

Antarctic biodiversity predictions through substrate qualities and environmental DNA

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Antarctic conservation science is crucial for enhancing Antarctic policy and understanding alterations to terrestrial Antarctic biodiversity. Antarctic conservation will have limited long-term impacts in the absence of large-scale biodiversity data, but if such data were available, it is likely to improve environmental protection regimes. To enable the prediction of Antarctic biodiversity across continental spatial scales through proxy variables, in the absence of baseline surveys, we linked Antarctic substrate-derived environmental DNA (eDNA) sequence data from the remote Antarctic Prince Charles Mountains to a selected range of concomitantly collected measurements of substrate properties. We achieved this through application of a statistical method commonly used in machine learning. Our analysis indicated that neutral substrate pH, low conductivity, and certain substrate minerals are important predictors of the presence of basidiomycetes, chlorophytes, ciliophorans, nematodes, and tardigrades. A bootstrapped regression revealed how variations in the identified substrate parameters influence probabilities of detecting eukaryote phyla across vast and remote areas of Antarctica. We believe that our work will improve future taxon distribution modeling and aid in developing more targeted surveys of biodiversity conducted under logistically challenging conditions.

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Although only 0.3% of continental Antarctica is ice-free, many organisms, including bacteria, unicellular eukaryotes, fungi, lichen, cryptogamic plants, and invertebrates, are scattered across the continent in island-like terrestrial habitats, soil-like substrates, lakes, and cryoconite holes (small depressions formed by melting of snow or ice below radiation-absorbing dark-colored dust or soot) (Convey *et al.* 2014; Chown *et al.* 2015). Anthropogenic threats, including climate change, pollution, and the introduction of invasive species, among others, imperil Antarctic biodiversity. Mitigation of these threats will rely on implementation of well-tailored management strategies across the continent's bioregions (Coetzee *et al.* 2017).

Effective continental-scale conservation management requires concomitant continental-scale data (Wauchope *et al.* 2019). However, knowledge of terrestrial Antarctic biodiversity remains limited because logistical difficulties exacerbated by harsh environmental conditions, along with funding constraints, impede research in Antarctica's ice-free areas. Environmental DNA (eDNA) analysis, despite shortcomings, represents one of the more practical and economical options for continental-wide surveys of terrestrial Antarctic biodiversity, given logistical challenges (Czechowski *et al.* 2017). Comparable

large-scale, systematic approaches to protect soil biodiversity are required globally but are often limited to charismatic taxa, such as those found in Arctic regions (Gillespie *et al.* 2020).

In this study, we linked commonly measured substrate properties to the cryptic eukaryotic biodiversity of terrestrial Antarctic ice-free regions. Soil nutrient status is the most important attribute of biodiverse soils (Geisen *et al.* 2019), and corresponding key variables can be, and are, routinely measured economically. We analyzed molecular data (eDNA) from an extremely remote Antarctic terrestrial region to clarify relationships between substrate properties and the presence of eukaryotic phyla. We believe that such an approach will be useful for predicting biodiversity encompassing a wide taxonomic spectrum across extensive areas of the Antarctic and especially for identifying regions worthy of lower-level taxonomic biodiversity surveys, possibly realized through the use of “barcoding” via mitochondrial DNA (such as with mitochondrial cytochrome oxidase 1) or logistically more challenging field biodiversity surveys.

The Prince Charles Mountains (PCMs), the most remote terrestrial area in eastern Antarctica, were first recorded by the US Operation Highjump (a US Navy initiative to establish an Antarctic research base conducted in 1946–1947) and mapped in greater detail during subsequent Australian (1954–1961) and Russian (1983–1991) expeditions. In 2011, we obtained eDNA samples from substrates collected throughout the PCMs and measured geochemical and mineral properties. Previously, Czechowski *et al.* (2016b) focused on invertebrates as the primary substrate-inhabiting metazoans and discovered major changes in their distribution over salinity gradients, as known from other Antarctic areas and taxa (Bottos

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et al. 2020). Here, we expanded our analyses of environmental variables to all eukaryotic phyla, thereby exploring approaches of inferring biodiversity presence that could be applied across the entirety of ice-free Antarctica. Beyond phylum-level surveys, our technique could be applied using other genetic markers and predictors to link future smaller scaled conservation projects anywhere in terrestrial Antarctica, aid taxon distribution modeling, and contribute toward improving conservation management strategies across Antarctic bioregions.

■ Methods

Fieldwork took place in the PCMs (East Antarctica; [Figure 1](#)) from 26 November 2011 to 21 January 2012, with soil sampling conducted near Mount Menzies (MM; 73°25'29.38"S, 62°0'37.61"E), at the Mawson Escarpment (ME; 73°19'16.91"S, 68°19'31.20"E), and at Lake Terrasovoje (LT; 70°32'23.58"S, 67°57'28.05"E), as described in [Czechowski *et al.* \(2016a,b\)](#). A total of 154 field samples (26 from MM, 70 from ME, and 58 from LT) were considered in this analysis ([WebTable 1](#)).

To infer climatic conditions in the PCMs, we used raster layers from Quantarctica v3 (<https://www.npolar.no/quantarctica>; [Matsuoka *et al.* 2021](#)) encoding annual mean precipitation (mm), wind speed (m s^{-1} at 10 m above the ground surface) and mean annual temperature ($^{\circ}\text{C}$ at 2 m above the ground surface, as this was the only temperature data available via Quantarctica). We disaggregated the layer rasterization from 35 kilometers per pixel (km px^{-1}) to 1 km px^{-1} through bilinear interpolation. We then extracted median values for the three variables from a 20-km buffer surrounding each sampling location ([WebFigure 1](#)).

As predictor data for the presence of eukaryotic phyla in substrates, geochemical composition (ammonium $[\text{NH}_4^+]$, carbon [C], soil density [ρ], nitrate $[\text{NO}_3^-]$, soil pH in water $[\text{pH}_{\text{H}_2\text{O}}]$, soil pH in calcium chloride $[\text{pH}_{\text{CaCl}_2}]$, phosphorus [P], potassium [K], sulfur [S], and soil texture) was analyzed by the agricultural soil testing service Australian Precision Ag Laboratory (www.apal.com.au). Measurements below detection levels were excluded to yield data completeness of at least 96.7% ([WebPanel 1](#)). The final analysis included K, S, ρ , and $\text{pH}_{\text{CaCl}_2}$ ($\text{pH}_{\text{H}_2\text{O}}$ was excluded as co-linear and texture was excluded as categorical). As additional predictors, substrate mineral compositions were considered through integration of X-ray diffraction spectra of quartz, calcite, feldspar, titanite, pyroxene/amphibole/garnet, micas, dolomite and kaolin/chlorite, and chlorite ([Czechowski *et al.* 2016b](#)). We handled the sum-to-unity constraint of the mineral compositions by excluding quartz, the most common mineral, from further analysis. As further predictors for most locations (MM: $n = 26$, ME: $n = 69$, LT: $n = 57$), we included unpublished measurements of soil-substrate adenosine 5'-triphosphate (ATP) (eg [Conklin and Macgregor 1972](#)), obtained with a Clean-Trace Luminometer (3M; Maplewood, Minnesota), and slope measurements. Prior to regression, all predictors were

standardized to zero mean and unit variance. Predictor densities are provided in [WebFigure 2](#).

Biological response data were prepared in QIIME 2020-2 ([Bolyen *et al.* 2019](#)) and R v4.0.0 (R Core Development Team 2019) from raw sequence data generated as described in [Czechowski *et al.* \(2016b, 2017\)](#). Briefly, 125 base pair (bp) eukaryotic 18S rDNA polymerase chain reaction (PCR) products (with 85 bp target region) were previously amplified using the primers “Euk1391f” and “EukBr” ([Caporaso *et al.* 2012](#)), which were established specifically for eukaryotic microbial surveying ([Thompson *et al.* 2017](#)). Following accepted recommendations, PCRs were conducted in triplicate, with each replicate carrying identical barcodes. The resulting eDNA libraries were combined for sequencing across two Illumina MiSeq runs ([WebFigure 3](#)). We redefined amplicon sequence variants (ASVs) ([Callahan *et al.* 2017](#)) from those data with QIIME: after pre-filtering (Phred score ≥ 25), we trimmed read pairs with Cutadapt v1.18 ([Martin 2011](#)), and denoised using DADA2 v1.6.0 ([Callahan *et al.* 2016](#)). We retained merged reads with an expected error value less than 3 that were not deemed chimeric.

Due to the shortness and slow evolution of the employed 18S marker, we elected to conduct our analyses at the phylum level, and to use species-level assignments solely to verify data credibility. Accordingly, we designed the retrieval of taxonomic annotations for our Antarctic DNA sequences to yield reliable species identifications in cases where Antarctic reference data were available, while still returning higher taxonomic (eg phylum-level) identifications when closely matching reference data were unavailable. Doing so enabled inclusion of a larger number of Antarctic sequences into our statistical analysis at the phylum level, but rendered species-level identifications as potentially unreliable, requiring verification at the alignment level. We identified eukaryotic sequences among our reads with a recent local copy (April 2020) of the entire National Center for Biotechnology Information (NCBI) nucleotide collection using the Basic Local Assignment Search Tool (BLAST) v2.10.0+. Taxonomic assignments were retrieved from reference sequences at least 50% identical to queries, with an assignment significance threshold (*e* value) of 10^{-10} , considering only matches with at least 90% coverage and excluding environmental sequences (*evaluate* $1e^{-10}$, *max_hsp* 5, *max_target_seqs* 5, *qcov_hsp_perc* 90, and *perc_identity* 50). For each Antarctic search query, we used the highest bit score among all NCBI-returned sequences for that query to choose the final taxonomic assignment. Subsequently, we used the R package *decontam* ([Davis *et al.* 2018](#)) to remove putatively contaminating reads, and likewise subtracted all sequences and taxa in negative controls from field samples. Because we focused on eukaryotes, all reads identified as non-eukaryotic were discarded ([WebFigure 4](#)). Post-filtering, we evaluated whether sufficient data were retained for subsequent analysis, both by bootstrapping simulated accumulation curves ($n = 1000$ per

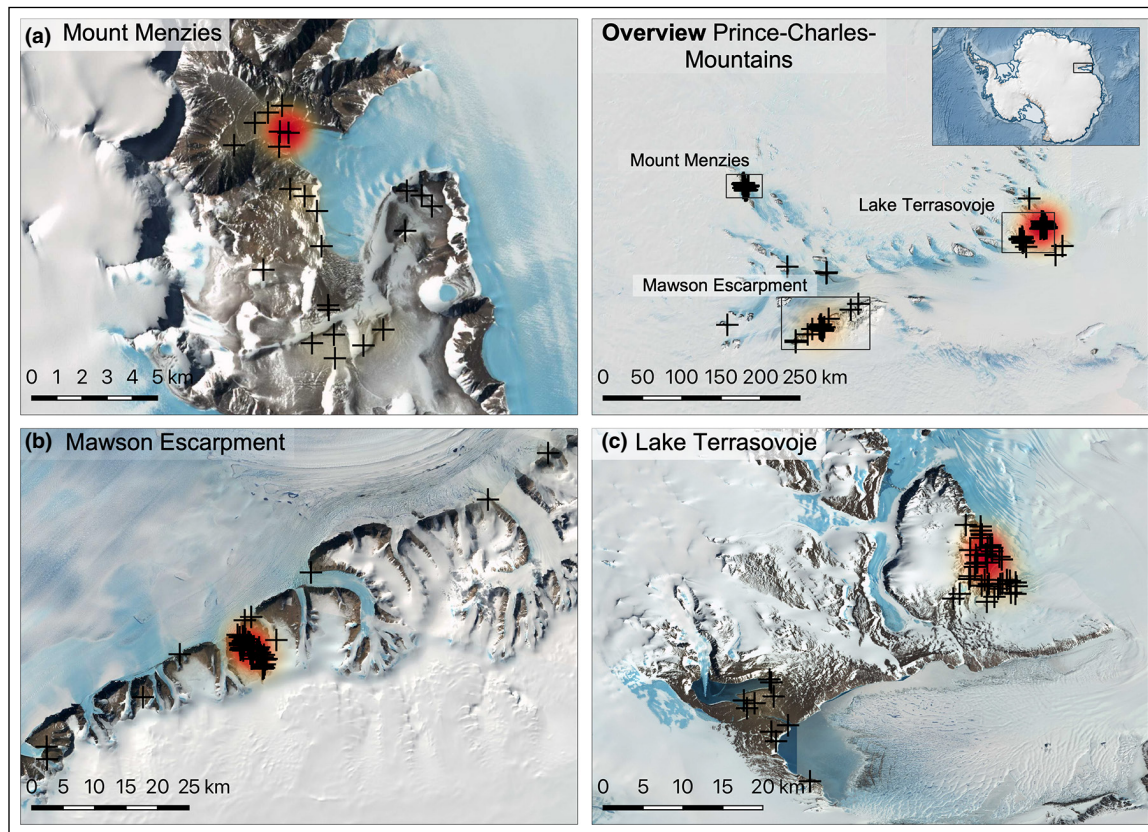


Figure 1. Sampling area in the Prince Charles Mountains, in eastern Antarctica. (a) Mount Menzies (MM), (b) the Mawson Escarpment (ME), and (c) Lake Terrasovoje (LT). All sampling locations are marked with a crosshair. Heat shading (at map scale) indicates density of 18S amplicon sequence variants (ASVs) (sensu Callahan *et al.* 2017) determined to be significantly influenced by substrate qualities as available. Base layers compiled by the Norwegian Polar Institute and distributed in the Quantarctica package (www.npolar.no/quantarctica). Base layers courtesy of Scientific Committee on Antarctic Research (SCAR) Antarctic Digital Database (© SCAR, 1993–2015); The National Snow and Ice Data Centre, University of Colorado, Boulder; the National Aeronautics and Space Administration's (NASA's) Visible Earth Team (<http://visibleearth.nasa.gov>); and the Australian Antarctic Division (© Commonwealth of Australia, 2006).

sample) and by analyzing phylum accumulation per sample with accumulating reads, as an absorbing Markov chain.

Using the *lasso* technique (Tibshirani 1996) of the R package *glmnet* (Friedman *et al.* 2010), we regressed each phylum present in at least 12 samples against the aforementioned predictors (WebFigure 5). In regressions, we disregarded sequence read abundances as meaningless due to inherent constraints of amplicon sequencing (Czechowski *et al.* 2017), analyzed presences instead, and used the most biodiverse of all locations (Czechowski *et al.* 2016b) as a reference location, such that predictor effects at MM and ME are reported as relative to those at LT. We initially retrieved the active set (variables not set to zero) estimated by *lasso*, repeated the regression of phylum presence against 1000 randomly chosen sample-sets of predictors, and calculated the number of times each variable was estimated to be non-zero: variables were considered significant if reported as non-zero more than 950 times. Accordingly, 95% non-parametric bootstrap confidence intervals (CIs) were also calculated for our estimates (ie 5% significance level). We did not adjust for multiple comparisons.

Hoping to find evidence of localized Antarctic invertebrate occurrences (Convey *et al.* 2014), we obtained putative species-level assignments among phyla significantly influenced by environmental predictors (see below) by querying the Global Biodiversity Information Facility (GBIF; www.gbif.org), iNaturalist (www.inaturalist.org), and Biodiversity Information Serving Our Nation (BISON; www.gbif.us, formerly [bison.usgs.gov](http://www.bison.usgs.gov)) databases with the R package *spocc* (see WebPanel 1 for detailed methods).

Results

Taking into consideration the coarse raster resolution and model-like character of the climate data, annual mean climate at MM was the coldest ($-32 \pm 0.3^\circ\text{C}$) and windiest ($10.2 \pm 0.05 \text{ ms}^{-1}$) of the three locations, but with intermediate precipitation ($86 \pm 1 \text{ mm}$) (WebFigure 1), whereas ME had the least precipitation ($55.3 \pm 7 \text{ mm}$), comparatively low wind speeds ($5.4 \pm 0.5 \text{ ms}^{-1}$), and slightly higher temperatures than MM ($-28.4 \pm 0.6^\circ\text{C}$). Closest to the coast and largely exposed, LT exhibited the highest precipitation

strong positive relationship of this phylum with dolomite ($\mu = 0.025\%$, $\sigma = 0.05\%$, $E[\text{present}_{\mu+1\sigma}] = 0.7$); (2) very low levels of chlorophytes (47 species; [Figure 2b](#)) at MM plausibly attributable to harsh environmental conditions (see [WebFigure 1](#)) ($E[\text{present}_{LT}] = 0.61$ and $E[\text{present}_{MM}] = 0.32$, including more alkaline substrates, $E[\text{present}_{\mu+1\sigma}] = 0.46$); (3) very low levels of ciliophorans (47 species; [Figure 2c](#)) at MM ($E[\text{present}_{LT}] = 0.70$ and $E[\text{present}_{MM}] = 0.39$), in sulfur-rich substrates ($\mu = 528 \text{ mg kg}^{-1}$, $\sigma = 1410 \text{ mg kg}^{-1}$, $E[\text{present}_{\mu+1\sigma}] = 0.61$), and in areas relatively rich in pyroxene, amphibole, or garnet ($\mu = 4\%$, $\sigma = 4\%$, $E[\text{present}_{\mu+1\sigma}] = 0.52$); (4) very low levels of nematodes (eight species; [Figure 2d](#)) at MM ($E[\text{present}_{LT}] = 0.47$ and $E[\text{present}_{MM}] = 0.28$), and in highly conductive substrates ($\mu = 0.55 \text{ decisiemens per meter [dS m}^{-1}]$, $\sigma = 1.07 \text{ dS m}^{-1}$, $E[\text{present}_{\mu+1\sigma}] = 0.35$); and (5) very low levels of tardigrades (nine species; [Figure 2e](#)) in alkaline substrates ($E[\text{present}_{\mu}] = 0.22$, $E[\text{present}_{\mu+1\sigma}] = 0.14$).

Observed fractions of non-zero coefficients are provided in [Table 1](#) and in [WebPanel 1](#) (95% non-parametric bootstrap CIs for non-zero estimates are also in [WebPanel 1](#)). Directions of all predictor effects on all analyzed taxa presences, including non-significant effects, are shown in [WebPanel 1](#).

For 66 of the 173 putative species assignments, 778 georeferenced records could be obtained (of those, 65% derived from GBIF, 27% from iNaturalist, and 7% from BISON). Of the obtained 123 locations, 4% were in Africa, 1.6% were in Antarctica, 13% were in Asia, 32% were in Europe, 21% were in North America, and 10% were in South America ([WebPanel 1](#)). The sole species recorded for Antarctica south of the polar circle (south of 66.56°S) was the nematode *S lindsayae*. Observations north of the polar circle (north of 66.56°S) included Basidiomycota (*Gloiocephala aquatica*, *Stereum*

rugosum, *M frigida*, *Rhodotorula mucilaginosa*), chlorophytes (*Haematococcus lacustris*, *Oophila amblystomatis*), and ciliophorans (*Furgasonia blochmanni*, *Chilodonella acuta*, *Tachysoma pellionellum*). Refer to [WebTable 3](#) for alignment qualities.

Discussion

Our work demonstrates two key technologies useful for performing baseline biodiversity surveys across large spatial scales in extremely remote environments: (1) robust statistics (such as *lasso*), often used in machine-learning algorithms (Muthukrishnan and Rohini 2016) and (2) biodiversity information derived from eDNA (Czechowski et al. 2017). To the best of our knowledge, our work is the first in associating eDNA data to environmental predictors using *lasso* to yield accurate detection probabilities for taxonomic groups, also in Antarctica. Thus, we present an analytical framework to identify areas for targeted species-level biodiversity surveys, using other markers, or predictors for Antarctica, and possibly for other locations that are difficult to access.

Our expanded analyses of the original data (Czechowski et al. 2016b) make use of new eDNA sequence processing algorithms (Callahan et al. 2016, 2017), along with more extensive reference databases for taxonomic assignment and new algorithms available with R. Our results align with earlier findings relating the distribution of eukaryotes to their environment in the PCMs and Antarctica (Czechowski et al. 2016a,b; Bottos et al. 2020), but improve the accuracy of those findings for five phyla.

A key strength of our analyses is the relatively easy retrieval of survey data encompassing many phyla (probably including many cryptic and unknown species) across many samples. Conversely, a weakness of the employed 18S marker is its

Table 1. Numerical summary of significant coefficient estimates for each phylum as obtained through *lasso* logistic regression

Phylum	Predictor	95% CI coefficient		95% CI odds ratio		Proportion of bootstrap replicates not zero
		Lower	Upper	Lower	Upper	
Basidiomycota	Dolomite	0	1.32	1	3.74	0.93
	PH	-1.54	-0.46	0.21	0.63	1.00
Chlorophyta	MM	-1.32	-0.10	0.27	0.90	0.99
	PH	-1.28	-0.10	0.28	0.90	0.99
Ciliophora	Garnet	-2.07	-0.11	0.13	0.90	0.99
	MM	-1.22	0.00	0.30	1.00	0.93
	Sulfur	-3.14	0.00	0.04	1.00	0.85
Nematoda	Conductivity	-2.17	0.00	0.11	1.00	0.99
	MM	-2.10	-0.26	0.12	0.77	0.99
Tardigrada	PH	-1.42	0.00	0.24	1.00	0.95

Notes: Confidence intervals (CIs) for the coefficients are provided on the logit scale. Units of the coefficient CI limits in logistic regression are the reciprocal of the unit of the predictor. Exponentiation of coefficient CI will result in the shown CIs for the odds ratios. Odds ratios are unitless: if they are greater than 1, then the predictor has a positive influence on the probability of finding the given taxon. Mount Menzies abbreviated as "MM".

limited ability to discern many distinct sequence variants at low taxonomic levels (eg at the species level). Regardless, identification of species with likely Antarctic occurrence, such as the nematode *S lindsayae* and the tardigrade *M furciger*, through the use of a relatively short and highly conserved pair of primers, highlights the ability of eDNA to retrieve species occurrence records, provided that sufficient sequence data are available for taxonomic assignment. Consequently, we believe that eDNA analysis should be the method of choice for obtaining biodiversity data from Antarctica, particularly when many samples are to be analyzed, but other markers and eDNA analysis techniques are needed to investigate fine-scaled endemism and to obtain better taxonomic resolution.

Georeferencing our putative species assignments with publicly accessible data had limited success, as expected (Cameron *et al.* 2018). The limitations of reference databases became obvious when species known to occur in the Antarctic, such as *Acutuncus antarcticus* (WebPanel 1), identified here through a perfect alignment (bit score 154.6), were absent from the reference databases, and only 38% of all putative Antarctic species assigned by us were georeferenced at all. High occurrence prevalence in North America and Europe not only indicates sampling bias in the GBIF, iNaturalist, and BISON databases, alongside substantial weaknesses of publicly accessible global biodiversity data concerning cryptic eukaryote species, but also supports our choice of merely conducting a phylum-level analysis despite overall well-matching species-level alignments.

Eukaryotic distribution patterns reported in previous Antarctic studies provide context for our observations from the PCMs. The rarity of chlorophytes, ciliophorans, and the otherwise ubiquitous nematodes at MM in relation to the two other, lower altitude, more northerly locations (ME, LT) seem to confirm trends of increasing eukaryotic richness and diversity with decreasing latitude and altitude (Czechowski *et al.* 2016a; Thompson *et al.* 2020; Zhang *et al.* 2020). However, such patterns are not always evident at the scales investigated here; rather, Antarctic biodiversity can be surprisingly regionalized (Convey *et al.* 2014), and our analysis revealed surprisingly high eukaryotic diversity to unexpectedly occur even in the harshest environments, such as local ice–soil substrate boundaries at MM (Figures 1a and 2). The absence of ciliophorans from sulfur-rich substrates and of nematodes from highly conductive soil interstices matches the findings from previous work employing non-eDNA approaches, in which distribution patterns were reported to be shaped by age-related salt accumulation at the surface–air interface of frozen soils (Velasco-Castrillón *et al.* 2014b; Lee *et al.* 2019).

In the absence of other predictors, the results of our analysis emphasize the importance of neutral substrate pH, low conductivity, and key minerals (dolomite, pyroxene, amphibole, or garnet) for predicting high eukaryotic density in Antarctic substrates and corroborate the negative influence of substrate alkalinity on Antarctic Basidiomycota reported by Arenz and Blanchette (2011). Bioregionalization notwithstanding, distance to coast appears to be a suitable proxy variable negatively

related to the presence of chlorophytes and ciliophorans, supporting previous findings (Thompson *et al.* 2020). In addition, soil alkalinity and substrate concentrations of sulfur, pyroxene, amphibole, or garnet also appear to constrain chlorophyte and ciliophoran distributions. Among nematodes, our results (ie perfect alignment between our Antarctic 18S sequence from MM and an annotated reference sequence) indicate that *S lindsayae* could occur in high altitude and high latitude environments such as MM, but in such environments would be influenced by the species' general indifference (rather than affinity; compare with Zawierucha *et al.* 2019) to alkaline substrates, and must be highly localized (at least at MM) if encountered at high abundance (Smykla *et al.* 2018; Zawierucha *et al.* 2019). Finally, we confirm the negative association between tardigrade occurrence and alkaline substrates previously observed in Victoria Land (Smykla *et al.* 2018).

Antarctic ice-free areas exhibiting high annual mean precipitation, low wind speeds, and relatively high temperatures, with substrates possessing neutral pH, low conductivity, abundance in dolomite, and scarcity in pyroxene, amphibole, or garnet, are likely to be highly biodiverse and should harbor candidates for focused conservation management. Furthermore, locations with more extreme environmental conditions may harbor endemic relic fauna equally warranting protection (Convey *et al.* 2014). Our results align with observations in other (including polar and alpine) ecosystems, where soil pH was an important factor determining bacterial and fungal community composition (Siciliano *et al.* 2014; Bottos *et al.* 2020). At the same time, Antarctic soil ecosystems are relatively simple and are assumed to largely lack complex biotic interactions, although such interactions may be more common in coastal terrestrial ecosystems (Velasco-Castrillón *et al.* 2014b; Lee *et al.* 2019). Consequently, the observed distribution patterns of soil eukaryotes (particularly at MM) are likely predominantly shaped by abiotic factors and would be gradually more influenced by limited biotic interactions, lower latitude substrates, or more coastal substrates (ME, LT).

■ Conclusions

Our analysis highlights the utility of environmental molecular data and predictive analysis algorithms to detect the presence of eukaryotes by means of relatively easily measured soil predictors, which can be combined with readily available climate data. Rather than recognizing trends, our analytical technique provides accurate detection probabilities for Basidiomycota, chlorophytes, nematodes, and tardigrades in relation to bedrock mineral composition, pH, conductivity, sulfur content, and, arguably, overall harshness of environmental conditions. These relationships, as quantified here, enable more precise modeling of phylum distributions over large spatial scales. Our approach may be used to identify regions worthy of species-level biodiversity surveys, possibly employing faster evolving molecular markers, other eDNA analysis techniques, or logistically more challenging

morphologic biodiversity assessments. We believe our approach to be valuable to better inform both Antarctic biogeography and delineation of conservation areas.

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Data Availability Statement

Some data were included in previous publications, which are cited in the text. All data were provided as in-confidence for peer review and have been revised during peer review. Data associated with this work are located at <https://doi.org/10.5281/zenodo.4579840>. External occurrence data are additionally registered via <https://doi.org/10.15468/dd.rkjads>. Code is available at <https://github.com/macrobioetus/pcm-eukaryotes> and <https://github.com/OldMortality/eukaryotes>.

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Supporting Information

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A story behind what's on the menu

For some species with a nonselective diet, which seem to eat everything, the story behind their menu selection warrants much investigation. Omnivores have a generalist diet governed by competitive pressures and resource availability. The degree of meat consumption by omnivores informs their trophic position within the community of sympatric species. To the extent that competition determines diet choice by omnivores, in regions experiencing declines in apex mammalian carnivores, which usually regulate ungulate prey populations, carnivory by omnivores may increase. In Ghana, preliminary evidence from stable isotope analysis suggested a trophic ascension in baboons (*Papio* spp) where populations of African lions (*Panthera leo*) have been extirpated (*Ecological and conservation implications of mesopredator release*. In: *Trophic cascades: predators, prey, and the changing dynamics of nature*. Washington, DC: Island Press; 2010). This photograph of ungulate carnivory by an olive baboon (*Papio anubis*) provides some evidence of such an ascension. The image – which was captured during an extensive camera trap survey investigating carnivores within the W-Arly-Pendjari (WAP) protected area complex in Burkina Faso, West Africa – depicts a neonatal ungulate that is normally consumed by lions (*Conserv Lett* 2019; doi.org/10.1111/conl.12667). Although small and largely fragmented populations of lions remain throughout West Africa, research and conservation efforts along with financial support for effective protected area management remain paramount. Future investigations need to determine the degree of dietary overlap between lions and baboons as well

as whether carnivory in baboons is highest in areas with reduced or extirpated lion populations. As African lion populations continue to decline, so does their regulatory capacity. As a result, new niches become available for more abundant species like baboons to begin the classic trophic cascade.

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