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# Phenol red dye removal from wastewater using TiO<sub>2</sub>-FSM-16 and Ni-FSM-16 photocatalysts

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#### ABSTRACT

In this study, the performance of Ni-FSM-16 and TiO<sub>2</sub>-FSM-16 photocatalysts in phenol red removal was explored. The XRD, FE-SEM, and BET tests were used to characterize the catalysts. All experiments were performed at ambient temperature and under UV (20 W). The parameters including dye concentration (20–80 mg/L), photocatalyst concentration (0–8 g/L), UV exposure duration, and contact time (0–160 min) were optimized using RSM software. BET values of Ni-FSM-16 and TiO<sub>2</sub>-FSM-16 were 718.63  $m^2/g$  and 844.93  $m^2/g$ , respectively. TiO<sub>2</sub>-FSM-16 showed better performance in dye removal than Ni-FSM-16. At pH 3, the maximum dye removal by TiO<sub>2</sub>-FSM-16/UV and Ni-FSM-16/UV was obtained 87% and 64%, respectively. The positive hole species had the main role in photocatalytic phenol red removal. The reusability study was done for up to 7 cycles, but the catalysts can be reused effectively for up to 3 cycles. The synergistic factor for the TiO<sub>2</sub>-FSM-16 and TiO<sub>2</sub>-FSM-16/UV processes were calculated to be 1.55 and 2.12, respectively. The dye removal efficiency by TiO<sub>2</sub>-carbon and Ni-carbon was slightly lower than those obtained by the FSM-16 ones. The TiO<sub>2</sub>-FSM-16 catalysts had a suitable surface and acceptable efficiency in phenol red removal.

#### 1. Introduction

Dyes are mainly used in the plastics, textile, paper, rubber, wood, and food industries [1]. The presence of colorant materials in streams is not only aesthetically unpleasant but can also cause problems for the environment by elevating toxicity, oxygen demand levels, and lagging the photosynthetic phenomena by reducing light penetration [2,3]. Dyes belonging to the groups based on anthraquinone, azo, and triphenylmethane are the main pollutants released from industrial effluents [4,5]. Most of the dye laden-effluents are produced by the textile industry, where 10–15% of the dye is wasted during the production process [4,6]. One of the dyes released from the textile is phenol red. Its solubility in water and ethanol is 0.77 g/L and 2.9 g/L, respectively [7]. Phenol red dye is water-soluble and can be utilized as a pH indicator in various laboratories [8,9]. This dye is used in industries such as

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pharmaceuticals, textiles, and chemicals, and the wastewater discharged from these industries must be properly treated [9].

Various physical, chemical, and biological methods have been tested to eliminate dyes. Physical and chemical methods can be used on a small scale [10]. Advanced oxidation methods (AOMs) have been considered by researchers for cleaning effluents containing resistant organic compounds [11,12]. AOMs are based on the production of active species for degrading pollutants using solar, chemical, or other forms of energy [13,14]. The photocatalytic oxidation process is a superficial phenomenon and light reactions happen on the catalyst surface like TiO<sub>2</sub> [11,15]. Titanium dioxide (TiO<sub>2</sub>) has so far been considered comparable to other semiconductors for many reasons, including its cheapness, availability, safety, and good reactivity [11,16]. Also, it can degrade organic compounds and as a result, the pollutants are converted to  $H_2O$  and  $CO_2$  [16]. Nickel is also a useful photocatalyst that has low cost and good stability in alkaline solutions [17]. However, it is difficult to use titanium dioxide and nickel alone because their separating from the solution is a time-consuming and costly process. Therefore, the photocatalyst should be stabilized on a support surface. Many supporting surfaces like activated carbon [18], graphene oxide [19], silicate [20], and glass spheres [21] have been used for titanium dioxide and nickel oxide.

The FSM-16 powder has outstanding features like high thermal stability, high specific surface area, and high pollutant sorption capacity, which makes it a suitable option for photocatalyst support. Although FSM-16 has been used to support photocatalysts to decompose some pollutants [22,23], its application as support of titanium dioxide and nickel oxide in dye removal is not yet clear.

Accordingly, the main aim of this study was to evaluate the photocatalytic activity of  $TiO_2$ -FSM-16 and Ni-FSM-16 in phenol red degradation. Phenol red decomposition using  $TiO_2$ -FSM-16 and Ni-FSM-16 was examined in the presence of UV rays and the effect of operating parameters was studied. The characteristics of photocatalysts, reaction kinetics, catalyst reusability, leaching of the photocatalyst components, and synergy of process components have also been investigated. In another part of the work, the dye removal efficiency of  $TiO_2$ -FSM-16 and Ni-FSM-16 was compared with  $TiO_2$ -carbon and Ni-carbon.

#### 2. Experimental

#### 2.1. Materials

Acetyl trimethylammonium bromide ( $C_{19}H_{42}BrN$ , purity >99.9%), hydrochloric acid (HCl, 37 wt%), sodium hydroxide (NaOH, purity >99.9%), silicon dioxide (SiO<sub>2</sub>, purity >99.8%), ammonium titanyl oxalate monohydrate ((NH<sub>4</sub>)<sub>2</sub>TiO(C<sub>2</sub>O<sub>4</sub>)<sub>2</sub>·H<sub>2</sub>O, purity: 98%), ethanol ( $C_{2}H_{6}O$ , purity ≥99.9%), nickel nitrate (Ni(NO<sub>3</sub>)<sub>2</sub>, purity> 99.8%) were purchased from Merck Company. Phenol red ( $C_{19}H_{14}O_5S$ , purity: 98%), *tert*-butanol alcohol (C₄H<sub>10</sub>O, purity ≥99.5%), *p*-benzoquinone (C<sub>6</sub>H<sub>4</sub>(=O), purity >98%), silver nitrate (AgNO<sub>3</sub>, purity ≥99%), and ammonium oxalate ((NH<sub>4</sub>)<sub>2</sub>C<sub>2</sub>O<sub>4</sub>·H<sub>2</sub>O, purity ≥99.9%) were provided from Sigma-Aldrich Company. The stock solution of the phenol red dye was prepared daily in doubly distilled water (DDW). All working solutions were prepared by diluting the stock with DDW.

#### 2.2. Measurements

Morphology of FSM-16, TiO<sub>2</sub>-FSM-16, and Ni-FSM-16 was done using the TeScan-Mira III Scanning Field Emission Microscope (FE-SEM, Czech Republic). Nitrogen absorption-desorption test was performed by BET instrument, Micrometric model ASAP2020 (USA). The XRD pattern was taken by the PANalytical manufacturer, Xpert-pro model (USA) at 20 equal to  $0.5-60^{\circ}$  and CuK $\alpha$  of 1.5406 nm. Phenol red content was determined by UV–vis spectrophotometer, PerkinElmer, Lambda 25 (USA) at  $\lambda_{max}$  423 nm. An example of a calibration curve to determine the concentration of phenol red is presented in Fig. S1 in the Supplementary Information. The amount of dye mineralization was measured by the TOC device (Shimadzu, model VCSH, Japan).

#### 2.3. Synthesis of FSM-16

First, 3.8 g NaOH was added to 100 mL of DDW and 6 g of SiO<sub>2</sub> was poured into the solution to make the SiO<sub>2</sub>/Na<sub>2</sub>O ratio equal to 2. It was then stirred for 3 h under normal conditions at ambient temperature (27 °C) to obtain a uniform solution. To remove the available solvent, the sample was placed in a rotary evaporator balloon. The evaporator temperature was set to 100 °C and the speed was 30 rpm. Approximately 60–70 mL of water was evaporated and resulted in a jelly-like liquid of sodium silicate. Finally, the sample was poured into a porcelain plate and transferred to a furnace to complete the synthesis and calcination process of layered sodium silicate. The furnace was programmed with a temperature rate of 2 °C/min at 700 °C for 6 h. The calcined sample was layered sodium silicate or Kanemite. This material was crushed in a mortar and 4 g of it was poured into 40 mL of DDW and agitated for 3 h at 27 °C. After the preparation of Kanemite paste, which was used as a source of silica in the synthesis of regular mesopores of FSM-16. Through a mechanism, the Kanemite silicate plates were folded and cross-linked to form a three-dimensional structure. To make a wet Kanemite paste, the resulting solution was filtered. The Kanemite was dispersed in 88 mL of n-hexadecyl trimethyl ammonium bromide solution (0.125 mol/L) and stirred for 3 h at 70 °C. In this step, pH should be adjusted between 11.5 and 12.5. After that, the pH was decreased to 8.5 using 2 mol/L HCl. The suspension solution was stirred for 3 h at 70 °C. The produced material was then washed at room temperature with a liter of distilled water and air-dried. The synthesized meso-cavity was placed in the Soxule device for washing. Washing was done with 20 mL of ethanol and 0.5 mL of HCl per gram of material. The meso-cavity material was placed in a Soxule device for 72 h. The material was then dried for 24 h at ambient temperature. Finally, it was placed in the furnace (950 °C, 9 h) to remove the surfactant [24].

#### 2.4. Synthesis of TiO<sub>2</sub>-FSM-16 and Ni-FSM-16

To deposit titanium dioxide on FSM-16, a solution with a concentration of 0.1 M of titanium oxalate ammonium salt was prepared. Then, 2 g of FSM-16 was added to 25 mL of titanium solution. The sample was stirred at room temperature for 24 h. After filtering and washing the samples with DDW and air-drying, the samples were calcined at 450 °C for 12 h [25].

To make the Ni-FSM-16 catalyst, initially, 1 g of FSM-16 mesoporous silica was dissolved in 20 mL of normal hexane and stirred at room temperature for 2 h (Solution A). The amount of 0.5675 g of nickel nitrate was added to another vessel with 1.14 mL of DDW (Solution B). After 2 h, Solution B was added dropwise to Solution A and the new mixture was stirred for 3 h. It was then air-dried for 24 h. The solids were subjected to argon gas for calcination at 550 °C for 5 h. The gradient for increasing and decreasing the temperature at this stage was 2.5 °C/min and the flow rate of argon gas injection into the furnace was 2 L/h [26].

#### 2.5. Synthesis of TiO2-carbon and Ni-carbon

To make TiO<sub>2</sub>-carbon and Ni-carbon, the same methods as TiO<sub>2</sub>-FSM-16 and Ni-FSM-16 were done. It should be noted that activated carbon in this study was purchased from Merck and according to the supplier's information, it had an active surface of 810  $m^2/g$ .



Fig. 1. FE-SEM images of FSM-16 (A) 200 nm (B) 10 µm, TiO<sub>2</sub>-FSM-16 (C) 200 nm (D) 10 µm, and Ni- FSM -16 (E) 500 nm (F) 10 µm.

#### 2.6. Design of experiments using a response surface methodology (RSM)

Four factors were analyzed by the response surface method. The effect of four experimental variables, namely TiO<sub>2</sub>-FSM-16 and Ni-FSM-16 quantity, solution pH, UV irradiation time, and dye concentration on the photocatalytic removal of phenol red was explored. The variables and their upper and lower levels are listed in Table S1 in the Supplementary Information.

A total of 30 tests were performed, including 16 experiments at factorial points, 8 experiments at axial points, and 6 replications at central points.

The efficiency of dye removal was calculated using the following equation:

$$Dye removal (\%) = \frac{\left( [Dye]_{initial} - [Dye]_{residual} \right)}{[Dye]_{initial}} \times 100$$
(1)

#### 2.7. Recovery test

The photocatalysts (TiO<sub>2</sub>-FSM-16 and Ni-FSM-16) were recovered after being used in dye removal. The photocatalysts were separated using a centrifuge, washed (with water and ethanol), dehydrated (at 105  $^{\circ}$ C), and then reused for dye degradation. The reusability tests were done 7 times.

#### 2.8. Effect of scavengers

Compounds of *tert*-butanol (500 mM), *p*-benzoquinone (10 mM), silver nitrate (10 mM), and ammonium oxalate (50 mM) were used as scavengers of active species and radicals in the process of phenol red removal [27]. This test was performed at optimal conditions (pH: 3, dye concentration: 20 mg/L, catalyst dose: 2 g/L, time: 120 min).

#### 3. Results and discussion

### 3.1. Main properties of photocatalysts

The micrographs (FESEM) of the FSM-16 are presented in Fig. 1A, B. Based on the FESEM graphs, nanoparticles with an approximate size of 200 nm were observed. In the study conducted by Hashemi et al., the FSM-16 nanoparticles have a spherical shape with a size of 100 nm [28]. The FESEM images of TiO<sub>2</sub>-FSM-16 are shown in Fig. 1C,D. As this figure shows, the photocatalyst particles are non-uniform and spherical. TiO<sub>2</sub>-FSM-16 particles have a similar shape to those reported in other works [22]. This particle size distribution with spherical morphology corresponded to van der Waals forces [29]. To decrease the surface energy, the primary particles tend to condense, with the formation of spherical masses, in the minimum surface-to-volume ratio, the minimum free surface energy can be obtained. The TiO<sub>2</sub>-FSM-16 composite powder is composed of nanoscale particles, which indicates that the prepared powder has a large specific surface area and volume. Therefore, the TiO<sub>2</sub>-FSM-16 composite can provide suitable active sites for photocatalysis. Fig. 1C,D shows that TiO<sub>2</sub> particles are dispersed on the surface of SiO<sub>2</sub> nanoparticles and have good stability.

The shape and size of crystals in Ni-FSM-16 are slightly different from  $TiO_2$ -FSM-16 (Fig. 1E,F). The Ni-FSM-16 particles have a small size and are aggregates of spherical microcrystal particles. Silicate plates are a combination of tens to hundreds of hexagonal cavities composed of Ni-FSM-16 catalyst particles. Adhesion between particles may be due to the small magnetism or the polymer between them [30,31].

The XRD image of FSM-16, TiO<sub>2</sub>-FSM-16, and Ni-FSM-16 is shown in Fig. 2. A wideband with an angle equal to  $20-30^{\circ}$  can be seen in the FSM-16 spectrum, revealing that the material is amorphous and has no specific crystalline peaks. The disappearance of the higher-order diffraction peaks indicates that the hexagonal arrangement of the channels in the TiO<sub>2</sub>-FSM-16 and Ni-FSM-16 mesoporous materials is slightly irregular. Therefore, it can be concluded that the addition of metal in the mesoporous structure of FSM-16 has a negative effect on the crystallinity of the material. This is because the use of metal ores weakens the self-assembly process and produces a less regular meso-cavity structure. In the XRD pattern of TiO<sub>2</sub>-FSM-16 and Ni-FSM-16, in addition to the four distinct peaks of the FSM-16 nanoparticles, the characteristic peaks of TiO<sub>2</sub> and NiO are also seen, indicating no noticeable change in the crystal



Fig. 2. X-ray diffraction pattern of FSM-16, TiO<sub>2</sub>-FSM-16, and Ni-FSM-16.

structure of the photocatalysts [32]. The low-angle XRD image of FSM-16 showed that the diffraction plate has a regular hexagonal structure [33]. An anatase-like reference pattern (JCPDS # 21–1272) was seen in TiO<sub>2</sub>-FSM-16 [34]. Five peaks at 37, 43, 48, 52, and 58 in Ni-FSM-16 photocatalyst are related to Miller indices of (111), (200), (220), (311), and (222), respectively, which confirms that the photocatalyst contains NiO [35–37]. Based on the literature [38], a main reflection at 2theta of 43.5°, corresponds to the (200) plane of cubic NiO (PDF-2, 01–071–1179).

The active area of the photocatalysts was computed using Brunauer-Emmett-Teller (BET) technique and the pore diameter and pore volume were calculated by the BJH technique (Fig. 3a–c). For the FSM-16 sample, the obtained BET was 1099.08 m<sup>2</sup>/g, and the absorption isotherm is almost synchronous with the desorption branch, indicating the mesoporous structure of the sample. Fig. 3b and c shows a similar pattern of BET for TiO<sub>2</sub>-FSM-16 (844.93 m<sup>2</sup>/g) and Ni-FSM-16 (718.63 m<sup>2</sup>/g), showing that the mesoporous feature of FSM-16 did not alter after the composition with Ni and TiO<sub>2</sub> [39].

The specific surface area of Ni-FSM-16 and TiO<sub>2</sub>-FSM-16 was lower than that of FSM-16, which is linked to the occupation of pores by Ni and TiO<sub>2</sub> (Fig. 3). However, the level of synthesized photocatalyst (844.93  $m^2/g$ ) is much bigger than the values stated for other photocatalysts in the literature. For example, Fatima and Supia prepared the TiO<sub>2</sub>-MCM-41 photocatalyst with a surface of 400.7  $m^2/g$  [40]. Sugiyama et al. have incorporated Ni into FSM-16 and MCM-41 structures to produce Ni-FSM-16 and Ni-MCM-41 and reported that the active surface of photocatalysts was lower than that of the base material [41].

#### 3.2. Interpretation of ANOVA and interaction of variables

ANOVA was utilized to determine the relationship between the removal rate and the variables. It is noteworthy that the judgment was based on the F-values and P-values for phenol red removal. The F-values of the model for  $TiO_2$ -FSM-16 and Ni-FSM-16 were obtained at 16.77 and 12.16, respectively, indicating that the models are significant. The P-value was <0.05 indicating the significance



Fig. 3. BET graphs for (a) FSM-16, (b) TiO<sub>2</sub>-FSM-16, and (c) Ni-FSM-16.

of the model expression. In this case, the 'A' variable in  $TiO_2$ -FSM-16 and 'A, B, and D' in Ni-FSM-16 were significant. In addition, the validity of the design with values of P and high value of correlation coefficients as well as the non-significance of "Lack of Fit" is confirmed [42]. P < 0.0001 reveals the importance of the model and the interaction of variables on phenol red elimination. The pH factor played an important role in phenol red removal using both photocatalysts (TiO\_2-FSM-16 and Ni-FSM-16) with the biggest F-value (60.56 and 133.57, respectively) compared to other variables (Table S2 and S3). To evaluate the quality of the models, the figures for the predicted response *vs* the actual values and the normal data distribution diagrams are shown in Fig. S2. The normal diagram of the residues is provided in Fig. S2a,c. Such a graph is very useful for optimizing complex systems such as multivariate optimization. The points are in a straight line and no deviations are seen in the distribution of data, indicating a suitable correlation and distribution between the values [43]. Since the specific trend is not related to variance changes (decrease or increase), the variance is fixed, which shows the scatter of points against the given values (Fig. S2b,d).

Investigation of the effect of the desired catalysts (TiO<sub>2</sub>-FSM-16 and Ni-FSM-16) on phenol red degradation efficiency under ultraviolet radiation is depicted in Fig. S3a. In this figure, the positive and negative effects and the magnitude of the effect of each variable on the response are identified. The sharpness of the slope in a factor indicates that it is a vital variable in the reaction [5]. Oppositely, a relatively flat line shows the insensitivity of the response to alteration in the given variable. As can be seen in Fig. S3a, the TiO<sub>2</sub>-FSM-16 curve shows the time and dye concentration of the slow curvature, showing that these factors have a small effect on the response. Accordingly, the significant sloping curvature at pH indicates that the phenol red removal was sensitive to this variable. Fig. S3b shows that the curvature time and dye concentration have a slow slope, indicating that the mentioned factors have a small impact on the response.





Fig. 4. 3D and 2D diagrams of the interactions of pH-photocatalyst dose on the phenol red removal in the presence of (a,b) TiO<sub>2</sub>- FSM-16 and (c,d) Ni-FSM-16.

#### 3.3. Effect of pH-photocatalyst dose

In Fig. 4a–d, the response surface (3D surfaces) and contour diagrams (2D contours) are depicted as a function of pH and photocatalyst dose. For both generated materials (TiO<sub>2</sub>-FSM-16 and Ni-FSM-16), while the irradiation time (120 min) and the phenol red concentration (20 mg/L) were constant, the amount of degradation decreases with increasing pH. In other words, for both catalysts, the highest efficiency occurred at pH 3. At acidic pHs, the phenol red molecule has the zwitterion form with two functional groups (sulfate with a negative charge and ketone with a positive charge) with an additional proton. The FSM-16-TiO<sub>2</sub> surface also has protonated functional groups at low pH. The dye molecule has positive and negative charges at low pH. Therefore, the maximum efficiency at the acidic pH of 3 is due to the electrostatic attraction between protonated groups and zwitterion negative charges [44]. If the pH of the solution is higher than the pKa of phenol red (about 7.9), the proton of the ketone group is lost and the molecule takes on a negative charge [45]. The higher the pH, the more negative the catalyst level becomes (pHzpc of both photocatalysts was about 5.4). In such a situation, the dye molecule is rejected from the surface of the catalysts, and as a result, the removal efficiency decreases. Kumar et al. [46] reported that at acidic pHs, the surface of titanium dioxide becomes positive, leading to higher dye removal.

With the increasing load of TiO<sub>2</sub>-FSM-16, a slight change in removal efficiency was observed. This has been linked to the tendency of particles to accumulate, which reduces the BET area of the catalyst and leads to less production of active radicals [47,48]. For the Ni-FSM-16, increasing the photocatalyst mass did not play a role in increasing the efficiency and the graph was smooth. Higher content



**Fig. 5.** 3D and 2D diagrams of the interactions of time-photocatalyst dose on the phenol red removal in the presence of **(a,b)** TiO<sub>2</sub>- FSM-16 and **(c, d)** Ni-FSM-16.

of TiO<sub>2</sub>-FSM-16 than Ni-FSM-16 can lead to more radical production and thus higher efficiency [25]. Similar observations have been reported by researchers for degrading cephalexin using NiS and NiS-support  $Fe_3O_4$ @PPY photocatalysts [49].

# 3.4. Effect of time-photocatalysts dose

The effect of treatment time and photocatalyst dose on the phenol red elimination is depicted in Fig. 5a–d. As shown in Fig. 5, increasing the dose of TiO<sub>2</sub>-FSM-16 photocatalyst has resulted in a slow increase in the removal efficiency, while with increasing Ni-FSM-16 the efficiency has remained almost constant. With more time of exposure of TiO<sub>2</sub> nanoparticles to light, the dye removal efficiency increases. Longer exposure to light implies the production of more hydroxyl radicals, which are responsible for the oxidation of the phenol red dye molecule [50].

With increasing filtration time, the amount of dye removal by the TiO<sub>2</sub>-FSM-16 photocatalyst increased due to the sufficient opportunity for radicals to attack the dye. But with increasing time, the removal efficiency by Ni-FSM-16 decreases, which is probably due to the desorption of absorbed dye from the catalyst surface. So, it can be confirmed that the Ni-FSM-16 photocatalyst may not have been able to oxidize the dye but rather remove it by the adsorption process.

#### 3.5. Photocatalyst reusability and leaching Ni and ti

A recyclable photocatalyst would be economically and environmentally beneficial. The photocatalyst recovery results are shown in Fig. 6a. Based on Fig. 6a, the reusability behavior of the two produced materials was different. The  $TiO_2$ -FSM-16 photocatalyst has higher usability and good strength.  $TiO_2$ -FSM-16 catalyst could be reused up to 3 times, while Ni-FSM-16 had up to 2 times good removal efficiency. The reduction in the efficiency of the recycled catalyst can be due to the leakage of its effective components. In these two photocatalysts, nickel and titanium are important components. The results of Fig. 6b show that these elements have leaked from the photocatalyst. Another significant point is that the amount of nickel leakage is lower than the drinking water standard (100  $\mu$ g/L). Among the two prepared catalysts, the amount of leakage in  $TiO_2$ -FSM-16 was lower, which indicates its stability and, as a result, its higher efficiency.

## 3.6. Synergy of system components in the removal of phenol red

In Fig. 7, the effect of system components on the removal efficiency of phenol red dye is compared with the whole system. As depicted in this figure, the efficiency of the system components and even UV along with nickel and titanium dioxides were much lower than the whole system. FSM-16 has been reported to have a 51% removal of phenol red, possibly due to the adsorption mechanism. The high surface of the FSM-16 material has also provided a suitable platform for dye adsorption. The synergy factor for the TiO<sub>2</sub>-FSM-16/UV was calculated based on the following formulas [51]:



Fig. 6. (a) Photocatalysts reusability (pH: 3, dye concentration: 20 mg/L, catalyst dose: 2 g/L, time: 120 min) and (b) leaching Ni and Ti from the photocatalysts (pH: 3, dye concentration: 20 mg/L, catalyst dose: 2 g/L, time: 120 min).



Fig. 7. Effect of the system components on phenol red removal (pH: 3, dye concentration: 20 mg/L, catalyst/adsorbent dose (if applicable): 2 g/L, time: 120 min).

Synergy factor<sub>TiO2-FSM-16</sub> = 
$$\frac{k_{obs}^{TiO2-FSM-16}}{k_{obs}^{FO2} + k_{obs}^{FSM-16}}$$
 (2)

Synergy factor<sub>TiO2-FSM-16/UV</sub> = 
$$\frac{k_{obs}^{\text{TiO2-FSM-16/UV}}}{k_{obs}^{\text{TiO2}} + k_{obs}^{\text{FSM-16}} + k_{obs}^{\text{UV}}}$$
(3)

The synergistic factor for the  $TiO_2$ -FSM-16 and  $TiO_2$ -FSM-16/UV processes were computed at 1.55 and 2.12, respectively. These findings reveal the effective interaction of UV,  $TiO_2$ , and FSM-16 components with each other that lead to elevated degradation of phenol red dye.

#### 3.7. Effect of scavengers

Components like hydroxyl radicals (OH•), superoxide ( $O_2^{-\bullet}$ ), photogenerated holes ( $h\vartheta^+$ ), and electrons (e<sup>-</sup>) can be effective in the photocatalytic dye removal process [52–54]. According to scientific papers [43,47], the compounds of *tert*-butanol, *p*-benzoquinone, silver nitrate, and ammonium oxalate are effective in abducting the active components of hydroxyl radicals (OH•), superoxide ( $O_2^{-\bullet}$ ), electrons, and photogenerated holes ( $h\vartheta^+$ ), respectively. To clarify the decomposition mechanism, the effect of these scavengers on dye removal efficiency was explored and the results are drawn in Fig. 8a. According to Fig. 8a, it is clear that the amount of dye removal has decreased with the addition of scavengers. The most severe decrease in phenol red removal efficiency was for ammonium oxalate,



**Fig. 8. (a)** Effect of scavengers on phenol red removal by the developed system (pH: 3, dye concentration: 20 mg/L, catalyst dose: 2 g/L, time: 120 min) and **(b)** kinetic of the dye degradation process (pH: 3, dye concentration: 20 mg/L, catalyst dose: 2 g/L).

followed by silver nitrate, *p*-benzoquinone, and *tert*-butanol. This indicates that photogenerated holes ( $h\theta^+$ ) were the most effective species in the photocatalytic process of phenol red removal. Similar results have been stated for the catalytic elimination of *p*-aminophenol and methylene blue dye by TiO<sub>2</sub>/RGO catalyst [47] and Brilliant green by MgFe<sub>2</sub>O<sub>4</sub> catalyst [43].

#### 3.8. Kinetic of the dye degradation process

The kinetics of the phenol red degradation process was evaluated under optimal conditions (see Fig. 8b). The first-order equation was utilized to assess the kinetic behavior of phenol red catalytic decontamination:

$$Ln \frac{[\text{final dye conc.}]}{[\text{initial dye conc.}]} = -k.t$$
(4)

If the graph  $Ln_{[initial dyc conc.]}^{[fnal dyc conc.]}$  vs time is drawn, a line is attained, and its slope is the reaction rate constant (k, min<sup>-1</sup>). As shown in Fig. 8b, the data follow first-order kinetics (R<sup>2</sup> > 0.96). The phenol red photodegradation rate constant using TiO<sub>2</sub>-FSM-16 and Ni-FSM-16 was calculated at 0.028 and 0.018 min<sup>-1</sup>, respectively. The reaction rate with TiO<sub>2</sub>-FSM-16 photocatalyst was about 1.5 times Ni-FSM-16. The kinetics of dye removal by various photocatalysts have followed the first-order model [43,52,55].

#### 3.9. Mineralization test

The amount of dye mineralization was investigated by two photocatalysts. The results showed that under optimal conditions (pH: 3, dye concentration: 20 mg/L, photocatalyst dose: 2 g/L, time: 120 min), the amount of dye mineralization by TiO<sub>2</sub>-FSM-16 and Ni-FSM-16 is 46% and 35%, respectively. It shows the effectiveness of the TiO<sub>2</sub>-FSM-16 photocatalyst in dye removal compared to Ni-FSM-16. The results of this study are in the range of values reported for acid orange removal using BiVO<sub>4</sub>/TiO<sub>2</sub> in the presence of H<sub>2</sub>O<sub>2</sub> and FeSO<sub>4</sub> [56].

#### 3.10. Literature comparison and comparison with TiO<sub>2</sub>-carbon and Ni-carbon

Several methods have been reported so far for removing dyes. Among these methods, researchers emphasize the oxidation and adsorption of dyes. Table 1 compares the methods mentioned in scientific texts for dye removal with our method. As can be seen, concerning the conditions of the tests, the photocatalysts in this paper are among the satisfactory catalysts for dye removal.

As it is clear in Table 1, the laboratory conditions of previously published work are different from our work, and an exact comparison cannot be made. Therefore, activated carbon as the most famous base for the catalyst was made under the same conditions as the catalysts of TiO<sub>2</sub>-FSM-16 and Ni-FSM-16 of this research and was investigated in the removal of the target pollutant. As shown in this table, the removal efficiency of carbon-based photocatalysts (BET: 810 m<sup>2</sup>/g) is lower than that of FSM-based ones. These differences are probably due to the lower surface area that carbon provided for the reaction compared to FSM.

#### Table 1

Comparison of the methods mentioned in the literature for dye removal with the method used in this work.

| Catalyst   | Dye conc.<br>(mg/L) | Dye               | Catalyst mass<br>(g/L) | UV irradiation time<br>(min) | Removal<br>(%) | Ref.                |
|--|---------------------|-------------------|------------------------|------------------------------|----------------|---------------------|
| TiO <sub>2</sub> nanosuspensions   | 3.2                 | Methylene<br>blue | 0.5                    | 120                          | 90             | [57]                |
| ZnO nanosuspensions  | 11                  | Methylene<br>blue | 1                      | 60                           | 100            | [57]                |
| Er/Pr-Codoped TiO <sub>2</sub> /Poly (vinylidene difluoride)–<br>Trifluoroethylene | 10                  | Methylene<br>blue | 0.1                    | 360                          | 95             | [58]                |
| ZnO-Graphene   | -                   | Methyl<br>orange  | -                      | 360                          | 89             | [59]                |
| BaTiO <sub>3</sub> /graphene oxide   | 5                   | Methylene<br>blue | 0.5                    | 180                          | 95             | [60]                |
| Nb(5)/TiO2 nanocomposite   | 20                  | Phenol red        | 0.1                    | 160                          | 94             | [ <mark>61</mark> ] |
| TiO <sub>2</sub>   | 20                  | Phenol red        | 0.1                    | 160                          | 79             | [ <mark>61</mark> ] |
| (α-FeOOH) Goethite composite <sup>a</sup>  | 1770                | Phenol red        | 1                      | 90                           | 96             | [62]                |
| TiO <sub>2</sub> +H <sub>2</sub> O <sub>2</sub> +O <sub>2</sub> gas + irradiation  | 10                  | Phenol red        | 0.5                    | 300                          | 94             | [63]                |
| TiO <sub>2</sub> -carbon   | 20                  | Phenol red        | 2                      | 120                          | 77             | This<br>work        |
| Ni-carbon  | 20                  | Phenol red        | 2                      | 120                          | 64             | This<br>work        |
| TiO <sub>2</sub> -FSM-16   | 20                  | Phenol red        | 2                      | 120                          | 84             | This<br>work        |
| Ni-FSM-16  | 20                  | Phenol red        | 2                      | 120                          | 66             | This<br>work        |

<sup>a</sup> This system has used irridation and H<sub>2</sub>O<sub>2</sub> beside the catalyst.

#### 4. Conclusions

In this study, TiO<sub>2</sub>-FSM-16 and Ni-FSM-16 photocatalysts were produced and used to remove phenol red dye in the presence of UV light. The TiO<sub>2</sub>-FSM-16 and Ni-FSM-16 photocatalyst had a BET area of 844.93 m<sup>2</sup>/g and 718.63 m<sup>2</sup>/g, respectively. Maximum phenol red removal using TiO<sub>2</sub>-FSM-16 and Ni-FSM-16 was obtained at 84% and 66%, respectively under the conditions of pH 3, dye concentration of 20 mg/L, catalyst dose of 2 g/L, and irradiation time of 120 min. The photocatalyst activity of FSM-16 was increased after the corporation with TiO<sub>2</sub> and Ni. The synergistic factor of TiO<sub>2</sub>-FSM-16 and TiO<sub>2</sub>-FSM-16/UV processes were found to be 1.55 and 2.12, respectively. The TiO<sub>2</sub>-FSM-16 photocatalyst with phenol red dye removal of 87% (TOC removal: 46%) had better activity than Ni-FSM-16 with 76% removal (TOC removal: 35%). Both photocatalysts, especially TiO<sub>2</sub>-FSM-16, had good capabilities in removing phenol red dye.

# Author contribution statement

Seyed Mohamadsadegh Mousavi: Conceived and designed the experiments; Performed the experiments. Seyed Hamed Meraji: Analyzed and interpreted the data; Wrote the paper. Ali Mohammad Sanati: Contributed reagents, materials, analysis tools or data; Performed the experiments. Bahman Ramavandi: Analyzed and interpreted the data; Wrote the paper.

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#### Data availability statement

Data will be made available on request.

#### Additional information

Supplementary content related to this article has been published online at [URL].

#### Declaration of interest's statement

The authors declare no conflict of interest.

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### Appendix A. Supplementary data

Supplementary data related to this article can be found at https://doi.org/10.1016/j.heliyon.2023.e14488.

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