

Data-Driven Modelling of Freshwater Ecosystems: A Multiscale Framework Based on Global Geospatial Data

Bruna Almeida^a and Pedro Cabral^b

Information Management School (NOVA IMS), Universidade Nova de Lisboa,
Campus de Campolide, 1070-312 Lisbon, Portugal

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Abstract: Freshwater ecosystems are primarily impacted by climate, land use and land cover changes, and over-abstraction. Satellite Earth observation (SEO) data and technologies are key in environmental modelling and support decisions. These technologies combined with machine learning (ML) are a powerful approach for modelling freshwater ecosystems at a multiscale level. The goal of this study is to present a set of reference data and guidelines that can be used to estimate the water and wetness probability index (WWPI) in different spatial and temporal scales. To find the best model's predictors, sensitivity analyses were carried out in a predictive ML model implemented in a transnational river basin district (Portugal – Spain), the Tagus Basin. Satellite imagery, satellite-derived data, biophysical variables, and landscape characteristics were the explanatory variables evaluated in the sensitivity analyses, and some of them were included in the framework as a reference source of spatial data.

1 INTRODUCTION

Anthropogenic and environmental changes threaten the existence of ecosystems that depend directly or indirectly on the presence of water in a landscape (Mpakairi et al., 2022). Those ecosystems are biodiversity hotspots that require access to water to maintain the communities of plants and animals, the ecological processes they support, and the services they provide (Shi et al., 2014). In dry periods, there is less water in rivers, and over-exploitation depletes the natural water table and affects these vulnerable ecosystems (Sharma et al., 2018).

The status of water resources degradation and scarcity gets worse as human activities directly cause an increase in the local drying initiated by climate (Kløve et al., 2014). Adverse atmospheric conditions impact water resource availability, compounding the challenge of integrated water management, concerning quality, quantity, and ecosystem support (Novo et al., 2018). In face of climate change and ever-more intensive land use, conceptual models, and quantitative assessments of surface and groundwater

interactions with the environment are needed (Yang et al., 2021).

The increasing global demand for water for agriculture, domestic, and industrial needs, requires integrated management of natural resources (Almeida & Cabral, 2021). Therefore, there is a lack of knowledge on the existence of conceptual frameworks and guidelines useful to model freshwater ecosystems (Cui et al., 2021). The process of acquiring geospatial data is the most challenging part of environmental modelling (Meddens et al., 2022). Frameworks to support data-driven modelling must incorporate the knowledge that facilitates spatial and spatiotemporal data acquisition and baselines that guide conceptualization and model implementation (Nti et al., 2022).

When there are few or no guidelines for sources of spatial data and environmental applications, searching and applying exhaustively all the possibilities is not a feasible option because, even with very sophisticated computers, the running time needed would be unacceptable, or could not add any value for the model in terms of quality. Several

^a <https://orcid.org/0000-0002-3349-1470>

^b <https://orcid.org/0000-0001-8622-6008>

studies have been developing methodologies for modelling water ecosystems using remote sensing and machine learning, such as Kundu et al. (2022) detecting water richness change, Mpakairi et al. (2022) assessing spatiotemporal variation of vegetation heterogeneity in groundwater-dependent ecosystems, and Sharma et al. (2018) estimating impacts on the water resources from crop irrigation. In synthesis, we observe that some of the previously proposed approaches solve the problem very well, but none of them was presenting possible sources of global spatial data and performed a sensibility analysis of their use and applications in modelling water ecosystems.

The Copernicus Land Monitoring Services (CLMS) provides the basis for integrated analysis of the main drivers of land use change to inform about Europe's natural resources and their status (Copernicus Programme, 2023). It offers information services based on satellite Earth observation freely and openly accessible to its users, only requiring registration and referencing. In the portfolio of the High-Resolution Layers (HRL) under the *Experts and National Products archive*, exist five thematic layers on land cover characteristics at 10m spatial resolution, covering 39 European countries. The HRL Water and Wetness (WAW) 2018 datasets are based on imagery covering the period 2012-2018. They were created based on satellite imagery, including ESA's Sentinel-1 and Sentinel-2 satellites and spectral indices such as the Normalised Difference Water Index (NDWI) and its modified version mNDWI, as well as the Normalised Difference Vegetation Index (NDVI).

The goal of this study is to present a set of reference data and application guidelines that support reproducing the WWPI with accuracy. For that, we developed machine learning (ML) models and performed sensitivity analyses to find the best predictors. The ML models were implemented in a transnational river basin district (Portugal – Spain), the Tagus Basin. Satellite imagery, satellite-derived data, biophysical variables, and landscape characteristics were the explanatory variables evaluated in the sensitivity analyses.

The framework comprises a set of global open-access data, that can be used to model freshwater ecosystems at multiscale (spatial and spatiotemporal). The outcomes of this research would increase our understanding of the use and replicability of the Copernicus data, and knowledge in implementing data-driven models based on SEO data in different years and locations. The developed framework comprises a set of baselines for predicting the spatial

distribution of water and wetness status and conditions allowing the assessment of the impacts of drivers of changes on freshwater ecosystems.

2 MATERIALS & METHODS

2.1 Study Area

The Tagus Basin in Portugal is the most important source of water in the country due to its productivity, and quality of water (Ribeiro, 1998). The river basin resources are responsible for economic and demographic expansion through the years (Mendonça, 1990).

The extensive network drainage promotes easy access to water and supports the development of intensive agriculture (Almeida, 2020), such as annually harvested plants including flooded crops such as rice fields and other inundated croplands, and permanent crops such as vineyards, fruit trees and olive groves (Novo et al., 2018). Forest and seminatural areas are represented by mixed forest and transitional woodland/shrub (Ramos et al., 2017).

Due to its strategic location in the surroundings of Lisbon, the most populated region of the country, there is a growing need for changes in land use to develop artificial surfaces such as urban fabric, road and rail networks, airports, dump sites and industrial areas (Mendes et al., 2015), which is compromising water resources and dependent ecosystems in the basin.

2.2 Data Source

2.2.1 Dependent Variable

The necessary data to implement a learning machine process is commonly divided into training, validation, and test datasets (Domingos, 2012). The training data are the predictors, the validation dataset is used to control the learning process and the test dataset is employed to assess the learner's performance. To predict continuous values such as the Water and Wetness Probability Index (WWPI), regression tasks were carried out. The target in the regression models is the response variable, also known as dependent variable.

The WWPI is a raster displaying the occurrence of water and wet areas as an index on a scale between 0 (only dry observations) to 100 (only water observations) (Copernicus Programme, 2023). It indicates the degree of wetness in a physical sense, independently of the actual vegetation cover. The

landscape elements included in the datasets are permanent and temporary open water bodies, temporarily inundated areas, wet agricultural fields, and transitional coastal water bodies (European Environment Agency, 2023).

2.2.2 Explanatory Variables

The main goal of this study is to find the best predictors to enable an accurate reproducibility of the WWPI dataset. A wide variety of data such as satellite imagery, satellite-derived data, biophysical variables, and landscape characteristics, were evaluated through sensitivity analysis.

Single bands from Sentinel-2 images and satellite-derived indices were the first predictors evaluated in the sensitivity analysis. This satellite has a high-resolution multispectral sensor acquiring and recording information in 13 bands: visible (VIS with 4 bands), near-infrared (NIR with 6 bands) and shortwave infrared (SWIR with 3 bands). This instrument can detect small differences in spectral signatures, as they have contiguous bands with small bandwidths (<20 nm), and consequently acquire more accurate information (Hunter et al., 2020). The Sentinel-2 mission is providing data from 2015 until 2025 with a temporal resolution of 5 days, and spatial resolution of 10, 20 and 60m.

The combination of algebraic operations in pre-established bands leads to the definition of indices that allow highlighting certain information (i.e., water, vegetation, soil, minerals, etc.). Spectral-derived indices result from combinations of two or more spectral bands designed for the enhancement of specific objects (Kuenzer et al., 2014). NDVI, NDWI and Normalized Difference Moisture Index (NDMI) were calculated and included as predictors.

Climate variables, such as precipitation and evapotranspiration combined with topography are important drivers in water modelling (Ringersma et al., 2003). The annual precipitation was obtained from the WorldClim database (Fick & Hijmans, 2017) and the Global Potential Evapotranspiration (Global-PET) and the Global Aridity Index (Global-Aridity) from the Consortium of Spatial Information, Global-Aridity and Global-PET Database (Trabucco & Zomer, 2019). Global data precipitation was obtained from spatial interpolation of nearly 60 000 weather stations providing monthly climate data. The datasets have a spatial resolution of approximately 1 km². Reference evapotranspiration and global aridity were derived from the WorldClim precipitation model and have the same spatial resolution and temporal scale.

Topographic data are globally available from NASA through the Shuttle Radar Topography Mission (SRTM). The dataset was released in the year 2000 with approximately 30m pixel resolution.

Table 1 lists the datasets used to estimate WWPI, specifying the source of data and the pixel resolution.

Table 1: Description of the datasets, its source and pixel resolution.

Data	Source	Pixel resolution
Satellite Data	Copernicus Open Access Hub (https://scihub.copernicus.eu/)	10m
Digital Elevation Model (DEM)	NASA Shuttle Radar Topography Mission (https://www.earthdata.nasa.gov/)	30m
Average Annual Precipitation (mm)	WorldClim (https://worldclim.org/)	1km
Global Potential Evapotranspiration (Global-PET) (mm)	Consortium of Spatial Information-CSI (http://www.cgiasr-csi.org)	1km
Global Aridity Index (Global-Aridity)	Consortium of Spatial Information-CSI (http://www.cgiasr-csi.org)	1km

2.3 Model Design

In regression problems, the target values are continuous, and the machine consistently improves learning to better fit the model. The expected output was a predictive spatial model implemented in a geographic information system using ArcGIS Pro 2.9.0 software (ESRI, 2022). The task was to build a regression model using a non-parametric approach that enables the replicability of the WWPI with accuracy. The relationship between explanatory variables and the response variable was modelled through the tool *Train Random Trees Regression Model*.

The Random Trees algorithm is an adaptation of the Decision trees (DT) (Breiman, 1984) in a GIS environment. It is a nonparametric approach that iteratively divides a dataset into increasingly smaller subgroups using the same splitting decision (Zhang et al., 2019). The decisions are made according to the rank order of importance, optimized by a randomized procedure (Mallinis et al., 2020). DT can inherently handle nonlinear relationships, mixed predictor categories, and data gaps, and are resistant to outliers and the effects of collinearity (Osborne & Alvares-Sanches, 2019). The disadvantage of the DT is related to the models' fitness and instability due to the

propagation of errors down through subsequent splits in the tree (Breiman, 2001).

The input datasets comprising the explanatory variables and the target were in raster data format. As the cell size affects the training result and the processing time, this parameter was set up in the environment settings to make sure that the training process will keep the pixel size of 10m, as same as the target. The *Percent Samples for Testing* was set to 20%, meaning that one-fifth of the training sample was used to measure the error for interpolation in space, called test location points. This parameter evaluates three types of errors: errors on training points, errors on test points, and errors on test location points. The maximum number of trees, the maximum tree depth and the maximum number of samples were set by default with values of 50, 30 and 100000, respectively. The maximum depth of each tree refers to the number of rules each tree is allowed to create to come to a decision.

Evaluating regression performance is crucial to understand how well the model is fitted and explained by independent variables. Coefficient of determination (R-squared) and regression error (Re) were the metrics used to detect bias and the proportion of variance of the response variable. A table containing information describing the importance of each predictor used in the model was provided as output by the tool, as well as scatterplots of training data, test data, and test location data, and the regression definition file contains attribute information, statistics, and model performance.

2.4 Sensitivity Analysis

Sensitivity analyses were carried out following the process shown in Figure 1.

The diagram flow has three main steps: modelling the relationship between the response variable and predictors, measuring the model’s performance, and optimizing the model. To build the best framework a set of global environmental variables were modelled, the goodness of fit was checked, and the model was optimized as needed. This process was applied twice, on a satellite image sensed on 25/04/2018 and another on 18/08/2018.

Table 1: Description of the explanatory variables included in each model test.

Model	Exploratory variable
M1	Satellite data (Single bands)
M2	Satellite-derived data (NDVI, NDWI, NDMI)
M3	M1 + M2
M4	M3 + Digital Elevation Model (m)
M5	M4 + Average annual precipitation (mm)
M6	M5 + Average annual potential evapotranspiration (mm)
M7	M6 + Annual aridity index

2.5 Models Deployment

The most skilled model resulting from the sensitivity analysis was deployed in two different scenarios. One was applying the framework to model landscape water resources on a Sentinel-2 image sensed on

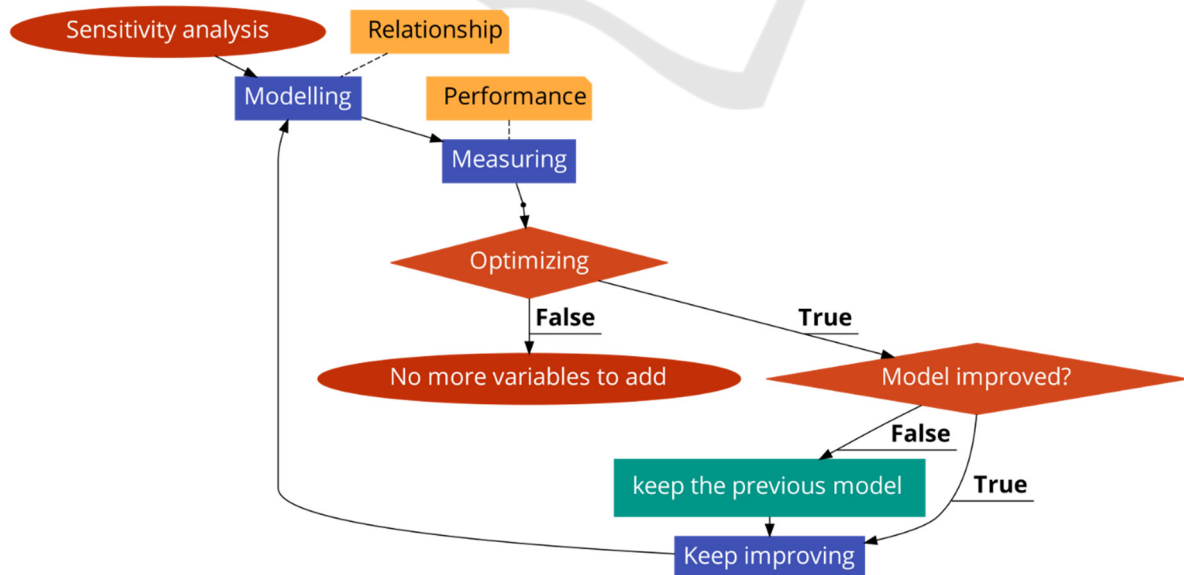


Figure 1: Diagram flow of sensitivity analysis.

29/01/2022, called temporal deployment. The second deployment was on an image taken on 23/08/2018 in another area (decimal geographic bounding box - Longitude: west = 350.99 and east = 352.28; and Latitude: south = 38.75 and north = 39.75), what was called a spatial deployment. Satellite-derived indices were calculated for both images to apply the framework.

3 RESULTS

3.1 Sensitivity Analysis

The outputs of the training process were analysed to measure the goodness of predictions. The tool checks for three types of errors: errors on training points, errors on test points, and errors on test location points. Table 3 shows the models' performance with values for the regression error (Re) at train locations (80% of all locations) and test locations (20% of all locations), and R-squared for training data, test data and test locations. The models are grouped by image date sensed, where sp refers to the image taken on 25/04/2018 and sm on 18/08/2018.

Table 3: Models performance grouped by image date sensed. Legend: TrTr (training data at training locations), TeTr (test data at training locations), and TeTe (test data at test locations), Re (regression error).

Model	TrTr		TeTr		TeTe	
	Re	R ²	Re	R ²	Re	R ²
M1 sp	1.289	0.960	2.904	0.872	3.913	0.894
M2 sp	1.827	0.924	3.751	0.777	5.258	0.809
M3 sp	1.278	0.960	2.925	0.871	3.908	0.892
M4 sp	1.004	0.957	2.490	0.820	3.413	0.851
M5 sp	0.901	0.961	2.134	0.841	3.001	0.850
M6 sp	0.801	0.962	1.980	0.800	2.706	0.811
M7 sp	0.801	0.966	1.980	0.820	2.827	0.825
M1 sm	1.218	0.954	2.488	0.848	3.541	0.878
M2 sm	1.675	0.909	3.189	0.739	4.491	0.779
M3 sm	1.191	0.951	2.455	0.835	3.431	0.886
M4 sm	0.887	0.953	2.220	0.785	3.108	0.772
M5 sm	0.771	0.958	1.950	0.805	2.652	0.824
M6 sm	0.725	0.962	1.816	0.814	2.571	0.829
M7 sm	0.722	0.962	1.830	0.789	2.490	0.839

The best performance was in model M7 for both dates. Models' predictors were the single bands, spectral indices, topography, precipitation, potential evapotranspiration, and aridity index. The image sensed on 18/08/18 has better results regarding Re than the first image.

Figure 2 is describing the importance of each predictor used in the models. The variables with the highest values are more correlated to the target and more relevant to the model. Values range between 0 and 1, and the sum of all the values equals 1.

Topography was the predictor that most contribute to both models (M7_sp and M7_sm). The predictors' importance varies widely between models. All variables showed to be important for the models, except for the single band B2 in the model M7_sm, and B3, B4 and NDMI for M7_sp.

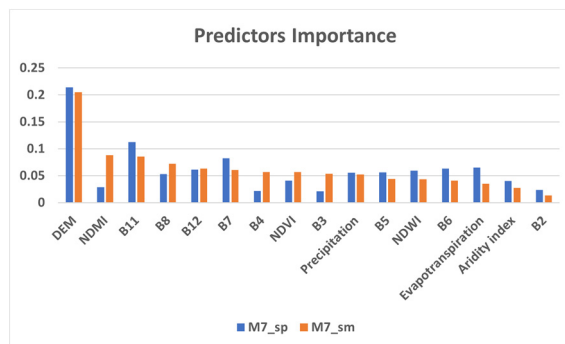


Figure 2: Predictors' importance for the best model in each group of sensitivity analysis tests.

3.2 Models Deployment

The framework was deployed to evaluate the temporal and spatial goodness of fit of the selected model. The ML regression tool measured R-squared and regression errors (Re) as the main evaluation measures of the goodness of predictions. The comparison of the true values with predicted values was done in terms of errors (standard error of the regression), the proportion of variance (R-squared) and plotting the residuals over targets. The first deployment was applying the framework at the study area but on another date (29/01/2022).

Figure 3 displays a scatterplot of predictions over-reference values and the residual plot. The R-squared is comparable with those obtained in the sensitivity analysis, with results showing very closed values, and the residual plot presenting few outliers, especially on the test data at test locations and training locations. Meaning that in locations where the WWPI detected no water (WWPI equals zero), the model predicted some degree of wetness. This can be explained by analysing the sensed date, which is from a Mediterranean wet season (29/01/2022), and on that day these locations could be flooded, or the soil moisture was higher than it used to be in the spring and summer.

The framework performance was also evaluated by deploying it using a satellite image from another area. Figure 4 shows the scatterplot of predictions over observed values and the residual plot of the spatial deployment.

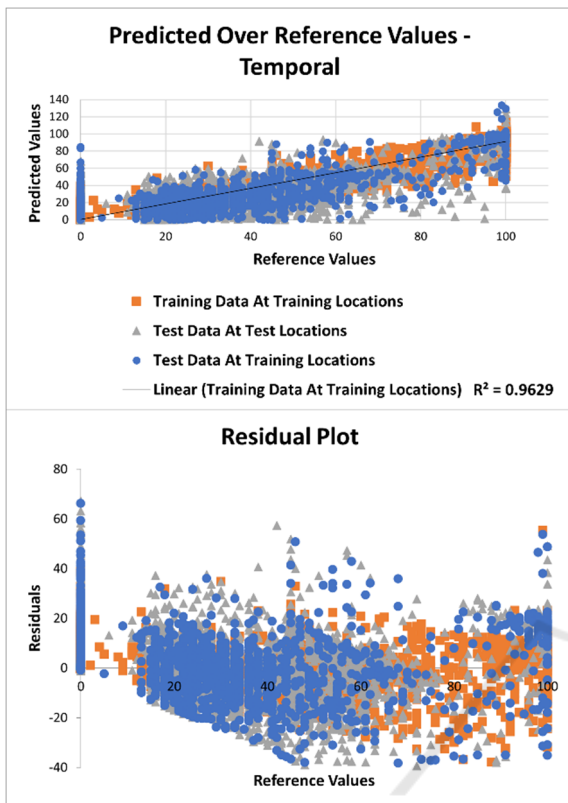


Figure 3: Scatterplot and residual plot of the temporal deployment applying the framework.

Comparing both deployments, the R-squared was better at applying the framework in the same study area rather than applying it to the other region. We also analysed R-squared and Re for test data at training locations and test data at test locations, and the overall results were better in the study area. The residuals are presenting some outliers, but more sparsely than the residual plot of the first deployment.

4 DISCUSSIONS

Water diversion and livestock grazing are threatening the functionality of many natural systems, which increases the impacts that a warmer climate may bring. When inland wetlands, rivers, and lakes, gets fragmented and become hydrologically inconsistent, the values they produce change, and can rapidly decrease. Assessing water resources at an appropriate management scale is a necessary step for the sustainable management of natural resources.

The HRL developed by CLMS are designed for use by a broad user community as the basis for environmental and regional analyses and for

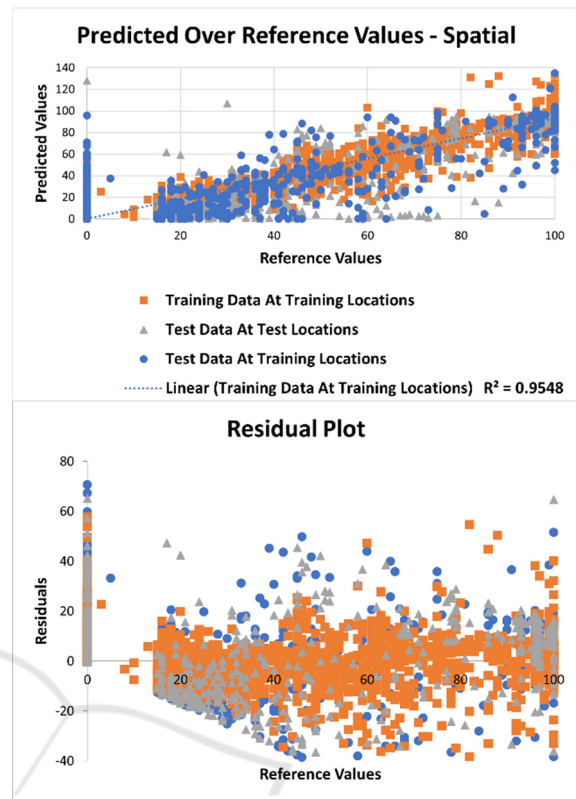


Figure 4: Spatial deployment at the study area applying the framework.

supporting political decision-making (European Environment Agency, 2023). They include geographical information on land cover and its changes, land use, vegetation state, water cycle and earth surface energy variables freely available at the European scale.

They were built based on high-resolution satellite data globally available. High spatial resolution sensors represent a smaller area of land, but with more detail and accuracy (Liang & Wang, 2019). However, higher resolution implies larger files, more loading, viewing, and processing time and more space taken up in databases, which can be a limitation. But also building environmental monitoring systems based on ML models require powerful computational resources, and most of the time a large amount of data.

ML models are only as accurate as the data they are fed in. The most known law of modelling saying, “garbage in, garbage out”, could be shown here if we did not carefully select the data to be included in the framework. Large margins of error in the predictors prevent the model from being considered correct. Models with lower error and higher R-squared are indications of higher skill predictions. Nevertheless,

none of those functions alone is enough to validate a model's predictions. Using a combination of evaluation metrics is recommended, to assess the differences between observed data and predictions, to compare models' performance, and quantify the explained variance.

Water yield and soil moisture content are some examples of ecosystem functions that are highly dependent on climatic and topographic drivers such as precipitation, evapotranspiration, temperature, and elevation. Topographic variables such as elevation and slope are very important drivers in ecological models. The topography creates specific habitats and influences the occurrence of certain plant species, as in wetlands, an ecosystem that predominantly occurs in flat landscapes, with a very low slope degree. The results show that elevation was the most important feature in both images tested, followed by the single band B11 (SWIR). The shortwave infrared bands highlight water content in vegetation.

The developed framework requires more deployment tests in different locations and time frames. Future work will expand the methodology to assess the relationship between ecosystem services and water resources at the national scale flagging the location and characteristics of water-dependent ecosystems across Portugal. But also, further research is needed to evaluate other ML algorithms and explore other spectral indices and landscape metrics that can be included as reference data in the framework. As well as testing other SEO data such as from the GRACE (Gravity Recovery and Climate Experiment) satellite mission that is used to identify regional trends of freshwater movement on the earth's surface and compare results.

5 CONCLUSIONS

The framework support studies related to the spatial estimation of water availability at the European level with the possibility of being implemented multiscale, as it includes global open-access data. The time capability of predictions depends on the launch date of the satellite data used. The comparison of models was important to build confidence in the selected predictors.

The outcomes of this study guide future research to make better use of the globally available data that can be used as predictors in data-driven water modelling. Also, to advance the knowledge of the datasets provided by CLSM at the European level. The developed framework requires more deployment tests, nevertheless, model deployment had shown

overall good results, and the capability of being used worldwide as a baseline for modelling freshwater ecosystems.

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REFERENCES

- Almeida, B. (2020). *Estudo Hidrogeológico no Sistema Aquífero Aluviões do Tejo: contributo para a sustentabilidade da massa de água subterrânea*. <http://hdl.handle.net/10451/45261>
- Almeida, B., & Cabral, P. (2021). Water yield modelling, sensitivity analysis and validation: A study for Portugal. *ISPRS International Journal of Geo-Information*, 10(8). <https://doi.org/10.3390/ijgi10080494>
- Breiman, L. (1984). *Classification and regression trees* (1st ed.). Routledge. <https://doi.org/10.1201/9781315139470>
- Breiman, L. (2001). Random forests. *Machine Learning*, 45(1), 5–32. <https://doi.org/10.1023/A:1010933404324>
- Copernicus Programme. (2023). *CLC 2018 — Copernicus Land Monitoring Service*. <https://land.copernicus.eu/pan-european/corine-land-cover/clc2018>
- Cui, Q., Ammar, M. E., Irvani, M., Kariyeva, J., & Faramarzi, M. (2021). Regional wetland water storage changes: The influence of future climate on geographically isolated wetlands. *Ecological Indicators*, 120. <https://doi.org/10.1016/j.ecolind.2020.106941>
- Domingos, P. (2012). A few useful things to know about machine learning. In *Communications of the ACM* (Vol. 55, Issue 10, pp. 78–87). <https://doi.org/10.1145/2347736.2347755>
- ESRI. (2022). *ArcGIS - ESRI (Environmental Systems Research Institute)*. <https://www.arcgis.com/index.html>
- European Environment Agency. (2023). *Copernicus Land Monitoring Service*.
- Fick, S. E., & Hijmans, R. J. (2017). WorldClim 2: new 1-km spatial resolution climate surfaces for global land areas. *International Journal of Climatology*, 37(12), 4302–4315. <https://doi.org/https://doi.org/10.1002/joc.5086>

- Hunter, F. D. L., Mitchard, E. T. A., Tyrrell, P., & Russell, S. (2020). Inter-seasonal time series imagery enhances classification accuracy of grazing resource and land degradation maps in a savanna ecosystem. *Remote Sensing*, 12(1), 198. <https://doi.org/10.3390/RS12010198>
- Kløve, B., Ala-Aho, P., Bertrand, G., Gurdak, J. J., Kupfersberger, H., Kværner, J., Muotka, T., Mykrä, H., Preda, E., Rossi, P., Uvo, C. B., Velasco, E., & Pulido-Velazquez, M. (2014). Climate change impacts on groundwater and dependent ecosystems. *Journal of Hydrology*, 518(PB), 250–266. <https://doi.org/10.1016/J.JHYDROL.2013.06.037>
- Kuenzer, C., Ottinger, M., Wegmann, M., Guo, H., Wang, C., Zhang, J., Dech, S., & Wikelski, M. (2014). Earth observation satellite sensors for biodiversity monitoring: potentials and bottlenecks. In *International Journal of Remote Sensing* (Vol. 35, Issue 18, pp. 6599–6647). Taylor and Francis Ltd. <https://doi.org/10.1080/01431161.2014.964349>
- Kundu, S., Pal, S., Mandal, I., & Talukdar, S. (2022). How far damming induced wetland fragmentation and water richness change affect wetland ecosystem services? *Remote Sensing Applications: Society and Environment*, 27, 100777. <https://doi.org/10.1016/j.rsase.2022.100777>
- Liang, S., & Wang, J. (2019). Advanced remote sensing: Terrestrial information extraction and applications. *Advanced Remote Sensing: Terrestrial Information Extraction and Applications*, 1–986. <https://doi.org/10.1016/C2017-0-03489-4>
- Mallinis, G., Chrysafis, I., Korakis, G., Pana, E., & Kyriazopoulos, A. P. (2020). A random forest modelling procedure for a multi-sensor assessment of tree species diversity. *Remote Sensing*, 12(7), 1210. <https://doi.org/10.3390/rs12071210>
- Meddens, A. J. H., Steen-Adams, M. M., Hudak, A. T., Mauro, F., Byasse, P. M., & Strunk, J. (2022). Specifying geospatial data product characteristics for forest and fuel management applications. *Environmental Research Letters*, 17(4). <https://doi.org/10.1088/1748-9326/ac5ee0>
- Mendes, M. P., Paralta, E., Batista, S., & Cerejeira, M. J. (2015). Vulnerabilidade, monitorização e risco na zona vulnerável do Tejo. *8º Congresso Da Água*.
- Mendonça, J. J. L. (1990). *Estudo Estatístico dos Parâmetros Hidráulicos do Sistema Aquífero Aluvionar do Tejo*.
- Mpakairi, K. S., Dube, T., Dondofema, F., & Dalu, T. (2022). Spatio-temporal variation of vegetation heterogeneity in groundwater dependent ecosystems within arid environments. *Ecological Informatics*, 69, 101667. <https://doi.org/10.1016/j.ecoinf.2022.101667>
- Novo, M. E., Oliveira, M., Martins, T., & Henriques, M. J. (2018). Projecto Bingo: O Impacto das Alterações Climáticas na Componente Subterrânea do Ciclo Hidrológico. *Revista Recursos Hidricos*, 39(2), 59–74. <https://doi.org/10.5894/rh39n2-cti3>
- Nti, E. K., Cobbina, S. J., Attafua, E. E., Opoku, E., & Gyan, M. A. (2022). Environmental sustainability technologies in biodiversity, energy, transportation and water management using artificial intelligence: A systematic review. *Sustainable Futures*, 4, 100068. <https://doi.org/10.1016/J.SFTR.2022.100068>
- Osborne, P. E., & Alvares-Sanches, T. (2019). Quantifying how landscape composition and configuration affect urban land surface temperatures using machine learning and neutral landscapes. *Computers, Environment and Urban Systems*, 76, 80–90. <https://doi.org/10.1016/j.compenvurbsys.2019.04.003>
- Ramos, T. B., Horta, A., Gonçalves, M. C., Pires, F. P., Duffy, D., & Martins, J. C. (2017). The INFOSOLO database as a first step towards the development of a soil information system in Portugal. *Catena*, 158(July), 390–412. <https://doi.org/10.1016/j.catena.2017.07.020>
- Ribeiro, M. M. S. (1998). *Contribuição para o conhecimento hidrogeológico do Cenozóico na Bacia do Baixo Tejo*.
- Ringersma, J., Batjes, N., & Dent, D. (2003). Green Water: definitions and data for assessment. *ISRIC – World Soil Information*, December, 83.
- Sharma, A., Hubert-Moy, L., Buvaneshwari, S., Sekhar, M., Ruiz, L., Bandyopadhyay, S., & Corgne, S. (2018). Irrigation History Estimation Using Multitemporal Landsat Satellite Images: Application to an Intensive Groundwater Irrigated Agricultural Watershed in India. *Remote Sensing*, 10(6), 893. <https://doi.org/10.3390/rs10060893>
- Shi, H., Li, L., Eamus, D., Cleverly, J., Huete, A., Beringer, J., Yu, Q., Van Gorsel, E., & Hutley, L. (2014). Intrinsic climate dependency of ecosystem light and water-use-efficiencies across Australian biomes. *Environmental Research Letters*, 9(10). <https://doi.org/10.1088/1748-9326/9/10/104002>
- Trabucco, A., & Zomer, R. (2019). *Global Aridity Index and Potential Evapotranspiration (ET0) Climate Database v2*. <https://doi.org/10.6084/m9.figshare.7504448.v3>
- Yang, X., Liu, S., Jia, C., Liu, Y., & Yu, C. (2021). Vulnerability assessment and management planning for the ecological environment in urban wetlands. *Journal of Environmental Management*, 298, 113540. <https://doi.org/10.1016/j.jenvman.2021.113540>
- Zhang, J., Okin, G. S., & Zhou, B. (2019). Assimilating optical satellite remote sensing images and field data to predict surface indicators in the Western U.S.: Assessing error in satellite predictions based on large geographical datasets with the use of machine learning. *Remote Sensing of Environment*, 233, 111382. <https://doi.org/10.1016/j.rse.2019.111382>