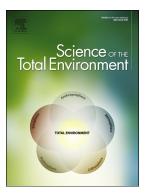
A framework for determining the life cycle GHG emissions of fossil marine fuels in countries reliant on imported energy through maritime transportation: A case study of South Korea



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A Framework for determining the life cycle GHG emissions of fossil

marine fuels in countries reliant on imported energy through maritime

transportation : A case study of South Korea

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#### Abstract

This research was motivated to address limitations in the current lifecycle assessment frameworks with the absence of proper guidelines for developing default lifecycle values of energies in consideration of supply chain activities and maritime transportation. Given this, it aims to evaluate the level of life cycle GHG emissions of heavy fuel oil, LNG, LPG and methanol as marine fuels produced and supplied in energy import-dependent countries, using South Korea as a case study. The analysis clearly shows that the impact of international shipping on Well-to-Tank (WtT) GHG emissions for energy carriers would be subject to several factors: propulsion system types, the quantify of energy transported, and the routes and distances of voyages. For instance, transportation

emissions from LNG carriers for LNG fuel vary significantly depending on the country of import, ranging from 2.26 g CO<sub>2</sub> eq./MJ (representing 12.2% of Well-to-Tank (WtT) emissions for Malaysia) to 5.97 g CO<sub>2</sub> eq./MJ (representing 33.3% of WtT emissions for Qatar). As a preliminary study, an enhancement on the quality of the input/inventory data is imperative for obtaining a reliability of results. Nevertheless, the comparative analysis of different fuels and life stages provides valuable insights for stakeholders to develop effective policies and energy refueling plans for reducing life cycle GHG emissions from marine fuels. These findings could also enhance the current regulatory framework and provide meaningful lifecycle carbon footprints of r. arine fuels for energy importing countries. The study results also strongly suggest that default values of GHG emission for different countries relying on energy imports via international maritime tran port should be further developed in consideration of the impact of regional differences, such as the analysis of LCA application on marine industry

Keywords: Life cycle assessment (LCA), Graniouse gas, Marine fuels, Default emission value, LCA regulatory framework, Case study, South Karea

# ABBREVIATIONS and TERMS

| AD                 | Average days at sea per year   |
|--------------------|--|
| AER                | Annual Efficiency Ratio  |
| A-LCA              | Attributional LCA modelling  |
| AML                | Kilowatts (kW) at average main and auxiliary engine load   |
| AP-X               | C3MR-Nitrogen hybrid cycle   |
| C3MR               | Propane precooled mixed refrigerant  |
| CfCO <sub>2</sub>  | Conversion factor between selected fuel consumptio $\gamma$ and CO <sub>2</sub> emission (t CO <sub>2</sub> /t fuel) |
| $\mathrm{CfCH}_4$  | Methane emission factor (t CH <sub>4</sub> /t fuel)  |
| CfN <sub>2</sub> O | Nitrous oxide emission factor (t N <sub>2</sub> O/t fuel)  |
| $\mathrm{CH}_4$    | Methane  |
| CII                | Carbon Intensity Indicator   |
| C-LCA              | Consequential LCA modelling  |
| CO <sub>2</sub>    | Carbon dioxide   |
| CO <sub>2</sub> eq | Carbon dioxide equivale. emissions   |
| DCS                | Data collection system   |
| DMR                | Dual mixed refrigerant   |
| Е                  | Estimated annual mission amount (t)  |
| EEDI               | Energy Ef iciei cy Design Index  |
| EEOI               | Energy Efficiency Operational Indicator  |
| EEXI               | Energy Efficiency Existing Ship Index  |
| EFGHG              | TtW GHG emission factor (t/t fuel)   |
| GHG                | Greenhouse gases   |
| FC                 | Fuel consumption (t)   |
| FCprimary          | Primary fuel consumption (t) for main engine   |
| FCpilot            | Pilot fuel consumption(t) for main engine  |
| FCaux              | Fuel consumption (t) for auxiliary engine  |

| GFS     | GHG Fuel Standard   |
|---------|---|
| HFO     | Heavy Fuel Oil  |
| ICAO    | International Civil Aviation Organization                                 |
| IMO     | International Maritime Organization                                       |
| LCA     | Life cycle assessment   |
| LCAF    | Lower carbon aviation fuels   |
| LHV     | Lower heating value   |
| LNG     | Liquefied natural gas   |
| LPG     | Liquefied petroleum gas   |
| LSD     | Low-speed diesel cycle engines  |
| LS-LPDF | Low-speed low-pressure dual-fuel engines                                  |
| LS-HPDF | Low-speed high-pressure dual-fuel engin .s                                |
| LSFO    | Low sulphur fuel oil  |
| MARPOL  | The International Convention of c. the Prevention of Pollution from Ships |
| MBMs    | Market-based measures   |
| MCR     | Maximum Continuous R'.ting  |
| MDO     | Marine diesel oil   |
| MGO     | Marine gas oil  |
| MSD     | Medium-s veed diesel cycle engines  |
| MS-LPDF | Medium-sp ed low-pressure dual-fuel engines                               |
| $N_2O$  | Nitrogen Oxides   |
| NCR     | Normal Continuous Rating  |
| NGLs    | Natural gas liquids   |
| SAF     | Sustainable aviation fuels  |
| SFOC    | Specific fuel consumption (g/kWh) for primary fuel at AM                  |
| SPOC    | Specific fuel consumption (g/kWh) for pilot fuel at AML                   |
| SEEMP   | Ship Energy Efficiency Management Plan                                    |
|         |   |

- SMR Single mixed refrigerant
- TtW Tank-to-Wake
- UNFCCC United Nations Framework Convention on Climate Change
- WtT Well-to-Tank
- WtW Well-to-Wake

# **1** Introduction

This Section provides an overview of the current issues with the IMO LCA regulatory framework on marine fuels. It also discusses challenges related to national and regional LCA regulatory frameworks for marine fuels. Lastly, the section outlines the research objectives and contributions of this study.

#### 1.1 Background

In international trade, maritime transport is considered the most efficient means to transport cargo and passengers overseas compared to other modes such as air, rail or total (Lenzen et al., 2023). Global GHG emissions from shipping activities account for above 370 of global annual CO2 emissions equivalent to the amount of annual CO<sub>2</sub> emissions in Germany, the world 6th largest emitter (Yuan et al., 2023). The emission level from international shipring is highly expected to increase from 50 to 250% by 2050 depending on economic and erargy scenarios (IMO, 2020). To improve ships' energy efficiency while reducing CO<sub>2</sub> levels, the Intrnational Maritime Organization (IMO) adopted the first-ever legally binding instrument ...tn.'d "Regulations on energy efficiency for ships" to MARPOL Annex VI in 2013. This package consists of technical and operational instruments which are known as the Energy Efficiency Design Index (EEDI) for new-built ships, and the Ship Energy Efficiency Management Plan (STCMP) for all ships. For evaluation of ship energy efficiency and CO<sub>2</sub> emissions from ship operation, the SEEMP proposes the use of a voluntary monitoring tool named Energy Efficiency Operational Indicator (EEOI) for each ship. With the adoption of the IMO Initial GHG Strategy in 2018, the IMO set a new target at  $CO_2$  reduction per transport work by at least 40% by 2030 compared to the 2008 level while pursuing toward 70% reduction by 2050, in parallel with the total 50 % GHG emission reduction in quantity by 2050 (IMO, 2018b). More recently, in 2021, the IMO has introduced a series of additional technical measures: the Energy Efficiency Existing Ship Index (EEXI) for existing ships and several different types of Carbon Intensity Indicators (CIIs) like Annual Efficiency Ratio (AER), Energy Efficiency Operational Indicator (EEOI), and other CIIs for implementing a rating scheme as operational measure (IMO, 2022d). As a basket of candidate mid-

term GHG reduction measures, the market-based measures (MBMs) have been also considered with combination of those technical measures such as a GHG fuel standard or IMO's carbon intensity measures (IMO, 2022b).

Despite the ambitious goals for reduction in GHG emissions established by the IMO, the introduction of alternative fuels to the shipping industry still remains at their brevity, as illustrated in Figure 1.

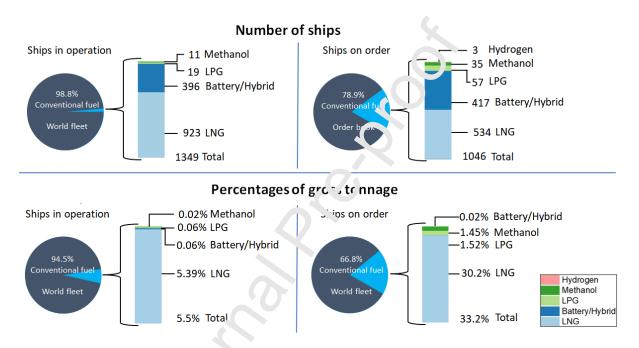
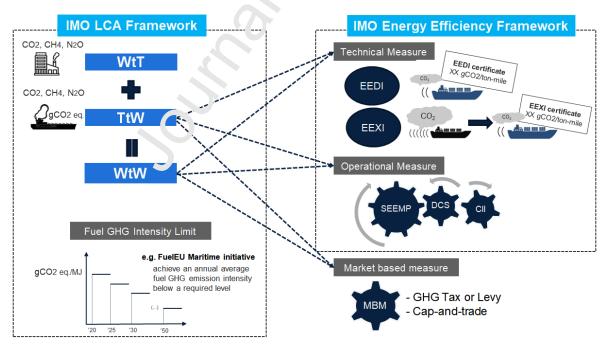


Figure 1 Alternative fue.' uptake in the world fleet by number of ships (DNV, 2022)

#### 1.1.1 Current issues on IMO LCA regulatory framework on marine fuels

The aforementioned IMO regulations and indicators aiming to curb air pollution are highly skewed by  $CO_2$  emissions from shipping activities. Despite vigorous efforts of the IMO and its Member States, waterborne transportation keeps contributing to the increment in emission levels by imposing greater burdens on other energy sectors. For example, the higher demand for hydrogen produced from natural gas for marine vessels reduces emissions in the shipping sector, but increases emissions in the energy production sector.

However, the current energy efficiency indexes established by the IMO, including EEDI, EEXI, AER and EEOI, have a limitation on evaluating the holistic GHG emission impact on marine fuels as they only consider emissions from ships and not the entire lifecycle of the fuel. To remedy current challenges while enhancing sustainability across sectors, IMO has proposed an urgent workstream for developing *"lifecycle GHG/carbon intensity guidelines for marine fuels"* with the primary work scope to develop Well-to-Tank (WtT) emission factor and Tank-to-Wake (TtW) default emission factors for marine fuels (IMO, 2018b). Nevertheless, due to the lack of unified LCA methodologies for marine fuels, a series of challenges are encountered in a process of developi.  $\sigma$  robust default emission values. These emission factors in these guidelines are likely to be used as leg n puts for enhancing the current IMO energy efficiency frameworks associated with EEDI, EE2 I, C I and other instruments. Figure 2 shows credible scenarios - identified in this research we is a well-to-wake (WtW), well-to-tank, or tank-to-wake scopes.



--- TtW/WtW into IMO instruments

Figure 2 Overview of credible IMO LCA applications to existing regulatory frameworks

The current IMO LCA framework considers introducing fuel certification schemes through new guidance on verification and certification for actual GHG emissions from different marine fuels. It is intended to grant flexibility in industrial choices and to incentivize the lower GHG emission fuels than the default values proposed in the LCA guidelines. This approach can allow fuel suppliers and ship operators to use actual emission values rather than default ones, whilst third-party verification and certification could also accommodate regional differences and specific feedstock. However, the absence of robust WtT and TtW emission default values, anticipated to be developed by a unified LCA methodology, may hinder the rule makers to apply the LCA a proach to the maritime industry. Robust default values for those marine fuels being widely used by the world fleet will play a crucial role in not only identifying the emission levels of marine fuels, i elping set achievable targets for decarbonization in the shipping sector but also incentivizing CHG reductions in lifecycle footprint across the fuel supply chain. In other words, especially for conventional marine fuels, it is an urgent issue to develop robust WtT and TtW emission value. through a unified methodology.

Another issue in determining robust defau' er ission values lies on the uncertainties and variations stemming from difference in geograph. al locations across nations and their supply routes and methods. Importantly, it needs to be claified how to tackle the geographical aspects in developing default emission factors for marine fuels for WtW emission. For instance, the lifecycle impacts of LNG supplied within LNG proclucing countries would be significantly different from the same volume of LNG imported from distant countries either by ship or road (IMO, 2022a). On the other hand, Norway et al argued that the GHG impacts caused by fuel import/export activities from and to different nations/regions are negligibly small, while insisting that larger variations in the GHG intensity could fall into the electricity mix across regions producing synthetic fuels (IMO, 2021). The identification and characterization of geographical differences in developing default emission factors for marine fuels is a crucial research gap that needs to be addressed.

#### 1.1.2 Challenges on national and regional LCA regulatory framework for marine fuels

From a national and regional regulation perspective, the European Union (EU), the most proactive bloc for curbing GHG, has established the FuelEU Maritime initiative that proposes the introduction

of lifecycle GHG impacts of maritime fuels; not only limiting the GHG intensity driven from onboard fuel consumptions, but also using a lifecycle analysis when assessing their GHG intensity (Marketa, 2022). A challenge can be observed through the fact that energy resources are not necessarily produced within the bloc or their own countries. It implies that there still require significant efforts to propose a unified LCA method and relevant data applicable to the EU initiative. Lack of clarification would leave greater challenges behind the rest of the world which are highly influenced by the EU's environmental policy. As a result, it is paramount to grasp the GHG footprint of those fuels by all countries; it is obvious that different countries have different levels of GHG impacts on marine fuels. For instance, South Korea's energy environment heavily relies on energy imports from handful energy producing countries. It accounts for approximately 98 % of the national fossil fuel demand, wherein 70% of Korea's petroleum was shipped from i.e. Middle East in 2019 (U.S. Energy Information Agency, 2019). Like South Korea, countries nat . re highly subject to energy import such as Japan, Taiwan or others, need wider energy publicles to achieve lifecycle decarbonization by tracking fuel types, applied production met<sup>1</sup> ods and supply chains of fuels from exporting places. In line with this, South Korea has introduced the national hydrogen economy roadmap for promoting hydrogen as a key national energy sou c ... targets not only decarbonization for all industry sectors but also the production and distribution of hydrogen. The roadmap also includes the future plan for overseas imports of green hydrog n from Australia and some others by 2030 (Kim et al., 2023). In this aspect, to assess the effectiveness of lifecycle policies aimed at promoting cleaner energy sources, it is imperative to establish baseline of lifecycle GHG emissions associated with conventional fossil fuels. Such a baseline can serve as a reference point for comparison with alternative fuels, allowing for a quantitative evaluation of the success of these policies in reducing overall GHG emissions.

Against the backdrop of mounting environmental concerns, policymakers have evinced growing interest in substituting conventional fossil-based fuels with alternative fuels in the marine sector. However, there has been a relative lack of effort from a policy perspective to determine the present state of life cycle GHG emissions, which is vital to evaluate the effectiveness of lifecycle policies towards cleaner energies. To address this gap, it is crucial to develop reliable default WtW emission

values of marine fuels and establish a database tailored to specific countries and circumstances. Such efforts can inform the formulation of future policies aimed at promoting sustainable and renewable fuels in the shipping sector.

#### 1.2 Research objective and contribution

This study aims to address two key research questions: How should default GHG life cycle emission values for marine fuels be determined in import-dependent countries? And, to what extent do regional or geographic differences impact the GHG lifecycle emission values of marine fuels, such as emissions from maritime transportation? The importance of these questions lies in the significant environmental impact of maritime transportation and its contribution to global GHG emissions, which are a key driver of climate change.

In this context, this paper contributes to the literature and ar ival of LCA to the shipping sector by achieving the following objectives: (1) to develop a model to calculate the present WtW emission level of conventional marine fuels produced and surplied in energy import-dependent countries, with a case study of South Korea; (2) to evaluate here the geographic difference influences on the overall WtW emission through case study; (3) a contribute to the development of robust default emission values based on fossil fuels in the LCA regulatory framework; and (4) to help a government establish future regulatory target based on with emission level of conventional fuels. To the best of the authors' knowledge, this study is first study is first study is contribute (i.e. South Korea). The contribution to future national and regional LCA regulatory framework of this study will be highlighted in discussion of Section 5.

To address the research questions, this paper is organized as follows: Section 2 reviews relevant literature that examined the life cycle GHG emissions of marine fuels, identifying gaps in the research related to the determination of default GHG life cycle emission values for marine fuels from the perspective of energy-importing countries like South Korea. Section 3 explains the main materials and methods used in this study. Section 4 presents the study's overall results, including an analysis of the

impact of maritime transportation on the lifecycle emission levels. In Section 5, we discuss the implications and limitations of our findings. Finally, Section 6 summarizes the key conclusions.

# **2** Literature review

Voluminous studies have been conducted to evaluate and compare the environmental performance of several marine fuels from a life cycle perspective as summarized in Table 1. In the literature, three key pollutants - CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O - were mainly considered within the GHG emission scope in the analysis since these three gases are the most significant contributors to overall GHG emissions. However, previous studies have self-demonstrated that the end no unified LCA approach to the fundamental methodology for estimating GHG emissions. As a result, research outcomes were subject to high ambiguity due to the lack of unified analysic scope, case study, assumptions, data usage, etc. In the LCA modelling, the selection of the attribution and consequential LCA (referred to as A-LCA and C-LCA) is a crucial element that greatly in fluences the determination of default emission values. Despite this fact, few LCA studies on marine fuels have clearly presented their methodologies; either attributional or consequential LCA is A-LCA modelling simply describes the immediate physical flows (e.g., energy, emissions and material) throughout the life cycle of a product and its subsystems. On the other hand, C-LCA is adding further considers how physical flows can be modified in response to changes in product demands with possible decisions (Chester and Cano, 2016; Earles and Halog, 2011; Moretti et al., 2022a; Moretti et al., 2022b; Vázquez-Rowe et al., 2014).

A review of literature on the topic of C-LCA and A-LCA reveals a diverse range of perspectives on the advantages and disadvantages of these two methodologies. Studies by (Brandão et al., 2014; Dale and Kim, 2014; Ekvall, 2019; Hertwich, 2014; Plevin et al., 2014; Prapaspongsa and Gheewala, 2017; Schaubroeck et al., 2021; Zamagni et al., 2012) have all contributed to this discourse, highlighting various pros and cons associated with each approach. Despite this, it is worth noting that the aviation sector similar to the shipping sector adopted a process-based attributional LCA approach along the

whole aviation fuel supply chain (ICAO, 2019). Also, the LCA regulatory framework for marine fuels should clearly define the unified selection of A-LCA and C-LCA.

With regard to variations in GHG emission impacts from the production and transportation of marine fuels across different regions, a number of studies have neglected to consider the sensitivity of GHG emissions from the maritime transportation and distribution of these fuels. These studies either focus on specific cases or do not address the impact of maritime transportation, making it challenging to use them as default values for policymaking. As demonstrated in Table 1, several studies have used default emission values for maritime transportation, as provided by he GREET or GaBi models. The fuel consumption in the GaBi model was estimated as linear or qua ratic polynomial functions of ship's deadweight. In addition, several studies have used defau t en ission values for energy transport by certain ships or from specific regions in their analyzer for instance, the study conducted by (Hwang et al., 2019) has limited scope, focusing privaria on MGO refined through crude oil imported from Qatar and the United States, and LUC from Qatar and the United States. It also employed genetic equations that show the r lationship between emission levels and distance traveled with a certain level of cargo. However, it's important to note that these emission models are not only based on the actual operation data of the snip but also on the overall energy import status of the importing countries. On the other ha. d, an LCA study shows that their LCA scope was subject to the exclusion of the emissions from tr. sportation and distribution process as those impacts were believed less than 2.5% of the tota GhG impacts of the study (Strazza et al., 2010). Despite this, it is still observed that there is a lack of reliable models derived from actual emission data from maritime transportation.

LCA studies on marine fuels tend to primarily focus on evaluating the environmental performance of alternative fuels in comparison to conventional fossil fuels, as highlighted in Table 1. In particular, a number of studies have been conducted to evaluate the environmental impacts of using LNG as a marine fuel, with varying results depending on the assumptions and scenarios used. (Lindstad and Rialland, 2020; Spoof-Tuomi and Niemi, 2020) compared the emission levels of LNG in shipping with conventional MDO. (Pavlenko et al., 2020) also argued that the increased use of LNG as marine

fuel would not contribute to emissions reductions and could potentially exacerbate the climate impacts of shipping if a 20-year global warming potential (GWP) was used as the impact assessment standard. Similar arguments were raised through some follow-up research. (Manouchehrinia et al., 2020) found that while natural gas (NG) engines with diesel cycles reduced GHG emissions by 2% compared to low sulphur petroleum diesel engines, other types of NG engines, such as lean-burn Otto cycle engines, resulted in 4% greater GHG emissions from a life cycle perspective. This study raises questions about WtT emissions and does not support the widespread adoption of NG fuel. (Hwang et al., 2019) also conducted a comparative analysis of WtW emissions from a ship using LNG and MGO. These studies all found that LNG is not always a cleaner option that conventional marine diesel oil (MDO) and it may not contribute to reducing GHG emissions. However, these studies appear to fall short of the objective of establishing the present level of  $1^{+}$ , cycle GHG emissions from marine fuels for the purpose of setting targets.

O STRONG

| Author(s)<br>and<br>publication<br>date | Type of Fuels   | GHG<br>emission<br>Scope  | Methodologic<br>al choice     | Geographical coverage for<br>production   | Maritime Transportation<br>(International)   |
|---|---|---|-------------------------------|---|--|
| (Strazza et<br>al., 2010)               | Methanol, Bio-<br>methanol, LNG,<br>Hydrogen in<br>Solid Oxide<br>Fuel Cells<br>(SOFC)                      | CO <sub>2</sub> ,<br>CH <sub>4</sub> , and<br>N <sub>2</sub> O                              | (no<br>indication)            | <ul> <li>Methanol: Global<br/>average</li> <li>LNG: Norwegian<br/>natural gas</li> <li>Hydrogen: average<br/>European plants</li> </ul>   | <ul> <li>(no indication*)</li> <li>*Excluded from the analysis (since fuel storage and bunkering phases, involving tank container and fuel distribution via pipeline, do not contribute for more than 2.5%)</li> </ul> |
| (Bengtsson<br>et al.,<br>2011)          | HFO, MGO,<br>LNG, GTL(gas-<br>to-liquid)  | $CO_2,$<br>$CH_4,$ and<br>$N_2O$  | Consequential                 | <ul> <li>Crude oil (HFO,<br/>MGO): European</li> <li>LNG, GTL: North Sea</li> </ul>   | <ul> <li>HFO, MGO: ELCD database</li> <li>LNG: 147K LNG carrier<br/>from Qatar to Rotterdam</li> <li>GTL: Product tanker from<br/>Qatar to Gothenburg</li> </ul>   |
| (Bengtsson<br>et al.,<br>2012)          | HFO, MGO,<br>Rapeseed<br>methyl ester<br>(RME),<br>Synthetic bio-<br>diesel (BTL),<br>LNG, Bio-LNG          | CO <sub>2</sub> ,<br>CH <sub>4</sub> , and<br>N <sub>2</sub> O                              | Consequential                 | <ul> <li>Crude oil (.1r)<br/>MGO), B', ':<br/>Europ :an</li> <li>LNC: .'ort', Sea</li> <li>RM: Bio-LNG:<br/>Swe, 'en</li> </ul>   | LNG: LNG carrier   |
| (Brynolf et<br>al., 2014)               | HFO, LNG, NG<br>based<br>Methanol, bio-<br>LNG, Bio-<br>methanol  | CO <sub>2</sub> ,<br>CH <sub>4</sub> , and<br>N <sub>2</sub> O                              | Consequential                 | <ul> <li>HFO. European</li> <li>LNG &amp; Methanol:<br/>natural gas from</li> <li>Norway and North</li> <li>African countries</li> <li>Bio-LNG &amp; Bio-<br/>methanol: Sweden</li> </ul> | <ul> <li>LNG: LNG carrier from<br/>North sea/Norway to<br/>Gothenburg</li> <li>Methanol: tanker from North<br/>Africa to Gothenburg via<br/>Rotterdam</li> </ul>   |
| (Gilbert et<br>al., 2018)               | HFO, MDO,<br>LNG,<br>Hydrogen,<br>Methanol, Bio-<br>LNG, Bio-<br>diesel, Straight<br>vegetable oil<br>(SVO) | CO <sub>2</sub> ,<br>CH <sub>4</sub> , a: d<br>N <sub>2</sub> O                             | A "ribu, onal                 | • All fuels: European   | • SVO: soybean grain from<br>transported by ship from<br>Argentina to Europe   |
| (El-<br>Houjeiri et<br>al., 2019)       | HFO, MGO,<br>LNG  | $CO_2,$<br>$CH_4,$ and<br>$N_2O$  | Attributional                 | <ul> <li>Crude oil (HFO and<br/>MGO): Saudi, North<br/>sea</li> <li>LNG: Australia, Qatar,<br/>USA</li> </ul>   | <ul> <li>LNG:138K LNG</li> <li>HFO: tanker from East Asia,<br/>Norway, UK to US, Japan<br/>refineries</li> </ul>   |
| (Hwang et<br>al., 2019)                 | MGO, LNG  | $\begin{array}{c} \text{CO}_2,\\ \text{CH}_4, \text{ and}\\ \text{N}_2\text{O} \end{array}$ | Attributional<br>(GaBi)       | <ul> <li>LNG: Qatar, USA</li> <li>Crude oil(MGO):<br/>Saudi Arabia, USA</li> </ul>  | <ul> <li>147K LNG Carrier from<br/>Qatar/USA to Korea</li> <li>57K Tanker from Saudi<br/>Arabia/ USA to Korea</li> </ul>   |
| (Thinkstep,<br>2019)                    | HFO, LSFO,<br>MGO, LNG  | $CO_2,$<br>$CH_4,$ and<br>$N_2O$  | Attributional<br>(GaBi)       | All fuels: Global   | Data from GaBi model   |
| (Sharafian<br>et al.,<br>2019)          | HFO, LNG  | $CO_2$ ,<br>$CH_4$ , and<br>$N_2O$  | (no<br>indication)<br>(GREET) | All fuels: Global<br>(based on North<br>America)  | Data from GREET model  |
| (Winebrak<br>e et al.,<br>2019)         | MDO,<br>Methanol, LNG   | $CO_2$ ,<br>$CH_4$ , and<br>$N_2O$  | (no<br>indication)<br>(GREET) | <ul> <li>All fuels: USA</li> <li>Methanol produced<br/>outside of North<br/>America</li> </ul>  | Methanol: ocean tanker<br>when methanol is produced<br>outside of North America  |

# Table 1 Previous LCA studies on marine fuels

| Author(s)<br>and<br>publication<br>date | Type of Fuels   | GHG<br>emission<br>Scope  | Methodologic<br>al choice     | Geographical coverage for production  | Maritime Transportation<br>(International)  |
|---|---|---|-------------------------------|---|---|
| (Lindstad<br>and<br>Rialland,<br>2020)  | HFO, LSFO,<br>MGO, LNG  | $CO_2,$<br>$CH_4,$ and<br>$N_2O$  | (no<br>indication)            | Literature review   | (no indication)   |
| (Perčić et<br>al., 2020)                | Methanol,<br>Dimethyl ether,<br>LNG,<br>Hydrogen,<br>Biodiesel,<br>Electricity  | $CO_2,$<br>$CH_4,$ and<br>$N_2O$  | (no<br>indication)<br>(GREET) | • USA: data from<br>GREET   | <ul> <li>Diesel: Tanker from Middle<br/>East</li> <li>Methanol: Tanker form<br/>Egypt</li> <li>LNG: LNG Carrier from<br/>Qatar</li> </ul> |
| (Spoof-<br>Tuomi and<br>Niemi,<br>2020) | MDO, LNG,<br>Bio-LNG  | $\begin{array}{c} \text{CO}_2,\\ \text{CH}_4, \text{ and}\\ \text{N}_2\text{O} \end{array}$ | (no<br>indication)            | <ul> <li>MDO: -</li> <li>LNG: extrac ~1 from<br/>North Sea</li> </ul>   | <ul> <li>MDO: dedicated tankers</li> <li>LNG: pipeline to the central<br/>hub in Finland</li> </ul>                                       |
| (Seithe et<br>al., 2020)                | HFO, LNG  | CO <sub>2</sub> ,<br>CH <sub>4</sub> , and<br>N <sub>2</sub> O                              | (no<br>indication)            | <ul> <li>HFO: P<sup></sup>Sta and<br/>North Sea</li> <li>LN<sup>-</sup>i: North sea and<br/>Carbon Algeria</li> </ul> | <ul> <li>HFO: pipeline to Rotterdam</li> <li>LNG: Pipeline and LNG carrier</li> </ul>   |
| (Pavlenko<br>et al.,<br>2020)           | HFO, LSFO,<br>MGO, LNG  | $CO_2,$<br>$CH_4,$ and<br>$N_2O$  | (no<br>indication)<br>(GREET) | UNG USA     Pet. Sleum (HFO,<br>LSFO, MGO): USA<br>and Canada   | Average international<br>transport value from<br>literature review  |
| (Manouche<br>hrinia et<br>al., 2020)    | LNG, MGO  | $CO_2,$<br>$CH_4,$ and<br>$N_2O$  | (r)<br>indica.'v.)            | <ul> <li>LNG: Canada</li> <li>MGO: data from<br/>GHGenius 5</li> </ul>  | <ul><li>LNG: pipeline</li><li>MGO: Data from GHGenius</li></ul>   |
| (Jang et al.,<br>2021)                  | HFO, LNG  | $\begin{array}{c} \text{CO}_2,\\ \text{CH}_4, \text{ and}\\ \text{N}_2\text{O} \end{array}$ | iı dication)                  | HFO, LNG: Literature     review   | (no indication)   |
| (Comer<br>and<br>Osipova,<br>2021)      | HFO, LSFO,<br>MGO, LNG  | CO <sub>2</sub> ,<br>CH <sub>4</sub> , N <sub>2</sub> U<br>and Blac                         | (no<br>indication)<br>(GREET) | <ul> <li>LNG: USA</li> <li>Petroleum (HFO,<br/>LSFO, MGO): USA<br/>and Canada</li> </ul>                              | Average international<br>transport value from<br>literature review  |
| (Malmgren<br>et al.,<br>2021)           | Bio-methanol,<br>Fossil<br>methanol,<br>Electro-<br>methanol<br>(eMeOH),<br>MGO | CO <sub>2</sub> ,<br>CH <sub>4</sub> , and<br>N <sub>2</sub> O                              | Attributional                 | <ul> <li>All fuels except<br/>MGO: North European</li> <li>MGO: Global average</li> </ul>                             | (no indication)   |
| (Lindstad<br>et al.,<br>2021)           | HFO,MGO,<br>LNG,<br>LPG,Methanol<br>Ammonia,Hydr<br>ogen                        | $CO_2,$<br>$CH_4,$ and<br>$N_2O$  | (no<br>indication)            | (no indication)   | (no indication)   |
| (Fernández<br>-Ríos et al.,<br>2022)    | Hydrogen  | $CO_2,$<br>$CH_4,$ and<br>$N_2O$  | (no<br>indication)            | (no indication)   | (no indication)   |
| (Chen and<br>Lam, 2022)                 | Diesel oil ,<br>Hydrogen  | $CO_2,$<br>$CH_4,$ and<br>$N_2O$  | (no<br>indication)            | <ul> <li>Diesel :<br/>European(ELCD<br/>database)</li> <li>Hydrgoen: Literature<br/>review</li> </ul>                 | • Diesel oil: ELCD and<br>Ecoinvent 3.6 database  |

| Author(s)<br>and<br>publication<br>date | Type of Fuels | GHG<br>emission<br>Scope           | Methodologic<br>al choice | Geographical coverage for production | Maritime Transportation<br>(International) |
|---|---------------|------------------------------------|---------------------------|--------------------------------------|--|
| (Seddiek<br>and<br>Ammar,<br>2023)      | HFO, Ammonia  | $CO_2$ ,<br>$CH_4$ , and<br>$N_2O$ | (no<br>indication)        | (no indication)                      | (no indication)                            |

#### 2.1 Research gaps

Numerous studies have aimed to compare the life cycle GHG emissions of conventional fossil fuels with alternative fuels. However, despite these efforts, the lack of the guidance or practice for establishing a baseline of life cycle GHG emission values, particula. Win countries that rely on energy imports, still remains. This research gap emphasizes the pressing need for default GHG emission values with a unified LCA approach and identificatio. of impact to importing countries due to regional or geographic differences values for enhancing the LCA frameworks of national and regional regulations. This study aims to fill the identified research gaps by providing a comprehensive understanding of the present level of life cycle. iHG emissions of conventional marine fuels produced and supplied in energy importing counties. The previous LCA studies, as shown in Table 1, have been inadequate in providing a clear answ... to the raised research question.

# 3 Material and method

The method was so proposed to compare the well-to-wake GHG emission impacts of HFO, LPG, LNG and methanol as a marine fuel bunkered in energy importing countries that the research findings could contribute to developing robust default WtW emission value for conventional marine fuels. In this research, South Korea was selected as the case region for the reason that it has great low primary energy production and heavily relies on energy imports from other countries. **Error! Reference source not found.** Figure 3 shows the methodology applied in this study. Life cycle assessment (LCA) is used to evaluate the well-to-wake environmental impact of fuel products and their process activities.

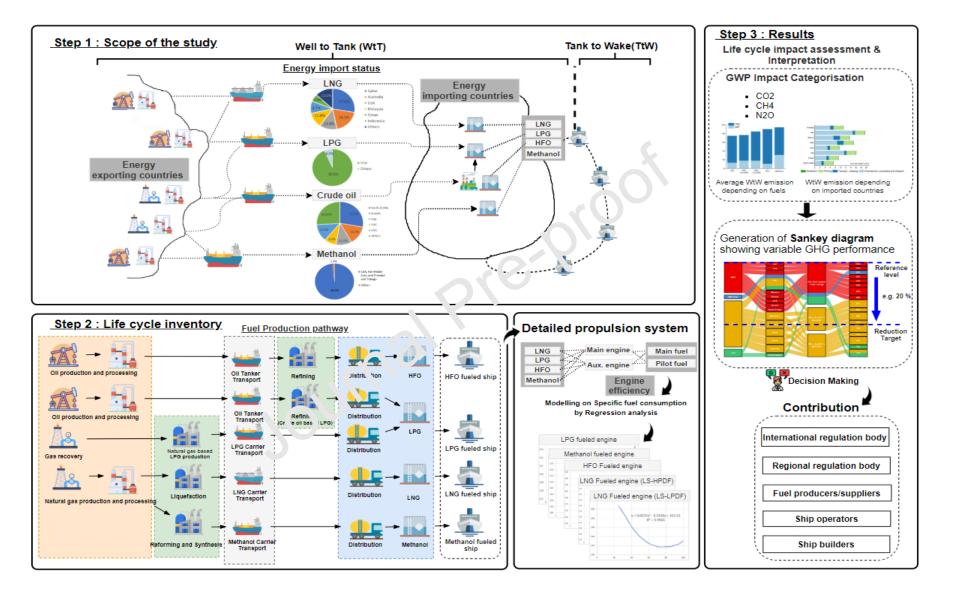


Figure 3 Life cycle assessment framework and methodology flowchart

#### 3.1 Scope of the study

Section 3.1 clarifies the scope of this study. First, the ship fuels to be considered in this study are reviewed, and the methodology to ensure that the study is properly conducted is specified.

#### 3.1.1 Selected marine fuels: HFO, LNG, LPG and methanol

Four fuels, namely heavy fuel oil (HFO), liquefied natural gas (LNG), liquefied petroleum gas (LPG), and methanol, have been identified as suitable candidates for the dev lopment of reliable default GHG emission values due to their high technical readiness levels (TRUs), 'echnical maturity, and robust safety standards, in contrast to other fuels with lower TRLs (Full-sailing Cepeda et al., 2019; Kouzelis et al., 2022; Ortega et al., 2021; Turnau et al., 2020). Amongst conventional fuels, this study solely focuses on HFO, as it remains the dominant fuel in the ship-ring sector, accounting for 79% of total fuel consumption (IMO, 2020)

Safety regulations for alternative fuels tend 'o ' e developed and applied after technical maturity and feasibility of technology on board ships and secured. Over the past decade, in order to minimize the risk to the ship and its crew due to the na ure of the fuels involved (see Table 2). In addition, IMO has developed safety regulations for LNG, LPG and methanol for their application to merchant vessels as shown in Table 3.

|                           | HFO       | LNG       | LPG                 | Methanol           |
|---------------------------|-----------|-----------|---------------------|--------------------|
|                           |           |           | (Propane/Butane)    |                    |
| Carrying temperature (°C) | Ambient   | -162      | -42                 | Ambient            |
| Flash point (°C)          | >61       | -188      | -104                | 11–12              |
| Auto ignition temperature | 230       | 537       | 410-580 (depending  | 470                |
| (°C)                      |           |           | on the composition) |                    |
| Flammability limits       | 0.6–7.5   | 5-15      | 1.8-10.1            | 6.7–36             |
| (volume % in air)         |           |           |                     |                    |
| Toxicity                  | Not toxic | Not toxic | Not toxic           | Low acute toxicity |
|                           |           |           |                     | (dangerous for     |
|                           |           |           |                     | humans)            |
| Energy density (MJ/L)     | 35.2      | 21.2      | 26.7                | 14.9               |

Table 2 Overview of kov p. ope ties of different fuels (Ampah et al., 2021; DNV, 2019a)

Table 3 The status of development for safety regulations on alternative fuels

| Type of Fuel | Safety regulations | Effective date |
|--------------|--------------------|----------------|
|              |                    |                |

| 1 | LNG      | MSC.391(95) International Code of Safety for        | January 2017          |
|---|----------|---|-----------------------|
|   |          | Ships using Gases or other Low-flashpoint Fuels     |                       |
|   |          | (IGF Code) - Part A-1                               |                       |
| 2 | Methanol | MSC.1/Circ.1621: Interim guidelines for the         | December 2020         |
|   |          | safety of ships using methyl/ethyl alcohol as fuel  |                       |
| 3 | LPG      | Interim guidelines for the safety of ships using    | June 2023             |
|   |          | LPG fuels   | (Approval at MSC 107) |
| 4 | -        | MSC.1/Circ.1647: Interim guidelines for the         | June 2022             |
|   |          | safety of ships using fuel cell power installations |                       |

#### (a) Heavy fuel oil (HFO)

Since the 1960s, HFO has been used primarily on board ships and a but 99% of the world fleets use conventional fuel by internal combustion engines as shown in Figure 1. Since it is a residual fuel from the distillation and cracking of petroleum, it contains various compounds such as sulphur and nitrogen that create more pollutants than other fuels (Carvalho e. ar. 2023). SO<sub>x</sub> and NO<sub>x</sub> emissions from international shipping were estimated 10-15% of the total global anthropogenic emissions and it accounts for about 3% of global CO<sub>2</sub> emissions (smith et al., 2015). These air pollutants from ships can cause serious health and environment <sup>1</sup> arm. In response, IMO has implemented stringent regulations related to sulphur content in the fuel oil and NO<sub>x</sub> emission from engines.

### (b) Liquefied natural gas (I. NG,

LNG-fueled ships are gradual, increasing to reduce GHG. LNG is mainly composed of methane  $(CH_4)$  which becomes liquid at a temperature of -160°C at atmospheric pressure (Holzer et al., 2017). In comparison to conventional fuels such as HFO, natural gas is known as a cleaner fuel to reduce  $SO_x$ ,  $NO_x$ , and Particulate Matter (PM) (Bilgili, 2021). As a result, the 4th IMO GHG Study shows a 150% increase in methane emissions from ships between 2012 and 2018 due to the increased number of LNG-fueled vessels. However, LNG has been challenged due to the emissions of unburned methane at the combustion process from LNG-fueled engines (Mavrelos and Theotokatos, 2018).

#### (c) Liquefied petroleum gas (LPG)

More than 71 ships using LPG fuel have been built or converted by 2022 (WLPGA, 2021). LPG is any mixture of propane ( $C_3H_8$ ) and butane ( $C_4H_{10}$ ) in liquid form. Unlike LNG, it can be easily handled at ambient temperature with a pressure of 10-20 bar (WLPGA, 2017). Using LPG as a fuel can reduce GHG emissions and other pollutants to the atmosphere compared to conventional fuels. It can eliminate SO<sub>x</sub> emissions and reduces GHG by approximately 17% compared to HFO (Brinks and Chryssakis, 2017). For a two-stroke diesel engine, NO<sub>x</sub> emissions can be expected to be reduced by 10-20% compared to HFO (Pham et al., 2021). LPG could serve as a transition fuel to ammonia since the energy conversion system fitted onboard using LPG may be con putible with system for ammonia through its minor modification (DNV, 2019b).

#### (d) Methanol

54 methanol-fueled ships are already in operation of on order according to DNV's online platform (DNV, 2022). Methanol (CH<sub>3</sub>OH) is a simple alcohol which is currently used to propel commercial vessels. It can just utilize existing shore into aructures for conventional fuel with small and minor modifications (de Fournas and Wei, 2022). From an infrastructure standpoint, methanol is already available worldwide for distribution and thorage capacity (Sun and Aziz, 2021). 88 out of the largest international ports already have the methanol bunkering infrastructure in place (Martin, 2021). It is mainly produced from natural gas but can also be made from coal and various agricultural wastes. Methanol is easier to header than LNG as it is liquid at atmospheric conditions (Thaler et al., 2022). However, it is a low flash point and toxic fuel, so it is important to handle it onboard with caution (Zhao et al., 2021).

#### 3.1.2 Methodological choices

The WtW emissions for selected fuels consist of two parts: WtT and TtW. This study was to consider WTW analyses for ship constructions and infrastructure development for fuel production out of scope. Instead, it was focused on the cradle-to-grave assessment of marine fuels under the current infrastructure and existing ships. For comparison between fuels investigated, the functional unit (g

 $CO_2$  eq./MJ) was defined as  $CO_2$  equivalent emission grams per MJ of produced fuel. For the evaluation of GHG emissions, three representative GHGs - $CO_2$ , CH<sub>4</sub>, and N<sub>2</sub>O- were investigated. The calculation of GHG emissions is based on the standard of a one-hundred-year time horizon impact assessment (GWP100), which is widely adopted as a standard method in international policy practices. Their global warming potentials were proposed at 1, 28, and 265 times, when compared to  $CO_2$ , respectively and those potentials were adopted to calculate the total  $CO_2$ -equivalent emissions referring to the overall WtW GHG emissions in this study. Due to geopolitical reasons, all the fuels in South Korea are imported via only maritime transportation. The stati-trical data for 2020 was used as a mostly updated one that could provide national energy data for the evaluation. The methodological choices applied in this study are detailed in Table 4. This study employs an attributional approach to evaluate the environmental implicit on fuels, which involves analyzing the resources and emissions that are directly associated with the production and utilization of the fuel.

| Methodological item   | Selection  |  |  |  |
|-----------------------|--|--|--|--|
| Selected fuels        | HFO, LNG, LPC and methanol   |  |  |  |
| Impact category       | Global War air is Dotential with the 100-year time frame                             |  |  |  |
| GHG emissions scope   | $CO_2, CH_2$ and $A_2O$  |  |  |  |
| Geographical coverage | From pt. ducing countries to South Korea   |  |  |  |
| System boundaries     | WtT p. rt covers fuel production and its transportation to ships onboard             |  |  |  |
|                       | while be ItW scope covers the emission   |  |  |  |
| Functional unit       | $CO_2 \circ \sigma$ ivalent emission grams per MJ of produced fuel (g $CO_2$ eq./MJ) |  |  |  |
| LCA methodology       | Atti butional approach (A-LCA)   |  |  |  |
| Allocation method     | Energy based allocation is prepared  |  |  |  |

| Table 4 | Summary | of the | methodological | chc ces |
|---------|---------|--------|----------------|---------|
|         |         |        |                |         |

#### 3.2 Data collection and generation

In this section, relevant data is collected and safeguarded to ensure accurate and precise research within the scope defined in Section 3.1. The analysis was performed by segmenting the fuel lifecycle into two stages: the upstream stage encompassing production to storage (Well-to-Tank), and the downstream stage (Tank-to-Wake), representing the use aspect.

#### 3.2.1 Well-to-Tank Inventory Analysis

Figure 4 summarizes the pathway and their ratios of the proposed fuels from the producing countries to South Korea. LNG, LPG, Crude oil, and methanol are imported from overseas, but crude oil-based LPG is produced in domestic refineries and small amounts of natural gas are also produced in South Korea.

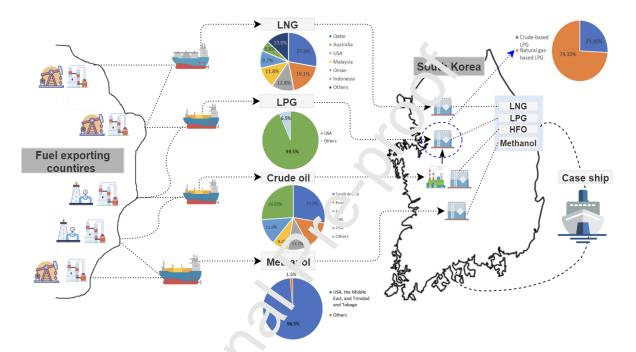


Figure 4 Imports Crude oil, LNG, LPG and methanol to South Korea

The WtT GHG emissions , re determined by geological characteristics, transportation means, and fuel production methods (Manouchehrinia et al., 2020). The fuel production pathways for HFO, LPG, LNG and methanol were identified and modelled as shown in Figure 5 in order to assess their WtW GHG emissions. The pathways were also categorized into four steps: Efs, Ec, Et, and Ed, as shown in equation (2). This was done to standardize the approach and enable comparison of the pathways. Furthermore, WtT inventory was also compiled, considering the specifics of energy producing/exporting countries. The more detailed data used in the analysis are summarized in the supplementary material. In this study, the weighted average WtT GHG emissions associated with imported energy sources or fuels were estimated using the following equation.

• Average WtT GHG emission = 
$$\frac{\sum_{i}^{n \ countries}(P_i \times WtT_i)}{\sum_{i}^{n \ countries}P_i}$$
 (1)

• 
$$WtT_i = Efs + Ec + Et + Ed$$
 (2)

where;  $P_i$ : the percentage(%) of a specific energy source or fuel imported to South Korea from a particular country, WtT<sub>i</sub>: WtT emission value(gCO<sub>2</sub>eq./MJ) of fuels imported from a particular country, based on production pathway in Figure 5, Efs : emissions associated with feedstock extraction, recovery, and transport, excluding international maritime transportation, Ec : emissions resulting from the conversion of the feedstock to the final fuel product, as well as emissions associated with the transportation and storage of the finished fuel. excluding international maritime transportation for the feedstock or the finished fuel, Ed : emissions associated with the distribution phase, encompassing local delivery, retail storage, and bunkering

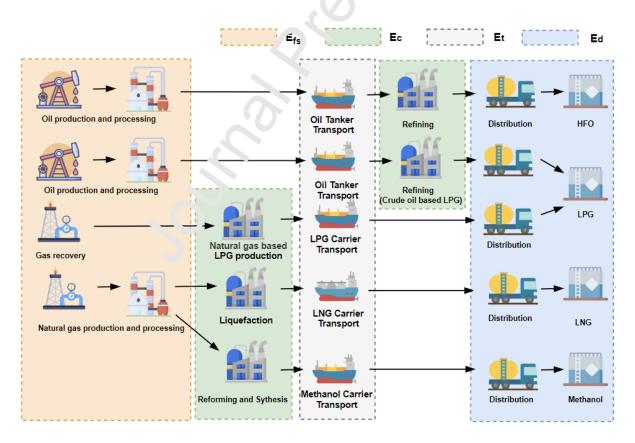


Figure 5 Production pathway modeling for HFO, LNG, LPG, and methanol in South Korea

#### (a) Heavy fuel oil (HFO)

South Korea imports crude oil from overseas via maritime transportation and HFO is produced from this oil production through domestic refineries. The data of crude oil imported were collected based on the eight major crude oil producing countries accounting for approximately 84.2 % of the total crude oil imports as listed: Saudi Arabia, Kuwait, Iraq, USA, UAE, Mexico, Iran and Russia (Korea Petroleum Association, 2022).

Case-specific emissions data for crude oil production across different countries was obtained from the "Oil Production Greenhouse Gas Emissions Estimator" (OPGEE), v. ich is currently considered the most reliable public data (El-Houjeiri et al., 2019; Masnadi et al., 2018, Thinkstep, 2019). It provides GHG emission data for some specific countries corresponding to production processes of both conventional and alternative fuels.

However, in relation to crude oil imports, OPGEE mc  $d^{-1}$  was adopted using identical default emission values for crude oil transportation through ocea. tanker with 250,000 tons for carriage and 8,000 miles for operation. In this study, the actual side operation data from a Korean shipping company was used to estimate the emissions by maritime transportation. Table 5 shows the details obtained from operational data (namely Abstract LCG) that includes the information of ship voyages such as calling in and out ports, navigation distance, average speed, and fuel consumption (MOL, 2022). The GHG emissions of crude oil, LNG and LPG during the maritime transportation were estimated with the following equation. The determination of emission factors(EF<sub>GHG</sub>) is achieved through the integration of per unit emissions of various greenhouse gases in relation to fuel consumption, as well as the consideration of their corresponding global warming potentials.

• 
$$Et = \frac{\sum_{i}^{n \text{ fuel}} \sum_{j}^{m \text{ engine}} M_{i,j} \times EF_{GHG}}{M_{cargo} \times LHV}$$
(3)

• 
$$EF_{GHG} = Cf_{CO2} + Cf_{CH4} \times GWP_{CH4} + Cf_{N2O} \times GWP_{N2O}$$
(4)

where;  $E_t$  : emissions during the maritime transportation(CO<sub>2</sub> eq./MJ),  $M_{i,j}$  : consumption of the specific fuel i oxidized in consumer j (t fuel) of ships,  $M_{cargo}$  : Mass of the specific cargo carried(t cargo), LHV : Lower heating value of fuel<sub>i</sub>, EF<sub>GHG</sub>: GHG emission factor (t/t fuel) corresponding to

each fuel,  $Cf_{CO2}$ : the conversion factor between selected fuel consumption and  $CO_2$  emission (t  $CO_2/t$  fuel),  $Cf_{CH4}$ : the methane emission factor (t  $CH_4/t$  fuel),  $Cf_{N2O}$ : the nitrous oxide emission factor (t  $N_2O/t$  fuel),  $GWP_{CH4}$ : Global warming potential for  $CH_4$ , equals to 28 for 100-year time horizon,  $GWP_{N2O}$ : Global warming potential for  $N_2O$ , equals to 265 for 100-year time horizon.

The data of emissions incurred in a refinery to distillate from crude oil to HFO was from published data which adopted process-level allocation method to calculate the refining energy use of individual petroleum products (Choi et al., 2020; Jang and Song, 2015).

|           | Maritime transportion |              |                    |                             |  |  |  |
|-----------|-----------------------|--------------|--------------------|-----------------------------|--|--|--|
| Type of   | Applied               | Main fuel    | Cargo car city m3) | Geographical coverage       |  |  |  |
| cargo     | propulsion            | types        |                    | (Imported countries)        |  |  |  |
|           | system                |              |                    |                             |  |  |  |
| Crude oil | LSD                   | HFO          | 30 )K              | Saudi Arabia, Kuwait, Iraq, |  |  |  |
|           |                       |              |                    | UAE, USA, Mexico and Iran   |  |  |  |
| LNG       | Steam turbine         | HFO,         | 125K               | Malaysia                    |  |  |  |
|           | (Main boiler)         | LNG          |                    | Indonesia                   |  |  |  |
|           |                       |              |                    | Australia                   |  |  |  |
|           |                       |              | 135K               | Qatar                       |  |  |  |
|           |                       |              |                    | Oman                        |  |  |  |
|           | LS-HPDF               |              | 174K               | USA                         |  |  |  |
| LPG       | LSD                   | <u>, 'FO</u> | 82K                | USA                         |  |  |  |
| Methanol  | LSD                   | HFO          | 55K                | USA                         |  |  |  |

Table 5 Ship specifications for maritime transportation

Note: Low-speed diesel cycle engines (LSD), Low-speed high-pressure dual-fuel engines (LS-HPDF), Low-speed low-pressure dual-tuni engines (LS-LPDF), Medium-speed low-pressure dual-fuel engines (MS-LPDF)

#### (b) LNG

Despite a small domestic production, most of the LNG consumed in South Korea is imported from several countries. Given this, this paper assumes all LNG to be produced overseas and imported to South Korea with the following ratios: Qatar (27%), Australia (19%), USA (14%), Malaysia (12%), Oman (9%), Indonesia (6%), Other (13%) (KOGAS, 2020).

The data pertaining to the emissions and energy consumptions (such as electricity, diesel oil and natural gas) during the production, liquefaction, and transportation of natural gas were obtained from GaBi database and the research products of the Natural Gas & bio Vehicle Association (NGVA)

(Schuller et al., 2019; Schuller et al., 2017). Data on fuel consumption and its emissions for LNG carrier transport were calculated during the actual ship's operation for approximately 100 voyages, based on the specifications in Table 5. The methane loss data for LNG terminal operations and LNG bunkering were included in distribution phase(Ed). These data were taken from the NGVA study and GaBi database (Schuller et al., 2019; Schuller et al., 2017). It should be noted that unlike other fuels, methane loss in the distribution phase only occurs during LNG terminal and bunkering operations.

#### (c) LPG

South Korean LPG mainly consists of crude oil-based LPG (75.9.5) made from domestic refineries and natural gas-based LPG (74.1%) imported from oversees (COSIS, 2022). The natural gas-based LPG imported from the USA accounts for about 93% while the rest is from other countries (Korea Petroleum Association, 2022). To simplify the analysis in this study, all imported LPG was assumed to be made from natural gas liquids (NGLs) produced in the USA.

For the inventory analysis of crude of based LPG, the emission data from domestic refineries, including liquefaction, were mainly cottaned from past publication (Choi et al., 2020). The actual operation data from the Korean shipping company was used to estimate the emissions by international transportation. The average Wt1 GhG emissions values of LPG were obtained as a weighted average based on the amount of domestic LPG produced and imported in 2019.

#### (d) Methanol

Given the mature methanol synthesis pathway, which accounts for 90% of global methanol demand, it is assumed that all methanol will be produced from natural gas. Its key process includes steam reforming of natural gas and methanol synthesis reactor (Adnan and Kibria, 2020). Approximately 98% of methanol shipped to South Korea originated from the USA, the Middle East, and Trinidad and Tobago (ARGUS, 2023). However, due to a lack of available data, the WtT GHG emissions for methanol production in the USA specifically were adopted from the latest version of the GREET

model (Wang et al., 2021). The feedstock for methanol production in the USA consists of 74.7% natural gas from shale production and 25.3% from conventional recovery practices. The transportation emissions data for methanol carrier was assumed based on the average emissions for a 55K methanol carrier traveling between the USA and South Korea.

#### 3.2.2 Tank-to-Wake Inventory Analysis

In the Tank-to-Wake analysis, which represents the downstream stage, a comprehensive analysis was conducted by subdividing it into case ship selection, propulsion engine, auxiliary engine, fuel consumption calculation/estimation, and GHG emissions assessment.

#### 3.2.2.1 Selection of case ship

In order to compare TtW and WtW emissions of the second fuels, a bulk carrier of M/V Ilshin Green Iris built in 2018 was selected. The main particulars were obtained from the Korean Register while ship's operational data corresponding to six months were collected from ILSHIN SHIPPING Co., Ltd. Table 6 summarizes general specifications and operational profiles of the selected ship.

| Specification                          | Details   |  |
|--|---|--|
| Length $\times$ Breadth $\times$ Depth | 190.6.` m × 32.26 m× 17.3 m                                     |  |
| Service speed                          | 14 knots  |  |
| Deadweight                             | 50 555 tons   |  |
| Main Engine                            | Low-speed high-pressure dual-fuel engines (LS-HPDF) (6G50ME-GI) |  |
| MCR/NCR 🔪                              | MCR: 7,250 kW × 88.7 RPM  |  |
|  | NCR: 5,597 kW × 81.4 RPM  |  |
| LNG fuel tank                          | $500 \text{ m}^3$   |  |
| Cruising range                         | Abt. 600 miles per one voyage                                   |  |
|  | from Donghae port to Gwangyang port located in south Korea      |  |
| Average main engine load               | 72.5 % of MCR   |  |

Table 6 Specifications and operational profiles of the case ship

The same specifications as MCR (7,250 kW  $\times$  88.7 RPM) and NCR (NCR: 5,597 kW  $\times$  81.4 RPM) were used in selecting the engine type using HFO, LPG, LNG, and methanol.

#### 3.2.2.2 Calculation of TtW GHG emissions and average GHG intensity

In order to estimate the annual GHG emissions from onboard ships, fuel consumption and GHG emission factors should be sought. The fuel consumptions are multiplied by the TtW emission factors with the following equation, which has been applied to IMO instruments: IMO DCS, CII and EEOI. The determination of TtW emission factors( $EF_{GHG}$ ) is achieved through the equation (4).

• 
$$E = \sum_{i}^{n \text{ fuel}} \sum_{j}^{m \text{ engine}} M_{i,j} \times EF_{GHG}$$
(5)

The average GHG intensity of the fuel used on board ships are estimated using the following equation.

• 
$$Et = \frac{E}{\sum_{i}^{n} M_{i} \times LHV_{i}}$$
(6)

where; E: Estimated annual emission amount (t),  $M_{i,j}$ : Consumption of the specific fuel i oxidized in consumer j (t fuel) of ships,  $EF_{GHG}$ : TtW GHG emission factors average GHG intensity of the fuel used (CO<sub>2</sub> eq./MJ) and LHV : Lower heating value of fuel<sub>i</sub>

#### 3.2.2.3 Types of propulsion engines

#### (a) LNG Fueled engine: low-speed h o'1-pressure dual-fuel engines(LS-HPDF)

The MEGI engine (M-type Electronically Controlled Gas Injected engine), whose first vessel was delivered in 2016, applies the all sel cycle with non-premixed combustion. As opposed to the Otto-cycle combustion, natural gas rule with high pressure of about 300 bar is injected into the combustion chamber together with 5% amount of pilot fuel to ensure optimal combustion (Domić et al., 2022).

#### (b) LNG Fueled engine: low-speed low-pressure dual-fuel engines (LS-LPDF)

The dual fuel engine employs low-pressure gas fuel and operates on the Otto-cycle combustion process utilizing premixed fuel/air and a relatively high air-to-fuel ratio. A minimal volume of pilot fuel, comprising approximately 1% of full load fuel consumption, is required for ignition of the premixed fuel/air. Unlike high-pressure gas injection engines, the utilization of low gas pressure,

approximately 13 bar, is sufficient to attain a homogenous air/gas mixture across the full range of engine loads due to the injection of gas at the start of compression (WIN GD, 2021).

#### (c) LPG-fueled engine and Methanol fueled engine

The dual-fuel ME-LGI engine is a novel propulsion system that is designed to operate using lowflashpoint liquid fuels, as opposed to the gaseous fuels utilized by LNG-fueled engines. The ME-LGI engine, which employs a diesel combustion cycle, is available in various versions, each optimized for a specific low-flashpoint fuel type (MAN Energy Solutions, 2020). . . r example, the ME-LGIP and ME-LGIM engines are specifically engineered for operation wit<sup>4</sup>. LPG and methanol fuels, respectively. The methanol fuel supply system employed in the ME-LGIM engine is similar in design to that of conventional heavy fuel oil (HFO) engine . The fuel is supplied at a pressure of approximately 10 bar and the injection pressure at the engine combustion cylinder is around 500-550 bar (MAN Energy Solutions, 2014). In contrast the ME-LGIP engine utilizes LPG fuel that is supplied at a pressure of 50 bar and is fueller pressurized to 600-700 bar by the high-pressure hydraulic oil system. Additionally, the ME-<sup>T</sup> GIP engines have the capability to operate in gas mode with minimal usage of pilot oil, typically and the operating on methanol with not only a low cetane number but also low self-ignition qual. (MAN Energy Solutions, 2021).

Based on selected propulsion angines above, the summary of emission data from propulsion engines can be found in Table 7. For all CO<sub>2</sub> emission factor between fuel consumption and CO<sub>2</sub> emission, it was taken by the 2018 EEDI Guidelines (IMO, 2018a). The emission factor for CH<sub>4</sub> varies by engine type. For LNG-fueled engines, the factors were chosen based on values of 2.50 g/kWh for LS-LPDF and 0.20 g/kWh for LS-HPDF (Pavlenko et al., 2020). They are weighted to represent E2 or E3 test cycles in IMO NO<sub>x</sub> Technical Code. The other CH<sub>4</sub> and N<sub>2</sub>O emission factors were from the 4<sup>th</sup> IMO GHG study. In particular, the factors for methanol were considered as 10% of HFO (IMO, 2020). Due to the lack of data for the LPG fuel, the LPG emission factors for CH<sub>4</sub> and N<sub>2</sub>O were substituted for those for methanol fuel.

|   | LNG Fueled engine   |   | LPG Fueled<br>engine                       | HFO fueled<br>engine                       | Methanol<br>engine                            |  |
|---|---|---|--|--|---|--|
| Main engine type  | Low-speed<br>high-<br>pressure<br>dual-fuel<br>engines<br>(LS-HPDF) | Low-speed<br>low-pressure<br>dual-fuel<br>engines (LS-<br>LPDF) | Low-speed<br>diesel cycle<br>engines (LSD) | Low-speed<br>diesel cycle<br>engines (LSD) | Low-speed<br>diesel cycle<br>engines<br>(LSD) |  |
| Engine Maker's<br>Model   | 6G50ME-<br>GI   | 6X-52DF   | 6G50ME-LGIP                                | 6G50ME                                     | 6G50ME-<br>LGIM                               |  |
| Average main<br>engine load at sea<br>and operation<br>days   | 72.5% / 250 I 'ays  |   |  |  |   |  |
| Engine thermal<br>efficiency (%)<br>at 72.5 % of MCR  | 55.30   | 50.20   | 5.5.7                                      | 53.9                                       | 53.9  |  |
| SFC (g/kWh) at<br>72.5 % of MCR<br>(Main /pilot fuel)   | 126.85/3.9  | 141.75/1.95   | (3),9/7.85                                 | 147.24/ -                                  | 307.5/13.1                                    |  |
| Emission factor (GHG t/t fue.) (IMO, 2018a; IMO, 2020; Pavlenko et al., 2020)   |   |   |  |  |   |  |
| CO <sub>2</sub>   | 2.75  | 2.75  | 3.015                                      | 3.144                                      | 1.375   |  |
| CH <sub>4</sub>   | 0.001449  | 0.)1/(83  | 0.00006                                    | 0.00006                                    | 0.000006                                      |  |
| N <sub>2</sub> O  | 0.000217  | <u>u.00u137</u>   | 0.00016                                    | 0.00016                                    | 0.000016                                      |  |
| Lower heating value (LHV) of LNG. <sup>19.2</sup> MJ/kg, Lower heating value of LPG: 46.0 MJ/kg, Lower  |   |   |  |  |   |  |
| heating value of HFO: 40.2 MJ/kg, <sup>1</sup> ower heating value of methanol: 19.9 MJ/kg, Emission per unit of   |   |   |  |  |   |  |
| fuel energy $(g/MJ \text{ fuel}) = 1/3.6 \times g^{1/2}$ Wh engine output x efficiency engine, Emissions per mass of fuel $(g/kg \text{ fuel}) = g/MI \text{ fuel} \times I HV$ . |   |   |  |  |   |  |
| $(g/kg \text{ fuel}) = g/MJ \text{ fuel} \times L^{\mu\nu} \iota_{\mu\nu} (MJ/kg)$  |   |   |  |  |   |  |

Table 7 The data for estimating annual TtW GHG emissions for main engines using LNG, LPG, HFO and methanol

#### 3.2.2.4 Auxiliary engines selected

In this study, auxiliary engines using LPG, HFO and methanol are medium-speed diesel cycle engines (MSD) while the LNG-fueled auxiliary engine employs an Otto combustion process. Table 8 presents specific fuel consumption and emission factors for the auxiliary engine. The emission factors obtained from the 4th IMO GHG study. Table 8 presents the auxiliary engine and boiler power outputs depending on operational mode. Considering ship type, size and operational mode for selected case ship, the power output of the auxiliary engine is assumed to be 260 kW during sea operation and 680 kW during maneuvering (IMO, 2020).

| Fuel  | HFO           | LNG                | LPG           | Methanol             |  |  |
|---|---------------|--------------------|---------------|----------------------|--|--|
| Auxiliary                                   | Medium-speed  | Medium-speed low-  | Medium-speed  | Medium-speed         |  |  |
| Engine type                                 | diesel cycle  | pressure dual-fuel | diesel cycle  | diesel cycle engines |  |  |
|   | engines (MSD) | engines (MS-LPDF)  | engines (MSD) | (MSD)                |  |  |
| SFOC  | 195           | 152                | 160           | 370                  |  |  |
| (g/kWh)                                     |               |                    |               |                      |  |  |
| Emission factors (GHG t/t fuel) (IMO, 2020) |               |                    |               |                      |  |  |
| $CO_2$                                      | 3.144         | 2.75               | 3.015         | 1.375                |  |  |
| CH <sub>4</sub>                             | 0.00006       | 0.036              | 0.000006      | 0.000006             |  |  |
| N <sub>2</sub> O                            | 0.00025       | 0.000131           | 0.000025      | 0.000025             |  |  |

Table 8 The emission factors for estimating annual TtW GHG emissions from auxiliary engine

#### 3.2.2.5 Fuel consumption estimation model

Emissions from the main engine depend on the selected rate a pover, load factor, fuel type, engine type and year the engine was built. The main engine power and load factor will change over time as a consequence of the vessel's operating and activity detail, such as speed, loading conditions, weather, etc (Smith et al., 2015). The power outputs required the ship propulsion are considered the results of operating speed trends. This can be represented as "loads" corresponding to the proportion of the overall installed maximum power output called MCR. In this study, based on ship's Abstract LOG data, average main engine load at sea wa: closen as 72.5% of MCR and its loads are converted to fuel consumption using the specific fiel consumption (SFC) for main fuel and pilot fuels. The dual fuel engines always operate on LNG, 1 PG and methanol as their primary fuels while the amount of pilot fuels injected changes depending on engine loads.

The annual fuel consum<sup>\*</sup> on of the main engine and auxiliary engine is calculated through the following equation.

- $FC_{primary} = AML \times SFOC \times AD$  (7)
- $FC_{pilot} = AML \times SPOC \times AD$  (8)
- $FC_{aux} = AML \times SFOC \times AD$  (9)

where;  $FC_{primary}$ : Primary fuel consumption (t) for main engine,  $FC_{pilot}$ : Pilot fuel consumption(t) for main engine,  $FC_{aux}$ : Fuel consumption (t) for auxiliary engine, AML: Kilowatts (kW) at average main

and auxiliary engine load, SFOC: Specific fuel consumption (g/kWh) for primary fuel at AML, SPOC: Specific fuel consumption (g/kWh) for pilot fuel at AML, AD: Average days at sea per year

To compare the performance of the GHG emission equivalently from engines using selected fuels, specific fuel oil consumption data (g/kWh) were collected from the engine manufacturer's engine selection software (MAN CEAS Engine Calculations and WinGD General Technical Data). The changes of SFOC and SPOC (g/kWh) were estimated as a function of engine load over the whole range. The study utilized regression analysis with R-squared value. (R<sup>2</sup>) to evaluate the fit of the model developed from the data provided by the engine manufacture. Based on this analysis, it was determined that a cubic function was the most suitable manematical model to describe the performance of methanol engines, whereas quadratic functions were found to be more appropriate for other types of engines. Table 7 also contains the specific fuel consumption values for each engine type at an average main engine load of 72.5% of t<sup>1</sup> and time continuous rating (MCR). It is noteworthy to mention that these formulas enable us to product fuel consumption based on the engine loads that are being used.

For LNG Fueled engine (LS-HPD.)

- SFOC (y) = $0.0045x^2 0.59x + 146.51$  (R<sup>2</sup> = 0.9566) (10)
- SPOC (y) =  $0.001 > -^2 0.3024x + 15.367$  (R<sup>2</sup> = 0.9414) (11)

For LNG Fueled engine (LS-LPDF),

- SFOC (y) =  $0.0033x^2 0.5148x + 161.61$  (R<sup>2</sup> = 0.9965) (12)
- SPOC (y) =  $0.0004x^2 0.0807x + 5.5128$  (R<sup>2</sup> = 0.9908) (13)

For LPG Fueled engine,

- SFOC (y) =  $0.0022x^2 0.2491x + 145.19$  (R<sup>2</sup> = 0.8209) (14)
- SPOC (y) =  $0.0037x^2 0.5995x + 30.602$  (R<sup>2</sup> = 0.9383) (15)

For HFO fueled engine,

• SFOC (y) = 
$$0.0065x^2 - 0.9238x + 198.72$$
 (R<sup>2</sup> =  $0.9674$ ) (16)

For Methanol fueled engine,

• SFOC (y) =
$$0.0002x^3 - 0.033x^2 + 1.8279x + 273.18(R^2 = 0.9933)$$
 (17)

• SPOC (y) = 
$$-0.0001x^3 + 0.0282x^2 - 2.0363x + 63.508(R^2 = 0.984)$$
 (18)

### 3.2.3 Impact analysis on maritime transportation

In addition to the aforementioned data collection and generation r ethels, this analysis was proposed to investigate the impact of geographical differences, as a key parametric variable, specifically transportation distances that reflect regional characteristics on the lifecycle emission levels. In this regard, the analysis was conducted in a way to compare three distinct functional units within the context of LNG transportation, depending on the import region. It utilized the actual data of the actual cargo (LNG) carried, voyage distances, ar 1 firel consumption of LNG carriers operating between South Korea and different importing countries. Key findings are presented in Section 4.4.

• 
$$E_c = \frac{\sum_{i}^{n \text{ fuel }} \sum_{j}^{m \text{ engine }} M_{i,j} \times F_{G_{i,j}}}{M_{cargo}}$$
(19)

• 
$$E_d = \frac{\sum_{i}^{n \text{ fuel } \sum_{j}^{m \text{ engine}} \cdots \times \alpha F_{GHG}}{L}$$
 (20)

• 
$$E_{tw} = \frac{\sum_{i}^{n \text{ fuel } \sum_{j}^{m \text{ engir } e} M_{i,j} \times EF_{GHG}}{M_{cargo} \times D}$$
 (21)

where;

 $E_c$ : emissions per LNG ton carried (tCO<sub>2</sub> eq./ton),  $E_d$ : emissions per voyage distance(tCO<sub>2</sub> eq./mile),  $E_{tw}$ : emissions per transport work(voyage distance multiplied by the ton of LNG carried) (tCO<sub>2</sub> eq./mile·ton), D : Total voyage distances, For further details and calculations, please refer to equations (3) and (4).

# **4** Results

Section 4 presents the findings of the study, with a specific focus on the analysis of the WtT GHG emissions of selected fuels in Section 4.1, the TtW GHG emissions from the main and auxiliary engines in Section 4.2, and the WtW GHG emissions, including a comparative assessment, in Section 4.3. The graphical representation of the results is presented in the main text, while the numerical values are provided in the supplementary material for reference.

### 4.1 Comparison of WtT GHG emissions from fuels

Figure 6 indicates the results of WtT GHG emission for H O nom seven imported countries. The identical trends were found that emissions for crude oil p. du tion and refining process account for a large proportion of the total WtT emissions wher as the transport emission by shipping and the emissions originated from distribution are that ely estimated to be smaller. During crude oil production, Iraq's case has about three times nore emissions when compared to Saudi Arabia. In the meantime, with regard to transport emissions between seven countries, the minimum emission is 0.91 g CO<sub>2</sub> eq./MJ for Saudi Arabia and the maximum is 2.5g CO<sub>2</sub> eq./MJ for USA while average value is 1.3g CO<sub>2</sub> eq./MJ. A key finding drough those figures is that it does not have significant impact on total emissions for HFO produced in South Korea. Unlike LNG and LPG, the reason for the relatively low transportation emissions by oil tanker is due to the very high transportation efficiency with large cargo capacity as this study assumed that all crude oils are transported by 300K VLCC.

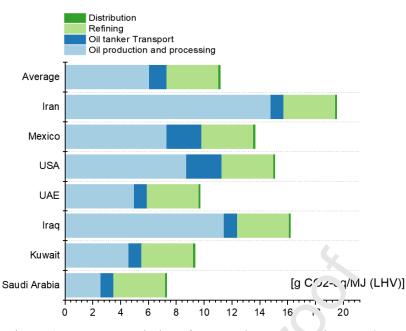


Figure 6 WtT GHG emissions for HFO depending on imported countries

Figure 7 shows WtT GHG emissions for LNG depending on imported countries. Although the emissions stemming from liquefaction process cont it at the most to the total WtT GHG emissions, their difference is not considerable. It can be not considerable. It can be not country with the efficiency of liquefaction is analogous across the technologie, applied in each country with the range from 90.4% to 92.9%, as indicated in Table 9 (Choi an Song, 2014).

Table 9 Percentage of Share and Etheriency of Various Liquefaction Technologies (Choi and Song, 2014)

|                       | Technology share (%) | Efficiency (%) |  |  |
|-----------------------|----------------------|----------------|--|--|
| C <sup>3</sup> MR     | 67.72                | 92.9           |  |  |
| C <sup>-</sup> .scade | 14.83                | 91.2           |  |  |
| SMR                   | 1.72                 | 91.6           |  |  |
| DMR                   | 3.93                 | 92.7           |  |  |
| AP-X                  | 11.79                | 90.4-92.9      |  |  |

On the other hand, transportation emissions by LNG carriers have a wide range of emissions from 2.26 g  $CO_2$  eq./MJ to 5.97 g  $CO_2$  eq./MJ. Notably, Indonesia and Malaysia demonstrate comparatively lower emissions, with values of 2.26 g CO2 eq./MJ and 2.43 g CO2 eq./MJ, respectively. In contrast, Australia's transportation emissions amount to 4.11 g CO2 eq./MJ. It is intriguing to observe that, despite employing the same conventional steam turbine system and cargo capacity for transporting

LNG, Australia's emissions differ from those of the other two countries. The further analysis on this difference was performed in section 4.4.

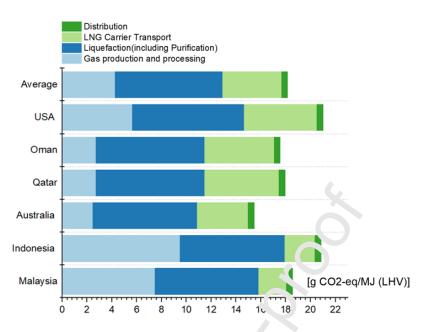


Figure 7 WtT GHG emissions for LNC Cenending on imported countries

As depicted in **Error! Reference source ne. hund**, the life cycle GHG emissions for natural gasbased LPG are slightly greater than one from LPG made from crude oil. This result suggests that the feedstock nature of marine fuels is on *correctional elements* to determine the level of GWP impact although they are fossil-based fuels.

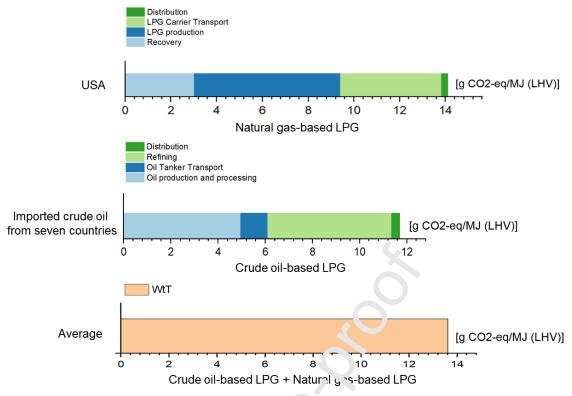


Figure 8 WtT GHG emissions for crude oi'-b'.s ed LPG and natural gas-based LPG

For methanol, the WtT emissions are determined to be approximately 20.72 g  $CO_2$  eq./MJ, as shown in Figure 9. These values are derived from the utilization of the GREET model developed at Argonne National Laboratory, combined with trais A station emissions data for a 55K methanol carrier.

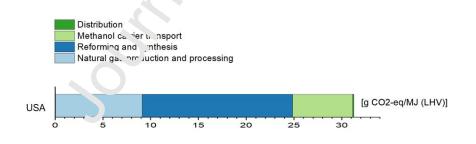


Figure 9 WtT GHG emissions for methanol

Among all fuels, the downstream GHG emission from HFO is the least whereas methanol is considered the most. This disparity can be attributed to the fact that methanol production is derived from fossil-based natural gas, and the energy required for the conversion of natural gas into methanol results in additional GHG emissions. However, it is important to note that the WtT emissions of HFO show the highest variability among all fuels as shown in Figure 6 and Table S6 in the supplementary

material.

#### 4.2 Comparison of TtW GHG emissions

Figure 10 compares estimates of the annual TtW emissions from ships using HFO, LNG, LPG and methanol fuels when using their main and auxiliary engines. The HFO-fueled ship was the greatest GHG emitter while using LNG with LS-HPDF was the lowest. Emissions from ships using these engines are approximately 17,689t, 12,463t (LS-HPDF),15,806t (LS-LPDF), 13,651t and 15,348t respectively. Overall, compared to emissions from main engine on room auxiliary engine have ranges of 5.2 to 7.3% of total TtW GHG emissions.

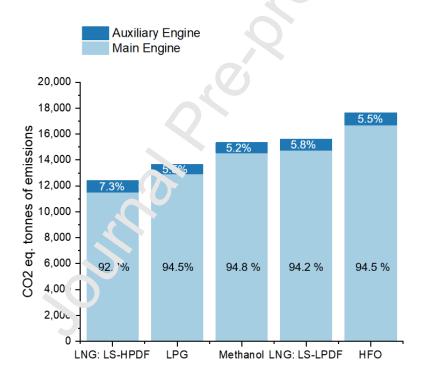


Figure 10 Annual TtW GHG emissions for HFO, LNG, LPG and methanol fuel (at average main engine load: 72.5% of MCR and operation days at sea: 250 days)

Figure 11 also illustrates average GHG intensity of the energy used on board ships for a given year when all emissions from ships are considered only from main and auxiliary engines. The trend of GHG intensity between fuels is analogous to that of annual GHG emissions indicated in Figure 10. As

a policy measure, the IMO is considering the establishment of a GHG Fuel Standard (GFS), which would require ships to use fuels or other energy sources with a WtW GHG intensity at or below a certain limit value over a compliance period (IMO, 2022c). This observation of the trends suggests that reducing the GHG intensity of the energy used on board ships through this measure will have a positive impact on achieving the annual GHG emission reduction levels in line with the ambitions of the IMO GHG Strategy. In addition, similar trends were observed across results that the emissions emitted from auxiliary engines using the same fuel as the main engines have no significant effect on the average total GHG intensity.

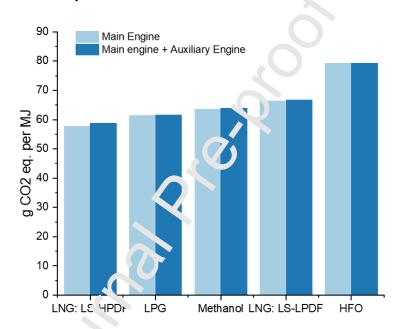


Figure 11 Average GHG intensity of the energy used on board ships

#### 4.2.1 Comparison of GHG emissions between main fuel and pilot fuel

Table 10 indicates the result of specific fuel consumption (g/kWh) for main fuel and pilot fuel depending on engine load, using developed quadratic functions per each fuel engine. It can be seen that the amount of pilot fuel supplied for ensuring a stable ignition and combustion is non-identical depending on the combustion characteristics of applied each fuel and combustion cycles (e.g., otto or diesel) of the engines. Especially, other fueled engines except methanol engine have smaller pilot fuel consumptions over the higher load operation but methanol engine with engine loads from 40% to

60%. It is also inferred that depending on engine type and its fuel, there are technical challenges in designing a robust injection system that is small and fast enough to inject a small amount of pilot oil into the engine while it is enabled to be efficient high load operation.

| Main<br>Engine | LNG Fueled<br>engine (LS-HPDF) |      | LNG Fueled engine<br>(LS-LPDF) |      | LPG Fueled engine |          | Methanol engine |      |
|----------------|--------------------------------|------|--------------------------------|------|-------------------|----------|-----------------|------|
| Load<br>(%)    | SFOC                           | SPOC | SFOC                           | SPOC | SFOC              | SPOC     | SFOC            | SPOC |
| 10             | 141.4                          | 12.5 | 156.8                          | 4.7  | 142.9             | 25.0     | 288.4           | 45.9 |
| 20             | 137.1                          | 10.1 | 152.6                          | 4.1  | 141.1             | 20.1     | 298.1           | 33.3 |
| 30             | 133.8                          | 8.0  | 149.1                          | 3.5  | 139.7             | . 5.9    | 303.7           | 25.1 |
| 40             | 131.4                          | 6.3  | 146.3                          | 2.9  | 138.7             | 17.5     | 306.3           | 20.8 |
| 50             | 129.8                          | 5.0  | 144.1                          | 2.5  | 138.2             | <u> </u> | 307.1           | 19.7 |
| 60             | 129.2                          | 4.1  | 142.6                          | 2.1  | 138.2             | 8.0      | 307.3           | 21.3 |
| 70             | 129.4                          | 3.5  | 141.7                          | 1.8  | 13:5              | 6.8      | 308.0           | 24.8 |
| 80             | 130.6                          | 3.3  | 141.5                          | 1.6  | 135.3             | 6.3      | 310.6           | 29.9 |
| 90             | 132.7                          | 3.5  | 142.0                          | 1.5  | 16                | 6.6      | 316.2           | 35.8 |
| 100            | 135.6                          | 4.1  | 143.1                          | 1.4  | 142.3             | 7.7      | 326.0           | 41.9 |

Table 10 Specific fuel consumption(g/kWh) for main fuel and pilot fuel depending on engine load

Based on the data presented in Table 10, the cal ulation of their GHG emissions was performed. The results are illustrated in Figure 12, which displays the contribution of main and pilot fuel to the total GHG emissions over the range of engine 1 ads. In particular, the GHG emissions due to the use of pilot fuel tend to be high at low engine load while their emissions have a different degree of impact on total GHG at that engine load det engine on the type of fuel or engine used. For instance, the emission emitted from LPG's pilot rue1 occupies about 17% at 10% load, whereas for LS LPDF with otto cycle, it is 3%. It is notewo, by that changing pilot fuels to renewable fuels like bio-fuel and synthetic fuel can also contribute to reducing GHG emissions. The extent of emissions reduction varies depending on the fuel, with potential reductions ranging from approximately 1% for LNG LS-LPDF to up to 9% for methanol at a realistic average main engine load of 70-80%.

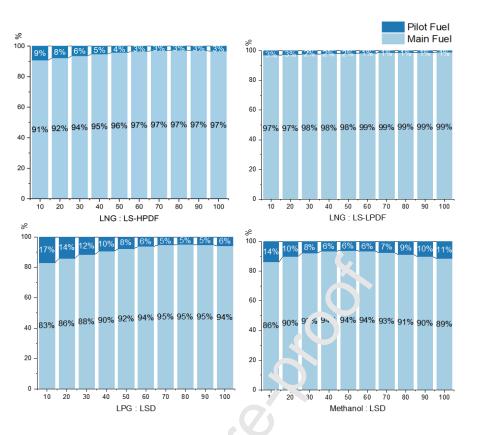
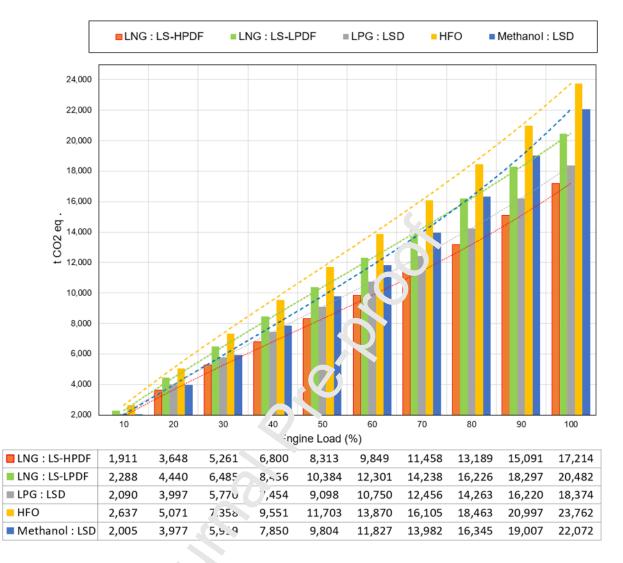


Figure 12 Ratio of GHG emissions emitted by . ain fuel and pilot fuel depending on engine load



4.2.2 Comparison of GHG emissions from main engines

Figure 13 GHG emissions from main engine using HFO, LNG, methanol and LPG fuel

Figure 13 presents the Gr.G emissions and their trends on main engines across four selected fuels, assuming they operate for the same period of 250 days at each engine load. At all engine loads, the main engine using HFO tends to emit the highest amount of GHGs, with the gap between HFO and other fuels in GHG emissions widening as engine load increases. It is also important to note that the emission characteristics of each fuel vary depending on the engine load. Specifically, at engine loads exceeding 80%, methanol tends to produce steeper increases in GHG emissions compared to LNG with LS-LPDF. For example, at an engine load of 70%, methanol and LS-LPDF emit 13981.95 and 14237.66 tons of greenhouse gases, respectively. At an engine load of 80%, the emissions for methanol and LS-LPDF are 16344.97 and 16225.77 tons, respectively. These findings suggest that the

load-dependent cubic function of specific fuel oil consumption for methanol is responsible for these differences, which distinguishes it from other fuels that exhibit quadratic functions Please see Table S26 in the supplement for more details.

#### 4.3 Comparison of WtW GHG emissions

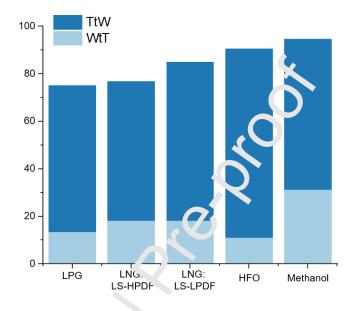


Figure 14 WtW GHG er as is ons for HFO, LNG, methanol and LPG fuel

Figure 14 shows the average WtW GHG emission that combines the results of Section 3.1 and 3.2. From the perspective of WtW GAG emission, LPG is the lowest while methanol counterpart is the largest. The results also clearly show that the emissions of selected fossil-based fuels for WtT phase have less influence on the overall WtW GHG emissions than TtW emissions, while their relative variances are higher. In particular, WtT contributions to the total ranged from 12.3% with 11.15 g/MJ (HFO) to 33.06% with 31.26 g/MJ (Methanol). This also inferred that GHG emissions are mainly derived from combustion processes from ship's engine using fossil fuel with significant carbon contents.

As a result, from a life cycle perspective, it is clear that LPG and LNG are fuels that can reduce GHG emissions compared to conventional fossil fuels (HFO) and methanol.

#### 4.4 Comparison on impact of maritime transportation

To investigate the impact of lifecycle emissions associated with the regions of import, the three functional units of LNG fuel were examined as mentioned in Section 3.2.3.

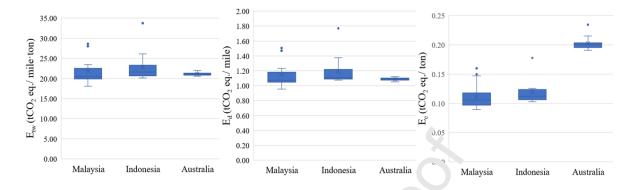


Figure 15 Comparison of GHG Emissions in LNG transportation to Korea via 125K LNGCs from Malaysia, Indonesia, and Australia

As indicated in Table 5, the LNG import to Ko er from Malaysia, Indonesia, and Australia is generally undertaken by 125K LNGCs with steen numbers. Figure 15 depicts the comparative analysis of average GHG emissions for these three units namely  $E_c$ ,  $E_d$  and  $E_{tw}$ .  $E_d$  and  $E_{tw}$  exhibit a similar trend across the three countries. In general there seem little differences on emission levels across the importing countries. However, when  $a_t$  plying  $E_c$  (emissions per LNG ton or MJ carried), commonly applied in LCA, Australia has a bigher tendency on emission levels compared to Malaysia and Indonesia. This suggests that the analyzed 125K LNG carriers operating between South Korea and Indonesia, or Malaysia in we shorter voyage distances compared to the operational range on the Australia route, as illustrated in Figure 16.

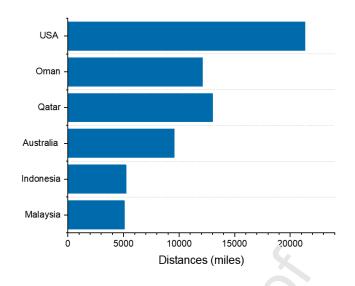


Figure 16 Variations in voyage distances based on regions of LNG import

On the other hand, ships equipped with greater cargo capacity and more efficient propulsion systems typically emit lower amounts of GHG per unit of transport work( $E_{tw}$ ). These ships also demonstrate greater energy efficiency in their operations (IMO, 2020). For example, the utilization of steam turbine propulsion is associated with several disadrantages, such as low efficiency and a relatively high level of GHG emissions (Fernández et al., 2017).

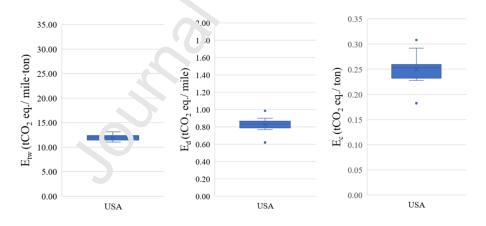


Figure 17 Comparison of GHG Emissions in LNG transportation to Korea via 174K LNGCs from USA

A straightforward comparison of Etw between 125K LNG carriers equipped with a steam turbine and 174K LNG carriers utilizing LS-HPDP for LNG transportation to Korea, as illustrated in Figure 15 and Figure 17, unequivocally demonstrates the higher efficiency of the latter. However, it should be noted that this study does not investigate the specific influence of cargo capacity and propulsion

systems on GHG emissions, and thus, their individual impacts remain unknown. This aspect is further discussed in Section 5.3 Limitations and Future Research Directions. Nevertheless, it is worth noting that the Ec of functional unit applied in the LCA, namely 'emissions per LNG ton or MJ carried' shows the opposite result: importing LNG from the USA via the 174K LNGC, which not only has high propulsion efficiency but also a large cargo capacity, results in higher GHG emission levels in terms of Ec. This increase in emissions is likely due to the higher voyage miles to South Korea, as shown in Figure 16. Hence, the distance from the energy source can have a relatively significant impact on GHG emissions per LNG ton or MJ carried.

# **5** Discussion

In this section, the contributions and limitations of the study, and the direction of future research are presented.

#### 5.1 Contribution to enhancing the ur .<sup>1</sup>erstanding on current GHG emission level

The decarbonization of the ship in industry plays a crucial role in tackling the global issue of climate change. Given the fact the the shipping industry has been focusing only on GHG emissions from ships, the life cycle approach to evaluating GHG emissions from ship fuel is regarded as an effective policy measure for promoting low-carbon and zero-carbon fuels in the industry, taking into consideration not only the fuel use emissions but also the upstream emissions. It is essential for policymakers to understand current WtW emissions values, specifically those associated with the use of fossil fuels, before setting targets for reducing greenhouse gas emissions. In this context, the outcomes from this study would serve as a new standard for future comparison with renewable sustainable fuels to be introduced and contribute to the assessment of policy effectiveness over time. This research has also demonstrated significant variations in upstream emissions across countries, despite the use of identical fossil-based fuels. It also highlights that the selection of environmentally beneficial fuels in terms of ship's emissions alone may not necessarily result in substantial reductions

in emissions levels, as these are also contingent on the production methods, source of energy, and countries of origin of the fuel. For instance, HFO WtT emission values, based on crude oil imported from Iraq and Saudi Arabia, show a notable discrepancy of about 220%, with values of 16.2 g/MJ and 7.3 g/MJ, respectively. This difference highlights the potential implications on the significance of considering the emission level of the produced countries of crude oil as part of the decarbonization efforts.

On the other hand, from the perspective of countries that rely on energy imports, such as South Korea, the growing implementation of alternative fuels in international shipping can also be perceived as a reduction in the GHG emissions arising from the transportation of 'uels into the country. By analyzing the proportion of GHG emission by international maritime transportation in total WtW emission of the available fuels, this study enables us to anticipate the excent to which the decarbonization of international shipping can enhance or contribute to the WtW copect of marine fuel produced in South Korea.

# 5.2 Contribution to suggesting the fut re national and regional LCA regulatory framework for marine fuels

Given that the uptake of sustainable alternative fuels is paramount for the decarbonization of the shipping industry, this study income driven by the strong needs of a government body, the Ministry of Oceans and Fisheries, which endeavors to determine the most suitable marine fuels. To satisfy these needs and answer the research questions, the study investigated the fuel pathways of currently available fuels and their relevant engine technologies, and analyzed their environmental performance to make the best choice.

Energy-importing countries like South Korea have lacked available and reliable life cycle emission data on marine fuels, resulting in a reliance on simple databases (e.g. GaBi and GREET) with varying situations and assumptions. Therefore, developing a customized database and proper emission model was critical work that could serve as a basis for formulating of firmly established policies for reducing GHG emissions in the shipping sector. Evidently, the case study has demonstrated the efficacy of the

proposed emission model through a thorough examination of reliable data and scenarios that are more suitable for its country like South Korea. The functionality for impact analysis in Section 4.4 has demonstrated excellent capabilities for understanding the influence of geographic differences (e.g., voyage distance) on overall maritime GHG emissions, although their impact on the WtW GHG emissions is relatively small.

In fact, there has been no research that proposed default value on marine fuel for the specific countries and international shipping sectors among LCA studies so far. In addition, as discussing the matters on the uncertainty and variations stemming from differences in geographical locations across nations and their supply routes and methods, the research findings are believ d to be valuable guidance for the government who are subjected to setting the proper baseline for regulating the proper limit of the lifecycle GHG emission considering various fuels options. In establishing baseline or default values, the use of conservative values rather than average perfor nance enables operators to provide certified actual values, leading to improved knowledge and 1, do ded uncertainty. For example, Figure 18 is a Sankey diagram showing variable GHG per orr ance of fuels applied in this study are produced. The baseline (100%) was based on the largest mitter of HFO produced from crude oil imported from Iran. The FuelEU Maritime set the target on GFG intensity of energy and fuel used by 2% in 2025 and increase it to 80% by 2050, with incrim targets of 14.5% in 2035 and 31% in 2040. If these targets were applied to marine fuels imported to South Korea, only the LPG fuel pathway and the LNG pathway with LS-HPDF are c pable of achieving their 2035 targets, regardless of the production region investigated. Moreover, the targets could also be met for LS-HPDF with the LNG produced from specific regions, as depicted in Figure 18.

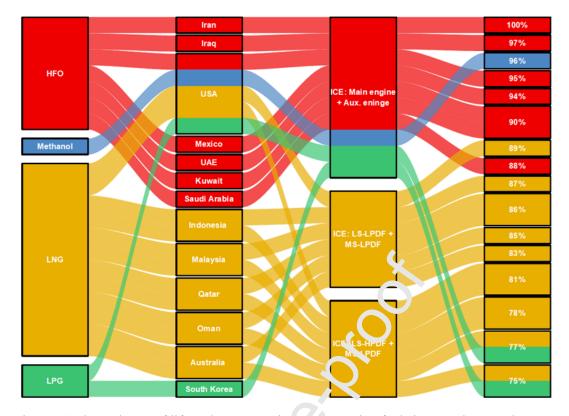


Figure 18 The pathway of lifecycle GHG emiss on for marine fuels imported to South Korea Therefore, the research results, with the San tey diagram, could be an excellent indicator for estimation the current levels of the emissions from use of fossil-based fuels and their potential reductions based on applied technologies with maturity. Additionally, this study provides fuel producers and engine makers with constructive recommendations for improving their emissions and achieving emission reduction targets. Although this paper is focused on South Korean case, it was suggested as an urgent task to tretch case studies on other countries to various regulatory bodies that require appropriate setting to regets for GHG reduction.

From an IMO regulatory standpoint, the IMO framework employs global default values for fossil fuel pathways in the LCA methodology as the initial option, with the potential for utilizing regional factors upon stabilization. In this context, it also contributes to establishing regional values if global values are deemed limiting.

Overall, this study affirms and argues, through a case study, that WtT and TtW emission values should be developed based on a unified methodology. This study presents a novel approach and direction for environmental evaluation in the life cycle aspect, which is being considered as an alternative to the current regulation that is solely centered on the user's point of view. Moreover, by

intuitively illustrating the significant environmental impacts of various import routes in the process of importing, refining, and utilizing fuel from a holistic perspective, this study demonstrates the urgent need for the introduction of a life cycle evaluation system for ship fuels. The outcomes of this study are significant and impactful, and they can be implemented in policies and regulations as an enhanced format of LCA application to the maritime sector. Meanwhile, it should be mentioned that the proposed LCA approach is useful for facilitating general observations on LCA impacts across various countries and regions. In addition to South Korean case, a series of other case studies should be followed to achieve this goal.

Last but not least, as a key contribution, research approaches and indings was documented and presented before IMO policy makers during IMO Marine Environment Protection Committee (MEPC) - 80th session. As a next step, it is paramount to take a further consideration on how the proposed method/model could be incorporated into IMO LCA griderines which are currently under further development.

#### 5.3 Limitation and future study directions

As depicted in Figure 18 The pathway of Infecycle GHG emission for marine fuels imported to South Korea , the utilization of fossil-anel-based marine fuels, which is the focal point of this study, indicates that there are restrictions to the transition to zero-emission shipping. It should be noted that selected fossil-based fuels exhibit significantly lower WtT emissions compared to TtW emissions. In other words, TtW emissions, which constitute the majority of WtW emissions, can only be reduced to a limited extent with fossil fuels (e.g., LPG has 25% lower emissions compared to HFO). To achieve significant GHG emission reductions by 2050, the use of alternative fuels is essential. Nonetheless, this paper provides knowledge not only for recognizing the current emission levels but also for identifying potential emission reductions in the near future, for example by opting for or converting to LNG and LPG or by constructing new ships with their respective propulsion systems. Additionally, it's worth noting that reducing GHG emissions does not solely rely on identifying emissions from

fossil fuels and setting reduction targets. It can also be accomplished by implementing quotas for alternative fuels (e.g., establishing a quota for synthetic aviation fuel in Europe).

The introduction of renewable fuels into the shipping sector has the capability to significantly mitigate life cycle GHG emissions (Deng et al., 2021). In particular, a shift from conventional fossil-based feedstock to renewable sources, such as biomass and renewable electricity, is imperative to achieve net zero emissions from the shipping sector by 2050 (Department for Transport (UK), 2021). However, due to the significant variations in the availability of renewable energy across regions and countries, further in-depth investigations are warranted to determine the most optimal renewable and sustainable fuel options from an LCA perspective, considering the unique energy scenarios of each nation's energy policy. Additionally, further research could be conducted to assess the potential impact for incorporating regional factors for renewable fuels into the LCA regulatory framework, based on emission level of fossil-based fuels presented in this tudy.

In addition, this study faced limitations in obtair in emission data, particularly regarding the WtT GHG emissions data for methanol production in the USA. To address this limitation, future research should prioritize the collection of comprehensine data on the WtT GHG emissions of methanol from different production regions, particularly the infiddle East and Trinidad and Tobago. Investigating variations in production methods and associated emissions profiles is crucial to provide a more accurate assessment of the environmental impart of methanol production and transportation to South Korea. Furthermore, further investigation of GH 3 emissions from methanol transport by international shipping based on actual vessel operational data is also necessary.

Another limitation of this study is that, while this study investigated the the general perspective on impact of geographical differences, specifically maritime transportation distances that reflect regional characteristics on the lifecycle emission levels, the specific influence of cargo capacity and propulsion systems on GHG emissions was not identified. Given that emissions associated with shipping activities vary depending on various parameters, including the transport route, ship capacity, total

transported quantity, and engine type, further research is required to quantitatively assess how individual factors affect GHG emissions.

It should be noted that the main purpose of this LCA approach is to assist in the decision-making process regarding the proper selection of fuels and propulsion systems for ships. Even though the WTT value has been extensively studied and is known to have a relatively smaller impact on LCA, compared to TTW impacts, this does not necessarily limit the novelty of this study. This study is intended to be applicable not only to conventional fossil fuels, but also to low-carbon or zero-carbon fuels. In the case of alternative fuels, such as hydrogen made from natural gas, where the weight of the WtT emission is significant (DNV, 2018), the importance of t' e key parameters discussed in this paper would be magnified. In other words, although the interpre ation of this case study may be limited, it would be beneficial to include in the discussion that when applying this LCA model to alternative fuels and their energy sources, where WTT has a decisive impact, the results will be significantly different and the impact will be even pre-tre.

Last but not least, as a key contribution, r search approaches and findings will be presented before IMO policy makers during IMO Marine Environment Protection Committee (MEPC). As a next step, it is paramount to take a further consideration on how the proposed method/model could be incorporated into IMO LCA guide'n, 's which are currently under further development.

# 6 Conclusion

The novelty of this research can be placed on addressing the importance of developing robust default WtW emission values and evaluating the GHG performance of various marine fuels across the world. The South Korean case study has successfully come into the following conclusions:

- Among HFO, LNG, LPG and methanol fuels, TtW GHG emissions from the HFO-fueled ship were the greatest while using LNG with LS-HPDF was the lowest.
- In the case of using dual fuel engines, changing pilot fuels to renewable fuels like bio-fuel and synthetic fuel can contribute to reducing TtW GHG emissions by 3% (LNG: LS-LPDF) to

17% (LPG), depending on the fuel and engine type, as well as engine load.

- The emissions generated by auxiliary engines account for between 5.2% and 7.3% of the total TtW GHG emissions. However, these emissions have minimal impact on the average total GHG intensity of the energy used on board ships, as long as the auxiliary engines and main engines use the same fuel.
- HFO comes to the forefront in terms of WtT GHG emission. Its GHG emission with 11.15 g/MJ would be 18.1%, 38.5%, and 64.3% percent lower than those of the LPG, LNG, and methanol. However, the emissions of WtT phase have leep in Tuence on the overall WtW GHG emissions than TtW emissions, while their relative versance is higher. In particular, WtT contributions to the total ranged from 12.3% with 115 g/MJ (HFO) to 33.1% with 31.26 g/MJ (Methanol).
- The average WtW GHG emissions of LF could LNG, which are fossil-based fuels, are generally lower than those of content and fossil fuel (HFO) and methanol. LPG has the lowest emission levels, followed 'v LNG (LS-HPDF), LNG (LS-LPDF), HFO and methanol.
- The impact of international shipting for energy carriers on WtT GHG emissions would be, to some large extent, influenced by various factors : the propulsion systems, the quantity of energy transported, and the routes and distances of voyages from the origin country of energy imports. For Ling Suel, transportation emissions from LNG carriers can vary, ranging from 12.2% of WtT emissions in the case of import from Malaysia (2.26 g CO<sub>2</sub> eq./MJ) to 33.3% of WtT emissions when it comes from Qatar (5.97 g CO<sub>2</sub> eq./MJ).
- Furthermore, for the estimated WtW GHG emission of LNG fuel, the lowest value is observed for fuel imported from Australia, which emits 74.25 g CO<sub>2</sub> eq./MJ (WtT :15.46, TtW: 58.8 using LS-HPDF). On the other hand, the highest value was found for LNG imported from the USA, with emissions of 87.85 g CO2 eq./MJ (WtT : 21.06, TtW: 66.84 using LS-LPDF).
- WtW GHG emissions from HFO show a greater variation depending on the country of origin

(100-88%), whereas this effect is relatively small for other fuels, including LNG with LS-LPDF (89-83%), methanol (96%), LNG with LS-HPDF (81-75%), and LPG (77-75%). However, it is evident that depending on the fuel, a substantial reduction in WtW emissions can be achieved through TtW emissions due to their greater impact (LNG with LS-LPDF: 79%, methanol: 67%, LNG with LS-HPDF: 76%, LPG: 82%). Consequently, while WtW considerations provide a detailed perspective, it is ultimately the TtW emissions that play a more decisive role in reducing emissions when using fossil-based fuels.

• Overall, this paper highlights the importance of developing a fault values of GHG emission for different countries. Research findings offers an insignt into future LCA guidelines and academic/industrial practices that need to be further incomported with the impact of regional or geographic characteristics, such as fuel production technologies and voyage distance in maritime transportation, in the WtT comported and the choice of propulsion system in the TtW component. These parameters will contribute to developing unique default LCA values for each nation so that the accuracy a. I precision of the holistic assessment will be improved. Finally, decision/policy-making processes including market-based measurement i.e. FuelEU Maritime can be enhanced the improved of the for curbing emissions from the transportation sector.

# **Declaration of Competing Interest**

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Author Contributions Statement

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Chybyung park: Writing - review & editing, Proof reading, Design, Supervision.

Byungchurl Ku: Investigation, proof reading and Resources.

## **Declaration of interests**

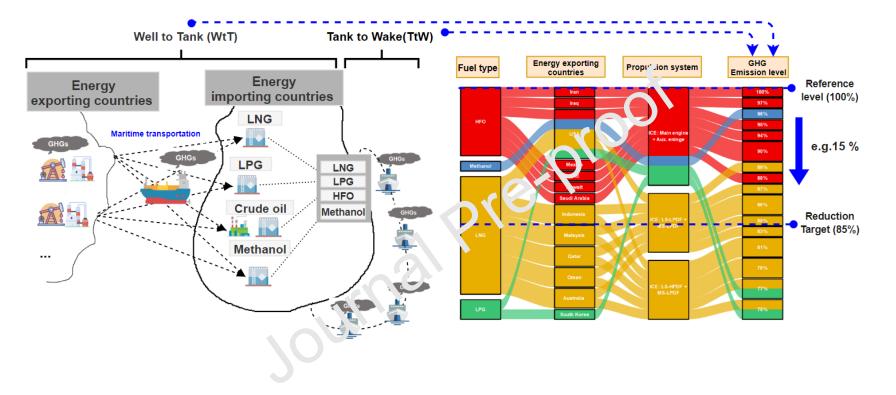
 $\boxtimes$  The authors declare that they have no known competing financial interests or personal

relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:



Graphical abstract



#### Highlights

• Developed framework for baseline of life cycle GHG emissions from marine fuels

• Demystified the current LCA framework for highlighting the potential impact of WtT areas neglected in regulation

- Compared the WtW emissions of selected marine fuels imported to South Korea
- Provided insights for future policies to reduce life cycle GHG emissions

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