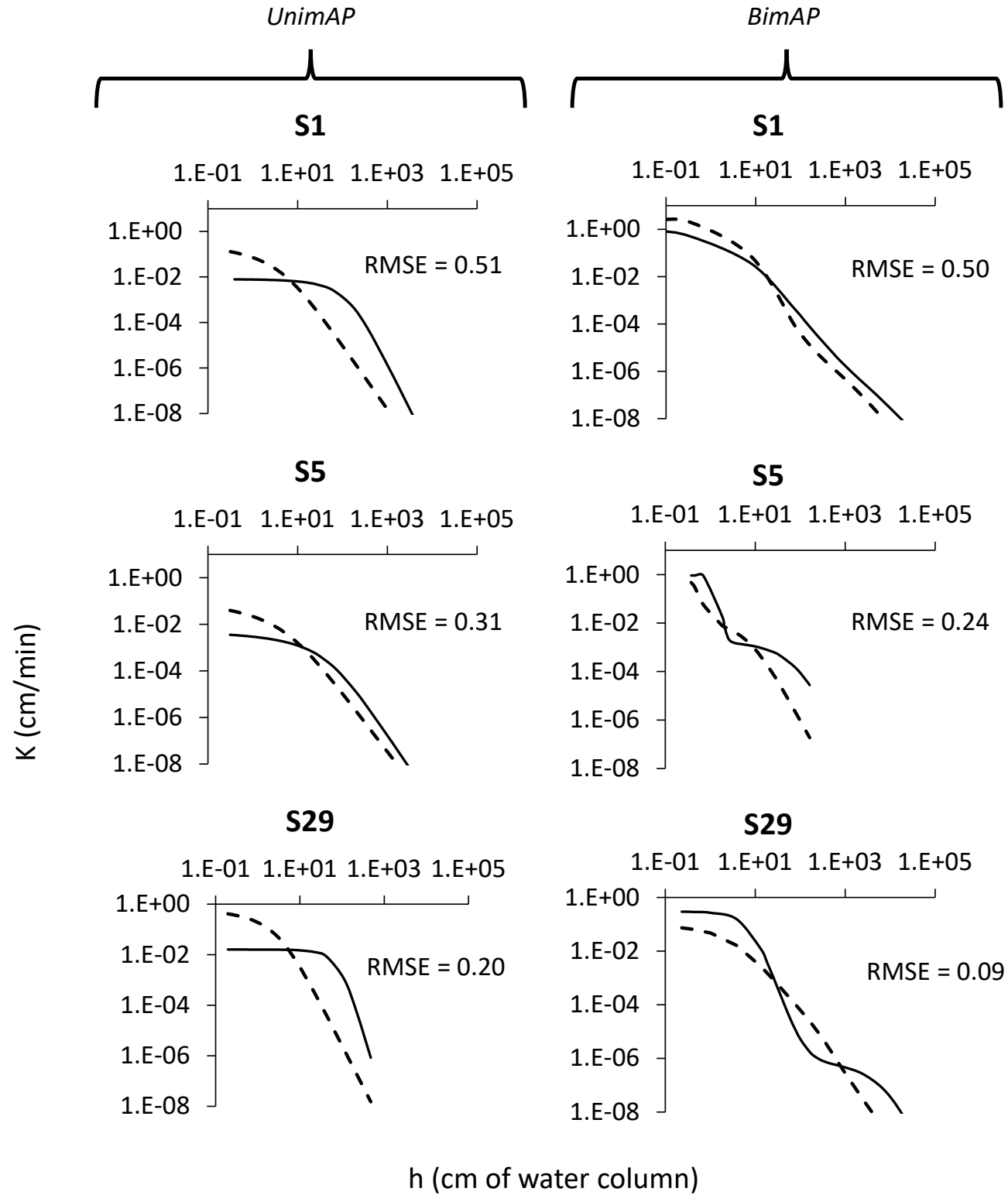


Figure 11

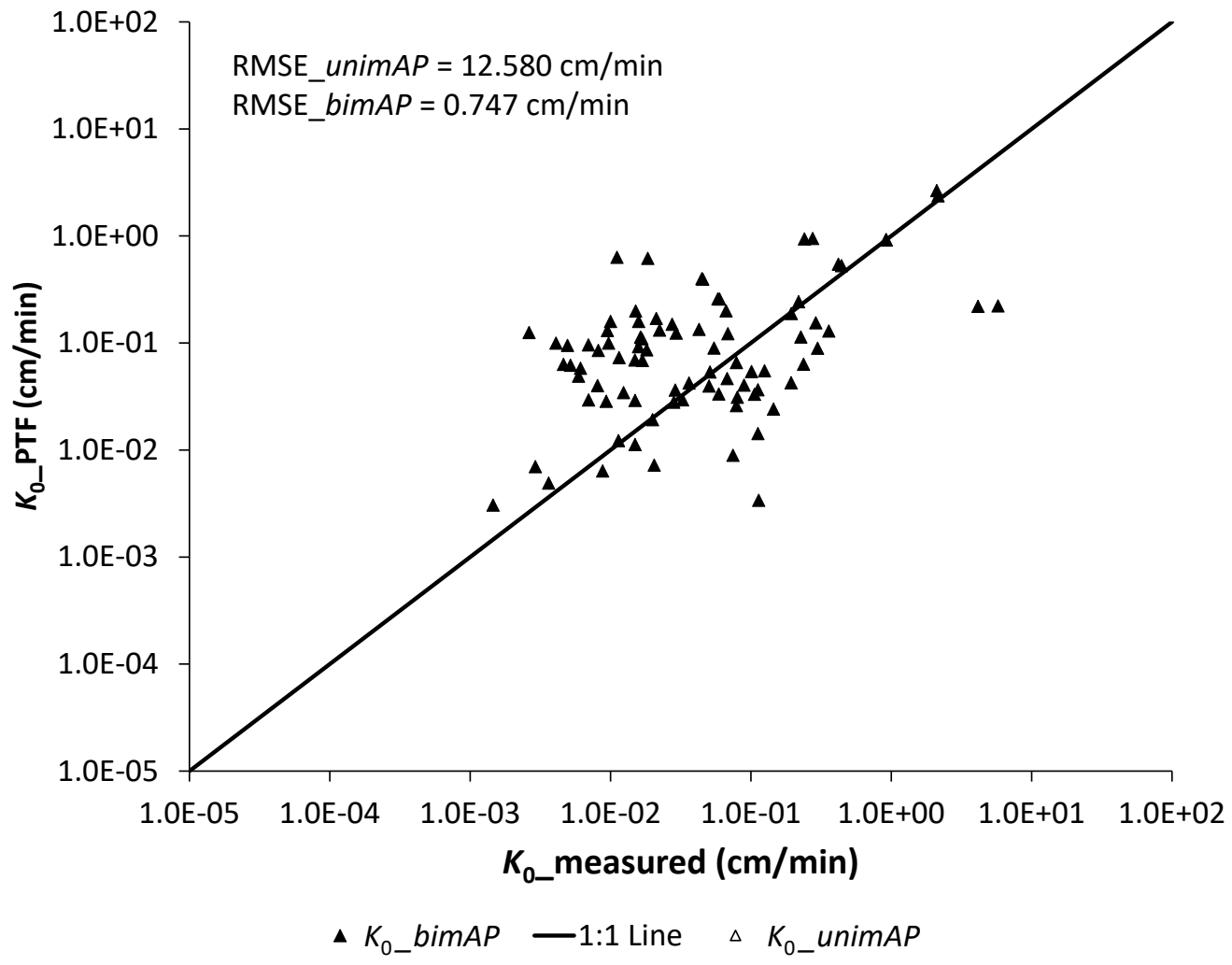


Figure 12

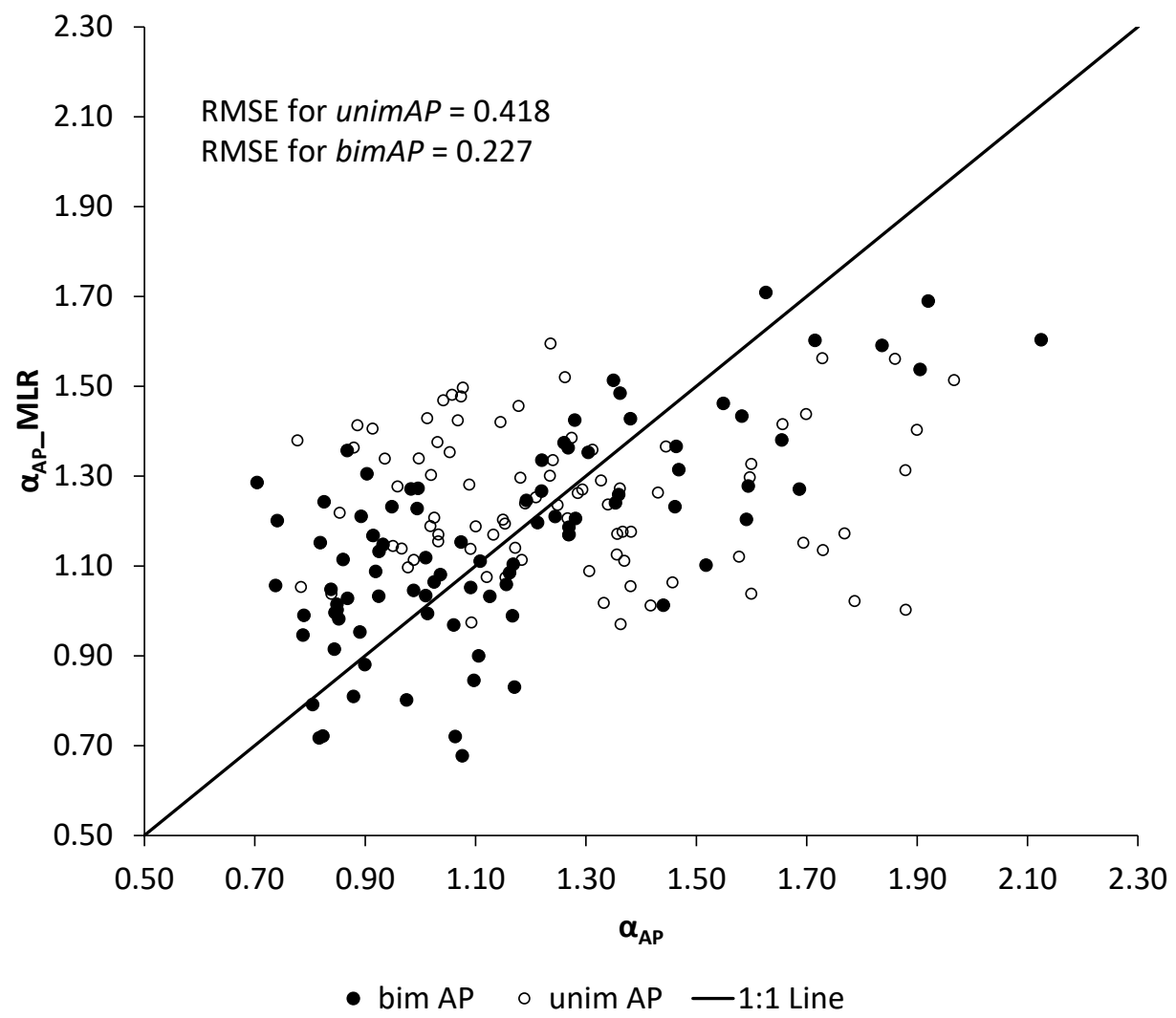


Figure 1. Schematic view of an undisturbed sample consisting: a) only of aggregates, without particle inclusions in the interspace among the aggregates; b) of aggregates with particle inclusions (orange circles) in the interspace among the aggregates

Figure 2. The bimAP WRC (symbols) and the measured bimodal WRC (solid line). The structural and textural parts of the WRC are clearly visible in both the curves. The two horizontal dashed lines indicate the porosity of the structural and textural regions of the bimAP WRC

Figure 3. The study area, the sector 6 of the district 10 in “Sinistra Ofanto” irrigation system

Figure 4. USDA textures of the 90 soil samples considered in this paper

Figure 5. Measured and fitted PSDs and ASDs for 7 of the samples used in the study. The divide between PSDs and ASDs is at 1 mm size. The symbols represent the measured data while the curves represent the curves fitted to equation (30).

Figure 6. Schematic view of steps followed to develop *bimAP* WRCs, HCCs, and predict *bimAP* scaling parameters (α_{AP})

Figure 7. Inverse solution results for three 3 locations in the study area. The symbols and solid lines represent the observed infiltration depths and the infiltration depths obtained from the inverse solution using DISC for the unimodal case (plots on the left) and HYDRUS 3D for the bimodal case (plots on the right)

Figure 8. Akaike’s Information Criterion (AIC) value resulting from fitting measured infiltrated depths to Richard’s infiltration model using the unimodal van Genuchten model (blank bars) and the bimodal Durner model (black bars) in all the ninety sites.

Figure 9. Comparison of the measured WRCs (solid lines - coming from the inversion of infiltration

experiments) to those obtained by both the *unimAP* and the *bimAP* (dashed lines) for three of the sites investigated. The graphs on the left side compare unimodal measured and estimated WRCs, those on the right side compare bimodal measured and estimated WRCs.

Figure 10. Comparison of the measured HCCs (solid lines - coming from the inversion of infiltration experiments) to those obtained by both the *unimAP* and the *bimAP* (dashed lines) for three of the sites investigated. The graphs on the left side compare unimodal measured and estimated HCCs, those on the right side compare bimodal measured and estimated HCCs.

Figure 11. K_0 obtained from *bimAP* and *unimAP* plotted against K_0 obtained from bimodal and unimodal inverse solutions respectively. The empty triangles represent the unimodal scenario and solid triangles represent the bimodal scenario. The solid line is a 1:1 line.

Figure 12. The α_{AP} values obtained by the MLR against the original values of α_{AP} for both the *unimAP* (white symbols) and *bimAP* (black symbols) approaches.

Table 1. Mean (μ) and standard deviations (σ) of measured (subscript *meas*) and AP PTF (subscript *PTF*) hydraulic parameters obtained from unimodal and bimodal inverse models. The scaling parameter, α_{AP} , is also reported only for the PTF case

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