

Smith ScholarWorks

Biological Sciences: Faculty Publications

Biological Sciences

11-14-2022

Extracellular DNAses Facilitate Antagonism and Coexistence in Bacterial Competitor-Sensing Interference Competition

Aoi Ogawa Smith College

Christophe Golé Smith College, cgole@smith.edu

Maria Bermudez Smith College

Odrine Habarugira Smith College

Gabrielle Joslin Smith College

See next page for additional authors

Follow this and additional works at: https://scholarworks.smith.edu/bio_facpubs

Part of the Biology Commons, and the Data Science Commons

Recommended Citation

Ogawa, Aoi; Golé, Christophe; Bermudez, Maria; Habarugira, Odrine; Joslin, Gabrielle; McCain, Taylor; Mineo, Autumn; Wise, Jennifer; Xiong, Julie; Yan, Katherine; and Vriezen, Jan A.C., "Extracellular DNAses Facilitate Antagonism and Coexistence in Bacterial Competitor-Sensing Interference Competition" (2022). Biological Sciences: Faculty Publications, Smith College, Northampton, MA. https://scholarworks.smith.edu/bio_facpubs/260

This Article has been accepted for inclusion in Biological Sciences: Faculty Publications by an authorized administrator of Smith ScholarWorks. For more information, please contact scholarworks@smith.edu

Authors

Aoi Ogawa, Christophe Golé, Maria Bermudez, Odrine Habarugira, Gabrielle Joslin, Taylor McCain, Autumn Mineo, Jennifer Wise, Julie Xiong, Katherine Yan, and Jan A.C. Vriezen



Extracellular DNAses Facilitate Antagonism and Coexistence in Bacterial Competitor-Sensing Interference Competition

Aoi Ogawa,^c Christophe Golé,^b Maria Bermudez,^a Odrine Habarugira,^a Gabrielle Joslin,^a Taylor McCain,^a Autumn Mineo,^a Jennifer Wise,^a Julie Xiong,^a Katherine Yan,^a ^(D) Jan A. C. Vriezen^a

^aDepartment of Biological Sciences, Smith College, Northampton, Massachusetts, USA ^bDepartment of Mathematical Sciences, Smith College, Northampton, Massachusetts, USA ^cDepartment of Statistics and Data Sciences, Smith College, Northampton, Massachusetts, USA

ABSTRACT Over the last 4 decades, the rate of discovery of novel antibiotics has decreased drastically, ending the era of fortuitous antibiotic discovery. A better understanding of the biology of bacteriogenic toxins potentially helps to prospect for new antibiotics. To initiate this line of research, we quantified antagonists from two different sites at two different depths of soil and found the relative number of antagonists to correlate with the bacterial load and carbon-to-nitrogen (C/N) ratio of the soil. Consecutive studies show the importance of antagonist interactions between soil isolates and the lack of a predicted role for nutrient availability and, therefore, support an in situ role in offense for the production of toxins in environments of high bacterial loads. In addition, the production of extracellular DNAses (exDNases) and the ability to antagonize correlate strongly. Using an in domum-developed probabilistic cellular automaton model, we studied the consequences of exDNase production for both coexistence and diversity within a dynamic equilibrium. Our model demonstrates that exDNase-producing isolates involved in amensal interactions act to stabilize a community, leading to increased coexistence within a competitor-sensing interference competition environment. Our results signify that the environmental and biological cues that control natural-product formation are important for understanding antagonism and community dynamics, structure, and function, permitting the development of directed searches and the use of these insights for drug discovery.

IMPORTANCE Ever since the first observation of antagonism by microorganisms by Ernest Duchesne (E. Duchesne, Contribution à l'étude de la concurrence vitale chez les microorganisms. Antagonism entre les moisissures et les microbes, These pour obtenir le grade de docteur en medicine, Lyon, France, 1897), many scientists successfully identified and applied bacteriogenic bioactive compounds from soils to cure infection. Unfortunately, overuse of antibiotics and the emergence of clinical antibiotic resistance, combined with a lack of discovery, have hampered our ability to combat infections. A deeper understanding of the biology of toxins and the cues leading to their production may elevate the success rate of the much-needed discovery of novel antibiotics. We initiated this line of research and discovered that bacterial reciprocal antagonism is associated with exDNase production in isolates from environments with high bacterial loads, while diversity may increase in environments of lower bacterial loads.

KEYWORDS competition models, exDNases, cellular automaton, coexistence, secondary metabolites, antibiotics, soil, antagonism, soil microbiology

The evolutionary arms race of effective antimicrobial development and subsequent development of antibiotic resistance in microbes has become a hindrance in our pursuit of new and effective antimicrobials (1). In a commentary in *Nature Microbiology*,

Editor Isaac Cann, University of Illinois at Urbana-Champaign

Copyright © 2022 Ogawa et al. This is an open-access article distributed under the terms of the Creative Commons Attribution 4.0 International license.

Address correspondence to Jan A. C. Vriezen, jvriezen@smith.edu.

The authors declare no conflict of interest.

Received 9 September 2022 Accepted 16 October 2022 Published 14 November 2022

Kolter and van Wezel (2) argued that the era of novel antibiotic discovery via brute force has ended. New approaches are needed to sample the underexploited niches in environments like soil, e.g., those developed by Ling et al. (3). In addition, approaches based on biological and ecological insights may lead us to antagonist bacteria with a reduced probability of rediscovering what is already known. Due to their disputed roles (4), many of these insights remain unknown. However, in recent years, several studies have addressed polymicrobial communities from different environments to further our understanding of dynamic interactions between bacteria and the consequences for diversity. Interactions can be cooperative, neutral, or antagonist, unidirectional or reciprocal, symmetrical or asymmetrical, social or asocial, transitive or nontransitive, sympatric or allopatric, structured or unstructured and are used in (mathematical) models to predict structure, stability, and diversity (5-15). Although varied in their specifics, all studies acknowledge inherent problems in testing that are hard to solve. For example, the culture dependence of the bacterial isolates may not be representative of in situ populations. Furthermore, the choice of media, nutrient availability, experimental design, and source of isolates vary between studies, resulting in often mixed or highly selective populations and making it difficult to directly compare results (5, 7, 11, 14). Studies attempting to more deeply understand the consequences of interactions within populations of antagonists are scarce, e.g., social versus asocial, sympatric and allopatric, the medium used, or in vitro versus in vivo studies (5, 8, 9, 11); however, comparative studies between populations are limited, and further exploration is warranted.

Although the role of bacteriogenic toxins in their environment is still up for debate, the anthropocentric point of view is that bioactive compounds are produced as offensive mechanisms against competitor bacteria. However, secreted bioactive compounds undergo diffusion and the resulting ambient concentrations are too low to inhibit growth on realistic length and time scales (16), while hormesis affects phenotypes (17). Regardless of the true *in situ* role, toxins that kill or inhibit growth at high concentrations lead to negative consequences to neighboring cells when in close proximity to antagonists (18, 19). Therefore, we hypothesize that an increased bacterial load will lead to greater benefits for the antagonists, resulting in an elevated presence of said antagonists. Additionally, recent work showed that interactions are overrepresented intergenerically and are inversely related to phylogenetic, metabolic, and functional distance (7, 11, 14). As a result, with an increasing bacterial load, one expects a decrease in the diversity of antagonists and a decrease in the ability to find novelty. These arguments echo the theoretical considerations published by Curtis and Sloan (20).

Various bacterial competition models exist; however, in this paper, we will focus on the following two models: exploitation and interference competition (21). Exploitation competition is practiced by bacteria able to efficiently use resources, while interference competition is practiced by those that produce toxins to ward off competitors. Russel et al. (11) studied the trade-off between exploitation competition and interference competition and found that antagonists that practice interference have a wider metabolic-niche space and a larger network. Within the general theoretical realm of interference competition, the cue for antagonism is either environmental or competitive. Theoretically, it is possible to distinguish between competitor-sensing interference competition (CSIC model) and nutrient deprivation-sensing interference competition (NDSIC model). In the case of the former, microbial populations in environments with higher bacterial loads are expected to have a higher connectance (10) than populations from environments with low bacterial loads. In contrast, in the latter, one expects an increase in antagonism with a decrease in available nutrients. These theoretical contemplations are addressed in this study.

Many bioactive compounds come in the form of enzymes, e.g., colicins and pyocins (22). These proteinaceous, toxic compounds are often produced against closely related strains; e.g., colicins kill *Escherichia coli* and pyocins kill pseudomonads. Many of these compounds have DNase activity (22), indicating a role for DNases in competition. This is further illustrated in several studies showing that extracellular DNases (exDNases)

and effectors of these have roles in hydrolyzing extracellular DNA (exDNA) in situations like biofilm development, toxin production in the presence of exDNA, in enriched plant root-associated bacteria (23–27), and are even differentially selected for by crop plants (28). Therefore, we further hypothesize that exDNase production has an important role in the structure of bacterial interaction networks.

In the study presented here, we compare two populations of antagonists and provide support for the CSIC model. In addition, we also determine the role of exDNase production in coexistence using mathematical modeling approaches.

RESULTS

Bacterial load and incidence of antagonism correlate positively. In order to obtain support for the CSIC or the NDSIC model, it was initially hypothesized that with an increasing bacterial load in an environment, the relative presence of antagonists increases due to closer proximity of competing bacteria, which would benefit antagonists. This would result in an increased presence of these antagonists. In order to probe whether the presence of competitors is a compulsive factor for toxin production, as would be the case in a CSIC model, the effect of the environmental bacterial load on the relative number of antagonists on Staphylococcus CWZ226 or E. coli MC4100 was determined. Soil samples from the Smith College MacLeish Field Station in Whately, MA (29), were retrieved from the grassland surface (GS; 3-cm depth), the grassland subsurface (GSS; 15- to 20-cm depth), the organic (O) horizon from hemlock forest soil (FS; 3-cm depth), and the forest subsurface (FSS; A/B horizon, 15- to 20-cm depth). To elucidate key factors distinguishing soil environments and their impacts on the bacterial load and percentage of antagonists, we determined edaphic characteristics like pH, percent nitrogen (%N), percent carbon (%C), carbon-to-nitrogen ratio (C/N), temperature, water content, CFU/g dry soil, and percentage of antagonists (Table S2-1 in the supplemental material). Principal component analysis (PCA) showed that principal component 1 (PC1) and PC2 explained 78.7% of the variation and that there was no overlap between FSS and GS samples (Fig. 1A). The bacterial load in each environment, expressed in CFU per gram (dry) weight, decreased from (4.41 \pm 0.26) imes 10⁶ (mean \pm standard error of the mean [SEM]) in GS to (3.12 \pm 0.29) \times 10 6 in GSS and from (1.68 \pm 0.73) imes 10⁶ in FS to (5.01 \pm 0.10) imes 10⁵ in FSS. The bacterial load in GSS was lower than that of soil taken from the surface (t test, P < 0.02). Similarly, the bacterial load in the hemlock O horizon was higher than that in the A/B horizon (one-sided heteroscedastic t test, P < 0.12). Furthermore, grassland samples always had a higher bacterial load than forest samples (Table S2-1). Moreover, culturability for GS and FSS was $0.53\% \pm 0.08\%$ and $1.12\% \pm 0.19\%$, respectively (P = 0.01, t test). The total viability results also indicate that GS had a higher bacterial load than FSS (Fig. 1C). The differences in CFU/g dry soil determined on 10% tryptic soy agar (TSA) plates, though compromised, reflected the in situ bacterial loads.

To determine the relative numbers of antagonists, 517, 494, 492, and 462 (n = 5 different dilution series per sample site) colonies from GS, GSS, FS, and FSS, respectively, were transferred to indicator lawns containing *Staphylococcus* or *E. coli*. Of these colonies, 10.3%, 9.3%, 5.7%, and 5.4% (53, 46, 28, and 25 isolates) inhibited the indicator lawn (P < 0.06 for GS versus FSS, *F* test) (Fig. 1B). The results obtained from mock communities showed that the method employed gave the expected ratios and, thus, allowed us to conclude that the level of antagonism reflected the soil community (Text S1).

The difference between GS and FSS was always substantial for pH, %N, %C, and the C/N ratio, with *P* values of <0.002 (heteroscedastic), <0.05, <0.01, and <0.001 respectively (*t* test). However, as expected, only the C/N ratio strongly reflected the sample sites and bacterial loads (Fig. 1B). The C/N ratios were 12.6 and 13.0 in GS and GSS, respectively (*P* = 0.24), whereas the samples from the forest soils had C/N ratios of 25.9 and 27.2 (*P* = 0.26). The C/N ratio is a well-known parameter indicating bacterial load, and the results support the hypothesis that available nitrogen and other quality nutrients allow a higher bacterial load (30–35), leading to increased competition that



FIG 1 Visualization of the quantification of soil edaphic characteristics. (A) PCA of the soil samples using the edaphic characteristics pH, %N, %C, C/N ratio, temperature, water content, CFU/g dry soil, and % antagonists. Colored ovals represent the 95% confidence interval (CI). (B) Averages of CFU/g dry soil of samples (light gray) and C/N ratio (dark gray). Error bars represent SEM (n = 3). Black bars represent the pooled % antagonist values. The data show strict correlation between CFU/g dry soil, pooled % antagonists, and C/N ratio. (C) Estimation of total viable bacterial cells in GS and FSS and the culturability of these populations on 10% TSA. Error bars represent SEM (n = 6). In *E. coli*, the culturability was 95.5% \pm 13.5% (n = 12).

DNases in Antagonism and Coexistence

Applied and Environmental Microbiology



FIG 2 Differentiation of the CSIC and NDSIC models. (A) Theoretical considerations and hypothetical outcome for connectance when competitor sensing (yellow bars) or environmental sensing (green bars) is the major cue for bacteriogenic toxin production. (B) The empirical connectance in both populations with interaction types in the total and exDNase-producing populations (black bars, GS_{48} ; white bars, FSS_{23}). The data indicate a higher connectance in the GS population than in the FSS population, thus giving support for the CSIC model. The rates of neutral (0,0), one way antagonism (0/-), and reciprocal antagonistic interactions (-/-) indicate more reciprocal interactions in the GS population than in the FSS population. (C) Theoretical considerations and hypothetical outcome to a response to a decrease in nutrients when competitor sensing (yellow line) or environmental sensing (green line) is the major cue for toxin production. (D) Responses of the GS and FSS populations to a decrease in nutrient strength and to nutrient complementation, as well as to DNA plates to test for DNA availability and DNase activity. Error bars represent the SEM (n = 3).

boosts the population of toxin-producing bacteria. Separating the parameters of competitor presence and nutrition depletion to more clearly understand the cues for bacterial toxin production warranted further study. To do so, we contrasted the GS and FSS populations for their connectance, response to nutrients, and ability to coexist, since these populations represented the two samples with the greatest deviations in terms of CFU/g dry soil, percentage of antagonists, and C/N ratio.

Support for CSIC. To further obtain support for the CSIC or the NDSIC model, the presence of a competitor and the role of available nutrients in toxin production by the GS and FSS populations were determined. To do so, we (i) determined the connectance of both populations and (ii) determined if these populations responded differentially to a reduction of available nutrients (Fig. 2). The connectance was $41.2\% \pm 0.9\%$ for the GS population and $30.6\% \pm 4.2\%$ for the FSS population (P = 0.03) (Fig. 2B). To address the response to nutrients (Fig. 2D), the isolates from both populations were tested on 100%, 50%, and 10% TSA. On 100% TSA, the difference between the FSS and

| TABLE 1 Aggressiveness index values of populations of isolates producing a zone of inhibition on | n different media |
|---------------------------------------------------------------------------------------------------------|-------------------|
|---------------------------------------------------------------------------------------------------------|-------------------|

| | | Value with ^a : | | | | | | |
|---------------------------------|----------------|---------------------------|-----|--------------------|------|-----|--------------------|----------------------|
| Medium or test | Soil source | No ZOI | | | ZOI | | | |
| | | AI | SEM | No. of isolates | AI | SEM | No. of isolates | P value ^b |
| % TSA | | | | | | | | |
| 100 | GS | -0.8 | 3.6 | 19 | 0.4 | 2.0 | 29 | NS |
| | FSS | -2.3 | 3.4 | 8 | 1.2 | 1.4 | 15 | 0.2 > P > 0.1 |
| 50 | GS | -3.5 | 2.0 | 24 | 3.4 | 3.0 | 24 | P < 0.05 |
| | FSS | -2.6 | 3.9 | 7 | 1.5 | 1.4 | 15 | 0.2 > P > 0.1 |
| 10 | GS | -1.3 | 1.9 | 42 | 8.7 | 6.7 | 6 | P < 0.05 |
| | FSS | 0.3 | 1.9 | 12 | -0.3 | 2.4 | 11 | NS |
| 10% TSA plus: | | | | | | | | |
| Glucose | GS | -1.2 | 2.4 | 20 | 0.7 | 2.7 | 28 | NS |
| | FSS | -3.4 | 2.7 | 8 | 1.8 | 1.6 | 15 | P < 0.05 |
| NH ₄ NO ₃ | GS | 0.4 | 2.1 | 42 | -4 | 2.3 | 5 | NS |
| | FSS | -4.7 | 1.8 | 9 | 3 | 1.7 | 14 | P < 0.01 |
| K ₃ PO ₄ | GS | 0.7 | 1.9 | 43 | -3.6 | 9.6 | 4 | NS |
| | FSS | -3.2 | 1.9 | 12 | 3.5 | 1.9 | 11 | P < 0.02 |
| DNA plate ^c | | | | | | | | |
| Test for killing | GS | -3.1 | 4.2 | 10 | 0.7 | 2.1 | 38 | NS |
| | FSS | 1.4 | 2.3 | 11 | -1.3 | 1.9 | 12 | 0.2 > P > 0.1 |
| Test for DNase | GS | -5 | 4.1 | 10 | 1.2 | 2.1 | 38 | 0.1 > P > 0.05 |
| activity | FSS | 1.3 | 2.6 | 10 | -1 | 1.8 | 13 | NS |
| Test for killing in | GS | -19 | 1 | 2 | 1.1 | 2 | 36 | P < 0.02 |
| DNase producing populations | FSS | 10 | NA | 1 | -0.8 | 1.1 | 12 | |

^aZOI, zone of inhibition; AI, aggressiveness index.

^{*b*}Return is from the homoscedastic, one-sided *t* test. NS, not significant with P > 0.2.

^cUsing DNA-plates seeded with *Staphylococcus* CWZ226 allows to test for a ZOI and the ability to hydrolyze DNA in the whole populations, as well as the exDNase producing populations.

GS populations was negligible (P = 0.39). The relative amount of antagonists in the FSS population decreased from 58.9% to 48.4% on 100% and 10% TSA, respectively (P < 0.14, sign and t test). The relative amount of antagonists in the GS population decreased from 55.6% to 20.8% on 100% and 10% TSA, respectively ($P < 1 \times 10^{-8}$, sign test, or P = 0.02, t test). The reduced nutrient content in a complex medium like TSA and the corresponding reduction in antagonist activity may be caused by depletion of a number of nutrients. Therefore, we tested whether the addition of 100 mM glucose, 120 mM ribose, 200 mM pyruvate, 50 mM NH₄NO₃, or 100 mM K₃PO₄ would complement the lack of C, N, or P in the 10% TSA plates and restore the levels of antagonism. Indeed, the addition of glucose or ribose to 10% TSA increased the antagonism to levels like those on 100% TSA, mainly for the GS population (P = 0.05 for GS on glucose). Interestingly, pyruvate did not complement the reduced sugar content in 10% TSA. Inorganic phosphate also failed to increase the levels of antagonism to the levels on 50% or 100% TSA. Ammonium nitrate also did not do so for the GS population (P = 0.5), and only an effect for the FSS population was observed (P = 0.06). Although there was a correlation with available nutrients, the data indicated that it was the increase of select nutrients that supported antagonist activity, not a decrease as expected in the NDSIC model.

In order to test whether the aggressiveness of a population correlated with its ability to produce a zone of inhibition (ZOI), we determined the aggressiveness index (AI) for all strains according to Zapien-Campos et al. (15) and contrasted the average AI of those strains producing a ZOI to the average AI of those not able to produce a ZOI (Table 1). We expected those populations producing a ZOI to have a higher aggressiveness index than those that did not. Accordingly, populations able to produce a ZOI almost always had a higher AI (0.4 to 8.7) than those strains not producing a ZOI on 100% TSA, 50% TSA, and 10% TSA (-0.8 to -3.5). Even though the number of GS isolates producing a ZOI on 10% TSA was limited (n = 6), the AI of this population stood out (8.7) compared to the AIs of the rest (0.4 to 3.4). The main outlier was the FSS population when tested on 10% TSA, for which the population producing a ZOI had a negative AI (-0.3) and the population not producing a ZOI had a positive AI (0.3, P = not significant). Only in this population, when tested on 10% TSA, was the expected correlation of ZOI and AI not observed.

The addition of glucose to 10% TSA restored the Als to the levels found for the populations producing a ZOI for the GS and FSS populations, with the Als found to be like those of populations tested on 50% and 100% TSA. In contrast, when NH₄NO₃ was added to 10% TSA, the Al of the FSS population was very much restored, and the difference in Als of populations able and unable to produce a ZOI increased even more than found on 100% TSA (P < 0.01). However, NH₄NO₃ strongly negatively affected the Al of the GS population, and the expected correlation between Al and the ability to produce a ZOI was not observed. Most interestingly, the GS populations even though the levels of antagonism were similar in both populations. Indeed, the two populations had only two isolates in common (Tables S3-1 to S3-3). Therefore, we concluded that although the addition of sugar to 10% TSA restored the Al for both populations, it was nitrogen availability under low-nutrient conditions that regulated the production of toxins in subpopulations derived from the GS soil.

While exDNA may act as an important source of nutrients (27), it also indicates the presence of competing bacteria (28). Given this association, we hypothesized that if competitor sensing was the primary instigator for toxin production (over nutrient availability) in the GS and FSS populations, then antagonism would increase when these populations were plated on DNA plates (Difco) compared to the antagonism on 100% TSA. We found substantial support for the idea that exDNase production was important in antagonism. First, a strong correlation between the ability to produce a ZOI on *Staphylococcus* and exDNase activity was observed in both GS and FSS populations ($P = 9.84 \times 10^{-12}$ for GS and $P = 7.75 \times 10^{-7}$ for FSS). Almost all isolates that antagonized on DNA plates were also exDNase producers (Tables S3-1 to S3-3). Second, substantially more GS isolates antagonized *Staphylococcus* on DNA plates than on TSA (76.5% versus 62.3% respectively), which was reversed for the FSS population (52.0% on DNA plates and 64.0% on 100% TSA) (Fig. 2D), indicating the larger role for exDNases in a high-bacterial-load and low-C/N-ratio environment.

The AI levels obtained for the GS populations producing a ZOI or hydrolyzing exDNA on DNA plates were higher than those of the populations not producing a ZOI (P = not significant) or hydrolyzing DNA, exactly as expected (P < 0.1). However, for the FSS population, this was reversed. Even when the DNase-producing populations only (GS₃₈ [the 38 exDNase-producing isolates in the GS population] and FSS₁₃) were analyzed, these relationships did not change.

Reciprocal amensal interactions are underrepresented in the FSS population. In addition to the high connectance in the GS population, the total amount of reciprocal (two-way amensal) interactions was higher in the GS population ($20.1\% \pm 0.5\%$) than in the FSS population ($10.3\% \pm 0.5\%$, P = 0.01) (Fig. 2B), while the level of amensal interactions was very similar in both populations (42.8% to 43.6%, P = 0.43) (Fig. 2B). However, in the GS population, the percentage of reciprocal interactions, estimated by taking the square of the amensal interactions ($0.436^2 = 19.0\% \pm 0.7\%$), was almost equal to the number of observed interactions ($20.1\% \pm 0.5\%$, P = 0.13). Interestingly, in the FSS population, the estimated percentage of reciprocal interactions ($0.428^2 = 18.3\% \pm 3.1\%$) was higher than the observed percentage of reciprocal interactions ($10.3\% \pm 2.4\%$, P = 0.05). Therefore, the reciprocal amensal interactions were underrepresented in the FSS population. In addition, exDNase activity was increased in the



FIG 3 Diversity and coexistence in GS and FSS populations pre- and postsimulation. (A) Estimated diversity and coexistence pre- and postsimulation for the whole (23 isolates) and the DNase-producing (13 isolates) population using a probabilistic cellular automaton (ProbCA) and the interaction matrices (Tables S3-1 to S3-3). The forest subsurface (FSS₂₃) population was composed of 23 isolates, and the grassland surface (GS₂₃) population was composed of the 23 best-surviving isolates selected after a pilot run with the full data set. Coexistence is the number of isolates present in the dynamic equilibrium. Shannon diversity was based on the resulting isolate distribution after simulation. Error bars represent the SEM (n = 100). SEMs in the GS₁₃ and FSS₁₃ populations were too small to show. (B) Interaction types and relative abundances of exDNase-producing isolates pre- and postsimulation in the 23-strain and 13-exDNase-producing-strain-only matrices. *, heteroscedastic t test; all others were homoscedastic.

isolates involved in reciprocal interactions, which for the FSS population were 10.3% ± 2.4% and 24.4% ± 3.8% (P < 0.02) of the total and DNase-producing populations, respectively. For the GS population, the amounts were 20.1% ± 0.5% and 25.4% ± 0.8% of the total and exDNase-producing population, respectively (P < 0.01) (Fig. 2D). The relative amounts of exDNase-producing isolates involved in reciprocal antagonism were very similar (Fig. 2B) (P < 0.46).

Coexistence is differentially affected by DNase production. Competitive (12, 14) and (a)social interactions in combination with media (5) positively affect coexistence. Additionally, killing is a means to promote diversity (16), and connectance as well as reciprocal interactions were higher in the GS than in the FSS population. Therefore, we predicted coexistence in the GS population to be higher than in the FSS population. In order to test for coexistence, we employed an *in domum*-developed probabilistic cellular automaton (ProbCA) (Text S4). Due to differences in starting sizes, the GS₄₈ population was simulated twice in series, as follows: (i) the population as a whole (48 \times 48

comparisons) and (ii) GS₂₃, consisting of the 23 best-surviving isolates, representing 99.2% of all cells occupied at the end of the first simulation. As shown by the data in Fig. 3A, a 350 \times 350 matrix seeded with $\rm GS_{23}$ or $\rm FSS_{23}$ led to a decrease in coexistence in both populations in which 16.9 \pm 0.2 and 15.8 \pm 0.1 isolates coexisted, respectively, although the difference in coexistence was only 1.1 isolates (7%, heteroscedastic onesided t test, n = 100, $P < 1 \times 10^{-7}$). Shannon diversity decreased in both populations relative to that in the not-simulated mock population of 23 at time zero (T = 0) and was lower in the FSS₂₃ population than in the GS₂₃ population (1.16 and 1.77 respectively, $P < 1 \times 10^{-136}$). Because the contributions of exDNase populations from GS and FSS to reciprocal interactions were essentially the same (P = 0.46) (Fig. 2B), we expected simulations of the exDNase-producing populations from both samples to lead to similar levels of coexistence. Indeed, the occupancy (number of cells occupied by exDNase-producing isolates divided by the total number $[350 \times 350 = 122,500]$) of exDNase isolates increased in both populations after the simulations. In the GS₂₃ population, this increased from 73.9% to 99.5% (a 34.6% increase), and in the FSS₂₃ population, it increased from 56.5% to 65.6% (a 16.1% increase) (Fig. 3B). Clearly, producing exDNases was an advantage; however, the advantage was greater in the GS population.

To examine the consequences of antagonism by exDNase-producing isolates only, simulations were run using the 13 exDNase isolates from the FSS population (FSS₁₃) and the best-surviving 13 isolates from the GS population (GS₁₃), representing 99.6% of the 38 exDNase-producing isolates in the GS population (GS₃₈; FSS₁₃ is 100%) after a pilot run. Compared to the results of the GS₂₃ and FSS₂₃ simulations, three major differences can be observed (Fig. 3A). (i) The coexistence of the GS₁₃ population was unchanged compared to the start situation and remained 13 (Fig. 3A, gray bars). (ii) The Shannon diversity of GS₂₃ (1.8 ± 0.002) increased in GS₁₃ despite starting with fewer isolates and was 2.08 ± 0.0004 ($P < 1 \times 10^{-97}$). Therefore, the GS₂₃ diversity was positively affected by the exDNase-producing population. In contrast, (iii) the Shannon diversity in FSS₂₃ (1.16 ± 0.001) was higher than that in FSS₁₃ (0.28 ± 0.003) ($P < 1 \times 10^{-123}$), while the coexistence was low (three strains). Two opposite effects were observed in both populations: in the GS population, exDNase production supported diversity and coexistence, while in the FSS population, it did not.

The low coexistence in the dynamic equilibrium in the FSS₁₃ simulations, not seen in the FSS₂₃ (15.8 strains), GS₂₃ (16.9 strains), and GS₁₃ (13 strains) simulations (Fig. 3B), was associated with a low rate of neutral interactions and more amensal and reciprocal interactions in the FSS₁₃ population only (Fig. 3B). In contrast, the higher rate of neutral interactions and lower rates of amensal and reciprocal interactions in the GS₂₃, FSS₂₃, and GS₁₃ populations led to higher levels of coexistence. Using the Dixon test, we identified the reciprocal interactions in the FSS₁₃ populations as lower than in any of the other sample origins (P > 0.30), as well as the rate of 28.2% of neutral interactions (0.05 > P < 0.10). For the amensal interactions, the outlier was GS₁₃ (0.2 > P < 0.1).

Phylogenetic distance negatively affects the rate of amensalism. To model patterns of amensalism and reciprocal amensalism, we first determined the genera of the isolates using 16S rRNA sequencing followed by BLAST searches of the complete-genome databases at NCBI. The results (Table S5-1) showed that the GS population consisted of 1 Gram-negative isolate and 47 Gram-positive isolates in four genera. The FSS population consisted of 8 Gram-negative isolates and 15 Gram-positive isolates in five genera. Based on genus, GS was less diverse than FSS, having 48 isolates from four genera while FSS contained 23 isolates from five genera (Fig. 4). In both populations, *Bacillus* was predominant, comprising 42 (87.5%) and 11 (47.8%) isolates in the GS and FSS populations, respectively. The second predominant genus was *Paenibacillus*, with three and four isolates from GS and FSS, respectively. The GS population contained two *Lysinibacillus* isolates and a *Variovorax* isolate, and the FSS population contained four *Paraburkholderia*, three *Collimonas*, and one *Dyella* isolate.

To compare the variations of patterns of inhibition of the indicator strains, PCA plots were created using the interaction matrices. As shown by the data in Fig. 4A and



FIG 4 Visualization of interactions in the GS_{48} and FSS_{23} populations using PCA (A and B) and schematic models (C, D, and E). The GS_{48} and FSS_{23} majority rule consensus interaction matrices were used as input for the PCA using default settings. (A) First and second component of the PCA analysis of the GS_{48} interaction matrix. All variation falls within the 95% CI of *Bacillus* (red line). (B) First and second components of the PCA analysis of the FSS_{23} interaction matrix. All variation falls within the 95% CI of *Bacillus* (red line). (C) Explanatory diagram of the meaning of the interactions between different genera in panels D and E. (D) Interaction diagram of the GS_{48} population. (E) Interaction diagram of the FSS_{23} population. A, amensal interactions; R, reciprocal interactions. In black are all interactions, and in purple are all exDNase-associated interactions.

B, PC1 and PC2 explained 44.2% and 43.3% of the variation for GS and FSS, respectively. The vast majority of isolates, regardless of genus, fell within the 95% confidence interval (CI) of *Bacillus* in both populations. We found 90% of all *Bacillus* isolates in both populations to produce exDNases. *Bacillus thuringiensis* is known to produce toxins in the presence of DNA and is a species in the *Bacillus cereus* group known for toxin producers to which many of our isolates are related (Table S5-1). Furthermore, many *Bacillus* species produce lipopeptides, compounds with antimicrobial activity involved in biofilm restructuring (36). When including the third component, >52% of the variation could be explained in both populations, and the same conclusions were supported (Fig. S6).

The interaction diagrams indicating inter- and intrageneric interactions, as well as those that were exDNase mediated, showed that the majority of intrageneric interactions was within the genus *Bacillus* in both populations (Fig. 4C to E) but was much more profound in the GS than in the FSS population. The number of interactions between Gram-positive producers on Gram-positive indicators was overrepresented. Gram-positive producers and Gram-negative indicators, as well as Gram-negative

producers and indicators, were underrepresented in both populations (χ^2 , df = 3, P < 0.001).

In both populations, the observed amensal and reciprocal interactions were different from the expected (χ^2 , df = 3, P < 0.001). Most intergeneric interactions took place between *Bacillus* and *Paenibacillus* in both populations. Furthermore, intrageneric reciprocal interactions were overrepresented in both populations (χ^2 , df = 3, P < 0.01). This was especially clear in the FSS population, where the reciprocal interactions between the *Bacillus* isolates (36.4%), as well as the *Paenibacillus* (33.3%) and the *Paraburkholderia* (16.7%) isolates, were overrepresented relative to the intergeneric interactions (4.9%) and total number of reciprocal interactions (12.6%). Surprisingly, in the GS population, all intrageneric reciprocal interactions were found in the *Bacillus* genus alone (24.2%). The intergeneric interactions were 9.4%, and total interactions were 20.0%.

Modeling did not affect the genera present even though the number of isolates present in the dynamic equilibrium decreased compared to the starting condition.

DISCUSSION

At the outset of this work, our theoretical contemplations based on theoretical soil microbial and diversity models, such as presented by Curtis and Sloan (20), and the competition models reviewed by Hibbing (21), were considered insufficient; however, they provided a framework for hypothesis development and testing. After the initiation of this project, we found a strong correlation between the CFU/g dry soil and the percentage of antagonists. Although this correlation supports a CSIC over an NDSIC model for the production of toxins, a correlation with the C/N ratio was also found. Though the C/N ratios strongly support the relative order of CFU/g dry soil, the results lay bare a conundrum: is the main cue to toxin production the nutrient status of the soil or the presence of competitors? Resolution of this conundrum is provided by (i) the increase in connectance in the population with a higher bacterial load (GS), (ii) the reduction of antagonist activity with decreasing nutrients, (iii) the higher rate of reciprocal amensal interactions in the population derived from a high bacterial load, and (iv) the frequency with which pathways for the production of secondary metabolites become cryptic outside the context of their natural environment—e.g., storage in glycerol or growth on agar plates (attenuation) lowers toxin expression (37–39), but toxin expression is reestablished when the correct conditions are met, including the presence of a competitor. Here, for example, of the 78 (53 + 25) isolates tested, 97.4% produced a ZOI on indicator isolates within the same population (Tables S3-1 to S3-3), which was an increase from the 44.0% to 64.4% on Staphylococcus only. (v) Finally, isolates producing a ZOI tended to have a higher aggressiveness index (AI) than the strains not producing a ZOI, especially for the GS population on TSA and DNA plates.

However, the complementation studies showed that nutrients did play a role at lower nutrient levels in the media. Glucose and ribose restored the levels of antagonism for mainly the GS population, while NH_4NO_3 did this for the FSS population. Interestingly, NH₄NO₃ addition reversed the AI in the GS population, with a higher AI for those strains not producing a ZOI than for those producing a ZOI. This may be explained by the C/N status of the soils the isolates were derived from. Because of the low C/N status of the GS soil, the addition of different nitrogen sources selectively leads to an antagonist population with low AI when available nutrients are scarce. This probably is a condition under which an antagonist with low AI can get the opportunity to proliferate in a competitive environment. This contrasts with the FSS population. Being isolated from a soil poor in available nitrogen already, nitrogen availability establishes the correlation of AI and the ability to produce a ZOI when available nutrients are limited. This is the condition under which antagonists with high Als can take the advantage in an otherwise less competitive environment. We have not been able to find support for these models in the literature. In addition, although we obtained an expected order of bacterial loads, which tended to correlate well with the active microbial biomass and enzyme activity (40), CFU/g dry soil on 10% TSA may not be an

Applied and Environmental Microbiology

accurate reflection of the total in situ bacterial load. And yet, our data do support that the CFU/g dry soil reflected the total viable bacterial populations in GS and FSS, although with low culturability. Furthermore, while the rates of amensal interactions were the same in the GS and FSS populations (Fig. 3A), more reciprocal interactions were observed in the GS population than in the FSS population, which can be explained as a consequence of random amensal interactions. In contrast, reciprocal amensalism in the FSS population is underestimated and may be selected against. Since participation in reciprocal antagonism is a consequence of the rate of amensalism and randomness, reciprocal antagonism is a more appropriate term than competition, which infers a purpose. This is counterintuitive to the hypothesis that the production of toxins is selected for (e.g., see reference 13). This discrepancy can be explained by the experimental design. De Vos et al. (14) and Kehe et al. (7) did their testing in nonstructured environments as opposed to testing within structured environments (8, 9; this study). In structured environments, the concentrations of toxin are distance dependent, explaining the phenomena we observed. When we assume the distribution between bacteria in soil is homogeneous and only 0.53% and 1.13% of the total soil microbial populations form colonies on 10% TSA, then the average distance between bacterial cells is 12.1 and 32.1 μ m for GS and FSS, respectively (Table S7-1). This is well within the range Raynaud and Nunan (41) estimated (0.3 to 532.4 μ m) using simulations of thin sections of soil. If the percentages of antagonists of the colony-forming populations are also an accurate estimate in the nonculturable populations, then the average distances between antagonists are 25.8 and 85.0 μ m for GS and FSS, respectively. When estimating the volume of a sphere using these distances as the radius, the volume in the FSS sphere is 35.8-fold larger than that of the GS sphere. If these bacteria produce the same amounts of toxin, and the toxin is evenly distributed in this sphere, then the toxin concentration in the FSS sphere is 2.8% (1/35.8) of that in the GS sphere. When the estimated connectance is also corrected for, the volume in the FSS sphere is 48.2-fold larger than that of the GS sphere and the toxin concentration only 2.1% of that in the GS sphere. Compellingly, to be an effective toxin producer requires substantially more effort for the FSS population. Furthermore, we observed that the number of interactions was increased intra- rather than intergenerically, which would further dilute the target in the populations that are more diverse, reducing effective toxin production even further.

Our observation that populations were reduced in antagonist activity under decreasing nutrient content (Fig. 2D) supports the observation by Russel et al. (11) of the tradeoff between exploitation and interference competition. However, most isolates were antagonistic at high nutrient content with a competitor present. Therefore, antagonists may be specialists under different conditions and the trade-off in cost is environment dependent (grassland versus forest soils) rather than primarily lifestyle dependent (metabolic, physiologic, or phylogenetic). This is exemplified by our complementation and AI studies, which also show differential responses to C, N, and P availability depending on the source of the population. Although the FSS environment may contain more specialists and exploiters and the GS population more generalists and antagonists, on the level of populations of antagonists, this discussion is futile: All isolates are antagonists. The observation that both populations showed a decrease in antagonism with decreasing nutrient content but that there was a more profound decrease in the GS population than in the FSS population (P = 0.02) suggests that antagonists in the FSS population are able to produce toxins at low nutrient availability better than those populations from an environment of high nutrient availability. This again illustrates that the trade-off cost between exploitation and antagonism is not one-size-fits-all.

Our data also indicate a role for exDNase production in the soil environment. The production of exDNases correlated with the ability to inhibit *Staphylococcus* and was higher in the GS population than in the FSS population. Furthermore, exDNase producers had higher Als and were competitive in simulations using the GS population, which was not seen in the FSS population. This was also exemplified when populations

were tested for antagonism and AI on DNA plates. The strong correlation between toxin production and exDNase activity suggests a functional linkage between, on one hand, increasing the pool of exDNA in the environment and, on the other hand, the utilization of available exDNA. That inorganic phosphorus did not affect the levels of antagonism in both populations supports the idea that organic phosphorus may be the main target of toxin production. In the GS population, this linkage was associated with competition and coexistence. Since our exDNase assay only determined the observable hydrolysis of exDNA, exDNA in soil is considered the target for these exDNases. Various sources of exDNA have been identified, e.g., the sloughing of plant cells from the root tip or, alternatively, the consequence of lysis of bacteria (42), therefore providing the functional linkage. At an increased bacterial load, more isolates may produce compounds with antimicrobial activity by lysing bacterial cells. Lipopeptides produced by Streptomyces, Pseudomonas, and Bacillus (36) kill by forming pores in membranes, lysing competing bacteria, which in turn provide the substrate for the exDNases. In Bacillus, lipopeptides are associated with biofilm restructuring and cannibalism (43). Disruption of biofilms allows better access for antimicrobials to otherwise recalcitrant cells. Variation in geographic lipopeptide production by Bacillus was observed previously (44), and they have a role in competition (36).

Although naturally competent, *Bacillus* is a genus not particularly known for the production of exDNases (24). However, their importance is illustrated by the following. With a decreasing C/N ratio, it is increasingly likely that phosphorus is limiting, resulting in competition for available organic phosphorus. Similar observations were made by Mulcahy et al. (27), Turk et al. (45), and Kamino and Gulden (28), who isolated exDNase-producing *Bacillus* strains mainly from soils of low C/N ratios and low phosphorus. Therefore, we hypothesize that a low C/N ratio results in generalist, exDNase-producing *Bacillus* isolates with high aggressiveness scavenging for organic phosphorus. A high C/N ratio results in fewer of these isolates. Consequently, more and fewer exDNase-mediated reciprocal amensal interactions between closely related bacteria were observed, respectively (7, 11, 14).

Although these explanations are plausible, genetic linkage or coregulation of the expression of toxin and exDNase and the consequences for competitiveness and coexistence are yet to be established. Studies with strains isogenic for exDNase activity have shown a role for exDNases in fitness and virulence (25–27).

On the genus level, diversity was higher in the FSS population than in the GS population pre- as well as postsimulation. However, since competitive (12, 14) and social (5) interactions positively affect diversity and both connectance and reciprocal amensal interactions were higher in the GS than in the FSS population, on the isolate level, we predicted coexistence in the GS population to be higher than in the FSS population. This is exactly what we found, but with only a marginal difference in coexistence (1.1 isolate). We found that coexistence increased in populations with more reciprocal ($GS_{23} > FSS_{23}$) (Fig. 4A) and fewer neutral interactions. In support of this, compared to the not-simulated mock community ($Mock_{13}$), in the GS_{13} simulations, the level of coexistence remained at 13 isolates. This indicates that the population of exDNase-producing antagonists coexists well, which also corresponds to a slight increase in reciprocal interactions compared to the level in GS_{23} .

In contrast, the large decrease in coexistence in the FSS₁₃ simulations, not seen in the FSS₂₃, GS₂₃, and GS₁₃ simulations, could potentially be caused by an increase in taxonomically different isolates, as proposed by de Vos et al. (14) and Kehe et al. (7). Alternatively, a change from neutral interactions in the FSS₂₃ population (from ~47% in FSS₂₃ to ~30% in FSS₁₃) (Fig. 3B) to reciprocal interactions in FSS₁₃ (from ~10% in FSS₂₃ to ~27% in FSS₁₃) (Fig. 3B), not observed in the GS₂₃ and GS₁₃ populations, may explain the decrease in coexistence.

The first phenomenon is unlikely since the main variation in antagonist activity in both populations was largely represented by *Bacillus* only, and thus, the populations were not taxonomically different on the genus level. In the second phenomenon, the

shift to higher reciprocal interactions and fewer neutral interactions was relatively small, and therefore, the high number of amensal interactions in the GS_{23} , FSS_{23} , and GS_{13} populations led to a high coexistence, while in the FSS_{13} population, it was the high rate of reciprocal and low rate of neutral interactions that led to low coexistence. This again does not support findings that competition is important for coexistence in an environment as structured as soil (5), but it supports the work by Mougi (10), whose *in silico* work showed that asymmetry in interactions supports stability. According to these models, the FSS_{23} population is expected to be relatively stable because of fewer reciprocal interactions in this population. In contrast, in the exDNase-only populations, the increase in reciprocal amensal interactions would lead to low coexistence. Both these consequences were observed. Therefore, we conclude that observable exDNase-producing isolates involved in amensal interactions stabilize a community, leading to an increase in coexistence in competitor-sensing interference competition in a structured environment.

MATERIALS AND METHODS

Strains, soil isolates, and culture conditions. All strains were stored in tryptic soy broth (TSB) (catalog number 211825; Difco) with 20% glycerol at -80° C. *Staphylococcus* sp. strain CWZ226 (46), *Escherichia coli* strain MC4100 (47), and *Serratia* sp. strain CWZ222 (Fig. S8-1) were provided by Dr. C. White-Ziegler (Smith College) and maintained on 100% tryptic soy agar (TSA) (catalog number DF0369-17-6; Difco). *Lysobacter antibioticus* strain CVAP#2 (*L. antibioticus* strain ATCC 29479 [48]; provided by Dr. J. Handelsman) and *Pseudomonas* sp. strain CVAP#3 (46) and all isolates from the soil were maintained on 10% TSA (from 100% TSA amended with agar [catalog number DF0812-17-9; Difco]; the final agar concentration was 1.5%). All strains were incubated for 36 to 48 h at 25°C and kept at 4°C until use for a maximum of 1 week.

Field sites and sampling. Samples were taken from the Smith College MacLeish Field Station in Whately, MA (29), on 21, 22, and 24 September in 2015. The geographic coordinates for the grassland soil samples are N42°26.983', W072°40.820'. The hemlock forest soil samples were taken at coordinates N42°27.328', W072°40.926' (Garmin eTrex 20x). Soil and air temperatures were taken at the moment of sampling. Soil samples were taken using a sterile spatula or spoon while wearing alcohol-sterilized gloves. The samples were stored in sterile wide-mouth Mason jars (450 mL) and frozen at -20° C upon arrival in the laboratory after subsamples were taken for bacterial counts.

Determination of CFU/g dry soil and identification of isolates that produced bioactive compounds. Within 6 h of taking the sample, 1 g of soil was weighed and a suspension was made in 9 mL of sterile phosphate-buffered saline (PBS) (product number 2810305; MP Biomedicals) in a sterile 15-mL conical tube (7). Amounts of 100 μ L of a 10-fold dilution series were spread on 10% TSA (6, 29) and incubated for 36 to 48 h at 25°C, after which CFU were counted and plates were stored at 4°C till further use. All plates, with an average number per dilution of $30 \leq CFU \leq 300$, were used to determine the bacterial load (49). The bacterial counts were corrected for water content, and the bacterial load in CFU/g dry soil was calculated for every plate. Subsamples, dilution series, and plate counts originating from the same soil sample were pooled. The resulting average values per sample site were used to calculate a grand average for an estimation of CFU/g dry soil for a sample period.

To identify the level of antagonism in a soil, colonies originating from the dilution plates used for counting were transferred to master plates made of 10% TSA. Colonies were picked randomly with a sterile flat toothpick. Only plates with fewer than 300 colonies were sampled. Up to 125 colonies from a dilution series representing one sample were transferred to the master plate (10% TSA) and incubated at 25°C for 36 to 48 h. Indicator plates were seeded as follows: a colony was resuspended in 1.0 mL PBS with a sterile synthetic-tipped applicator (catalog number 23-400-122; Fisherbrand), spread over the surface of an agar plate, and dried. Drying is required to prevent swarming of colonies and contamination of neighboring colonies. After drying, *Staphylococcus* and *E. coli* were transferred onto the indicator plates as negative controls. L antibioticus and Pseudomonas sp. strain CVAP#3, which inhibit *Staphylococcus*, were used as positive controls. Plates were stored at 4°C for 3 to 5 days and incubated at 25°C overnight. Isolates inhibiting an indicator lawn were purified at least twice, retested, stored at -80° C in TSB with 20% glycerol, and given an identifying number (Chris Vriezen antibiotic producer number [CVAP#]). The relative number of antagonists was determined by dividing the final number of isolates producing a ZOI by the original number transferred and tested. All data from the same sample period and site were pooled (7, 12, 14).

Estimation of the total viable cells and culturability. Culturability is the fraction of CFU relative to the total population of viable bacterial cells in a sample (50). In order to estimate the culturability in the samples derived from the grassland surface, the forest subsurface, and *E. coli* suspensions in PBS, we determined the number of viable cells using live/dead stain (BacLight bacterial viability kit, product number L7012; Molecular Probes). On 26 and 28 August 2022, soil samples were taken and dry weight determined as described above. Soil suspensions (1:10) were made in 1× PBS. The suspensions were vortexed for 30 s, and CFU/g dry soil determined. In addition, 100 μ L resuspended soil was mixed 1:1 in PBS containing 3 μ L Syto9 and 3 μ L propidium iodide per mL. Using a Neubauer counting chamber, the

| Annotation of | |
|--------------------------------------|------------------------------------------------------------------------------------------------|
| populations | Explanation ^a |
| GS ₅₃ , FSS ₂₅ | Total populations (53 and 25 isolates) of isolated antagonists |
| GS ₄₈ , FSS ₂₃ | Populations of antagonists after QC |
| GS ₂₃ , FSS ₂₃ | Representatives of the quality-controlled population of antagonists used to seed the ProbCA |
| GS ₃₈ , FSS ₁₃ | Total population of exDNase-producing antagonists after QC |
| GS ₁₃ , FSS ₁₃ | Representatives of the exDNase-producing population of antagonists used to seed the ProbCA |

^aQC, quality control.

number of bright green cells was counted at $\times 200$ or $\times 400$ total magnification on a Leica DM5500B using the I3 filter cube. Only bright fluorescent green cells were counted. At least three different subsamples were counted for every suspension. At least three frames per sample were completely counted, with a minimum of 100 cells per prep. This resulted in six estimates per soil. Control suspensions of *E. coli* strain MC4100 growing on 10% TSA in PBS were treated in a similar manner.

Determination of basic soil edaphic properties. To determine the water content, three soil subsamples for each original sample were weighed, stored at 55°C for 2 to 4 days, and weighed again. The water content is the weight lost after drying divided by the original weight and expressed as a percentage. Dried samples were sieved (product number 04-881G; Fisher Scientific), and the percent nitrogen (%N), percent carbon (%C), and C/N ratio determined using a vario micro select CHNOS element analyzer (51). These measurements were carried out at the Center for Aqueous Biogeochemistry Research at Smith College. For pH determination, dried soil was diluted 1:9 (wt/wt) in demineralized water. After 2 h, the pH was measured using an Accumet model 10 with Accumet probe 13-620-285.

Nutrient and DNase activity determination. The following media were used to test the response of the soil isolates: 100% TSA (catalog number DF0369-17-6; Difco) and 50% and 10% TSA (from 100% TSA amended with agar [catalog number DF0812-17-9; Difco], final agar concentration of 1.5%). Nutrient complementation studies were performed by the addition of 100 mM glucose, 120 mM ribose, 200 mM pyruvate, 50 mM NH₄NO₃, or 100 mM K₃PO₄ to 10% TSA, and the pH was set at 7.3 using HCl or NaOH as needed.

To test for antagonist activity toward *Staphylococcus* and for exDNase activity, DNA plates (catalog number 263220; Difco) were seeded with *Staphylococcus* as described above, and soil isolates were tested for their ability to inhibit *Staphylococcus* and to break down DNA in the agar plates. The positive control for exDNase activity was *Serratia* sp. CWZ222, or *Pseudomonas* sp. strain CVAP#3 for a ZOI on *Staphylococcus* (Fig. S8-1), and the negative-control *Staphylococcus* colonies were transferred and spotted on the plate using a sterile flat toothpick. Plates were incubated overnight at 25°C, and the zone of inhibition (ZOI) was determined, as well as the exDNase activity by flooding the plate with 1 N HCI (23).

Determination of connectance and AI. To determine the connectance (10) of a population, a colony used as the indicator was used to seed a 100% TSA plate as described above. Soil isolates were transferred onto this indicator strain along with controls, and plates were stored at 4°C for 3 to 5 days. After overnight incubation at 25°C, the plates were analyzed for the appearance of a ZOI. Data were recorded as positive (+) when a ZOI was produced or negative (-) when no ZOI was produced. This procedure was done with the GS₅₃ population (53 × 53) and the FSS₂₅ population (25 × 25). The experiments were repeated three times and quality control (QC) performed, leading to the inclusion of 48 isolates in the GS₄₈ population and 23 isolates in the FSS₂₃ population (Table 2). The connectance was determined by the number of positive results (an isolate producing a ZOI on an indicator) divided by the total tests performed.

The aggressiveness index (AI) was determined for every strain by taking the number of other strains antagonized by a given strain and subtracting the number of other strains antagonizing it (15).

Estimation of coexistence using a ProbCA. We developed a probabilistic cellular automaton (ProbCA) in Mathematica (11.3.0.0). The details of the development and code are provided in Text S4. The initial grid (350×350) was randomly and uniformly seeded. Each cell was chosen as the focal cell in random order and interacted with a randomly chosen neighbor cell according to the interaction matrix, and 100 simulations of 100 iterations were run.

Coexistence (richness) is the number of isolates present in the dynamic equilibrium at the end of the simulation and is used in microbiology as a measure of diversity that seems to correlate with other regularly used diversity indices. In addition, for reasons described in Text S4, the distribution of isolates in the population in the dynamic equilibrium was used to estimate Shannon diversity.

The input matrices and simulations. The three replicate interaction matrices were used to create a consensus binary interaction matrix using the majority rule. After quality control, five strains from the GS and two from the FSS matrices were removed for being compromised in quality (>10% of the data inconclusive or not tested) or lacking antagonist activity on any indicator, compromising the simulation. Preliminary results showed that coexistence is not necessarily a reliable measure, since coexistence increases with increasing grid size (Text S4). Due to this concern, isolates not involved in any antagonist interactions were excluded from the analysis. The forest subsurface (FSS₂₃) population was composed of 23 isolates (529 tests), and the grassland surface (GS₄₈) population was composed of 48 isolates (2,304

tests). To ensure a good direct comparison between simulations using the GS₄₈ and FSS₂₃ populations (Table 2), the simulations were first seeded using the consensus interaction matrix to determine the 23 best-surviving isolates, followed by running the simulation using these 23 isolates (Table 2). Preliminary experiments indicated that no significant differences were found in ranking and relative presence of the isolates if only this subset of isolates was reseeded and the simulations run again. A similar procedure was applied to identify 13 exDNase-producing isolates from the 38 exDNase isolates in the GS population (GS₃₈) (Table 2).

Microbial identification using a partial 16S rRNA sequence. To determine the genera of the soil isolates, we amplified the 16S rRNA genes and determined their sequences. To amplify the 16S rRNA gene, cells from a single colony were suspended in 100 μ L sterile 1× PBS using a sterile toothpick. Five microliters of this suspension was used as the template in colony PCR using Illustra PuReTaq Ready-To-Go PCR beads (catalog number 46-001-014 [Fisher Scientific]) with 1 μ L of forward primer (pA, 27F, or bac8F, 20 μ M, 5'-AGAGTTTGATCCTGGCTCAG-3') (12, 52, 53), 1 μ L of reverse primer (1492R, 20 μ M, 5'-GGTTACCTTGTTACGACTT-3') (53), and 23 μ L water, totaling 30 μ L. Primers were purchased from IDT. Amplification was achieved using the following program: 94°C for 10 min, 30 cycles of 94°C for 30 s, 58°C for 30 s, 72°C for 1 min 50 s, and finally, 72°C for 10 min. All reactions were performed in an MJ Research PTC-200 Peltier thermal cycler in the Center for Molecular Biology at Smith College. After amplification, a 5-µL sample was tested for the correct fragment size using 1.0% (wt/vol) agarose (CAS Registry Number [CAS RN] 9012-36-6; AmericaBio) gels in 1× Tris-acetate-EDTA (TAE) electrophoresis buffer with SYBR green (catalog number S3312; Invitrogen). The products were cleaned using an EdgeBio Performa DTR gel filtration cartridge (product number 42451; EdgeBio) and Sanger sequenced with the following reaction mixture and program: 1 µL BigDye (catalog number 4337454; Applied Biosystems), 5 μ L double-distilled water (ddH₂O), 0.5 μ L primer (15 pM), and 3.5 μ L template. The primers used were pA or 27F (5'-AGAGTTTGATCCTGGCTCAG-3') (12, 52, 53) and 806R (5'-GGACTACHVGGGTWTCTAAT-3') (54). The thermocycler program for labeling was 96°C for 5 min, 0.7°C/s to 96°C, 96°C for 10 s, 0.7°C/s to 50°C, 50°C for 5 s, 0.7°C/s to 60°C, 60°C for 4 min, repeat 27 times, 1°C/s to 4°C, 4°C forever. After the labeling reaction, the samples were cleaned using gel filtration and the nucleotide sequence determined using an Applied Biosystems 3130xl Genetic Analyzer in the Center for Molecular Biology at Smith College. 16S rRNA gene sequence quality control (QC) was completed using 4Peaks 1.7.1 (nucleobytes.com). Alignments of the 16S rRNA gene sequences were made in the Lasergene package SegMan Pro (version 15.3.0 Intel), and a consensus sequence was generated from at least three sequences from at least two different PCRs and sequenced in reverse and forward directions, unless indicated otherwise. A majority rule with quality weights for consensus calling of 66% was used, and an average of 470 nucleotides/isolate was obtained. The consensus was used as the query for molecular identification using the NCBI databases. The Basic Local Alignment Search Tool was used to navigate through databases and compare nucleotide sequences from the bacterial 16S rRNA gene consensus sequences to a library of published sequences. The NCBI search option "complete genomes" was employed. The first published result with the highest percent identity, highest maximum score, and highest total score was recorded (55-57).

Mathematical manipulations and statistical analysis. All statistical tests confirmed the theory explained in Kanji's *100 Statistical Tests* (58). All statistical tests involving the *t* test were executed in Excel using the one-sided, homoscedastic *t* test, unless mentioned differently. The *F* test was used to estimate significance in variation, and the Poisson test was used to estimate the significance between two observations. Furthermore, the binomial sign test and the Dixon *Q* test for outliers were performed in Excel. To determine if the interactions were equally distributed or differed from an expected population, the expected random distributions were calculated, and the observed data were tested against the expected using the χ^2 test in Excel. PCA was done online using default settings (59).

SUPPLEMENTAL MATERIAL

Supplemental material is available online only. SUPPLEMENTAL FILE 1, PDF file, 6.5 MB. SUPPLEMENTAL FILE 2, MP4 file, 0.6 MB.

ACKNOWLEDGMENTS

Many thanks go out to Bill Peterson, Kevin Shea, and Tom Richardson at Smith College for providing the support from the Office of Faculty Development and the Science Center. In addition, funding for course development was also provided by Joanne Benkley and Paul Wetzel at the Center for the Environment, Ecological Design and Sustainability (CEEDS) and by the Albert F. Blakeslee Trust. Funders had no role in the study design, data collection and interpretation, nor the decision to submit the work for publication.

The CURE Transformations project, which is supported by the National Science Foundation (NSF) through an NSF DUE IUSE grant to the Council on Undergraduate Research (grant number 16-25354), provides the intellectual framework for the implementation of research in undergraduate courses. Personal thanks go out to

Christine White-Ziegler for providing the infrastructure, the intellectual input, and the opportunities for developing this line of research and to Mariana Abarca for proofreading. The jump-start provided by Simon Hernandez, Tiffany Tsang, Nichole Broderick, and Jo Handelsman at the Tiny Earth was also highly appreciated. J.A.C.V. certainly hopes Jo Handelsman can appreciate this paper, which is a result of several years believing in CURE and equity in the classroom (60, 61). Furthermore, thanks go out to Louie Bierwert and Ricardo Racicot, directors of the Center for Molecular Biology at Smith College, for sequencing. In addition, thanks to Mark Anderson, the director of the Center for Aqueous Biogeochemical Research at Smith College, for use of the equipment. Furthermore, Alexandra Hill and Siphokazi Kargbo for their help in data acquisition for %N and %C content of soil. Special thanks go out to Orielle Rollinson for all her hours spent on laboratory prepping, data acquisition, and the many valuable discussions. Moreover, the very valuable comments we received from three anonymous reviewers that helped us quide to a better synthesis of our data were highly appreciated. Lastly, thanks to the many students in the courses BIO133 and BIO205 from Spring 2014 till Spring 2019 at Smith College, who all contributed to this project through many hours of working together, reflections, and discussions.

Aoi Ogawa: connectance, development of cellular automaton, sequencing, writing. Maria Bermudez: sequence analysis. Odrine Habarugira: method development and control experiments, CFU, and % antagonists. Christophe Golé: development of cellular automaton. Gabrielle Joslin: method development and control experiments, CFU, and % antagonists. Taylor McCain: connectance, writing. Autumn Mineo: basic soil edaphic characteristics, preliminary statistics, and writing. Jennifer Wise: exDNase activity, simulations, writing. Katherine Yan: connectance and writing. Julie Xiong: sequencing and analysis of CWZ222. Jan A. C. Vriezen: conception and execution of project, funding, training and assisting with experimentation, data analysis, writing.

REFERENCES

- Renwick MJ, Brogan DM, Mossialos E. 2016. A systematic review and critical assessment of incentive strategies for discovery and development of novel antibiotics. J Antibiot (Tokyo) 69:73–88. https://doi.org/10.1038/ja.2015.98.
- Kolter R, van Wezel GP. 2016. Goodbye to brute force in antibiotic discovery? Nat Microbiol 1:15020. https://doi.org/10.1038/nmicrobiol.2015.20.
- Ling LL, Schneider T, Peoples AJ, Spoering AL, Engels I, Conlon BP, Mueller A, Schäberle TF, Hughes DE, Epstein S, Jones M, Lazarides L, Steadman VA, Cohen DR, Felix CR, Fetterman KA, Millett WP, Nitti AG, Zullo AM, Chen C, Lewis K. 2015. A new antibiotic kills pathogens without detectable resistance. Nature 517:455–459. https://doi.org/10.1038/nature14098.
- Ratcliff WC, Denison RF. 2011. Alternative actions for antibiotics. Science 332:547–548. https://doi.org/10.1126/science.1205970.
- Abrudan MI, Smakman F, Grimbergen AJ, Westhoff S, Miller EL, van Wezel GP, Rozen DE. 2015. Socially mediated induction and suppression of antibiosis during bacterial coexistence. Proc Natl Acad Sci U S A 112: 11054–11059. https://doi.org/10.1073/pnas.1504076112.
- Freilich S, Zarecki R, Eilam O, Segal ES, Henry CS, Kupiec M, Gophna U, Sharan R, Ruppin E. 2011. Competitive and cooperative metabolic interactions in bacterial communities. Nat Commun 2:589. https://doi.org/10 .1038/ncomms1597.
- Kehe J, Ortiz A, Kulesa A, Gore J, Blainey PC, Friedman J. 2021. Positive interactions are common among culturable bacteria. Sci Adv 7:eabi7159. https://doi.org/10.1126/sciadv.abi7159.
- Kerr B, Riley MA, Feldman MW, Bohannan BJM. 2002. Local dispersal promotes biodiversity in a real-life game of rock-paper-scissors. Nature 418: 171–174. https://doi.org/10.1038/nature00823.
- Kirkup BC, Riley MA. 2004. Antibiotic-mediated antagonism leads to a bacterial game of rock-paper-scissors in vivo. Nature 428:412–414. https://doi .org/10.1038/nature02429.
- Mougi A. 2016. The roles of amensalistic and commensalistic interactions in large ecological network stability. Sci Rep 6:29929. https://doi.org/10 .1038/srep29929.
- Russel J, Røder HL, Madsen JS, Burmølle M, Sørensen SJ. 2017. Antagonism correlates with metabolic similarity in diverse bacteria. Proc Natl Acad Sci U S A 114:10684–10688. https://doi.org/10.1073/pnas.1706016114.

- Tyc O, van den Berg M, Gerards S, van Veen JA, Raaijmakers JM, de Boer W, Garbeva P. 2014. Impact of interspecific interactions on antimicrobial activity among soil bacteria. Front Microbiol 5:567. https://doi.org/10 .3389/fmicb.2014.00567.
- Vetsigian K, Jajoo R, Kishony R. 2011. Structure and evolution of Streptomyces interaction networks in soil and in silico. PLoS Biol 9:e1001184. https://doi.org/10.1371/journal.pbio.1001184.
- de Vos MGJ, Zagorski M, McNally A, Bollenbach T. 2017. Interaction networks, ecological stability, and collective antibiotic tolerance in polymicrobial infections. Proc Natl Acad Sci U S A 114:10666–10671. https://doi .org/10.1073/pnas.1713372114.
- Zapien-Campos R, Olmedo-Älvarez G, Santillan M. 2015. Antagonistic interactions are sufficient to explain self-assemblage of bacterial communities in a homogeneous environment: a computational modeling approach. Front Microbiol 6:489. https://doi.org/10.3389/fmicb.2015.00489.
- Abrudan MI, Brown S, Rozen DE. 2012. Killing asmeans of promoting biodiversity. BiochemSoc Trans 40:1512–1516. https://doi.org/10.1042/ BST20120196.
- Iavicoli I, Fontana L, Agathokleous E, Santocono C, Russo F, Vetrani I, Fedele M, Calabrese EJ. 2021. Hormetic dose responses induced by antibiotics in bacteria: a phantom menace to be thoroughly evaluated to address the environmental risk and tackle the antibiotic resistance phenomenon. Sci Total Environ 798:149255. https://doi.org/10.1016/j.scitotenv.2021.149255.
- Stubbendieck RM, Straight PD. 2016. Multifaceted interfaces of bacterial competition. J Bacteriol 198:2145–2155. https://doi.org/10.1128/JB.00275-16.
- Stubbendieck RM, Vargas-Bautista C, Straight PD. 2016. Bacterial communities: interactions to scale. Front Microbiol 7:1234. https://doi.org/10 .3389/fmicb.2016.01234.
- 20. Curtis TP, Sloan WT. 2005. Exploring microbial diversity—a vast below. Science 309:1331–1333. https://doi.org/10.1126/science.1118176.
- Hibbing ME, Fuqua C, Parsek MR, Peterson SB. 2010. Bacterial competition: surviving and thriving in the microbial jungle. Nat Rev Microbiol 8: 15–25. https://doi.org/10.1038/nrmicro2259.

- 22. Cascales E, Buchanan SK, Duché D, Kleanthous C, Lloubès R, Postle K, Riley M, Slatin S, Cavard D. 2007. Colicin biology. Microbiol Mol Biol Rev 71:158–229. https://doi.org/10.1128/MMBR.00036-06.
- 23. Andrade JP, de Souza HG, Ferreira LC, Cnockaert M, De Canck E, Wieme AD, Peeters C, Gross E, De Souza JT, Marbach PAS, Góes-Neto A, Vandamme P. 2021. *Burkholderia perseverans* sp. nov., a bacterium isolated from the Restinga ecosystem, is a producer of volatile and diffusible compounds that inhibit plant pathogens. Braz J Microbiol 52:2145–2152. https://doi.org/10.1007/s42770-021-00560-w.
- Logan NA, DeVos P. 2015. Bacillus, p 1–163. In Whitman WB, Rainey F, Kämpfer P, Trujillo M, Chun J, DeVos P, Hedlund B, Dedysh S (ed). Bergey's manual of systematics of Archaea and Bacteria, 1st ed. Wiley, Hoboken, NJ.
- Ma L-S, Hachani A, Lin J-S, Filloux A, Lai E-M. 2014. Agrobacterium tumefaciens deploys a superfamily of type VI secretion DNase effectors as weapons for interbacterial competition in planta. Cell Host Microbe 16:94–104. https://doi.org/10.1016/j.chom.2014.06.002.
- Minh Tran T, MacIntyre A, Khokhani D, Hawes M, Allen C. 2016. Extracellular DNases of *Ralstonia solanacearum* modulate biofilms and facilitate bacterial wilt virulence: extracellular DNases modulate *R. solanacearum* biofilm and virulence. Environ Microbiol 18:4103–4117. https://doi.org/10 .1111/1462-2920.13446.
- Mulcahy H, Charron-Mazenod L, Lewenza S. 2010. *Pseudomonas aeruginosa* produces an extracellular deoxyribonuclease that is required for utilization of DNA as a nutrient source. Environ Microbiol 12:1621–1629. https://doi.org/10.1111/j.1462-2920.2010.02208.x.
- 28. Kamino LN, Gulden RH. 2021. The effect of crop species on DNase-producing bacteria in two soils. Ann Microbiol 71:14. https://doi.org/10 .1186/s13213-021-01624-w.
- Zukswert JM, Bellemare J, Rhodes AL, Sweezy T, Gallogly M, Ms A, Taylor RS. 2014. Forest community structure differs, but not ecosystem processes, 25 years after eastern hemlock removal in an accidental experiment. Southeast Nat 13:61–87.
- Girvan MS, Bullimore J, Pretty JN, Osborn AM, Ball AS. 2003. Soil type is the primary determinant of the composition of the total and active bacterial communities in arable soils. Appl Environ Microbiol 69:1800–1809. https://doi.org/10.1128/AEM.69.3.1800-1809.2003.
- Hassink J, Bouwman LA, Zwart KB, Bloem J, Brussaard L. 1993. Relationships between soil texture, physical protection of organic matter, soil biota, and C and N mineralization in grassland soils. Geoderma 57: 105–128. https://doi.org/10.1016/0016-7061(93)90150-J.
- Hossain MZ, Aziz CB, Saha ML. 2012. Relationships between soil physicochemical properties and total viable bacterial counts in Sunderban mangrove forests, Bangladesh. Dhaka Univ J Biol Sci 21:169–175. https://doi .org/10.3329/dujbs.v21i2.11515.
- Kaiser K, Wemheuer B, Korolkow V, Wemheuer F, Nacke H, Schöning I, Schrumpf M, Daniel R. 2016. Driving forces of soil bacterial community structure, diversity, and function in temperate grasslands and forests. Sci Rep 6:33696. https://doi.org/10.1038/srep33696.
- Lauber CL, Strickland MS, Bradford MA, Fierer N. 2008. The influence of soil properties on the structure of bacterial and fungal communities across land-use types. Soil Biol Biochem 40:2407–2415. https://doi.org/10 .1016/j.soilbio.2008.05.021.
- 35. Zhalnina K, Dias R, de Quadros PD, Davis-Richardson A, Camargo FAO, Clark IM, McGrath SP, Hirsch PR, Triplett EW. 2015. Soil pH determines microbial diversity and composition in the park grass experiment. Microb Ecol 69:395–406. https://doi.org/10.1007/s00248-014-0530-2.
- Raaijmakers JM, De Bruijn I, Nybroe O, Ongena M. 2010. Natural functions of lipopeptides from *Bacillus* and *Pseudomonas*: more than surfactants and antibiotics. FEMS Microbiol Rev 34:1037–1062. https://doi.org/10 .1111/j.1574-6976.2010.00221.x.
- Van Lanen SG, Shen B. 2006. Progress in combinatorial biosynthesis for drug discovery. Drug Discov Today Technol 3:285–292. https://doi.org/10 .1016/j.ddtec.2006.09.014.
- Sánchez S, Chávez A, Forero A, García-Huante Y, Romero A, Sánchez M, Rocha D, Sánchez B, Ávalos M, Guzmán-Trampe S, Rodríguez-Sanoja R, Langley E, Ruiz B. 2010. Carbon source regulation of antibiotic production. J Antibiot (Tokyo) 63:442–459. https://doi.org/10.1038/ja.2010.78.
- Vriezen JAC, Valliere M, Riley MA. 2009. The evolution of reduced microbial killing. Genome Biol Evol 1:400–408. https://doi.org/10.1093/gbe/evp042.
- Blagodatskaya E, Kuzyakov Y. 2013. Active microorganisms in soil: critical review of estimation criteria and approaches. Soil Biol Biochem 67: 192–211. https://doi.org/10.1016/j.soilbio.2013.08.024.

- 41. Raynaud X, Nunan N. 2014. Spatial ecology of bacteria at the microscale in soil. PLoS One 9:e87217. https://doi.org/10.1371/journal.pone.0087217.
- Levy-Booth DJ, Campbell RG, Gulden RH, Hart MM, Powell JR, Klironomos JN, Pauls KP, Swanton CJ, Trevors JT, Dunfield KE. 2007. Cycling of extracellular DNA in the soil environment. Soil Biol Biochem 39:2977–2991. https://doi.org/10.1016/j.soilbio.2007.06.020.
- López D, Vlamakis H, Losick R, Kolter R. 2009. Cannibalism enhances biofilm development in *Bacillus subtilis*: cannibalism and biofilm matrix. Mol Microbiol 74:609–618. https://doi.org/10.1111/j.1365-2958.2009.06882.x.
- Price NPJ, Rooney AP, Swezey JL, Perry E, Cohan FM. 2007. Mass spectrometric analysis of lipopeptides from *Bacillus* strains isolated from diverse geographical locations: MALDI-TOF MS analysis of *Bacillus* sp. biomarkers. FEMS Microbiol Lett 271:83–89. https://doi.org/10.1111/j.1574-6968.2007.00702.x.
- 45. Turk V, Rehnstam AS, Lundberg E, Hagstrom A. 1992. Release of bacterial DNA by marine nanoflagellates, an intermediate step in phosphorus regeneration. Appl Environ Microbiol 58:3744–3750. https://doi.org/10 .1128/aem.58.11.3744-3750.1992.
- 46. McCall BL, Vriezen JAC. 2021. Draft genome sequences of *Staphylococcus* sp. strain CWZ226, of unknown origin, and *Pseudomonas* sp. strain CVAP#3, antagonistic to strain CWZ226. Microbiol Resour Announc 10: e00688-21. https://doi.org/10.1128/MRA.00688-21.
- Kanegusuku GA, Stankovic IN, Cote-Hammarlof PA, Yong PH, White-Ziegler CA. 2021. A shift to human body temperature (37°C) rapidly reprograms multiple adaptive responses in *Escherichia coli* that would facilitate niche survival and colonization. J Bacteriol 203:e00363-21. https://doi .org/10.1128/JB.00363-21.
- Christensen P, Cook FD. 1978. *Lysobacter*, a new genus of nonfruiting, gliding bacteria with a high base ratio. Int J Syst Evol 28:367–393. https:// doi.org/10.1099/00207713-28-3-367.
- 49. Sutton S. 2011. Accuracy of plate counts. J Valid Technol 17:42–46.
- Vriezen JA, de Bruijn FJ, Nüsslein KR. 2012. Desiccation induces viable but non-culturable cells in *Sinorhizobium meliloti* 1021. AMB Express 2:6. https:// doi.org/10.1186/2191-0855-2-6.
- 51. Park EJ. 2015. Metal speciation, mixtures and environmental health impacts. PhD dissertation. Harvard T.H. Chan School of Public Health, Boston, MA.
- Edwards U, Rogall T, Blöcker H, Emde M, Böttger EC. 1989. Isolation and direct complete nucleotide determination of entire genes. Characterization of a gene coding for 16S ribosomal RNA. Nucleic Acids Res 17: 7843–7853. https://doi.org/10.1093/nar/17.19.7843.
- Fierer N, Jackson RB. 2006. The diversity and biogeography of soil bacterial communities. Proc Natl Acad Sci U S A 103:626–631. https://doi.org/ 10.1073/pnas.0507535103.
- Bokulich NA, Joseph CML, Allen G, Benson AK, Mills DA. 2012. Next-generation sequencing reveals significant bacterial diversity of botrytized wine. PLoS One 7:e36357. https://doi.org/10.1371/journal.pone.0036357.
- Altschul S, Gish W, Miller W, Myers E, Lipman D. 1990. Basic Local Alignment Search Tool. J Mol Biol 215:403–410. https://doi.org/10.1016/S0022 -2836(05)80360-2.
- Ciufo S, Kannan S, Sharma S, Badretdin A, Clark K, Turner S, Brover S, Schoch CL, Kimchi A, DiCuccio M. 2018. Using average nucleotide identity to improve taxonomic assignments in prokaryotic genomes at the NCBI. Int J Syst Evol Microbiol 68:2386–2392. https://doi.org/10.1099/ijsem.0.002809.
- 57. Earl JP, Adappa ND, Krol J, Bhat AS, Balashov S, Ehrlich RL, Palmer JN, Workman AD, Blasetti M, Sen B, Hammond J, Cohen NA, Ehrlich GD, Mell JC. 2018. Species-level bacterial community profiling of the healthy sinonasal microbiome using Pacific Biosciences sequencing of full-length 16S rRNA genes. Microbiome 6:190. https://doi.org/10.1186/s40168-018-0569-2.
- 58. Kanji GK. 2006. 100 statistical tests, 3rd ed. Sage Publications, Thousand Oaks, CA.
- Metsalu T, Vilo J. 2015. ClustVis: a web tool for visualizing clustering of multivariate data using principal component analysis and heatmap. Nucleic Acids Res 43:W566–W570. https://doi.org/10.1093/nar/gkv468.
- Handelsman J, Elgin S, Estrada M, Hays S, Johnson T, Miller S, Mingo V, Shaffer C, Williams J. 2022. Achieving STEM diversity: fix the classrooms. Science 376:1057–1059. https://doi.org/10.1126/science.abn9515.
- 61. Hurley A, Chevrette MG, Acharya DD, Lozano GL, Garavito M, Heinritz J, Balderrama L, Beebe M, DenHartog ML, Corinaldi K, Engels R, Gutierrez A, Jona O, Putnam JHI, Rhodes B, Tsang T, Hernandez S, Bascom-Slack C, Blum JE, Price PA, Davis D, Klein J, Pultorak J, Sullivan NL, Mouncey NJ, Dorrestein PC, Miller S, Broderick NA, Handelsman J. 2021. Tiny Earth: a big idea for STEM education and antibiotic discovery. mBio 12:e03432-20. https://doi.org/10.1128/mBio.03432-20.