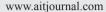


European Journal of Remote Sensing

An official journal of the Italian Society of Remote Sensing





Classifying silvicultural systems (coppices vs. high forests) in Mediterranean oak forests by Airborne Laser Scanning data

Francesca Bottalico¹, Davide Travaglini¹, Gherardo Chirici^{2*}, Marco Marchetti², Enrico Marchi¹, Susanna Nocentini¹ and Piermaria Corona³

¹Dipartimento di Gestione dei Sistemi Agrari, Alimentari e Forestali, University of Florence, via San Bonaventura, 13, 50145 Firenze, Italy

²Dipartimento di Bioscienze e Territorio, University of Molise,
Contrada Fonte Lappone, 86090 Pesche (Isernia), Italy

³Consiglio per la Ricerca e la sperimentazione in Agricoltura, Forestry Research Centre,
Viale S. Margherita, 80, 52100 Arezzo, Italy

*Corresponding author, e-mail address: gherardo.chirici@unimol.it

Abstract

Forest classification by silvicultural systems (coppices vs. high forests) is important for forest resource assessment as such systems relate to a wide variety of ecosystem services. In this paper the potential of Airborne Laser Scanning (ALS) data for Mediterranean oak forests classification of coppices with standards vs. high forests was investigated in three study areas in Italy. We addressed the following issues: can coppices and high forests be distinguished using a raster Canopy Height Model (CHM)? Which are the most efficient CHM-derived metrics? Does the scale of analysis influence the classification potential of CHM metrics? Our results show that CHM in grid format (1-m² pixel) provides support information to classify silvicultural systems.

Keywords: ALS; Canopy Height Model; image segmentation; classification.

Introduction

Forest management determines rotation length and the presence of different stages in forest succession. It implies important consequences in ecosystem functioning and services. Silvicultural systems are among the main drivers of differences in forest management and are mainly classified into high forest systems and coppice systems according to the method of tree regeneration. In high forest systems (conifer and broadleaved species) the stand is regenerated from seedlings, either natural or planted, or a combination of both, and rotation period is generally long. In coppice systems the regeneration of broadleaved species consists mainly of sprouts originating from

cut stumps, and rotation is generally shorter. The conversion of a coppice to a high forest is the transition from a stand based on vegetative reproduction to a stand that regenerates by seed [Matthews, 1991]. In Europe coppice systems account for 16% of the area of forest available for wood supply, covering extensive areas in several countries, including France, Italy, Greece, Turkey, Spain and Bulgaria [UN/ECE-FAO, 2000].

In Italy forest and other wooded lands extend over 10.5 million hectares [INFC, 2005]; high forest systems represent 42% of total forest area, with a slight majority of evenaged stands; coppice systems represent the other 58% of total forest area. Most coppices are located along the lower slopes of the Alps and all along the Italian peninsula and the islands, from the coast to the upper mountain zone in the Apennines. The majority are coppices with standards, i.e. even-aged stands with 40-150 trees ha-1 of two to three times the rotation age which are released at coppicing [Ciancio and Nocentini, 2004; Ciancio et al., 2006]. Felling is carried out by clearcutting at the end of rotation (usually 15-35 years), on areas from a few square hundred meters up to 10-20 hectares, but clearcut size is most usually in the range of 1-5 hectares. Coppice with standards is the only silvicultural system for which clearcut is allowed in Italy [Chirici et al., 2011]. Only small areas are covered by selection coppices (uneven aged coppices) or compound coppices (a mixture of coppice and high forest). Oaks (Quercus cerris L., Quercus pubescens Willd, Quercus ilex L.), sweet chestnut (Castanea sativa Miller) and beech (Fagus sylvatica L.) are the most common species in coppices, and often form pure stands [Ciancio and Nocentini, 2004].

The main product from coppice systems is fuelwood and polewood, the latter from chestnut stands. In the last years increasing oil prices and growing interest towards energy from renewable sources have further increased fuelwood demand, for both domestic and industrial uses [Lasserre et al., 2010; Picchio et al., 2009] and coppicing has again become a profitable management option in many private forests. An example of the renewed interest towards coppices in Europe is COST Action FP 1301 [EuroCoppice, http://www.eurocoppice.uni-freiburg.de] started in 2013, which has the aim of bringing together European scientists, experts and young scholars to exchange knowledge on coppice forestry and to develop innovative management and harvesting concepts/techniques for modern multifunctional coppice management systems.

Classification of high forest vs. coppices is important for forest resource assessment as such different systems relate to a different variety of ecosystem processes and services, and it is necessary for sustainable forest management at the stand level, since they are characterized by different forest structures and require different felling patterns.

Under management framework, coppice and high forest classification is carried out at the forest compartment level and it is traditionally based on ground observation or is performed by manual delineation of aerial photographs with the support of local field knowledge and observation. Even if these methods provide accurate classification, they are costly and time consuming.

Recently, the key importance of silvicultural systems for sustainable forestry has been confirmed by the development of methods for mapping potential forest management, especially in relation to the difficulties of detecting such systems [Hengeveld et al.,

2012]. Several studies have investigated the potential of multispectral sensor for forest structure variable estimation [e.g., Cohen and Spies, 1992; Hall et al., 2006; Lamonaca et al., 2008; Gebreslasie et al., 2010; Ozdemir and Karnieli, 2011]. However, as a general trend, passive optical sensors have limitations for characterizing vertical forest structure due do their relative insensitivity to canopy structure, especially when compared to Airborne Laser Scanning (ALS) [e.g., Lefsky et al., 2001; Hudak et al., 2002; Lu, 2006; Pflugmacher et al., 2012; Ozdemir and Donoghue, 2013].

ALS is an active remote sensing technology that can provide detailed information on forest canopy structure [Lefsky et al., 2002]. Several papers have documented the promising use of ALS data to predict relevant and basic forest structural attributes in boreal [e.g., Lim et al., 2003; Næsset, 2004; Hyyppä et al., 2008; van Leeuwen and Nieuwenhuis, 2010], Alpine [e.g., Hollaus et al., 2006; Clementel et al., 2012; Alberti et al., 2013], temperate and Mediterranean environments [Corona et al., 2012; Cartisano et al., 2013; Montaghi et al., 2013]. For instance, ALS data (alone or with spectral data) were used to estimate forest height [e.g., Magnussen and Boudewyn, 1998; Næsset and Økland, 2002], to predict biomass and volume of trees [e.g., Næsset, 1997; van Aardt et al., 2006; Corona and Fattorini, 2008] and shrubs [e.g., Estornell et al., 2012], to assess vertical stratification of forest vegetation [e.g., Zimble et al., 2003; Morsdorf et al., 2010; Ferraz et al., 2012], to map tree species composition [e.g., Ke et al., 2010; Cho et al., 2012], and to assess tree size and species diversity [e.g., Simons et al., 2012, 2013; Ozdemir and Donoghue, 2013].

Lefsky et al. [2005] found that mean height and height variability assessed by ALS data were useful to examine the relationships between comprehensive assemblages of forest canopy and stand structure indices. Pascual et al. [2008] found that median and standard deviation of height computed from ALS-derived digital Canopy Height Model (CHM) were slightly better than mean and standard deviation of height for distinguishing among forest structure types based on cluster analysis, especially in case of horizontally heterogeneous forests. Falkowski et al. [2009] used a variety of ALS metrics in conjunction with the Random Forests algorithm to classify forest successional stage across a structurally diverse mixed-species forest, and found that canopy cover and mean height were the most important ALS metrics for the classification.

Detecting silvicultural systems by ALS under temperate and Mediterranean conditions (i.e. the discrimination between high forest and coppice stands) has never been investigated (at least in our knowledge), even if this problem is still unresolved by conventional remote sensing approaches by optical imagery.

On the other hand, at least in Europe, it is quite frequent that ALS surveys are committed on large territories by governmental (national or local) administrations for purposes like topographical or hydrogeological investigations, and the ALS-derived CHM data are then released to the general public at low or no cost (the so called open data) [e.g., Corona et al., 2012; Montaghi et al., 2013].

Objectives

Taking into account the above considerations, in this study the potential of ALS-derived CHM data for Mediterranean oak forests classification into coppice stands and high forest

stands was investigated. We have addressed the following issues: 1) can coppices and high forests be distinguished using a raster CHM? 2) Which are the most efficient CHM metrics? 3) Does the scale of analysis influence the classification potential of CHM metrics to map coppices vs. high forests?

To our knowledge, this is the first attempt of using CHM metrics for coppice vs. high forest classification.

Material and methods

Study area

The work was carried out in three study areas in Italy, hereafter named A, B and C (Fig. 1). The study areas were selected on the basis of the availability of laser scanning data in forest regions with coppices and high forests.

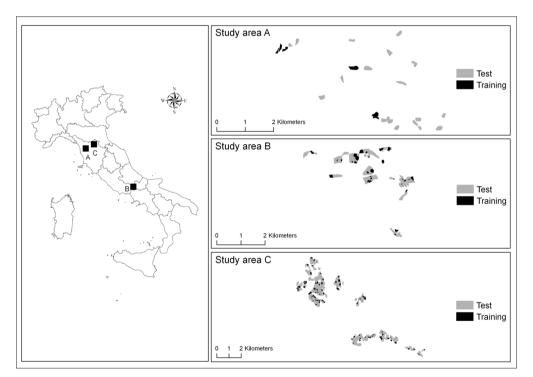


Figure 1 - Location of the study areas (on the left side) and distribution of forest stands according to their partitioning into training and test sites (on the right side).

Area A is located in the Tuscany Region, in the Province of Pisa (43° 43' - 43° 45' N, 10° 40' - 10° 44' E). It is a hilly area with gentle slopes and altitudes ranging between 0 and 100 m a.s.l. The average yearly temperature is 14 °C, and the average yearly precipitation is 910 mm. The forest landscape is characterized by maritime pine (*Pinus pinaster* Aiton) forests and mixed broadleaved forests dominated by turkey oak (*Quercus cerris* L.), the latter associated with other deciduous species, especially

manna ash (*Fraxinus ornus* L.), sessile oak (*Quercus petraea* Matt. Liebl.) and checker tree (*Sorbus torminalis* Crantz). The understory layer is composed by shrubs with density inversely related to tree cover [Travaglini et al., 2011].

Area B is located in the Molise Region, in the Province of Isernia (41° 33' - 41° 35' N, 14° 02' - 14° 05' E). The average yearly temperature is of 12 °C, and the average yearly precipitation is of 1180 mm. Altitude varies between 400 and 1000 m a.s.l. Turkey oak, downy oak (*Quercus pubescens* Willd.) and hop hornbeam (*Ostrya carpinifolia* Scop.) are the prevailing forest species [Garfi and Marchetti, 2011].

Area C is located in the Tuscany Region, in the Province of Florence (43° 55' - 44° 00' N, 11° 16' - 11° 24' E). It is a hilly area at the base of the Apennines mountain, with altitudes ranging between 150 and 550 m a.s.l. The average yearly temperature is 13 °C, and the average yearly precipitation is 1025 mm. Forests are dominated by oak stands (mainly turkey oak associated with manna ash, chestnut and hop hornbeam), chestnut stands, and European black pine stands (*Pinus nigra* Arn.), the latter originated from plantations.

In each study area the distribution of even-aged forest stands dominated by Turkey oak and downy oak were extracted from existing forest type maps (scale of 1:10,000) [Garfi and Marchetti, 2011; Nocentini et al., 2011]. A total forest surface of 45 ha, 107 ha, and 506 ha was considered for the study areas A, B and C, respectively. The silvicultural systems adopted in these forests are high forest and coppice with standards (herewith, coppice) under a regular felling regime, mainly addressed towards bioenergy assortments which have an increasing market value. Forest development stages varies from very young (i.e., stand after final cut) to adult (i.e., stand close to the rotation age) for coppices, while no young stands occur for high forests. This is a common situation for broadleaved forests in Italy (Table 1).

| | O | | 0 | • | |
|------------|---------------|--------------|---------------|--------------|--|
| | Сор | pice | High forest | | |
| Study area | Forest height | Canopy cover | Forest height | Canopy cover | |
| | m | % | m | % | |
| A | 4-16 | 30-80 | 15-25 | >80 | |
| В | 3-12 | 30-80 | 12-22 | >90 | |
| С | 3-15 | 30-80 | 12-25 | >90 | |

Table 1 - Forest height and canopy cover of coppices and high forests in the study areas.

Airborne laser scanning data and pre-processing

In area A, ALS survey was performed in winter 2008; an ALS ALTM (Airborne Laser Terrain Mapper) Gemini sensor was used; the sensor recorded two echoes per pulse with an average laser point density of about 1.6 laser points per m². In area B, ALS dataset was acquired in June 2010 by Partenavia P68 I-GIFE aircraft carrying an Optech Pegasus sensor which recorded two echoes per pulse; the average laser point density was about 1.5 laser points per m². In area C, ALS survey was performed in winter 2007 by an ALS ALTM 3033 sensor which registered two echoes per pulse; the

average laser point density was about 1.0 laser points per m².

The ALS data were filtered and classified in order to separate points belonging to forest and natural vegetation cover from ground points. In order to do so, outlying pulses due to sensor errors were first identified and erased from the ALS dataset using TerraScan software [Terrasolid, 2005]. Subsequently, ground level echoes were separated from echoes of the outer forest canopy using the TerraScan software and the Axelsson [2000] algorithm. Both ground and canopy echoes, once separated, were spatially interpolated and resampled in a grid format with a geometric resolution of 1 m to produce, respectively, a Digital Terrain Model (DTM) and a Digital Surface Model (DSM).

A CHM was generated as result of the algebraic subtraction of DTM from DSM. Later, pixels with height < 2 m in the CHM were set to zero to avoid the influence of the shrub understory [Floris et al., 2010; Gonzáles-Ferreiro, 2012]. Power lines were also masked on the basis of a manual delineation of electric lines over the raster CHM. Finally, pixels outside the selected study areas were excluded.

Ground data

In each study area forest cover was classified into high forest stands (T1) and coppice stands on the basis of ground observations with the support of a GPS receiver Trimble Juno 3B Handheld with 2-5 m positional accuracy. Fieldwork was carried out in winter 2012 by visual interpretation of forest polygons generated on the basis of CHM (see Polygon delineation section). In case of coppice stands, two age classes were considered taking into account the time lag between ALS survey and ground observation: coppice with age > 10 years (called adult coppices, T2), and coppice with age < 10 years (called young coppices, T3). The surface covered by coppice stands and high forest stands in the three study areas is reported in Table 2.

| young | young coppice stands (13) in the study areas. | | | | | | | | | | |
|-------|---|-------|-------|------|-------|--|--|--|--|--|--|
| St. I | | Т1 | Т2 | Т3 | Total | | | | | | |
| Stu | dy area | ha | ha | ha | ha | | | | | | |
| | A | 15.0 | 15.0 | 15.0 | 45.0 | | | | | | |
| | В | 36.1 | 35.5 | 35.8 | 107.4 | | | | | | |
| | С | 130.4 | 287.4 | 88.2 | 506.0 | | | | | | |

Table 2 - Surface (in hectares) of high forest stands (T1), adult coppice stands (T2) and young coppice stands (T3) in the study areas.

Mapping silvicultural systems

The mapping procedure involved five consecutive steps (Fig. 2): 1) polygon delineation from the CHM, 2) selection of training sites and test sites, 3) CHM-derived metrics extraction and analysis, 4) polygon classification, 5) accuracy assessment.

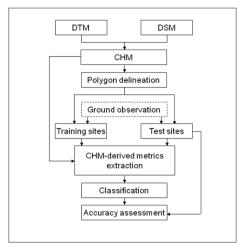


Figure 2 - Flow diagram of the proposed silvicultural systems classification procedure.

Polygon delineation

The aim of this step was to divide forest cover into forest regions with similar characteristics in terms of canopy structure. The object-oriented segmentation tool available in the Definiens (vers. 5) software was used to delineate polygons on the basis of CHM.

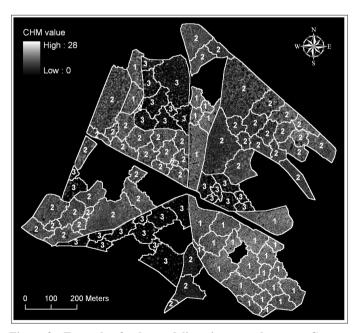


Figure 3 - Example of polygon delineation over the raster Canopy Height Model (CHM) generated by a scale parameter of 50. Numbers represent high forests (1), adult coppices (2) and young coppices (3).

Segmentation is the subdivision of an image into separated regions. Definiens uses a bottomup approach to generate polygons based upon a multiresolution segmentation technique and criteria of homogeneity in term of colour, smoothness and compactness; the so-called scale parameter influences the average object size: a larger value leads to bigger objects and vice versa [Baatz and Shäpe, 2000]. In this study six consecutive segmentations were applied to the CHM. The CHM-based segmentation was expected to provide differentiation between forest stands with different height. After some preliminary tests which were carried out to assess the effect of different user-specified parameters through a trial-and-error approach, a first segmentation level with a scale parameter of 50 was derived, assigning a weight of 0.5 to colour, smoothness and compactness (Fig. 3). This setting was chosen empirically by visual inspection as it was judged able to delineate homogeneous forest regions representing coppices and high forest stands. These regions were later sub-segmented into five hierarchical levels with a scale parameter of 30, 20, 15, 10 and 5. Colour, smoothness and compactness were unchanged during the multiple level segmentation process. More details on LiDAR-based segmentation and related issues can be found in Ke et al. [2010]. At the first segmentation level a total of 30, 191, and 778 polygons were generated in the study areas A, B and C, respectively, with an average size of 1.19 ha, 0.56 ha and 0.65 ha, respectively. Total number and average size of polygons generated with the multiple level object-oriented segmentation are given in Table 3.

Table 3 - Total number and average size (in hectares), with standard deviation in brackets, of the polygons generated by the multiple level object-oriented segmentation.

| Study area | Level 1 (scale parameter = 50) | | Level 2 (scale parameter = 30) | | Level 3 (scale parameter = 20) | | Level 4 (scale parame | | | ter = 10) | Level 6 (scale paramete | er = 5) |
|---------------|--------------------------------------|----------------------|--------------------------------------|----------------------|--------------------------------------|----------------------|-----------------------------|----------------------|-------|----------------------|-------------------------------|----------------------|
| | Num. | Average size (ha) | Num. | Average size (ha) | Num. | Average size (ha) | Num. | Average size (ha) | Num. | Average size (ha) | Num | Average size (ha) |
| A | 30 | 1.190 (0.821) | 224 | 0.203 (0.176) | 652 | 0.069 (0.066) | 1311 | 0.034 (0.030) | 3122 | 0.014 (0.011) | 11346 | 0.004 (0.003) |
| В | 191 | 0.560 (0.430) | 564 | 0.190 (0.137) | 1259 | 0.085 (0.061) | 2345 | 0.046 (0.033) | 5556 | 0.019 (0.015) | 22341 | 0.005 (0.004) |
| С | 778 | 0.650 (0.531) | 2340 | 0.216 (0.179) | 5799 | 0.087 (0.069) | 11295 | 0.045 (0.034) | 29144 | 0.017 (0.013) | 132682 | 0.004 (0.003) |

Training sites and test sites

In each study area training sites and test sites were selected from polygons delineated at the first segmentation level using a random selection method; 20% of the forest area was selected for training sites and the remaining 80% was used as test sites (Fig. 1). At the lower hierarchical levels training and test site distribution was the same as that adopted for the first segmentation level. Training and test site polygons were classified into silvicultural systems by observations in the field (see Ground data section).

Metrics extraction and analysis

For each polygon delineated by the multiple level object-oriented segmentation process, the following ALS metrics were computed on the basis of the CHM: minimum height, maximum height, mean height, range of the heights, standard deviation (SD) of the heights, and coefficient of variation (CV) of the heights. These metrics were selected as they are known to be related to forest stand structural characteristics [e.g., Zimble et al., 2003; Parker and Russ, 2004; Lefsky et al., 2005; Pascual et al., 2008; Barbati et al., 2009; Falkowski et al., 2009; Monnet et al., 2010; Corona et al., 2012]. Indeed, silvicultural systems are expected to be distinguished on the basis of forest height and height variations: the former should be lower in coppices which are, on average, younger than high forests, the latter should be higher in coppices due to the presence of standards which are taller than the shoots, even though these dissimilarities progressively decrease as coppice age increases. To investigate the capability of each metric in distinguishing between high forest stands, adult coppice stands and young coppice stands, a statistical analysis was performed on CHM metrics computed on training sites using SPSS for Windows. To this end two non parametric tests (Kruskal-Wallis test and Median test) were used to check whether the selected metrics varied significantly among T1, T2 and T3. In case of statistically significant differences (p < 0.001) the Mann-Whitney U test (a Two-Independent-Samples Test) was used to determine the differences between groups.

Polygon classification and accuracy assessment

Polygon classification was performed using the Maximum Likelihood algorithm available in Idrisi Selva software, a well known classification method for remote sensing applications [Lillesand and Kiefer, 2000], with reference to the training sites described above (see Training sites and test sites section).

The thematic accuracy of the classification was estimated over the test sites by means of a comparison between the classification of remote sensing data and ground truth. The following indexes were computed: overall accuracy (OA), producer's accuracy (PA), user's accuracy (UA), and kappa index of agreement (KIA) [Congalton, 1991].

Results

Results of the statistical analysis for study areas A, B and C are shown in Tables 4, 5 and 6. They indicate that the CHM metrics are able to discriminate between coppice and high forest stands. Some metrics seem to be influenced by the scale of the analysis, as they produced different results in relation to the scale parameter adopted in the segmentation process. The ALS metrics that showed such behaviour were: minimum height (in all the areas), SD (in A and B), maximum height and range (in B and C). Minimum height was able to distinguish types T1, T2 and T3 when scale parameter was \leq 15 in areas A and C, and \leq 30 in area B; with a scale parameter of 20 the types T2 and T3 were not discriminated by minimum height in the study area A. These results can be explained by the fact that when larger polygons are delineated, gaps in the forest cover with CHM heights close to zero are not detected by the segmentation process, so that coppices and high forests cannot be discriminated. As shown by CHM height histograms computed on training sites (Figs. 4, 5 and 6), CHM heights equal to zero occur in every silvicultural system, even thought with a smaller proportion in high forest than in coppice stands. Maximum height distinguishes the target forest classes when

scale parameter is ≤ 15 in area B and ≤ 30 in area C, while it is not able to separate T2 from T3 when a greater scale parameter is used: when larger polygons are delineated, maximum height in young coppices corresponded to the height of the tallest standard, which is similar to the maximum height determined in adult coppices as shown by CHM height histograms

to the maximum height determined in adult coppices as shown by CHM height histograms (Figs. 5 and 6). The results provided by range are similar to those reported for maximum height: range worked well in area C when scale parameter was \leq 30, while in area B it was able to discriminate the types T1, T2 and T3 only with a scale parameter of 5.

Table 4 - Study area A: results of the statistical analysis (Kruskal-Wallis test, Median test and Mann-Whitney test) carried out on CHM-derived metrics computed on training site polygons for high forest (T1), adult coppice (T2) and young coppice (T3) (SD = standard deviation, CV = coefficient of variation).

| Level of segmentation (scale parameter) | segmentation (scale polygons | | CHM- derived metric | derived on training site | | | Kruskal- Wallis test (p-level) | Median test (p-level) | Mann-Whitney test | |
|--|---------------------------------|-----|---------------------------|--------------------------|------|------|---|-----------------------------|-------------------|----------------------|
| | T1 | T2 | T3 | | T1 | T2 | Т3 | | | |
| | | | | Min | 0.0 | 0.0 | 0.0 | - | - | - |
| | | | | Max | 26.5 | 20.5 | 13.0 | 0.000 | 0.001 | T1≠T2; T2≠T3 |
| L1 (50) | 7 | 8 | 9 | Mean | 18.9 | 6.4 | 2.2 | 0.000 | 0.001 | T1≠T2; T2≠T3 |
| L1 (30) | ' | 0 | ' | Range | 26.5 | 20.5 | 13.0 | 0.000 | 0.001 | T1≠T2; T2≠T3 |
| | | | | SD | 4.3 | 4.3 | 3.1 | 0.002 | 0.002 | T1=T2; T2≠T3 |
| | | | | CV | 0.2 | 0.7 | 1.4 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | Min | 0.4 | 0.0 | 0.0 | 0.157 | 0.151 | T1=T2; T2=T3 |
| | | | | Max | 26.8 | 20.1 | 12.9 | 0.000 | 0.000 | T1\neq T2; T2\neq T3 |
| L2 (30) | 10 | 16 | 21 | Mean | 19.5 | 7.0 | 2.3 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| L2 (30) | 10 | 10 | 21 | Range | 26.3 | 20.1 | 12.9 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | SD | 4.3 | 4.2 | 3.1 | 0.000 | 0.000 | T1=T2; T2≠T3 |
| | | | | CV | 0.2 | 0.7 | 1.5 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | Min | 1.7 | 0.5 | 0.0 | 0.000 | 0.000 | T1≠T2; T2=T3 |
| | | | | Max | 23.5 | 16.7 | 10.4 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| L3 (20) | 37 | 48 | 42 | Mean | 17.5 | 7.1 | 2.0 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| 25 (20) | 37 | .0 | | Range | 21.9 | 16.3 | 10.4 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | SD | 3.6 | 3.5 | 2.6 | 0.000 | 0.001 | T1=T2; T2≠T3 |
| | | | | CV | 0.2 | 0.7 | 1.6 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | Min | 3.7 | 0.8 | 0.0 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | Max | 24.0 | 15.4 | 9.9 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| L4 (15) | 73 | 102 | 80 | Mean | 18.4 | 7.2 | 2.2 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| 2.(10) | ,,, | 102 | | Range | 20.2 | 14.5 | 9.9 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | SD | 3.4 | 3.3 | 2.5 | 0.000 | 0.000 | T1=T2; T2≠T3 |
| | | | | CV | 0.2 | 0.7 | 1.6 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | Min | 6.8 | 1.6 | 0.1 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | Max | 23.3 | 13.4 | 9.0 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| L5 (10) | 169 | 235 | 212 | Mean | 18.3 | 7.3 | 2.6 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| - () | | | | Range | 16.5 | 11.8 | 8.9 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | SD | 3.2 | 2.8 | 2.4 | 0.000 | 0.029 | T1≠T2; T2≠T3 |
| | | | - | CV | 0.2 | 0.6 | 1.6 | 0.000 | 0.000 | T1\neq T2\neq T3 |
| | | | | Min | 12.1 | 2.9 | 0.6 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | Max | 21.9 | 10.7 | 6.2 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| L6 (5) | 770 | 785 | 882 | Mean | 18.4 | 7.0 | 2.8 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| (-) | | | | Range | 9.8 | 7.8 | 5.6 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | SD | 2.4 | 2.0 | 1.6 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | CV | 0.2 | 0.5 | 1.1 | 0.000 | 0.000 | T1≠T2; T2≠T3 |

Table 5 - Study area B: results of the statistical analysis (Kruskal-Wallis test, Median test and Mann-Whitney test) carried out on CHM-derived metrics computed on training site polygons for high forest (T1), adult coppice (T2) and young coppice (T3) (SD = standard deviation, CV = coefficient of variation).

| Level of segmentation (scale parameter) | ation site polygons | | CHM- derived metric | Average value of the CHM-derived metric computed on training site polygons | | Kruskal- Wallis test (p-level) | Median test (p-level) | Mann- Whitney test | | |
|--|---------------------|------|---------------------------|--|------|---|-----------------------------|-----------------------|-------|--------------|
| | T1 | T2 | Т3 | | T1 | T2 | Т3 | | | |
| | | | | Min | 2.4 | 0.4 | 0.0 | 0.132 | 0.143 | T1=T2; T2=T3 |
| | | | | Max | 25.0 | 17.9 | 16.5 | 0.000 | 0.000 | T1≠T2; T2=T3 |
| 1.1 (50) | 13 | 7 | 10 | Mean | 17.8 | 9.3 | 4.4 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| L1 (50) | 13 | ' | 10 | Range | 22.6 | 17.4 | 16.5 | 0.018 | 0.023 | T1=T2; T2=T3 |
| | | | | SD | 2.5 | 3.2 | 3.5 | 0.053 | 0.296 | T1=T2; T2=T3 |
| | | | | CV | 0.1 | 0.3 | 0.8 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | Min | 4.5 | 1.1 | 0.0 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | Max | 23.5 | 15.6 | 14.8 | 0.000 | 0.000 | T1≠T2; T2=T3 |
| L2 (30) | 37 | 33 | 43 | Mean | 17.2 | 8.7 | 4.4 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| 22 (30) | " | | | Range | 19.0 | 14.6 | 14.8 | 0.000 | 0.003 | T1≠T2; T2=T3 |
| | | | | SD | 2.5 | 2.6 | 2.9 | 0.001 | 0.011 | T1=T2; T2=T3 |
| | | | | CV | 0.1 | 0.3 | 0.7 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | Min | 6.6 | 1.8 | 0.0 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | 87 | Max | 22.2 | 15.3 | 13.2 | 0.000 | 0.000 | T1≠T2; T2=T3 |
| L3 (20) | 79 | 77 | | Mean | 16.9 | 9.0 | 4.2 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| L3 (20) | ' | ' ' | | Range | 15.6 | 13.5 | 13.2 | 0.022 | 0.149 | T1=T2; T2=T3 |
| | | | | SD | 2.2 | 2.5 | 2.8 | 0.000 | 0.000 | T1=T2; T2≠T3 |
| | | | | CV | 0.1 | 0.3 | 0.9 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | Min | 8.4 | 2.5 | 0.1 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | Max | 21.0 | 14.8 | 12.5 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| 14(15) | 159 | 162 | 131 | Mean | 16.4 | 9.2 | 4.2 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| L4 (15) | 139 | 102 | 131 | Range | 12.6 | 12.3 | 12.4 | 0.184 | 0.000 | T1=T2; T2=T3 |
| | | | | SD | 2.0 | 2.5 | 2.7 | 0.000 | 0.000 | T1=T2; T2≠T3 |
| | | | | CV | 0.1 | 0.3 | 0.9 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | Min | 10.5 | 3.7 | 0.6 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | Max | 20.3 | 13.5 | 11.2 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| 1.5 (10) | 339 | 356 | 387 | Mean | 16.6 | 9.0 | 4.7 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| L5 (10) | 339 | 330 | 38/ | Range | 9.8 | 9.8 | 10.6 | 0.000 | 0.000 | T1=T2; T2≠T3 |
| | | | | SD | 1.7 | 2.1 | 2.4 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | CV | 0.1 | 0.3 | 0.8 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | Min | 13.2 | 5.8 | 2.0 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | Max | 19.2 | 12.2 | 8.5 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| 16 (5) | 1221 | 1240 | 1742 | Mean | 16.7 | 9.2 | 5.0 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| L6 (5) | 1231 | 1349 | 1742 - | Range | 6.0 | 6.3 | 6.5 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | SD | 1.3 | 1.5 | 1.7 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | CV | 0.1 | 0.2 | 0.6 | 0.000 | 0.000 | T1≠T2; T2≠T3 |

Table 6 - Study area C: results of the statistical analysis (Kruskal-Wallis test, Median test and Mann-Whitney test) carried out on CHM-derived metrics computed on training site polygons for high forest (T1), adult coppice (T2) and young coppice (T3) (SD = standard deviation, CV = coefficient of variation).

| Level of segmentation (scale parameter) | | er of tra olygons | ining | derived metric computed on training site polygons | | Kruskal- Wallis test (p-level) | Wallis Median test test | | | |
|--|------|----------------------|-------|---|------|---|----------------------------|-------|-------|--------------|
| | T1 | T2 | Т3 | | T1 | T2 | T3 | | | |
| | | | | Min | 0.0 | 0.0 | 0.0 | - | - | - |
| | | | | Max | 28.5 | 21.46 | 19.2 | 0.000 | 0.000 | T1≠T2; T2=T3 |
| L1 (50) | 20 | 45 | 13 | Mean | 15.3 | 8.2 | 2.4 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| 121 (30) | 20 | 13 | 13 | Range | 28.5 | 21.4 | 19.2 | 0.000 | 0.000 | T1≠T2; T2=T3 |
| | | | | SD | 6.5 | 4.4 | 3.7 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | CV | 0.4 | 0.6 | 1.8 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | Min | 0.0 | 0.0 | 0.0 | - | - | - |
| | | | | Max | 26.5 | 18.8 | 16.9 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| L2 (30) | 104 | 261 | 62 | Mean | 15.4 | 8.0 | 2.8 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| 1.2 (50) | 104 | 201 | 02 | Range | 26.5 | 18.8 | 16.9 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | SD | 6.3 | 4.3 | 3.7 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | CV | 0.4 | 0.6 | 1.7 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | 233 | Min | 0.2 | 0.0 | 0.0 | 0.011 | 0.012 | T1=T2; T2=T3 |
| | | | | Max | 25.0 | 17.6 | 15.1 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| L3 (20) | 294 | 639 | | Mean | 15.0 | 8.1 | 3.0 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| L3 (20) | 274 | 037 | | Range | 24.8 | 17.5 | 15.1 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | SD | 5.9 | 4.1 | 3.4 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | CV | 0.4 | 0.6 | 2.1 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | Min | 0.6 | 0.2 | 0.1 | 0.000 | 0.000 | T1≠T2; T2=T3 |
| | | | | Max | 24.1 | 16.7 | 13.3 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| L4 (15) | 582 | 1202 | 491 | Mean | 14.9 | 8.1 | 3.3 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| L4 (13) | 362 | 1202 | 491 | Range | 23.5 | 16.5 | 13.3 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | SD | 5.7 | 3.9 | 3.2 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | CV | 0.4 | 0.6 | 2.2 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | Min | 1.9 | 0.7 | 0.3 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | Max | 22.9 | 15.3 | 10.5 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| L5 (10) | 1514 | 3129 | 1324 | Mean | 14.9 | 8.2 | 3.6 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| L3 (10) | 1314 | 312) | 1324 | Range | 20.9 | 14.6 | 10.2 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | SD | 5.2 | 3.5 | 2.6 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | CV | 0.4 | 0.6 | 1.9 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | Min | 6.6 | 2.6 | 0.8 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | Max | 20.9 | 13.1 | 7.4 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| L6 (5) | 8739 | 1/1119 | 4612 | Mean | 15.1 | 8.4 | 3.6 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| TO (2) | 0/39 | 14118 | 4012 | Range | 14.3 | 10.5 | 6.6 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | SD | 3.9 | 2.7 | 1.8 | 0.000 | 0.000 | T1≠T2; T2≠T3 |
| | | | | CV | 0.3 | 0.5 | 1.2 | 0.000 | 0.000 | T1≠T2; T2≠T3 |

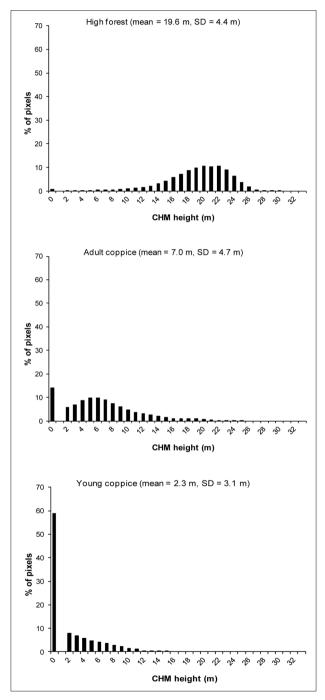


Figure 4 - Area A: CHM height histograms on training sites.

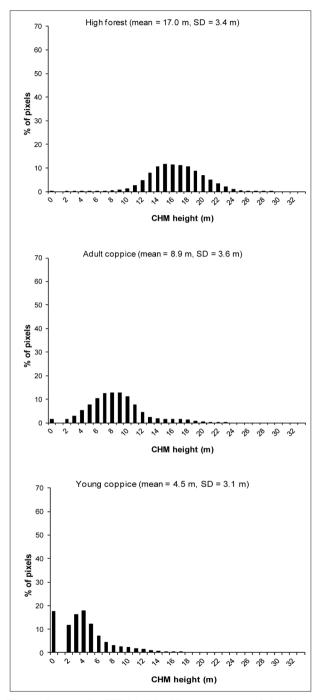


Figure 5 - Area B: CHM height histograms on training sites.

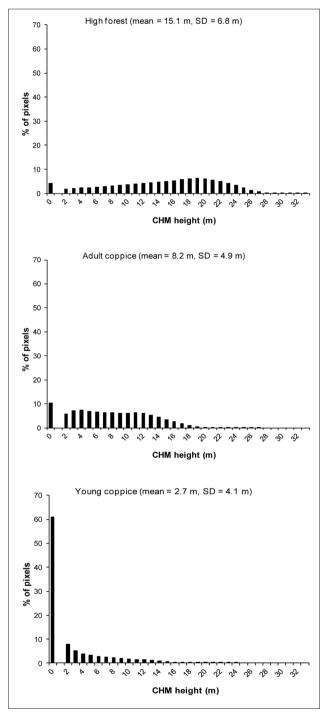


Figure 6 - Area C: CHM height histograms on training sites.

Conversely, the metrics which were not affected by the scale of the analyses were: mean height and CV (in all the areas), range and maximum height (in A), and SD (in C). Such results clearly indicated that mean height and CV are the only metrics capable of separating the target forest classes in all the investigated study areas independently from the scale of the analysis.

Mean height and CV were finally used to map high forests, adult coppices and young coppices. The classification was performed on polygons delineated at the first level of segmentation (scale parameter = 50). The final maps are shown in Figure 7. Thematic accuracy of the classification is shown in Tables 7, 8 and 9: overall accuracy was 85.9% in area C, 86.0% in area A and 92.7% in the area B. KIA was 0.75 in the area C, 0.79 in the area A and 0.89 in area B.

Classification errors between T1 and T2 and between T2 and T3 can be attributed to stand development (e.g., young high forests classified as mature coppices and mature coppices classified as high forests) and stand density (on one hand, a low tree density determines a reduction of the mean height computed by the raster CHM and, on the other hand, an increase of height variability). Other possible influencing factors are soil fertility, which influences stand height, or silvicultural practices such as leaving, at coppicing, a certain number of older standards, i.e. with ages > 2-3 times rotation period [Ciancio and Nocentini, 2004]. Since in the study areas soil conditions were quite uniform, this factor should have a limited influence on classification errors. Misclassification between T1 and T3 occurred only in area C, and on a very limited portion of the area characterized by large openings in the forest cover.

It is worth noting that thematic accuracy of the classification increases when no distinction is done between young (< 10 years) and adult (> 10 years) coppices: in this case the overall accuracy becomes greater than 91% in all the study areas (Tabs. 10, 11 and 12), and KIA is 0.76 in area C, 0.89 in area B, and 1 in area A.

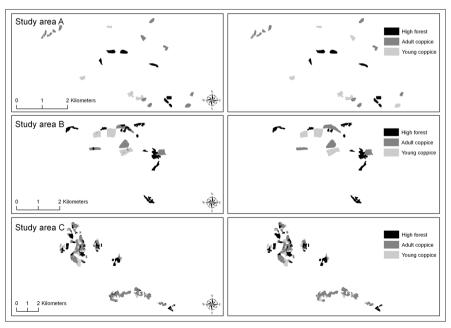


Figure 7 - Classification of silvicultural systems by ALS data (on the left side) and ground truth (on the right side).

Table 7 - Study area A: thematic accuracy (OA = overall accuracy, PA = producer's accuracy; UA = user's accuracy) of coppice (T2 = adult coppice, T3 = young coppice) and high forest (T1) classification computed on test sites.

| | | | UA (%) | | |
|----------------|---------|---------|---------|---------|--------------|
| | | T1 (ha) | T2 (ha) | T3 (ha) | UA (76) |
| | T1 (ha) | 11.7 | 0.0 | 0.0 | 100.0 |
| Classification | T2 (ha) | 0.0 | 11.9 | 4.9 | 70.6 |
| | T3 (ha) | 0.0 | 0.0 | 6.8 | 100.0 |
| PA (%) | | 100.0 | 100.0 | 57.9 | OA(%) = 86.0 |

Table 8 - Study area B: thematic accuracy (OA = overall accuracy, PA = producer's accuracy; UA = user's accuracy) of coppice (T2 = adult coppice, T3 = young coppice) and high forest (T1) classification computed on test sites.

| | | | UA (%) | | |
|----------------|---------|---------|---------|---------|--------------|
| | | T1 (ha) | T2 (ha) | T3 (ha) | UA (70) |
| | T1 (ha) | 28.4 | 4.8 | 0.0 | 85.5 |
| Classification | T2 (ha) | 0.3 | 23.5 | 1.1 | 94.3 |
| | T3 (ha) | 0.0 | 0.0 | 27.4 | 100.0 |
| PA (%) | | 99.0 | 83.0 | 96.1 | OA(%) = 92.7 |

Table 9 - Study area C: thematic accuracy (OA = overall accuracy, PA = producer's accuracy; UA = user's accuracy) of coppice (T2 = adult coppice, T3 = young coppice) and high forest (T1) classification computed on test sites.

| | | | Ground truth | | |
|----------------|---------|---------|--------------|---------|--------------|
| | | T1 (ha) | T2 (ha) | T3 (ha) | UA (%) |
| | T1 (ha) | 80.6 | 11.4 | 0.0 | 87.6 |
| Classification | T2 (ha) | 23.6 | 209.7 | 13.4 | 85.0 |
| | T3 (ha) | 0.1 | 8.3 | 56.8 | 87.2 |
| PA (%) | | 77.3 | 91.4 | 80.9 | OA(%) = 85.9 |

Table 10 - Study area A: thematic accuracy (OA = overall accuracy, PA = producer's accuracy; UA = user's accuracy) of coppice (T2 = adult coppice, T3 = young coppice) and high forest (T1) classification computed on test sites.

| | | Groun | UA (%) | |
|----------------|------------|---------|------------|---------------|
| | | T1 (ha) | T2+T3 (ha) | UA (70) |
| Classification | T1 (ha) | 11.7 | 0.0 | 100.0 |
| Classification | T2+T3 (ha) | 0.0 | 23.6 | 100.0 |
| PA (%) | | 100.0 | 100.0 | OA(%) = 100.0 |

Table 11 - Study area B: thematic accuracy (OA = overall accuracy, PA = producer's accuracy; UA = user's accuracy) of coppice (T2 = adult coppice, T3 = young coppice) and high forest (T1) classification computed on test sites.

| | | Groun | UA (%) | |
|----------------|------------|---------|------------|--------------|
| | | T1 (ha) | T2+T3 (ha) | UA (70) |
| Classification | T1 (ha) | 28.40 | 3.85 | 88.1 |
| | T2+T3 (ha) | 0.30 | 52.95 | 99.4 |
| PA (%) | | 99.0 | 93.2 | OA(%) = 95.1 |

Table 12 - Study area C: thematic accuracy (OA = overall accuracy, PA = producer's accuracy; UA = user's accuracy) of coppice (T2 = adult coppice, T3 = young coppice) and high forest (T1) classification computed on test sites.

| | | Groun | Ground truth | | | |
|----------------|------------|---------|--------------|--------------|--|--|
| | | T1 (ha) | T2+T3 (ha) | UA (%) | | |
| Classification | T1 (ha) | 80.6 | 11.4 | 87.6 | | |
| Classification | T2+T3 (ha) | 23.7 | 288.1 | 92.4 | | |
| PA (%) | | 77.3 | 96.2 | OA(%) = 91.3 | | |

Discussion and conclusions

In this study we have assessed the performance of ALS-derived CHM metrics to classify silvicultural systems in Mediterranean deciduous oak forests. Such classification is traditionally carried out by means of ground observations. We believe our results show the potential of raster CHM (1-m² pixel) for classifying coppices vs. high forests.

We used a methodological approach similar to that by Pascual et al. [2008] in conifer stands. At least to our knowledge, this is the first time that ALS data is used for broadleaved forests classification by silvicultural systems.

The methodology is implicitly based on the assumption that even-aged coppices and high forests present different heights and height variations due to differences in age and structure determined by forest management in term of rotation length and felling patterns. Our work shows that mean and coefficient of variation of heights extracted from raster CHM derived from low point density ALS data (1-1.6 laser points per m²) acquired in different season (winter and summer) provide effective support information to discriminate between high forest and coppice stands. The same two metrics have been proved relevant for forest structure classification by Falkowski et al. [2009], Lefsky et al. [2005] and Pascual et al. [2008]. The classification efficiency of other four CHM metrics have proven to be influenced by the scale of the analysis (minimum height, maximum height, range, and standard deviation). Conversely, mean and coefficient of variation of heights provide robust results for scale parameters up to 50, which is the maximum value considered in our work. It is worth noting that the scale parameter of 50, which in the examined study areas proved to successfully separate coppice polygons from high forest polygons, falls in the optimal scale range 20-150 suggested by Ke et al. [2010] for forest species classification.

In the examined study cases, classification errors have proven to be quite low: the overall accuracy of the semiautomatic classification produced by Maximum Likelihood algorithm is > 85% when three classes (high forest, adult coppice, young coppice) are considered, and it is > 91 % when only two classes are considered (high forest vs. coppice). However, it is worth noting that the high accuracy achieved in our study is influenced by the low number of considered classes. In addition, it is interesting to note that the overall accuracy in the study areas A and C was similar and lower than in the area B, probably because in A and C ALS data was acquired during leaf-off periods, and the derived CHM underestimated the canopy height more than in B where ALS was acquired during leaf-on conditions (Wasser et al., 2013).

The procedure presented in this paper can be used by forest managers for stand classification when drafting management plans, and also for stratified estimation of forest attributes to increase the precision of estimates under forest inventory applications [e.g., McRoberts et al., 2005; Corona and Marchetti, 2007; Corona et al., 2011], and even for generating auxiliary

information relevant for mapping fire risk [e.g., González-Olabarria et al., 2012].

It is important to stress the simplicity and feasibility of the proposed methodological approach, which exploits: (i) the raster ALS-derived CHM data, which are often available to the general public as open data, i.e. at no (or very low) cost (see Objectives section), (ii) conventional image segmentation techniques, and (iii) well known classification algorithm available in many commercial and open-source software.

On the other hand, further studies on this topic are needed to assess the efficiency of the proposed method for classifying silvicultural systems on large areas, in different forests types with different forest developmental stages; to test the eventual advantage of more complex classification algorithms; to investigate whether the original ALS point cloud might contain additional information to such an end [e.g., Jakubowski et al., 2013], and to assess the eventual added value provided by the integration of ALS data with remotely sensed multispectral information.

Acknowledgements

This work was partially supported by ITALID project (scientific coordinator: M. Marchetti), PRIN2009 "in_FLAMING" project ("Developing innovative models and techniques for integrated fuel management for fire prevention in Mediterranean and temperate forests", national coordinator: P. Corona), and by the University of Florence (Fondi di ricerca di Ateneo 2011, scientific coordinator: D. Travaglini). We wish to thank the Province of Pisa, BLOM ITALIA and the Tuscany Region for having provided Airborne Laser Scanning data for the study areas A, B and C, respectively.

In addition, we wish to thank two anonymous reviewers for their helpful suggestions and comments on an early version of the manuscript.

References

- Alberti G., Boscutti F., Pirotti F., Bertacco C., De Simon G., Sigura M., Cazorzi F., Bonfanti P. (2013) A LiDAR-based approach for a multi-purpose characterization of Alpine forests: an Italian case study. iForest, 6: 156-168. doi: http://dx.doi.org/10.3832/ifor0876-006.
- Axelsson P. (2000) *DEM generation from laser scanner data using adaptive TIN models*. International Archives of Photogrammetry and Remote Sensing, 33: 111-118 (B4/1; Part 4).
- Baatz M., Shäpe A. (2000) Multiresolution segmentation an optimization approach for high quality multi-scale image segmentation. In: Strobl, J., Blaschke, T., Greisebener, G. (Eds.), Angewandte Geographische Informationsverarbeitung XII. Wichmann-Verlag, Heidelberg, Germany, pp. 12-23.
- Barbati A., Chirici G., Corona P., Montaghi A., Travaglini D. (2009) *Area-based assessment of forest standing volume by field measurements and airborne laser scanner data*. International Journal of Remote Sensing, 30: 5177-5194. doi: http://dx.doi.org/10.1080/01431160903023017.
- Cartisano R., Mattioli W., Corona P., Scarascia Mugnozza G., Sabatti M., Ferrari F., Cimini D., Giuliarelli D. (2013) Assessing and mapping biomass potential productivity from poplar-dominated riparian forests: A case study. Biomass & Bioenergy, 54: 293-302.

- - doi: http://dx.doi.org/10.1016/j.biombioe.2012.10.023.
- Chirici G., Giuliarelli D., Biscontini D., Tonti D., Mattioli W., Marchetti M., Corona P. (2011) Large-scale monitoring of coppice forest clearcuts by multitemporal very high resolution satellite imagery. A case study from central Italy. Remote Sensing of Environment, 115: 1025-1033. doi: http://dx.doi.org/10.1016/j.rse.2010.12.007.
- Cho M.A., Mathieu R., Asner G.P., Naidoo L., van Aardt J., Ramoelo A., Debba P., Wessels K., Main R., Smit I.P.J., Erasmus B. (2012) *Mapping tree species composition in South African savannas using an integrated airborne spectral and Lidar system*. Remote Sensing of Environment, 125: 214-226. doi: http://dx.doi.org/10.1016/j.rse.2012.07.010.
- Ciancio O., Nocentini S. (2004). *The coppice forest. Silviculture, regulation, management*. In: Ciancio O., Nocentini S., Il bosco ceduo. Selvicoltura, assestamento, gestione. Accademia Italiana di Scienze Forestali, Firenze, Italy, pp. 679-701.
- Ciancio O., Corona P., Lamonaca A., Portoghesi L., Travaglini D. (2006) Conversion of clearcut beech coppices into high forests with continuous cover: A case study in central Italy. Forest Ecology and Management, 224: 235-240. doi: http://dx.doi.org/10.1016/j.foreco.2005.12.045.
- Clementel F., Colle G., Farruggia C., Floris A., Scrinzi G., Torresan C. (2012) *Estimating forest timber volume by means of "low-cost" LiDAR data*. Italian Journal of Remote Sensing, 44: 125-140. doi: http://dx.doi.org/10.5721/ItJRS201244110.
- Cohen W.B., Spies T.A. (1992) Estimating structural attributes of Douglas-Fir/Western Hemlock forest stands from Landsat and Spot Imagery. Remote Sensing of Environment, 41: 1-17. doi: http://dx.doi.org/10.1016/0034-4257(92)90056-P.
- Congalton R. (1991) A review of assessing the accuracy of classifications of remotely sensed data. Remote Sensing of Environment, 37: 35-46. doi: http://dx.doi.org/10.1016/0034-4257(91)90048-B.
- Corona P., Marchetti M. (2007) Outlining multi-purpose forest inventories to assess the ecosystem approach in forestry. Plant Biosystems, 141: 243-251. doi: http://dx.doi.org/10.1080/11263500701401836.
- Corona P., Fattorini L. (2008) *Area-based lidar-assisted estimation of forest standing volume*. Canadian Journal of Forest Research, 38: 2911-2916. doi: http://dx.doi.org/10.1139/X08-122.
- Corona P., Chirici G., McRoberts R.E., Winter S., Barbati A. (2011) *Contribution of large-scale forest inventories to biodiversity assessment and monitoring*. Forest Ecology and Management, 262: 2061-2069. doi: http://dx.doi.org/10.1016/j.foreco.2011.08.044.
- Corona P., Cartisano R., Salvati R., Chirici G., Floris A., Di Martino P., Marchetti M., Scrinzi G., Clementel F., Travaglini D., Torresan C. (2012) *Airborne Laser Scanning to support forest resource management under alpine, temperate and Mediterranean environments in Italy*. European Journal of Remote Sensing, 45: 27-37. doi: http://dx.doi.org/10.5721/EuJRS20124503.
- Estornell J., Ruiz L.A., Velázquez-Martí B., Hermosilla T. (2012). *Estimation of biomass and volume of shrub vegetation using Lidar and spectral data in a Mediterranean environment*. Biomass & Bioenergy, 46: 710-721. doi: http://dx.doi.org/10.1016/j.biombioe.2012.06.023.
- Falkowski M.J., Evans J.S., Martinuzzi S., Gessler P.E., Hudak A.T. (2009) Characterizing forest succession with lidar data: An evaluation for the Inland Northwest, USA.

- Remote Sensing of Environment, 113: 946-956. doi: http://dx.doi.org/10.1016/j.rse.2009.01.003.
- Ferraz A., Bretar F., Jacquemoud S., Gonçalves G., Pereira L., Tomé M., Soares P. (2012) 3-D mapping of a multi-layered Mediterranean forest using ALS data. Remote Sensing of Environment, 121: 210-223. doi: http://dx.doi.org/10.1016/j.rse.2012.01.020.
- Floris A., Clementel F., Farruggia C., Scrinzi G. (2010) *Forest volume estimate based on LiDAR data: a study for the Paneveggio Forest (NE Alpine area, Italy)*. Italian Journal of Remote Sensing, 42: 15-32. doi: http://dx.doi.org/10.5721/ItJRS20104232.
- Garfi V., Marchetti M. (2011) *Tipi forestali e preforestali della Regione Molise*. Edizioni dell'Orso S.r.l., Alessandria, Italy. ISBN 978-88-6274-280-1.
- Gebreslasie M.T., Ahmed F.B., van Aardt J.A.N. (2010) *Predicting forest structural attributes using ancillary data and ASTER satellite data*. International Journal of Applied Earth Observation and Geoinformation, 12S: S23-S26. doi: http://dx.doi.org/10.1016/j.jag.2009.11.006.
- González-Ferreiro E., Diéguez-Aranda U., Miranda D. (2012) Estimation of stand variables in Pinus radiata D. Don plantations using different LiDAR pulse densities. Forestry, 85: 2. doi: http://dx.doi.org/10.1093/forestry/cps002.
- González-Olabarria J.R., Rodríguez F., Fernández-Landa A., Mola-Yudego B. (2012) Mapping fire risk in the Model Forest of Urbión (Spain) based on airborne Lidar measurements. Forest Ecology and Management, 282: 149-156. doi: http://dx.doi.org/10.1016/j.foreco.2012.06.056.
- Hall R.J., Skakun R.S., Arsenault E.J., Case B.S. (2006) Modeling forest stand structure attributes using Landsat ETM+ data: Application to mapping of aboveground biomass and stand volume. Forest Ecology and Management, 225: 378-390. doi: http://dx.doi.org/10.1016/j.foreco.2006.01.014.
- Hengeveld G.M., Nabuurs G.-J., Didion M., Van den Wyngaert I., Clerkx A.P.P.M., Schelhaas M.-J. (2012) *A forest management map of European forests*. Ecology and Society, 17: 53. doi: http://dx.doi.org/10.5751/ES-05149-170453.
- Hollaus M., Wagner W., Eberhofer C., Karel W. (2006) *Accuracy of large-scale canopy heights derived from LiDAR data under operational constraints in a complex alpine environment*. ISPRS Journal of Photogrammetry and Remote Sensing, 60: 323-338. doi: http://dx.doi.org/10.1016/j.isprsjprs.2006.05.002.
- Hudak A.T., Lefsky M.A., Cohen W.B., Berterretche M. (2002) *Integration of lidar and Landsat ETM*+ *data for estimating and mapping forest canopy height.* Remote Sensing of Environment, 82: 397-416. doi: http://dx.doi.org/10.1016/S0034-4257.
- Hyyppä J., Hyyppä H., Leckie D., Gougeon F., Yu X., Maltamo M. (2008) *Review of methods of small-footprint airborne laser scanning for extracting forest inventory data in boreal forests*. International Journal of Remote Sensing, 29: 1339-1366. doi: http://dx.doi.org/10.1080/01431160701736489.
- INFC (2005) *Inventario Nazionale delle Foreste e dei Serbatoi Forestali di Carbonio*. Ministero delle Politiche Agricole Alimentari e Forestali, Ispettorato Generale Corpo Forestale dello Stato. Consiglio per la Ricerca e Sperimentazione in Agricoltura Unità di ricerca per il Monitoraggio e la Pianificazione Forestale (CRA-MPF).
- Jakubowski M.K., Li W., Guo Q., Kelly M. (2013) Delineating Individual Trees from Lidar Data: A Comparison of Vector- and Raster-based Segmentation Approaches.

- - Remote Sensing 5, 4163-4186. doi: http://dx.doi.org/10.3390/rs5094163.
- Ke Y., Quackenbush L.J., Im J. (2010) *Synergistic use of QuickBird multispectral imagery and LIDAR data for object-based forest species classification*. Remote Sensing of Environment, 114: 1141-1154. doi: http://dx.doi.org/10.1016/j.rse.2010.01.002.
- Lamonaca A., Corona P., Barbati A. (2008) *Exploring forest structural complexity by multi-scale segmentation of VHR imagery*. Remote Sensing of Environment, 112: 2839-2849. doi: http://dx.doi.org/10.1016/j.rse.2008.01.017.
- Lasserre B., Chirici G., Chiavetta U., Garfi V., Tognetti R., Drigo R., DiMartino P., Marchetti M. (2010) Assessment of potential bioenergy from coppice forests trough the integration of remote sensing and field surveys. Biomass & Bioenergy, 35: 716-724. doi: http://dx.doi.org/10.1016/j.biombioe.2010.10.013.
- Lefsky M.A., Cohen W.B., Spies T.A. (2001) An evaluation of alternate remote sensing products for forest inventory, monitoring, and mapping of Douglas-fir forests in western *Oregon*. Canadian Journal of Forest Research, 31: 78-87. doi: http://dx.doi.org/10.1139/cjfr-31-1-78.
- Lefsky M.A., Cohen W.B., Parker G.G., Harding D.J. (2002) *Lidar remote sensing for ecosystem studies*. Bioscience, 52: 19-30.
- Lefsky M.A., Hudak A.T., Cohen W.B., Acker S.A. (2005) *Patterns of covariance between forest stand and canopy structure in the Pacific Northwest*. Remote Sensing of Environment, 95: 517-531. doi: http://dx.doi.org/10.1016/j.rse.2005.01.004.
- Lillesand T.M., Kiefer R.W. (2000) Remote Sensing and Image Interpretation. Wiley.
- Lim K., Treitz P., Wulder M., St-Onge B., Flood M. (2003) *Lidar remote sensing of forest structure*. Progress in Physical Geography, 27: 88-106. doi: http://dx.doi.org/10.1191/0309133303pp360ra.
- Lu D. (2006) *The potential and challenge of remote sensing-based biomass estimation*. International Journal of Remote Sensing, 27: 1297-1328. doi: http://dx.doi.org/10.1080/01431160500486732.
- Magnussen S., Boudewyn P. (1998) Derivations of stand heights from airborne laser scanner data with canopy-based quantile estimators. Candian Journal of Forest Research, 28: 1016-1031. doi: http://dx.doi.org/10.1139/x98-078.
- Matthews J.D. (1991) Silvicultural Systems. Oxford University Press.
- McRoberts R.E., Wendt D.G., Liknes G.C. (2005) Stratified estimation of forest inventory variables using spatially summarized stratifications. Silva Fennica, 39: 559-571.
- Monnet J.M., Mermin E., Chanussot J., Berger F. (2010) *Estimation of forestry parameters in mountainous coppice stands using airborne laser scanning*. In: 10th International Conference on Lidar Applications for Assessing Forest Ecosystems (Silvilaser 2010), Freiburg, Germany, pp. 8.
- Montaghi A., Corona P., Dalponte M., Gianelle D., Chirici G., Olsson H. (2013) *Airborne laser scanning of forest resources: An overview of research in Italy as a commentary case study.* International Journal of Applied Earth Observation and Geoinformation, 23: 288-300. doi: http://dx.doi.org/10.1016/j.jag.2012.10.002.
- Morsdorf F., Mårell A., Koetz B., Cassagne N., Pimont F., Rigolot E., Allgöwer B. (2010) Discrimination of vegetation strata in a multi-layered Mediterranean forest ecosystem using height and intensity information derived from airborne laser scanning. Remote Sensing of Environment, 114: 1403-1415. doi: http://dx.doi.org/10.1016/

- j.rse.2010.01.023.
- Næsset E. (1997) Estimating Timber Volume of Forest Stands Using Airborne Laser Scanner Data. Remote Sensing of Environment, 61: 246-253. doi: http://dx.doi.org/10.1016/S0034-4257(97)00041-2.
- Næsset E. (2004) Practical large-scale forest stand inventory using a small-footprint airborne scanning laser. Scandinavian Journal of Forest Research, 19: 164-179. doi: http://dx.doi.org/10.1080/02827580410019544.
- Næsset E., Økland T. (2002) Estimating tree height and tree crown properties using airborne scanning laser in a boreal nature reserve. Remote Sensing of Environment, 79: 105-115. doi: http://dx.doi.org/10.1016/S0034-4257(01)00243-7.
- Nocentini S., Bottalico F., Faraoni L., Fiorentini S., Lisa C., Marchi E., Neri F., Paffetti D., Piemontese F.P., Puletti N., Travaglini D. (2011) *Piano di Gestione del Consorzio Forestale delle Cerbaie 2012-2026*. Università degli Studi di Firenze, Dipartimento di Economia, Ingegneria, Scienze e Tecnologie Agrarie e Forestali, Firenze, pp. 255.
- Ozdemir I., Karnieli A. (2011) Predicting forest structural parameters using the image texture derived from WorldView-2 multispectral imagery in a dryland forest, Israel. International Journal of Applied Earth Observation and Geoinformation, 13: 701-710. doi: http://dx.doi.org/10.1016/j.jag.2011.05.006.
- Ozdemir I., Donoghue D.N.M. (2013) *Modelling tree size diversity from airborne laser scanning using canopy height models with image texture measures*. Forest Ecology and Management, 295: 28-37. doi: http://dx.doi.org/10.1016/j.foreco.2012.12.044.
- Parker G.G., Russ M.E. (2004) The canopy surface and stand development: assessing forest canopy structure and complexity with near-surface altimetry. Forest Ecology and Management, 189: 307-315. doi: http://dx.doi.org/10.1016/j.foreco.2003.09.001.
- Pascual C., García-Abril A., García-Montero L.G., Martín-Fernández S., Cohen W.B. (2008) Object-based semi-automatic approach for forest structure characterization using lidar data in heterogeneous Pinus sylvestris stands. Forest Ecology and Management, 255: 3677-3685. doi: http://dx.doi.org/10.1016/j.foreco.2008.02.055.
- Pflugmacher D., Cohen W.B., Kennedy R.E. (2012) *Using Landsat-derived disturbance history (1972-2010) to predict current forest structure*. Remote Sensing of Environment, 122: 146-165. doi: http://dx.doi.org/10.1016/j.rse.2011.09.025.
- Picchio R., Maesano M., Savelli S., Marchi E. (2009) *Productivity and energy balance in the conversion into high forest system of a Quercus cerris L. coppice in Central Italy.* Croatian Journal of Forest Engineering, 1: 15-26.
- Simonson W.D., Allen H.D., Coomes D.A. (2012) *Use of an airborne lidar system to model plant species composition and diversity of Mediterranean oak forests*. Conservation Biology, 26: 840-850. doi: http://dx.doi.org/10.1111/j.1523-1739.2012.01869.x.
- Simonson W.D., Allen H.D., Coomes D.A. (2013) Remotely sensed indicators of forest conservation status: Case study from a Natura 2000 site in southern Portugal. Ecological Indicators, 24: 636-647. doi: http://dx.doi.org/10.1016/j.ecolind.2012.08.024.
- Terrasolid (2005) TerraScan User's Guide. Terrasolid Ltd., Jyvaskyla, Finland, 169 pp.
- Travaglini D., Bottalico F., Fiorentini S., Lisa C., Marchi E., Mottola S., Neri F., Nocentini S., Puletti N. (2011) *I boschi delle Cerbaie: gestione, conservazione e uso sostenibile*. Pacini Editore SpA, Ospedaletto (Pisa), Italy. pp. 61. ISBN 978-88-6315-297-5.
- UN/ECE-FAO (2000) Forest resources of Europe, CIS, North America, Australia, Japan

- - and New Zeland. Main Report. Geneva Timber and Forest Study Papers 17, Geneva, Switzerland.
- van Aardt J.A.N., Wynne R.H., Oderwald R.G. (2006) Forest volume and biomass estimation using small-footprint lidar-distributional parameters on a per-segment basis. Forest Science, 52: 636-649.
- van Leeuwen M., Nieuwenhuis M. (2010) *Retrieval of forest structural parameters using Lidar remote sensing*. European Journal of Forest Research, 129: 749-770. doi: http://dx.doi.org/10.1007/s10342-010-0381-4.
- Wasser L., Day R., Chasmer L., Taylor A. (2013) *Influence of Vegetation Structure on Lidar-derived Canopy Height and Fractional Cover in Forested Riparian Buffers During Leaf-Off and Leaf-On Conditions*. PLoS ONE, 8(1): e54776. doi: http://dx.doi.org/10.1371/journal.pone.0054776.
- Zimble D.A., Evans D.L., Carlson G.C., Parker R.C., Grado S.C., Gerard P.D. (2003) Characterizing vertical forest structure using small-footprint airborne Lidar. Remote Sensing of Environment, 87: 171-182.