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15	Interactive effects of a bacterial parasite and the insecticide carbaryl to life-
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38 Abstract

39 Natural and chemical stressors occur simultaneously in the aquatic environment. Their combined 40 effects on biota are usually difficult to predict from their individual effects due to interactions between 41 the different stressors. Several recent studies have suggested that synergistic effects of multiple 42 stressors on organisms may be more common at high compared to low overall levels of stress. In this 43 study, we used a three-way full factorial design to investigate whether interactive effects between a 44 natural stressor, the bacterial parasite Pasteuria ramosa, and a chemical stressor, the insecticide 45 carbaryl, were different between two genetically distinct clones of Daphnia magna that strongly differ in 46 their sensitivity to carbaryl. Interactive effects on various life-history and physiological endpoints were 47 assessed as significant deviations from the reference Independent Action (IA) model, which was 48 implemented by testing the significance of the two-way carbaryl x parasite interaction term in two-way 49 ANOVA's on log-transformed observational data for each clone separately. Interactive effects (and 50 thus significant deviations from IA) were detected in both the carbaryl-sensitive clone (on survival, 51 early reproduction and growth) and in the non-sensitive clone (on growth, electron transport activity 52 and proPhenolOxidase activity). No interactions were found for maturation rate, filtration rate, and 53 energy reserve fractions (carbohydrate, protein, lipid). Furthermore, only antagonistic interactions were 54 detected in the non-sensitive clone, while only synergistic interactions were observed in the carbaryl sensitive clone. Our data clearly show that there are genetically determined differences in the 55 56 interactive effects following combined exposure to carbaryl and Pasteuria in D. magna.

57

58 Keywords: mixture, interactive effects, bacterial parasite, carbaryl, Daphnia magna

59 1. Introduction

60 The study of combined effects of multiple chemical stressors is becoming increasingly important in ecotoxicology. This is because the toxicity of a given mixture of chemical stressors can usually not be 61 predicted in a straightforward way from the toxicity of the different individual stressors in that mixture 62 63 due to non-additive (i.e., interactive) effects. This considerably complicates environmental risk 64 assessment of chemical mixtures (Van Gestel et al., 2010). In addition, chemical stressors can also 65 interact with (biotic and abiotic) 'natural' stressors. It is well-documented that 'natural' stressors such as temperature and food limitation may modify the effects of chemicals on organisms and vice versa 66 67 (see recent reviews of Heugens et al., 2001; Holmstrup et al., 2010; Laskowski et al., 2010). A meta-68 analysis of interactions between natural stressors and toxic chemicals in 61 studies by Laskowski et 69 al. (2010) showed a significant interaction in 62.3% cases, indicating the importance of the occurrence 70 of such interactions in natural ecosystems. Moreover, these authors showed that the null hypothesis assuming no interactions between chemical and natural stressors should be rejected at $p=2.7\times10^{-82}$. 71 72 The review by Holmstrup et al. (2010) evaluating the interactive effects of binary combinations of 73 natural and chemical stressors as reported in more than 150 studies (covering natural stressors 74 including heat, cold, desiccation, oxygen depletion, pathogens and immunomodulatory factors) 75 revealed similar results. In this set of studies, synergistic interactions, i.e. with the effect of the 76 combination of two stressors being stronger than expected based on their non-interactive combined 77 action, were reported in more than 50% of the cases. These authors also report antagonistic 78 interactions, i.e. were the effect of combined stressors is smaller than expected, but these interactions 79 were found in much fewer cases. Holmstrup et al. (2010) also pointed out that synergistic effects of 80 chemical and natural stressors appear to be more likely with increasing levels of stress caused by one 81 or both stressors. The aim of the present study was to start testing this hypothesis from a slightly 82 different angle by investigating whether a clone of the water flea Daphnia magna that is more sensitive to a given chemical, and thus experiences a higher level of stress, would also experience more 83 84 pronounced synergistic effects during a combined exposure to a natural stressor and that chemical 85 compared to a less sensitive clone.

We chose the insecticide carbaryl and the bacterium *Pasteuria ramosa*, a bacterial endoparasite of *D. magna*, as our model system for a combined analysis of a chemical and natural stressor. Earlier work

found synergistic effects for these two stressors in *D. magna*. Coors et al. (2008) and Coors and De Meester (2008, 2011) exposed a single clone of *D. magna* to the insecticide carbaryl and *P. ramosa* and found that sublethal concentrations of carbaryl enhanced the virulence of the parasite: i.e. sterilization of *D. magna* by *P. ramosa* was accelerated under carbaryl exposure. In addition, Jansen et al. (2011a) showed in an experimental evolution trial that the evolution of increased resistance to the pesticide carbaryl resulted in an increased susceptibility to infection by *P. ramosa*.

94 We performed a 10-day exposure experiment according to a full-factorial 2 x 2 x 2 design, using two 95 D. magna clones (one clone sensitive to carbaryl, denoted 'S', and one clone non-sensitive to carbaryl, denoted 'NS'), P. ramosa (absence vs. presence) and carbaryl (absence vs. presence) as factors. 96 97 During this experiment we recorded several life-history endpoints (survival, growth and early 98 reproduction). In addition, we included several physiological endpoints such as filtration rate, energy 99 reserves, electron transport system activity, acetylcholinesterase- and phenoloxidase activity. 100 Including these endpoints does not only broaden the set of endpoints but also may help in pinpointing 101 mechanistic causes of interaction effects. Three-way and two-way ANOVA on log-transformed 102 observational data were then used to test for interactive effects between carbaryl and P. ramosa on all 103 recorded D. magna endpoints and to test whether interactions differed between the sensitive (S) and 104 non-senstive (NS) clone.

105

106 2. Material and methods

107 2.1. Organisms and stressors

Daphnia magna is a planktonic cyclic parthenogenetic crustacean and a keystone species in
freshwater lakes and ponds (Lampert, 2011; Stollewerk, 2010). It is a frequently-used model organism
in ecotoxicology (Altshuler et al., 2011) and for host-parasite studies (e.g. Ebert et al., 2004;
Decaestecker et al., 2007; Coors et al., 2008).

The gram-positive bacterium *Pasteuria ramosa* is an obligate endoparasite of *D. magna* that irreversibly sterilizes its host within 5 to 15 days after infection (Ebert, 2005). The energetic resources that become available through suppression of reproduction are channeled towards the production of new parasite endospores, which can infect new hosts through horizontal transmission from decaying hosts (Ebert et al., 2004). Susceptibility to *P. ramosa* may depend on genetically and environmentally
determined host immunity (Little and Ebert, 2000).

The methyl carbamate insecticide carbaryl is a model substance that is representative for insecticides with mode of action class 1a, i.e. carbamate acetylcholinesterase inhibitors, according to the Insecticide Resistance Action Committee (http://www.irac-online.org/eClassification). Carbaryl acts as a quasi-irreversible inhibitor of acetylcholinesterase, an enzyme which hydrolyses the neurotransmitter acetylcholine. Inhibition of acetylcholinesterase results in the accumulation of acetylcholine at the postsynaptic receptor, which results both in repetitive firing and blocking of other neuronal transmissions (Corbett et al., 1984).

125

126 2.2. Experimental design

127 Two different *D. magna* clones with a known difference in their sensitivity to carbaryl (based on earlier experiments, Jansen et al., 2011a), and further denoted as clone S (sensitive) and NS (non-sensitive), 128 129 were cultured parthenogenetically under controlled laboratory conditions (20°C ± 1°C, 16:8h light:dark 130 cycle; 1000 lux) for multiple generations prior to the experiment. The chemically defined ADaM medium (Klüttgen et al., 1994) was used as both the culture and the test medium. Stock cultures as 131 well as experimental animals were fed daily with 2×10⁵ cells per mL of the green alga 132 Pseudokirchneriella subcapitata, corresponding with 1.25 mgC·L⁻¹. Both clones originated from the 133 dormant egg bank of a pond in Oud-Heverlee Zuid, Belgium (50°50'22" N, 4°39'18" E), also described 134 135 by Coors et al. (2009). A three-way full factorial experiment was conducted with parasite challenge 136 (absent or present), carbaryl exposure (absent or present) and clone (S or NS) as factors, resulting in 137 four exposure treatments per clone. Three independent replicates of 320 animals per treatment were 138 set-up in 10 L glass aquaria holding ADaM medium using pooled second to fourth brood juveniles 139 (<24h old). The population density of the daphnids was maintained at one individual per 5 mL medium 140 during the first four days of the exposure and then changed to one daphnid per 30 mL until the end of 141 the experiment (day 10) by adapting the volume in the aquaria. The densities used during the 142 exposures are realistic for the field, where densities of >300 individuals/L can be observed (add refs). 143 The exposures took place under diffuse light conditions (40cd, 16:8h light:dark cycle) and under 144 controlled temperature conditions (20 °C ± 1 °C). The medium was renewed every other day. Temperature (mean \pm SD: 19.3 \pm 0.4 °C), oxygen concentration (mean \pm SD: 9.05 \pm 0.27 mg L⁻¹), pH (mean \pm SD: 7.63 \pm 0.10) and conductivity (mean \pm SD: 892 \pm 20 μ S cm⁻¹) did not differ systematically among treatments or replicates. The experiment was terminated after 10 days of exposure. At this point in time, most animals had released their first brood in the control treatment, allowing a reliable assessment of effects on early reproduction.

150

151 2.3. Stressor exposures

152 2.3.1. Parasite challenge

153 D. magna neonates from an isoclonal stock culture of clone K6 (originating from a pond in Kiel, 154 Antwerp, Belgium and cultured in our laboratory in Ghent for over 20 years) were exposed to sediment from a pond in Knokke, Belgium (Knokke In, 51°20'6"N, 3°20'54"E), which is known to contain P. 155 156 ramosa spores (Jansen et al., 2010). After 22 days the infected hosts were collected and ground. The 157 resulting suspension was filtered over a 60 µm nylon filter (Millipore) and then diluted with deionized water to a concentration of 5×10⁶ spores mL⁻¹. A placebo-suspension for parasite-free treatments was 158 159 prepared in the same way by grounding the same amount of uninfected stock culture daphnids, in such a way that it contained an equal weight of ground daphnia tissue per mL. Prior to challenging the 160 161 daphnids with the bacterial spores, the suspension was examined under a phase-contrast microscope 162 at a 400x magnification to determine the presence of spores from other parasites that may have been present in the sediment. Only P. ramosa spores were observed. Daphnids were challenged with 163 3.75×10⁴ mature *P. ramosa* spores per mL medium during the first six days of the experiment. More 164 specifically, spores were added to fresh medium at the start of the experiment (day 0) and at the time 165 166 of media renewals, i.e. on day two and day four (Jansen et al., 2011b). All parasite-free treatments 167 received the same amount of placebo solution. No spores or placebo-solution were added to the 168 medium later on in the experiment.

169

170 2.3.2. Pesticide challenge

Daphnids were challenged with 8 µg·L⁻¹ carbaryl (1-naphthyl methylcarbamate, 99.8% purity, Sigma-171 172 Aldrich) during the first six days of the exposure. Carbaryl was added to fresh medium at the start of 173 the experiment (day 0) and at the time of the media renewals. The US Environmental Protection 174 Agency report on the ecological risk assessment of carbaryl reports measured surface water concentrations of carbaryl up to 5.5 μ g L⁻¹ and estimated peak concentrations ranging between 23 and 175 153 µg·L⁻¹ (US EPA, 2003), pointing to the environmental relevance of the carbaryl concentration 176 177 used in the present study. Carbaryl stock solutions were prepared in ethanol and the ethanol 178 concentration in the exposure was set to the same level in all treatments, including treatments without carbaryl (50 µL·L⁻¹). Three mixed samples of 250 mL of each fresh and 48-hour old medium (i.e. 179 sample taken immediately after transferring daphnids to fresh media) separately were taken and 180 181 stored in brown glass bottles at -20°C for later verification of carbaryl concentrations. Analysis of 182 carbaryl concentrations was done by GC-MS (Trace GC 2000 series, Thermoquest; Polaris, Finnigan/Thermoguest) on an apolar SLB[™]-5ms column (Supelco, Sigma-Aldrich). Extraction and 183 elution was performed on Solid Phase Extraction according to the manufacturer's application notes 184 (Waters and Phenomenex). Propoxur was used as the internal standard at a concentration of 4 µg·L⁻¹ 185 186 to control and correct for extraction losses. Recovery was always >90%. Proxopur belongs to the same functional class of pesticides as carbaryl, i.e. the carbamates. Immediately before injection of 187 the sample, a recovery standard was also applied, to control for the injection itself. The carbaryl 188 concentration was 8.85 \pm 0.21 μ g·L⁻¹ (mean \pm standard deviation) in freshly prepared medium and 189 $6.57 \pm 0.40 \ \mu g L^{-1}$ in 48-hour old medium (immediately after renewal). 190

191

192 2.4. Life-history endpoints

Maturation rate is reported as the percentage of egg-carrying individuals on day 8 of the exposure. No offspring were released from the brood pouches before this day in none of the experimental cultures. Investment in early reproduction is reported as the number of offspring produced between media renewals on day 8 and day 10 divided by the number of egg-carrying individuals counted on day 8. Body length on day 10 of six to eight animals was measured from the top of the head to the base of the spine by analyzing a microscopic image with the Image Tool 3.0 software (San Antonio, TX, USA). 199

200 2.5. Physiological endpoints

201 2.5.1. Feeding rate

Filtration rate was measured at the end of the experiment (day 10) according to the method described in Muyssen et al. (2006) with minor modifications. Three replicates of one individual daphnid per treatment were set up and three 'blancs' without daphnids (but with algal food added) per treatment were used to be able to account for algal growth when calculating filtration rate. The algal concentrations (*P. subcapitata*) were measured using a Coulter Counter (Z1 Coulter Particle Counter, Beckman Coulter) at the beginning of the feeding period and after 24 hours.

208

209 2.5.2. Energy reserves and electron transport system activity

210 Energy reserves were measured on day 10 as three separate energy fractions: protein, lipid and 211 carbohydrate content of the organisms. For each fraction seven daphnids were collected and flash-212 frozen in liquid nitrogen on day 10. Samples were stored at -80°C until analysis. The different fractions 213 were measured spectrophotometrically in triplicate and transformed into energetic equivalents as 214 described in De Coen and Janssen (1997). The energy consumption was estimated by measuring the electron transport system (ETS) activity at the mitochondrial level as described in De Coen and 215 216 Janssen (1997). ETS activity was measured as an alternative to oxygen consumption measurements 217 as these could not be performed due to a broken probe.

218

219 2.5.3. Acetylcholinesterase activity

Pools of seven flash-frozen daphnids collected on day 10 were homogenized in 0.02M ice-cold sodium hydrogen phosphate buffer (PB), pH 8.0, containing 1% Triton-X-100 (Sigma-Aldrich) with a motordriven Teflon pestle for 45s. Ice-cold PB (without Triton-X-100) was added to the initial homogenate in a 10:1 ratio. The final homogenates were mixed and centrifuged at 3000g at 2-4°C for 10min. Supernatants were collected in a clean, pre-cooled Eppendorf tube and assayed immediately. The enzyme activity was determined in triplicate for each sample according to the colorimetric method described by Ellman et al. (1961). Briefly, 100 μL of 8 mM 5-5'-dithiobis-2-nitrobenzoate (DTNB)
(Sigma-Aldrich) in PB supplemented with sodium hydrogen carbonate (Sigma-Aldrich) at 0.75 mg·mL⁻¹
and 50 μL of supernatant were added to a 96-well microtiter plate. Measurement of enzyme activity
was initiated by adding 50 μL of 16 mM acethylthiocholine iodide (Sigma-Aldrich) in PB. Spontaneous
hydrolysis of the substrate was assessed using a blank in triplicate, containing PB with 0.1% Triton-X100 instead of the supernatant. After an incubation period of 10 minutes at 20°C, absorbances at 405
nm and 20°C were measured every 60s during 10min with intermittent shaking.

The enzyme activity was expressed in nmol·min⁻¹·mg⁻¹ as activity = $(\Delta OD/min) / (\varepsilon \times I \times C)$ where $\Delta OD/min$ is the change in optical density per minute (min⁻¹), ε is the molar extinction coefficient of DTNB (= $1.34 \cdot 10^6$ nM⁻¹cm⁻¹), *I* is the length of the light path (cm), *C* the protein concentration in the supernatans (mg·L⁻¹).

Protein concentration in the homogenate supernatant was determined using the Bradford method
(Bradford, 1976), with bovine serum albumin (Sigma-Aldrich) as a standard.

239 Quality control of the assay was assessed using a quality control enzyme standard of electric eel 240 cholinesterase (Sigma-Aldrich) in ice-cold PB containing 1 mg·mL⁻¹ bovine serum albumin (Sigma-241 Aldrich). The reaction rate of the quality control enzyme was confirmed at a change of 55-60 242 mOD·min⁻¹.

243

244 2.5.4. Phenoloxidase activity

A major component of the invertebrate innate immune system is the prophenoloxidase (proPO) activation system, providing immunity against a large range of pathogens (Soderhall and Cerenius, 1998; Cerenius et al., 2008). Upon infection, the inactive proenzyme proPO is activated and transformed into the active form phenoloxidase (PO), which oxidizes phenols and thus leads to the formation of melanin, which is believed to play an important role in encapsulation and neutralization of bacteria (Soderhall and Cerenius, 1998).

251 Measurement of phenoloxidase (PO) activity normally uses extracted haemolymph as described by 252 Mucklow and Ebert (2003). However, because carbaryl-treated daphnids of clone S at the end of the

253 exposure were too small to extract sufficient amounts of haemolymph, we choose to use real-time qPCR gene expression analysis of the proPO gene as an alternative. We used the method described 254 255 by Labbé and Little (2009). Daphnids collected on day 10 were shock-frozen in liquid nitrogen prior to 256 total RNA isolation. Total RNA isolation was performed using RNeasykit and Qiashredder kit (Qiagen) 257 following manufacturer's instructions. Contaminating DNA was removed by a DNAse treatment 258 (Qiagen). Prior to cDNA transcription, RNA quality and quantity were determined with a Nanodrop 259 spectrophotometer. RNA aliquots for reverse transcriptase were stored at -80°C and afterwards 260 reverse transcribed to cDNA using 1 µg of RNA and the MessageAmpTM II mRNA Amplification kit 261 (Applied Biosystems) according to manufacturer's protocol. Only first strand cDNA synthesis was 262 performed. Sample quality and yield were again assessed using the Nanodrop spectrophotometer. Samples were stored at -20°C until qPCR analysis, which was performed on a Corbett RotorGene 263 264 3000 during 45 cycles (30s at 95°C; 30s at 58°C; 35s at 72°C). Further qPCR analysis was performed 265 as described in Labbé and Little (2009).

266

267 2.6. Data treatment and statistical analyses

268 In all statistical tests performed, all data were balanced, i.e. an equal number of replicate observations 269 was available for each treatment for each endpoint. All statistics were performed with Statistica 7.0 270 software (Statsoft, Tulsa, OK, USA). All endpoints were log₁₀-transformed prior to statistical analysis to 271 ensure compliance with assumptions of normality (Shapiro-Wilkinson's W test) and homoscedasticity 272 (Levene's test) for all endpoints. This transformation also allowed us to interpret findings of a 273 statistically significant carbaryl x parasite interaction term in a two-way ANOVA as a statistically 274 significant deviation from the independent action (IA) model of joint stressor effects (Sih et al., 1998; 275 Fournier et al., 2006; see further).

First, we performed three-way ANOVA to determine the significance of the main effects and two-way and three-way interaction terms for all endpoints. All analyses were performed at a significance level of 95% (p < 0.05). Of particular interest were findings of significant clone × carbaryl interaction (confirming a different effect of carbaryl between the two clones) and significant three-way clone × carbaryl × parasite interaction. While three-way interactions can be interpreted in different ways (Kutner et al., 2005) one possible interpretation in our study is that it indicates that the carbaryl ×

parasite interaction is different between the two clones, which is exactly what we wanted to test in relation to the aims of our study. Therefore we also performed a more detailed analysis of the carbaryl x parasite interaction with two-way ANOVA's for each clone separately to aid the validation of such an interpretation (e.g., if this interaction would be significant in one clone but not in the other). At the same time, the same two-way ANOVA analysis provided a formal statistical test of the independent action (IA) model (see below).

Second, we investigated for each endpoint and for each clone separately if the effect observed in the combined P+C treatment followed the Independent Action (IA) model. This is the recommended reference model for predicting combined effects of dissimilarly acting stressors (Jonker et al., 2004) and thus the logical choice in the present study based on the biologically fundamentally different mechanisms of action of carbaryl and *Pasteuria* infection in *Daphnia* (Coors et al., 2008). This model, originally formulated by Bliss (1939), predicts combined effects of binary stressors from observed effects in the individual stressor treatments as follows (Faust et al., 2003):

- 295 $E_{PC,predicted} = E_P + E_C E_P \times E_C (Eq. 1)$
- 296 where

297
$$E_i = (Y_{control} - Y_i) / Y_{control} (Eq. 2)$$

298 with E_i the observed fractional effect of treatment *i* on endpoint Y relative to the control treatment, 299 where *i* is either P (*Pasteuria*), C (carbaryl) or PC (combined *Pasteuria* + carbaryl treatment). It should 300 be noted that E_i can be both positive (in case of a decrease of the endpoint compared to the control) 301 and negative (in case of an increase of the endpoint compared to the control). Algebraically 302 combining Eq. 1 and Eq. 2, also allows predictions of the value of each endpoint in the combined carbaryl + parasite treatment (Y_{PC}, depicted in Figures 1 and 2), based on the arithmetic mean of the 303 304 values observed in the control (Y_{control}), the carbaryl only treatment (Y_C) and the parasite only 305 treatment (Y_P) :

306 $Y_{PC, predicted} = Y_P \times Y_C / Y_{control}$ (Eq. 3)

307 The actual statistical testing of the hypothesis of independent action for each clone separately was 308 implemented by determining the significance of the *Pasteuria* × carbaryl interaction term in the 2-way 309 ANOVA's carried out on log₁₀-transformed observational data for each clone separately. In other words, a significant interaction term at the 95% significance level (p < 0.05) found with this ANOVA 310 311 implies a statistically significant deviation from IA. This approach has infrequently been used in the 312 field of ecotoxicology for testing deviations from the IA model for binary chemical stressor mixtures. In 313 contrast, it is already being used for more than a decade in the field of ecology for detecting significant 314 departures from independent action of binary combinations of ecological stressors, for instance prey 315 stressed by two predators (review by Sih et al., 1998) or plants stressed by two parasites (Fournier et al., 2006). 316

317 Third, when the 2-way ANOVA revealed a statistically significant Pasteuria x carbaryl interaction, we 318 classified it as synergistic if the observed effect in the combined treatment was 'higher' than the effect predicted with the IA model (Eq. 1) (Faust et al., 2003). In terms of the calculations made with Eq 1 319 320 and Eq 2., this occurs if E_{PC,observed} > E_{PC,predicted} in cases where E_{PC,observed} > 0 (i.e., where the 321 combined treatment causes a reduction of the endpoint compared to the control, e.g. survival, see Fig. 322 1A) or if E_{PC,observed} < E_{PC, predicted} in cases where E_{PC,observed} <0 (i.e., where the combined treatment 323 causes an increase of the endpoint compared to the control, e.g. proPO expression, see Fig. 2G). 324 When the observed effect was 'smaller' than the predicted effect, i.e. if E_{PC.observed} < E_{PC. predicted} in 325 cases where E_{PC,observed} > 0, or if E_{PC,observed} > E_{PC, predicted} in cases where E_{PC,observed} <0, the interaction 326 was classified as antagonistic.

327

328 **3. Results**

Results for all measured endpoints are presented in Figures 1 and 2. Results of three-way ANOVA analyses are given in Table 1 and Table 2 for life-history and physiological endpoints, respectively. Table 1 and Table 2 also contain results of the two-way ANOVA analysis of the parasite × carbaryl interaction for clones NS and S separately. Complete two-way ANOVA results are listed in Table S1 and Table S2 in supplementary material.

Main effects of clone, parasite and carbaryl were detected in most endpoints, with few exceptions (Tables 1 and 2). Sterilization already reached 100% on day 10 in all parasite and parasite + carbaryl treatments in both clones, thus making any further testing of parasite × carbaryl interactionsimpossible for this endpoint.

338

339 3.1. Life history endpoints

With three-way ANOVA, significant clone × carbaryl interactions were observed for all life-history endpoints, showing stronger reductions of survival, investment in early reproduction, body length and maturation rate after carbaryl exposure in the carbaryl sensitive clone S than in the non-sensitive clone NS (Figure 1, Table 1). No significant clone × parasite interactions were observed for any of the lifehistory endpoints (Table 1). Our observations on the variables scored indicate that the two studied clones show differences in their sensitivity towards carbaryl but not towards the parasite.

346 Significant three-way clone x carbaryl x parasite interactions suggest that there are clonal differences 347 in carbaryl x parasite interactions for three of the four tested life-history endpoints (i.e. survival, 348 investment in early reproduction and body length) (Table 1, Figure 1A, 1B, 1C). No significant three-349 way interaction was observed for maturation rate (Table 1, Figure 1D). Detailed follow-up analyses of 350 the three significant three-way interactions with two-way ANOVA indicated that no interactive effect 351 between parasite and carbaryl on survival was found for clone NS, while a synergistic interaction was 352 detected for clone S (Table 1; Figure 1A). Clone S also showed a synergistic parasite x carbaryl 353 interaction effect for early reproduction, while clone NS did not (Table 1, Figure 1B). Finally, for body 354 length an antagonistic interaction was observed for clone NS (Table 1, Figure 1C), while a synergistic 355 interaction was detected for clone S (Table 1, Figure 1C).

356

357 3.2. Physiological endpoints

With three-way ANOVA, differences in response to carbaryl among the two clones were observed for some measured physiological endpoints, with significant clone × carbaryl interaction terms for protein and carbohydrate reserves (Table 2). While carbaryl has no effect in clone NS, it has a strong negative effect on the total protein and carbohydrate reserves in clone S (Figure 2B and C). In addition, twoway ANOVA revealed significant clone × carbaryl interactions in the absence of parasites for ETS (p < 363 0.001) and AChE (p = 0.01). For the latter two endpoints carbaryl has a strong positive effect in clone 364 S, while no effect was detected in clone NS (Figure 2E and F). No clone x carbaryl interaction was detected with three-way ANOVA for filtration rate (Figure 2A), lipid reserves (Figure 2D) and proPO 365 366 expression (Figure 2G). Together, our observations indicate that carbaryl elicits a very different 367 response of the physiological endpoints studied in both clones, with an overall more pronounced effect 368 in clone S. None of the measured physiological endpoints showed significant three-way interactions. 369 However, in two-way ANOVAs carried out for both clones separately, a significant, antagonistic 370 parasite x carbaryl interaction was detected in clone NS for the proPO expression endpoint, while no 371 interaction was detected for clone S (Table 2). This suggests a tendency for differences in the 372 response to a combined effect of parasite and carbaryl exposure between these two clones for this 373 endpoint. No carbaryl × parasite interactions were observed with two-way ANOVA for ETS and AChE 374 in clone S, while a significant, antagonistic carbaryl x parasite interaction was detected for ETS in 375 clone NS.

377 4. Discussion

378 Susceptibility to the adverse effects of parasite infection can increase with host environmental stress (Gerard et al., 2008). Evidence of chemical stressors interacting with parasites is mounting, but mostly 379 380 limited to vertebrate species (Holmstrup et al., 2010). Kramarz et al. (2007) showed that the snail 381 Canthareus aspersus exposed simultaneously to cadmium and the nematode Phasmarhabditis 382 hermaphrodita accumulated cadmium to higher concentrations than control snails. Scarab grubs 383 (Cyclocephala hirta and C. pasadenae) exposed to a combination of a biopesticide and nematodes 384 showed additive or greater than additive mortalities (Koppenhöffer and Kaya, 1997). Synergistic 385 interactions are also reported for the pesticide imidacloprid applied together with entomopathogenic 386 nematodes in white grubs (C. hirta, C. borealis and Popillia japonica) in (Koppenhöffer et al., 2000). 387 Finally, Cuthbertson et al. (2003) showed increased mortality of sweet potato whitefly larvae after 388 exposure to a combination of imidacloprid and the nematode Steinernema feltiae. However, none of 389 these studies investigated possible genotype-based differences in chemical x pathogen interactions 390 between different genotypes of the same species that differ in their sensitivity to one of the stressors. 391 One study by Salice and Roesijadi (2002) points in the direction of such differences: they found higher 392 mortality due to cadmium in a parasite-resistant strain of the freshwater snail Biomphalaria glabrata 393 compared to a parasite-susceptible strain. Yet, these authors did not expose both strains to a 394 combination of cadmium and the parasite, making the assessment of interactions between both 395 stressors impossible.

396 To be able to test whether the carbarayl-sensitive clone would experience different interactive effects 397 between the insecticide carbaryl and the bacterial parasite P. ramosa compared to the less sensitive clone, we needed to verify first if one clone was indeed more sensitive to carbaryl than the other and 398 399 whether both clones were overall equally sensitive to the parasite. For life-history endpoints, the three-400 way ANOVA showed clear clone x carbaryl interactions (Table 1). Clonal differences of daphnids in 401 sensitivity to various stressors such as pesticides have been shown before (e.g. Calow et al., 1990; 402 Warming et al., 2009). Both clones investigated here are affected by the parasite, with no significant clone x parasite interaction for the studied life-history endpoints (Three-way ANOVA, Table 1). 403 404 Maturation rate decreased, while 'investment in early reproduction' increased following parasite 405 exposure in both clones (Figure 1B and D). In other words, fewer animals reached maturity in the 406 parasite treatment after 8 days of exposure (compared to the control), but those that did reach maturity 407 produced more juvenile offspring per animal on average. The observed decrease in maturation rate is 408 most likely a direct effect of the parasite infection process (Ebert, 2005), while increased investment in 409 the first brood has been described before as an adaptive defensive mechanism of *Daphnia* to 410 sterilizing parasites (Ebert, 2005; Hall et al., 2007). The above observations in the three-way ANOVA's 411 of the presence of a clone × carbaryl interaction and the absence of a clone × parasite interaction for 412 the observed life-history endpoints, justifies the choice made regarding the clones to work with.

413 Overall, we observed strong differences in carbaryl x parasite interactions among the two studied 414 clones for life history endpoints (Two-way ANOVA's, Table 1). First, while no interaction between 415 parasite and carbaryl for survival was found for the less sensitive clone, a synergistic interaction was 416 detected with the sensitive one. Coors et al. (2008) and Coors and De Meester (2008) also found a 417 synergistic interaction between parasite and carbaryl for survival in another D. magna clone. Coors et 418 al. (2008) also found a synergistic interaction on sterilization at day 10, but, as noted earlier, 419 interactive effects on this endpoint could not be assessed in the present study because sterilization 420 had already reached 100% in all parasite and combined treatments at day 10. Second, differences in 421 interactions between both clones were also noted for early reproduction: a synergistic effect was 422 detected for the carbaryl-sensitive clone while no significant interaction was detected for the less 423 sensitive clone. Third, an antagonistic interaction was observed for body length for the less carbaryl 424 sensitive clone, while a synergistic interaction was detected for the sensitive clone. In summary, only 425 synergistic interactions were found for the carbaryl sensitive clone while either no or anatagonistic 426 interactions were observed for the less sensitive clone. The latter is interesting, as Holmstrup et al. 427 (2010) note in their review that all interactions between pathogens and chemicals reported so far are synergistic and none were antagonistic. The difference with our study might be related to a different 428 429 exposure scenario: exposures to the stressors in the studies reported in the review by Holmstrup et al. 430 (2010) generally lasted for a longer period than in our study. In our study, exposure to both P. ramosa 431 spores and carbaryl ended on day 6, while endpoints were recorded on day 10.

With respect to the measured physiological endpoints, differences in response to carbaryl were detected among the two clones with respect to protein and carbohydrate reserves (Three-way ANOVA's, Table 2), confirming the difference in sensitivity to carbaryl between the two clones. Mobilization and shifts in energy reserves have previously been reported after exposure to different 436 stressors, including pesticides (Calow, 1991; De Coen and Janssen, 2003; Nath et al., 1997). 437 Interestingly, parasite exposure also induced significant shifts in energy reserves in both clones 438 (Figure 2B, 2C, Table 2, three-way ANOVA): carbohydrate reserves increased and protein reserves 439 decreased. This shift could be related to the parasite channeling and using different forms of reserves 440 of the Daphnia into its own development (Ebert, 2005; Ebert et al., 2004). The strong positive effect of 441 carbaryl in the sensitive clone on ETS reflects an increase in energy demand (Figure 2E). This 442 suggests an increased investement of energy in mechanisms to cope with the effect of carbaryl on the 443 organism. The two clones also differed in the activity of acetylcholinesterase (AChE) upon carbaryl 444 exposure. Carbaryl acts as a guasi-irreversible inhibitor of AChE. However, in contrast to what would 445 be expected based on this inhibition, an increase in AChE activity was observed for the carbaryl 446 sensitive clone (Figure 2F). This could be an indirect effect, related to the negative effect carbaryl had 447 on the body length of this clone. This is supported by a recent review by Domingues et al. (2010) which considers size as an important factor affecting AChE, with smaller individuals of species, 448 449 including Daphnia magna and Daphnia similis, generally exhibiting higher AChE activity than their 450 larger conspecifics. In addition, Xuereb et al. (2009) reported a strong negative correlation between 451 AChE activity and body weight for the aquatic invertebrate Gammarus fussarum. In a similar context, 452 Chandrasekara and Pathiratne (2007) reported body size-related differences in the inhibition of brain 453 AChE activity in juvenile tilapia. Similar to our observations for life-history endpoints, we did not detect 454 any significant clone x parasite interactions for the physiological endpoints with the three-way 455 ANOVA's, indicating that both clones responded in a similar way to P. ramosa infection. We detected a significant parasite x carbaryl interaction for proPO expression (an antagonistic one), but only in the 456 457 less carbaryl-sensitive clone (Table 2, Two-way ANOVA). The latter observation again points to 458 differences in parasite x carbaryl interactions between both clones.

Interactions between genotype and the response to two different environmental stressors have not often been studied yet. Muyssen et al. (2010) studied interactions between temperature and cadmium on both life-history and physiology in three different *D. magna* clones. They detected significant clone × cadmium × temperature interaction effects for two out of three endpoints for which there was a difference in cadmium susceptibility among the three clones, which was reflected by a significant clone × cadmium interaction term. In other words, they too found that stressor interactions (cadmium ×

465 temperature) were different among clones that differed in susceptibility to one of these stressors466 (cadmium).

467 Different physiological mechanisms may underlie the differences in parasite x carbaryl interactions in 468 life history endpoints between the two clones. Based on the limited available knowledge regarding 469 parasite x chemical interactions, these mechanisms could be related to (i) toxicant-induced reduction 470 in filtration rate resulting indirectly in reduced intestinal exposure to and infection risk by the parasites 471 (Restif and Kaltz, 2006; Auld et al., 2012; Ebert et al. 1996) or (ii) the immuno-modulatory action of 472 chemicals (e.g. Galloway and Depledge, 2001; Galloway and Handy, 2003; Ville et al., 1997). The 473 first mechanism, however, has likely not played a role in our model system, as the absence of a clone 474 × carbaryl interaction for filtration rate (Table 2, three-way ANOVA) indicates no difference among the 475 two clones in how filtration rate is affected by carbaryl. In addition, such a mechanism would likely 476 result in antagonistic interactions in both clones, which is not what we observed here. In contrast, the 477 second mechanism is more plausible. Indeed, different carbaryl x parasite interactive effects between 478 the two clones were observed for proPO expression, i.e. antagonism in the non-sensitive clone and no 479 interaction in the sensitive clone. This is a key enzyme of the immune system response which has 480 been suggested as a useful indicator of immunocompetence in arthropods (Adamo et al., 2001; Kurtz 481 and Sauer, 2001) and which has been shown earlier to be responsive to P. ramosa infection (Labbé 482 and Little, 2009, Mucklow et al., 2004). Furthermore, Coors et al. (2008) already argued that their 483 observed synergistic interactions between P. ramosa and carbaryl for sterilization and survival (in 484 another D. magna clone) could have been the result of such immuno-suppressive activity of carbaryl. 485 All this suggests that differences in the immuno-modulatory activity of carbaryl in both clones may 486 indeed be at the basis of the very different parasite x carbaryl interactive effect between both clones, as observed across the variety of life-history endpoints investigated here. 487

Collectively, our data show differences in parasite × carbaryl interaction effects among two *Daphnia* clones, with the clone having a higher sensitivity to the chemical stressor exhibiting only synergistic interaction effects and the clone with lower sensitivity to the chemical stressor only exhibiting antagonistic interaction effects. We observed this both for life-history and physiological endpoints. This is in agreement with our hypothesis, derived from the findings of Holmstrup et al. (2010), that sensitive genotypes would be more prone to synergistic effects upon exposure to combined stressors. As our study involved only two clones, however, it does not provide a solid test of this hypothesis, but rather it

495 illustrates that genetically different genotypes (clones) of the same species may strongly differ in their 496 response to mixed stressors and, in our study, do so in a pattern that corresponds to expectations 497 generated by this hypothesis. Further, it should be emphasized that the conclusions drawn in this 498 study are only demonstrated for the particular combination of concentrations or stressor levels of both 499 stressors used. Thus, follow-up studies with multiple clones and a broader range of concentrations or 500 stressor levels would be required to test the broader validity of our hypothesis.

501

502 5. Conclusions

503 When analyzing carbaryl × parasite interaction effects between a clone that is highly sensitive and one 504 that is much less sensitive to carbaryl, we observed significant interactions in three out of nine 505 endpoints tested in the sensitive clone and three out of ten endpoints for the less sensitive clone. The 506 interaction effects observed for the less sensitive clone were all antagonistic, while only synergistic 507 interactions were detected in the carbaryl sensitive clone.

508

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518 References

- Adamo, S.A., Jensen, M., Younger, M. 2001. Changes in lifetime immunocompetence in male and female *Gryllus texensis* (formerly G. integer): trade-offs between immunity and reproduction. Anim. Behav. 62, 417–425.
- 521 Altshuler, I., Demiri, B., Xu, S., Constantin, A., Yan, N.D., Cristescu, M.E. 2011. An integrated multi-disciplinary
- 522 approach for studying multiple stressors in freshwater ecosystems: *Daphnia* as a model organism. Integr. Comp.
- 523 Biol. 51, 623-633.
- 524 Bliss, C.I. 1939. The toxicity of poisons applied jointly. Ann. Appl. Biol. 26, 585-615.
- 525 Bradford, M.M. 1976. A rapid and sensitive method for the quantification of microgram quantities of protein, 526 utilizing the principle of protein-dye binding. Anal. Biochem. 72, 248-254.
- 527 Calow, P. 1991. Physiological costs of combating chemical toxicants: ecological implications. Comp. Biochem.
 528 Physiol. C Comp. Pharmacol. 100, 3-6.
- 529 Calow, P., Baird, D.J., Barber, I. 1990. Clonal variation in general responses of *Daphnia magna Straus* to toxic 530 stress: II. Physiological effects. Func. Ecol. 4, 409-414.
- 531 Cerenius, L., Lee, B.L., Söderhall, K. 2008. The proPOsystem: Pros and cons for its role in invertebrate immunity.
 532 Trends. Immunol. 29, 263-271.
- 533 Chandrasekara, L.W.H.U., Pathiratne, A. 2007. Body size-related differences in the inhibition of brain
 534 acetylcholinesterase activity in juvenile Nile tilapia (*Oreochromis niloticus*) by chlorpyrifos and carbosulfan.
 535 Ecotoxicol. Environ. Safety 67, 109-119.
- 536 Coors, A., Decaestecker, E., Jansen, M., De Meester, L. 2008. Pesticide exposure strongly enhances parasite
 537 virulence in an invertebrate host model. Oikos. 117, 1840-1846.
- 538 Coors, A., De Meester, L. 2008. Synergistic, antagonistic and additive effects of multiple stressors: predation 539 threat, parasitism and pesticide exposure in *Daphnia magna*. J. Appl. Ecol. 45, 1820-1828.
- 540 Coors, A., De Meester, L. 2011. Fitness and virulence of a bacterial endoparasite in an environmentally stressed
 541 crustacean host. Parasitology. 138, 122-131.
- 542 Coors, A., Vanoverbeke, J., De Bie, T., De Meester, L. 2009. Land use, genetic diversity and toxicant tolerance in 543 natural populations of *Daphnia magna*. Aquat. Toxicol. 95, 71-79.
- 544 Corbett, J.R., Wright, K., Baillie, A.C. 1984. The biochemical mode of action of pesticides. Academic Press, 545 London.

- 546 Cuthbertson, A.G.S., Head, J., Walters, K.F.A., Murray, K.F.A. The integrated use of chemical insecticides and 547 the entomopathogenic nematode, *Steinernema feltiae*, for the control of sweetpotato whitefly, *Bemisia tabaci*. 548 Nematology. 5, 713-720.
- 549 Decaestecker, E., Gaba, S., Raeymaekers, J.A.M., Stoks, R., Van Kerckhoven, L., Ebert, D., De Meester, L.
 550 2007. Host–parasite 'Red Queen' dynamics archived in pond sediment. Nature. 450, 870-873.
- 551 De Coen, W.M., Janssen, C.R. 1997. The use of biomarkers in *Daphnia magna* toxicity testing. IV. Cellular 552 Energy Allocation: a new methodology to assess the energy budget of toxicant-stressed *Daphnia* populations. J.
- 553 Aquat. Ecosyst. Stress Recov. 6, 43-55.
- 554 De Coen, W.M., Janssen, C.R. 2003. The missing biomarker link: Relationships between effects on the cellular
- 555 energy allocation biomarker of toxicant-stressed Daphnia magna and corresponding population characteristics.
- 556 Environ. Toxicol. Chem. 22,1632-1641.
- 557 Ebert, D. 2005. Ecology, epidemiology and evolution of parasitism in Daphnia [Internet]. Bethesda (MD): National 558 Librarv of Medicine (US), National Center for Biotechnology Information. Available from: 559 http://www.ncbi.nlm.nih.gov/entrez/query.fcgi?db=Books
- Ebert, D., Carius, H.J., Little, T., Decaestecker, E. 2004. The evolution of virulence when parasites cause host
 castration and gigantism. Amer. Nat. 164, S19-S32.
- 562 Ellers, J., van Alphen, J.J.M. 1997. Life history evolution in *Asobara tabida*: plasticity in allocation of fat reserves
 563 to survival and reproduction. J. Evol. Biol. 10,771-785.
- Ellman, G.L., Courtney, K.D., Adres, V.J., Featherstone, A.M. 1961. A new and rapid colorimetric determination of
 acetylcholinesterase activity. Biochem. Pharmacol. 7, 88-95.
- 566 Galloway, T.S., Depledge, T.H. 2001. Immunotoxicity in invertebrates: Measurement and ecological relevance.
 567 Ecotoxicology. 10, 5-23.
- 568 Galloway, T.S., Handy, R. 2003. Immunotoxicity of oragnophosphorus pesticides. Ecotoxicology. 12, 345-363.
- 569 Gèrard, C., Carpentier, A., Paillisson, J.-M. 2008. Long-term dynamics and community structure of freshwater 570 gastropods exposed to parasitism and other environmental stressors. Freshw. Biol. 53, 470-484.
- Hall, S.R., Becker, C., Cáceres, C.E. 2007. Parasitic castration: a perspective from a model of dynamic energy
 budgets. Integr. Comp. Biol. 47, 295-309.
- 573 Heugens, E.H.W., Hendriks, A.J., Dekker, T., Van Straalen, N.M., Admiraal, W. 2001. A review of the effects of
- 574 multiple stressors on aquatic organisms and analysis of uncertainty factors for use in risk assessment. Crit. Rev.
- 575 Toxicol. 31,247-284.

- 576 Holmstrup, M., Bindesbøl, A.-M., Oostingh, G.J., Duschl, A., Scheil, V., Köhler, H.-R., Loureiro, S., Soares,
- 577 A.M.V.M., Ferreira, A.L.G., Kienle, C., Gerhardt, A., Laskowski, R., Kramarz, P.E., Bayley, M., Svendsen, C.,
- 578 Spurgeon, D.J. 2010. Interactions between effects of environmental chemicals and natural stressors: A review.
- 579 Sci. Total. Environ. 408, 3746-3762.
- Jansen, M., Stoks, R., Decaestecker, E., Coors, A., Van De Meutter, F., De Meester, L. 2010. Local exposure
 shapes spatial patterns in infectivity and community structure of *Daphnia* parasites. J. Anim. Ecol. 79, 1023-1033.
- Jansen, M., Coors, A., Stoks, R., De Meester, L. 2011a. Evolutionary ecotoxicology of pesticide resistance: a
 case study in *Daphnia*. Ecotoxicology. 20, 543-551.
- 584 Jansen, M., Stoks, R., Coors, A., Van Doorslaer, W., De Meester, L. 2011b. Collateral damage: Rapid exposure-
- 585 induced evolution of pesticide resistance leads to increased susceptibility to parasites. Evolution. 65, 2681-2691.
- 586 Jonker, D., Freidig, A.P., Groten, J.P., de Hollander, A.E.M., Stierum, .R.H, Woutersen, R.A., Feron, V.J. 2004.
- 587 Safety evaluation of chemical mixtures and combinations of chemical and non-chemical stressors. Rev. Environ.
- 588 Health. 19, 83-139.
- 589 Klüttgen, B., Dülmer, U., Engels, M., Ratte, H.T. 1994. ADaM, an artifical freshwater for the culture of 590 zooplankton. Water. Res. 28, 743-746.
- Koppenhöffer, A.M., Kaya, H.K. 1997. Additive and synergistic interaction between entomopathogenic nematodes
 and *Bacillus thuringiensis* for scarab grub control. Biol. Control. 8, 131-137.
- 593 Koppenhöffer, A.M., Grewal, P.S., Kaya, H.K. Synergism of imidacloprid and entomopathogenic nematodes 594 against white grubs: the mechanism. Entomol. Exp. Appl. 94, 283-293.
- 595 Kramarz, P.E., De Vaufleury, A., Zygmunt, P.M.S., Verdun, C. Increased response to cadmium and Bacillus
- 596 thuringiensis maize toxicity in the snail Helix aspersa infected by the nematode Phasmarhabditis hermaphrodita.
- 597 Environ. Toxicol. Chem. 26, 73-79.
- Kurtz, J., Sauer, K.P. 2001. Gender differences in phenoloxidase activity of *Panorpa vulgaris* hemocytes. J.
 Invert. Pathol. 78, 53-55.
- Labbé, P., Little, T.J. 2009. ProPhenolOxidase in *Daphnia magna*: cDNA sequencing and expression in relation to
 resistance to pathogens. Dev. Comp. Immunol. 33, 674-680.
- 602 Lampert, W. 2011. Daphnia: Development of a model organism in ecology and evolution. In: Kinne O. (Ed.),
- 603 Excellence in ecology, Book 21. International Ecology Institute, Oldendorf/Luhe.

- Laskowski, R., Bednarska, A.J., Kramarz, P.E., Loureiro, S., Scheil, V., Kudłek, J., Holmstrup, M. 2010.
 Interactions between toxic chemicals and natural environmental factors: A meta-analysis and case studies. Sci.
 Total. Environ. 408, 3763-3774.
- Little, T., Ebert, D. 2000. The cause of parasitic infection in natural populations of *Daphnia* (Crustacea:
 Cladocera): the role of host genetics. Proc. R. Soc. Lond. B Biol. Sci. 267, 2037-2042.
- Mucklow, P.T., Ebert, D. 2003. Physiology of immunity in the water flea *Daphnia magna*: Environmental and
 genetic aspects of phenoloxidase activity. Physiol. Biochem. Zool. 76, 836-842.
- 611 Mucklow, P.T., Vizoso, D.B., Jensen, K.H., Refardt, D., Ebert, D. 2004. Variation in phenoloxidase activity and its
- relation to parasite resistance within and between populations of *Daphnia magna*. Proc. R. Soc. Lond. B. 271,1175-1183.
- Muyssen, B.T.A., De Schamphelaere, K.A.C., Janssen, C.R. 2006. Mechanisms of chronic waterborne Zn toxicity
 in *Daphnia magna*. Aquat. Toxicol. 77, 393-401.
- Muyssen, B.T.A., Messiaen, M., Janssen, C.R. 2010. Combined cadmium and temperature acclimation in
 Daphnia magna: Physiological and sub-cellular effects. Ecotox. Environ. Safe. 73, 735-742.
- 618 Nath, B.S., Suresh, A., Varma, B.M., Kumar, R.P. 1997. Changes in protein metabolism in hemolymph and fat
- 619 body of the silkworm, Bombyx mori (Lepidoptera: Bombycidae) in response to organophosphorus insecticides
- 620 toxicity. Ecotox. Environ. Safe. 36, 169-173.
- Piegorsch, W., Bailor, J.A. 1997. Statistics for environmental biology and toxicology. First ed, Chapman & Hall,
 New York.
- Restif, O., Kaltz, O. 2006. Condition-dependent virulence in a horizontally and vertically transmitted bacterial
 parasite. Oikos. 114, 148-158.
- Rivero, A., Casas, J. 1999. Incorporating physiology into parasitoid behavioral ecology: the allocation of nutritional
 resources. Res. Popul. Ecol. 41, 39-45.
- Salice, C.J., Roesijadi, G. Resistance to cadmium and parasite infection are inversely related in two strains of a
 freshwater gastropod. Environ. Toxicol. Chem. 21, 1398-1403.
- Söderhall, K., Cerenius, L. 1998. Role of the prophenoloxidase-activating system in invertebrate immunity. Curr.
 Opin. Immunol. 10, 23-28.
- Stollewerk, A. 2010. The water flea *Daphnia* a "new" model system for ecology and evolution? J. Biol. 9, article
 21.

- US EPA. 2003. Environmental fate and ecological risk assessment for the re-registration of carbaryl. US
 Environmental Protection Agency, Washington, DC.
- van Gestel, C.A.M., Jonker, M., Kammenga, J.E., Laskowski, R., Svendsen, C. 2010. Mixture toxicity: Linking
 approaches from ecological and human toxicology. First ed., CRC Press, Florida.
- 637 Ville, P., Roch, P., Cooper, E.L., Narbonne, J.-F. 1997. Immuno-modulator effects of carbaryl and 2,4 D in the
- 638 earthworm *Eisenia fetida* Andrei. Arch. Environ. Contam. Toxicol. 32, 291-297.
- 639 Warming, T.P., Mulderij, G., Christoffersen, K.S. 2009. Clonal variation in physiological responses of Daphnia
- 640 *magna* to the strobilurin fungicide azoxystrobin. Environ. Toxicol. Chem. 28,374-380.
- Kuereb, B., Chaumot, A., Mons, R., Garric, J., Geffard, O. 2009. Acetylcholinesterase activity in Gammarus
- 642 *fossarum* (Crustacea Amphipoda): Intrinsic variability, reference levels and a reliable tool for field surveys. Aquat.
- 643 Toxicol. 93, 225-233.

645 Figures



Figure 1: Life-history endpoints (A: survival; B: early reproduction; C: body length and D: maturation 647 rate) for clones NS (carbaryl non-sensitive) and S (carbaryl sensitive) for all treatments. Predicted 648 649 values of each endpoint for the combined treatment according to the Independent Action Model, 650 calculated with equation 3, are also depicted. (A/S) indicates significant antagonistic and synergistic 651 interactions, respectively. Early reproduction is expressed as the number of juveniles per egg-carrying 652 individual on day 8 and maturation rate as the percentage egg-carrying individuals on day 8. Ctrl: 653 control; P: parasite exposure; C: carbaryl exposure; PxC; combined parasite and carbaryl exposure. 654 Error bars indicate standard deviation.



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Figure 2: Physiological endpoints (A: filtration rate; B: total protein content; C: total carbohydrate content; D: total lipid content; E: electron transport system activity; F: acetylcholinesterase activity and G: prophenoloxidase expression) for clones NS (carbaryl non-sensitive) and S (carbaryl sensitive) for all treatments. Predicted values of the endpoints in the combined treatment according to the

- 662 Independent Action Model, calculated with equation 3, are also depicted. (A/S) indicates significant
- antagonistic and synergistic interaction, respectively. Ctrl: control; P: parasite exposure; C: carbaryl
- 664 exposure; PxC; combined parasite and carbaryl exposure. Error bars indicate standard deviation.

666 Supplementary Data

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