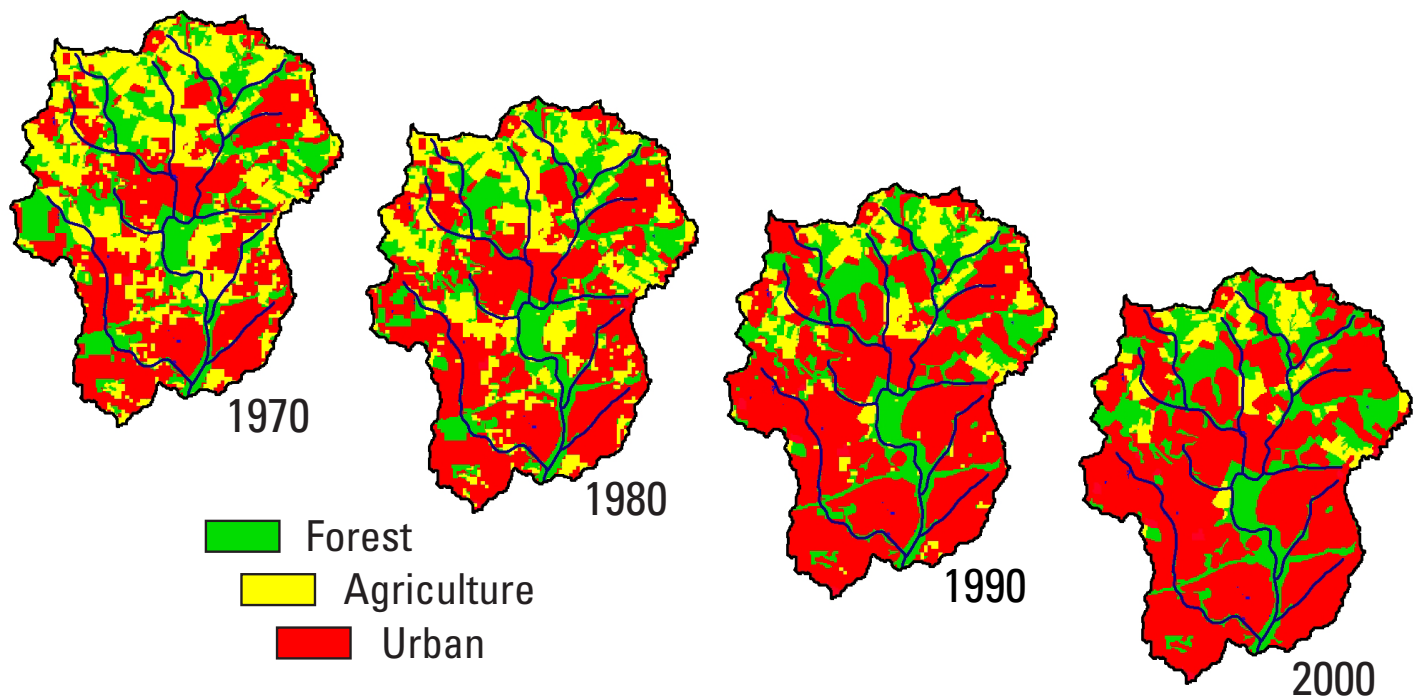


Methods for Adjusting U.S. Geological Survey Rural Regression Peak Discharges in an Urban Setting



Scientific Investigations Report 2006-5270

Cover. Land use change in the Northwest Branch Anacostia River near Colesville, Maryland (USGS streamgage 01650500). The images show the progression of urban development in this watershed over the period 1970 to 2000.

Methods for Adjusting U.S. Geological Survey Rural Regression Peak Discharges in an Urban Setting

By Glenn E. Moglen and Dorianne E. Shivers

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Contents

Abstract.....	1
Introduction.....	1
Previous Investigations.....	2
Overview of Peak Discharge Adjustment Method.....	4
Assumptions of Peak Discharge Adjustment Method.....	4
Enumeration of Steps in Peak Discharge Adjustment Method.....	4
Analytical Presentation of Adjustment Method.....	5
Adjustment of the Observed Annual Maximum Time Series for Nonstationary Urbanization.....	7
The Adjustment Models.....	7
Censoring of Streamgages with Non-elevated Flood-Frequency Values.....	9
Example—the Calibrated 5-year Simple Imperviousness Adjustment Model.....	9
Methods.....	11
Development of Rural Peak Discharge Estimates.....	11
Delineation of Watershed Boundaries.....	11
Estimation of Watershed Characteristics.....	11
Development of Flood-frequency Estimates from Gage Record.....	11
Selection of Gages.....	11
Observed Annual Maximum Time Series.....	12
Use of the PeakFQ Program.....	12
Skew.....	12
PeakFQ Run-Time Parameters.....	12
Identifying the High-Outlier Threshold.....	12
Identifying the Historic Period for the Largest Flood in the Time Series.....	13
Example Application of Outlier and Historical Period Identification Procedures.....	13
Land-Use Change Time Series.....	14
Population Density.....	14
Imperviousness.....	16
Estimating Imperviousness from Census Data.....	16
Implementation of Historical Census Data.....	16
Results.....	18
Comparison to Urban Equations Developed by Sauer and others (1983).....	32
Application of Calibrated Models to Streamgages with Local Urban Equations.....	34
Interpretation of Results.....	35
Model Strengths.....	35
Model Weaknesses.....	36
Choosing a Model.....	36
Summary and Conclusions.....	37
Notation.....	38
Discharge Adjustment Model Definitions:.....	39
Acknowledgements.....	40
References.....	40

Appendix 1.....	44
Appendix 2.....	48
Appendix 3.....	52
Appendix 4.....	53
Appendix 5.....	54
Appendix 6.....	55

Figures

Figure 1. Locations of urban and rural streamgages used in current (2006) study of rural regression peak discharges.....	3
Figure 2. Flow chart for the process for calibrating adjustment equations.....	6
Figure 3. The relation between observed and scaled imperviousness for the scaled values of $I^*=12.1$ and $c_d=0.2$	8
Figure 4. Annual maximum time series for U.S. Geological Survey streamgage 02037800, Falling Creek near Midlothian, Virginia.....	14
Figure 5. Population density in Montgomery County, Maryland, as a function of census tract area from the 2000 census.....	15
Figure 6. Distribution of population density as a fraction of the total watershed area for U.S. Geological Survey streamgage 01645000, Seneca Creek at Dawsonville, Maryland.....	15
Figure 7. Relations of imperviousness to population density as measured by Stankowski (1972) and as measured in this study.....	17
Figure 8. Estimated imperviousness from census tract data for U.S. Geological Survey streamgage 01645000, Seneca Creek at Dawsonville, Maryland.....	17
Figure 9. Application of the (A) simple imperviousness model, (B) imperviousness distribution model, and (C) scaled imperviousness model to the ratio of urban to rural discharges as a function of return period and imperviousness at U.S. Geological Survey streamgage 01645000, Seneca Creek at Dawsonville, Maryland.....	28
Figure 10. Application of the (A) simple density model, (B) density distribution model, and (C) scaled density model to the ratio of urban to rural discharges as a function of return period and imperviousness at U.S. Geological Survey streamgage 01645000, Seneca Creek at Dawsonville, Maryland.....	29
Figure 11. Application of the (A) simple imperviousness model, (B) imperviousness distribution model, and (C) scaled imperviousness model to the ratio of urban to rural discharges as a function of return period and imperviousness at U.S. Geological Survey streamgage 01585200, West Branch Herring Run at Idlewylde, Maryland.....	30
Figure 12. Application of the (A) simple density model, (B) density distribution model, and (C) scaled density model to the ratio of urban to rural discharges as a function of return period and imperviousness at U.S. Geological Survey streamgage 01585200, West Branch Herring Run at Idlewylde, Maryland.....	31

Tables

Table 1.	Example of the adjustment of the observed annual maximum time series using the 5-year simple imperviousness adjustment model for U.S. Geological Survey streamgage 01645000, Seneca Creek at Dawson, Maryland.....	10
Table 2.	Number of gages used in the regression analysis as a function of return period.....	18
Table 3.	Goodness-of-fit characteristics for the models evaluated.....	19
Table 4.	Calibrated values of the adjustment model coefficients and exponents.....	20
Table 5.	Rankings of model performance based on the goodness-of-fit values shown in table 3.....	21
Table 6.	Summed rankings of model calibration statistics based on goodness-of-fit values given in table 3.....	21
Table 7.	Smoothed values of the adjustment model coefficients and exponents.....	23
Table 8.	Performance of U.S. Geological Survey urban equations from an earlier study and urban equations developed in this report.....	33
Table 9.	Relative Standard error (S_e/S_y) for three imperviousness models developed in this report as normalized by the relative standard error determined by Sauer and others (1983).....	33
Table 10.	Mean bias, in cubic feet per second, for new urban equations used on data from 10 U.S. Geological Survey streamgages for which localized urban equations are available.....	34
Table 11.	Relative standard error (S_e/S_y) for new urban equations used on data from 10 U.S. Geological Survey streamgages for which localized urban equations are available....	35
Table 12.	Summed and ranked overall performance of new urban equations used on data from 10 U.S. Geological Survey streamgages for which localized urban equations are available.....	35

Conversion Factors

Inch/Pound to SI

Multiply	By	To obtain
Length		
inch (in.)	2.54	centimeter (cm)
inch (in.)	25.4	millimeter (mm)
Area		
square mile (mi ²)	259.0	hectare (ha)
square mile (mi ²)	2.590	square kilometer (km ²)
Flow rate		
cubic foot per second (ft ³ /s)	0.02832	cubic meter per second (m ³ /s)
Hydraulic gradient		
foot per mile (ft/mi)	0.1894	meter per kilometer (m/km)

Temperature in degrees Celsius (°C) may be converted to degrees Fahrenheit (°F) as follows:

$$^{\circ}\text{F}=(1.8\times^{\circ}\text{C})+32$$

Temperature in degrees Fahrenheit (°F) may be converted to degrees Celsius (°C) as follows:

$$^{\circ}\text{C}=(^{\circ}\text{F}-32)/1.8$$

Vertical coordinate information is referenced to the North American Vertical Datum of 1988 (NAVD 88).

Horizontal coordinate information is referenced to the North American Datum of 1983 (NAD 83).

Altitude, as used in this report, refers to distance above the vertical datum.

The following SI unit of measure is given in this report in reference to impervious mapping resolution.

Multiply	By	To obtain
Length		
meter (m)	3.281	foot (ft)

Methods for Adjusting U.S. Geological Survey Rural Regression Peak Discharges in an Urban Setting

By Glenn E. Moglen and Dorianne E. Shivers¹

Abstract

A study was conducted of 78 U.S. Geological Survey gaged streams that have been subjected to varying degrees of urbanization over the last three decades. Flood-frequency analysis coupled with nonlinear regression techniques were used to generate a set of equations for converting peak discharge estimates determined from rural regression equations to a set of peak discharge estimates that represent known urbanization. Specifically, urban regression equations for the 2-, 5-, 10-, 25-, 50-, 100-, and 500-year return periods were calibrated as a function of the corresponding rural peak discharge and the percentage of impervious area in a watershed. The results of this study indicate that two sets of equations, one set based on imperviousness and one set based on population density, performed well. Both sets of equations are dependent on rural peak discharges, a measure of development (average percentage of imperviousness or average population density), and a measure of homogeneity of development within a watershed. Average imperviousness was readily determined by using geographic information system methods and commonly available land-cover data. Similarly, average population density was easily determined from census data. Thus, a key advantage to the equations developed in this study is that they do not require field measurements of watershed characteristics as did the U.S. Geological Survey urban equations developed in an earlier investigation.

During this study, the U.S. Geological Survey PeakFQ program was used as an integral tool in the calibration of all equations. The scarcity of historical land-use data, however, made exclusive use of flow records necessary for the 30-year period from 1970 to 2000. Such relatively short-duration streamflow time series required a nonstandard treatment of the historical data function of the PeakFQ program in comparison to published guidelines. Thus, the approach used during this investigation does not fully comply with the guidelines set forth in U.S. Geological Survey Bulletin 17B, and modifications may be needed before it can be applied in practice.

Introduction

The U.S. Geological Survey (USGS) maintains and publishes a set of rural regression equations for use in estimating peak discharges of varying return periods for each State in the United States (U.S.; Jennings and others, 1994; Ries and Crouse, 2002). Sauer and others (1983) developed a method for transforming these rural discharge estimates to estimates for urban watersheds based on several watershed characteristics, most notably two different indices of urbanization—the basin development factor (BDF) and the percentage of impervious area. A study was conducted to reassess the method of Sauer and others (1983) by taking advantage of 20 additional years of streamflow data, using geographic information system (GIS) data and techniques, and considering new approaches in the development of a set of urban regression equations. As in the earlier study, the goal of this study was to develop flood-frequency equations for urbanized sites by applying urban adjustments to the T -year flood values obtained from the regional equations for rural ungaged sites. The T -year flood is the flood that, on average, is equaled or exceeded once in any given number (T) of years.

The USGS typically conducts flood-frequency analyses consistent with Bulletin 17B (Interagency Advisory Committee on Water Data, 1982). For the study, the USGS flood-frequency analysis program, PeakFQ, was incorporated into an optimization program to calibrate several sets of urban regression equations. Because of limited-duration streamflow time series used in the analysis, however, the approach used in this study for the historical period input to the PeakFQ program does not comply with Bulletin 17B guidelines.

¹ University of Maryland, Department of Civil and Environmental Engineering, College Park, MD 20742.

2 Methods for Adjusting USGS Rural Regression Peak Discharges in an Urban Setting

Previous Investigations

The effects of land-use changes on stream discharge are well documented (for example Carter, 1961; James, 1965; Viessman, 1966; Leopold, 1968; Andersen, 1970). The effects of urbanization, in particular, typically are related to characteristic changes in watersheds. Urbanization generally is associated with increased areas of impervious surfaces, such as pavement and rooftops. These surfaces usually have drainage features that quickly convey water away from structures or road surfaces and into nearby streams. In addition, urbanization generally reduces temporary stormwater storage, such as depression storage and tree-leaf interception. These urbanization processes create large volumes of stormwater that generally travel swiftly to the surface-drainage network and result in high flood peaks, which are the focus of this study.

Sauer and others (1983) presented a regression-equation approach of using rural flood-frequency estimates, such as those resulting from the equations summarized by Jennings and others (1994), and scaling these estimates upwards based on several measures of urbanization. Such rural flood-frequency estimates are currently (2006) calculated using the USGS National Flood-Frequency Program (version 3; Ries and Crouse, 2002).

Although Sauer and others (1983) calibrated several different sets of adjustment equations, their best model included seven parameters as presented in the following form:

$$UQ_T = c_{1,T} A^{c_{2,T}} SL^{c_{3,T}} (RI2 + 3)^{c_{4,T}} (ST + 8)^{c_{5,T}} (13 - BDF)^{c_{6,T}} IA^{c_{7,T}} RQ_T^{c_{8,T}}, \quad (1)$$

where UQ_T is the T -year urban peak discharge, in cubic feet per second;
 A is the drainage area, in square miles;
 SL is the main channel slope, in foot per mile;
 $RI2$ is the 2-hour, 2-year rainfall, in inches;
 ST is the basin storage in percent;
 BDF is the basin development factor (an urbanization index that quantifies channel improvements, channel linings, storm drains/sewers, and curb-and-gutter streets);
 IA is the impervious area, in percent; and
 RQ_T is the T -year rural discharge as predicted from the appropriate USGS rural-regression equation.

(Note: The difficulty in determining the BDF index based solely on remotely sensed data was a motivating factor for conducting this study.)

In some States, the USGS has developed localized urban equations that supersede the use of the national urban equations developed by Sauer and others (1983; shaded areas in fig. 1). The localized urban equations vary in sophistication, but all generally require some measure of urbanization, such as the percentage of impervious or residential area. The general equations for the States of New Jersey (Stankowski, 1974), Pennsylvania (Stuckey and Reed, 2000), and Wisconsin (Conger, 1986) are presented here to illustrate the available range of localized urban equations.

New Jersey:

$$Q_T = c_{1,T} A^{c_{2,T}} SL^{c_{3,T}} ST^{c_{4,T}} IA^{c_{5,T}}, \quad (2)$$

where Q_T is the T -year urban peak discharge, in cubic feet per second;
 A is the drainage area, in square miles;
 SL is the channel slope, in foot per mile;
 ST is the basin storage, such as lakes and swamps, in percent; and
 IA is the impervious area, in percent, based on population density.

Pennsylvania (Region A):

$$Q_T = c_{1,T} A^{c_{2,T}} (1 + 0.01F)^{c_{2,T}} (1 + 0.01U)^{c_{2,T}} (1 + 0.01C)^{c_{3,T}} (1 + 0.01CA)^{c_{4,T}}, \quad (3)$$

where Q_T is the T -year urban peak discharge, in cubic feet per second;
 A is the drainage area, in square miles;
 F is forest cover, in percent;
 U is urban development, in percent;
 C is the watershed, in percent, underlain by carbonate rock; and
 CA is the watershed, in percent, controlled by lakes, swamps, or reservoirs.

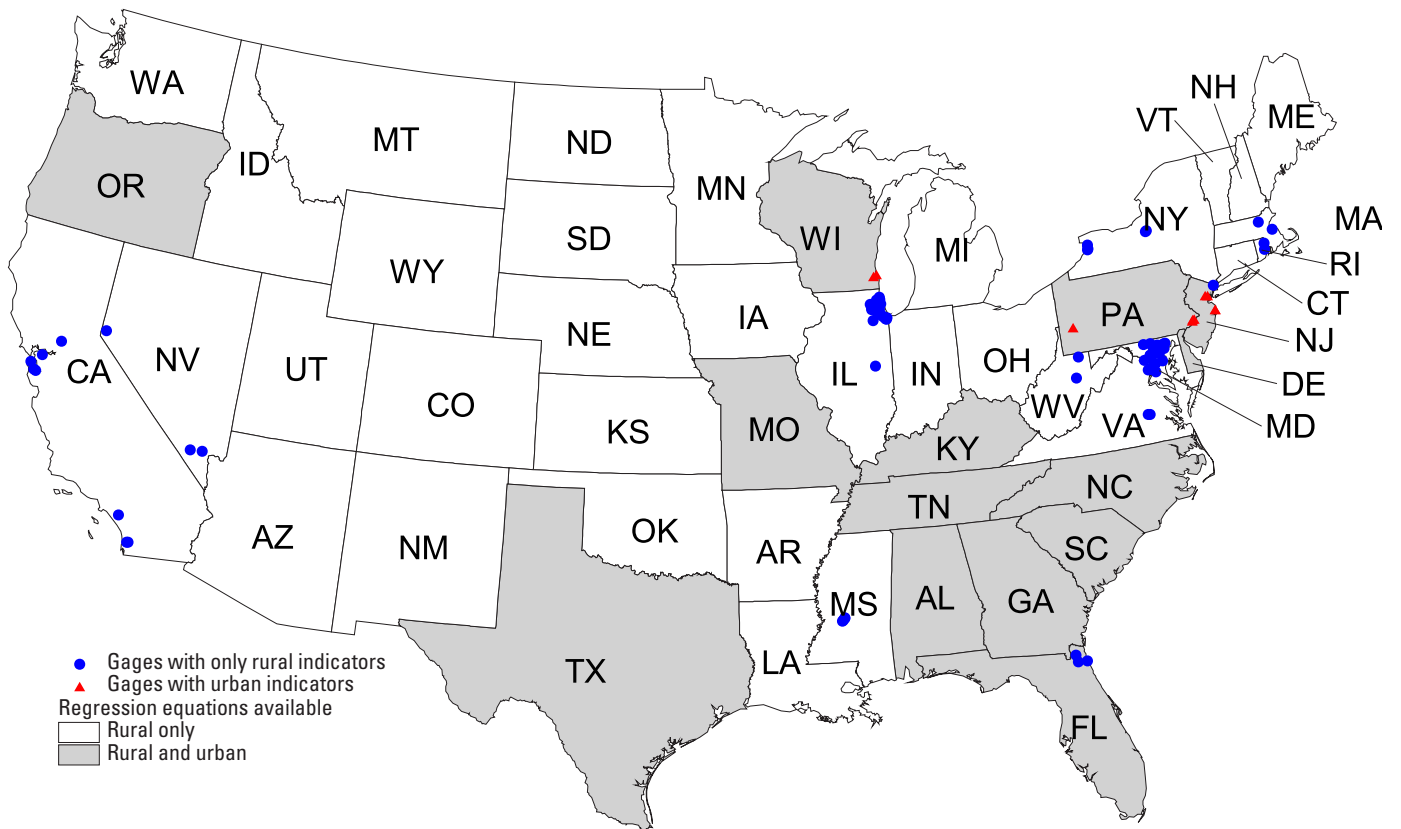


Figure 1. Locations of urban and rural streamgages used in current (2006) study of rural regression peak discharges. [Shaded states have urban prediction regression equations that apply either statewide or in selected areas.]

Wisconsin (Milwaukee County urban areas):

$$Q_T = c_{1,T} A^{c_{2,T}} IA^{c_{3,T}}, \tag{4}$$

where Q_T is the T -year urban peak discharge, in cubic feet per second;
 A is the drainage area, in square miles; and
 IA is the impervious area, in percent.

Overview of Peak Discharge Adjustment Method

The central premise of the study is that observed peak discharges at USGS streamgages are a function of watershed characteristics that exist at the time of the observed flood. Specifically, urbanization, measured as impervious area, varies over time and needs to be properly accounted for when performing flood-frequency analysis on an observed annual maximum time series.

Assumptions of Peak Discharge Adjustment Method

- **Assumption 1:** The USGS rural regression equations for a State or region represent the best estimate of rural discharges for that particular location.
- **Assumption 2:** In a watershed undergoing urbanization, the observed flood peak in any given year (t) depends on the amount of impervious area in the watershed. Because urbanization is an ongoing process, it may be necessary to obtain annual time series of impervious areas in rapidly changing watersheds.
- **Assumption 3:** At streamgages influenced by urbanization, the observed annual maximum time series can be adjusted downward to be consistent with the predicted flood frequency obtained by applying the USGS rural regression equations for the gage location.
- **Assumption 4:** The equations used to adjust the observed annual maximum time series can be inverted for use in scaling a USGS rural flood-frequency regression equation upward to account for urbanization effects. (Note: The objective function of a regression equation can be minimized only in the form in which it was calibrated, and the inversion process may cause a loss in optimality of the objective function.)
- **Assumption 5:** Because calibrated coefficients and exponents vary with each return period, separate optimizations are required for each return period considered. This assumption is consistent with the approach in Sauer and others (1983), which produced different coefficients and exponents for each return period.
- **Assumption 6:** Stormwater-management methods either are not present or do not have a measurable effect on flooding caused by urbanization. Although stormwater-management methods can mitigate increased flooding caused by urbanization, much of the urbanization quantified in this report predates the implementation of stormwater-management methods. Therefore, this report does not attempt to quantify any effects from implementing stormwater-management methods.

Enumeration of Steps in Peak Discharge Adjustment Method

The general steps in the adjustment method steps are described here. Because this is a calibration exercise, the process outlined here is iterative and is dependent on the minimization of an objective function that quantifies the difference between the observed rural flood-frequency values and the flood-frequency results obtained by using the PeakFQ program and adjusted annual maximum discharge data for urban streams.

- Step 1.** Assume (a) a functional form for an adjustment equation and (b) initial trial values for the coefficients and exponents of this adjustment equation.
- Step 2.** Adjust the annual maximum time series for the selected gage (i) as a function of the observed annual maximum time series and the impervious area time series. This produces an adjusted annual maximum time series.
- Step 3.** Write the adjusted annual maximum time series in the USGS National Water Data Storage and Retrieval System (WATSTORE) format.
- Step 4.** Use the PeakFQ program to determine the flood-frequency distribution at the selected gage (i) for the adjusted time series.
- Step 5.** Scan the output of the PeakFQ program for the T -year flood and incrementally adjust the objective function by the squared difference between the PeakFQ T -year flood and the rural regression value for the same flood-frequency. (Note that these differences will be in the logarithmic space to give similar weight to both small and large watersheds.)
- Step 6.** Modify the trial coefficients and exponents assumed in Step 1.

Steps 2 through 6 are repeated for all gages in the database and then performed iteratively until the objective function described in Step 5 is minimized. The entire process is conducted separately for the 2-, 5-, 10-, 25-, 50-, 100-, and 500-year flood-frequencies. A detailed flow chart of the general process described here is shown in figure 2. (Note: Steps 2-6 are automated by a nonlinear optimization program modified from McCuen (1993). The objective function that is minimized is the sum of the squared errors between the PeakFQ output for the T -year flood and the equivalent T -year flood estimated by the rural regression equation.)

After an equation is calibrated for each return period, the calibrated urban equation is determined in the following step (7):

Step 7. Substitute the calibrated coefficients from the urban-to-rural adjustment function (f -inverse) into the original rural-to-urban function (f).

From the user's standpoint, there likely is only an interest in the actual use of the function f from Step 7. The user's experience will be limited to the following two steps:

- 1) Using the National Flood-Frequency (NFF) program (Ries and Crouse, 2002) or direct application of the USGS rural regression equation, estimate the rural flood-frequency for the ungaged location; and
- 2) Use the rural-to-urban function (f), transform rural T -year floods (from the rural regression equations previously determined) into corresponding T -year urban floods.

A critical challenge is the determination of the appropriate annual maximum discharge adjustment equation. The approach used during this study was to explore several functional forms and select the form that produced the best goodness-of-fit statistics across all frequencies.

Analytical Presentation of Adjustment Method

Pursuant to Bulletin 17B (Interagency Advisory Committee on Water Data, 1982), $FF(x, Q_o, HP, SS)$ is defined as the flood-frequency operator. This operator returns the flood-frequency vector, (Q_{ff} for the 2-, 5-, 10-, 25-, 50-, 100-, and 500-year floods) for a time series of annual maximum floods (x), for an identified high-outlier discharge threshold (Q_o), for a historical return period associated with the largest flood exceeding Q_o (HP), and assuming the station skew option (SS). An individual member of the flood-frequency vector, say the T -year flood, is identified as $Q_{ff}(T)$ or Q_T .

The observed annual maximum time series is defined as $Q_{obs}(t)$. The adjusted annual maximum time series is defined as $Q_{adj}(t)$. Performing flood-frequency analyses on these two time series produced the flood-frequency vectors, $Q_{ff,obs}$ and $Q_{ff,adj}$, respectively. Thus, the resulting equations take the following forms:

$$Q_{ff,obs} = FF(Q_{obs}(t), Q_o, HP_{obs}, SS), \quad (5)$$

and

$$Q_{ff,adj} = FF(Q_{adj}(t), Q_o, HP_{adj}, SS), \quad (6)$$

where HP_{obs} and HP_{adj} are the return periods associated with the largest floods in the respective annual maximum time series exceeding Q_o . For simplification, the dependency of the flood-frequency operator on a high-outlier threshold, a specified historical period, and a skew option was omitted from the notation. Thus, equations 5 and 6, respectively, reduce to simpler forms:

$$Q_{ff,obs} = FF[Q_{obs}(t)], \quad (7)$$

and

$$Q_{ff,adj} = FF[Q_{adj}(t)]. \quad (8)$$

Two other flood-frequency vectors were relevant to this study— $Q_{ff,rural}$ and $Q_{ff,urban}$, which are the rural and urban flood-frequency vectors, respectively. Because these vectors are determined from regression equations, they exist only as flood-frequency values and have no time series directly associated with them.

6 Methods for Adjusting USGS Rural Regression Peak Discharges in an Urban Setting

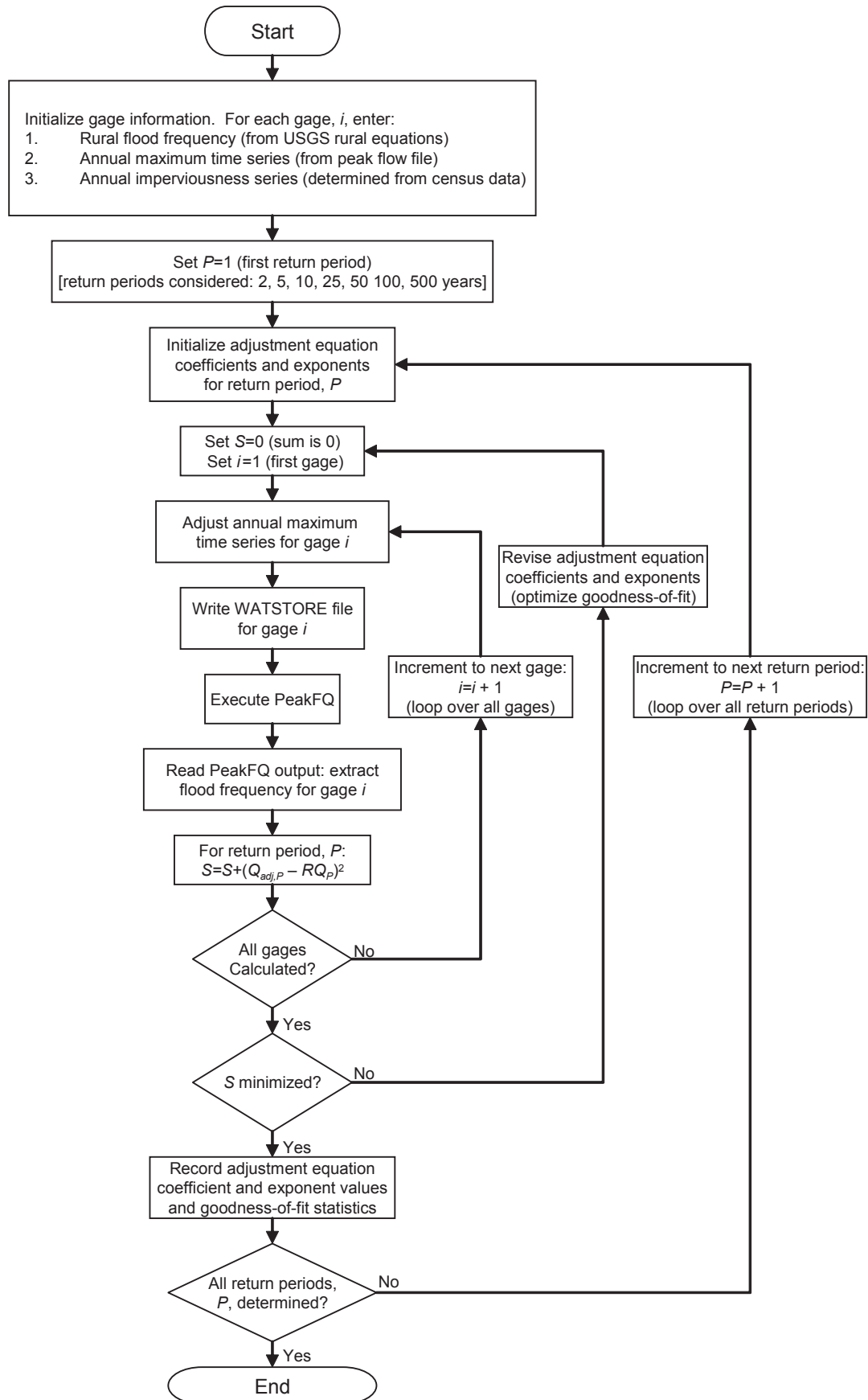


Figure 2. Flow chart for the process for calibrating adjustment equations.

Adjustment of the Observed Annual Maximum Time Series for Nonstationary Urbanization

A central premise in the study by Sauer and others (1983) was that the observed flood-frequency record for a selected streamgauge was stationary; in other words, the record was unaffected by changes in urbanization over the period of record analyzed. To compensate for this premise, Sauer and others (1983) applied the rule that a streamgauge record was usable if the amount of impervious area in the watershed had not increased more than 50 percent during the period of record.

In the current (2006) study, a different approach was taken. As discussed previously, an annual time series of impervious area, $IA(t)$, was developed for each gage studied in order to relate a unique value of impervious area to each observed annual maximum flood value.

The observed and adjusted annual maximum time series are related by a set of nonlinear functions, $f_T(\cdot)$, such that

$$Q_{adj,T}(t) = f_T[Q_{obs}(t), IA(t)], \quad (9)$$

where $Q_{adj,T}(t)$ is the adjusted time series of annual maximum flows that is keyed to the T -year flood. Each of the seven equations for the 2-, 5-, 10-, 25-, 50-, 100-, and 500-year floods ($f_2, f_5, f_{10}, f_{25}, f_{50}, f_{100},$ and f_{500} , respectively) must be calibrated.

When the flood-frequency operator is applied to each adjusted time series, it produces the rural flood-frequency value for the return period, (T) :

$$Q_{ff,rural}(T) = RQ_T = FF\{f_T[Q_{obs}(t), IA(t)]\}. \quad (10)$$

Once all $f_T(\cdot)$ are determined and assumption 4 was applied, these functions were inverted such that the independent variables were the rural flood-frequency values and impervious area, and the dependent variable was the urban flood-frequency value:

$$Q_{ff,urban}(T) = UQ_T = f_T^{-1}[Q_{ff,rural}(T), IA], \quad (11)$$

where $f_T^{-1}(\cdot)$ is the inverted form of $f_T(\cdot)$.

The Adjustment Models

Several forms of $f_T(\cdot)$ were investigated—the null model, the simple imperviousness model, the simple density model, the imperviousness distribution model, the density distribution model, the scaled imperviousness model, and the scaled density model. These model forms are represented by the following equations (12 - 18, respectively).

Null model:

$$Q_{adj,T}(t) = c_{1,T} [Q_{obs}(t)]^{c_{2,T}}. \quad (12)$$

Simple imperviousness model:

$$Q_{adj,T}(t) = \frac{c_{1,T} [Q_{obs}(t)]^{c_{2,T}}}{[IA(t) + 1.0]^{c_{3,T}}}. \quad (13)$$

Simple density model:

$$Q_{adj,T}(t) = \frac{c_{1,T} [Q_{obs}(t)]^{c_{2,T}}}{[PD(t) + 0.001]^{c_{3,T}}}. \quad (14)$$

Imperviousness distribution model:

$$Q_{adj,T}(t) = \frac{c_{1,T} [Q_{obs}(t)]^{c_{2,T}} [\Delta IA(t) + 0.01]^{c_{4,T}}}{[IA(t) + 0.01]^{c_{3,T}}}. \quad (15)$$

Density distribution model:

$$Q_{adj,T}(t) = \frac{c_{1,T} [Q_{obs}(t)]^{c_{2,T}} [\Delta PD(t) + 0.001]^{c_{4,T}}}{[PD(t) + 0.001]^{c_{3,T}}}. \quad (16)$$

8 Methods for Adjusting USGS Rural Regression Peak Discharges in an Urban Setting

Scaled imperviousness model:

$$Q_{adj,T}(t) = \frac{c_{1,T} [Q_{obs}(t)]^{c_{2,T}}}{\left[1 + \frac{99}{1 + e^{c_{4,T} \{I_T^* - IA(t)\}}} \right]^{c_{3,T}}} \quad (17)$$

Scaled density model:

$$Q_{adj,T}(t) = \frac{c_{1,T} [Q_{obs}(t)]^{c_{2,T}}}{\left[1 + \frac{99}{1 + e^{c_{4,T} \{PD_T^* - PD(t)\}}} \right]^{c_{3,T}}} \quad (18)$$

The “distribution” models (eqs. 15 and 16) deserve a brief discussion. The terms $\Delta IA(t)$ and $\Delta PD(t)$ represent the difference between the 10th and 90th percentiles of the impervious area and population density, respectively. These terms quantify the homogeneity or uniformity of development within a watershed. Watersheds with relatively small values of either of these terms are considered to be uniform. Consideration of uniformity of development is helpful in predicting flood behavior.

The “scaled” models (eqs. 17 and 18) also deserve a brief explanation. The term “scaled imperviousness” varies non-linearly from 1 to 100 as IA increases from 0 to 100 percent. For $IA = I^*$, the scaled imperviousness equals 50.5. The scaled imperviousness function for typical parameter values is shown in figure 3. Scaled imperviousness is used because it could be argued that each incremental increase in imperviousness may not have an equivalent effect on flood-frequency. In figure 3, the effect of imperviousness increases most rapidly within the range of 5 to 15 percent imperviousness. This is consistent with studies of the effects of imperviousness on ecological health (for example, Schueler, 1987).

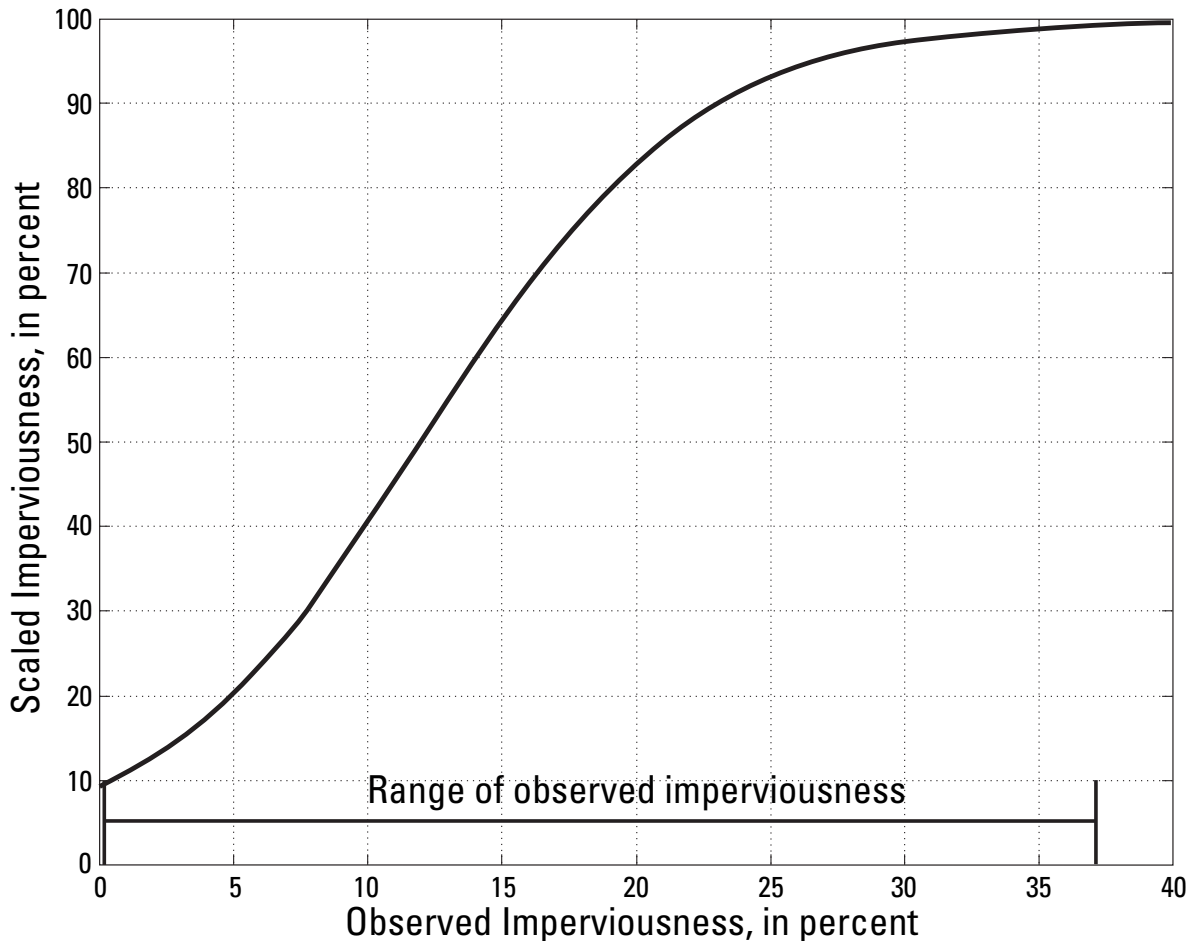


Figure 3. The relation between observed and scaled imperviousness for the scaled values of $I^*=12.1$ and $c_f=0.2$.

Censoring of Streamgages with Non-elevated Flood-Frequency Values

Using PeakFQ, a preliminary flood-frequency analysis was performed for the observed annual maximum time series at each of the 78 streamgages. These analyses were conducted identically to those that are conducted in the optimization model with respect to identifying high outliers and determining the length of the historical record (discussed later). The flood-frequency results from these preliminary analyses were compared with the rural regression flood-frequency values described later in this report. If, at a specific streamgage, the flood-frequency from the observed annual maximum time series produced estimates for the T -year flood such that

$$Q_{ff,obs}(T) < Q_{ff,rural}(T), \quad (19)$$

this gage was eliminated from the database for the development of the T -year regression equation. The rationale for such data elimination is that if a streamgage does not have an increase in flood-frequency values relative to the rural regression equations for peak flows, then it does not make sense to use such a gage to calibrate an urban-scaling factor. The number of gages censored from the analysis will be presented later in the “Results” section.

Example—the Calibrated 5-year Simple Imperviousness Adjustment Model

As an example of the flow-adjustment process, consider the calibrated form for the 5-year, simple imperviousness adjustment model determined in this study:

$$f_5 = 0.331[Q_{obs}(t)]^{1.15}[IA(t) + 1.0]^{-0.173}. \quad (20)$$

Included in table 1 are the imperviousness time series, $IA(t)$, the observed annual maximum time series, $Q_{obs}(t)$, and the 5-year adjusted annual maximum time series, $Q_{adj,5}(t)$ for USGS streamgage 01645000, Seneca Creek at Dawsonville, Maryland. The far right column of the table contains the ratios of the adjusted discharges to the observed discharges. Notice that the ratios are less than 1 for all by the three largest discharges in 1971, 1972, and 1975, indicating that the adjusted annual maximum time series generally is less than the observed annual maximum time series, which would be expected if the adjusted series approximates the comparatively small rural discharges that would have occurred in these years had urbanization not occurred. The adjustment ratios also are clearly smaller for smaller discharges and approach (and slightly exceed) 1 as the discharges become large. This indicates that relatively small discharges that correspond to more frequent 2- and 5-year flows are subject to adjustment from the effects of imperviousness. Such adjustments are consistent with the expectation that urbanization has the greatest effect on flood flows from storms of small magnitude. Further, in general, note that the adjustment ratios decrease with time, indicating the need to adjust the more recent discharges that correspond to greater urbanization in the watershed.

The behavior of the adjustment equation can be observed clearly by comparing years 1973 and 1999 (table 1). In 1973, the peak discharge was 3,020 ft³/s, and imperviousness was about 5.6 percent. The adjusted discharge for 1973 is 2,397 ft³/s. In comparison, the peak discharge in 1999 was 3,060 ft³/s, and imperviousness was 12.8 percent. The adjusted discharge for 1999 is 2,144 ft³/s. Although the observed discharge in 1999 is only slightly greater (40 ft³/s) than the observed discharge in 1973, the adjusted discharge for 1973 is greater by 253 ft³/s than the adjusted discharge for 1999. This clearly indicates the role of the annual imperviousness time series in this study. It is worth noting that imperviousness in the watershed between 1970 and 2000 increased from 4.8 percent to almost 13 percent, which represents more than a doubling of impervious area. For this reason, USGS streamgage 01645000 would have been eliminated from the Sauer and others (1983) study because their threshold of a 50-percent increase in imperviousness.

Finally, it is instructive to look at the 5-year flood-frequency values and compare these to the 5-year rural regression equation value of 5,876 ft³/s determined from Dillow (1996). While the observed 5-year peak discharge is 1.27 times the rural regression estimate, the adjusted 5-year peak is only 1.05 times this estimate. The similarity of the adjusted 5-year peak to the rural regression estimate is indicative of the success of the calibrated adjustment equation in quantifying the effect of imperviousness on the annual maximum discharge.

10 Methods for Adjusting USGS Rural Regression Peak Discharges in an Urban Setting

Table 1. Example of the adjustment of the observed annual maximum time series using the 5-year simple imperviousness adjustment model for U.S. Geological Survey streamgage 01645000, Seneca Creek at Dawson, Maryland.

[%, percent; ft³/s, cubic feet per second]

Year	Imperviousness (%)	Observed Q (ft ³ /s)	Adjusted Q (ft ³ /s)	$\frac{Q_{adj,5}(t)}{Q_{obs}(t)}$
1970	4.84	2,200	1,702	0.774
1971	5.10	25,900	28,787	1.111
1972	5.37	26,100	28,831	1.105
1973	5.63	3,020	2,397	0.794
1974	5.90	3,160	2,508	0.794
1975	6.16	16,000	16,094	1.006
1976	6.42	4,900	4,101	0.837
1977	6.69	3,770	3,016	0.800
1978	6.95	7,850	6,969	0.888
1979	7.22	16,000	15,716	0.982
1980	7.48	10,800	9,946	0.921
1981	7.85	1,340	896	0.668
1982	8.23	3,160	2,385	0.755
1983	8.60	3,260	2,455	0.753
1984	8.98	3,010	2,225	0.739
1985	9.35	3,620	2,734	0.755
1986	9.73	1,070	669	0.625
1987	10.10	4,950	3,871	0.782
1988	10.48	7,410	6,120	0.826
1989	10.85	8,250	6,886	0.835
1990	11.23	2,270	1,553	0.684
1991	11.40	5,120	3,947	0.771
1992	11.58	1,750	1,146	0.655
1993	11.76	3,350	2,412	0.720
1994	11.93	9,160	7,651	0.835
1995	12.11	2,080	1,388	0.667
1996	12.28	11,000	9,399	0.854
1997	12.46	3,880	2,829	0.729
1998	12.64	5,280	4,023	0.762
1999	12.81	3,060	2,144	0.701
2000	12.99	1,910	1,244	0.651
Q_5		7,435	6,157	

Methods

The database used for this investigation contains watershed characteristics for 78 USGS streamgages across the United States (Appendix 1). The two most critical elements in this database are the flood-frequency values obtained directly from the gage record and the flood-frequency values obtained from the current USGS rural discharge equations that apply at each gage site. In this section, the method for determining these two sets of values for the gages used in this investigation is discussed. Other ancillary data, such as outlier values, development of impervious-cover time series from census data, and determination of watershed characteristics also are discussed.

Development of Rural Peak Discharge Estimates

The premise of the current study is that the USGS rural flood-frequency regression equations for each State represent the best estimate of the nonurbanized flood-frequency at any gage analyzed in the study. Thus, it is necessary to apply these equations to each gage included in the study and use the flood-frequency results in the development of the urban adjustment methods.

Delineation of Watershed Boundaries

For each gage used in the study, the digital elevation model (DEM) covering the extent of the gaged watershed was obtained from the National Elevation Dataset (NED; U.S. Geological Survey, 2004a). Flow directions were constrained by “burning in” stream locations from the 1:100,000-scale National Hydrography Dataset (NHD; U.S. Geological Survey and U.S. Environmental Protection Agency, 2004). Standard GIS techniques were used to fill depressions and determine flow directions using a D8 algorithm (Jenson and Domingue, 1988). The watershed boundaries were delineated automatically by indicating the location of the streamgage to the GIS and initiating the delineation algorithm. An automated delineation is considered acceptable and consistent with USGS standards if the delineated drainage area is within 10 percent of the drainage area reported by the USGS (U.S. Geological Survey, 2004b).

Estimation of Watershed Characteristics

Using the watershed boundaries generated as described in the previous section other watershed characteristics were determined as needed from additional interpretation of the DEM (for example, calculating the channel slope). If additional land-use and land-cover descriptors were needed, such as percentage of forest cover, these data were determined by using the most current land-use and land-cover data available. In most cases this meant using the 1992 National Land Cover Dataset (NLCD; Vogelmann and others, 1998a, 1998b; U.S. Geological Survey, 2005), although the 2001 NLCD (Homer and others, 2004; Multi-Resolution Land Characterization Consortium, 2005) was available for some locations during this study.

Development of Flood-frequency Estimates from Gage Record

Flood-frequency estimates were developed from the observed annual maximum time series recorded as peak streamflow by the USGS (U.S. Geological Survey, 2004b). In this section, the methods will be discussed for selecting the gages used in the analysis, obtaining the observed annual maximum time, and using the PeakFQ program (Flynn and others, 2005) to analyze the data and produce the flood-frequency values.

Selection of Gages

The following list of criteria was used for selecting gages for this study:

- The gage must have been active for 20 or more years during 1970-2000, the first full period for which census population data are available in digital format.
- The gage must not be subject to flow regulation.
- The gage must be equipped to report measured peak discharges, not just stage measurements.
- The current peak-flow regression equations for the gage cannot be dependent on a measure of urbanization for flood prediction.

12 Methods for Adjusting USGS Rural Regression Peak Discharges in an Urban Setting

- The drainage area for the gage, as determined by GIS, must be within plus or minus 10 percent of the reported USGS drainage area for the gage, thus providing a high degree of confidence in the application of GIS methods for determining the rural discharge estimates.
- The gage record must not have a large proportion (approximately 20 percent) of streamflow values identified as outliers by the PeakFQ program.
- The gage drainage area should contain more than 5 percent impervious area as quantified by the NLCD 2001 (Multi-Resolution Land Characteristics Consortium, 2005).

Applying these criteria resulted in the final selection of 78 gages in 12 States for the study analyses. The States in which the gages are located are California, Florida, Illinois, Maryland, Massachusetts, Michigan, Mississippi, Nevada, New York, Rhode Island, Virginia, and West Virginia (fig. 1).

Observed Annual Maximum Time Series

The observed annual maximum time series was obtained from the USGS (U.S. Geological Survey, 2004b). Observed peak discharge values for each study gage were obtained for all recorded floods occurring between 1970 and 2000. Data were collected in WATSTORE format, which can be read directly by the PeakFQ program.

Use of the PeakFQ Program

The USGS developed the PeakFQ computer program (Flynn and others, 2005) to conduct flood-frequency analyses based on the guidelines in Bulletin 17B (Interagency Advisory Committee on Water Data, 1982). The PeakFQ program uses the sample-moment method to fit the Pearson Type III frequency distribution to the logarithms of annual flood peaks. The skew that is used may be (a) user-developed generalized skew for a region, (b) the skew map from Bulletin 17B, (c) computed from the data, or (d) weighted between the generalized skew and station skew computed from the data. Adjustments can be made for high and low outliers and historic information. Qualification codes can be used to censor data from the analysis.

Skew

The PeakFQ program offers several options for defining a skew value to use in developing a flood-frequency estimate. In developing rural flood-frequency estimates for a given State, a regionalized skew typically is developed that is weighted with the local skew at the gaging station being analyzed. For urban estimates, this method is problematic because of fewer urban watersheds from which to develop a regionalized value and the varying extent of urbanization. In this study, therefore, each gaging station was considered individually and the “station skew” option was used throughout the analyses presented in this report.

PeakFQ Run-Time Parameters

Because of the relatively short periods of record (31 years or less) for the flood-frequency analysis, there is a greater possibility of time-sampling errors than may be expected when examining longer periods of record. To reduce the potential effects of these errors, a systematic method for applying the high-outlier option of the PeakFQ program was applied. To apply this option, however, two quantities must be defined—the discharge threshold that defines the minimum high-outlier discharge and the return frequency of the largest discharge in the systematic record analyzed. Because of the nature of the model calibration process, the procedures outlined here were performed repeatedly (for each iteration, for each gage, for each return period examined) in calibrating various trial-adjustment models.

Identifying the High-Outlier Threshold

The high-outlier threshold is the largest discharge that is considered within the normal range of discharge for a particular time series being analyzed. Discharges greater than this threshold were not attributed a probability using the Weibull plotting position formula in the analyses. The high-outlier discharge threshold was determined by the following method: given an annual maximum time series of n discharges at a particular gage, it was arbitrarily decided that these n values would represent the flood-frequency characteristics for that location for a period four times as long. Although this is an arbitrary assumption, it adequately produced the desired effect of filtering out large extreme floods that far exceeded the periods of record being analyzed while allowing the consideration of moderate-sized floods. Thus, the approach for identifying outlier discharges for this study is *not* consistent with the approach outlined in Bulletin 17B; however, this outlier-identification approach was a repeatable criterion that worked effectively and could be applied uniformly.

The discharges determined from the rural flood-frequency regression equations were used to determine the high-outlier threshold. The flood-frequency distribution was treated as a piecewise log-normal distribution, and the high-outlier threshold was determined by linear interpolation within the bounds of these values by applying the following equation:

$$\log(Q_o) = \left[\frac{\log(Q_u) - \log(Q_l)}{\log\{z(T_u)\} - \log\{z(T_l)\}} \right] [\log\{z(4n)\} - \log\{z(T_l)\}] + \log(Q_l), \quad (21)$$

where $T_l < 4n \leq T_u$ and $z(T_x)$ is the standard normal deviate corresponding to a flood of return period, T_x . Thus, the high-outlier threshold (Q_o), is determined simply as

$$Q_o = 10^{\log(Q_o)}. \quad (22)$$

Identifying the Historic Period for the Largest Flood in the Time Series

In a similar manner, the largest flood in the systematic record was assigned a value for the historic period that it represents given the rural regression flood-frequency equation estimates that apply to a specific watershed. The largest flood was compared to the rural flood-frequency distribution, and a return period was determined by interpolating from this distribution, thus treating the flood-frequency distribution as a piecewise log-normal distribution. If the largest flood in the systematic record exceeded the 500-year rural flood, a return period of 500 years was arbitrarily assigned to the flood. The interpolation was performed as follows:

$$z(Q_o) = \left[\frac{\log(Q_o) - \log(Q_l)}{\log(Q_u) - \log(Q_l)} \right] [z(Q_u) - z(Q_l)] + z(Q_l), \quad (23)$$

where $Q_l < Q_o \leq Q_u$ and $z(Q_x)$ is the standard normal deviate corresponding to a flood of magnitude, Q_x . With $z(Q_o)$ known, the cumulative probability ($P(Q_o)$) associated with this value was determined from the standard normal distribution. Finally, the return period ($T(Q_o)$) was determined by using

$$T(Q_o) = \frac{1}{1 - P(Q_o)}. \quad (24)$$

Example Application of Outlier and Historical Period Identification Procedures

To illustrate the application of the outlier and historical period identification procedures, the procedures were applied to the record of USGS streamgage 02037800, Falling Creek near Midlothian, Virginia. The period of record for this gage is 1951-2003, with several breaks in the record. Because of limited census data, only the period 1970-2000 was considered. Further, the gage was inactive in 1978 and again from 1994 to 2000; therefore, only the records for the periods 1970-1977 and 1979-1993 were available for this study. The largest flood during these two periods had a magnitude of 5,170 ft³/s and occurred in 1979. The next largest flood had a magnitude of 1,400 ft³/s and occurred in 1985 (fig. 4).

The rural peak discharges determined from Bisese (1995) for the Southern Piedmont region produced peak-discharge estimates for the streamgage at Falling Creek of 380 ft³/s, 647 ft³/s, 862 ft³/s, 1,217 ft³/s, 1,552 ft³/s, 1,923 ft³/s, and 3,054 ft³/s for the 2-, 5-, 10-, 25-, 50-, 100-, and 500-year floods, respectively. To identify the high-outlier threshold (Q_m) and determine the return period corresponding to the largest flood, the following criteria were applied. The length of the time series used in the analysis was $n=23$ years, representing the 1970-1977 and 1979-1993 periods of record. Using $4n=92$ years, the log-normally interpolated 92-year flood from equations 21 and 22 is 1,876 ft³/s. No interpolation is necessary to determine the return period corresponding to the largest flood for either the observed or adjusted (discussed later) annual maximum time series because the observed (5,170 ft³/s) and adjusted (4,220 ft³/s) values exceed the rural 500-year flood of 3,054 ft³/s. A 500-year return period is arbitrarily attached to the largest flood during the periods of record analyzed. Thus, the PeakFQ program was executed for both the observed and adjusted annual maximum time series using station skew, high-outlier threshold (1,876 ft³/s), and historical period (500 years).

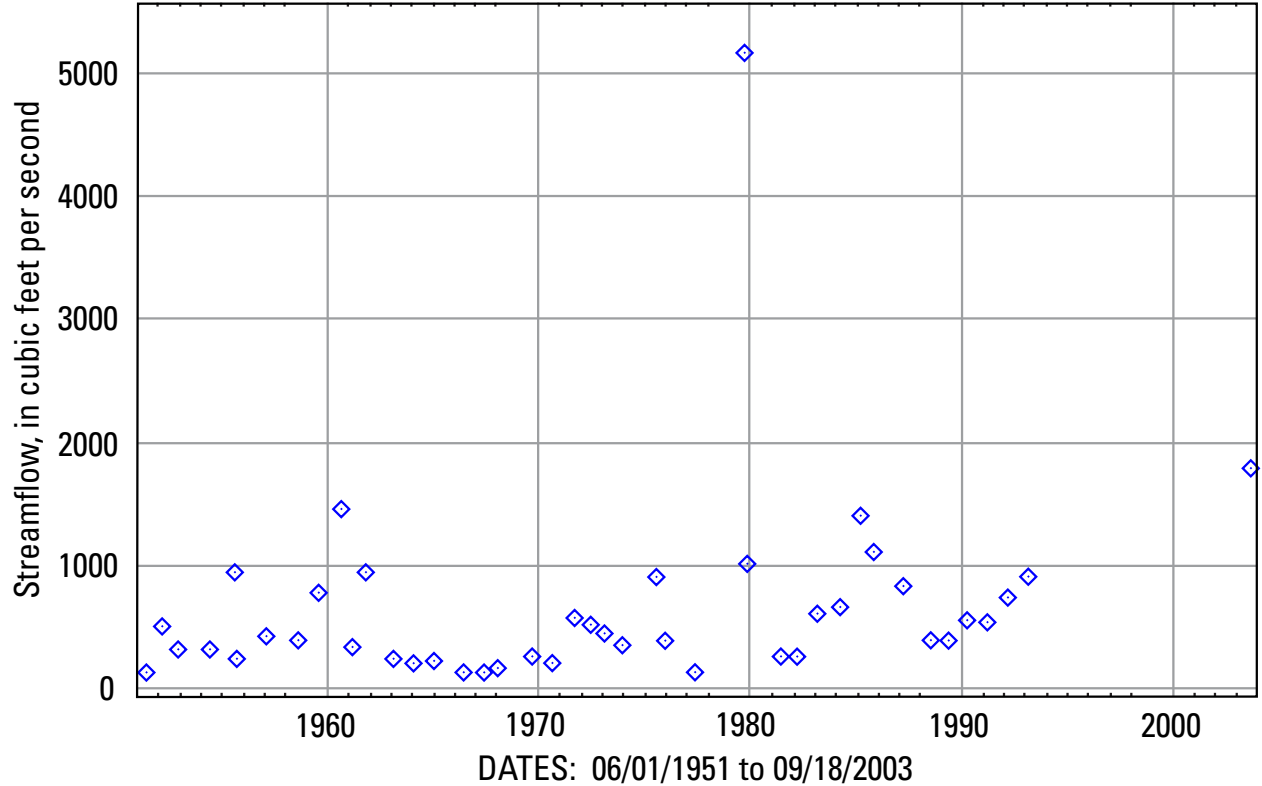


Figure 4. Annual maximum time series for U.S. Geological Survey streamgage 02037800, Falling Creek near Midlothian, Virginia.

Land-Use Change Time Series

This study draws heavily on the idea of developing a time series of changes in the spatial characteristics of the landscape. As a single descriptor of land-use change, imperviousness is the most simple characteristic that can be identified as having a strong influence on flood magnitude. Historical records of imperviousness are scarce, however, in both the spatial and temporal dimensions. For this reason, historical census-derived population-density estimates were used as a surrogate of imperviousness and as a predictor of imperviousness from which to develop the necessary space-time estimates of the changing landscape.

Population Density

Census data are readily available in 10-year increments and digital format for 1970 through 2000. Assuming linear changes in population density between incremental census data, it is possible to develop space-time estimates of population density for any location in the United States at the resolution of a census tract. Census tracts vary considerably in scale depending on population density. For this study, the focus was on urban areas for which the census data tend to delineate small-scale census tracts. For example, Montgomery County, Maryland, is a densely populated county directly north of Washington, D.C. In the 2000 census, this county was composed of 177 census tracts (GeoLytics, 2003), ranging in size from 0.13 mi² to 70 mi² and averaging 2.86 mi². Population density is not uniformly distributed among these tracts, however, and smaller tracts tend to have greater population densities, as shown in figure 5.

Landscape changes, such as increases in imperviousness, introduction of curb and gutter drainage, and channelization can vary with population density. Using a single, average population-density value and the landscape changes for which it is a surrogate to quantify human influences within a watershed may be a poor model when a watershed is large or if human influences vary considerably. To quantify such variability or nonhomogeneity within a watershed, the distribution of population densities was quantified as a fraction of the total watershed area as shown in the example for USGS streamgage 01645000, Seneca Creek at Dawsonville, Maryland (fig. 6). The population densities that correspond to the 10th and 90th percentiles of the watershed area were tagged, and the difference between these densities was recorded. This quantity (ΔPD) was defined as follows:

$$\Delta PD = PD_{10} - PD_{90} \quad (25)$$

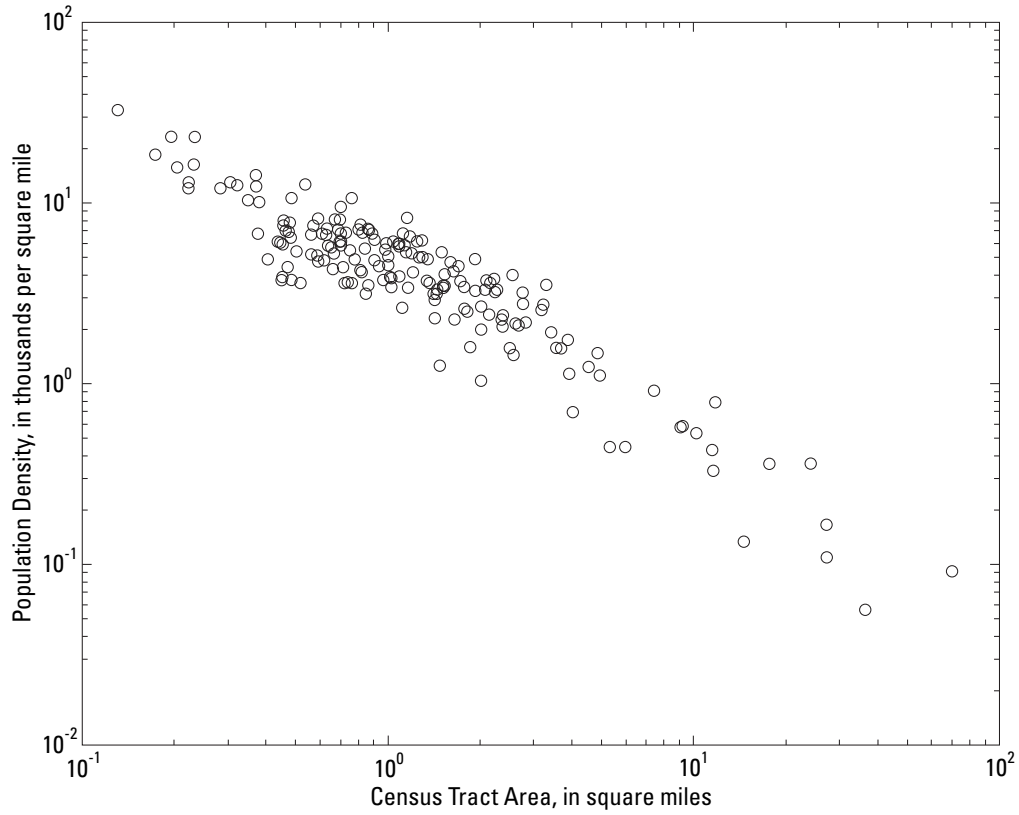


Figure 5. Population density in Montgomery County, Maryland, as a function of census tract area from the 2000 census.

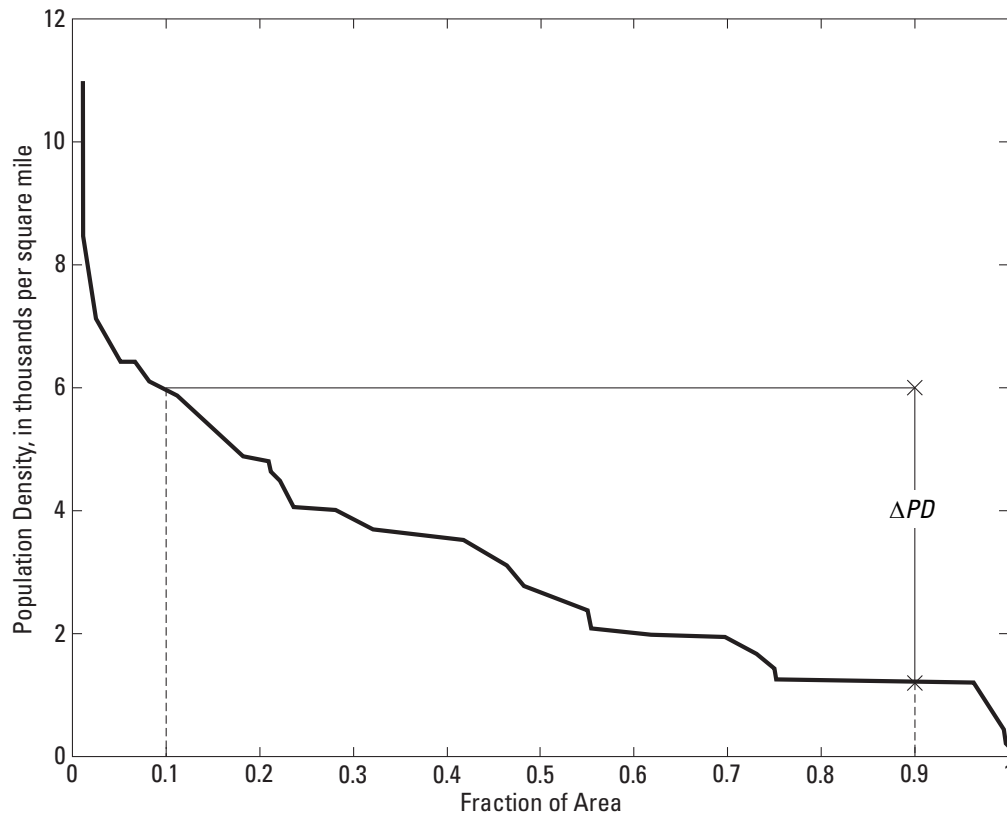


Figure 6. Distribution of population density as a fraction of the total watershed area for U.S. Geological Survey streamgage 01645000, Seneca Creek at Dawsonville, Maryland. [ΔPD is the difference in population density between the 10th and 90th percentiles of the distribution of population density in the watershed.]

The more homogeneous the distribution of population within a watershed, the smaller ΔPD will be. ΔPD is a potentially useful predictor, as discussed in the earlier section on Adjustment Models.

Imperviousness

The 2001 NLCD includes imperviousness mapping product at 30-meter resolution. Each pixel reports a value of imperviousness as a percentage of the total area of the pixel. The determination of the percentage of imperviousness across an entire watershed is straightforward and requires simply calculating the average imperviousness from the set of pixels within the watershed.

Estimating Imperviousness from Census Data

A time series of historical land-use (or imperviousness) data covering several decades and uniformly available at the national scale does not exist. Because of the need to develop estimates of imperviousness for each year for each watershed being studied, it was necessary to develop a procedure that could be uniformly applied to each watershed.

The imperviousness layer from the 2001 NLCD was obtained for the central region of Maryland because it was available and because this region includes two highly urbanized areas—Baltimore, Maryland, and Washington, D.C. Similarly, tract data from the 2000 census were obtained. For each census tract, the population density (in thousands of people per square mile) was determined based on the reported total population and the area of the tract as determined by using GIS. Similarly, the tract outlines were used to sample the 2001 NLCD imperviousness layer and to determine the average imperviousness for each tract. Several different models were investigated for describing the relation between population density and imperviousness. A simple power model was selected to provide the best compromise between statistical power and model rationality. The calibrated equation took the following form:

$$IA = 12.1953(PD)^{0.5195}, \quad (26)$$

where IA is imperviousness, in percent; and
 PD is population density in thousands of people per square mile.

This equation was determined for $0.0002 < PD < 176.4$ and $0.08 < IA < 96.66$. The median PD and IA values were 3.87 and 21.8, respectively. The regression was performed using data from 998 census tracts. The ratio of the standard error (S_e) of equation 26 to the standard deviation (S_d) of observed imperviousness was $S_e/S_d = 0.6163$, with an explained variance of 62.1 percent.

The regression equation developed by Stankowski (1972) to estimate imperviousness from population density in New Jersey had a different form from that of equation 26:

$$IA = 0.117PD^{0.792-0.039 \log PD}, \quad (27)$$

where IA is imperviousness, in percent; and
 PD is population density, in persons (not thousands of persons) per square mile.

The results of both equations, however, are quite comparable, as shown in figure 7.

In focusing on predicted imperviousness between 0 and 50 percent, which is the typical range observed in most watersheds, the maximum departure between the two methods is about 1.2 percent (fig. 7). The estimated imperviousness obtained by using equation 27 is 31.0 percent and 29.8 percent by using equation 26 for a population density of approximately 5,660 persons per square mile. While the difference is small, it is even less significant considering that most of the watersheds in this study had lower population densities and, thus, smaller estimated imperviousness. The typical departure in estimates of imperviousness between the two methods for the watersheds included in this study was less than 1.0 percent.

Implementation of Historical Census Data

Historical census tract data were obtained for 1970, 1980, 1990, and 2000. Population density and estimated imperviousness were determined, using equation 26, for each census tract in each study watershed for each census period. The watershed boundary was then used to select all intersecting census tracts, and an area-weighted average of imperviousness for each census period was determined for each watershed. Imperviousness was assumed to vary linearly between each census measure.

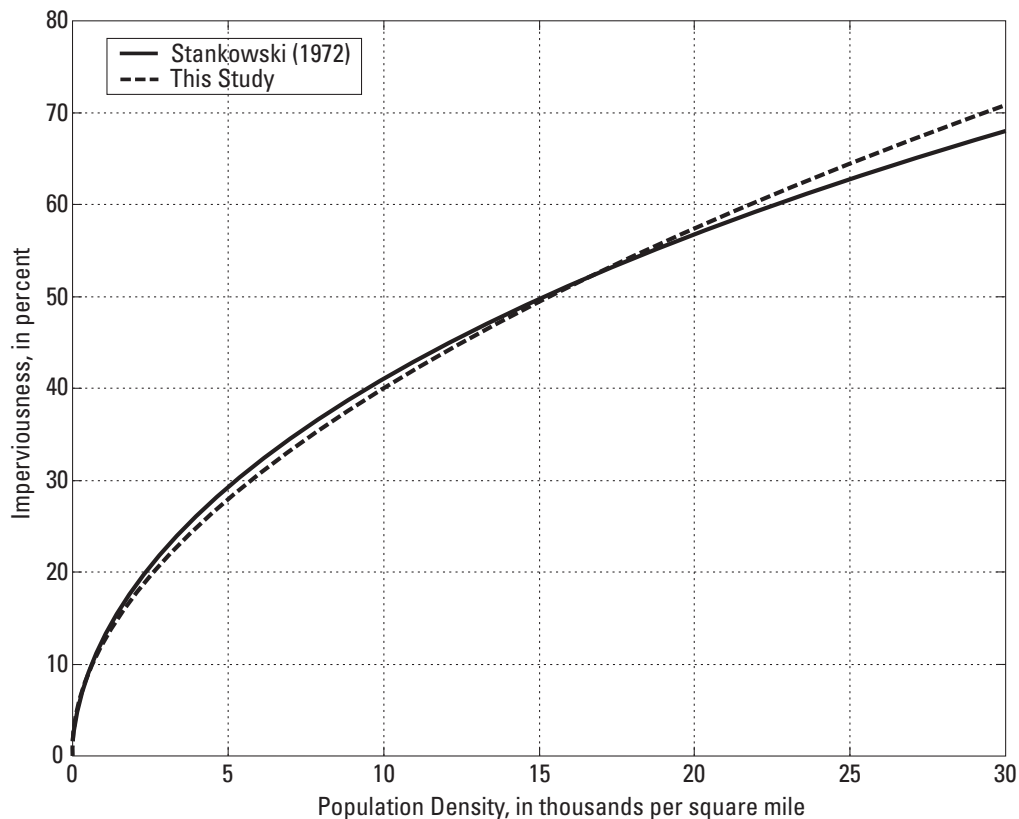


Figure 7. Relations of imperviousness to population density as measured by Stankowski (1972) and as measured in this study. [Note: Stankowski's (1972) equation was based on density in persons per square mile while this study's equation was based on density in thousands of persons per square mile. For consistency in this figure, Stankowski's (1972) equation was revised to be based on thousands of persons per square mile.]

In the event of decreasing population density over time, the estimate of imperviousness naturally would decline (eq. 26). Although population could decrease, imperviousness was assumed to be a strictly nondecreasing function with time and, therefore, was assumed to remain at the previous value. Of the 78 watersheds studied, 10 had decreasing population estimates and nonvarying estimates of imperviousness during the entire gaged period.

To illustrate the use of historical census data in estimating imperviousness, census tract data were obtained for the watershed of USGS streamgage 01645000, Seneca Creek at Dawsonville, Maryland (fig. 8). Equation 26 was applied to population densities for 1970, 1980, 1990, and 2000. The resulting imperviousness estimations for these years were 4.84 percent, 7.48 percent, 11.23 percent, and 12.99 percent, respectively.

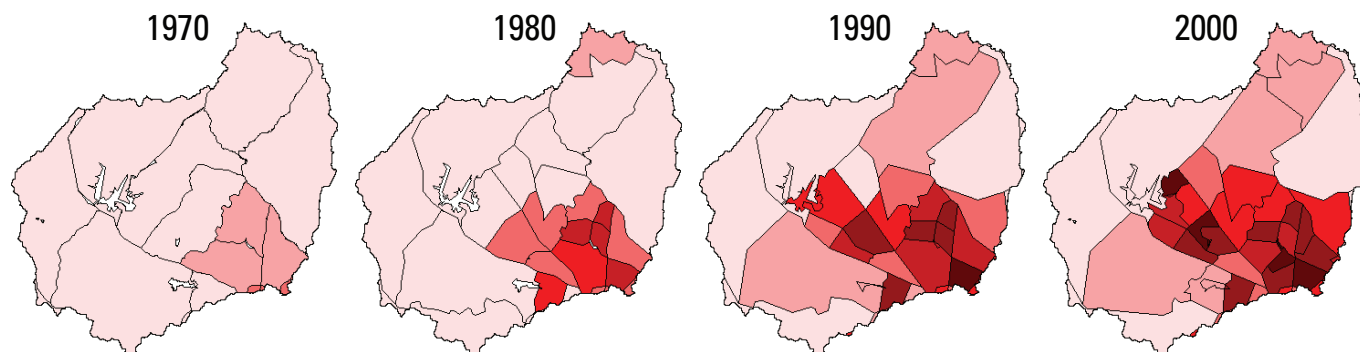


Figure 8. Estimated imperviousness from census tract data for U.S. Geological Survey streamgage 01645000, Seneca Creek at Dawsonville, Maryland. [Imperviousness in this figure ranges from 0 to 40 percent. Darker shades correspond to high degrees of imperviousness.]

Results

Previously, in the “Methods” section the need to censor gages with non-elevated flood-frequency was discussed. The actual numbers of streamgages that were used in the calibration of the adjustment equations as a function of the flood-frequencies being considered are given in table 2.

Table 2. Number of gages used in the regression analysis as a function of return period.

Return period (years)	2	5	10	25	50	100	500
Number of streamgages used	52	52	50	48	43	41	28

In total, information was obtained from 61 of the 78 streamgages (identified by *italics* in appendix 2) for at least one of the seven frequencies for which urban regression equations were calibrated in this study. In other words, of the 78 streamgages originally selected for use in this study, observed flood-frequency estimates for 17 of the streamgages were no greater than the rural regression estimates for any of the seven frequencies examined. The explanation for this is attributed collectively to error inherent in the rural regression equations, possible error in the discharge measurement or corresponding rating curve, error in the flood-frequency analysis, and other inherent characteristics (such as manmade storage or missing basin attributes) that may have served to reduce the flood magnitudes as predicted by the rural regression equations.

The goodness-of-fit characteristics for the model calibrations are presented in table 3, and the calibrated coefficients are shown in table 4. Table 3 presents standard error (S_e) calculated as

$$S_e = \sqrt{\frac{1}{n-m} \sum (y - \hat{y})^2}, \quad (28)$$

where n is the number of observations,
 m is the number of coefficients or exponents being calibrated,
 y is the observed discharge (from the PeakFQ output), and
 \hat{y} is the predicted output calibrated by the nonlinear regression tool.

Standard deviation (S_y) is calculated as

$$S_y = \sqrt{\frac{1}{n-1} \sum (y - \bar{y})^2}, \quad (29)$$

where \bar{y} is the mean of the discharges for the return period (T).

Explained variance (R^2) is calculated as

$$R^2 = \frac{1}{n^2 \cdot S_e^2 \cdot S_x^2} \left[\sum (y - \hat{y}) \cdot (y - \bar{y}) \right]^2, \quad (30)$$

where

$$S_x = \sqrt{\frac{1}{n-1} \sum (\hat{y} - \bar{\hat{y}})^2} \quad (31)$$

in which $\bar{\hat{y}}$ is the mean of the predicted discharges for the return period (T).

As shown in table 3, the null model performed almost as well as the more sophisticated models, especially in regard to the standard error ratio measure. It must be noted, however, that only urban streamgages were used in this study. In particular,

the calibrated equations were developed only for urban gages where the observed flood-frequency was associated with greater discharges than those predicted by the rural regression equations. Given such a dataset from which to calibrate an equation, the null model was certain to perform well simply because the data were already screened for the tendency to produce elevated flood magnitudes. The null model could be expected to have a poorer performance if both rural and urban streamgages were in the dataset. Further, the major shortcoming of the null model is that it will not predict larger urban discharges in relation to larger amounts of urbanization (imperviousness or density). The value of the null model is to provide some information about the additional predictive power of the models that include imperviousness in the flow-adjustment scheme.

Provided that a model that accounts quantitatively for urbanization is needed, the question now turns to which model of those examined performs “best.” An absolute best model is not indicated in table 3 because only seven models were examined, and it is impossible to prove logically that the very best mathematical form was investigated. Nevertheless, the models examined encompass several different approaches to adjusting flows, and their performance probably is not far removed from the hypothetical best adjustment model. Further, the relative performance of the seven models varied, depending on the measure (explained variance or relative standard error) and the flood-frequency being considered (table 3). The goodness-of-fit values in table 3 were ranked within each measure and return period and the best performing model received a ranking of “1” with subsequent rankings increasing as model performance decreased (table 5). The rankings then were summed across both the explanatory measures and flood-frequencies, and the model producing the overall smallest sum was considered to be the best overall model among those considered. These summed rankings are presented in table 6.

This ranking scheme indicates that the best overall performing model was the imperviousness distribution model (table 6). This assertion can be confirmed by a closer examination of table 3 in which the imperviousness distribution model produced the largest explained variance in three of the seven flood-frequencies examined and was among the top three performing models in this measure. With respect to relative standard error, the imperviousness distribution model is the best model in two of the seven flood-frequencies and was always among the top three performing models in this measure. In ranking overall performance, the imperviousness distribution model performed the best (slightly better than the density distribution model) with regard to explained variance and relative standard error (table 5).

In regard to the remaining models, one could ask two questions. First, does population density serve as a stronger predictor than imperviousness? Second, does the “simple” approach, such as simple imperviousness or simple density, outperform the “distribution” or “scaled” approach?

With regard to the first question, the tabulated results (table 5) indicate a modest advantage for imperviousness over density as a predictor. The imperviousness distribution model ranked first overall, whereas the density distribution model ranked second. The simple imperviousness model ranked fourth overall, whereas the simple density model ranked fifth. The oppo-

Table 3. Goodness-of-fit characteristics for the models evaluated.

[All statistics presented are for log-transformed discharges: $\log_{10}(Q)$. Bold values are best values for the return period.]

Model	Explained variance (R^2)							Standard error ratio (S_e/S_y)						
	Return period (years)							Return period (years)						
	2	5	10	25	50	100	500	2	5	10	25	50	100	500
Null	0.821	0.859	0.864	0.902	0.899	0.889	0.814	0.390	0.378	0.341	0.265	0.261	0.277	0.446
Simple imperviousness	0.818	0.866	0.879	0.909	0.902	0.884	0.801	0.392	0.389	0.348	0.281	0.277	0.310	0.480
Simple density	0.818	0.866	0.877	0.907	0.897	0.881	0.804	0.397	0.363	0.328	0.257	0.298	0.311	0.478
Imperviousness distribution	0.832	0.867	0.874	0.909	0.901	0.897	0.855	0.385	0.374	0.338	0.292	0.268	0.271	0.356
Density distribution	0.833	0.870	0.874	0.902	0.893	0.889	0.779	0.383	0.368	0.335	0.278	0.289	0.297	0.526
Scaled imperviousness	0.822	0.873	0.869	0.904	0.900	0.884	0.810	0.409	0.396	0.365	0.288	0.280	0.310	0.528
Scaled density	0.823	0.861	0.873	0.904	0.900	0.887	0.821	0.409	0.414	0.361	0.289	0.287	0.307	0.464

Table 4. Calibrated values of the adjustment model coefficients and exponents.

[Variables for all models are defined in equations 12-18.]

Model	Variable	Return period (years)						
		2	5	10	25	50	100	500
Null	c_1	0.275	0.489	0.305	0.330	0.305	0.330	0.330
	c_2	1.10	1.04	1.10	1.10	1.10	1.10	1.10
Simple imperviousness	c_1	0.298	0.331	0.33	0.33	0.228	0.225	0.263
	c_2	1.16	1.15	1.16	1.16	1.15	1.15	1.12
	c_3	0.177	0.173	0.176	0.176	0.0383	0.0446	0.035
Simple density	c_1	0.27	0.315	0.314	0.368	0.35	0.35	0.308
	c_2	1.10	1.10	1.10	1.10	1.10	1.10	1.10
	c_3	0.05	0.0987	0.0977	0.0642	0.172	0.172	0.0474
Imperviousness distribution	c_1	0.33	0.42	0.405	0.435	0.363	0.293	0.251
	c_2	1.10	1.10	1.10	1.10	1.10	1.10	1.10
	c_3	0.134	0.159	0.135	0.101	0.0819	0.0148	0.0229
	c_4	0.0687	0.025	0.0166	0.0425	0.0388	0.0666	0.108
Density distribution	c_1	0.27	0.313	0.312	0.315	0.304	0.345	0.332
	c_2	1.10	1.10	1.10	1.10	1.10	1.10	1.10
	c_3	0.171	0.228	0.111	0.155	0.140	0.136	0.198
	c_4	0.11	0.062	0.0264	0.0739	0.0963	0.126	0.113
Scaled imperviousness	c_1	0.288	0.315	0.3	0.3	0.3	0.315	0.285
	c_2	1.15	1.15	1.15	1.15	1.15	1.15	1.15
	c_3	0.11	0.148	0.0997	0.1	0.1	0.109	0.0898
	c_4	0.131	0.29	0.2	0.167	0.105	0.149	0.204
	I^*	13.9	20.1	9	12.5	7.54	10.2	13.1
Scaled density	c_1	0.287	0.3	0.3	0.3	0.3	0.296	0.284
	c_2	1.15	1.15	1.15	1.15	1.15	1.15	1.15
	c_3	0.11	0.1	0.125	0.0944	0.108	0.108	0.109
	c_4	0.271	0.505	0.765	0.213	0.000488	0.00469	0.00907
	PD^*	1.05	1.2	3.04	0.960	0.284	0.369	0.519

Table 5. Rankings of model performance based on the goodness-of-fit values shown in table 3.

Model	Explained variance (R^2)							Standard error ratio (S_e/S_y)							
	Return period (years)							Return period (years)							
	2	5	10	25	50	100	500	2	5	10	25	50	100	500	
Null	5	6	7	6	5	3	3	3	4	4	4	4	1	2	2
Simple imperviousness	6	4	1	1	1	5	6	4	5	5	5	3	5	5	
Simple density	6	4	2	3	6	7	5	5	1	1	1	7	7	4	
Imperviousness distribution	2	3	3	1	2	1	1	2	3	3	3	2	1	1	
Density distribution	1	2	3	6	7	2	7	1	2	2	2	6	3	6	
Scaled imperviousness	4	1	6	4	3	5	4	6	6	7	7	4	5	7	
Scaled density	3	7	5	4	3	4	2	6	7	6	6	5	4	3	

Table 6. Summed rankings of model calibration statistics based on goodness-of-fit values given in table 3.

[Bold values are best values for each column.]

Model	Explained variance, R^2 sum	Standard error ratio, S_e/S_y sum	Overall sum (Overall rank)
Null	35	20	55 (3)
Simple imperviousness	24	32	56 (4)
Simple density	33	26	59 (5)
Imperviousness distribution	13	15	28 (1)
Density distribution	28	22	50 (2)
Scaled imperviousness	27	42	69 (7)
Scaled density	28	37	65 (6)

22 Methods for Adjusting USGS Rural Regression Peak Discharges in an Urban Setting

site trend occurred for the scaled models—the imperviousness-based model ranked seventh (last) and the density-based model ranked sixth. Thus, the results slightly favor imperviousness over population density as a predictor of flood-frequency magnitude.

With regard to the second question, the “distribution” approach (eqs. 15, 16) ranked first, followed by the “simple” approach (eqs. 13, 14); the “scaled” models (eqs. 17, 18) performed the worst. This result is interesting because the simple models require the calibration of three parameters; the distribution models require four parameters, and the scaled models require five parameters. Generally, a more complex model would be expected to perform best because it has more flexibility to conform to the observed data. In this case, the moderately complex model performs best. The ΔPD or ΔIA term of the distribution model structure apparently is an effective way of quantifying urban heterogeneity within a watershed and, therefore, is useful in predicting flood behavior.

The calibration results for all models are presented in table 4. (Appendix 3 contains plots for comparing predicted and observed discharges for the calibrated simple imperviousness model, which is fairly representative of other model calibrations as well.) While the values shown in table 4 represent optimum calibration results, a brief examination of this table reveals somewhat erratic trends in many calibration parameters as return period varies. It would be expected that these parameters vary smoothly from one return period to the next. Towards this end, a secondary smoothing was performed for each calibration parameter using a linear regression model:

$$\hat{c}_{x,T} = \alpha \log_{10}(T) + \beta, \quad (32)$$

where T is the return period and α and β are determined by regressing the calibrated values of $c_{x,T}$ on $\log_{10}(T)$.

These smoothed parameter values were used in all subsequent analyses, because it was felt these values were more representative of the true relations between rural and urban flood magnitudes (table 7). Appendix 4 contains the calibrated and smoothed coefficients plotted against return periods for each adjustment model.

As stated previously, the functions, $f_n(\cdot)$ can be rearranged to express the urban peak discharge as a function of the rural peak discharge and imperviousness. For simplicity, the equation 11 can be revised as follows:

$$f_T^{-1} = UQ_T. \quad (33)$$

Thus, the null model is rearranged as

$$UQ_T = \left(\frac{1}{c_{1,T}} \right)^{\frac{1}{c_{2,T}}} RQ_x^{\frac{1}{c_{2,T}}}, \quad (34)$$

the simple imperviousness model is rearranged as

$$UQ_T = \left(\frac{1}{c_{1,T}} \right)^{\frac{1}{c_{2,T}}} RQ_T^{\frac{1}{c_{2,T}}} (IA + 1)^{\frac{c_{3,T}}{c_{2,T}}}, \quad (35)$$

the simple density model is rearranged as

$$UQ_T = \left(\frac{1}{c_{1,T}} \right)^{\frac{1}{c_{2,T}}} RQ_T^{\frac{1}{c_{2,T}}} (PD + 0.001)^{\frac{c_{3,T}}{c_{2,T}}}, \quad (36)$$

the imperviousness distribution model is rearranged as

$$UQ_T = \left(\frac{1}{c_{1,T}} \right)^{\frac{1}{c_{2,T}}} RQ_T^{\frac{1}{c_{2,T}}} (IA + 0.01)^{\frac{c_{3,T}}{c_{2,T}}} (\Delta IA + 0.01)^{\frac{-c_{4,T}}{c_{2,T}}}, \quad (37)$$

the density distribution model is rearranged as

Table 7. Smoothed values of the adjustment model coefficients and exponents.

[Variables for all models are defined in equations 12-18.]

Model	Variable	Return period (years)						
		2	5	10	25	50	100	500
Null	c_1	0.289	0.297	0.303	0.311	0.317	0.324	0.338
	c_2	1.10	1.10	1.10	1.10	1.10	1.10	1.10
Simple imperviousness	c_1	0.327	0.312	0.301	0.287	0.276	0.264	0.239
	c_2	1.16	1.16	1.16	1.15	1.15	1.14	1.13
	c_3	0.200	0.170	0.147	0.117	0.095	0.072	0.019
Simple density	c_1	0.305	0.312	0.318	0.325	0.330	0.336	0.349
	c_2	1.10	1.10	1.10	1.10	1.10	1.10	1.10
	c_3	0.0856	0.0910	0.0950	0.100	0.104	0.108	0.118
Imperviousness distribution	c_1	0.414	0.393	0.378	0.357	0.341	0.325	0.289
	c_2	1.10	1.10	1.10	1.10	1.10	1.10	1.10
	c_3	0.161	0.136	0.118	0.0927	0.0740	0.0552	0.0115
	c_4	0.0269	0.0361	0.0431	0.0523	0.0592	0.0662	0.0823
Density distribution	c_1	0.289	0.297	0.304	0.313	0.320	0.326	0.342
	c_2	1.10	1.10	1.10	1.10	1.10	1.10	1.10
	c_3	0.166	0.165	0.164	0.163	0.162	0.161	0.159
	c_4	0.066	0.073	0.079	0.087	0.093	0.098	0.112
Scaled imperviousness	c_1	0.303	0.302	0.301	0.300	0.300	0.299	0.298
	c_2	1.15	1.15	1.15	1.15	1.15	1.15	1.15
	c_3	0.123	0.117	0.113	0.108	0.104	0.100	0.0911
	c_4	0.189	0.185	0.182	0.178	0.175	0.172	0.165
	I^*	14.4	13.7	13.1	12.3	11.8	11.2	9.88
Scaled density	c_1	0.298	0.297	0.296	0.295	0.295	0.294	0.292
	c_2	1.15	1.15	1.15	1.15	1.15	1.15	1.15
	c_3	0.108	0.108	0.108	0.108	0.108	0.108	0.107
	c_4	0.512	0.418	0.347	0.253	0.182	0.111	-0.0539
	PD^*	1.67	1.45	1.28	1.06	0.893	0.725	0.336

$$UQ_T = \left(\frac{1}{c_{1,T}} \right)^{\frac{1}{c_{2,T}}} RQ_T^{\frac{1}{c_{2,T}}} (PD + 0.001)^{\frac{c_{3,T}}{c_{2,T}}} (\Delta PD + 0.001)^{\frac{-c_{4,T}}{c_{2,T}}}, \quad (38)$$

the scaled imperviousness model is rearranged as

$$UQ_T = \left(\frac{1}{c_{1,T}} \right)^{\frac{1}{c_{1,T}}} RQ_T^{\frac{1}{c_{2,T}}} \left[1 + \frac{99}{1 + e^{c_{4,T}(I_T^* - IA)}} \right]^{\frac{c_{3,T}}{c_{2,T}}}, \quad (39)$$

and the scaled density model is rearranged as

$$UQ_T = \left(\frac{1}{c_{1,T}} \right)^{\frac{1}{c_{1,T}}} RQ_T^{\frac{1}{c_{2,T}}} \left[1 + \frac{99}{1 + e^{c_{4,T}(PD_T^* - PD)}} \right]^{\frac{c_{3,T}}{c_{2,T}}}, \quad (40)$$

where IA is the impervious area, in percent, for the watershed conditions at which the urban discharge estimate is desired.

24 Methods for Adjusting USGS Rural Regression Peak Discharges in an Urban Setting

The coefficients and (or) exponents, $c_{x,T}$, refer to the coefficients outlined in the models presented in equations 12-18 for the T -year prediction equation. Note that equations 12-18 refer to adjustment models that convert urbanized discharges to equivalent rural ones, whereas equations 33-40 convert rural discharges to urban ones.

Based on the *smoothed* coefficients and (or) exponents given in table 7, seven sets of urban equations are given based on the calibrated null, simple imperviousness, simple density, imperviousness distribution, density distribution, scaled imperviousness, and scaled density adjustment models. For the null model, the resulting urban equations are as follows:

$$UQ_2 = 3.091RQ_2^{0.909}, \quad (41)$$

$$UQ_5 = 3.014RQ_5^{0.909}, \quad (42)$$

$$UQ_{10} = 2.959RQ_{10}^{0.909}, \quad (43)$$

$$UQ_{25} = 2.889RQ_{25}^{0.909}, \quad (44)$$

$$UQ_{50} = 2.838RQ_{50}^{0.909}, \quad (45)$$

$$UQ_{100} = 2.790RQ_{100}^{0.909}, \quad (46)$$

and

$$UQ_{500} = 2.683RQ_{500}^{0.909}. \quad (47)$$

The calibrated urban equations for the simple imperviousness model are as follows:

$$UQ_2 = 2.614RQ_2^{0.859} (IA + 1)^{0.172}, \quad (48)$$

$$UQ_5 = 2.866RQ_5^{0.862} (IA + 1)^{0.147}, \quad (49)$$

$$UQ_{10} = 2.827RQ_{10}^{0.866} (IA + 1)^{0.128}, \quad (50)$$

$$UQ_{25} = 2.965RQ_{25}^{0.870} (IA + 1)^{0.102}, \quad (51)$$

$$UQ_{50} = 3.080RQ_{50}^{0.873} (IA + 1)^{0.0825}, \quad (52)$$

$$UQ_{100} = 3.206RQ_{100}^{0.876} (IA + 1)^{0.0628}, \quad (53)$$

and

$$UQ_{500} = 3.541RQ_{500}^{0.883} (IA + 1)^{0.0166}. \quad (54)$$

The calibrated urban equations for the simple density model are as follows:

$$UQ_2 = 2.941RQ_2^{0.909} (PD + 0.001)^{0.0778}, \quad (55)$$

$$UQ_5 = 2.880RQ_5^{0.909} (PD + 0.001)^{0.0827}, \quad (56)$$

$$UQ_{10} = 2.835RQ_{10}^{0.909} (PD + 0.001)^{0.0864}, \quad (57)$$

$$UQ_{25} = 2.778RQ_{25}^{0.909} (PD + 0.001)^{0.0912}, \quad (58)$$

$$UQ_{50} = 2.737RQ_{50}^{0.909} (PD + 0.001)^{0.0948}, \quad (59)$$

$$UQ_{100} = 2.697RQ_{100}^{0.909} (PD + 0.001)^{0.0985}, \quad (60)$$

and

$$UQ_{500} = 2.607RQ_{500}^{0.909} (PD + 0.001)^{0.107}. \quad (61)$$

The calibrated urban equations for the imperviousness distribution model are as follows:

$$UQ_2 = 2.230RQ_2^{0.909} (IA + 0.01)^{0.147} (\Delta IA + 0.01)^{-0.0245}, \quad (62)$$

$$UQ_5 = 2.336RQ_5^{0.909} (IA + 0.01)^{0.124} (\Delta IA + 0.01)^{-0.0328}, \quad (63)$$

$$UQ_{10} = 2.424RQ_{10}^{0.909} (IA + 0.01)^{0.107} (\Delta IA + 0.01)^{-0.0392}, \quad (64)$$

$$UQ_{25} = 2.552RQ_{25}^{0.909} (IA + 0.01)^{0.0843} (\Delta IA + 0.01)^{-0.0475}, \quad (65)$$

$$UQ_{50} = 2.659RQ_{50}^{0.909} (IA + 0.01)^{0.0673} (\Delta IA + 0.01)^{-0.0538}, \quad (66)$$

$$UQ_{100} = 2.775RQ_{100}^{0.909} (IA + 0.01)^{0.0502} (\Delta IA + 0.01)^{-0.0602}, \quad (67)$$

and

$$UQ_{500} = 3.091RQ_{500}^{0.909} (IA + 0.01)^{0.0105} (\Delta IA + 0.01)^{-0.0748}. \quad (68)$$

The calibrated urban equations for the density distribution model are as follows:

$$UQ_2 = 3.095RQ_2^{0.909} (PD + 0.001)^{0.151} (\Delta PD + 0.001)^{-0.0598}, \quad (69)$$

$$UQ_5 = 3.011RQ_5^{0.909} (PD + 0.001)^{0.150} (\Delta PD + 0.001)^{-0.0667}, \quad (70)$$

$$UQ_{10} = 2.951RQ_{10}^{0.909} (PD + 0.001)^{0.149} (\Delta PD + 0.001)^{-0.0720}, \quad (71)$$

$$UQ_{25} = 2.875RQ_{25}^{0.909} (PD + 0.001)^{0.148} (\Delta PD + 0.001)^{-0.0789}, \quad (72)$$

$$UQ_{50} = 2.820RQ_{50}^{0.909} (PD + 0.001)^{0.147} (\Delta PD + 0.001)^{-0.0841}, \quad (73)$$

$$UQ_{100} = 2.767RQ_{100}^{0.909} (PD + 0.001)^{0.146} (\Delta PD + 0.001)^{-0.0894}, \quad (74)$$

and

$$UQ_{500} = 2.653RQ_{500}^{0.909} (PD + 0.001)^{0.145} (\Delta PD + 0.001)^{-0.102}. \quad (75)$$

For the scaled imperviousness model, the calibrated urban equations are as follows:

$$UQ_2 = 2.828RQ_5^{0.870} \left[1 + \frac{99}{1 + e^{0.189(14.4 - IA)}} \right]^{0.107}, \quad (76)$$

$$UQ_5 = 2.834RQ_5^{0.870} \left[1 + \frac{99}{1 + e^{0.185(13.7 - IA)}} \right]^{0.102}, \quad (77)$$

$$UQ_{10} = 2.839 \cdot RQ_{10}^{0.870} \cdot \left[1 + \frac{99}{1 + e^{0.182(13.1 - IA)}} \right]^{0.0985}, \quad (78)$$

$$UQ_{25} = 2.846RQ_{25}^{0.870} \left[1 + \frac{99}{1 + e^{0.178(12.3 - IA)}} \right]^{0.0940}, \quad (79)$$

$$UQ_{50} = 2.851RQ_{50}^{0.870} \left[1 + \frac{99}{1 + e^{0.175(11.8 - IA)}} \right]^{0.0905}, \quad (80)$$

$$UQ_{100} = 2.855RQ_{100}^{0.870} \left[1 + \frac{99}{1 + e^{0.172(11.2 - IA)}} \right]^{0.0871}, \quad (81)$$

and

$$UQ_{500} = 2.866RQ_{500}^{0.870} \left[1 + \frac{99}{1 + e^{0.165(9.88 - IA)}} \right]^{0.0792}. \quad (82)$$

Finally, for the scaled density model, the calibrated urban equations are as follows:

$$UQ_2 = 2.868RQ_2^{0.870} \left[1 + \frac{99}{1 + e^{0.512(1.67 - PD)}} \right]^{0.0942}, \quad (83)$$

$$UQ_5 = 2.876RQ_5^{0.870} \left[1 + \frac{99}{1 + e^{0.418(1.45 - PD)}} \right]^{0.0940}, \quad (84)$$

$$UQ_{10} = 2.881RQ_{10}^{0.870} \left[1 + \frac{99}{1 + e^{0.347(1.28 - PD)}} \right]^{0.0939}, \quad (85)$$

$$UQ_{25} = 2.888RQ_{25}^{0.870} \left[1 + \frac{99}{1 + e^{0.253(1.06 - PD)}} \right]^{0.0937}, \quad (86)$$

$$UQ_{50} = 2.894RQ_{50}^{0.870} \left[1 + \frac{99}{1 + e^{0.182(0.893 - PD)}} \right]^{0.0936}, \quad (87)$$

$$UQ_{100} = 2.900RQ_{100}^{0.870} \left[1 + \frac{99}{1 + e^{0.111(0.725 - PD)}} \right]^{0.0935}, \quad (88)$$

and

$$UQ_{500} = 2.913RQ_{500}^{0.870} \left[1 + \frac{99}{1 + e^{-0.0539(0.336 - PD)}} \right]^{0.0931}. \quad (89)$$

A positive exponent on any predictor variable indicates a positive relation between that predictor and the urban discharge; a negative exponent indicates a negative relation. All of the model forms have positive exponents on all predictors except the ΔIA and ΔPD terms (eqs. 62-75), which have negative exponents. (One other exception, which will be discussed later, occurred for c_4 in equation 89.) It can be concluded quickly that all exponents are rational because the urban flood magnitude is expected

to increase with increasing rural flood magnitude, increasing imperviousness, and increasing population density. Although the negative sign on the ΔIA and ΔPD terms is less obvious, it indicates that as the watershed becomes more varied spatially in urbanization, the predicted discharge decreases. In essence, the ΔIA and ΔPD terms serve to temper the IA or PD terms, which convey only the average imperviousness or population density in a watershed but not variability. Finally, the positive signs on the two parameters that control the scaled sigmoid function presented in figure 8 are worth noting.

Trends in the parameter values as return period increases also are important to examine. Generally, as the return period increases, one might expect to see diminishing effects of urbanization on flood magnitude. For example, the 2-year flood would be expected to have greater dependence on urbanization than the 100-year flood, and trends in the parameter values would be expected to support this. The following discussion focuses on the rationality of the trends for each model:

- **Null model:** The evaluation of trends in this model is not applicable.
- **Simple imperviousness model:** Trends in this model are entirely rational. The model has a negative trend in the IA exponent.
- **Simple density model:** The positive trend in the PD exponent for this model is not rational. This is of concern.
- **Imperviousness distribution model:** Trends in this model are entirely rational. The model has a positive trend in ΔIA exponent and a negative trend in the IA exponent.
- **Density distribution model:** The trends in this model follow the same pattern as in the imperviousness distribution model, although the negative trend in the PD exponent is weak.
- **Scaled imperviousness:** Trends in this model are entirely rational. The model has a negative trend in the scaled IA exponent. It is not clear whether a positive or negative trend in the two scaled imperviousness terms is expected but both terms have negative trends, which is consistent with the trends in the analogous terms in the scaled density model.
- **Scaled density:** This model has a weak positive trend in the scaled PD exponent. This is not rational, but because the trend is so weak, the non-rationality similarly is weak. The negative trend in c_4 actually leads to a slightly negative value for the 500-year flood. The physical interpretation of this result is not clear, and the trends for this model are of concern. Again, the trends in both terms in the scaled density model are negative, consistent with the analogous terms in the scaled imperviousness model.

The results of the rationality analysis are summarized as follows:

1. All models had rational signs in the urbanization (imperviousness (IA) or population density (PD)) exponents.
2. All imperviousness models (simple, distribution, and scaled) had rational trends in the exponents.
3. Two population density models (simple and scaled) had one or more non-rational trends in the exponents. The results for these two models should be used with caution, if at all.

Each of the above equations (eqs. 41-89) was divided by RQ_T to produce the ratio of urban to rural discharge (UQ/RQ). Assuming a vector of rural flood-frequency values ($Q_{ff,rural}$), the UQ/RQ was determined as a function of varying imperviousness. To visualize results of this approach, each model was applied to values in two watersheds, one fairly large and one fairly small. The fairly large watershed (102 mi²) was represented by USGS streamgage 01645000, Seneca Creek at Dawsonville, Maryland; the fairly small watershed (2.2 mi²) was represented by USGS streamgage 01585200, West Branch Herring Run at Idlewylde, Maryland. The UQ/RQ ratios are shown in figures 9-12.

The three imperviousness models that were investigated have several common characteristics (figs. 9, 11). First, all models indicate a general decline in UQ/RQ , approaching 1, as the return period increases. Because the effects of urbanization are most profound for small magnitude and high-frequency flooding events, the greatest adjustment ratios are required for the small return periods. Also, all models had a greater UQ/RQ value as imperviousness increased. This was not surprising because, all things being equal, higher values of imperviousness would be expected to lead to greater urban discharges. In most cases, the null model results were roughly centered within the spread of imperviousness values and were considered representative of average imperviousness across the datasets evaluated (figs. 9, 11).

The three calibrated population density models were applied to discharge data from the two selected gaging stations to obtain the UQ/RQ values across all return periods (figs. 10, 12). The irrationality in the trend in the population density exponent in the simple density model is shown in figure 10A by the increasing spread of the constant population density traces in relation to return period increases. The tight clustering of the constant population density traces in figure 10C indicates that the calibrated parameters for the scaled population density term do not lead to greatly varying estimates of scaled population density. This is not irrational, but it diminishes the credibility of the scaled population density model.

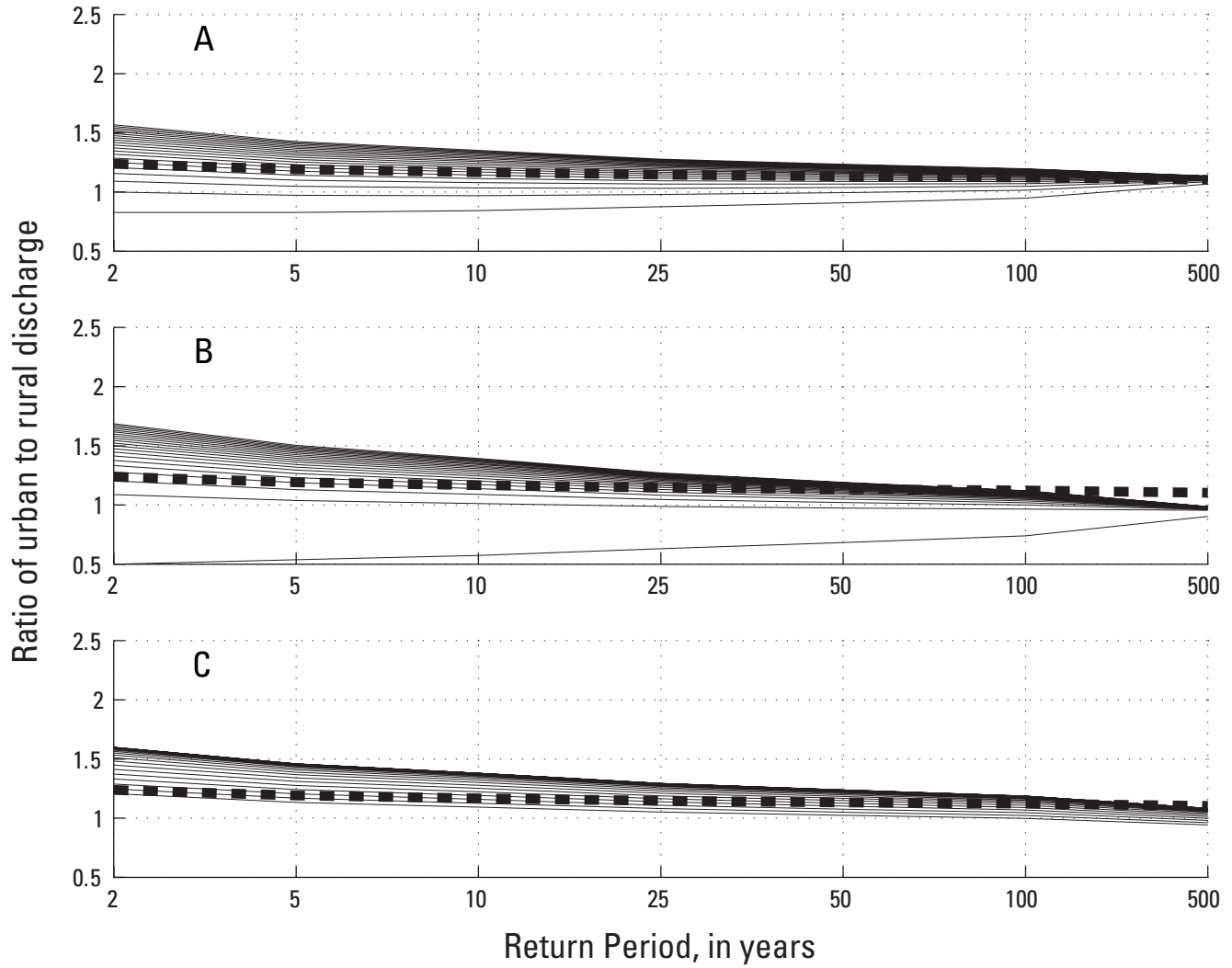


Figure 9. Application of the (A) simple imperviousness model, (B) imperviousness distribution model, and (C) scaled imperviousness model to the ratio of urban to rural discharges as a function of return period and imperviousness at U.S. Geological Survey streamgage 01645000, Seneca Creek at Dawsonville, Maryland. [The heavy dashed line shows the ratio for the “null” model. The light lines indicate imperviousness increasing in 2-percent increments starting from 0 percent for the lowest trace.]

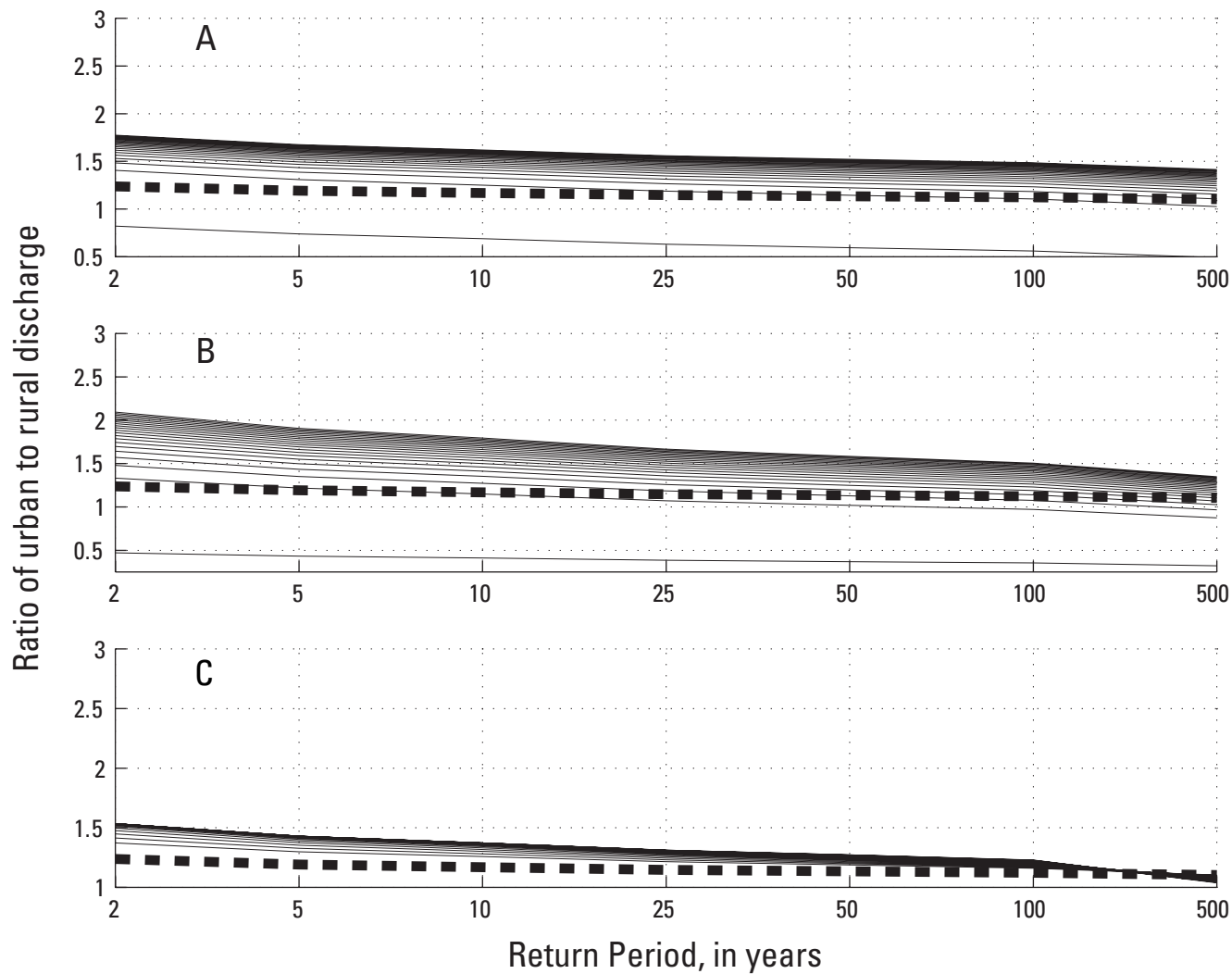


Figure 10. Application of the (A) simple density model, (B) density distribution model, and (C) scaled density model to the ratio of urban to rural discharges as a function of return period and imperviousness at U.S. Geological Survey streamgage 01645000, Seneca Creek at Dawsonville, Maryland. [The heavy dashed line shows the ratio for the "null" model. The light lines indicate population density increasing in increments of 1,000 persons per square mile starting from 0 persons per square mile for the lowest trace.]

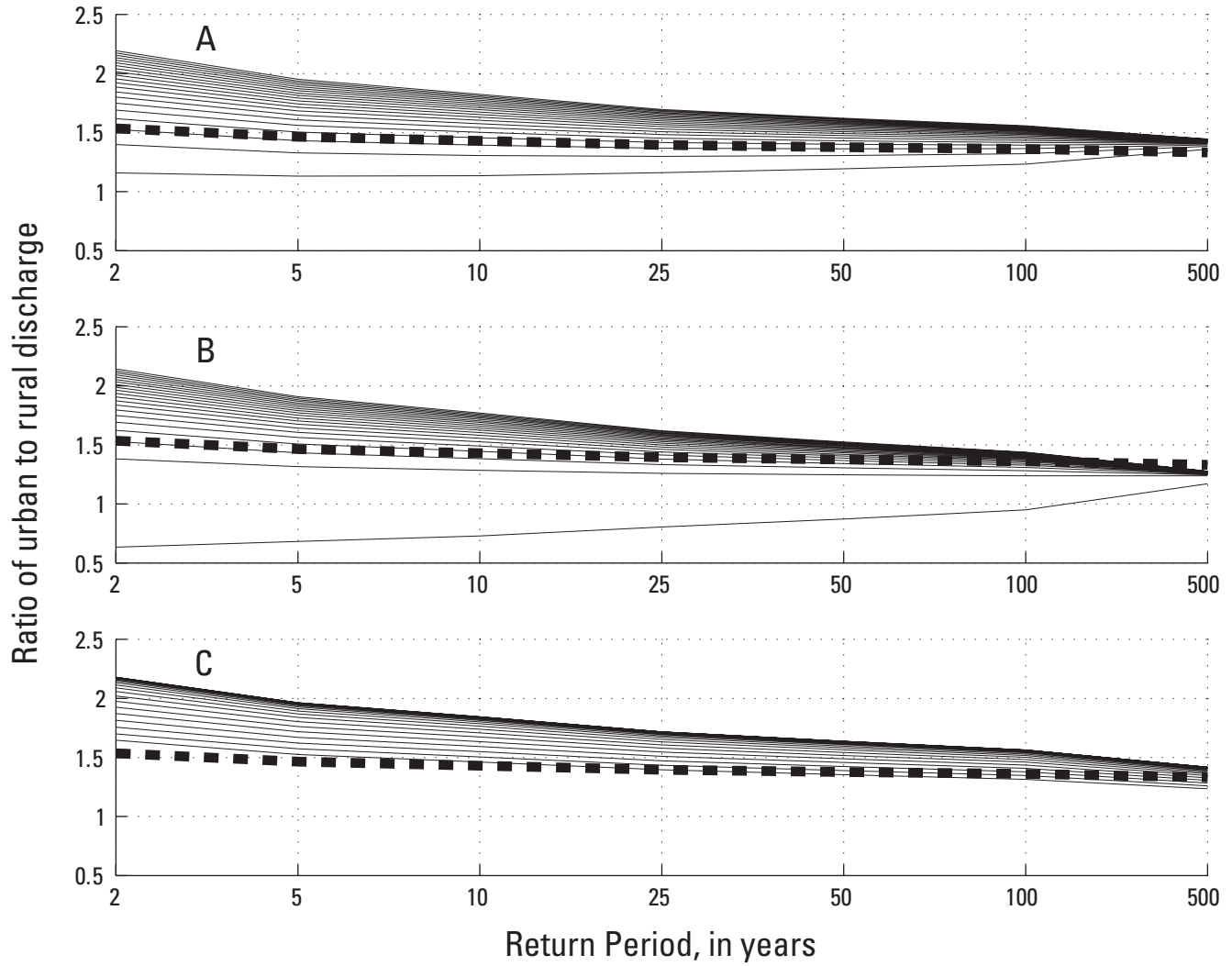


Figure 11. Application of the (A) simple imperviousness model, (B) imperviousness distribution model, and (C) scaled imperviousness model to the ratio of urban to rural discharges as a function of return period and imperviousness at U.S. Geological Survey streamgage 01585200, West Branch Herring Run at Idlewyde, Maryland. [The heavy dashed line shows the ratio for the “null” model. The light lines indicate imperviousness increasing in 2-percent increments starting from 0 percent for the lowest trace.]

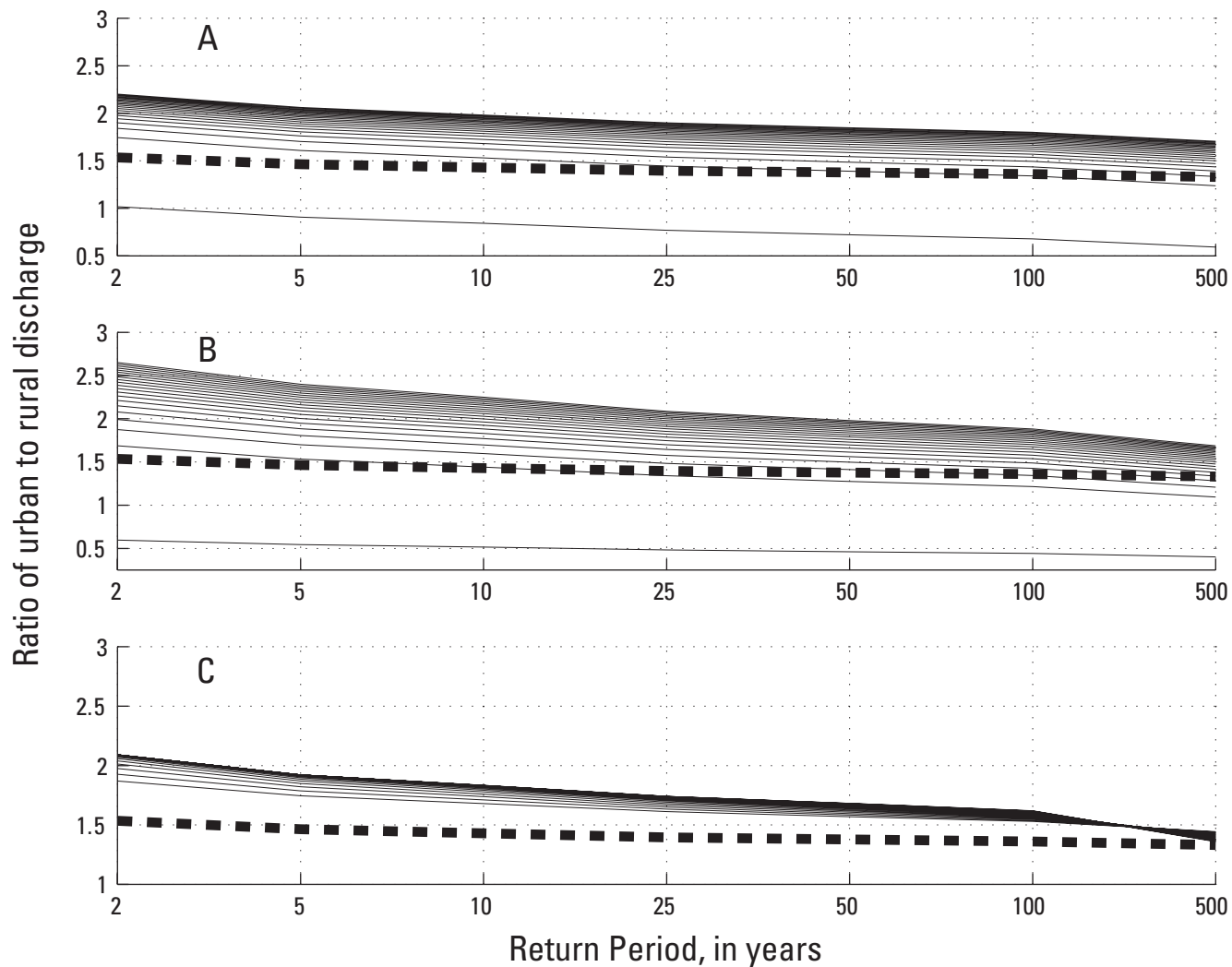


Figure 12. Application of the (A) simple density model, (B) density distribution model, and (C) scaled density model to the ratio of urban to rural discharges as a function of return period and imperviousness at U.S. Geological Survey streamgage 01585200, West Branch Herring Run at Idlwylyde, Maryland. [The heavy dashed line shows the ratio for the “null” model. The light lines indicate population density increasing in increments of 1,000 persons per square mile starting from 0 persons per square mile for the lowest trace.]

By comparing the UQ/RQ ratios for imperviousness (figs. 9, 11) and population density (figs. 10, 12) in the two watersheds, the effect of watershed scale is apparent between the two datasets. Herring Run is roughly two orders of magnitude smaller than Seneca Creek, which results in a greater range in UQ/RQ ratios for this smaller watershed. In comparing figures 9A and 11A for the 2-year return period, a UQ/RQ range is observed of 0.8 to 1.6 for Seneca Creek and approximately 1.2 to 2.2 for Herring Run, respectively. The greater range in UQ/RQ ratios for the smaller watershed is because the exponent on RQ , which is positive but less than 1, has the effect of the UQ/RQ ratio varying more widely for small discharges associated with small watersheds.

Comparison to Urban Equations Developed by Sauer and others (1983)

Application of the urban equations in Sauer and others (1983) to the database assembled for this current study was not possible because all of the equations developed in the earlier work included many watershed characteristics that were not gathered for the current study, most notably the basin development factor (BDF). It is possible, however, to return to the dataset from Sauer and others (1983) and apply three of the equations generated in the current study.

The three equations represent the null model, the simple imperviousness model, and the scaled imperviousness model. The adjustment models that included population density and (or) imperviousness distribution as predictors could not be compared because these data were not measured or presented in the dataset generated by Sauer and others (1983). Another model was added, however; the “Sauer (null)” model is the analog of the null model presented in the current study. The Sauer (null) model was calibrated as a simple power model of the rural discharge,

$$UQ_T = c_{1,T} RQ_T^{c_{2,T}}, \quad (90)$$

using a logarithmic transformation of the rural and urban discharges and calibrating a linear model through least-squares regression. The purpose of the Sauer (null) model was to provide some perspective on the merits of the seven-parameter Sauer and others (1983) model relative to the simplest possible model.

The mean bias and relative standard error were examined for the seven-parameter model (Sauer and others, 1983), the Sauer (null) model, and the three models developed in the current study as applied to the seven return periods (table 8). For the purposes of this report, mean bias was calculated as the average difference between predicted discharges (from the equations developed in this report) and observed urban flood-frequency (reported by Sauer and others, 1983).

In general, the equations from Sauer and others (1983) performed best in predicting urban discharges across the 203 streamgages in the study, especially in terms of standard error (table 8). This would be expected because Sauer and others (1983) used these gages to calibrate their equations. Although the three new models calibrated in the current study did not make use of the database from Sauer and others (1983), the prediction ability of these new models was respectable as evidenced by similar, albeit slightly poorer, bias and standard error values (table 8). This is somewhat surprising in that this comparison is not exactly fair given the possibility for systematic differences in the way imperviousness was measured in the two studies and the fact that this is the calibration set for the Sauer and others (1983) equations. Further, the Sauer (null) model shows comparable relative standard error performance for the 25- through 500-year return periods. The complexity of the seven-parameter model from Sauer and others (1983) seems to yield meaningfully better predictions relative to the other models only for the 2-, 5-, and 10-year return periods. Finally, the performance of the new models developed in this study is even more impressive when considering the use of only two predictors (rural discharge and imperviousness) in these models.

To further quantify the performance of the three models developed in this (current) study, the relative standard errors of each of the new models (null, simple, and scaled) were normalized by the relative standard error from the seven-parameter Sauer and others (1983) model for the same return period. Each of the three new models performed similarly across all return periods (table 9). On average, the standard error of the three new models was about 6 to 9 percent greater than the equivalent standard error from Sauer and others (1983). Standard error is not uniform across return periods, however, but tends to decrease as the return period increases.

This finding mirrors the performance of the Sauer (null) model, which produced poorer relative standard errors for the 2-, 5-, and 10-year return periods and then was essentially equivalent in relative standard error to the seven-parameter model (Sauer and others, 1983) for all larger flood events.

Table 8. Performance of U.S. Geological Survey urban equations from an earlier study and urban equations developed in this report.

[Bold values indicate overall best model performance. Italic values indicate best model performance developed in this report. ft³/s, cubic feet per second]

Return period (years)	Mean bias (ft ³ /s)					Relative standard error: S_e/S_y				
	Sauer (null)	Sauer*	Null	Simple	Scaled	Sauer (null)	Sauer*	Null	Simple	Scaled
2	-127	-111	-33	-111	32	0.462	0.381	0.486	<i>0.461</i>	<i>0.461</i>
5	-171	-178	-80	-214	6	0.463	0.425	<i>0.466</i>	0.477	0.467
10	-235	-141	-174	-285	-65	0.474	0.449	<i>0.474</i>	0.496	0.481
25	-357	-289	-388	-455	-241	0.493	0.490	<i>0.496</i>	0.523	0.507
50	-537	-281	-553	-644	<i>-377</i>	0.520	0.518	<i>0.521</i>	0.550	0.534
100	-675	-300	-867	-734	<i>-658</i>	0.532	0.533	<i>0.540</i>	0.573	0.555
500	-1,200	-1,150	-1,748	-1,106	-1,458	0.600	0.600	0.600	<i>0.586</i>	0.612

*Sauer and others, 1983.

Table 9. Relative Standard error (S_e/S_y) for three imperviousness models developed in this report as normalized by the relative standard error determined by Sauer and others (1983).

Return period (years)	Null	Simple	Scaled	Mean
2	1.244	1.218	1.205	1.223
5	1.094	1.146	1.106	1.115
10	1.047	1.120	1.078	1.082
25	1.004	1.073	1.040	1.039
50	1.002	1.052	1.034	1.029
100	1.016	1.048	1.048	1.037
500	1.000	0.976	1.020	0.999
Mean	1.058	1.090	1.076	1.075

Application of Calibrated Models to Streamgages with Local Urban Equations

A total of 10 USGS streamgages were identified in New Jersey (Stankowski, 1974), Pennsylvania (Stuckey and Reed, 2000), and Wisconsin (Walker and Krug, 2003) that were subject to substantial urbanization. A set of USGS peak flow regression equations that include an imperviousness or urbanization descriptor is available for each of these States. For each of the 10 urban streamgages, watershed characteristics were determined and regression equations were applied twice—first, to obtain the best estimate of the current urban flow, and second, to set the urban descriptor (either imperviousness or urban area) at zero to develop an equivalent rural flow. The watershed characteristics for the 10 streamgages are provided in appendix 5, and the rural and urban flood-frequency results are provided in appendix 6.

The following example is given for USGS streamgage 01393450, Elizabeth at Ursino Lake at Elizabeth, New Jersey. The 2-year USGS regression equation for New Jersey is as follows:

$$Q_2 = 25.6A^{0.89}S^{0.25}(ST + 1)^{-0.56}(I + 1)^{0.25}, \quad (91)$$

where the constants of 1 are added by default on the chance that either of these quantities would be zero for a particular watershed being studied. For USGS streamgage 01393450, the drainage area is 17.6 mi², the channel slope is 22.4 feet per mile, the storage is 1.6 percent, and the imperviousness in 2000 was 41.9 percent. The result of applying equation 91 to these values is an urban 2-year discharge of 1,071 ft³/s. If imperviousness is set at zero (actually 1, as applied in the National Flood-Frequency program), the equivalent rural discharge is 550 ft³/s.

To test the quality of the urban equation models calibrated in this study, the urban equations were applied to the equivalent rural discharge, and the result was compared with the urban value of 1,071 ft³/s. Continuing this example, the simple imperviousness model was applied as follows:

$$UQ_2 = 2.614(550)^{0.859}(41.9 + 1)^{0.172} = 1,127 \text{ ft}^3/\text{s}. \quad (92)$$

The result of 1,127 ft³/s was compared with the estimated 1,071 ft³/s from equation 91. Similar comparisons were made for the other nine streamgages, and the mean bias results are presented in table 10 and the relative standard error results are presented in table 11.

The performance of the models was ranked in the same way as the model calibrations. The sums of ranks for both the mean bias and relative standard error for the performance of the models for the 10 streamgages with local urban equations are given in table 12. Here, mean bias is defined as the average difference between the predicted discharges determined by applying the equations developed in this study and the discharges determined by applying the appropriate (published) State urban regression equations. The strongest performing models were models based on population density; however, all models outperform the null model, which indicates the genuine predictive capabilities that these models offer beyond a simplistic scaling up of a rural discharge.

Table 10. Mean bias, in cubic feet per second, for new urban equations used on data from 10 U.S. Geological Survey streamgages for which localized urban equations are available.

[Bold values indicate overall best model performance.]

Return period (years)	Null	Simple		Distribution		Scaled	
		Impervious- ness	Density	Impervious- ness	Density	Impervious- ness	Density
2	-481	-428	-391	-390	-331	-403	-359
5	-587	-431	-417	-506	-392	-491	-383
10	-592	-503	-345	-541	-380	-473	-305
25	-557	-418	-191	-575	-349	-404	-135
50	-495	-288	-10	-584	-305	-317	64
100	-385	-77	248	-569	-231	-180	341
500	62	803	1165	-406	55	336	-1,708

Table 11. Relative standard error (S_e/S_y) for new urban equations used on data from 10 U.S. Geological Survey streamgages for which localized urban equations are available.

[Bold values indicate overall best model performance.]

Return period (years)	Null	Simple		Distribution		Scaled	
		Imperviousness	Density	Imperviousness	Density	Imperviousness	Density
2	0.486	0.454	0.409	0.417	0.378	0.500	0.428
5	0.361	0.284	0.252	0.311	0.249	0.376	0.280
10	0.298	0.262	0.178	0.271	0.200	0.311	0.202
25	0.234	0.192	0.139	0.238	0.166	0.243	0.138
50	0.201	0.160	0.174	0.221	0.162	0.209	0.148
100	0.177	0.156	0.236	0.202	0.162	0.184	0.196
500	0.218	0.307	0.439	0.184	0.209	0.222	0.361

Table 12. Summed and ranked overall performance of new urban equations used on data from 10 U.S. Geological Survey streamgages for which localized urban equations are available.

[Smaller sums and lower ranks indicate best performance.]

Performance measure	Null	Simple		Distribution		Scaled	
		Imperviousness	Density	Imperviousness	Density	Imperviousness	Density
Sum of mean bias ranks	41	29	22	40	17	28	19
Sum of relative standard error ranks	34	25	25	33	14	42	23
Overall Sum	75	54	47	73	31	70	42
Overall Rank	7	4	3	6	1	5	2

Interpretation of Results

The equations developed in this study appear to be comparable in predictive capability to those developed by Sauer and others (1983). Three models developed during this current study—the null, simple imperviousness, and scaled imperviousness models—were compared with the best model from the Sauer and others (1983) study. The simple and scaled imperviousness models performed, on average, within about 6 to 9 percent of the results obtained by Sauer and others (1983). The model performance for the more frequent return periods, however, was not as strong. The relative standard error was about 20 to 24 percent greater for the 2-year return period, 10 to 15 percent greater for the 5-year return period, and 5 to 12 percent greater for the 10-year return period than the results obtained by Sauer and others (1983). These results are considered comparable because the statistics from Sauer and others (1983) were based on performance using the same dataset from which the equations were calibrated, and the new models were calibrated using a completely different dataset.

Model Strengths

The models developed in this study have two key strengths

- **Simplicity:** The equations require three predictors or less. The rural discharge is a predictor in all models. Imperviousness or population density is used as a predictor in all but the null model. The quantities ΔIA and ΔPD are used as third predictors in the distribution models to quantify the heterogeneity of urbanization within a watershed. Although ΔIA and ΔPD qualify as third predictors, they are based directly on the same quantities as either imperviousness or population

density. A user of any of these models needs access only to the rural discharges and a representation of either imperviousness or population density.

- **Ease of Use/GIS friendly:** These models are applied easily by using available data from remote locations. The water, environmental, and ecological communities increasingly recognize the value of imperviousness as a strong predictor of human effects; simultaneously, the availability and quality of GIS coverages of imperviousness are increasing (Homer and others, 2004). Similarly, GIS coverages of census data are available commonly and render the determination of population density and ΔPD to an easily automated task. In contrast, the equations from Sauer and others (1983) require the determination of seven predictors, of which at least one (BDF) requires a physical reconnaissance of the watershed for accurate determination.

Model Weaknesses

The principal weakness of the models developed in this study is the relatively mild dependency of flood magnitude on imperviousness or population density. This mild dependency is underscored by the good performance of the calibrated null model and on the non-smoothness of the calibrated adjustment model parameters shown in table 6. If this method were expanded to more gages or repeated with longer time series as additional data become available, it may be the case that the functional dependency on urbanization measures would become more clear and the need for calibrated model parameter smoothing (shown in table 7) would diminish.

Another weakness of the models, which is apparent from the comparison to data from Sauer and others (1983; see table 8) and to the locally available urban equations (table 10), is a clearly negative bias in the errors of all models for the small return periods. A shift to a positive bias occurred for the largest return periods (table 10). The origin of this bias is not known with certainty, although it may be a result of the model calibration method which depends on a logarithmic transformation of the observed rural discharges. Because calibration takes place in the logarithmic space, positive and negative errors measured in the arithmetic space are unequal, and a negative bias can result.

Choosing a Model

Ultimately, the questions one faces are these: “Which is the best model form to use?” and “Are there models that are clearly superior or inferior to the others?”

- **Population density approach:** Among the models based on population density, the density distribution model (eqs. 69-75) was the best performer during calibration and testing with the local urban equations. Further, the other two sets of density-based equations had irrational exponent values or trends, which cast doubt on their application.
- **Imperviousness approach:** The choice of the best imperviousness model is less certain. The imperviousness distribution model (eqs. 62-68) performed the best overall during model calibration, but the performance of this model in testing the local urban equations was not as strong as for the simple imperviousness model (eqs. 48-54). This is perhaps attributable to the small sample size (10 streamgages) used for testing. Because of the clearly strong performance during model calibration and consistency with the population density approach, the imperviousness distribution models appear to be good choices if imperviousness data are used.

Summary and Conclusions

In this study, a new approach was developed, using data from 78 urbanized streamgages across the United States, for calibrating a set of USGS urban regression equations. The rural flood-frequency was determined for each streamgage by using the best available GIS data. The annual maximum time series for the 31-year period, 1970–2000, formed the observed peak discharge portion of the dataset. GIS techniques were used to develop an annual time series of imperviousness over this period using census-derived population density as a surrogate and predictor of imperviousness. The PeakFQ program¹ was embedded in a modified nonlinear optimization program, and several adjustment equations were investigated for calibrating adjustment equations that could scale-back the observed annual maximum time series to an equivalent rural annual maximum time series closely representative of the rural regression equation flood frequency. Seven adjustment models of varying degrees of complexity were calibrated. These models then were compared with earlier USGS urban equations and with urban flood-frequency equations for individual States.

The results indicate that the urban equations developed in this study are comparable in performance to the earlier USGS urban equations but are simpler in that they are dependent only on rural discharge and measures of either imperviousness or population density. Further, the new urban equations can be applied readily by using only remotely sensed data rather than requiring data that can only be measured during site visits to the watershed of interest.

The highest performing models emerging from this study are the imperviousness distribution models (eqs. 62-68) and the population density distribution models (eqs. 69-75). These models depend on three predictors each—rural discharge, imperviousness or population density, and delta imperviousness or delta population density. The imperviousness or population density predictor serves to scale up the rural discharge, and the delta imperviousness or delta population density predictor scales down the discharge. This delta predictor quantifies the homogeneity of the development in a watershed and is calculated readily by GIS techniques using the same spatial data used in determining the imperviousness or population density predictor.

Several directions are worth pursuing in future efforts to develop urban equations.

- **Expand the model calibration database:** The data collection for 78 gages used in this study was labor intensive but crucial to the development of nationally applicable regression equations. In contrast, the database for the Sauer and others (1983) urban equation study was slightly in excess of 200 streamgages. Although expanding the current database may be impractical (or impossible), more streamgages could be identified that have been active in recent (1970 to present) times in States for which no local urban equations have been developed. Alternatively, this dataset likely will improve with time as periods of record for urbanization data increase.
- **Expand the test application database:** Data for only 10 streamgages were used to compare locally available urban equations to predictions from the urban equations developed in this study. The results of this application are, therefore, only as sound as the representativeness of these 10 streamgages and the respective local urban equations. Local urban equations have been developed for approximately 13 States. Using data from streamgages in all of these States may lead to more definitive or possibly different results from those documented here.
- **Investigate additional trial adjustment models:** As previously stated, it is a logical impossibility to consider all possible models for relating urban floods to rural floods. In this study, a creative attempt was made in this regard. Other models, however, could be investigated and may prove more effective.
- **Recalibrate model in compliance with Bulletin 17B:** It is advisable that additional investigative work, using the embedded PeakFQ-optimization approach developed in this study, be done in compliance with the guidelines set forth in Bulletin 17B (Interagency Advisory Committee on Water Data, 1982).
- **Investigate different optimization and calibration algorithms:** The tendency for calibrated models (developed in this study and in the Sauer and others (1983) urban equation study) to have a negative bias for small return periods and a neutral or positive bias for large return periods indicates a weakness in the current log-transformation method. Using a different model calibration approach that avoids this log-transformation method may produce unbiased calibrations and ultimately smaller standard errors.

¹ In this study, a nonstandard treatment of the historical period was used in the PeakFQ program, and this approach does not fully comply with the procedures set forth in Bulletin 17B, as previously noted. Therefore, modifications may be needed before this approach can be applied in practice.

Notation

A	drainage area of watershed (in square miles)
BDF	basin development factor – a number between 0 and 12 (used in the Sauer and others (1983) urban equations) that represents the degree of channel improvements, channel linings, storm drains or storm sewers, and curb and gutter streets within the upper, middle, and lower third of the watershed
C	percentage of watershed underlain by carbonate rock (in Pennsylvania Region A flood regression equations (in percent))
CA	percentage of the basin controlled by lakes, swamps, or reservoirs (in Pennsylvania Region A flood regression equations (in percent))
$c_{x,T}$	calibrated coefficients or exponents corresponding to parameter x and the T -year flood in one of the adjustment equations
F	forest cover (in percent)
$FF(.)$	flood-frequency operator in Bulletin 17B—produces a vector of flood-frequency discharge estimates (in units of cubic feet per second)
$f_T(.)$	a discharge adjustment model for return period T
$f_T^{-1}(.)$	a discharge adjustment model for return period T
HP	historical period used as input in Bulletin 17B flood-frequency analysis (in years)
IA	impervious area within the watershed (in percent)
IA_x	impervious area corresponding to the x^{th} percentile of the distribution of imperviousness within the watershed (in percent)
$IA(t)$	impervious area within the watershed in year t (in percent)
ΔIA	difference between the 10 th and 90 th percentile imperviousness values (in percent)
$\Delta IA(t)$	difference between the 10 th and 90 th percentile imperviousness values in year t (in percent)
I_T^*	threshold impervious area for return period T in equation 39 (in percent)
m	the number of coefficients and (or) exponents being calibrated in a discharge adjustment model
n	number of years in a gage record (in years)
n	number of observations (gages) used in the calibration of a discharge adjustment model
$P(Q_x)$	cumulative probability associated with having an annual maximum flood of magnitude Q_x
PD	population density (thousands of people per square mile)
PD_x	population density corresponding to the x^{th} percentile of the distribution of population density within the watershed (thousands of people per square mile)
$PD(t)$	population density in year t (thousands of people per square mile)
PD_T^*	threshold population density for return period T in equation 40 (thousands of people per square mile)
ΔPD	difference between the 10 th and 90 th percentile population density values (thousands of people per square mile)
$\Delta PD(t)$	difference between the 10 th and 90 th percentile population density values in year t (thousands of people per square mile)
P_T	cumulative probability associated with the T -year annual maximum flood
$Q_{adj}(t)$	the generic adjusted annual maximum flood in year t (in cubic feet per second)
$Q_{adj,T}(t)$	the T -year adjusted annual maximum flood in year t (in cubic feet per second)
$Q_{ff}(T)$	the T -year flood from a flood-frequency vector (in cubic feet per second)
$Q_{ff,rural}$	the rural flood-frequency vector (2-, 5-, 10-, 25-, 50-, 100-, and 500-year floods) (in cubic feet per second)
$Q_{ff,urban}$	the urban flood-frequency vector (2-, 5-, 10-, 25-, 50-, 100-, and 500-year floods) (in cubic feet per second)
$Q_{ff,x}$	the x (rural or urban) flood-frequency vector (2-, 5-, 10-, 25-, 50-, 100-, and 500-year floods) (in cubic feet per second)
$Q_{ff}(T)$	the T -year flood discharge from flood-frequency analysis (in cubic foot per second)
Q_T	the T -year flood discharge (in cubic foot per second)
$Q_{obs}(t)$	the observed annual maximum flood in year t (in cubic foot per second)
Q_l	lower bound flood-frequency discharge used in piecewise log-normal interpolation of high-outlier discharge threshold (in cubic foot per second)
Q_{max}	the largest annual maximum flood in the systematic record of a streamgage (in cubic feet per second)
Q_o	in flood-frequency analysis, the high-outlier threshold flood discharge (in cubic feet per second)
Q_u	upper-bound flood-frequency discharge used in piecewise log-normal interpolation of high-outlier discharge threshold (in cubic feet per second)
R^2	explained variance of a calibrated regression equation

$RI2$	the 2-year, 2-hour rainfall depth (used in the Sauer and others (1983) urban equations in inches)
RQ_T	rural discharge for the T -year return period (in cubic feet per second)
S_d	standard deviation of imperviousness estimates from a regression equation (in percent)
S_y	standard deviation of discharge estimates from a regression equation (in cubic feet per second)
S_e	standard error of discharge (or imperviousness) estimates from a regression equation (in cubic feet per second or in percent)
S_e/S_d	relative standard error of imperviousness estimated from population density
SL	the main channel slope measured between the points that are 10 percent and 85 percent of the main channel length upstream from the study site (in feet per mile)
SS	station skew option—an input to a Bulletin 17B flood-frequency analysis
ST	basin storage—the percentage of the drainage basin occupied by lakes, reservoirs, swamps, and wetlands (in percent)
T	return period of flood in question (in years)
T_l	lower bound return period used in piecewise log-normal interpolation of high-outlier discharge threshold (in years)
T_u	upper bound return period used in piecewise log-normal interpolation of high-outlier discharge threshold (in years)
$T(Q_x)$	the return period or frequency of a flood of magnitude Q_x (in years)
U	urban development (in Pennsylvania Region A flood regression equations (in percent))
UQ_T	urban discharge for the T -year return period (in cubic feet per second)
UQ/RQ	ratio of urban peak discharge to rural peak discharge (A value greater than 1 indicates urban amplification of flood peak)
x	the time series of annual maximum floods reported at a streamgage (reported in cubic feet per second)
y	a single, rural flood-frequency value from a statewide rural regression equation calculated for an arbitrary streamgage and return period (in cubic feet per second)
\hat{y}	a single, predicted value from an adjusted annual maximum time series calculated for an arbitrary streamgage (in cubic feet per second)
\bar{y}	the mean of the set of rural flood-frequency values from statewide rural regression equations calculated across all such values for a single return period (in cubic feet per second)
$\bar{\hat{y}}$	the mean of the set of predicted flood-frequency values from adjusted annual maximum time series calculated across all such values for a single return period (in cubic feet per second)
$z(T_x)$	the standard normal deviate corresponding to a flood of return period T_x
$z(Q_x)$	the standard normal deviate corresponding to a flood of magnitude Q_x

Discharge Adjustment Model Definitions:

Null: Urban discharges are a power function of rural discharges. The null model has no dependency on imperviousness, population density, or any other watershed characteristic. Because the model does not depend on any watershed characteristics, it is of little practical value by itself but is useful in comparison to other models being calibrated.

Simple Imperviousness: Urban discharges are a power function of rural discharges and percent imperviousness in a watershed.

Simple Density: Urban discharges are a power function of rural discharges and population density in a watershed.

Imperviousness Distribution: Urban discharges are a power function of rural discharges, percentage of imperviousness, and the difference in percentage of imperviousness between the 10th and 90th percentiles of the watershed area. This last quantity captures the homogeneity of the distribution of imperviousness within a watershed. A watershed with a large difference in percentages of imperviousness between these two percentiles has a wide range of urbanization within its domain. A watershed with a small difference is fairly uniform in degree of urbanization.

Density Distribution: Urban discharges are a power function of rural discharges, population density, and the difference in population density between the 10th and 90th percentiles of a watershed area. This last quantity captures the homogeneity of the distribution of population density within a watershed. A watershed with a large difference in population density between these

two percentiles has a wide range of urbanization within its domain. A watershed with a small difference is fairly uniform in degree of urbanization.

Scaled Imperviousness: Urban discharges are a power function of rural discharges and scaled imperviousness. This model assumes a nonlinear effect of imperviousness such that the marginal effect of increasing imperviousness is greatest close to the value I_T^* .

Scaled Density: Urban discharges are a power function of rural discharges and scaled population density. This model assumes a nonlinear effect of population density such that the marginal effect of increasing population density is greatest close to the value PD_T^* .

Sauer (null): Analogous to the “null” model described above, urban discharges are a power function of rural discharges for the dataset assembled by Sauer and others (1983).

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42 **Methods for Adjusting USGS Rural Regression Peak Discharges in an Urban Setting**

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Appendix 1

Appendix 1: Watershed characteristics of streamgages with only rural indicators.

[dd, decimal degrees; mi², square miles; ft/mi, feet per mile; ft, feet; thousand ft, thousands of feet; %, percent; mi²/mi², square miles per square mile]

Streamgage number	Streamgage name	State	Hydrologic region	2-year, 24-hour rainfall (inch)	Mean annual precipitation (inch)	Latitude (dd)	Area (mi ²)	Contributing area (mi ²)	Slope (ft/mi)
11023330	LOS PENASQUITOS C BL POWAY C NR POWAY	CA	South Coast		14		31.2767		
11023340	LOS PENASQUITOS C NR POWAY	CA	South Coast		13.5		42.3671		
11075800	SANTIAGO C AT MODJESKA	CA	South Coast		23		13.0287		
11162720	COLMA C AT S SAN FRANCISCO	CA	Central Coast		22.5		10.8595		
11162800	REDWOOD C AT REDWOOD CITY	CA	Central Coast		25		1.68449		
11166000	MATADERO C AT PALO ALTO	CA	Central Coast		20		7.05069		
11183600	WALNUT C AT CONCORD	CA	Central Coast		23		84.1331		
11336580	MORRISON C NR SACRAMENTO	CA	Central Coast		18.5		46.8109		
02231280	THOMAS C NR CRAWFORD	FL	Region A				34.9028		15.7775
02246300	ORTEGA R AT JACKSONVILLE	FL	Region A				28.1		12.9732
02246828	PABLO C AT JACKSONVILLE	FL	Region A				22.6575		15.0657
03337000	BONEYARD C AT URBANA	IL	Region 1	3			4.20883		37.931
05528500	BUFFALO C NR WHEELING	IL	Region 1 & 2	2.7			23.8388		19.8662
05529500	MC DONALD C NR MT PROSPECT	IL	Region 1	2.7			8.47548		21.7277
05530000	WELLER C AT DES PLAINES	IL	Region 1	2.7			12.8585		19.2446
05531500	SALT C AT WESTERN SPRINGS	IL	Region 1	2.7			114.996		6.41933
05532000	ADDISON C AT BELLWOOD	IL	Region 1	2.7			16.3686		17.4477
05533000	FLAG C NR WILLOW SPRINGS	IL	Region 1	2.7			16.8802		19.2316
05533400	SAWMILL C NR LEMONT	IL	Region 1	2.75			12.1458		27.5769
05534500	NB CHICAGO R AT DEERFIELD	IL	Region 1	2.7			19.6466		10.7933
05535000	SKOKIE R AT LAKE FOREST	IL	Region 1	2.7			12.7473		14.6833
05535500	WF OF NB CHICAGO R AT NORTHBROOK	IL	Region 1	2.7			11.6297		17.5063
05536000	NB CHICAGO R AT NILES	IL	Region 1	2.7			99.2577		5.79078
05536215	THORN C AT GLENWOOD	IL	Region 1	2.8			18.2409		18.583
05536255	BUTTERFIELD C AT FLOSSMOOR	IL	Region 1	2.8			23.3021		12.023
05536275	THORN C AT THORNTON	IL	Region 1	2.8			102.905		10.0069
05536340	MIDLOTHIAN C AT OAK FOREST	IL	Region 1	2.8			12.8987		18.1147
05536500	TINLEY C NR PALOS PARK	IL	Region 1	2.75			11.2544		21.7328
05539900	WB DU PAGE R NR W CHICAGO	IL	Region 2	2.75			29.629		12.3176
05540060	KRESS C AT W CHICAGO	IL	Region 2	2.8			19.9741		17.5559
05540160	E BR DU PAGE R NR DOWNERS GROVE	IL	Region 1	2.75			25.3031		16.0058
05540500	DU PAGE R AT SHOREWOOD	IL	Region 2	2.8			328.971		4.45675
05550500	POPLAR C AT ELGIN	IL	Region 2	2.75			34.625		13.7089
01097300	NASHOBA BROOK NR ACTON	MA	Eastern				11.3877		32.5419
01105600	OLD SWAMP R NR S WEYMOUTH	MA	Eastern				4.83577		34.5835
01581700	WINTERS RUN NR BENSON	MD	Piedmont					34.7	
01583500	WESTERN RUN AT WESTERN RUN	MD	Piedmont					60.3	
01584050	LONG GREEN C AT GLEN ARM	MD	Piedmont					9.3	
01584500	LITTLE GUNPOWDER FALLS AT LAUREL BROOK	MD	Piedmont					36.1	

Main channel length (mi)	Main channel elevation (ft)	Main channel elevation (thousand ft)	Basin elevation (ft)	Basin elevation (thousand ft)	Basin relief (ft)	Storage (%)	Lake area (%)	Forest cover (%)	Limestone (%)	Shape index (mi ² /mi ²)	Start imperviousness (%)	End imperviousness (%)	Period of record (years)
		0.51706									6.29	11.98	1970-1993 (23)
		0.493116									5.74	15.34	1970-2000 (31)
		1.08444									1.06	2.17	1970-2000 (31)
		0.239556									23.95	30.62	1970-1996 (26)
		0.152659									10.76	10.76	1970-1997 (28)
		0.195519									10.73	10.73	1970-2000 (30)
		0.357578									10.58	14.29	1970-1997 (24)
		0.103499									11.91	12.93	1970-2000 (21)
							4.79563				2.59	2.59	1970-2000 (28)
							1.7356				4.67	7.54	1970-2000 (31)
							24.2448				13.17	14.84	1974-2000 (24)
											28.68	28.94	1970-2000 (31)
											7.06	27.29	1970-2000 (31)
											19.78	24.97	1970-2000 (31)
											28.31	28.36	1970-2000 (31)
											16.49	21.01	1970-2000 (31)
											23.44	24.36	1970-2000 (31)
											18.91	21.81	1970-2000 (31)
											10.42	18.97	1970-2000 (25)
											11.3	12.81	1970-2000 (31)
											14.43	17.39	1970-2000 (31)
											13.29	15.53	1970-2000 (31)
											17.37	18.04	1970-2000 (31)
											18.16	18.91	1970-2000 (31)
											36.75	36.75	1970-2000 (31)
											12.67	14.81	1970-2000 (31)
											12.51	22.08	1970-2000 (31)
											9.85	20.55	1970-2000 (31)
											9.95	22.11	1970-2000 (31)
											4.52	9.57	1970-2000 (26)
											18.32	23.49	1970-2000 (18)
											9.57	9.57	1970-2000 (31)
											11.64	16.76	1970-2000 (31)
	136.09					6.93514					5.06	6.56	1970-2000 (31)
	88.4992					6.17764					15.07	15.97	1970-2000 (31)
								31.9			4.79	8.64	1970-2000 (31)
								33.6			3.25	4.55	1970-2000 (31)
								21.2			5.39	5.8	1970-2000 (25)
								31.2			4.13	5.86	1970-2000 (17)

Appendix 1: Watershed characteristics of streamgages with only rural indicators—continued.

[dd, decimal degrees; mi², square miles; ft/mi, feet per mile; ft, feet; thousand ft, thousands of feet; %, percent; mi²/mi², square miles per square mile]

Streamgage number	Streamgage name	State	Hydrologic region	2-year, 24-hour rainfall (inch)	Mean annual precipitation (inch)	Latitude (dd)	Area (mi ²)	Contributing area (mi ²)	Slope (ft/mi)
01585100	WHITEMARSH RUN AT WHITE MARSH	MD	Western Coastal Plain					7.6	
01585200	W B HERRING RUN AT IDLEWYLDE	MD	Piedmont					2.2	
01585500	CRANBERRY B NR WESTMINSTER	MD	Piedmont					3.4	
01586000	N B PATAPSCO R AT CEDARHURST	MD	Piedmont					56.3	
01589100	E B HERBERT RUN AT ARBUTUS	MD	Western Coastal Plain					2.4	
01589300	GWYNNS FALLS AT VILLA NOVA	MD	Piedmont					32.6	
01589330	DEAD RUN AT FRANKLINTOWN	MD	Piedmont					5.6	
01589440	JONES FALLS AT SORRENTO	MD	Piedmont					25.2	
01590500	BACON RIDGE B AT CHESTERFIELD	MD	Western Coastal Plain					6.9	
01591000	PATUXENT R NR UNITY	MD	Piedmont					34.8	
01591400	CATTAIL C NR GLENWOOD	MD	Piedmont					22.9	
01591700	HAWLINGS R NR SANDY SPRING	MD	Piedmont					26.2	
01639500	BIG PIPE C AT BRUCEVILLE	MD	Blue Ridge and Great Valley					102.8	
01645000	SENECA C AT DAWSONVILLE	MD	Blue Ridge and Great Valley					102	
01649500	NE B ANACOSTIA R AT RIVERDALE	MD	Western Coastal Plain					73.1	
01651000	NW B ANACOSTIA R NEAR HYATTSVILLE	MD	Western Coastal Plain					49.3	
01653600	PISCATAWAY C AT PISCATAWAY	MD	Western Coastal Plain					38.6	
02485650	PURPLE C AT JACKSON	MS	East				6.11212		27.858
02485700	HANGING MOSS C NR JACKSON	MS	East				17.1624		22.3212
02485950	TOWN C AT JACKSON	MS	East				9.49435		26.0154
02486100	LYNCH C AT JACKSON	MS	East				11.9325		28.9894
09419663	LAS VEGAS WASH TR S OF NELLIS AIR FORCE BASE	NV	Region 10			36.19444	1.04392	1.04392	
09419670	RED ROCK WASH NR BLUE DIAMOND	NV	Region 10			36.15833	8.04635	8.04635	
10311200	ASH CANYON C NR CARSON CITY	NV	Region 5			39.17639	5.38204	5.38204	
01376500	SAW MILL R AT YONKERS	NY	Region 3		48		25.2228		15.8832
04214500	BUFFALO C AT GARDENVILLE	NY	Region 6		38		142.638		10.6801
04215500	CAZENOVIA C AT EBENEZER	NY	Region 6		39		135.705		16.0498
04218518	ELLCOTT C BELOW WILLIAMSVILLE	NY	Region 6		35		80.5685		11.4892
04240010	ONONDAGA C AT SPENCER ST, SYRACUSE	NY	Region 7		37		127.722		27.3073
04240100	HARBOR BROOK AT SYRACUSE	NY	Region 7		37		10.0487		71.1392
01114000	MOSHASSUCK R AT PROVIDENCE	RI	-				15.5125		
01117000	HUNT R NR E GREENWICH	RI	-				23.0741		
01652500	FOURMILE RUN AT ALEXANDRIA	VA	Northern Piedmont				14.0295	14.0295	37.1694
01653000	CAMERON RUN AT ALEXANDRIA	VA	Northern Piedmont				33.9277	18.7489	25.8906
01654000	ACCOTINK C NR ANNANDALE	VA	Northern Piedmont				23.8425	0.781281	22.5089
02037800	FALLING C NR MIDLOTHIAN	VA	Southern Piedmont				18.7191	13.768	22.3087
02038000	FALLING C NR CHESTERFIELD	VA	Southern Piedmont				33.8736	6.35192	15.65
03052500	SAND RUN NR BUCKHANNON	WV	North Region				14.3806		
03062400	COBUN CREEK AT MORGANTOWN	WV	North Region				10.9832		

Main channel length (mi)	Main channel elevation (ft)	Main channel elevation (thousand ft)	Basin elevation (ft)	Basin elevation (thousand ft)	Basin relief (ft)	Storage (%)	Lake area (%)	Forest cover (%)	Limestone (%)	Shape index (mi ² /mi ²)	Start imperviousness (%)	End imperviousness (%)	Period of record (years)
								12.7			13.58	23.82	1970-2000 (29)
								15.5			30.06	30.06	1970-2000 (20)
								23			4.1	7.09	1970-2000 (31)
								25.9			4.78	6.98	1970-2000 (31)
								8.3			26.52	26.79	1970-2000 (22)
								24.3			15.29	20.51	1970-2000 (23)
								5			16.88	21.57	1970-2000 (20)
								46.5			7.64	11.09	1970-2000 (23)
								52.8			7.93	7.93	1970-1990 (21)
								38.2			3.52	5.38	1970-2000 (31)
								17.9			3.65	4.99	1979-2000 (22)
								26.7			6.22	8.37	1979-2000 (22)
					292.6				0		3.54	5.06	1970-2000 (31)
					253.8				0		4.8	12.89	1970-2000 (31)
								28.4			15.84	18.48	1970-2000 (31)
								20.8			22.2	24.96	1970-2000 (31)
								49.9			7.15	10.23	1970-2000 (31)
7.32016											10.71	12.13	1970-2000 (28)
8.14158											9.79	10.75	1970-2000 (31)
6.89985											20.63	20.91	1970-2000 (27)
6.70905											16.39	17.54	1970-2000 (29)
			883.491								0.46	5.53	1970-1998 (28)
			1936.4								0.37	0.64	1970-1999 (30)
			2312.27								1.74	2.89	1977-2000 (23)
	162.236					0.711219		35.7406		21.1296	19.01	19.89	1970-1999 (24)
	288.828					0.879581		42.0923		13.356	3.99	5.37	1970-2000 (31)
	424.913					0.507045		56.7619		11.9188	5.03	5.2	1970-2000 (31)
	356.182					1.21751		33.944		20.8887	8.5	8.5	1973-2000 (28)
	393.303					0.762828		51.5949		5.96086	8.22	8.22	1971-2000 (30)
	310.797					0.296921		35.3556		5.02514	11.9	12.24	1970-2000 (31)
				0.152314				29.8198			22.88	24.31	1970-2000 (31)
				0.154381				62.5405			8.07	8.88	1970-2000 (31)
8.22381	122.926							20.4007			33.81	37.05	1970-2000 (30)
12.1626	129.507							32.3356			19.85	19.85	1970-2000 (31)
11.1018	153.17							38.1093			20.53	24.51	1970-2000 (30)
9.67723	133.053							61.0473			3.7	13.32	1970-1993 (23)
14.5847	128.416							61.2425			4.64	13.87	1970-2000 (29)
											0	2.07	1970-2000 (31)
											0	5.95	1970-2000 (31)

Appendix 2

Appendix 2: Actual and estimated rural discharges for streamgages with only rural indicators.

[ft³/s, cubic feet per second; *italics*, numbers in italics satisfy equation 19 and indicate their use in the regression equation analyses in this study]

Streamgage number	Streamgage name	2-year actual (ft ³ /s)	2-year rural (ft ³ /s)	5-year actual (ft ³ /s)	5-year rural (ft ³ /s)
11023330	LOS PENASQUITOS C BL POWAY C NR POWAY	775	120	1757	490
11023340	LOS PENASQUITOS C NR POWAY	1027	141	2352	582
11075800	SANTIAGO C AT MODJESKA	338	143	1188	578
11162720	COLMA C AT S SAN FRANCISCO	1775	717	2524	1385
11162800	REDWOOD C AT REDWOOD CITY	218	277	348	445
11166000	MATADERO C AT PALO ALTO	400	447	746	872
11183600	WALNUT C AT CONCORD	4005	3209	7874	6789
11336580	MORRISON C NR SACRAMENTO	1115	4170	1809	6871
02231280	THOMAS C NR CRAWFORD	594	552	1095	1061
02246300	ORTEGA R AT JACKSONVILLE	802	663	1421	1260
02246828	PABLO C AT JACKSONVILLE	418	144	647	297
03337000	BONEYARD C AT URBANA	644	451	814	805
05528500	BUFFALO C NEAR WHEELING	447	540	602	904
05529500	MC DONALD C NEAR MOUNT PROSPECT	237	322	380	543
05530000	WELLER C AT DES PLAINES	892	422	1105	708
05531500	SALT C AT WESTERN SPRINGS	1425	1407	1873	2255
05532000	ADDISON C AT BELLWOOD	505	488	666	814
05533000	FLAG C NR WILLOW SPRINGS	779	524	1085	876
05533400	SAWMILL C NR LEMONT	707	558	1075	956
05534500	NB CHICAGO R AT DEERFIELD	417	447	583	734
05535000	SKOKIE RIVER AT LAKE FOREST	277	368	377	612
05535500	WF OF NB CHICAGO R AT NORTHBROOK	542	373	712	623
05536000	NB CHICAGO R AT NILES	1270	1192	1632	1905
05536215	THORN C AT GLENWOOD	1275	721	1752	1225
05536255	BUTTERFIELD C AT FLOSSMOOR	801	709	1254	1188
05536275	THORN C AT THORNTON	2024	2099	2884	3473
05536340	MIDLOTHIAN C AT OAK FOREST	232	541	323	921
05536500	TINLEY C NR PALOS PARK	661	469	969	796
05539900	WB DU PAGE R NR WEST CHICAGO	559	419	721	697
05540060	KRESS CREEK AT WEST CHICAGO	289	412	398	699
05540160	E BR DU PAGE R NR DOWNERS GROVE	612	768	863	1287
05540500	DU PAGE R AT SHOREWOOD	3697	1948	4965	3128
05550500	POPLAR C AT ELGIN	539	499	735	832
01097300	NASHOBA BROOK NR ACTON	202	191	329	283
01105600	OLD SWAMP R NR S WEYMOUTH	161	106	249	159
01581700	WINTERS RUN NR BENSON	2610	1590	4792	2850
01583500	WESTERN RUN AT WESTERN RUN	2332	2230	4260	3940
01584050	LONG GREEN C AT GLEN ARM	695	743	1475	1380

10-year actual (ft³/s)	10-year rural (ft³/s)	25-year actual (ft³/s)	25-year rural (ft³/s)	50-year actual (ft³/s)	50-year rural (ft³/s)	100-year actual (ft³/s)	100-year rural (ft³/s)	500-year actual (ft³/s)	500-year rural (ft³/s)
2639	969	4008	2123	5206	3331	6549	4724	10260	9586
3587	1156	5582	2542	7393	3994	9493	5678	15610	11575
2116	1156	3700	2566	5151	4069	6802	5779	11310	11754
2980	1871	3511	2511	3877	3040	4219	3521	4945	4738
423	553	504	684	555	776	598	866	678	1080
1018	1195	1401	1631	1711	1993	2040	2332	2878	3205
10560	9470	13830	13068	16120	16320	18250	19040	22600	26013
2249	8621	2766	10857	3119	12750	3446	14123	4120	17373
1553	1476	2309	2082	3022	2580	3884	3116	6625	4428
1882	1742	2505	2442	2992	3016	3493	3633	4713	5192
807	427	1018	622	1180	784	1345	961	1744	1392
906	1062	1005	1395	1069	1652	1126	1903	1238	2502
700	1150	820	1456	908	1678	994	1888	1190	2367
494	693	660	880	800	1017	956	1146	1386	1443
1221	901	1346	1142	1427	1318	1500	1483	1645	1863
2202	2806	2655	3486	3021	3961	3413	4410	4442	5426
780	1034	931	1308	1049	1507	1172	1695	1481	2125
1306	1115	1605	1414	1843	1631	2094	1837	2738	2309
1329	1231	1658	1581	1907	1840	2160	2089	2764	2666
692	924	826	1158	924	1325	1020	1482	1242	1839
437	775	504	977	550	1123	593	1260	681	1573
830	792	987	1002	1110	1154	1237	1298	1557	1626
1869	2367	2168	2934	2391	3328	2615	3701	3146	4541
2072	1573	2481	2013	2790	2338	3101	2651	3848	3377
1627	1512	2194	1919	2693	2216	3263	2499	4927	3154
3500	4400	4333	5573	4992	6418	5684	7234	7449	9117
389	1182	481	1512	555	1755	635	1989	843	2530
1179	1022	1450	1305	1655	1515	1863	1714	2360	2176
817	880	926	1108	1001	1269	1072	1421	1221	1765
474	893	574	1136	652	1313	732	1480	932	1864
1062	1640	1354	2084	1603	2408	1882	2718	2669	3431
5649	3884	6371	4821	6823	5466	7215	6077	7954	7457
853	1053	989	1328	1082	1524	1169	1709	1354	2128
429	359	576	471	701	567	838	676	1217	963
308	204	382	270	436	326	490	391	614	564
6478	3970	8828	5690	10710	7210	12690	8970	17680	14200
5852	5440	8227	7740	10260	9760	12530	12100	18810	19000
2207	1960	3416	2870	4550	3680	5903	4630	10080	7490

50 Methods for Adjusting USGS Rural Regression Peak Discharges in an Urban Setting

Appendix 2: Actual and estimated rural discharges for streamgages with only rural indicators—continued.

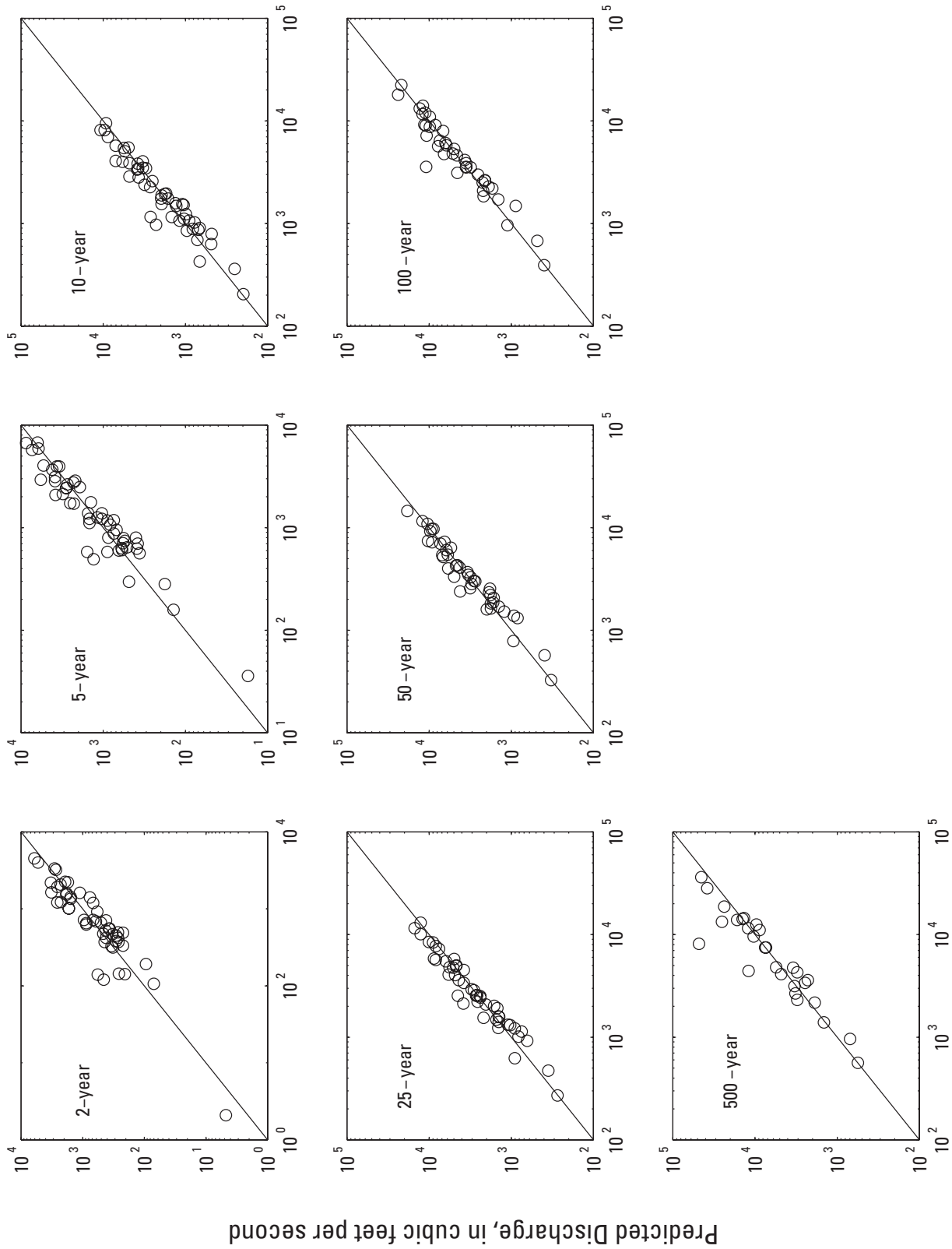
[ft³/s, cubic feet per second; *italics*, numbers in italics satisfy equation 19 and indicate their use in the regression equation analyses in this study]

Streamgage number	Streamgage name	2-year actual (ft³/s)	2-year rural (ft³/s)	5-year actual (ft³/s)	5-year rural (ft³/s)
01584500	LITTLE GUNPOWDER FALLS AT LAUREL BROOK	3856	1630	6022	2930
01585100	WHITEMARSH RUN AT WHITE MARSH	1559	625	2248	1116
01585200	WB HERRING RUN AT IDLEWYLDE	718	314	1141	606
01585500	CRANBERRY B NR WESTMINSTER	380	386	845	740
01586000	NB PATAPSCO R AT CEDARHURST	2685	2250	4232	3960
01589100	EB HERBERT RUN AT ARBUTUS	663	329	932	602
01589300	GWYNNNS FALLS AT VILLA NOVA	1940	1610	3303	2880
01589330	DEAD RUN AT FRANKLINTOWN	1587	654	2336	1220
01589440	JONES FALLS AT SORRENTO	1111	1190	2006	2180
01590500	BACON RIDGE B AT CHESTERFIELD	196	240	311	457
01591000	PATUXENT R NR UNITY	1491	1530	2896	2750
01591400	CATTAIL C NR GLENWOOD	2169	1360	3413	2450
01591700	HAWLINGS R NR SANDY SPRING	1329	1370	2625	2480
01639500	BIG PIPE C AT BRUCEVILLE	3318	3864	5544	6874
01645000	SENECA C AT DAWSONVILLE	4031	3366	7435	5876
01649500	NE B ANACOSTIA R AT RIVERDALE	4953	2204	7121	4039
01651000	NW B ANACOSTIA R NR HYATTSVILLE	3645	2063	5594	3670
01653600	PISCATAWAY C AT PISCATAWAY	975	922	1852	1770
02485650	PURPLE C AT JACKSON	1091	692	1416	1168
02485700	HANGING MOSS C NR JACKSON	2771	1527	3693	2638
02485950	TOWN C AT JACKSON	2725	1008	3288	1718
02486100	LYNCH C AT JACKSON	3316	1229	3883	2118
09419663	LAS VEGAS WASH TR S OF NELLIS AIR FORCE BASE	1.40	12.0	24.3	87.0
09419670	RED ROCK WASH NR BLUE DIAMOND	5.40	40.0	112	291
10311200	ASH CANYON C NR CARSON CITY	14.3	2.10	38.3	36.0
01376500	SAW MILL R AT YONKERS	687	852	965	1371
04214500	BUFFALO C AT GARDENVILLE	6251	4005	7938	5747
04215500	CAZENOVIA C AT EBENEZER	7115	4541	9333	6674
04218518	ELLCOTT C BELOW WILLIAMSVILLE	1787	2190	2347	3204
04240010	ONONDAGA C AT SPENCER ST, SYRACUSE	1907	6132	2643	9086
04240100	HARBOR BROOK AT SYRACUSE	227	950	342	1453
01114000	MOSHASSUCK R AT PROVIDENCE	913	369	1116	632
01117000	HUNT R NR E GREENWICH	434	334	629	564
01652500	FOURMILE RUN AT ALEXANDRIA	3048	1010	3918	1737
01653000	CAMERON RUN AT ALEXANDRIA	3908	1221	5077	2093
01654000	ACCOTINK C NR ANNANDALE	2651	1429	3961	2444
02037800	FALLING C NR MIDLOTHIAN	534	380	852	647
02038000	FALLING C NR CHESTERFIELD	830	477	1374	803
03052500	SAND RUN NEAR BUCKHANNON	858	951	1472	1518
03062400	COBUN CREEK AT MORGANTOWN	470	782	855	1264

10-year actual (ft ³ /s)	10-year rural (ft ³ /s)	25-year actual (ft ³ /s)	25-year rural (ft ³ /s)	50-year actual (ft ³ /s)	50-year rural (ft ³ /s)	100-year actual (ft ³ /s)	100-year rural (ft ³ /s)	500-year actual (ft ³ /s)	500-year rural (ft ³ /s)
7367	4080	8930	5850	9994	7410	10970	9210	12990	14600
2736	1545	3385	2220	3892	2836	4419	3560	5738	5816
1458	881	1899	1320	2256	1710	2636	2180	3622	3600
1250	1070	1862	1590	2385	2060	2959	2610	4493	4280
5494	5480	7386	7800	9033	9840	10900	12200	16260	19200
1128	847	1395	1239	1609	1594	1836	2006	2423	3291
4502	4020	6420	5760	8187	7310	10280	9090	16770	14500
2882	1740	3628	2560	4224	3300	4854	4160	6474	6800
2811	3060	4119	4410	5338	5610	6797	6980	11360	11100
388	668	484	1040	553	1410	621	1880	774	3510
4225	3840	6473	5510	8643	6990	11310	8690	20030	13800
4183	3430	5074	4940	5677	6290	6230	7850	7357	12600
3719	3470	5364	5000	6775	6350	8342	7920	12640	12600
7839	9593	12080	13947	16550	17846	22520	22395	45190	36241
10440	8092	15210	11526	19550	14499	24640	18017	40030	28327
8661	5701	10720	8387	12340	10951	14020	13983	18270	23874
7167	5089	9510	7291	11540	9277	13820	11645	20350	18755
2783	2590	4550	4020	6459	5420	9050	7200	19140	13300
1622	1540	1873	1983	2056	2348	2235	2647	2645	3476
4192	3501	4721	4508	5053	5333	5343	6018	5894	7902
3590	2270	3913	2918	4121	3452	4306	3889	4674	5097
4195	2811	4539	3615	4767	4279	4975	4819	5403	6320
89.8	205	318	411	673	607	1262	876	3987	1837
485	729	2106	1551	5179	2386	11260	3583	49570	8164
69.9	181	142	1072	232	3026	372	7003	1043	38288
1167	1795	1441	2433	1660	2959	1891	3531	2488	5142
8897	6940	9968	8480	10680	9696	11340	10974	12680	13982
10680	8154	12280	10065	13390	11583	14460	13181	16800	17015
2738	3898	3255	4786	3658	5478	4076	6194	5125	7844
3153	11177	3823	13775	4341	15704	4874	17595	6195	22190
420	1801	519	2218	594	2507	668	2771	845	3402
1248	694	1416	1015	1541	1385	1667	1830	1966	3219
779	628	993	919	1171	1253	1367	1656	1900	2912
4432	2381	5024	3394	5431	4299	5813	5349	6630	8455
5819	2867	6728	4086	7387	5176	8035	6441	9522	10197
4786	3345	5770	4765	6459	6036	7112	7514	8518	11909
1068	862	1341	1217	1542	1552	1740	1923	2193	3054
1735	1061	2177	1479	2491	1868	2792	2293	3443	3576
2011	1944	2869	2536	3657	3012	4588	3518	7455	4795
1238	1631	1919	2143	2613	2556	3510	2997	6712	4120

Appendix 3

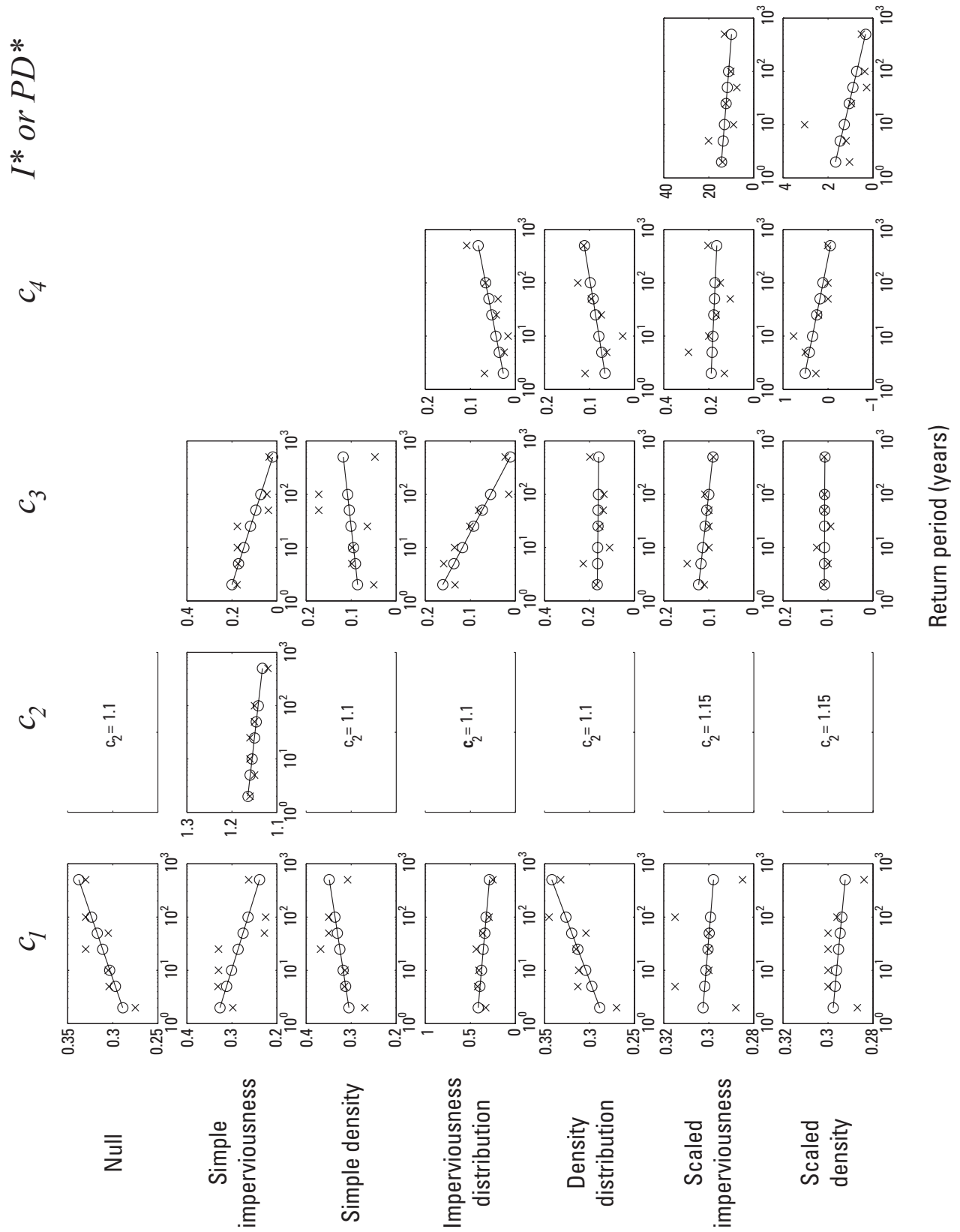
Appendix 3. Plots of predicted and observed flood discharges. Predicted discharges are from calibrated simple imperviousness adjustment model. Observed are from appropriate U.S. Geological Survey rural regression equations.



Observed Discharge, in cubic feet per second

Appendix 4

Appendix 4. Calibrated (x) and smoothed (o) coefficient and exponents for the discharge adjustment models and return period.



Appendix 5

Appendix 5. Watershed characteristics for 10 U.S. Geological Survey streamgages with localized urban equations.

[mi², square miles; ft/mi, feet per mile; %, percent]

Streamgage number	Streamgage name	State	Area (mi ²)	Slope (ft/mi)	Storage (%)	Imperious area (%)	Forest cover (%)	Urban area (%)	Carbonate area (%)	Controlled area (%)
01393450	ELIZABETH R AT URSINO LAKE AT ELIZABETH NJ	NJ	17.6	22.4	1.6	41.9				
01394500	RAHWAY R NR SPRINGFIELD NJ	NJ	25.3	30.6	1.3	22.2				
01407705	SHARK R NR NEPTUNE CITY NJ	NJ	9.8	29.4	17.6	9.0				
01407760	JUMPING BROOK NR NEPTUNE CITY NJ	NJ	6.3	35.2	8.4	10.1				
01465798	POQUESSING C AT GRANT AVE AT PHILADELPHIA PA	PA	21.1				16.6	66.5	1.1	14.9
01467048	PENNYPACK C AT LOWER RHAWN ST BDG, PHILA PA	PA	50.0				29.2	59.4	1.5	6.1
01467087	FRANKFORD C AT CASTOR AVE, PHILADELPHIA PA	PA	29.8				17.4	77.2	0.0	0.0
03049800	LITTLE PINE C NR ETNA PA	PA	5.7				79.8	8.2	0.0	0.0
04087204	OAK C AT S MILWAUKEE WI	WI	25.4			16.8				
04087220	ROOT R NR FRANKLIN WI	WI	49.8			19.6				

Appendix 6

Appendix 6. Rural and urban discharges, in cubic feet per second for 10 U.S. Geological Survey streamgages with localized urban equations.

[Entries in *italics* were not available directly from the regression equations and were determined from log-normal extrapolation.]

Streamgage number	Streamgage name	State	Return Period (years)							
			2	5	10	25	50	100	500	
Urban Discharges										
01393450	ELIZABETH R AT URSINO LAKE AT ELIZABETH NJ	NJ	1,071	1,516	1,992	2,552	2,993	3,528	4,922	
01394500	RAHWAY R NR SPRINGFIELD NJ	NJ	1,452	2,088	2,784	3,580	4,219	5,016	7,120	
01407705	SHARK R NR NEPTUNE CITY NJ	NJ	157	244	338	460	568	694	1,039	
01407760	JUMPING BROOK NR NEPTUNE CITY NJ	NJ	166	254	350	476	584	715	1,075	
01465798	POQUESSING C AT GRANT AVE AT PHILADELPHIA PA	PA	2,739	3,984	4,845	5,970	7,009	8,140	11,180	
01467048	PENNYPACK C AT LOWER RHAWN ST BDG, PHILADELPHIA PA	PA	5,019	7,095	8,503	10,312	11,971	13,742	18,381	
01467087	FRANKFORD C AT CASTOR AVE, PHILADELPHIA PA	PA	4,520	6,213	7,337	8,761	10,060	11,433	14,968	
03049800	LITTLE PINE C NR ETNA PA	PA	347	611	821	1,124	1,415	1,752	2,742	
04087204	OAK C AT S MILWAUKEE WI	WI	941	1,301	1,585	1,912	2,207	2,440	2,991	
04087220	ROOT R NR FRANKLIN WI	WI	1,847	2,518	3,062	3,691	4,266	4,716	5,777	
Rural Discharges										
01393450	ELIZABETH R AT URSINO LAKE AT ELIZABETH NJ	NJ	550	863	1,217	1,672	2,104	2,674	4,343	
01394500	RAHWAY R NR SPRINGFIELD NJ	NJ	903	1,411	1,993	2,713	3,386	4,287	6,912	
01407705	SHARK R NR NEPTUNE CITY NJ	NJ	91	152	220	313	405	517	850	
01407760	JUMPING BROOK NR NEPTUNE CITY NJ	NJ	97	159	229	327	421	540	896	
01465798	POQUESSING C AT GRANT AVE AT PHILADELPHIA PA	PA	1,132	2,089	2,879	4,051	5,220	6,626	11,062	
01467048	PENNYPACK C AT LOWER RHAWN ST BDG, PHILADELPHIA PA	PA	2,237	3,932	5,281	7,233	9,142	11,383	18,204	
01467087	FRANKFORD C AT CASTOR AVE, PHILADELPHIA PA	PA	1,676	3,012	4,090	5,670	7,227	9,075	14,791	
03049800	LITTLE PINE C NR ETNA, PA	PA	303	553	757	1,059	1,415	1,752	2,737	
04087204	OAK C AT S MILWAUKEE, WI	WI	454	651	792	966	1,095	1,219	1,514	
04087220	ROOT R NR FRANKLIN, WI	WI	796	1,127	1,360	1,645	1,859	2,062	2,544	

