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## Ocean acidification impacts bacteria–phytoplankton coupling at low-nutrient conditions

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**Abstract.** The oceans absorb about a quarter of the annually produced anthropogenic atmospheric carbon dioxide (CO<sub>2</sub>), resulting in a decrease in surface water pH, a process termed ocean acidification (OA). Surprisingly little is known about how OA affects the physiology of heterotrophic bacteria or the coupling of heterotrophic bacteria to phytoplankton when nutrients are limited. Previous experiments were, for the most part, undertaken during productive phases or following nutrient additions designed to stimulate algal blooms. Therefore, we performed an in situ large-volume mesocosm (~55 m<sup>3</sup>) experiment in the Baltic Sea by simulating different fugacities of CO<sub>2</sub> (*f*CO<sub>2</sub>) extending from present to future conditions. The study was conducted in July–August after the nominal spring bloom, in order to maintain low-nutrient conditions throughout the experiment. This resulted in phytoplankton communities dominated by small-sized functional groups (picophytoplankton). There was no consistent *f*CO<sub>2</sub>-induced effect on bacterial protein production (BPP), cell-specific BPP (csBPP) or biovolumes (BVs) of either free-living (FL) or particle-associated (PA) heterotrophic bacteria, when considered as individual components (univariate analyses). Permutational

Multivariate Analysis of Variance (PERMANOVA) revealed a significant effect of the *f*CO<sub>2</sub> treatment on entire assemblages of dissolved and particulate nutrients, metabolic parameters and the bacteria–phytoplankton community. However, distance-based linear modelling only identified *f*CO<sub>2</sub> as a factor explaining the variability observed amongst the microbial community composition, but not for explaining variability within the metabolic parameters. This suggests that *f*CO<sub>2</sub> impacts on microbial metabolic parameters occurred indirectly through varying physicochemical parameters and microbial species composition. Cluster analyses examining the co-occurrence of different functional groups of bacteria and phytoplankton further revealed a separation of the four *f*CO<sub>2</sub>-treated mesocosms from both control mesocosms, indicating that complex trophic interactions might be altered in a future acidified ocean. Possible consequences for nutrient cycling and carbon export are still largely unknown, in particular in a nutrient-limited ocean.

## 1 Introduction

Since the industrial revolution the oceans have absorbed ca. one half of anthropogenic carbon dioxide (CO<sub>2</sub>). This has resulted in a shift in carbonate equilibria and pH (e.g. Caldeira and Wickett, 2003) with potential consequences for organismal physiology (Taylor et al., 2012). In principal, autotrophs should be fertilised by an enhanced CO<sub>2</sub> availability, increasing the production of particulate (POM) and dissolved organic matter (DOM; Hein and Sand-Jensen, 1997; Riebesell et al., 2007). However, most CO<sub>2</sub> enrichment experiments studying natural plankton assemblages under variable nutrient conditions do not reveal a consistent response of primary production to elevated CO<sub>2</sub> (e.g. Engel et al., 2005; Riebesell et al., 2007; Bach et al., 2016). Both amount and stoichiometric composition of algal DOM and POM can be affected by changes in *f*CO<sub>2</sub>. For example, Riebesell et al. (2007) and Maat et al. (2014) reported an increased stoichiometric draw-down of inorganic carbon (C) to nitrogen (N) at higher levels of *f*CO<sub>2</sub>, most likely as a result from C overconsumption (Toggweiler, 1993).

Heterotrophic bacteria, in oligotrophic systems, are largely dependent on phytoplankton-derived organic carbon (e.g. Azam, 1998), and as such respond to alterations in both the quantity and quality of phytoplankton-derived DOM and POM (e.g. Allgaier et al., 2008; Grossart et al., 2006a). Availability and competition for nutrients, however, can substantially impact *f*CO<sub>2</sub>-induced changes in activity and biomass of phytoplankton and subsequently of heterotrophic bacteria. In nutrient-depleted or nutrient-limited systems, bacteria are restricted in their utilization of phytoplankton-derived organic carbon (Hoikkala et al., 2009; Lignell et al., 2008). Consequently, *f*CO<sub>2</sub>-dependent increases in inorganic C availability for autotrophs may not stimulate heterotrophic activity, causing a decoupling of heterotrophic and autotrophic processes (Thingstad et al., 2008). The accumulation of bioavailable dissolved organic carbon (DOC) and particulate organic carbon (POC), as a consequence of this decoupling in nutrient-limited oceanic surface waters, may have profound consequences for nutrient cycling and the nature of the oceanic carbon pump (e.g. Thingstad et al., 1997). Given that various studies have reported on limitation of bacterial growth by inorganic nutrients in several parts of the Baltic Sea (e.g. Hoikkala et al., 2009; Kuparinen and Heinänen, 1993), we sought to evaluate the effects of enhanced *f*CO<sub>2</sub> on activity and biomass of free-living (FL) as well as particle-associated (PA) bacteria during a period characterized by low nutrients and low productivity.

## 2 Methods

### 2.1 Experimental setup, CO<sub>2</sub> manipulation and sampling

Nine floating, pelagic KOSMOS (Kiel Off-Shore Mesocosms for future Ocean Simulations) mesocosms (cylindrical, 2 m diameter, 17 m long with conical sediment trap extending to 19 m depth) were moored (day  $-10 = t - 10$ ; 10 days before CO<sub>2</sub> manipulation) at 59°51.5' N, 23°15.5' E, in the Baltic Sea at Tvärminne Storfjärden on the southwestern coast of Finland on 12 June 2012. Mesocosm bags were rinsed for a period of 5 days, covered on the top and bottom with a 3 mm net to exclude larger organisms. Thereby, the containing water was fully exchanged with the surrounding water masses. Sediment traps were attached to the bottom of each mesocosm at 17 m depth 5 days prior the start of the experiment ( $t - 5$ ). In addition, submerged mesocosm bags were drawn 1.5 m above the water surface, enclosing and separating ~55 m<sup>3</sup> of water from the surrounding Baltic Sea, and meshes were removed. Mesocosms were covered by a photosynthetic active radiation (PAR) transparent roof to prevent nutrient addition from birds and freshwater input from rain. Additionally, existing haloclines were removed in each mesocosm as described in Paul et al. (2015), thereby creating a fully homogeneous water body.

The experiment was conducted between 17 June ( $t - 5$ ) and 4 August ( $t 43$ ) 2012. To minimize environmental stress on enclosed organisms, CO<sub>2</sub> addition was performed stepwise over 3 days commencing on day  $t 0$ . CO<sub>2</sub> addition was repeated on  $t 15$  in the upper mixed 7 m of water to compensate for outgassing. Different *f*CO<sub>2</sub> treatments were achieved by equally distributing filtered (50 µm), CO<sub>2</sub>-saturated seawater into the treated mesocosms with a water distributor as described by Paul et al. (2015). Control mesocosms were also manipulated with the water distributor and 50 µM pre-filtered water without CO<sub>2</sub>. CO<sub>2</sub> amendments resulted in ca. 0.04–0.35 % increases in the total water volume across mesocosms (Paul et al., 2015). Integrated water samples (0–17 m) were collected from each mesocosm and the surrounding seawater using depth-integrated water samplers (IWS, HYDROBIOS, Kiel). Samples for activity measurements were directly subsampled from the IWS on the sampling boat without headspace to maintain in situ *f*CO<sub>2</sub> levels during incubation.

Unfortunately, three mesocosms failed during the experiment as a consequence of welding faults, resulting in unquantifiable water exchanges with the surrounding waters. Therefore, with reference to the six remaining mesocosms, CO<sub>2</sub> concentrations defining each treatment are reported as the mean *f*CO<sub>2</sub> determined over the initial 43 days ( $t 1 - t 43$ ) as described in Paul et al. (2015). The control mesocosms (two replicates) had 365 and 368 µatm *f*CO<sub>2</sub>. The four treatment mesocosms each had 497, 821, 1007 and 1231 µatm *f*CO<sub>2</sub>. Detailed descriptions on the study site, mesocosm

deployment, system performance of the mesocosm facility throughout the experiment, CO<sub>2</sub> addition, carbonate chemistry, cleaning of the mesocosm bags, as well as sampling frequencies of single parameters are given in Paul et al. (2015).

## 2.2 Physical and chemical parameters

Temperature and salinity were measured using a CTC60M memory probe (Sea and Sun Technology GmbH, Trappenkamp, Germany) and are calculated as the mean integrated over the total depth. Photosynthetic active radiation (PAR) was measured with a PAR sensor (LI-COR LI-192) on the roof of Tvärminne Zoological Station.

Samples for dissolved inorganic carbon (DIC) and total pH were gently pressure-filtered (Sarstedt Filtropur PES, 0.2 µm pore size) using a membrane pump (Stepdos). As described in Dickson et al. (2007), total pH was determined on a Cary 100 (Varian) spectrophotometer in a temperature-controlled 10 cm cuvette using *m*-cresol indicator dye. DIC concentrations were determined by infrared absorption using a LI-COR LI-7000 on an AIRICA system (MARIANDA e.K., Kiel). Total pH and DIC were used to calculate carbonate chemistry speciation using the stoichiometric equilibrium constants for carbonic acid of Mehrbach et al. (1973) as re-fitted by Lueker et al. (2000).

Samples for dissolved organic carbon (DOC), total dissolved nitrogen (TDN), dissolved silica (DSi) and dissolved inorganic phosphate (DIP) were filtered through pre-combusted (450 °C, 6 h) GF/F filters (Whatman, nominal pore size of 0.7 µm). Concentrations of DOC and TDN were determined using a high-temperature catalytic combustion technique with a Shimadzu TOC-TN V analyser following Badr et al. (2003). DSi concentrations were determined using standard colorimetric techniques (Grasshoff et al., 1983) at the micromolar level with a nutrient autoanalyser (Seal Analytical, QuAAtro). DIP concentrations were determined with a colorimetric method using a 2 m liquid waveguide capillary cell (Patey et al., 2008) with a miniaturized detector (Ocean Optics Ltd).

Total particulate carbon (TPC), particulate organic nitrogen (PON) and total particulate phosphorus (TPP) samples were collected onto pre-combusted (450 °C, 6 h) GF/F filters (Whatman, nominal pore size of 0.7 µm) using gentle vacuum filtration and stored in glass Petri dishes at –20 °C. Biogenic silica (BSi) samples were collected on cellulose acetate filters (0.65 µm, Whatman) using gentle vacuum filtration (<200 mbar) and stored in glass Petri dishes at –20 °C. Filters for TPC/PON analyses were dried at 60 °C, packed into tin capsules and measured on an elemental analyser (EuroEA) according to Sharp (1974), coupled by either a ConFlo II to a Finnigan Delta<sup>Plus</sup> isotope ratio mass spectrometer or a ConFlo III to a Thermo Finnigan Delta<sup>Plus</sup> XP isotope ratio mass spectrometer. Filters for TPP were treated with oxidizing decomposition reagent (MERCK, catalogue no. 112936) to oxidize organic phosphorus to orthophos-

phate. Particulate silica was leached from filtered material. Concentrations of dissolved inorganic phosphate as well as dissolved silica were determined spectrophotometrically according to Hansen and Koroleff (1999).

Samples for chlorophyll *a* (Chl *a*) were filtered on GF/F filters (Whatman, nominal pore size of 0.7 µm) and stored at –20 °C. Chl *a* was extracted in acetone (90 %) and samples were homogenized. After centrifugation (10 min, 800 × *g*, 4 °C) the supernatant was analysed on a fluorometer (TURNER 10-AU) to determine concentrations of Chl *a* (Welschmeyer, 1994).

Further details on the determination of physical parameters, concentration of Chl *a*, as well as dissolved and particulate nutrients can be obtained from Paul et al. (2015).

## 2.3 Microbial standing stock

Abundance of free-living (FL) heterotrophic prokaryotes (HP) and photoautotrophic prokaryotic (*Synechococcus* spp.) as well as eukaryotic cells (<20 µm) were determined by flow cytometry (Crawford et al., 2016). Briefly, phytoplankton were discriminated based on their chlorophyll red and/or phycoerythrin orange autofluorescence (Marie et al., 1999). In combination with their side scatter signal and size fractionation, the phytoplankton community could be divided into six clusters, varying in size from 1 to 8.8 µm average cell diameter (Crawford et al., 2016). Three groups of picoeukaryotic phytoplankton (Pico I–III), 1 picoprokaryotic photoautotroph (*Synechococcus* spp., SYN) and 2 nanoeukaryotic phytoplankton groups (Nano I–II) were detected. Biovolume (BV) estimations were based on cell abundance and average cell diameters by assuming a spherical cell shape. The BV sum of *Synechococcus* and Pico I–III is expressed as BV<sub>Pico</sub>. The BV sum of Nano I and II will be referred to as BV<sub>Nano</sub>.

Abundances of FL prokaryotes were determined from 0.5 % glutaraldehyde fixed samples after staining with the nucleic acid-specific dye SYBR green I (Crawford et al., 2016). Unicellular cyanobacteria (*Synechococcus* spp.) contributed at maximum 10 % of total counts. Two additional groups were identified based on their low (LDNA) and high (HDNA) fluorescence. This identification was based on gating of SYBR green I fluorescence against the side scatter signal (Crawford et al., 2016). Particle-associated (PA) prokaryotes were enumerated by epifluorescence-microscopy on a Leica Leitz DMRB fluorescence microscope with UV- and blue light excitation filters (Leica Microsystems, Wetzlar, Germany). Fresh samples were gently mixed to prevent particle settling and a 15 mL subsample was filtered on a 0.1 % Irgalan Black coloured 5.0 µm polycarbonate filter (Whatman, Maidstone, UK; Hobbie et al., 1977). Filters were fixed with glutaraldehyde (Carl Roth, Karlsruhe, Germany; final conc. 2 %) and stained for 15 min with 4-6-diamidino-2-phenylindole (DAPI, final conc. 1 µg mL<sup>-1</sup>; Porter and Feig, 1980) directly on the filtration device and rinsed twice with

sterile filtered habitat water before air-drying and embedding in Citifluor AF1 (Citifluor Ltd, London, UK) on a microscopic slide (Rieck et al., 2015). Counts were made based on 15 random unique squares as observed at a magnification of 1000 $\times$ . The total number of heterotrophic PA prokaryotes was enumerated by subtracting Chl *a* autofluorescent cells from DAPI-stained cells (Rieck et al., 2015).

BV of FL and PA prokaryotes was calculated separately. For FL prokaryotes we estimated BVs on the basis of an average cell volume of 0.06  $\mu\text{m}^3$  (Hagström et al., 1979). BV of PA prokaryotes was calculated from measurements of 1600 cells across three different mesocosms (346, 868, 1333  $\mu\text{atm}$ ) and three time points ( $t_0$ ,  $t_{20}$ ,  $t_{39}$ ) throughout the experiment (Massana et al., 1997). A resulting average BV of 0.16  $\mu\text{m}^3$  per cell was used to calculate BV of PA prokaryotes derived from cell abundances. We subsequently adopted the term “heterotrophic bacteria”, since bacteria account for the majority of non-photosynthetic prokaryotes in surface waters (Kirchman et al., 2007).

## 2.4 Metabolic parameters

Rates of bacterial protein production (BPP) were determined by incorporation of  $^{14}\text{C}$ -leucine ( $^{14}\text{C}$ -Leu; Simon and Azam, 1989) according to Grossart et al. (2006a). Triplicates and a formalin-killed control were incubated with  $^{14}\text{C}$ -Leu (213 mCi  $\text{mmol}^{-1}$ ; Hartmann Analytic GmbH, Germany) at a final concentration of 165 nM, which ensured saturation of the uptake systems of both FL and PA bacteria. Incubation was performed in the dark at in situ temperature (between 7.8 and 15.8  $^{\circ}\text{C}$ ) for 1.5 h. After fixation with 2 % formalin, samples were filtered onto 5.0  $\mu\text{m}$  (PA bacteria) nitrocellulose filters (Sartorius, Germany) and extracted with ice-cold 5 % trichloroacetic acid (TCA) for 5 min. Thereafter, filters were rinsed twice with ice-cold 5 % TCA, once with ethanol (50 % *v/v*) and dissolved in ethyl acetate for measurement by liquid scintillation counting (Wallac 1414, Perkin Elmer). Afterwards, the collected filtrate was filtered on 0.2  $\mu\text{m}$  (FL bacteria) nitrocellulose filters (Sartorius, Germany) and processed in the same way as the 5.0  $\mu\text{m}$  filters. Standard deviation of triplicate measurements was usually < 15 %. The amount of incorporated  $^{14}\text{C}$ -Leu was converted into BPP by using an intracellular isotope dilution factor of 2. A conversion factor of 0.86 was used to convert the produced protein into carbon (Simon and Azam, 1989). Cell-specific BPP rates (csBPP) were calculated by dividing BPP rates by abundances of FL prokaryotes and PA HP.

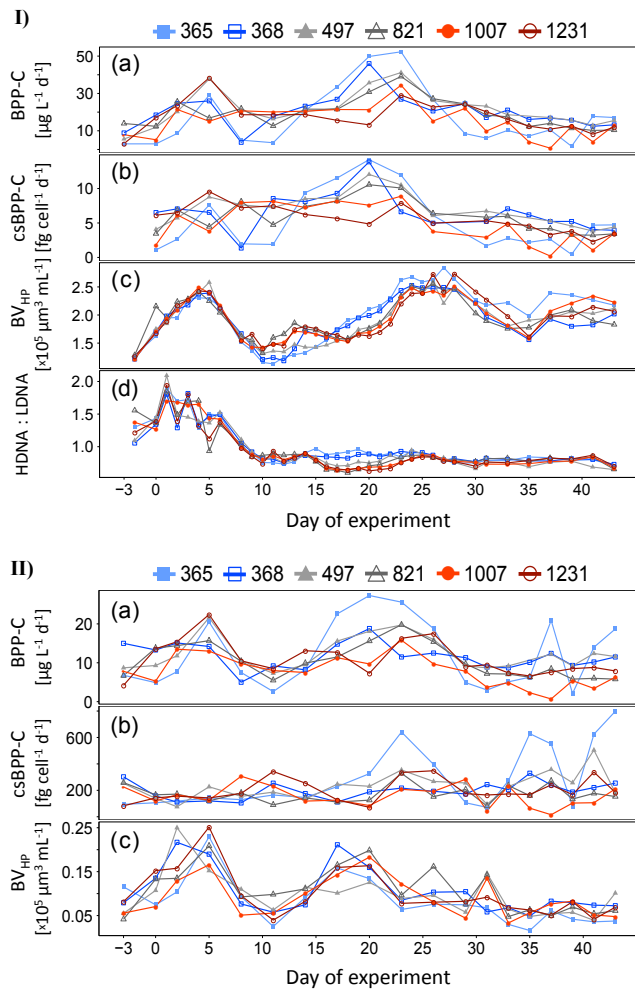
Community respiration (CR) rates were calculated from oxygen consumption during an incubation period of 48 h at in situ temperature in the dark by assuming a respiratory quotient of 1 (Berggren et al., 2012). Thereby, oxygen concentrations were measured in triplicate in 120 mL  $\text{O}_2$  bottles without headspace, using a fiber optical dipping probe (PreSens, Fibox 3), which was calibrated against anoxic and air saturated water.

Primary production (PP) was measured using radio-labelled  $\text{NaH}^{14}\text{CO}_3$  (Steeman-Nielsen, 1952) in 0–10 m depth-integrated samples. After incubation of duplicate samples with 10  $\mu\text{L}$  of  $^{14}\text{C}$  bicarbonate solution (DHI Lab, 20  $\mu\text{Ci mL}^{-1}$ ) in 8 mL vials at 2, 4, 6, 8 and 10 m for 24 h, samples were acidified with 1 M HCl to remove remaining inorganic  $^{14}\text{C}$ . Radioactivity was determined by using a scintillation counter (Wallac 1414, Perkin Elmer). PP was calculated knowing the dark-control corrected  $^{14}\text{C}$  incorporation and the fraction of the  $^{14}\text{C}$  addition to the total inorganic carbon pool according to Gargas (1975). Further details on the measurement of CR and PP are given by Spilling et al. (2016a).

## 2.5 Statistical analyses

Permutational multivariate analysis of variance – PERMANOVA (Anderson, 2001) was used to determine associations between physical/chemical variables and biotic variables. PERMANOVA (perm = 9999) was performed to test for significant differences in variance over time and between  $f\text{CO}_2$ -treated mesocosms (Anderson et al., 2008). Environmental data were normalized according to Clarke and Gorley (2006). Biotic abundance data were  $\log(x + 1)$  transformed (Clarke and Gorley, 2006). PERMANOVA partitions the total sum of squares based on the experimental design and calculates a distance-based pseudo- $F$  statistic for each term in the model. Distance-based linear modelling (DistLM) was implemented to relate physical/chemical predictor variables and the multivariate assemblage of biotic variables (Supplement Table S1; Legendre and Anderson, 1999; Anderson et al., 2008). The DistLM routine was based on the AIC model selection criterion (see Anderson et al., 2008) using a stepwise selection procedure. In case of equally AIC-ranked models (difference < 1), a model with fewer parameters was preferred. A Principal Component Analysis (PCA) was performed on normalized chemical data to identify chemical gradients and patterns between the differently  $f\text{CO}_2$ -treated mesocosms over time (Clarke and Gorley, 2006). Distance-based redundancy analysis (dbRDA) was used for visual interpretation of the DistLM in multidimensional space (Anderson et al., 2008). Multivariate analyses of physicochemical, metabolic and community data were performed on a reduced data set comprising 10 time points ( $t_5$ – $t_{29}$ , every 3 days;  $t_{31}$ ), containing all measured activity variables (BPP, areal PP and CR). Missing values of nutrient data or abundance data (based on every-other-day measurements) were estimated as means of the preceding and following measurement day. In general, no activity data were interpolated or extrapolated.

Cluster analyses were performed based on Spearman's rank correlation coefficient calculated for each mesocosm between all possible combinations of LDNA, HDNA, pico- and nanophytoplankton abundances as well as total Chl *a*. Thereafter,  $p$  values were corrected for multiple testing ac-



**Figure 1.** I: Free-living (0.2–5.0  $\mu\text{m}$ ) and II: particle-associated (> 5.0  $\mu\text{m}$ ), size-fractionated rates of (a) bacterial protein production (BPP) [ $\mu\text{g CL}^{-1} \text{d}^{-1}$ ] as well as (b) cell-specific bacterial protein production (csBPP) [ $\text{fg C cell}^{-1} \text{d}^{-1}$ ] and (c) heterotrophic prokaryotic biovolume ( $\text{BV}_{\text{HP}}$ ) [ $\times 10^5 \mu\text{m}^3 \text{mL}^{-1}$ ] during the course of the experiment. I: (d) Ratio of high vs. low nucleic acid-stained prokaryotic heterotrophs (HDNA : LDNA). Colours and symbols indicate average  $f\text{CO}_2$  [ $\mu\text{atm}$ ] between  $t1 - t43$ .

cording to Benjamini and Hochberg (1995). The R package pvclust was used to assess the uncertainty in hierarchical cluster analysis (Suzuki and Shimodeira, 2015). For each cluster, AU (approximately unbiased)  $p$  values (between 0 and 1) were calculated via multiscale bootstrap resampling (Suzuki and Shimodeira, 2015).

PERMANOVA, distLM and dbrDA were carried out using Primer 6.0 and PERMANOVA + for PRIMER software (Clarke and Gorley, 2006; Anderson et al., 2008). All other analyses and the visualization of results were performed with R 3.2.5 (R Core Team, 2016) using packages Hmisc (Harrell et al., 2016), vegan (Oksanen et al., 2016), pvclust (Suzuki

and Shimodeira, 2015), gplots (Warnes et al., 2016) and ggplot2 (Wickham, 2009).

## 3 Results

### 3.1 Bacterial dynamics

Heterotrophic bacterial BV was predominantly comprised of FL bacteria. PA bacteria contributed at maximum  $2 \pm 0.7$ – $10 \pm 0.7$  % (mean  $4.8 \pm 0.6$  %) of total bacterial BV. PA bacteria, however, accounted for a substantial fraction of overall BPP ( $27 \pm 1$ – $59 \pm 7$  %, mean  $39 \pm 4$  %). There was no significant effect of  $f\text{CO}_2$  on BPP, csBPP or BV of neither FL nor PA heterotrophic bacteria ( $p_{\text{perm}} > 0.05$ ); however, a significant temporal effect was observed ( $p_{\text{perm}} < 0.05$ ). Both bacterial size-fractions had distinct dynamics in abundance, BPP and csBPP during the course of the experiment (Fig. 1). BPP and bacterial abundances were closely related to Chl  $a$  and BV of nano- and picophytoplankton, trending along with Chl  $a$  until  $t10$  and then continuing to increase with BVs of nano- and photoautotrophs and Chl  $a$ . The period between  $t16$  and  $t26$ , following a sharp decrease in Chl  $a$  at  $t16$ , revealed the highest BPP rates across the experiment with lower rates at higher  $f\text{CO}_2$  for PA as well as FL bacteria. CsBPP rates were lower at elevated  $f\text{CO}_2$  for only the FL bacteria during this period. Additionally, BVs of FL and PA bacteria revealed contrasting dynamics (Figs. 1, S1 in the Supplement). PA bacterial BVs declined with the decay of Chl  $a$ , whereas FL bacterial BVs increased strongly associated with an increase in BV of picophotoautotrophs during this period. The ratio of HDNA : LDNA prokaryotes, which both make up FL bacteria, also showed differences between the experimental treatments. Between  $t14 - t25$  the ratio of HDNA : LDNA was lower at higher  $f\text{CO}_2$ .

### 3.2 Phytoplankton dynamics

Chl  $a$  concentration exhibited distinct maxima at two time periods ( $t5$  and  $t16$ ). The second maximum was associated with an increase in the BV of nanophotoautotrophs ( $\text{BV}_{\text{Nano}}$ , Fig. 2). This increase was reduced in mesocosms containing higher levels of  $f\text{CO}_2$  between  $t13 - t17$ . The differences in  $\text{BV}_{\text{Nano}}$  between the treatments were reflected in lower concentrations of Chl  $a$  in the three highest  $f\text{CO}_2$ -treated mesocosm at  $t16$ . Chl  $a$  and  $\text{BV}_{\text{Nano}}$  concentrations declined after  $t16$ . In contrast, BV of picophotoautotrophs ( $\text{BV}_{\text{Pico}}$ ) increased after  $t11$ , associated with an increase in BV of *Synechococcus* spp., which accounted for  $31 \pm 2$  to  $59 \pm 2$  % of  $\text{BV}_{\text{Pico}}$  across the period of this study (Fig. S2). All four groups of picoautotrophs distinguished by flow cytometry, exhibited time-dependent positive or negative relationships with  $f\text{CO}_2$  (Figs. 3, S2, S3). The Pico I ( $\sim 1 \mu\text{m}$ ) and Pico II taxa infrequently exhibited strong fertilization effects in response to the  $f\text{CO}_2$  treatment. In contrast, *Synechococcus*



**Table 1.** Results of two-factor permutational multivariate analysis of variance (PERMANOVA)<sup>a</sup> on a resemblance matrix (Euclidian distance) of normalized chemical variables (Phosphate, DOC, TDN, DSi, TPC, PON, POP, BSi). Degrees of freedom (*df*); sum of squares (SS); mean square (MS).

Source of variation	<i>df</i>	SS	MS	Pseudo- <i>F</i>	<i>p</i> (perm)	Unique perms
Time	9	309.93	34.436	11.118	0.0001	9920
<i>f</i> CO <sub>2</sub> <sup>b</sup>	4	31.974	7.9936	2.5808	0.0246	9936
Time × <i>f</i> CO <sub>2</sub>	36	80.177	2.2271	0.71906	0.8794	9904
Residuals	10	30.973	3.0973			
Total	59	472				

<sup>a</sup> Permutation was performed with unrestricted permutation of raw data. <sup>b</sup> Pairwise test could only be performed for control mesocosms (*n* = 2) with each *f*CO<sub>2</sub> treatment (*n* = 1), due to missing replication for each *f*CO<sub>2</sub> treatment. Pairwise comparison was only significant between controls and the highest *f*CO<sub>2</sub> treatment (*p*<sub>perm</sub> = 0.029).

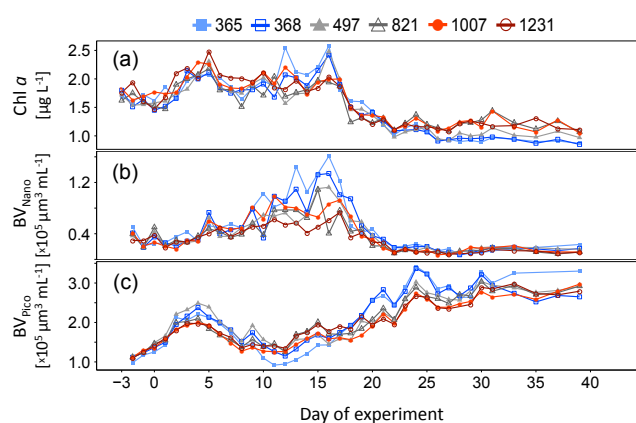
**Table 2.** Eigenvectors and explained variation of the first four axes of a PCA on normalized variables of dissolved and particulate nutrients. Ordination of the PCA is visualized in Fig. 6.

Variable	PC1	PC2	PC3	PC4
DOC	−0.4	−0.23	0.04	0.68
TDN	0.39	0.21	0.21	0.47
Phosphate	−0.1	0.48	−0.74	0.35
DSi	0.3	0.52	−0.03	−0.24
TPC	0.48	−0.06	0.03	0.13
PON	0.46	−0.05	−0.05	0.16
POP	0.36	−0.39	−0.04	0.21
BSi	0.17	−0.51	−0.63	−0.22
% variation	49.2	19.7	11.4	7.2
cum. % variation	49.2	68.9	80.4	87.6

spp. and Pico III were infrequently negatively affected by the *f*CO<sub>2</sub> treatment.

### 3.3 Relation between functional heterotrophic and autotrophic groups

A cluster analysis of pairwise Spearman correlations between functional bacterial and phytoplankton groups revealed a separation based on *f*CO<sub>2</sub> treatment. Specifically the four CO<sub>2</sub>-amended mesocosms were readily distinguishable from the control treatments. Multiple bootstrap resampling (Suzuki and Shimodaira, 2015) supported this, but only significantly for the three highest *f*CO<sub>2</sub>-treated mesocosms. The two highest *f*CO<sub>2</sub> treatments revealed a positive correlation between LDNA bacteria and Pico I, which could not be observed in any other experimental treatment. In all CO<sub>2</sub>-treated mesocosms we observed positive correlations between *Synechococcus* spp. and Pico III as well as *Synechococcus* spp. and Pico I, which were not present in both control mesocosms. In contrast, positive correlations between LDNA and HDNA were not detected in any *f*CO<sub>2</sub> treatment. Additionally, positive correlations between Pico I and Nano II as well as HDNA and Cyanobacteria were only



**Figure 2.** (a) Concentration of chlorophyll *a* [ $\mu\text{g L}^{-1}$ ], (b) biovolume of nanophytoplankton (Nano I and Nano II) [ $\times 10^5 \mu\text{m}^3 \text{mL}^{-1}$ ] and (c) biovolume of picophytoplankton (*Synechococcus* spp., Pico I–III) [ $\times 10^5 \mu\text{m}^3 \text{mL}^{-1}$ ] during the course of the experiment. Colours and symbols indicate average *f*CO<sub>2</sub> [ $\mu\text{atm}$ ] between *t*1 and *t*43.

present in both controls and the lowest *f*CO<sub>2</sub> treatment (Fig. 4).

After *t*10, the ratio between heterotrophic prokaryotic BV and Chl *a* varied between the *f*CO<sub>2</sub> treatments, but did not show a consistent pattern. After *t*17, however, the control mesocosms revealed a higher ratio compared to all *f*CO<sub>2</sub>-treated mesocosms (Fig. 5).

### 3.4 Multivariate physicochemical characterization

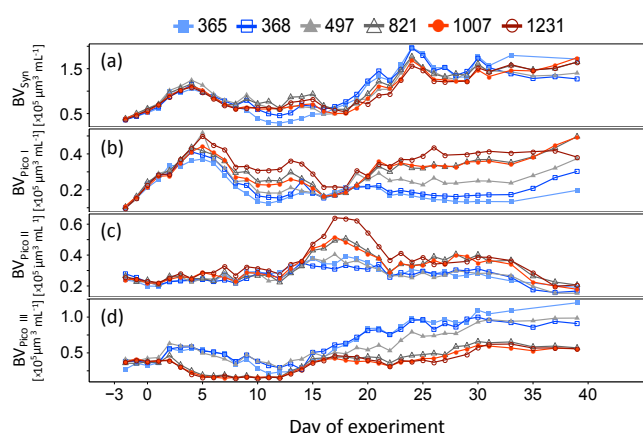
Integrated water temperature and PAR ranged between 8.0–15.9 °C and 11.2–66.8 mol m<sup>−2</sup> day<sup>−1</sup> during the experiment respectively. Integrated water temperature reached the maximum at *t*15 and dropped again to 8.2 °C at *t*31.

PERMANOVA results (Table 1) on a multivariate assemblage of dissolved (DOC, TDN, Phosphate, BSi) and particulate (TPC, PON, POP, BSi) nutrients showed significant temporal (Time-*F*<sub>9,10</sub> = 11.1, *p* = 0.0001) and spa-

**Table 3.** Results of two-factor permutational multivariate analysis of variance (PERMANOVA)<sup>a</sup> on a resemblance matrix (Euclidian distance) based on normalized metabolic variables (bacterial protein production (BPP), areal primary production (PP) and community respiration (CR)). Degrees of freedom (df); sum of squares (SS); mean square (MS).

Source of variation	df	SS	MS	Pseudo- <i>F</i>	<i>p</i> (perm)	Unique perms
Time	9	92.128	10.236	6.73	0.001	9931
$f\text{CO}_2^b$	4	16.044	4.011	2.637	0.023	9944
Time $\times$ $f\text{CO}_2$	36	42.721	1.1867	0.78018	0.792	9904
Residuals	10	15.21	1.521			
Total	59	182.46				

<sup>a</sup> Permutation was performed with unrestricted permutation of raw data. <sup>b</sup> Pairwise test could only be performed for control mesocosms ( $n = 2$ ) with each  $f\text{CO}_2$  treatment ( $n = 1$ ), due to missing replication for each  $f\text{CO}_2$  treatment. Pairwise comparisons were significant between control and all  $f\text{CO}_2$  treatments ( $p_{\text{perm}} < 0.04$ ).



**Figure 3.** (a) Biovolume of *Synechococcus* spp. [ $\times 10^5 \mu\text{m}^3 \text{mL}^{-1}$ ] and (b–d) biovolume of picoeukaryote groups I–III (Pico I–III) [ $\times 10^5 \mu\text{m}^3 \text{mL}^{-1}$ ] during the course of the experiment. Colours and symbols indicate average  $f\text{CO}_2$  [ $\mu\text{atm}$ ] between  $t_1$  and  $t_{43}$ .

tial variations along the  $f\text{CO}_2$  gradient ( $f\text{CO}_2\text{-}F_{4,10} = 2.6$ ,  $p = 0.02$ ). PCA ordination of the same chemical dataset strongly reflects the temporal pattern, separating the initial time points before  $t_{11}$  from other time points of the experiment along the first PCA axis (Fig. 6). Thereby, eigenvectors of TPC and PON were highest on PCA axis 1 (Table 2). PCA axis two was mainly characterized by high eigenvectors of dissolved phosphate as well as dissolved and particulate silica. The first two PCA axes explained 69 % of variation, and cumulatively 80 % including axis three (Table 2).

### 3.5 Multivariate characterization of metabolic parameters

PERMANOVA on the resemblance matrix of normalized metabolic variables (BPP, areal PP, CR) revealed significant temporal (Time- $F_{9,10} = 6.7$ ,  $p = 0.0002$ ) and spatial variations along the  $f\text{CO}_2$ -gradient ( $f\text{CO}_2\text{-}F_{4,10} = 2.64$ ,  $p < 0.03$ ) (Table 3). DistLM identified significant effects of temperature ( $p < 0.03$ ), phosphate ( $p < 0.02$ ), DOC

**Table 4.** Summary of a DistLM procedure for modelling the relationship between physicochemical variables and a resemblance matrix based on a multivariate assemblage comprising normalized data of bacterial protein production (BPP), areal primary production (PP) and community respiration (CR). Redundant physicochemical variables were removed prior to analysis. Therefore, PON and pH were excluded from the subsequent analysis due to high correlations ( $r_s > 0.9$ ) to TPC and  $f\text{CO}_2$  respectively.

Variable	SS (trace)	Pseudo- <i>F</i>	<i>p</i>	Prop.
$f\text{CO}_2$	5.0551	1.6527	0.1759	0.03
Temp*	10.209	3.4376	0.0229	0.055
PAR*	6.2466	2.056	0.1067	0.034
DOC*	8.6228	2.8769	0.0474	0.047
TDN	4.7628	1.5545	0.1984	0.026
Phosphate*	12.319	4.1994	0.0111	0.068
DSi	0.26167	0.083	0.9648	0.001
TPC	7.7827	2.5842	0.0613	0.004
POP	5.0171	1.6399	0.1818	0.027
BSi	11.688	3.9696	0.0111	0.064

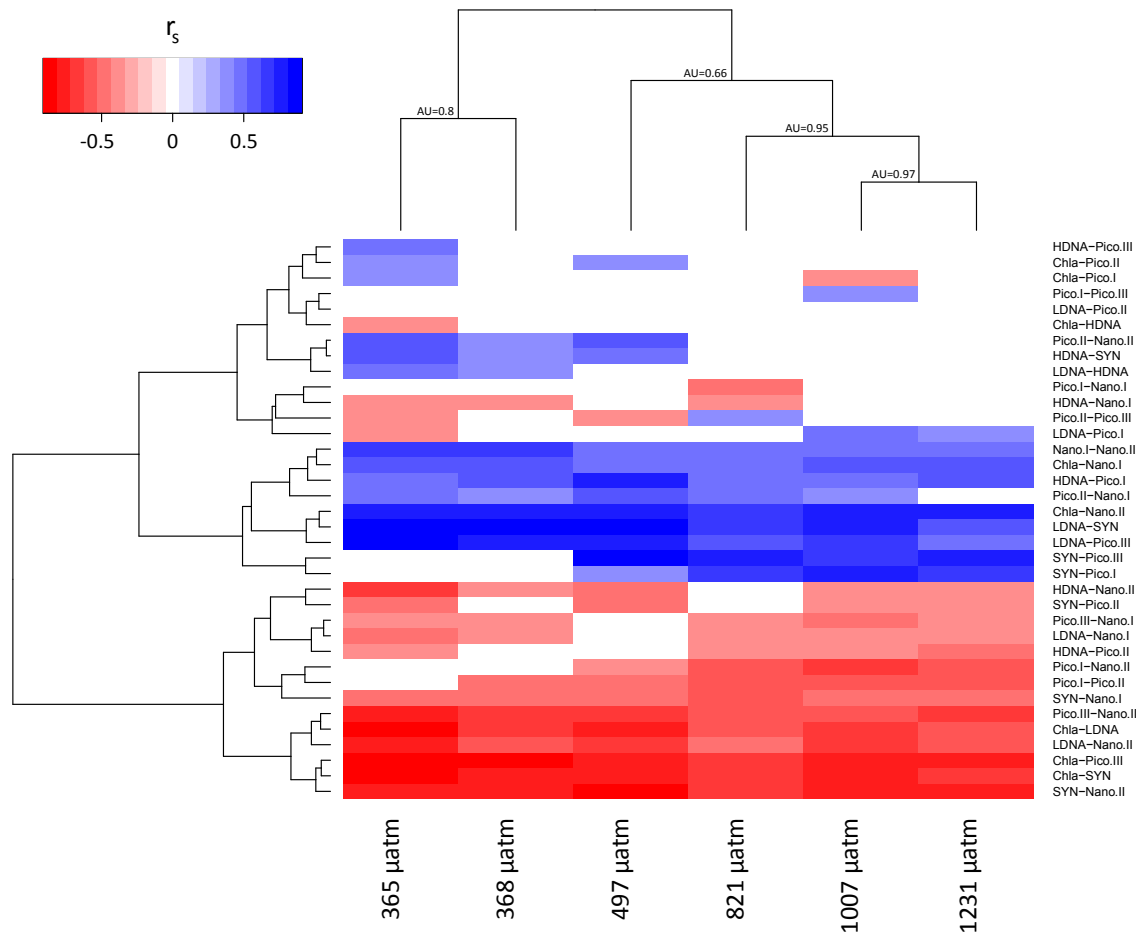
\* Variables selected in stepwise procedure based on AIC.

( $p < 0.05$ ) and BSi ( $p < 0.02$ ) on the multivariate assemblage of metabolic variables (Table 4). The stepwise procedure selects PAR, temperature, DOC and phosphate as determining factors (AIC = 59.6,  $R^2 = 0.26$ , number of variables = 4). The dbRDA ordination separates the temporal development. Thereby, 92 % of the variability in the fitted model and 24 % of the total variation is explained by the first two dbRDA axes (Fig. 6).

### 3.6 Multivariate characterization of the bacterioplankton and phytoplankton community

PERMANOVA on the resemblance matrix of a multivariate assemblage comprising variables of bacterial and phytoplankton communities (abundances of Pico I–III, Nano I–II, FL bacteria (HDNA, LDNA), PA bacteria, *Synechococcus* spp. and Chl *a*) revealed significant tempo-



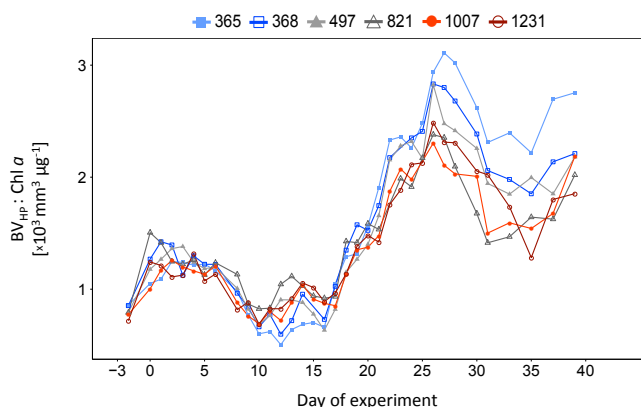


**Figure 4.** Heatmap and cluster analysis based on significant Spearman's rank correlation coefficients calculated for each mesocosm between all possible combinations of abundances between different functional heterotrophic prokaryotic and phytoplanktonic groups (high and low nucleic acid-stained prokaryotic heterotrophs (HDNA : LDNA), *Synechococcus* spp. (SYN), picoeukaryotes I-III (Pico I-III), nanophytoplankton I-II (Nano I-II)) and Chl *a* based on daily measurements between  $t_1$  and  $t_{39}$ . Colours indicate the Spearman's  $r_s$  rank coefficient between two variables.  $P$  values of correlations were corrected for multiple testing according to Benjamini and Hochberg (1995). Uncertainty in hierarchical clustering was assessed with multiscale bootstrap resampling using approximately unbiased (AU)  $p$  values (between 0 and 1; Suzuki and Shimodeira, 2015). Clusters of the three highest  $f\text{CO}_2$  treatments are significantly different at the 0.05 level. Numbers indicate the  $f\text{CO}_2$  treatment with average  $f\text{CO}_2$  [ $\mu\text{atm}$ ] between  $t_1$  and  $t_{43}$ .

ral (Time- $F_{9,10} = 56.8$ ,  $p = 0.0001$ ) and spatial variations along the  $f\text{CO}_2$ -gradient ( $f\text{CO}_2$ - $F_{4,10} = 14.9$ ,  $p = 0.0001$ ) (Table 5). DistLM identified significant effects of  $f\text{CO}_2$  ( $p < 0.02$ ), temperature ( $p < 0.001$ ), phosphate ( $p < 0.003$ ), TPC ( $p < 0.001$ ), BSi ( $p < 0.001$ ) and POP ( $p < 0.001$ ) on the multivariate assemblage of bacterial and phytoplankton communities (Table 6). The stepwise procedure selects  $f\text{CO}_2$ , temperature, TPC and phosphate as determining factors (AIC = 67.2;  $R^2 = 0.44$ ; number of variables = 4). The dbrDA reveals a separation along the gradient of  $f\text{CO}_2$  on the second dbrDA axis. The first dbrDA axis represents the overall temporal development. Thereby, the first two dbrDA axes capture 74 % of the variability in the fitted model and 32 % of the total variation.

#### 4 Discussion

Although OA and its ecological consequences have received growing recognition during the last decade (Riebesell and Gattuso, 2015), surprisingly little is known about the ecological effects on heterotrophic bacterial biomass or the production and coupling of bacterio- and phytoplankton at nutrient-limited conditions. Previous experiments were, for the most part, conducted during productive phases of the year (e.g. phytoplankton blooms), under eutrophic conditions (e.g. coastal areas) or with nutrient additions (e.g. Grossart et al., 2006a; Allgaier et al., 2008; Brussaard et al., 2013; Bach et al., 2016). However, large parts of the oceans are nutrient limited or experience extended nutrient-limited periods during the year (Moore et al., 2013). Thus, we conducted our



**Figure 5.** Ratio of heterotrophic prokaryotic biovolume and Chl *a* ( $BV_{HP} : Chl\ a$ ) during the course of the experiment. PA  $BV_{HP}$  was interpolated using splines with software R (R Core Team, 2016) for time-points, where no data were available. Colours and symbols indicate average  $fCO_2$  [ $\mu atm$ ] between  $t1$  and  $t43$ .

experiment in July–August, when nutrients and phytoplankton production were relatively low in the northeastern Baltic Sea (Hoikkala et al., 2009; Lignell et al., 2008), and exposed a natural plankton community to different levels of  $CO_2$ .

#### 4.1 Phytoplankton–bacterioplankton coupling at low nutrient conditions

Heterotrophic bacteria are important recyclers of autochthonous DOM in aquatic systems and play an important role in nutrient remineralization in natural plankton assemblages (Kirchman, 1994). BV and production of heterotrophic bacteria are highly dependent on quantity and quality of phytoplankton-derived organic carbon and are usually tightly related to phytoplankton development (e.g. Grossart et al., 2006b; Allgaier et al., 2008). During this study, low nitrogen availability limited overall autotrophic production (Paul et al., 2015; Nausch et al., 2016). This resulted in a post spring bloom phytoplankton community, dominated by picophytoplankton (Paul et al., 2015). This is consistent with previous reports of picophytoplankton accounting for a large fraction of total phytoplankton biomass in oligotrophic, nutrient-poor systems (e.g. Agawin et al., 2000). Chl *a* dynamics indicated two minor blooms of larger phytoplankton during the first half of the experiment, although picophytoplankton still accounted for mostly > 50 % of the total Chl *a* during this period (Paul et al., 2015; Spilling et al., 2016b). The phytoplankton development was also reflected in the PCA ordination of dissolved and particulate nutrients, clearly separating the preceding period before  $t11$ , including the first peak of Chl *a*, from the other observations during the experiment (Fig. 6). The separation was primarily driven by concentrations of particulate matter (Table 2), which decreased until  $t11$  and subsequently sank out of the water column (Paul et al., 2015).

Bacterial BV and BPP paralleled phytoplankton development during this period. With the decay of the initial phytoplankton bloom, a second bloom event resulted, comprised primarily of nanophytoplankton and picophytoplankton (Crawford et al., 2016). A decrease in nanophytoplankton BV and Chl *a* concentrations after  $t16/t17$  benefited both FL heterotrophic bacteria and picophotoautotrophs. The increased availability of DOM, resulting from cell lysis and remineralization of POM, was associated with increases in the BV of both groups and bacterial production levels (Figs. 1, S1). We attributed these increases to the cells of Picoplankton, which, due to their high volume-to-surface ratio as well as a small boundary layer surrounding these cells, are generally favoured compared to larger cells in terms of resource acquisition at low nutrient conditions (Raven, 1998; Moore et al., 2013). If cell size is the major factor determining the access to dissolved nitrogen and phosphorous, bacteria should be able to compete equally or better with picophytoplankton at low concentrations (Suttle et al., 1990; Drakare et al., 2003). However, when phytoplankton is restricted in growth due to the lack of mineral nutrients, a strong commensalistic relationship between phytoplanktonic DOM production and bacterioplanktonic DOM utilization may evolve (Azam et al., 1983; Bratbak and Thingstad, 1985; Joint et al., 2002). Although heterotrophic microbes may indirectly limit primary production by depriving phytoplankton of nutrients, they would not be able to outcompete autotrophs completely since this would remove their source of substrates for carbon and energy (Bratbak and Thingstad, 1985; Joint et al., 2002). Such a relationship might explain the paralleled increase in FL bacterial and picophytoplankton BV.

PA bacteria are typically impacted to a lesser extent by nutrient limitation due to consistently higher nutrient availability at particle surfaces (e.g. Grossart and Simon, 1993). This was reflected in this study by the maintenance of high csBPP rates associated with PA heterotrophic bacteria throughout the experiment. Overall, PA bacteria contributed only a minor fraction (at maximum  $10 \pm 0.7\%$ ) to the overall bacterial BV, which is typical for oligotrophic or mesotrophic ecosystems (Lapoussière et al., 2011). Nevertheless, their substantial contribution to overall BPP emphasizes their importance, especially during such low productive periods (e.g. Grossart, 2010). PA heterotrophic bacteria are essential for the remineralization of nutrients from autotrophic biomass, which would otherwise sink down from surface waters (Grossart, 2010). Leakage of hydrolysis products and the attachment and detachment of bacteria to and from particles stimulate production amongst free-living bacteria (Smith et al., 1992; Grossart, 2010) and picophytoplankton.

**Table 5.** Results of two-factor permutational multivariate analysis of variance (PERMANOVA)<sup>a</sup> on a resemblance matrix (Bray–Curtis similarity coefficient) based on  $\log(X+1)$  transformed abundances of Pico I–III, Nano I–II, FL bacteria (HDNA, LDNA), PA bacteria, SYN and Chl *a*. Degrees of freedom (df); sum of squares (SS); mean square (MS).

Source of variation	df	SS	MS	Pseudo- <i>F</i>	<i>p</i> (perm)	Unique perms
Time	9	201.83	22.426	56.754	0.0001	9923
$f\text{CO}_2^b$	4	23.631	5.9077	14.951	0.0001	9940
Time $\times$ $f\text{CO}_2$	36	19.859	0.55164	1.396	0.151	9915
Residuals	10	3.9515	0.39515			
Total	59	271.01				

<sup>a</sup> Permutation was performed with unrestricted permutation of raw data. <sup>b</sup> Pairwise test could only be performed for control mesocosms ( $n=2$ ) with each  $f\text{CO}_2$  treatment ( $n=1$ ), due to missing replication for each  $f\text{CO}_2$  treatment. Pairwise comparisons were significant between control and all  $f\text{CO}_2$  treatments ( $p_{\text{perm}} < 0.01$ ).

**Table 6.** Summary of a DistLM procedure for modelling the relationship between physicochemical variables and a multivariate assemblage, comprising variables of the bacterial and phytoplankton community. The resemblance matrix (Bray–Curtis similarity coefficient) was based on  $\log(X+1)$  transformed abundances of Pico I–III, Nano I–II, FL bacteria (HDNA, LDNA), PA bacteria, *Synechococcus* spp. and Chl *a*. Redundant physicochemical variables were removed prior analysis. Therefore, PON and pH were excluded from the subsequent analysis due to high correlations ( $r_s > 0.9$ ) to TPC and  $f\text{CO}_2$  respectively.

Variable	SS (trace)	Pseudo- <i>F</i>	<i>p</i>	Prop.
$f\text{CO}_2^*$	20.469	4.7386	0.0119	0.075
Temp*	51.838	13.718	0.0001	0.191
PAR	10.791	2.4051	0.0813	0.039
DOC	11.14	2.4864	0.0769	0.041
TDN	9.4456	2.0945	0.1078	0.034
Phosphate*	25.649	6.063	0.0029	0.095
DSi	9.5766	2.1246	0.103	0.035
TPC*	36.038	8.8955	0.0002	0.133
POP	52.171	13.827	0.0001	0.193
BSi	36.439	9.01	0.0005	0.134

\* Variables selected in stepwise procedure based on AIC.

#### 4.2 Effects of $f\text{CO}_2$ /pH on phytoplankton–bacterioplankton coupling at low nutrient conditions

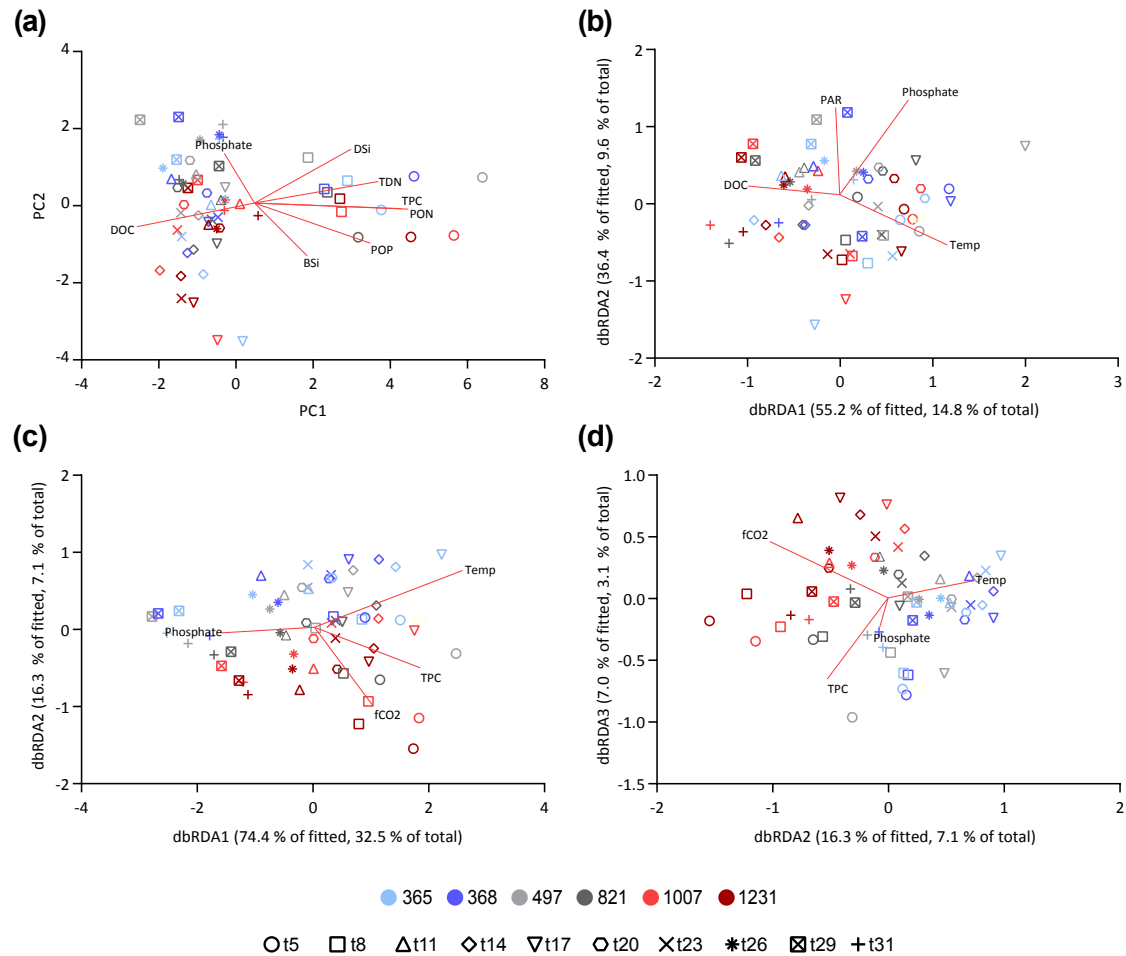
The response of heterotrophic bacteria to changes in  $f\text{CO}_2$  has been previously shown to be related to phytoplankton rather than being a direct effect of pH or  $\text{CO}_2$  (e.g. Allgaier et al., 2008; Grossart et al., 2006a). Here, neither BPP nor BV of neither FL nor PA bacteria suggested a direct effect of  $\text{CO}_2$  on these variables (PERMANOVA). Differences in FL bacterial BV, BPP and the ratio of HDNA : LDNA occurred along the gradient of  $f\text{CO}_2$ , but were limited to short time periods. Furthermore, these changes were not consistent with  $f\text{CO}_2$ , resulting in both increases and decreases of a particular variable at specific times (Fig. 1). Periods where  $f\text{CO}_2$ -related

effects were apparent comprised periods with high organic matter turnover (e.g. breakdown of Chl *a* maximum). However, Paul et al. (2015) could not reveal any effect of  $f\text{CO}_2$  on the export of carbon, neither across the study period nor at individual time points. Thus, it is reasonable to assume that these small  $f\text{CO}_2$ -related differences in bacterial variables are a consequence of other altered components of the aquatic food web, and do not necessarily manifest as changes in carbon export.

Given the inability to relate individual aspects of microbial metabolism or community composition to  $f\text{CO}_2$  concentrations, we sought to determine whether an impact was evident using a multivariate approach. Chemical, metabolic and community matrices exhibited large variations in relation to a strong temporal effect throughout the whole sampling period ( $p < 0.01$ ; Tables 1, 3, 5). In addition, an effect of the  $f\text{CO}_2$  treatment was also evident in all three multivariate assemblages, albeit explaining far less of the observed variability in chemical and metabolic variables ( $p < 0.03$ ; Tables 1, 3, 5). However, when relating physicochemical to metabolic variables (DistLM, Table 4), neither  $f\text{CO}_2$  nor pH were suitable to explaining the observed variability. In contrast,  $f\text{CO}_2$  contributed to explaining the variability amongst the bacterioplankton–phytoplankton community dynamics (DistLM, Table 6). Taken together, this suggests that effects of  $f\text{CO}_2$  treatments manifest indirectly, through either altering physicochemical parameters or more likely the composition of the microbial community with possible but so far hidden consequences for microbial metabolism.

#### 4.3 $f\text{CO}_2$ /pH effects on phytoplankton indirectly alter phytoplankton–bacterioplankton coupling at low-nutrient conditions

Autotrophic organisms can be fertilized by an enhanced  $\text{CO}_2$  availability, altering growth conditions of phytoplankton and increasing the production of particulate (POM) and dissolved organic matter (DOM; Hein and Sand-Jensen, 1997; Riebesell et al., 2007). As a consequence of this increased photosynthetic fixation rate, both quantity and quality of dis-



**Figure 6.** (a) First and second axis of a principal component analysis (PCA) calculated on normalized variables of dissolved and particulate nutrients ( $n = 60$ ). The set of variables and the eigenvectors of the first four axes can be derived from Table 2. (b) Ordination of a distance-based redundancy analysis (dbRDA) for visual interpretation of distance-based linear modelling (DistLM) between physical/chemical predictor variables and metabolic variables as well as (c–d) abundances of functional bacterial and phytoplankton groups. A table comprising the set of variables used for DistLM can be derived from the Supplement (Table S1).

solved organic matter (DOM) available for heterotrophic bacteria are impacted, with potential implications for the nature of coupling between phytoplankton and bacterioplankton at low-nutrient conditions (Azam et al., 1983; Bratbak and Thingstad, 1985). So far, CO<sub>2</sub> enrichment experiments examining natural plankton assemblages (e.g. Engel et al., 2005; Riebesell et al., 2007; Bach et al., 2016) have not revealed a consistent pattern of species response or primary production to elevated CO<sub>2</sub>. During this experiment, Spilling et al. (2016a) could not detect any effect of increased CO<sub>2</sub> on total primary production, even though Crawford et al. (2016) reported effects of CO<sub>2</sub> on several groups of picophytoplankton. Although one larger picoeukaryote (Pico III) was negatively impacted by  $f\text{CO}_2$  during our study, two small picoeukaryotes (Pico I, Pico II) benefitted from the CO<sub>2</sub> addition, yielding significantly higher growth rates and BVs at higher  $f\text{CO}_2$  (Crawford et al., 2016). This is consistent

with recent evidence suggesting a positive impact of enhanced  $f\text{CO}_2$  on the abundance of small picoeukaryotic phytoplankton (Brussaard et al., 2013; Newbold et al., 2012; Sala et al., 2015). Both picoeukaryotic groups were identified as variables explaining the separation along the  $f\text{CO}_2$  gradient on the second and third dbRDA axes in the DistLM ordination of the bacteria–phytoplankton community. Specifically, Pico I was highly negatively correlated ( $r_s = -0.67$ ) to dbRDA axis two. However, dbRDA also indicated opposing effects of  $f\text{CO}_2$  on Pico II ( $r_s = 0.54$ ) and HDNA prokaryotes ( $r_s = -0.31$ ), being positively or negatively correlated with axis three. Indeed, sharp increases in  $\text{BV}_{\text{Pico II}}$  at high  $f\text{CO}_2$  between  $t14$  and  $t17$  were associated with decreases in  $\text{BV}_{\text{HDNA}}$ .

Although we are not able to draw solid conclusions on the interaction of these two particular groups of organisms, a cluster analysis of pairwise Spearman correlations between

functional groups of bacteria and phytoplankton revealed a distinct clustering with mesocosms based on  $f\text{CO}_2$  levels (Fig. 4). We also detected a change in the ratio of heterotrophic bacterial BV to Chl *a* between the different  $f\text{CO}_2$  treatments, though this change was not visible for the entire study duration and not consistent with  $f\text{CO}_2$ . These results strongly suggest that trophic interactions between functional groups of bacteria and phytoplankton might be changing in a future acidified ocean.

In nutrient-poor systems, variable growth rates of phytoplankton, DOM quality and quantity, and also loss of phytoplankton and bacterioplankton due to grazing or viral lysis may potentially contribute to this observed decoupling of phytoplankton and bacterioplankton at high  $f\text{CO}_2$  (Azam et al., 1983; Bratbak and Thingstad, 1985; Sheik et al., 2014). The viral shunt or bacterivory may release phytoplankton from competition with bacteria for limited nutrients (e.g. Bratbak and Thingstad, 1985; Caron and Goldman, 1990). How increased  $f\text{CO}_2$  will affect these processes (e.g. viral lysis and bacterial grazing) under nutrient-limited conditions remains uncertain so far. Bacterial grazing by mixotrophs, which would also directly benefit from increased  $\text{CO}_2$  availability (Rose et al., 2009), may provide a mechanism for recycling of inorganic nutrients, otherwise bound in bacterial biomass, as a means for supporting phytoplankton growth (e.g. Mitra et al., 2014). However, other studies examining bacterial grazing under different nutrient conditions reported conflicting positive and negative results of increased  $f\text{CO}_2$  (e.g. Brussaard et al., 2013; Rose et al., 2009). Although we are unable to draw defined conclusions on how this myriad of complex biological processes are impacted by  $f\text{CO}_2$ , it is very likely that there is an impact on trophic interactions which may account for the portion of unexplained variance we observed in our multivariate analyses.

## 5 Conclusions

The use of large-volume mesocosms allowed us to test for multiple  $f\text{CO}_2$ -related effects on dynamics of heterotrophic bacterial activity and their biovolume in a near-realistic ecosystem by including trophic interactions from microorganisms up to zooplankton. Complex interactions between various trophic levels, which can only be properly addressed at the scale of whole ecosystems, are important for understanding and predicting  $f\text{CO}_2$ -induced effects on aquatic food webs and biogeochemistry in a future, acidified ocean. We examined these impacts in a nutrient-depleted system, which is representative for large parts of the oceans (Moore et al., 2013). Heterotrophic bacterial productivity was, for the most part, tightly coupled to the availability of phytoplankton-derived organic matter. When accounting for temporal development and taking into account trophic interactions using multivariate statistics, changes in nutrient composition, metabolic parameters and bacteria–phytoplankton

communities revealed a significant effect of the  $f\text{CO}_2$  treatment. Although not consistent throughout the experiment, differences in the ratio of heterotrophic bacterial BV to Chl *a* during the last half of the experiment suggest that a future ocean will become more autotrophic during low productive periods as a result of altered trophic interactions between functional groups of bacteria and phytoplankton. There is additional support for this conclusion from examining the atmospheric exchange of  $\text{CO}_2$  (Spilling et al., 2016b). During the limited timescale of this study, the observed effects of  $f\text{CO}_2$  did not manifest as altered carbon export (Paul et al., 2015). However, over several years, maintained changes in nutrient cycling, as a consequence of a permanent decoupling between bacteria and phytoplankton, are likely to arise and impact the nature of the carbon pump.

## 6 Data availability

Data of primary production and respiration can be obtained from Spilling et al. (2016c). Other variables from the experiment (e.g. total particulate and dissolved nutrients) can be found in Paul et al. (2016). Flow cytometry data can be obtained from Crawford et al. (2016). Bacterial protein production rates and abundances of particle-associated bacteria can be obtained from Hornick et al. (2016).

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## References

- Agawin, N. S. R., Duarte, C. M., and Agusti, S.: Nutrient and temperature control of the contribution of picoplankton to phytoplankton biomass and production, *Limnol. Oceanogr.*, 45, 591–600, 2000.
- Allgaier, M., Riebesell, U., Vogt, M., Thyraug, R., and Grossart, H.-P.: Coupling of heterotrophic bacteria to phytoplankton bloom development at different  $p\text{CO}_2$  levels: a mesocosm study, *Biogeosciences*, 5, 1007–1022, doi:10.5194/bg-5-1007-2008, 2008.
- Anderson, M. J.: A new method for non-parametric multivariate analysis of variance, *Aust. Ecol.*, 35, 32–46, 2001.
- Anderson, M. J., Gorley, R. N., and Clarke, K. R.: PERMANOVA+ for PRIMER: Guide to Software and Statistical Methods, PRIMER-E, Plymouth, UK, 214 pp., 2008.
- Azam, F.: Microbial Control of Oceanic Carbon Flux: The Plot Thickens, *Science*, 280, 694–696, doi:10.1126/science.280.5364.694, 1998.
- Azam, F., Fenchel, T., Field, J. G., Gray, J. S., Meyer-Reil, L. A., and Thingstad, F.: The Ecological Role of Water-Column Microbes in the Sea, *Mar. Ecol.-Prog. Ser.*, 10, 257–263, 1983.
- Bach, L. T., Taucher, J., Boxhammer, T., Ludwig, A., The Kristineberg KOSMOS Consortium, Achterberg, E. P., Algueró-Muizñiz, M., Anderson, L. G., Bellworthy, J., Büdenbender, J., Czerny, J., Ericson, Y., Esposito, M., Fischer, M., Haunost, M., Hellemann, D., Horn, H. G., Hornick, T., Meyer, J., Sswat, M., Zark, M., and Riebesell, U.: Influence of Ocean Acidification on a Natural Winter-to-Summer Plankton Succession: First Insights from a Long-Term Mesocosm Study Draw Attention to Periods of Low Nutrient Concentrations, *PLoS ONE*, 11, e0159068, doi:10.1371/journal.pone.0159068, 2016.
- Badr, E.-S. A., Achterberg, E. P., Tappin, A. D., Hill, S. J., and Braungardt, C. B.: Determination of dissolved organic nitrogen in natural waters using high-temperature catalytic oxidation, *TrAC-Trend, Anal. Chem.*, 22, 819–827, doi:10.1016/S0165-9936(03)01202-0, 2003.
- Benjamini, Y. and Hochberg, Y.: Controlling the false discovery rate: a practical and powerful approach to multiple testing, *J. Roy. Stat. Soc. B*, 57, 289–300, 1995.
- Berggren, M., Lapierre, J.-F., and del Giorgio, P. A.: Magnitude and regulation of bacterioplankton respiratory quotient across freshwater environmental gradients, *ISME J.*, 6, 984–993, 2012.
- Bratbak, G. and Thingstad, T. F.: Phytoplankton-bacteria interactions: an apparent paradox? Analysis of a model system with both competition and commensalism, *Mar. Ecol.-Prog. Ser.*, 25, 23–30, 1985.
- Brussaard, C. P. D., Noordeloos, A. A. M., Witte, H., Collenteur, M. C. J., Schulz, K., Ludwig, A., and Riebesell, U.: Arctic microbial community dynamics influenced by elevated  $\text{CO}_2$  levels, *Biogeosciences*, 10, 719–731, doi:10.5194/bg-10-719-2013, 2013.
- Caldeira, K. and Wickett, M. E.: Anthropogenic carbon and ocean pH, *Nature*, 425, 365–365, 2003.
- Caron, D. A. and Goldman, J. C.: Protozoan nutrient regeneration, in: *Ecology of marine protozoa*, edited by: Capriulo, G. M., Oxford University Press, New York, 283–306, 1990.
- Clarke, K. R. and Gorley, R. N.: PRIMER v6: User manual/tutorial, PRIMER-E, Plymouth, UK, 115 pp., 2006.
- Crawford, K. J., Brussaard, C. P. D., and Riebesell, U.: Shifts in the microbial community in the Baltic Sea with increasing  $\text{CO}_2$ , *Biogeosciences Discuss.*, doi:10.5194/bg-2015-606, in review, 2016.
- Dickson, A. G., Sabine, C., and Christian, J. (Eds.): Guide to best practices for ocean  $\text{CO}_2$  measurements, *PICES Special Publication 3*, 191 pp., <http://aquaticcommons.org/1443/> (last access: 16 October 2012), 2007.
- Drakare, S., Blomqvist, P., Bergström, A.-K., and Jansson, M.: Relationships between picophytoplankton and environmental variables in lakes along a gradient of water colour and nutrient content, *Freshwater Biol.*, 48, 729–740, 2003.
- Engel, A., Zondervan, I., Aerts, K., Beaufort, L., Benthien, A., Chou, L., Delille, B., Gattuso, J.-P., Harlay, J., and Heemann, C.: Testing the direct effect of  $\text{CO}_2$  concentration on a bloom of the coccolithophorid *Emiliania huxleyi* in mesocosm experiments, *Limnol. Oceanogr.*, 50, 493–507, doi:10.4319/lo.2005.50.2.0493, 2005.
- Gargas, E.: A manual for phytoplankton primary production studies in the Baltic, *The Baltic Marine Biologist*, Hørsholm, Denmark, 88 pp., 1975.
- Grasshoff, K., Ehrhardt, M., Kremling, K., and Almgren, T.: Methods of seawater analysis, Wiley Verlag Chemie GmbH, Weinheim, Germany, 1983.
- Grossart, H.-P.: Ecological consequences of bacterioplankton lifestyles: changes in concepts are needed, *Environ. Microbiol. Rep.*, 2, 706–714, doi:10.1111/j.1758-2229.2010.00179.x, 2010.
- Grossart, H.-P. and Simon, M.: Limnetic macroscopic organic aggregates (lake snow): Occurrence, characteristics, and microbial dynamics in Lake Constance, *Limnol. Oceanogr.*, 38, 532–546, 1993.
- Grossart, H.-P., Allgaier, M., Passow, U., and Riebesell, U.: Testing the effect of  $\text{CO}_2$  concentration on the dynamics of marine heterotrophic bacterioplankton, *Limnol. Oceanogr.*, 51, 1–11, 2006a.
- Grossart, H.-P., Czub, G., and Simon, M.: Specific interactions of planktonic algae and bacteria: Implications for aggregation and organic matter cycling in the sea, *Environ. Microbiol.*, 8, 1074–1084, 2006b.
- Hagström, Å., Larsson, U., Hörstedt, P., and Normark, S.: Frequency of Dividing Cells, a New Approach to the Determination of Bacterial Growth Rates in Aquatic Environments, *Appl. Environ. Microbiol.*, 37, 805–812, 1979.
- Hansen, H. P. and Koroleff, F.: Determination of nutrients, in: *Methods of Seawater Analysis*, edited by: Grasshoff, K., Kremling, K., and Ehrhardt, M., Wiley Verlag Chemie GmbH, Weinheim, Germany, 159–228, 1999.
- Harrell Jr., F. E., Dupont, C. and many others: Hmisc: Harrell Miscellaneous, R package version 3.17-4, <http://CRAN.R-project.org/package=Hmisc>, last access: 2 September 2016.
- Hein, M. and Sand-Jensen, K.:  $\text{CO}_2$  increases oceanic primary production, *Nature*, 388, 526–527, doi:10.1038/41457, 1997.
- Hobbie, J. E., Daley, R. J., and Jasper, S.: Use of nucleopore filters for counting bacteria by fluorescence microscopy, *Appl. Environ. Microbiol.*, 33, 1225–1228, 1977.
- Hoikkala, L., Aarnos, H., and Lignell, R.: Changes in Nutrient and Carbon Availability and Temperature as Factors Controlling Bacterial Growth in the Northern Baltic Sea., *Estuar. Coast.*, 32, 720–733, doi:10.1007/s12237-009-9154-z, 2009.



- Hornick, T., Bach, L. T., Crawford, K. J., Spilling, K., Achterberg, E. P., Woodhouse, J. N., Schulz, K. G., Brussaard, C. P. D., Riebesell, U., and Grossart, H.-P.: KOSMOS Finland 2012 mesocosm study: Size-fractionated bacterial protein production (BPP) of free-living and particle associated bacteria and abundance of particle associated heterotrophic prokaryotes, doi:10.1594/PANGAEA.868621, 2016.
- Joint, I., Henriksen, P., Fonnes, G. A., Bourne, D., Thingstad, T. F., and Riemann, B.: Competition for inorganic nutrients between phytoplankton and bacterioplankton in nutrient manipulated mesocosms, *Aquat. Microb. Ecol.*, 29, 145–159, 2002.
- Kirchman, D. L.: The Uptake of Inorganic Nutrients by Heterotrophic Bacteria, *Microb. Ecol.*, 28, 255–271, 1994
- Kirchman, D. L., Elifantz, H., Dittel, A. I., Malmstrom, R. R., and Cottrell, M. T.: Standing stocks and activity of Archaea and Bacteria in the western Arctic Ocean, *Limnol. Oceanogr.*, 52, 495–507, 2007.
- Kuparinen, J. and Heinänen, A.: Inorganic Nutrient and Carbon Controlled Bacterioplankton Growth in the Baltic Sea, *Estuar. Coast. Shelf Sci.*, 37, 271–285, 1993.
- Lapoussière, A., Michel, C., Starr, M., Gosselin, M., and Poulin, M.: Role of free-living and particle-attached bacteria in the recycling and export of organic material in the Hudson Bay system, *J. Mar. Syst.*, 88, 434–445, 2011.
- Legendre, P. and Anderson, M. J.: Distance-based redundancy analysis: testing multispecies responses in multifactorial ecological experiments, *Ecol. Monogr.*, 69, 1–24, 1999.
- Lignell, R., Hoikkala, L., and Lahtinen, T.: Effects of inorganic nutrients, glucose and solar radiation treatments on bacterial growth and exploitation of dissolved organic carbon and nitrogen in the northern Baltic Sea, *Aquat. Microb. Ecol.*, 51, 209–221, 2008.
- Lueker, T. J., Dickson, A. G., and Keeling, C. D.: Ocean  $p\text{CO}_2$  calculated from dissolved inorganic carbon, alkalinity, and equations for  $K_1$  and  $K_2$ : validation based on laboratory measurements of  $\text{CO}_2$  in gas and seawater at equilibrium, *Mar. Chem.*, 70, 105–119, doi:10.1016/S0304-4203(00)00022-0, 2000.
- Maat, D. S., Crawford, K. J., Timmermans, K. R., and Brussaard, C. P. D.: Elevated  $\text{CO}_2$  and Phosphate Limitation Favor *Micromonas pusilla* through Stimulated Growth and Reduced Viral Impact, *Appl. Environ. Microbiol.*, 80, 3119–3127, doi:10.1128/AEM.03639-13, 2014.
- Marie, D., Brussaard, C. P. D., Thyraug, R., Bratbak, G., and Vault, D.: Enumeration of marine viruses in culture and natural samples by flow cytometry, *Appl. Environ. Microbiol.*, 65, 45–52, 1999.
- Massana, R., Gasol, J. M., Bjørnsen, P. K., Blackburn, N., Hagström, Å., Hietanen, S., Hygum, B. H., Kuparinen, J. and Pedrós-Alió, C.: Measurement of bacterial size via image analysis of epifluorescence preparations: description of an inexpensive system and solutions to some of the most common problems, *Sci. Mar.*, 61, 397–407, 1997.
- Mehrbach, C., Culbertson, C. H., Hawley, J. E., and Pytkowicz, R. M.: Measurement of apparent dissociation constants of carbonic acid in seawater at atmospheric pressure, *Limnol. Oceanogr.*, 18, 897–907, 1973.
- Mitra, A., Flynn, K. J., Burkholder, J. M., Berge, T., Calbet, A., Raven, J. A., Granéli, E., Glibert, P. M., Hansen, P. J., Stoecker, D. K., Thingstad, F., Tillmann, U., Våge, S., Wilken, S., and Zubkov, M. V.: The role of mixotrophic protists in the biological carbon pump, *Biogeosciences*, 11, 995–1005, doi:10.5194/bg-11-995-2014, 2014.
- Moore, C. M., Mills, M. M., Arrigo, K. R., Berman-Frank, I., Bopp, L., Boyd, P. W., Galbraith, E. D., Geider, R. J., Guieu, C., Jaccard, S. L., Jickells, T. D., La Roche, J., Lenton, T. M., Mahowald, N. M., Marañón, E., Marinov, I., Moore, J. K., Nakatsuka, T., Oschlies, A., Sato, M. A., Thingstad, T. F., Tsuda, A., and Ulloa, O.: Processes and patterns of oceanic nutrient limitation, *Nat. Geosci.*, 6, 701–710, doi:10.1038/NNGEO1765, 2013.
- Nausch, M., Bach, L. T., Czerny, J., Goldstein, J., Grossart, H.-P., Hellemann, D., Hornick, T., Achterberg, E. P., Schulz, K.-G., and Riebesell, U.: Effects of  $\text{CO}_2$  perturbation on phosphorus pool sizes and uptake in a mesocosm experiment during a low productive summer season in the northern Baltic Sea, *Biogeosciences*, 13, 3035–3050, doi:10.5194/bg-13-3035-2016, 2016.
- Newbold, L., Oliver, A. E., Booth, T., Tiwari, B., DeSantis, T., Maguire, M., Andersen, G., van der Gast, C. J., and Whiteley, A. S.: The response of marine picoplankton to ocean acidification, *Environ. Microbiol.*, 14, 2293–2307, 2012.
- Oksanen, J., Blanchet, F. G., Friendly, M., Kindt, R., Legendre, P., McGinn, D., Minchin, P. R., O’Hara, R. B., Simpson, G. L., Solymos, P., Stevens, M. H. H., Szoecs, E., and Wagner, H.: vegan: Community Ecology Package, R package version 2.4-0, <https://CRAN.R-project.org/package=vegan>, last access: 2 September 2016.
- Patey, M. D., Rijkenberg, M. J. A., Statham, P. J., Stinchcombe, M. C., Achterberg, E. P., and Mowlem, M.: Determination of nitrate and phosphate in seawater at nanomolar concentrations, *TrAC-Trend, Anal. Chem.*, 27, 169–182, doi:10.1016/j.trac.2007.12.006, 2008.
- Paul, A. J., Bach, L. T., Schulz, K.-G., Boxhammer, T., Czerny, J., Achterberg, E. P., Hellemann, D., Trense, Y., Nausch, M., Sswat, M., and Riebesell, U.: Effect of elevated  $\text{CO}_2$  on organic matter pools and fluxes in a summer Baltic Sea plankton community, *Biogeosciences*, 12, 6181–6203, doi:10.5194/bg-12-6181-2015, 2015.
- Paul, A., Schulz, K. G., Achterberg, E. P., Hellemann, D., Nausch, M., Boxhammer, T., Bach, L. T., and Trense, Y.: KOSMOS Finland 2012 mesocosm study: carbonate chemistry, particulate and dissolved matter pools, and phytoplankton community composition using marker pigments (CHEMTAX), doi:10.1594/PANGAEA.863032, 2016.
- Porter, K. G. and Feig, Y. S.: Dapi for identifying and counting aquatic microflora, *Limnol. Oceanogr.*, 25, 943–948, 1980.
- Raven, J. A.: The twelfth Tansley Lecture. Small is beautiful: the picophytoplankton, *Funct. Ecol.*, 12, 503–513, 1998.
- R Core Team R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria, URL <http://www.R-project.org/>, last access: 2 September 2016.
- Riebesell, U. and Gattuso, J.-P.: Lessons learned from ocean acidification research, Reflection on the rapidly growing field of ocean acidification research highlights priorities for future research on the changing ocean, *Nature Climate Change*, 5, 12–14, doi:10.1038/nclimate2456, 2015.
- Riebesell, U., Schulz, K. G., Bellerby, R. G. J., Botros, M., Fritsche, P., Meyerhöfer, M., Neill, C., Nondal, G., Oschlies, A., Wohlers, J., and Zöllner, E.: Enhanced biological carbon consumption in a high  $\text{CO}_2$  ocean, *Nat. Lett.*, 450, 545–548, doi:10.1038/nature06267, 2007.

- Rieck, A., Herlemann, D. P. R., Jürgens, K., and Grossart, H.-P.: Particle-Associated Differ from Free-Living Bacteria in Surface Waters of the Baltic Sea, *Front. Microbiol.*, 6, doi:10.3389/fmicb.2015.01297, 2015.
- Rose, J. M., Feng, Y., Gobler, C. J., Gutierrez, R., Hare, C. E., Leblanc, K., and Hutchins, D. A.: Effects of increased  $p\text{CO}_2$  and temperature on the North Atlantic spring bloom, II. Microzooplankton abundance and grazing, *Mar. Ecol.-Prog. Ser.*, 388, 27–40, 2009.
- Sala, M. M., Aparicio, F. L., Balagué, V., Boras, J. A., Borull, E., Cardelús, C., Cros, L., Gomes, A., López-Sanz, A., Malits, A., Martínez, R. A., Mestre, M., Movilla, J., Sarmiento, H., Vázquez-Dominguez, E., Vaqué, D., Pinhassi, J., Calbet, A., Calvo, E., Gasol, J. M., Pelejero, C., and Marrasé, C.: Contrasting effects of ocean acidification on the microbial food web under different trophic conditions, *ICES J. Mar. Sci.*, 73, 670–679, doi:10.1093/icesjms/fsv130, 2015.
- Sharp, J.: Improved analysis for particulate organic carbon and nitrogen from seawater, *Limnol. Oceanogr.*, 19, 984–989, 1974.
- Sheik, A. R., Brussaard, C. P. D., Lavik, G., Lam, P., Musat, N., Krupke, A., Littmann, S., Strous, M., and Kuypers M. M. M.: Responses of the coastal bacterial community to viral infection of the algae *Phaeocystis globosa*, *ISME J.*, 8, 212–225, doi:10.1038/ismej.2013.135, 2014.
- Simon, M. and Azam, F.: Protein content and protein synthesis rates of planktonic marine bacteria, *Mar. Ecol.-Prog. Ser.*, 51, 201–213, 1989.
- Smith, D. C., Simon, M., Alldredge, A. L., and Azam, F.: Intense hydrolytic enzyme activity on marine aggregates and implications for rapid particle dissolution, *Nature*, 359, 139–142, 1992.
- Spilling, K., Paul, A. J., Virkkala, N., Hastings, T., Lischka, S., Stuhr, A., Bermúdez, R., Czerny, J., Boxhammer, T., Schulz, K. G., Ludwig, A., and Riebesell, U.: Ocean acidification decreases plankton respiration: evidence from a mesocosm experiment, *Biogeosciences*, 13, 4707–4719, doi:10.5194/bg-13-4707-2016, 2016a.
- Spilling, K., Schulz, K. G., Paul, A. J., Boxhammer, T., Achterberg, E. P., Hornick, T., Lischka, S., Stuhr, A., Bermúdez, R., Czerny, J., Crawford, K., Brussaard, C. P. D., Grossart, H.-P., and Riebesell, U.: Effects of ocean acidification on pelagic carbon fluxes in a mesocosm experiment, *Biogeosciences*, 13, 6081–6093, doi:10.5194/bg-13-6081-2016, 2016b.
- Spilling, K., Paul, A., Virkkala, N., Hastings, T., Lischka, S., Stuhr, A., Bermúdez, R., Czerny, J., Boxhammer, T., Schulz, K. G., Ludwig, A., and Riebesell, U.: KOSMOS Finland 2012 mesocosm study: primary production and respiration, doi:10.1594/PANGAEA.863933, 2016c.
- Steeman-Nielsen, E.: The use of radioactive carbon for measuring organic production in the sea, *J. Cons. Int. Explor. Mer.*, 18, 117–140, 1952.
- Suttle, C. A., Fuhrman, J. A., and Capone, D. G.: Rapid ammonium cycling and concentration-dependent partitioning of ammonium and phosphate: Implications for carbon transfer in planktonic communities, *Limnol. Oceanogr.*, 35, 424–433, 1990.
- Suzuki, R. and Shimodaira, H.: pvcust: Hierarchical Clustering with p-values via Multiscale Bootstrap Resampling, R package version 2.0-0., <https://CRAN.R-project.org/package=pvcust> (last access: 2 September 2016), 2015.
- Taylor, A. R., Brownlee, C., and Wheeler, G. L.: Proton channels in algae: reasons to be excited, *Trends Plant Sci.*, 17, 675–684, doi:10.1016/j.tplants.2012.06.009, 2012.
- Thingstad, T. F., Hagström, Å., and Rassoulzadegan, F.: Accumulation of degradable DOC in surface waters: Is it caused by a malfunctioning microbial loop?, *Limnol. Oceanogr.*, 42, 398–404, 1997.
- Thingstad, T. F., Bellerby, R. G. J., Bratbak, G., Borsheim, K. Y., Egge, J. K., Heldal, M., Larsen, A., Neill, C., Nejtgaard, J., Norland, S., Sandaa, R.-A., Skjoldal, E. F., Tanaka, T., Thyrrhaug, R., and Töpper, B.: Counterintuitive carbon-to-nutrient coupling in an Arctic pelagic ecosystem, *Nat. Lett.*, 455, 387–391, doi:10.1038/nature07235, 2008.
- Toggweiler, J. R.: Carbon overconsumption, *Nature*, 363, 210–211, 1993.
- Warnes, G. R., Bolker, B., Bonebakker, L., Gentleman, R., Liaw, W. H. A., Lumley, T., Maechler, M., Magnusson, A., Moeller, S., Schwartz, M., and Venables, B.: gplots: Various R Programming Tools for Plotting Data, R package version 3.0.1, <https://CRAN.R-project.org/package=gplots>, last access: 2 September 2016.
- Welschmeyer, N. A.: Fluorometric analysis of chlorophyll a in the presence of chlorophyll b and pheopigments, *Limnol. Oceanogr.*, 39, 1985–1992, doi:10.4319/lo.1994.39.8.1985, 1994.
- Wickham, H.: ggplot2: Elegant graphics for data analysis, Springer-Verlag New York, 2009.