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1 **Altitude and temperature drive anuran community assembly in a Neotropical**
2 **mountain region**

3

4 Altitude and temperature drive anuran communities (**running title**)

5

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25

26 **Abstract**

27 **Aim:** Understanding the spatial variation in species richness and the mechanisms that
28 limit species range sizes along geographical gradients belong to the central research
29 issues in macroecology. Here, we aim to test the topographic and climatic effects on
30 anuran species richness and community composition in mountainous regions in the
31 Brazilian Atlantic Forest biome.

32 **Location:** Atlantic Forest, Southeastern Brazil.

33 **Taxon:** Amphibia: Anura.

34 **Methods:** We used Individual-Based Rarefaction Curve (interpolation and
35 extrapolation), Generalized Additive Model (GAM), Midpoint method and Principal
36 Coordinates Analysis (PCoA) to analyze the topographic and climatic effects on anuran
37 community composition, richness, and range sizes in a global biodiversity hotspot.

38 **Results:** Our results showed that altitude and annual mean temperature and temperature
39 seasonality were the main drivers of species altitudinal range sizes and community
40 assembly in mountainous regions.

41 **Main conclusion:** Anuran community richness peaked at intermediate altitudes
42 following thus a hump-shaped pattern and corroborated the Rapoport`s altitudinal rule
43 as range sizes increased with altitude in mountainous regions from the Atlantic Forest
44 biome. Study revealed new insights into the patterns and drivers of Neotropical anuran
45 communities.

46 **Keywords:** Amphibians, Atlantic Forest, Biodiversity hotspots, Neotropical region,
47 Poços de Caldas Plateau, Range size, Rapoport altitudinal rule.

48

49

50

51 **1. INTRODUCTION**

52

53 Understanding the spatial variability in species richness and the mechanisms that
54 limit species range sizes along geographical gradients is an important issue in
55 biogeography and macroecology (Gaston & Blackburn, 2000; Gaston, 2003; Sexton et
56 al., 2009). Such research questions have continued to fascinate biogeographers and
57 macroecologists for several decades (Rahbek, 1995; McCain & Grytnes, 2010; Sanders
58 & Rahbek, 2012), being still a current challenge (Gaston, 2000).

59 Several studies conducted with different taxa and geographic regions support the
60 view that biodiversity (typically species richness) often shows a predictable variation
61 along mountainous regions typically showing decreasing, increasing or hump-shaped
62 patterns (Rahbek, 1995; McCain & Grytnes, 2010; Sanders & Rahbek, 2012). Some
63 mechanisms such as climatic, biological, evolutionary, and historical process have been
64 proposed to explain patterns of species richness in mountainous regions (Sanders et al.,
65 2003; McCain & Grytnes, 2010; Sanders & Rahbek, 2012), but currently the most
66 important processes underlying species richness patterns are still poorly understood for
67 many taxa and ecosystems (Brown, 2001; Grau et al., 2007; Gallou et al., 2017).

68 One of the biogeographical hypotheses proposed to explain species richness
69 patterns focuses on species range size variation along geographical gradients, namely
70 the Rapoport's rule (Stevens, 1989; 1992). Although Rapoport's rule has been initially
71 proposed to explain species range size patterns along latitudinal gradients (Steven,
72 1989), later it was extended to explain species ranges size along the altitudinal gradients
73 too (Stevens, 1992; Teittinen et al., 2016). According to the Rapoport altitudinal rule,
74 there is a positive correlation between altitude and species altitudinal range size
75 (Stevens, 1992), with higher altitude species tending to have broader altitudinal ranges

76 due to wider climatic tolerances than lower altitude species and consequently covering
77 wider altitudinal ranges. On the other hand, lower altitude species have narrower
78 climatic tolerances and hence narrower altitudinal ranges due to more stable
79 environmental conditions in low altitude regions. It is worth noting that unlike the mid
80 domain effect based on a null model (Colwell & Lees, 2000), Rapoport altitudinal rule
81 (Stevens, 1992) attempts to explain species spatial distribution patterns based on the
82 climatic variation effects. Thus, for explaining Rapoport altitudinal rule, Stevens (1992)
83 proposed the climatic variability hypothesis or simply Rapoport “rescue” hypothesis,
84 which predicts that species that have broad physiological tolerance are able to become
85 more widely distributed (Stevens, 1992; Gaston & Chown, 1999).

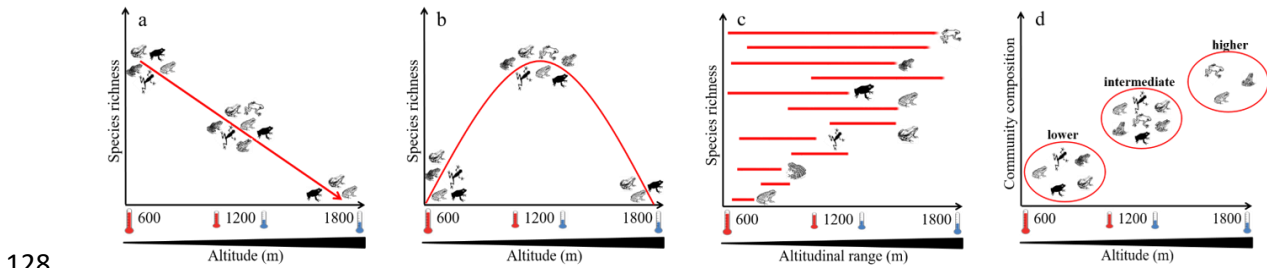
86 Anurans are considered an interesting animal group to test Rapoport altitudinal
87 rule due to their complex life cycles (Duellman & Trueb, 1986), limited dispersal ability
88 (Duellman & Trueb, 1986) and because they are ectothermic vertebrates highly
89 sensitive to rainfall and temperature levels, which influence their development,
90 metabolism, behaviour, and reproductive rate (Duellman & Trueb, 1986; Buckley &
91 Jetz, 2007; Vitt & Caldwell, 2009). Studies carried out specifically with anurans show
92 that species richness peaks can be found in lower (Hofer et al., 1999 [Cameroon];
93 Goyannes-Araújo et al., 2015 [Brazil]), intermediate (Fauth et al., 1989 [Costa Rica]; Fu
94 et al., 2006 [China]; Kozak & Wiens, 2010 [North America]; Hutter et al.,
95 2013[Andes]) and higher altitudes (Giaretta et al., 1999 [Brazil]; Naniwadekar &
96 Vasudevan, 2007 [India]), depending on the region (McCain & Grytnes, 2010).
97 However, we expected to find a peak of anuran species richness at intermediate altitudes
98 showing thus “unimodal or hump-shaped pattern” because it is the most common trend
99 for all vertebrate taxa (Rahbek, 1995; McCain & Grytnes, 2010).

100 The Atlantic Forest biome is one of the most biodiverse regions worldwide
101 including rare, endemic and threatened anuran species (Haddad et al., 2013; Rossa-
102 Feres et al., 2017) and considered one of the top-five global biodiversity hotspots in the
103 world (Myers et al., 2000; Mittermeier et al., 2011). Despite the high species richness
104 and endemism level, the Atlantic Forest has been under great human impact for about
105 half a century due to habitat fragmentation resulting in severe biodiversity loss (Ribeiro
106 et al., 2009; Haddad et al., 2015).

107 Climate change is also expected to force species distributions towards higher
108 altitudes (Chen et al., 2011; Tingley et al., 2012; Freeman and Freeman, 2014) and
109 species range shifts may accelerate in the future (Elsen et al., 2018). This would lead to
110 the extinction of small-ranged species and weak dispersal (Janzen, 1967; Davies et al.,
111 2009; Sandel et al., 2011) of small and isolated populations from their current
112 geographical ranges (Midgley et al., 2002; Thomas et al., 2004), especially in tropical
113 mountainous regions, where the risk of decline may be higher and more severe (McCain
114 and Colwell, 2011; Davidson et al., 2013).

115 Here, we aim to analyze the topographic and climatic effects on anuran
116 community composition, richness, and range sizes in a mountainous region in the
117 Atlantic Forest biome. Specifically, we: (1) determine the shape of the distribution
118 pattern of anuran species richness along the altitudinal gradient; (2) test if the Rapoport
119 rule explains anuran range size variation along the altitudinal gradient; and (3)
120 investigate how anuran community composition varies along the altitudinal gradient.
121 Our hypotheses are: 1) Anuran species richness will exhibit either a decreasing or a
122 unimodal pattern along the altitudinal gradient (Figure 1A and 1B); 2) Species recorded
123 at higher altitudes will show wider range size than lower altitudes species along the
124 altitudinal gradient (Figure 1C), and 3) Anuran community composition will show three

125 distinct groups (lower, intermediate, and higher altitudes) along the altitudinal gradient
126 mainly because of different environmental conditions at different elevations (Figure
127 1D).



128
129 **Figure 1.** Hypothetical relationships of species richness and community
130 composition in response to the topographic and climatic variables along an
131 altitudinal gradient in Poços de Caldas Plateau region, Southeastern Brazil. In
132 panel [a] we hypothesize that species richness will decline with altitude as a
133 consequence of temperature reduction; in [b] we hypothesize that in the
134 intermediate altitudes, we will detect higher richness, because this range will
135 share species from both lower and higher altitudes; in [c] we hypothesize that
136 species that are found in higher altitudes will also have higher plasticity in their
137 thermal tolerance, thus being also found in lower altitudes; in [d] we present an
138 alternative hypothesis, where community composition differs among lower,
139 intermediate, and high altitudes with highest richness in intermediate altitudes,
140 as presented in [b].

141

142 2. METHODS

143

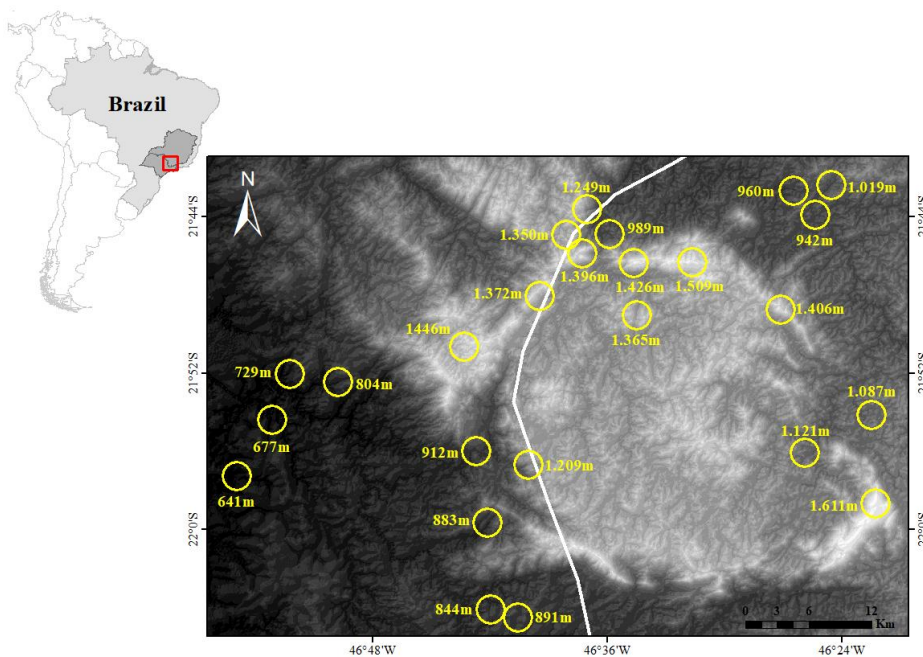
144 2.1. Study area

145 The study was carried out in Poços de Caldas Plateau region, considered one of
146 the largest set of alkaline rocks on Earth (Figure 2). The Plateau is located on the

147 extreme western edge of the Serra da Mantiqueira, a mountain range in eastern South
148 America, Southeastern Brazil (Christofoletti, 1973).

149 The Plateau region is characterized by a main volcanic caldera almost complete
150 with 800 km² (Christofoletti, 1973) formed from the intrusion of alkaline rocks during
151 the Upper Cretaceous (Ellert, 1959), about 80 million years ago (Thomaz-Filho &
152 Rodrigues, 1999) on the border between Minas Gerais and São Paulo States (Figure 2).

153



154

155 **Figure 2.** Study area covering 25 sampled landscapes distributed along an
156 altitudinal gradient in Poços de Caldas Plateau region, Southeastern Brazil.
157 White line indicates the border between Minas Gerais (*right*) and São Paulo
158 (*left*) States. Yellow circles (*buffers 1 km*) represent the sampled landscapes.
159 Altitudes represent the mean elevations within each landscape sampled (buffers
160 1 km).

161

162 The climate of the region is Cwb of Köppen class “subtropical of altitude” being
163 mesothermic with a dry winter from April to September and with mild summers and a
164 rainy season from October to March. Mean annual temperatures vary between 17.6°C
165 and 18.4°C, with the mean annual temperature of the coolest month around 16.5°C and
166 of the warmest month not exceeding 22°C. Mean annual precipitation varies between
167 1,300 and 1,700 mm (Pell et al., 2007). The regional altitude from the Poços de Caldas
168 Plateau varies from 600 to 1,800 m a.s.l. (Fraenkel et al., 1985). The region belongs to
169 the Atlantic Forest biome (Ab’Saber, 1989) considered one of top-five global
170 biodiversity hotspot (Myers et al., 2000; Mittermeier et al., 2011). Sampling areas are
171 located in a transition area between the Atlantic Semi-deciduous Forest, Cerrado
172 patches, and high-altitude fields enclaves (above 1,200 m a.s.l) in the interior of the
173 Poços de Caldas Plateau region (Veloso et al., 1991).

174

175 **2.2. Anuran sampling**

176 We sampled the anuran communities in 25 landscapes distributed along an
177 altitudinal gradient of 600 to 1,800 m a.s.l. with a minimum distance of 3 km between
178 each landscape (Figure 2). We defined our altitudinal boundaries as lower (600 to 900
179 m), intermediate (900 to 1,200 m), and higher altitudes (above 1,200 m a.s.l.). The
180 sampled landscapes were delineated by a 1 km radius around the landscapes’ centroid as
181 this radius is considered a reasonable sized area comprising dispersion and migration
182 movements for most anuran species (Guerry & Hunter, 2002; Collins & Fahrig, 2017).
183 Moreover, we argue that the use of 1-km radius size ensures that landscapes can be
184 regarded as independent sampling areas in data analyses.

185 Anuran sampling was performed monthly in the rainy season and bi-monthly in
186 the dry season from December 2016 to December 2017. Visual (young and adults) and

187 acoustic search (calling males) methods were used simultaneously to maximize the
188 number of detected species (Heyer et al., 1994). Each landscape was investigated at
189 night between 1800 and 2400h always by two researchers (RM and JMO). Different
190 areas in the landscapes and all potential microhabitat present (rocks, fallen logs, leaf
191 litter, branches of trees or forest vegetation, and lentic and lotic water bodies when
192 presents) were sampled during the research. All the anurans encountered during the
193 fieldwork were captured, identified in situ, and then most individuals were released (but
194 see below). In order to standardize the sampling effort among the landscapes and to
195 minimize the number of false absences, the landscapes were sampled 10 times in a
196 randomized order and standardized by time (1 h/night), totalizing 10 hours per each
197 landscape (Heyer et al., 1994).

198 Specimens that could not be identified in the field were collected (under SISBio
199 license number #48526-1), anesthetized and euthanized with xylocaine 20%, fixed in
200 10% formalin and preserved in alcohol 70%. Vouchers were deposited in the Coleção
201 Herpetológica do Laboratório de Zoologia de Vertebrados da ESALQ/USP (acronym
202 VESALQ).

203

204 **2.3. Topographic and bioclimatic variables**

205 We extracted topographic variables (altitude and slope) from the Brazil
206 Geomorphometric Database (TOPODATA) with spatial resolution 30 m – downloaded
207 from (<http://www.dsr.inpe.br/topodata>) and 19 standardized climatic variables (Bio 1 to
208 Bio 19) obtained from the Global Climate Data (WorldClim version2) with spatial
209 resolution of 30 seconds-arc “~1 km²” (Fick and Hijmans 2017) – downloaded from
210 (<http://www.worldclim.org>). These variables were obtained from the interpolation of
211 average climate measurements from weather stations (Fick and Hijmans 2017). We

212 included these variables into data analyses as topography and climate represent the most
213 important factors that determine anuran distributions (Duellman, 1999; Menin et al.,
214 2007; Vasconcelos et al., 2010).

215

216 **2.4. Data analyses**

217 To test whether anuran species richness shows a unimodal hump-shaped pattern
218 along the altitudinal gradient, we first used an individual-based rarefaction curve to
219 control for confounding effects of species abundance on richness (Gotelli & Colwell,
220 2001). Such rarefaction is needed along the altitudinal gradients because abundance
221 typically affects anuran population density and size, which in turn have a positive effect
222 on species richness (Chao et al., 2014). Furthermore, we combined rarefaction
223 (interpolation) and prediction (extrapolation) to make a meaningful comparison
224 standardized by identical sampling effort (Chao et al., 2014). Whereas the interpolation
225 was performed based on the landscape with the smallest number of individuals ($n =$
226 105), the extrapolation was based on a larger number of individuals ($n = 1,145$). Chao et
227 al. (2014) have extended the classical models of rarefaction for the most common
228 diversity measures (species richness, Shannon index, and Simpson's index) throughout
229 Hill numbers. We used the nearly unbiased Shannon estimation ($q = 1$, Jost, 2007),
230 which represents the alpha diversity with proportional weight to rare and common
231 species.

232 In addition, we also tested whether topographic (altitude and slope) and climatic
233 variables (19 bioclimatic variables) affect species rarefied richness. Because we
234 expected a non-linear relationship between altitude and species richness, we used a
235 Generalized Additive Model (GAM) that is able to fit non-linear patterns using a
236 smooth function (Zuur et al., 2009). To simplify the statistical model, we started with a

237 global model including uncorrelated predictors (see below) and compared this model
238 with nested, simpler models. In addition, this model simplification compare linear
239 effects with smoothed effects to obtain the best fit, as demonstrated in an example
240 model:

```
241 gam.global <- gam (rarefied_richness ~ bio1 + bio4 + bio15 + alt)
```

```
242 summary (gam.global)
```

```
243 gam.mod1 <- gam (rarefied_richness ~ bio1 + bio4 + bio15 + s(alt))
```

```
244 summary (gam.mod1)
```

245 The function compares whether the smoothed (s) effects represent a best fit with
246 the linear effects. Furthermore, to avoid the undesirable effects of multicollinearity on
247 model parameters, we first compared with a multiple correlation all predictor variables
248 and removed those highly correlated ($r > 0.6$). We also retained only variables with a
249 recognized biological meaning for our hypothesis.

250 To test the Rapoport rule (that is, anuran range size variation) along the
251 altitudinal gradient, we used the midpoint method (Rohde et al., 1993) and calculated
252 the weighted occurrence mean (between lower and upper limits of distribution of each
253 species) to understand the altitudinal trends of anuran ranges size. Species that were
254 recorded only at one site were excluded from the analyses.

255 To visualize differences in anuran community composition along the altitudinal
256 gradient, we performed a Principal Coordinates Analysis (PCoA) using the abundance-
257 based distance method of Bray-Curtis (Legendre & Legendre, 2012). PCoA ordines
258 the samples so that the distance between points represents the dissimilarity between
259 samples. Scores of PCoA represent a measure of species composition weighed by
260 species abundance.

261 Finally, to test whether community composition varies significantly among
262 altitudinal zones (i.e., among lower, intermediate, and higher altitudes) we performed a
263 permutational multivariate analysis of variance (PERMANOVA – Anderson, 2001).
264 PERMANOVA is a routine analysis for testing the simultaneous response of one or
265 more variables to one or more factors, thus allowing the analysis of multivariate data in
266 the context of more complex sampling structures. Probabilities were calculated with
267 permutation method randomized 9,999 times. We also used the BETADISPER method
268 to investigate if there is a variance heterogeneity affecting the results obtained from the
269 PERMANOVA method (Anderson & Walsh, 2013). This is an essential step because
270 PERMANOVA confounds location (a measure of composition) and group dispersion (a
271 measure of within-group compositional variation). All analyses were conducted using
272 the R software (R Development Core Team, 2017).

273

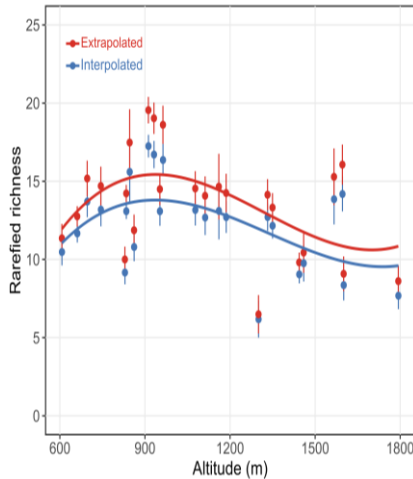
274 **4. RESULTS**

275

276 **4.1. Patterns of species richness and abundance**

277 We recorded 10,935 specimens belonging to 10 families, 20 genera, and 70
278 species along the altitudinal gradient (Table S1), representing approximately 11% of
279 Atlantic Forest (Rossa-Feres et al., 2017) and 7% of the Brazilian anuran fauna (Segalla
280 et al., 2019). Hylidae was the richest family (37 species or 53%) followed by
281 Leptodactylidae (12 or 17,1%), Bufonidae (5 or 7,1%), Brachycephalidae (5 or 7,1%),
282 Hylodidae (4 or 6%), Odontophrynidae (3 or 4,2%), Craugastoridae (1 or 1,4%),
283 Phyllomedusidae (1 or 1,4%), Microhylidae (1 or 1,4%), and Ranidae (1 or 1,4%). The
284 most abundant species was *Dendropsophus minutus* (N=1,123 or 10% of sampled
285 individuals), followed by *Physalaemus cuvieri* (N=1,099 or 10%) and *Ischnocnema*

286 *juipoca* (N=894 or 8%). These three species represent approximately 28% of all
287 sampled individuals (Table S1). Anuran rarefied richness (interpolated and
288 extrapolated) showed a unimodal pattern along the altitudinal gradient, with the highest
289 species richness found in midlands at ca. 900 m elevation (Figure 3).



290

291 **Figure 3.** Individual-based rarefaction curve on rarefied (interpolated and
292 extrapolated) richness along an altitudinal gradient in Poços de Caldas Plateau
293 region, Southeastern Brazil.

294

295 According to GAM model, altitude ($p=0.004$), annual mean temperature
296 ($p=0.004$) and temperature seasonality ($p=0.005$) were the main drivers influencing
297 anuran species richness along the altitudinal gradient (Table 1). The other variables did
298 not contribute significantly ($P > 0.05$) to variation in anuran richness.

299

300 **Table 1:** Influence of the topographic and climatic variables on species rarefied
301 (interpolated and extrapolated) richness through a Generalized Additive Model (GAM)
302 along an altitudinal gradient in Poços de Caldas Plateau region, Southeastern Brazil.

303

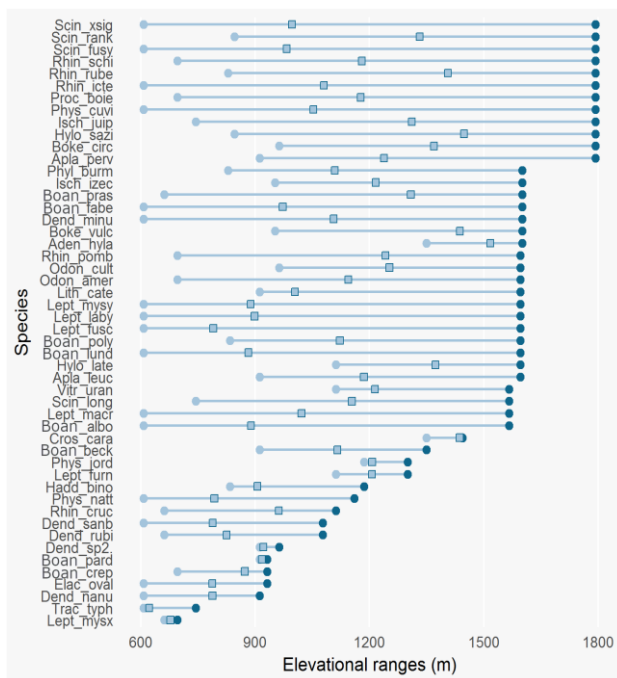
304

INTERPOLATED RICHNESS				
Non linear models	Edf	Ref.df	F	P-value
sAltitude (Alt)	1.000.000	1.000.000	10.464.824	0.004
sTemperature Seasonality (Bio 04)	3.461.033	4.251.915	4.951.471	0.005
EXTRAPOLATED RICHNESS				
Linear models	Estimate	Std. Error	t-value	P-value
(Intercept)	-9.510.342	72.005.629	-1.320.778	0.201
Annual Mean Temperature (Bio 01)	1.235.705	0.3845988	3.212.972	0.004
Non linear models	Edf	Ref.df	F	P-value
sTemperature Seasonality (Bio 04)	3.450.829	4.246.575	4.978.777	0.005

305

306 4.2. Rapoport altitudinal rule

307 Anurans altitudinal range sizes increased with altitude, with higher-altitude
 308 species showing wider ranges than lower-altitude species, thus corroborating the
 309 Rapoport altitudinal rule (Figure 4).



310

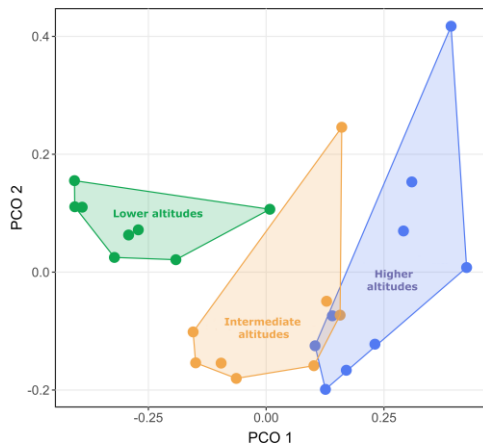
311 **Figure 4.** Altitudinal range size distributions of anurans in Poços de Caldas

312 Plateau region, Southeastern Brazil. Bars show altitudinal range size along an
 313 altitudinal gradient. Square: median abundance. The range profiles were based
 314 on the lowest and highest elevation at which a species was observed.

315

316 **4.3. Altitudinal variation in anuran composition**

317 The Principal Coordinates Analysis (PCoA) revealed that anuran
318 community composition clustered into three distinct groups along the altitudinal
319 gradient (axis 1 = 41.43% and axis 2 = 14.84%): (i) first group comprised “lower
320 altitude species”, (ii) second “intermediate altitude species” and (iii) third “higher
321 altitude species” (Figure 5). These results were supported by the PERMANOVA ($F =$
322 12.867 , $R^2 = 0.35874$, $P < 0.001$).



323

324 **Figure 5.** Principal Coordinates Analysis (PCoA) ordination of anurans
325 communities' composition along an altitudinal gradient in Poços de Caldas
326 Plateau region, Southeastern Brazil. Ordination was based on dissimilarity
327 matrix calculated with the Bray-Curtis index accounting for species abundances.

328

329 **5. DISCUSSION**

330

331 **5.1. Patterns of species richness and abundance**

332 Mountains are topographically and geologically highly heterogeneous
333 maintaining high biological diversity (Körner 2004; Spehn & Körner, 2005). Moreover,

334 the high anuran species richness and abundance found here may partly stem from the
335 high number of landscapes sampled along the altitudinal gradient.

336 Several empirical studies on anuran species richness in different tropical and
337 subtropical mountainous regions worldwide have documented from 16 to 60 species
338 (Giaretta et al., 1999; Siqueira et al., 2011; Malonza & Veith, 2012; Zancolli et al.,
339 2014; Goyannes-Araújo et al., 2015; Villacampa et al., 2019; Carvalho-Rocha et al.,
340 2021; Siqueira et al., 2021). In our study, we found 70 anuran species at Poços de
341 Caldas Plateau region. This species pool is an important subset of the anuran fauna from
342 the Brazilian Atlantic Forest (625 species **in total** - Rossa-Feres et al., 2017), considered
343 one of the 34 biodiversity hotspots in the world (Myers et al., 2000; Mittermeier et al.,
344 2011) with high anuran species richness and endemism (Haddad et al., 2013; Rossa-
345 Feres et al., 2017). However, direct comparisons with other studies carried out in
346 mountainous regions in Brazil or in other regions around the worldwide are not
347 straightforward because of different survey methods, protocols, and scales (Rahbek
348 2005; Haider et al., 2018).

349 Our finding of a hump-shaped richness pattern along the altitudinal
350 gradient agrees with several empirical studies and recent meta-analysis conducted with
351 different taxonomic groups and regions and indicates that hump-shaped pattern is
352 perhaps the most common richness pattern along mountainsides (Rahbek, 1995;
353 McCain & Grytnes, 2010; Sanders & Rahbek, 2012). On the other hand, despite these
354 similar hump-shaped richness patterns recorded in different mountainous regions with
355 equivalent altitudes worldwide, these mountainsides may vary substantially in their
356 climatic and vegetation features (Safford, 1999a, 1999b), and consequently, can
357 influence altitudinal anuran range size, the peaks of species richness, and community
358 segregation patterns in mountainous regions worldwide.

359

360 **5.2. Rapoport altitudinal rule**

361 Our results also revealed that anuran altitudinal range sizes increased with
362 altitude corroborating Rapoport altitudinal rule (Stevens, 1992). Our finding is
363 consistent with other empirical studies encompassing different taxa such as butterflies
364 (Fleishman et al., 1998), ants (Sanders, 2002), spiders (Chatzaki et al., 2005),
365 harvestmen (Almeida-Neto et al., 2006), and dung beetles (Herzog et al., 2013), which
366 also documented Rapoport altitudinal rule to occur in different regions. We note,
367 however, that this rule is highly controversial (Hawkins & Diniz-Filho, 2006; McCain
368 & Knight, 2013), with many studies disagreeing with the rule (Ribas & Schoereder
369 2006; McCain & Knight, 2013; Wang & Soininen, 2017). Nevertheless, Stevens (1992)
370 did not claim Rapoport's rule (latitudinal or altitudinal) to be valid for all taxa, and the
371 lack of this consensus as a universal rule is a common basis for criticism (Gaston et al.,
372 1998; Ruggiero & Werenkraut, 2007).

373 One of the most common mechanisms proposed by Stevens (1992) to explain
374 Rapoport altitudinal rule is the climatic variability hypothesis or simply Rapoport
375 "rescue" hypothesis. It predicts that species that can withstand broad climatic variability
376 (i.e. have broad climatic niche) can become more widely distributed (Stevens, 1992;
377 Gaston & Chown, 1999). Our results indeed confirmed that annual mean temperature
378 and temperature seasonality were the main drivers on anuran altitudinal range sizes.
379 This is consistent with previous studies that also supported the climatic variability
380 hypothesis with different taxa, such as dung beetles (Gaston & Chown, 1999), ants
381 (Sanders et al., 2003), birds (McCain, 2009a), and moths (Beck et al., 2016).

382 Tropical biodiversity is strongly influenced by climatic stability (Barron, 1995),
383 especially because ectothermic organisms such as anurans are highly dependent on the

384 environment temperatures to maintain their body temperature (Bakken & Angilletta,
385 2014). Temperature has a profound influence on behaviour, physiology, and ecological
386 performance of anurans, as well as on growth rates, metabolism, species diversity, and
387 geographic ranges (Navas, 2006; Navas et al., 2008; Angilletta, 2009). These factors
388 impose selective pressures on the geographical ranges of terrestrial vertebrate (McCain,
389 2009b; Sheldon et al., 2015; Chan et al., 2016), having a direct impact on the extinction
390 probability. According to Whitton et al. (2012) and Pintor et al. (2015), the current
391 climate patterns are consistently the best predictors for amphibian range sizes on a
392 regional scale. Overall, Rapoport altitudinal rule may help to understand the species
393 richness peak at the middle of the altitudinal gradient rather than promoting a monotonic
394 decrease in richness with altitude.

395

396 **5.3. Altitudinal variation in anuran composition**

397 Anuran community composition exhibited a high species turnover revealing a
398 clear segregation pattern along the altitudinal gradient forming three distinct
399 communities. Some species were recorded only in lower (*Boana pardalis*,
400 *Leptodactylus mystaceus*, *Trachycephalus typhonius*), intermediate (*Boana Beckeri*,
401 *Crossodactylus caramaschii*, *Haddadus binotatus*, *Leptodactylus furnarius*,
402 *Physalaemus jordanensis*), and higher altitudes (*Adenomera hylaedactyla*,
403 *Bokermannohyla vulcaniae*, *Trachycephalus imitatrix*, *Ischnocnema gr. lactea*). This
404 emphasize the contribution of altitude and temperature on explaining species
405 distribution in mountainous regions. Such community segregation has been documented
406 in birds (Rahbek, 1997), dung beetles (Davis et al., 1999), gastropods (Presley et al.,
407 2011), small mammals (Andrade & Monjeau, 2014), and anurans (Malonza & Veith,
408 2012) in some tropical montane ecosystems.

409 Some previous empirical studies have revealed that dispersal limitation
410 dominates at regional and large scales (Condit et al., 2002; Tuomisto et al., 2003) while
411 environmental filtering dominates at smaller scales (Davidar et al., 2007; Hardy et al.,
412 2011). Dispersal limitation occurs mainly with organisms less vagile such as anurans
413 studied here (Smith & Green, 2005; Buckley & Jetz, 2007; Qian & Ricklefs, 2012),
414 which corroborates our results. However, the dispersal limitation and environmental
415 filtering can vary with study region and spatial scales (Soininen et al., 2007; Wang et
416 al., 2015). On the other hand, it is worth noting that other factors can also influence
417 anuran community assembly in our study region, such as isolation, complex topography
418 (physical barriers), climatic history and historical, and evolutionary processes (Currie,
419 1991; Brown, 2001; Hawkins et al., 2003; Steinbauer et al., 2016; Laiolo et al., 2018).
420 Finally, anuran community composition from the Poços de Caldas Plateau region
421 revealed a high species turnover and a clear community segregation pattern corroborates
422 the idea that the altitude (dispersal limitation) and temperature (environmental filtering)
423 are important drivers to shape community assembly from the Brazilian Atlantic Forest
424 mountainous regions.

425

426 **6. Implications for conservation**

427 Montane ecosystems are expected to experience high biodiversity losses in the
428 next decades (McCain & Colwell, 2011; Davidson et al., 2013; Peters et al., 2019). This
429 has been occurring mainly because of the intensification of hard anthropogenic
430 activities (Barnosky et al., 2011; Dirzo et al., 2014) as well as an accelerated climate
431 change (Chen et al., 2011; Bellard et al., 2012; Catenazzi, 2015). These negative effects
432 are jointly imposing unprecedented pressures on the world's biodiversity (Díaz et al.,
433 2019) and some taxa are currently experiencing range shifts (Chen et al., 2009; Elsen &

434 Tingler, 2015; Badgley et al., 2017). This may accelerate in the future (Elsen et al.,
435 2018) resulting higher extinction probability. It may be most severe especially in
436 tropical montane ecosystems (McCain & Colwell, 2011; Davidson et al., 2013), and
437 particularly affecting montane species (Elsen and Tingley 2015).

438 Preserving biodiversity in an era considered the sixth mass extinction (Barnosky
439 et al., 2011) will be the greatest challenge of this century (Sala et al., 2000; Cardinale et
440 al., 2012; Ruffell et al., 2017). Our research revealed a high species turnover because
441 several species presented narrow altitudinal range size, resulting in a clear anuran
442 community segregation pattern along the mountainsides from the Poços de Caldas
443 Plateau region. These results demonstrated that, in order to maintain high biodiversity
444 levels, different conservation strategy would be needed in tropical mountainous regions.
445 Montane ecosystems with high species turnover among sites can support several and
446 spatially separated species pools and, therefore, only large enough protected areas can
447 maximize biodiversity conservation (Franklin, 1993). On the other hand, montane
448 ecosystems with high richness difference among sites (that is, high degree of
449 nestedness) would suggest that it is better to protect biodiversity hotspots (Cutler,
450 1994). Our findings revealed that beta diversity patterns in mountainous regions also
451 have important implications for species conservation. Furthermore, our rarefaction
452 curves highlight that the distributions of common, rare, endemics, and threatened
453 species all deserve attention, and conservation plans need to encompass the entire
454 mountainous region and not just total diversity hotspots (Orme et al., 2005; Grenyer et
455 al., 2006) or simply only one altitudinal zone.

456 Some studies have reported that the new protected areas also need to focus on
457 species' range shifts in order to minimize the effects of climate change on species
458 distribution (Hannah, 2010; Mawdsley, 2011; Lemes and Loyola, 2013). Small-ranged

459 species constitute most to Earth's species diversity (Gaston, 2003) and maybe overall
460 more vulnerable and threatened by climate change effects, thus presenting higher risk of
461 extinction (Chen et al., 2009; Davies et al., 2009; Bellard et al., 2012; Elsen et al.,
462 2018). Our findings revealed that protecting the whole montane ecosystem (lower,
463 intermediate, and higher altitudes) is important to preserve the whole regional species
464 pool, species with narrow range sizes, and diverse communities in Tropical
465 mountainous regions. Therefore, studies on Rapoport altitudinal rule may not only help
466 us to understand the variability in species altitudinal range sizes, but also help to
467 conserve the species that occupy different narrow altitudinal ranges. To understand the
468 species richness spatial patterns and the potential climate change effects are crucial to
469 target conservation actions properly in the long-term (Zhang et al., 2012; Zancolli et al.,
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471

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486

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490

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492

493 **Data availability statement:** The data supporting the findings of this study are
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495

496 **7. REFERENCES**

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985 **SEPARATE FIGURE LEGEND:**

986 Figure 1. Hypothetical relationships of species richness and community composition in
987 response to the topographic and climatic variables along an altitudinal gradient in
988 Poços de Caldas Plateau region, Southeastern Brazil. In panel [a] we hypothesize
989 that species richness will decline with altitude as a consequence of temperature
990 reduction; in [b] we hypothesize that in the intermediate altitudes, we will detect
991 higher richness, because this range will share species from both lower and higher

992 altitudes; in [c] we hypothesize that species that are found in higher altitudes will
993 also have higher plasticity in their thermal tolerance, thus being also found in lower
994 altitudes; in [d] we present an alternative hypothesis, where community
995 composition differs among lower, intermediate, and high altitudes with highest
996 richness in intermediate altitudes, as presented in [b].

997 Figure 2. Study area covering 25 sampled landscapes distributed along an altitudinal
998 gradient in Poços de Caldas Plateau region, Southeastern Brazil. White line
999 indicates the border between Minas Gerais (right) and São Paulo (left) States.
1000 Yellow circles (buffers 1 km) represent the sampled landscapes.

1001 Figure 3. Individual-based rarefaction curve on rarefied (interpolated and extrapolated)
1002 richness along an altitudinal gradient in Poços de Caldas Plateau region,
1003 Southeastern Brazil.

1004 Figure 4. Altitudinal range size distributions of anurans in Poços de Caldas Plateau
1005 region, Southeastern Brazil. Bars show altitudinal range size along the altitudinal
1006 gradient. Square: median abundance. The range profiles were based on the lowest
1007 and highest elevation at which a species was observed.

1008 Figure 5. Principal Coordinates Analysis (PCoA) ordination of anurans communities'
1009 composition along an altitudinal gradient in Poços de Caldas Plateau region,
1010 Southeastern Brazil. Ordination was based on dissimilarity matrix calculated with
1011 the Bray-Curtis index accounting for species abundances.

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