#### FOCUS ARTICLE



Check for updates

## Streambed migration frequency drives ecology and biogeochemistry across spatial scales

Ute Risse-Buhl<sup>1,2</sup> <sup>(D)</sup> | Shai Arnon<sup>3</sup> <sup>(D)</sup> | Edo Bar-Zeev<sup>3</sup> <sup>(D)</sup> | Anna Oprei<sup>4</sup> <sup>(D)</sup> | Aaron I. Packman<sup>5</sup> <sup>(D)</sup> | Ignacio Peralta-Maraver<sup>6,7</sup> <sup>(D)</sup> | Anne Robertson<sup>8</sup> <sup>(D)</sup> | Yoni Teitelbaum<sup>3</sup> <sup>(D)</sup> | Michael Mutz<sup>4</sup>

<sup>1</sup>Department of River Ecology, Helmholtz Centre for Environmental Research, Magdeburg, Germany

<sup>2</sup>Institute for Environmental Sciences, Rheinland-Pfälzische Technische Universität Kaiserslautern-Landau, Landau, Germany

<sup>3</sup>Zuckerberg Institute for Water Research, The Jacob Blaustein Institutes for Desert Research, Ben-Gurion University of the Negev, Midreshet Ben-Gurion, Israel

<sup>4</sup>Department of Aquatic Ecology, Brandenburg University of Technology, Bad Saarow, Germany

<sup>5</sup>Northwestern Center for Water Research, Department of Civil and Environmental Engineering, Northwestern University, Evanston, Illinois, USA

<sup>6</sup>Departamento de Ecología e Instituto del Agua, University of Granada, Granada, Spain

<sup>7</sup>Research Unit Modeling Nature, University of Granada, Granada, Spain

Revised: 16 December 2022

<sup>8</sup>Department of Life & Health Sciences, University of Roehampton, London, UK

#### Correspondence

Ute Risse-Buhl, Department of Biology, Rheinland-Pfälzische Technische Universität Kaiserslautern-Landau, Erwin-Schroedinger-Straße 14, 67663 Kaiserslautern, Germany. Email: ute.risse-buhl@rhrk.uni-kl.de

#### Funding information

Carl-Zeiss-Stiftung, Grant/Award Number: P2021-00-004; Deutsche Forschungsgemeinschaft, Grant/Award Numbers: MU 1464/7-1, RI 2093/2-1; Israel Science Foundation, Grant/Award Number: 682/17; Israeli Science Foundation, Grant/Award Number: 944 \21; NSF-BSF, Grant/Award Number: EAR-1734300; UK-Israel Science Fellowship Scheme 2018-2019

Edited by: Jan Seibert, Co-Editor-in-Chief

#### Abstract

The bed of fluvial ecosystems plays a major role in global biogeochemical cycles. All fluvial sediments migrate and although responses of aquatic organisms to such movements have been recorded there is no theoretical framework on how the frequency of sediment movement affects streambed ecology and biogeochemistry. We here developed a theoretical framework describing how the moving-resting frequencies of fine-grained sediments constrain streambed communities across spatial scales. Specifically, we suggest that the most drastic impact on benthic and hyporheic communities will exist when ecological and biogeochemical processes are at the same temporal scale as the sediment moving-resting frequency. Moreover, we propose that the simultaneous occurrence of streambed patches differing in morphodynamics should be considered as an important driver of metacommunity dynamics. We surmise that the frequency of patch transition will add new dimensions to the understanding of biogeochemical cycling and metacommunities from micro-habitat to segment scales. This theoretical framework is important for fluvial ecosystems with frequent sediment movement, yet it could be applied to any other dynamic habitat.

#### This article is categorized under:

Water and Life > Nature of Freshwater Ecosystems

This is an open access article under the terms of the Creative Commons Attribution-NonCommercial-NoDerivs License, which permits use and distribution in any medium, provided the original work is properly cited, the use is non-commercial and no modifications or adaptations are made. © 2023 The Authors. *WIREs Water* published by Wiley Periodicals LLC.

#### **KEYWORDS**

benthic and hyporheic communities, environmental filtering, hierarchical spatial and temporal scaling, process time scale, sediment transport frequency

#### **1** | THE IMPORTANCE OF STREAMBEDS IN FLUVIAL ECOSYSTEMS

The metabolic hotspots of fluvial ecosystems often occur at and below the water-sediment interface, making the bed of streams and rivers an important zone for global biogeochemical transformation processes (Battin et al., 2016; Krause et al., 2017; Peralta-Maraver et al., 2021). Streams and rivers (hereafter streams) have a disproportionate impact on global biogeochemical cycles given their small surface area relative to other aquatic and terrestrial environments (Cole et al., 2007; Drake et al., 2018). Emission of carbon dioxide related to streams was recently estimated to range between 2% and 7% (1.1–3.9 Pg C year<sup>-1</sup>) of terrestrial emissions (54 Pg C year<sup>-1</sup>), even though streams cover only 0.6% of the non-glaciated land surface (Allen & Pavelsky, 2018; Regnier et al., 2013). Streams also recycle nutrients at much higher rates than other aquatic systems, as they are imported from terrestrial environments and rapidly mineralized by diverse benthic and hyporheic communities (Battin et al., 2016; Romaní et al., 2008).

Despite the rigidity of individual sediment grains, most streambeds migrate continuously or episodically when the shear stress of flow exceeds the force needed to move individual grains (Bridge, 2003). Most streambeds migrate, but sediment transport is expected to be much more frequent in lowland streams with sizes of grain that range between fine sand to very fine gravel (0.063–4 mm, respectively), defined herein as "fine sediments" (Buffington & Montgomery, 2021; Frings, 2008; Kondolf et al., 2016). Sediment grains spend variable amounts of time moving (rolling, saltating, or suspended) and resting under different flow conditions. Consequently, the structure of stream channels and floodplains develops through the long-term effects of geology (e.g., sediment inputs), hydrology (e.g., water flow), and biology (e.g., stabilization by vegetation) (Buffington & Montgomery, 2021; Dietrich & Perron, 2006; Kondolf et al., 2016; Montgomery & Buffington, 1997; Perucca et al., 2007). Hereafter we refer to periods where streambed sediments are in motion as the "moving phase" and when not in motion as the "resting phase" (Nikora et al., 2002; Zhang et al., 2012).

The ecological and biogeochemical consequences of morphodynamics are well recognized and have generated several scientific concepts on disturbance (Lake, 2000; Resh et al., 1988) and ecological memory (Ortmann-Ajkai et al., 2014). However, the concept of disturbance in stream ecology focuses primarily on catastrophic flood events that occur infrequently (recurrence interval of years or decades) and over large spatial scales (typically catchment scales). This binary view of normal low flow (resting phase) versus infrequent disturbance (moving phase) is overly simplistic and does not capture processes occurring in streams with frequent movement of fine sediments (Kondolf et al., 2016; Ryo et al., 2019; Segura et al., 2011). Here we extend this concept of disturbance in stream ecology to streambeds with fine sediments, and introduce a new theoretical framework on how moving-resting frequency affects stream ecology and biogeochemistry. We first describe the physical environment in migrating streambeds. Afterwards we discuss the ecological and biogeochemical consequences of sediment migration across a range of frequencies and spatial scales by incorporating insights from freshwater and marine environments. Finally, we explore new research perspectives that this viewpoint generates.

## 2 | MORPHODYNAMICS OF FINE-GRAINED FLUVIAL ECOSYSTEMS

Streambed habitat reflects the physical structure and dynamics from a single sediment grain to the whole channel morphodynamics over spatial scales from micrometers to kilometers (Figure 1). Modeling frameworks evaluate sediment transport of single grains in terms of the frequency, intensity, and duration of moving versus resting phases (Figure 1a) (Nikora et al., 2002; Zhang et al., 2012). Sediment grains are mobilized at the streambed–water interface via turbulent eddies or collision with other moving grains. At meso-habitat scale (1 cm to 100 m), coupling between flow, sediment transport, and channel produces migrating bedforms such as upper stage plane bed, ripples, dunes and bars (Figure 1b), which can be classified based on the grain size, bed and water level slope, water depth, flow velocity, and degree of lateral channel confinement (Andreotti et al., 2012; Garcia, 2008; Montgomery & Buffington, 1997; Rinaldi et al., 2016). Single sediment grains within migrating bedforms experience a moving-resting cycle with specific migration frequency and resting duration. High migration frequency and short resting duration are typical for an upper stage

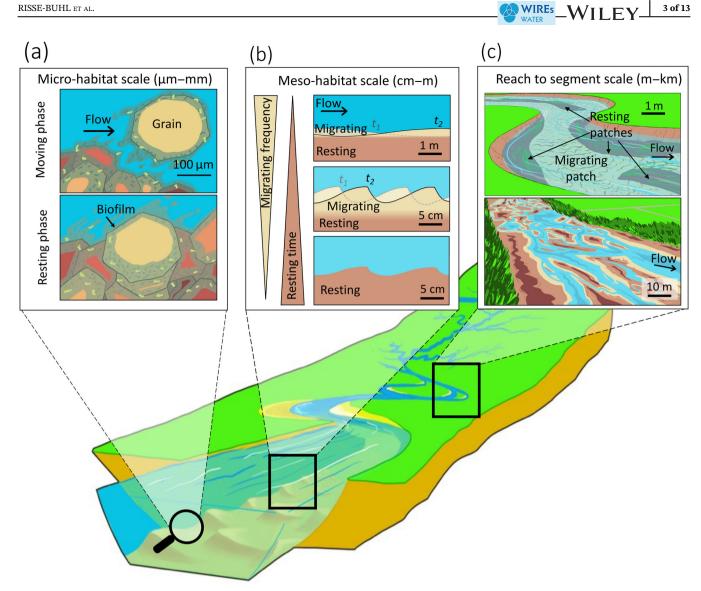


FIGURE 1 Morphodynamics of fine sediments (fine sand to very fine gravel) at different spatial scales. (a) Micro-habitat scale (µm to mm) where the colonization of a single sediment grain by biofilms may have a large impact on the grain scale environment and migration. These sediment grains can move (top panel) or rest (bottom panel) according to the flow regime. (b) Meso-habitat scale (cm to m) highlights that bedforms are established at a specific range of flow velocities and sediment grain sizes. Migration frequency and resting duration of sediment grains are consequences of bedform dynamics. The panels illustrate the sideview of the streambed at the upper stage plane bed (top panel), where a thin layer of grains is almost continuously in saltation and bounce over the streambed that undulates in waves (few mm in height), migrating ripples and dunes (central panel), where the mobile streambed layer corresponds to the height of the bedform (about 1 cm in height), and the resting streambed sediments without a mobile layer (bottom panel). Bedform height is not to scale and displayed higher than in reality. (c) Streambed heterogeneity at reach to segment scale (m to km) often constitutes a mosaic of migrating and resting sediment patches. This streambed heterogeneity is controlled by the history of sediment inputs, flow conditions, and lateral interactions within the active channel and between the channel, banks, and floodplain.

plane bed, in which the surficial sediment grains are almost continuously in saltation and bounce downstream (Naqshband et al., 2017), while lower migration frequency and intermediate resting duration typically occurs in ripples and dunes, in which sediment grains become buried and remobilized by the passage of bedforms (Buffington & Montgomery, 2021; Kondolf et al., 2016; Raudkivi, 2006).

Stream velocity varies both laterally and longitudinally, producing patches of different migrating bedforms that evolve over time at the reach scale (100 m to km). Even small changes of discharge generate fluctuations in the size and spatial distribution of the migrating and resting patches (Figure 1c). The proportion of the streambed in motion is modulated by in-stream features such as boulders and woody debris providing local streambed stability (Krause et al., 2014).

3 of 13

At the segment scale (10–100 km), resting times of sediments in the range of months to decades are set by the frequency of large flood events, which are sometimes described as channel-forming flows (Phillips & Jerolmack, 2014; Tockner et al., 2003). These infrequent events yield large-scale morphodynamic processes extending over the entire channel such as deep sediment relocation, meander scrolling, and channel avulsion (David et al., 2017; Venditti et al., 2017). During such high flow events, grains can be transported between the active stream channel and the floodplain, as well as within the main channel (Buffington & Montgomery, 2021; Frings, 2008; Lewin et al., 2017).

4 of 13

# 3 | ECOLOGICAL IMPLICATIONS OF MOVING AND RESTING PHASE AT MICRO-HABITAT SCALE

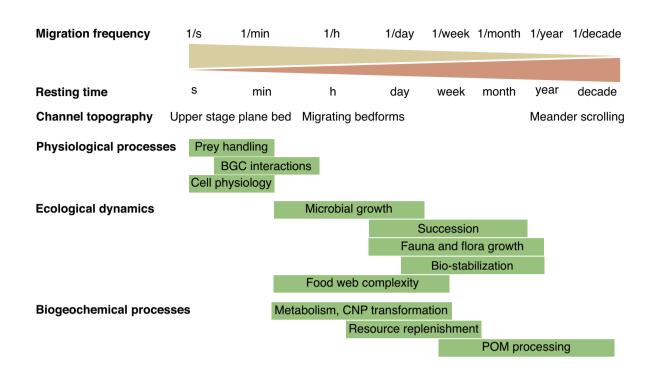
Phototrophic and heterotrophic microbes (including both prokaryotes and protists) are the building blocks of biofilms that grow on all types of aquatic surfaces (micro-habitat scale; Figure 1a). Stability of sediment grains during the resting phase enables the growth and succession of biofilms in the benthic and hyporheic zone (Figure 1a), which in turn supports diverse food web interactions. Phototrophic and heterotrophic microbes utilize the available resources from the sediment pore water, such as oxygen, nutrients and dissolved organic carbon, which are then transferred to micro-grazers such as protists (Augspurger et al., 2008; Norf et al., 2009; Risse-Buhl et al., 2012). Metazoa, including both meio- and macro-fauna, contribute to food web interactions during prolonged resting phases, when they have time to reproduce (Majdi & Traunspurger, 2015; Weitere et al., 2018). In turn, grazing promotes energy transfer and boosts microbial activity by keeping the community in a productive state (Oprei et al., 2019; Peralta-Maraver et al., 2019; Risse-Buhl et al., 2015).

Excessive microbial growth that includes secretions of extracellular polymeric substances (EPS) during prolonged resting phases (weeks to months) stabilizes the streambed, reduces the frequency of sediment remobilization and decreases the exchange of resources between the overflowing water and the streambed (Caruso et al., 2017; Cook et al., 2020; Mendoza-Lera & Mutz, 2013). The processes of bio-stabilization and bio-clogging are caused not only by bacterial biofilms and algal mats, which produce multicellular networks and EPS (Gerbersdorf et al., 2020; Malarkey et al., 2015; Paterson et al., 2018), but also by macrofaunal structures such as mucus, silk secretions, cases and nets (Albertson et al., 2019; Majdi et al., 2014).

The biofilm community colonizing single sediment grains can endure significant changes in environmental conditions as migration frequencies change. Migration frequency of once per minute to once per day occurring in migrating bedforms affects both structure and growth rates of the benthic and hyporheic communities (Scheidweiler et al., 2021; Zlatanović et al., 2017). Microbial doubling times are generally in the range of minutes to days (Aufrecht et al., 2019; Bar-Zeev et al., 2012; Sinsabaugh et al., 2015; Zlatanović et al., 2017; Figure 2). Consequently, microbial growth will be inhibited when growth rates are longer than the sediment migration frequency, owing to the physical disturbance of the growing community. As growth rates are species-specific, sediment migration results in the selection of species with growth rates shorter than migration frequency and a shift in the benthic and hyporheic community structure compared to resting streambed patches. Furthermore, diatom frustules, biofilms as well as benthic and hyporheic fauna endure mechanical damages and dispersal by abrasion when colliding with moving grains (Delgado et al., 1991; Luce et al., 2010; Luce et al., 2013; Miller, 1989; Probandt et al., 2018). Motile bacteria and diatoms are mainly found to colonize fissures and crevices of sand grains, where they are subject to lower mechanical forces (Ahmerkamp et al., 2020; Jewson et al., 2006; Krejci & Lowe, 1986; Miller et al., 1987; Weise & Rheinheimer, 1978). The mechanical forces thus act as an environmental filter constraining the community to the habitable area of sediment grains.

We expect biological traits such as small cell size, motility, robust shells or diatom frustules, as well as greater adhesion (e.g., via stalks and EPS) to be advantageous for the colonization of moving grains in migrating patches because of the greater mechanical challenge associated with this habitat (DeJonge, 1985; Delgado et al., 1991; Hou et al., 2018; Risse-Buhl et al., 2014). High shear forces cause changes in EPS properties (e.g., stiffer and more adhesive polysaccharides) and production (e.g., over production of polysaccharides), resulting in stronger cohesion and surface adhesion (Chew et al., 2014; Hou et al., 2018; Stoodley et al., 2002). Despite this protective mechanism, sediment migration can impair cell–cell interactions (Remis et al., 2010; Rivera-Yoshida et al., 2018), and consequently limit interactions between microbial species. Important limitations could include: impaired quorum sensing when flow causes autoinducer molecules to disperse (Kim et al., 2016), decreased proximity of fungi and bacteria repressing transfer of nutrients (Arnon et al., 2010; Lock et al., 1984), and decreased bacterial utilization of labile DOC released by phototrophic protists (Romaní et al., 2004).

Doubling times of meio- and macro-fauna and -flora range from weeks to months (Olsen & Townsend, 2005; Riis & Sand-Jensen, 2006; Utz & Hilderbrand, 2011), impairing the ability of these organisms to colonize migrating sediments with shorter resting phases. However, larger organisms use strategies such as burrowing, adhesion or clinging to



**FIGURE 2** Time scales of the migration frequency and resting duration of sediments at which they impact different physiological processes, ecological dynamics and biogeochemical processes. Biogeochemical (BGC) interactions include inter-species communications (e.g., bacterial quorum sensing) as well as multi-species interaction (e.g., exchange of labile DOC components, competition). Temporal scales at which sediment migration may affect the different cellular, ecological, and biogeochemical processes are indicated by the green boxes.

surfaces, and case construction to resist the effects of mechanical forces, allowing them to survive in such conditions (de la Fuente et al., 2018; Kryvokhyzhyna et al., 2022; Swan & Palmer, 2000; Wood et al., 2010). Indeed, these strategies might act to stabilize the streambed and reduce sediment transport as pointed out above. Similarly, excessive development of biofilms, density-dependent growth of macrophytes (e.g., Heffernan, 2008), or regrowth of macrophytes from rhizomes resulting in bio-stabilization and bio-clogging of sediments. As a consequence, this process causes longer sediment resting duration and accelerates community recovery from streambed migration. Thus, ecological mechanisms of resistance and resilience to sediment migration at micro-habitat scale introduce non-linearity into the relationship between migration frequency at meso-habitat scale and ecological processes.

## 4 | CONCEPT OF MIGRATION FREQUENCY AT MESO-HABITAT SCALE

Based on the current state of knowledge related to migrating streambeds, we hypothesize that physiological processes, ecological dynamics and biogeochemical processes are affected primarily when their time scale and the migration frequency of the sediment coincide (Figure 2). The different migration frequencies and resting durations of the sediment grains considered in Figure 2 can be aligned with the channel topography of fine sediment streambeds at meso-habitat scale (Figure 1b). Migration frequency acts synergistically as an environmental filter (Poff, 1997) for the establishment of streambed communities and their associated biogeochemical functions. Considering physiological processes at the cellular level (Figure 2), prey handling time of flagellates (heterotrophic protists) ranges between 2 and 300 s (Boenigk & Arndt, 2002) indicating inhibition of food uptake at sediment migration frequencies in the range of 1/s and 1/min. Thus, at shorter resting phases, inter-species interactions in food webs or during succession and biological modification of habitat are likely to be less important relative to mechanical challenges.

Considering ecological dynamics (Figure 2), the frequency of sediment migration generally correlates negatively with biofilm biomass (Luce et al., 2010). We expect a shift in microbial community composition from eukaryotic phototrophs to prokaryotic phototrophs and heterotrophs with increasing migration frequency. Biofilms regrowing within hours can colonize migrating streambed patches (Aufrecht et al., 2019), while macrophytes with a slower growth rate (doubling

WIRES\_WILFY 5 of 13

time of aboveground biomass 10–100 days) require long periods of sediment stability to re-establish following floods (Nielsen & Sand-Jensen, 1991; Riis & Sand-Jensen, 2006). Microcrustacean densities are higher in low shear stress patches (and thus lower frequency of sediment movement) where they also recover faster after a flood, than in high shear stress patches (Robertson et al., 1995). Slower growth rates of some meio- and macro-fauna species are expected to truncate food webs, reducing diversity and ecological interactions in migrating sediments compared to resting sediments (Hatton et al., 2019; Xuehua et al., 2009). Biogeochemical processes are differentially affected by migration frequency depending on their processing time or bioavailability, microbial and fauna growth (Figure 2). These examples draw on existing ecological knowledge of biota and sediment morphodynamics. To our knowledge there are no studies explicitly designed to provide proof of our postulated theoretical framework on the ecological and sediment morphodynamic time scales, calling for further complementary research.

# 5 | CONCEPT OF MIGRATING-RESTING PATCH TRANSITION AT LARGER SCALE

Frequent migrating and resting sediment patches occur simultaneously at larger spatial scales (Figure 1c) and should be considered as an important driver of metacommunity dynamics. Within the streambed metacommunity (Leibold et al., 2004), sediment migration horizontally homogenizes the landscape, thus, reducing meiofauna species diversity (Schratzberger & Larcombe, 2014), while supporting their passive horizontal displacement (Levin & DiBacco, 1995). Simultaneously, sediment migration enhances the vertical mixing of benthic and hyporheic interstitial communities (Kryvokhyzhyna et al., 2022). Phototrophs as well as small vermiform meio- and macro-fauna actively move and burrow deeper into the streambed to take refuge avoiding the mechanical challenges in the upper migrating sediment layers (Dole-Olivier et al., 1997; Fegley, 1987; Gaudes et al., 2010; Robertson & Wood, 2010). In addition to the passive displacement and active movement within the streambed, organisms that are unable to burrow might be dispersed into the water column, either as individual organisms or aggregates of streambed sediments and biofilms (protobiofilms), and consequently drift downstream with the flow (Bar-Zeev et al., 2012; Risse-Buhl et al., 2014; Shimeta et al., 2003). Drifting organisms and protobiofilms generally settle in areas of low flow and resting sediment (James et al., 2009; Palmer, 1992) and rapidly recolonize the streambed within minutes to days (Arndt et al., 2003; Bar-Zeev et al., 2012; Robertson, 2000). Thus, the local communities of distinct streambed patches are connected both via downstream drift and dispersal within the streambed. Hence, the heterogeneous distribution of migrating and resting patches within the streambed are expected to increase  $\beta$ -diversity at reach to segment scale.

Species-specific requirements, local patch-specific factors (environmental filters) and dispersal processes all drive  $\beta$ -diversity (Bo et al., 2020; Soininen & Teittinen, 2019; Tolonen et al., 2018). We hypothesize that metacommunity dynamics are structured by the distribution of migrating and resting patches in the reach that includes their spatial arrangement and areal coverage as well as the frequency of migrating-resting patch transition. Consequently, it is likely that benthic and hyporheic communities will experience transport-resting cycles at nested spatio-temporal scales: (1) The moving-resting cycle of single sediment grains occurs at a gradient of frequencies at meso-habitat scale, and (2) the transition of migrating-resting patches occurs at a much lower frequency driven by larger-scale fluvial constraints such as discharge variability, water depth and sediment load at reach to segment scale. The nestedness of spatio-temporal scales highlights the need for cross-scale observations in future research.

It is still unknown to what extent individual populations or communities adapt to new conditions after migratingresting patch transition. Previous studies suggest that communities from migrating ripples can adjust to resting conditions within a few days (Atkinson et al., 2008; Miller, 1989; Zlatanović et al., 2017), although the antecedent sediment transport conditions can affect community responses (Scheidweiler et al., 2021). This behavior suggests that these communities are not only adapted to the most-prevalent environmental conditions of their local habitat, but also to the larger scale frequency of migrating-resting patch transitions of fluvial systems. This idea extends the intermediate disturbance hypothesis that focused on temporal scales (Connell, 1978) to spatio-temporal patterns in streambed habitats driven by larger-scale fluvial processes. Therefore, we suggest that  $\gamma$ -diversity peaks at specific transition frequency and areal coverage of migrating and resting patches, and thus is lower when the whole streambed is composed either of resting or migrating sediment patches.

## **6** | IMPLICATIONS ON BIOGEOCHEMICAL PROCESSES

Availability of resources (e.g., dissolved gases, dissolved and particulate organic matter and nutrients) becomes the primary rate-limiting factor for biological uptake rates during prolonged resting phases at slow flow velocities (Battin et al., 2003; De Falco et al., 2018; Figure 2). Increasing stream water velocity induces bedform migration and improves resource availability by increasing the advective exchange flux between the flowing water and the streambed (Arnon et al., 2007; Wolke et al., 2020). During bedform migration, porewater exchange is also influenced by the movement of the grains. Porewater is released from the upstream part of the bedform and may be trapped in the downstream part of the bedform. This process is commonly termed "turnover" and may become dominant over advection driven exchange at relatively high bedform migration velocities (Elliott & Brooks, 1997; Packman et al., 2000).

The heterotrophic activity and resource recycling is ultimately affected by a combination of resource availability, porewater exchange and local reaction rates (Zheng et al., 2019). For example, local respiration rates and resource removal decrease as bedform migration velocity increases (Ahmerkamp et al., 2017; Wolke et al., 2020; Zheng et al., 2019). In comparison to resting sediments, lower bacterial abundance in migrating sediments could potentially explain the lower heterotrophic respiration rates (Scheidweiler et al., 2021; Zlatanović et al., 2017). In contrast, suppressed primary production in migrating sediments cannot be explained by changes in phototroph biomass (chlorophyll a) (Scheidweiler et al., 2021). Frequent burial within the bedform and resulting light limitation for the phototrophs (Izagirre et al., 2009; Pilditch & Miller, 2006) might impair primary production.

Migrating sediments are in general well connected with the stream water. Exchange flux increases with stream water velocity and bedform migration velocity. However, delivery of nutrients into deeper streambed layers located below the migrating layer (i.e., the upper layer of the streambed that migrates; Figure 1b) is reduced as compared to non-migrating bedforms. The penetration of water, nutrients and particles is impeded when turnover is becoming dominant under fast bedform migration velocities (Teitelbaum et al., 2022; Zheng et al., 2019). Thus, resource transfer to hyporheic communities below the migrating layer is reduced (Elliott & Brooks, 1997; Wolke et al., 2020; Zheng et al., 2019). At the same time, suspended particulate matter is trapped by migrating sediments and accumulates in the zone below the migrating layer (Dallmann et al., 2020; Dallmann et al., 2021; Harvey et al., 2012; Teitelbaum et al., 2022). This accumulation can lead to clogging of the streambed by clay particles (Teitelbaum et al., 2021) but also to increased biogeochemical activity utilizing organic matter (Phillips et al., 2019).

Most of the recent process-based studies focus on the bedform scale (meso-habitat), yet, bedform migration varies spatially in channel cross-section and longitudinally (Figure 1c). The frequency of migrating-resting patch transition thus adds new dimensions to the understanding of biogeochemical cycling at reach and segment scales. Although it is acknowledged that the transport of fine sediments in fluvial ecosystems is a common phenomenon in lowland alluvial and glacial landscapes, the global distribution and statistics of streambed migration is still unclear. In Germany, approximately 16% of all stream length have sandy sediments (source of data: https://www.gewaesser-bewertung.de/index.php?article\_id=11&clang=0) and flow velocities that can cause streambed migration under base flow conditions. Some local studies from England, Wales and US suggest that sandy streambeds are migrating about 50% of the time (Booker & Dunbar, 2008; Verzano et al., 2012), but the impact of streambed migration on biogeochemical processes in

#### **BOX 1** Future research needs

The major research gaps related to the coupling between fine sediment transport in fluvial ecosystems, dynamics of benthic and hyporheic communities, and biogeochemical processes across spatio-temporal scales are:

- 1. Estimates of the global distribution and statistics of streambed migration are needed to improve our understanding on how this streambed heterogeneity created by the moving-resting frequency of sediment grains (micro-habitat scale) and migrating-resting patch transitions (meso-habitat and reach scale) can provide important control points for ecosystem biogeochemistry and affect global biogeochemical cycles.
- 2. The theoretical framework we postulate on the coupling between ecological and biogeochemical consequences across a range of sediment migration frequencies at micro- and meso-habitat scale remains to be demonstrated.
- 3. Cross-scale observations are needed because nonlinear relationships may exist when the response of ecological dynamics and biogeochemical processes to the frequency of sediment migration at the micro-habitat are scaled up and translated to larger spatial scales where communities experience the transition of migrating-resting patches.

those streams was not evaluated. Global data on the distribution of mobile-bed streams and the frequency of streambed migration in these systems is still lacking, as most prior studies have focused on either morphological classification or total catchment sediment yields. Worldwide agricultural intensification and urban area expansion are substantially increasing catchment erosion and quantities of fine sediments in fluvial ecosystems (Wilkes et al., 2019; Zhang et al., 2014). Climate change is also increasing the frequency of high-flow events that remobilize streambed sediments (Buffington & Montgomery, 2021; Kondolf et al., 2016). Stream ecosystems are particularly sensitive to both of these perturbations (Kondolf et al., 2016; Luce et al., 2013; Ormerod, 2009). Thus, there is a pressing need to improve our understanding on how the moving-resting frequency of sediment grains and migrating-resting patch transitions affect fluvial ecosystems and global biogeochemical cycles (Box 1).

## 7 | CONCLUSION

We here focus on the morphodynamics of fine-grain sediments while identifying their impact on ecological dynamics and biogeochemical processes at the micro to meso-habitat, reach and segment scale. Despite the enormous geophysical, ecological and biogeochemical implications of morphodynamics, a conceptual framework for sediment movingresting frequency has not yet been developed. The presented concept is an extension of existing ecological theory, that is, the disturbance theory focusing on flood events, toward more frequent movement of fine-grain sediments. Transport-resting cycles affecting benthic and hyporheic communities are nested across spatio-temporal scales: (1) At the micro-habitat scale, sediment grains are moving and resting at different frequencies. We propose that the most severe effect on benthic and hyporheic communities will occur when ecological and biogeochemical processes are at the same temporal scale as the migration frequency of the sediment. In addition, ecological mechanisms introduce non-linearity into the relationship between migration frequency and ecological or biogeochemical processes. (2) At reach to segment scale, migrating and resting patches occur simultaneously and their transition is driven by larger-scale fluvial constraints such as discharge variability, water depth and sediment load. We provide evidence that benthic and hyporheic communities are not only adapted to the most-prevalent environmental conditions of their local habitat, but also to the larger scale frequency of migrating-resting patch transitions. Thus, migrating-resting patch transition adds new dimensions to the understanding of biogeochemical cycling at larger spatial scales. The nestedness of spatiotemporal scales highlights the need for cross-scale observations to improve our understanding of fine-grain fluvial ecosystems and global biogeochemical cycles.

## **AUTHOR CONTRIBUTIONS**

**Ute Risse-Buhl:** Conceptualization (lead); funding acquisition (equal); project administration (lead); visualization (lead); writing – original draft (lead). **Shai Arnon:** Conceptualization (lead); funding acquisition (equal); visualization (supporting); writing – original draft (lead). **Edo Bar-Zeev:** Conceptualization (lead); funding acquisition (equal); visualization (supporting); writing – original draft (lead). **Anna Oprei:** Conceptualization (supporting); visualization (supporting); writing – original draft (supporting). **Aaron Packman:** Conceptualization (supporting); funding acquisition (equal); visualization (supporting); writing – original draft (supporting). **Ignacio Peralta-Maraver:** Conceptualization (equal); funding acquisition (equal); visualization (lead); writing – original draft (equal). **Anne Robertson:** Conceptualization (equal); funding acquisition (equal); visualization (supporting); writing – original draft (supporting); writing – original draft (equal). **Anne Robertson:** Conceptualization (equal); funding acquisition (equal); visualization (supporting); writing – original draft (supporting); writing – original draft (equal). **Anne Robertson:** Conceptualization (equal); funding acquisition (equal); visualization (supporting); writing – original draft (equal). **Yoni Teitelbaum:** Conceptualization (supporting); visualization (supporting); writing – original draft (lead). **Michael Mutz:** Conceptualization (lead); funding acquisition (equal); visualization (supporting); writing – original draft (lead).

#### ACKNOWLEDGMENT

Open Access funding enabled and organized by Projekt DEAL.

## FUNDING INFORMATION

This project was supported by grants to Ute Risse-Buhl and Michael Mutz (German Research Foundation joint funding grant RI 2093/2-1 and MU 1464/7-1), Ute Risse-Buhl (Carl Zeiss Foundation, P2021-00-004), Shai Arnon (Israel Science Foundation, grant 682/17), Aaron I. Packman and Shai Arnon (NSF-BSF joint funding grant EAR-1734300), Anne Robertson, Ignacio Peralta-Maraver, Shai Arnon and Edo Bar-Zeev (UK-Israel Science Fellowship Scheme 2018–2019) and Edo Bar-Zeev (Israeli Science Foundation, grant 944\21).

## **CONFLICT OF INTEREST**

The authors have declared no conflicts of interest for this article.

#### DATA AVAILABILITY STATEMENT

Data sharing is not applicable to this article as no new data were created or analyzed in this study.

#### ORCID

Ute Risse-Buhl <sup>®</sup> https://orcid.org/0000-0002-7219-5172 Shai Arnon <sup>®</sup> https://orcid.org/0000-0002-7109-8979 Edo Bar-Zeev <sup>®</sup> https://orcid.org/0000-0001-6575-2068 Anna Oprei <sup>®</sup> https://orcid.org/0000-0001-5529-6644 Aaron I. Packman <sup>®</sup> https://orcid.org/0000-0003-3172-4549 Ignacio Peralta-Maraver <sup>®</sup> https://orcid.org/0000-0002-3467-7664 Anne Robertson <sup>®</sup> https://orcid.org/0000-0001-8398-3556 Yoni Teitelbaum <sup>®</sup> https://orcid.org/0000-0001-5411-2954

## **RELATED WIRES ARTICLES**

Physical and biological controls on fine sediment transport and storage in rivers

#### REFERENCES

- Ahmerkamp, S., Marchant, H. K., Peng, C., Probandt, D., Littmann, S., Kuypers, M. M. M., & Holtappels, M. (2020). The effect of sediment grain properties and porewater flow on microbial abundance and respiration in permeable sediments. *Scientific Reports*, 10(1), 3573. https://doi.org/10.1038/s41598-020-60557-7
- Ahmerkamp, S., Winter, C., Kramer, K., de Beer, D., Janssen, F., Friedrich, J., Kuypers, M. M. M., & Holtappels, M. (2017). Regulation of benthic oxygen fluxes in permeable sediments of the coastal ocean. *Limnology and Oceanography*, 62(5), 1935–1954. https://doi.org/10. 1002/lno.10544
- Albertson, L. K., Sklar, L. S., Cooper, S. D., & Cardinale, B. J. (2019). Aquatic macroinvertebrates stabilize gravel bed sediment: A test using silk net-spinning caddisflies in semi-natural river channels. *PLoS One*, 14(1), e0209087. https://doi.org/10.1371/journal.pone.0209087
- Allen, G. H., & Pavelsky, T. M. (2018). Global extent of rivers and streams. Science, 361(6402), 585–587. https://doi.org/10.1126/science. aat0636
- Andreotti, B., Claudin, P., Devauchelle, O., Duran, O., & Fourriere, A. (2012). Bedforms in a turbulent stream: Ripples, chevrons and antidunes. Journal of Fluid Mechanics, 690, 94–128. https://doi.org/10.1017/jfm.2011.386
- Arndt, H., Schmidt-Denter, K., Auer, B., & Weitere, M. (2003). Protozoans and biofilms. In W. E. Krumbein, D. M. Paterson, & G. A. Zavarzin (Eds.), Fossil and recent biofilms (pp. 161–179). Kluwer Academic Publisher.
- Arnon, S., Gray, K. A., & Packman, A. I. (2007). Biophysicochemical process coupling controls nitrate use by benthic biofilms. *Limnology and Oceanography*, 52(4), 1665–1671. https://doi.org/10.4319/lo.2007.52.4.1665
- Arnon, S., Marx, L. P., Searcy, K. E., & Packman, A. I. (2010). Effects of overlying velocity, particle size, and biofilm growth on streamsubsurface exchange of particles. *Hydrological Processes*, 24(1), 108–114. https://doi.org/10.1002/hyp.7490
- Atkinson, B. L., Grace, M. R., Hart, B. T., & Vanderkruk, K. E. N. (2008). Sediment instability affects the rate and location of primary production and respiration in a sand-bed stream. Journal of the North American Benthological Society, 27(3), 581–592. https://doi.org/10.1899/07-143.1
- Aufrecht, J. A., Fowlkes, J. D., Bible, A. N., Morrell-Falvey, J., Doktycz, M. J., & Retterer, S. T. (2019). Pore-scale hydrodynamics influence the spatial evolution of bacterial biofilms in a microfluidic porous network. PLoS One, 14(6), e0218316. https://doi.org/10.1371/journal.pone.0218316
- Augspurger, C., Gleixner, G., Kramer, C., & Küsel, K. (2008). Tracking carbon flow in a 2-week-old and 6-week-old stream biofilm food web. Limnology and Oceanography, 53(2), 642–650. https://doi.org/10.4319/lo.2008.53.2.0642
- Bar-Zeev, E., Berman-Frank, I., Girshevitz, O., & Berman, T. (2012). Revised paradigm of aquatic biofilm formation facilitated by microgel transparent exopolymer particles. Proceedings of the National Academy of Sciences of the United States of America, 109(23), 9119–9124. https://doi.org/10.1073/pnas.1203708109
- Battin, J. T., Kaplan, L. A., Newbold, J. D., & Hansen, C. M. E. (2003). Contributions of microbial biofilms to ecosystem processes in stream mesocosms. *Nature*, 426, 439–442. https://doi.org/10.1038/nature02152
- Battin, T. J., Besemer, K., Bengtsson, M. M., Roman, A. M., & Packmann, A. I. (2016). The ecology and biogeochemistry of stream biofilms. *Nature Reviews Microbiology*, 14(4), 251–263. https://doi.org/10.1038/nrmicro.2016.15
- Bo, T., Doretto, A., Levrino, M., & Fenoglio, S. (2020). Contribution of beta diversity in shaping stream macroinvertebrate communities among hydro-ecoregions. Aquatic Ecology, 54(4), 957–971. https://doi.org/10.1007/s10452-020-09786-6
- Boenigk, J., & Arndt, H. (2002). Bacterivory by heterotrophic flagellates: Community structure and feeding strategies. Antonie van Leeuwenhoek International Journal of General and Molecular Microbiology, 81(1–4), 465–480. https://doi.org/10.1023/a:1020509305868
- Booker, D. J. & Dunbar, M. J. (2008). Predicting river width, depth and velocity at ungauged sites in England and Wales using multilevel models. *Hydrological Processes*, 22(20), 4049–4057. https://doi.org/10.1002/hyp.7007

WIRES WILEY 9 of 13

Bridge, J. S. (2003). Rivers and floodplains. Forms, processes, and sedimentary record (1st ed.). Blackwell Science.

- Buffington, J. M., & Montgomery, D. R. (2021). Geomorphic classification of rivers: An updated review. *Treatise on Geomorphology*, 2, 1–47. Caruso, A., Boano, F., Ridolfi, L., Chopp, D. L., & Packman, A. (2017). Biofilm-induced bioclogging produces sharp interfaces in hyporheic flow, redox conditions, and microbial community structure. *Geophysical Research Letters*, 44(10), 4917–4925. https://doi.org/10.1002/2017gl073651
- Chew, S. C., Kundukad, B., Seviour, T., van der Maarel, J. R. C., Yang, L., Rice, S. A., Doyle, P., & Kjelleberg, S. (2014). Dynamic remodeling of microbial biofilms by functionally distinct exopolysaccharides. *Mbio*, 5(4), e01536-14. https://doi.org/10.1128/mBio.01536-14
- Cole, J. J., Prairie, Y. T., Caraco, N. F., McDowell, W. H., Tranvik, L. J., Striegl, R. G., Duarte, C. M., Kortelainen, P., Downing, J. A., Middelburg, J. J., & Melack, J. (2007). Plumbing the global carbon cycle: Integrating inland waters into the terrestrial carbon budget. *Ecosystems*, 10(1), 171–184. https://doi.org/10.1007/s10021-006-9013-8
- Connell, J. H. (1978). Diversity in tropical rain forests and coral reefs—High diversity of trees and corals is maintained only in a non-equilibrium state. *Science*, 199(4335), 1302–1310. https://doi.org/10.1126/science.199.4335.1302
- Cook, S., Price, O., King, A., Finnegan, C., van Egmond, R., Schfer, H., Pearson, J. M., Abolfathi, S., & Bending, G. D. (2020). Bedform characteristics and biofilm community development interact to modify hyporheic exchange. *Science of the Total Environment*, 749, 141397. https://doi.org/10.1016/j.scitotenv.2020.141397
- Dallmann, J., Phillips, C. B., Teitelbaum, Y., Saavedra Cifuentes, E. Y., Sund, N., Schumer, R., Arnon, S., & Packman, A. I. (2021). Bedform segregation and locking increase storage of natural and synthetic particles in rivers. *Nature Communications*, 12(1), 1–7.
- Dallmann, J., Phillips, C. B., Teitelbaum, Y., Sund, N., Schumer, R., Arnon, S., & Packman, A. I. (2020). Impacts of suspended clay particle deposition on sand-bed morphodynamics. *Water Resources Research*, 56(8), e2019WR027010. https://doi.org/10.1029/2019WR027010
- David, S. R., Edmonds, D. A., & Letsinger, S. L. (2017). Controls on the occurrence and prevalence of floodplain channels in meandering rivers. Earth Surface Processes and Landforms, 42(3), 460–472. https://doi.org/10.1002/esp.4002
- De Falco, N., Boano, F., Bogler, A., Bar-Zeev, E., & Arnon, S. (2018). Influence of stream–subsurface exchange flux and bacterial biofilms on oxygen consumption under nutrient-rich conditions. *Journal of Geophysical Research—Biogeosciences*, 123(7), 2021–2034. https://doi.org/ 10.1029/2017jg004372
- de la Fuente, M., Bonada, N., Bêche, L., Dahm, C. N., Mendez, P. K., Tockner, K., Uehlinger, U., & Acuña, V. (2018). Evolutionary responses of aquatic macroinvertebrates to two contrasting flow regimes. *Hydrobiologia*, 808(1), 353–370. https://doi.org/10.1007/s10750-017-3437-3
- de Jonge, V. N. (1985). The occurrence of epipsammic diatom populations—A result of interaction between physical sorting of sediment and certain properties of diatom species. Estuarine Coastal and Shelf Science, 21(5), 607–622. https://doi.org/10.1016/0272-7714(85)90061-7
- Delgado, M., de Jonge, V. N., & Peletier, H. (1991). Effect of sand movement on the growth of benthic diatoms. *Journal of Experimental Marine Biology and Ecology*, 145, 221–231. https://doi.org/10.1016/0022-0981(91)90177-X
- Dietrich, W. E., & Perron, J. T. (2006). The search for a topographic signature of life. *Nature*, 439(7075), 411–418. https://doi.org/10.1038/ nature04452
- Dole Olivier, M. J., Marmonier, P., & Beffy, J. L. (1997). Response of invertebrates to lotic disturbance: Is the hyporheic zone a patchy refugium? Freshwater Biology, 37(2), 257–276. https://doi.org/10.1046/j.1365-2427.1997.00140.x
- Drake, T. W., Raymond, P. A., & Spencer, R. G. M. (2018). Terrestrial carbon inputs to inland waters: A current synthesis of estimates and uncertainty. *Limnology and Oceanography Letters*, 3(3), 132–142. https://doi.org/10.1002/lol2.10055
- Elliott, A. H., & Brooks, N. H. (1997). Transfer of nonsorbing solutes to a streambed with bed forms: Theory. *Water Resources Research*, 33(1), 123–136. https://doi.org/10.1029/96wr02784
- Fegley, S. R. (1987). Experimental variation of near-bottom current speeds and its effects on depth distribution of sand-living meiofauna. *Marine Biology (Berlin)*, 95(2), 183–192.
- Frings, R. M. (2008). Downstream fining in large sand-bed rivers. Earth-Science Reviews, 87(1-2), 39-60.
- Garcia, M. (2008). Sedimentation engineering: Processes, measurements, modeling, and practice. American Society of Civil Engineers.
- Gaudes, A., Artigas, J., & Muñoz, I. (2010). Species traits and resilience of meiofauna to floods and drought in a Mediterranean stream. Marine and Freshwater Research, 61(11), 1336–1347. https://doi.org/10.1071/Mf10044
- Gerbersdorf, S. U., Koca, K., de Beer, D., Chennu, A., Noss, C., Risse-Buhl, U., Weitere, M., Eiff, O., Wagner, M., Aberle, J., Schweikert, M., & Terheiden, K. (2020). Exploring flow-biofilm-sediment interactions: Assessment of current status and future challenges. *Water Research*, 185, 116182. https://doi.org/10.1016/j.watres.2020.116182
- Harvey, J. W., Drummond, J. D., Martin, R. L., McPhillips, L. E., Packman, A. I., Jerolmack, D. J., Stonedahl, S. H., Aubeneau, A. F., Sawyer, A. H., Larsen, L. G., & Tobias, C. R. (2012). Hydrogeomorphology of the hyporheic zone: Stream solute and fine particle interactions with a dynamic streambed. *Journal of Geophysical Research-Biogeosciences*, 117, G00n11. https://doi.org/10.1029/2012jg002043
- Hatton, I. A., Dobson, A. P., Storch, D., Galbraith, E. D., & Loreau, M. (2019). Linking scaling laws across eukaryotes. Proceedings of the National Academy of Sciences of the United States of America, 116(43), 21616–21622. https://doi.org/10.1073/pnas.1900492116
- Heffernan, J. B. (2008). Wetlands as an alternative stable state in desert streams. Ecology, 89(5), 1261–1271. https://doi.org/10.1890/07-0915.1
- Hou, J. P., Veeregowda, D. H., van de Belt-Gritter, B., Busscher, H. J., & van der Mei, H. C. (2018). Extracellular polymeric matrix production and relaxation under fluid shear and mechanical pressure in *Staphylococcus aureus* biofilms. *Applied and Environmental Microbiology*, 84(1), e01516-17. https://doi.org/10.1128/AEM.01516-17
- Izagirre, O., Serra, A., Guasch, H., & Elosegi, A. (2009). Effects of sediment deposition on periphytic biomass, photosynthetic activity and algal community structure. Science of the Total Environment, 407(21), 5694–5700. https://doi.org/10.1016/j.scitotenv.2009.06.049
- James, A. B. W., Dewson, Z. S., & Death, R. G. (2009). The influence of flow reduction on macroinvertebrate drift density and distance in three New Zealand streams. Journal of the North American Benthological Society, 28(1), 220–232. https://doi.org/10.1899/07-146.1

- Jewson, D. H., Lowry, S. F., & Bowen, R. (2006). Co-existence and survival of diatoms on sand grains. European Journal of Phycology, 41(2), 131–146. https://doi.org/10.1080/09670260600652903
- Kim, M. K., Ingremeau, F., Zhao, A. S., Bassler, B. L., & Stone, H. A. (2016). Local and global consequences of flow on bacterial quorum sensing. *Nature Microbiology*, 1(1), 15005. https://doi.org/10.1038/Nmicrobiol.2015.5
- Kondolf, G. M., Piégay, H., Schmitt, L., & Montgomery, D. R. (2016). Geomorphic classification of rivers and stream. In G. M. Kondolf, & H. Piégay (Eds.), *Tools in fluvial geomorphology* (pp. 133–158). John Wiley & Sons, Ltd.
- Krause, S., Klaar, M. J., Hannah, D. M., Mant, J., Bridgeman, J., Trimmer, M., & Manning-Jones, S. (2014). The potential of large woody debris to alter biogeochemical processes and ecosystem services in lowland rivers. WIREs Water, 1, 263–275. https://doi.org/10.1002/wat2.1019
- Krause, S., Lewandowski, J., Grimm, N. B., Hannah, D. M., Pinay, G., McDonald, K., Mart, E., Argerich, A., Pfister, L., Klaus, J., Battin, T., Larned, S. T., Schelker, J., Fleckenstein, J., Schmidt, C., Rivett, M. O., Watts, G., Sabater, F., Sorolla, A., & Turk, V. (2017). Ecohydrological interfaces as hot spots of ecosystem processes. *Water Resources Research*, 53(8), 6359–6376. https://doi.org/10.1002/2016wr019516
- Krejci, M. E., & Lowe, R. L. (1986). Importance of sand grain mineralogy and topography in determining micro-spatial distribution of epipsammic diatoms. *Journal of the North American Benthological Society*, 5(3), 211–220. https://doi.org/10.2307/1467708
- Kryvokhyzhyna, M., Majdi, N., Oprei, A., Mutz, M., & Risse-Buhl, U. (2022). Response of meiobenthos to migrating ripples in sandy lowland streams. *Hydrobiologia*, 849, 1905–1921. https://doi.org/10.1007/s10750-022-04836-6
- Lake, P. S. (2000). Disturbance, patchiness, and diversity in streams. Journal of the North American Benthological Society, 19(4), 573–592.
- Leibold, M. A., Holyoak, M., Mouquet, N., Amarasekare, P., Chase, J. M., Hoopes, M. F., Holt, R. D., Shurin, J. B., Law, R., Tilman, D., Loreau, M., & Gonzalez, A. (2004). The metacommunity concept: A framework for multi-scale community ecology. *Ecology Letters*, 7(7), 601–613. https://doi.org/10.1111/j.1461-0248.2004.00608.x
- Levin, L. A., & DiBacco, C. (1995). Influence of sediment transport on short-term recolonization by seamount infauna. Marine Ecology Progress Series, 123(1), 163–175. https://doi.org/10.3354/MEPS123163
- Lewin, J., Ashworth, P. J., & Strick, R. J. P. (2017). Spillage sedimentation on large river floodplains. Earth Surface Processes and Landforms, 42(2), 290–305. https://doi.org/10.1002/esp.3996
- Lock, M. A., Wallace, R. R., Costerton, J. W., Ventullo, R. M., & Charlton, S. E. (1984). River epilithon: Toward a structural-functional model. *Oikos*, 42, 10–22. https://doi.org/10.2307/3544604
- Luce, J. J., Lapointe, M. F., Roy, A. G., & Ketterling, D. B. (2013). The effects of sand abrasion of a predominantly stable stream bed on periphyton biomass losses. *Ecohydrology*, *6*(4), 689–699.
- Luce, J. J., Steele, R., & Lapointe, M. F. (2010). A physically based statistical model of sand abrasion effects on periphyton biomass. *Ecological Modelling*, 221(2), 353–361.
- Majdi, N., Boiche, A., Traunspurger, W., & Lecerf, A. (2014). Predator effects on a detritus-based food web are primarily mediated by nontrophic interactions. Journal of Animal Ecology, 83(4), 953–962. https://doi.org/10.1111/1365-2656.12189
- Majdi, N., & Traunspurger, W. (2015). Free-living nematodes in the freshwater food web: A review. Journal of Nematology, 47(1), 28-44.
- Malarkey, J., Baas, J. H., Hope, J. A., Aspden, R. J., Parsons, D. R., Peakall, J., Paterson, D. M., Schindler, R. J., Ye, L. P., Lichtman, I. D., Bass, S. J., Davies, A. G., Manning, A. J., & Thorne, P. D. (2015). The pervasive role of biological cohesion in bedform development. *Nature Communications*, 6, 6257. https://doi.org/10.1038/ncomms7257
- Mendoza-Lera, C., & Mutz, M. (2013). Microbial activity and sediment disturbance modulate the vertical water flux in sandy sediments. *Freshwater Science*, 32(1), 26–38. https://doi.org/10.1899/11-165.1
- Miller, A. R., Lowe, R. L., & Rotenberry, J. T. (1987). Succession of diatom communities on sand grains. Journal of Ecology, 75(3), 693–709. https://doi.org/10.2307/2260200
- Miller, D. C. (1989). Abrasion effects on microbes in sandy sediments. Marine Ecology Progress Series, 55(1), 73–82. https://doi.org/10.3354/ meps055073
- Montgomery, D. R., & Buffington, J. M. (1997). Channel-reach morphology in mountain drainage basins. Geological Society of America Bulletin, 109(5), 596–611. https://doi.org/10.1130/0016-7606(1997)109<0596:Crmimd>2.3.Co;2
- Naqshband, S., Hoitink, A. J. F., McElroy, B., Hurther, D., & Hulscher, S. J. M. H. (2017). A sharp view on river dune transition to upper stage plane bed. *Geophysical Research Letters*, 44(22), 11437–11444. https://doi.org/10.1002/2017gl075906
- Nielsen, S. L., & Sand-Jensen, K. (1991). Variation in growth rates of submerged rooted macrophytes. Aquatic Botany, 39(1–2), 109–120. https://doi.org/10.1016/0304-3770(91)90026-2
- Nikora, V., Habersack, H., Huber, T., & McEwan, I. (2002). On bed particle diffusion in gravel bed flows under weak bed load transport. Water Resources Research, 38(6), 1081. https://doi.org/10.1029/2001wr000513
- Norf, H., Arndt, H., & Weitere, M. (2009). Responses of biofilm-dwelling ciliate communities to planktonic and benthic resource enrichment. *Microbial Ecology*, 57(4), 687–700. https://doi.org/10.1007/s00248-008-9470-z
- Olsen, D. A., & Townsend, C. R. (2005). Flood effects on invertebrates, sediments and particulate organic matter in the hyporheic zone of a gravel-bed stream. *Freshwater Biology*, *50*(5), 839–853. https://doi.org/10.1111/j.1365-2427.2005.01365.x
- Oprei, A., Zlatanović, S., & Mutz, M. (2019). Grazers superimpose humidity effect on stream biofilm resistance and resilience to dry-rewet stress. Science of the Total Environment, 659, 841–850. https://doi.org/10.1016/j.scitotenv.2018.12.316
- Ormerod, S. J. (2009). Climate change, river conservation and the adaptation challenge. *Aquatic Conservation: Marine and Freshwater Ecosystems*, 19(6), 609–613.
- Ortmann-Ajkai, A., Loczy, D., Gyenizse, P., & Pirkhoffer, E. (2014). Wetland habitat patches as ecological components of landscape memory in a highly modified floodplain. *River Research and Applications*, *30*(7), 874–886. https://doi.org/10.1002/rra.2685

## 12 of 13 WILEY WIRES

- Packman, A. I., Brooks, N. H., & Morgan, J. J. (2000). A physicochemical model for colloid exchange between a stream and a sand streambed with bed forms. Water Resources Research, 36(8), 2351–2361. https://doi.org/10.1029/2000wr900059
- Palmer, M. A. (1992). Incorporating lotic meiofauna into our understanding of faunal transport processes. *Limnology and Oceanography*, *37*, 329–341. https://doi.org/10.4319/lo.1992.37.2.0329
- Paterson, D. M., Hope, J. A., Kenworthy, J., Biles, C. L., & Gerbersdorf, S. U. (2018). Form, function and physics: The ecology of biogenic stabilisation. *Journal of Soils and Sediments*, 18(10), 3044–3054. https://doi.org/10.1007/s11368-018-2005-4
- Peralta-Maraver, I., Robertson, A. L., & Perkins, D. M. (2019). Depth and vertical hydrodynamics constrain the size structure of a lowland streambed community. *Biology Letters*, 15(7), 20190317. https://doi.org/10.1098/rsbl.2019.0317
- Peralta-Maraver, I., Stubbington, R., Arnon, S., Kratina, P., Krause, S., de Mello Cionek, V., Leite, N. K., Lemes da Silva, A. L., Thomaz, S. M., Posselt, M., Milner, V. S., Momblanch, A., Moretti, M. S., Nóbrega, R. L. B., Perkins, D. M., Petrucio, M. M., Reche, I., Saito, V., Sarmento, H., ... Robertson, A. L. (2021). The riverine bioreactor: An integrative perspective on biological decomposition of organic matter across riverine habitats. *Science of the Total Environment*, 772, 145494. https://doi.org/10.1016/j.scitotenv.2021.145494
- Perucca, E., Camporeale, C., & Ridolfi, L. (2007). Significance of the riparian vegetation dynamics on meandering river morphodynamics. Water Resources Research, 43(3), W03430. https://doi.org/10.1029/2006wr005234
- Phillips, C. B., Dallmann, J. D., Jerolmack, D. J., & Packman, A. I. (2019). Fine-particle deposition, retention, and resuspension within a sand-bedded stream are determined by streambed morphodynamics. Water Resources Research, 55(12), 10303–10318. https://doi.org/10.1029/2019wr025272
- Phillips, C. B., & Jerolmack, D. J. (2014). Dynamics and mechanics of bed-load tracer particles. Earth Surface Dynamics, 2(2), 513–530. https://doi.org/10.5194/esurf-2-513-2014
- Pilditch, C. A., & Miller, D. C. (2006). Phytoplankton deposition to permeable sediments under oscillatory flow: Effects of ripple geometry and resuspension. *Continental Shelf Research*, 26(15), 1806–1825. https://doi.org/10.1016/j.csr.2006.06.002
- Poff, N. L. (1997). Landscape filters and species traits: Towards mechanistic understanding and prediction in stream ecology. Journal of the North American Benthological Society, 16(2), 391–409. https://doi.org/10.2307/1468026
- Probandt, D., Eickhorst, T., Ellrott, A., Amann, R., & Knittel, K. (2018). Microbial life on a sand grain: From bulk sediment to single grains. ISME Journal, 12(2), 623–633. https://doi.org/10.1038/ismej.2017.197
- Raudkivi, A. J. (2006). Transition from ripples to dunes. Journal of Hydraulic Engineering—ASCE, 132(12), 1316–1320. https://doi.org/10. 1061/(Asce)0733-9429(2006)132:12(1316)
- Regnier, P., Friedlingstein, P., Ciais, P., Mackenzie, F. T., Gruber, N., Janssens, I. A., Laruelle, G. G., Lauerwald, R., Luyssaert, S., Andersson, A. J., Arndt, S., Arnosti, C., Borges, A. V., Dale, A. W., Gallego-Sala, A., Godderis, Y., Goossens, N., Hartmann, J., Heinze, C., ... Thullner, M. (2013). Anthropogenic perturbation of the carbon fluxes from land to ocean. *Nature Geoscience*, 6(8), 597– 607. https://doi.org/10.1038/Ngeo1830
- Remis, J. P., Costerton, J. W., & Auer, M. (2010). Biofilms: Structures that may facilitate cell-cell interactions. ISME Journal, 4(9), 1085– 1087. https://doi.org/10.1038/ismej.2010.105
- Resh, V. H., Brown, A. V., Covich, A. P., Gurtz, M. E., Li, H. W., Minshall, G. W., Reice, S. R., Sheldon, A. L., Wallace, J. B., & Wissmar, R. C. (1988). The role of disturbance in stream ecology. *Journal of the North American Benthological Society*, 7(4), 433–455.
- Riis, T., & Sand-Jensen, K. (2006). Dispersal of plant fragments in small streams. Freshwater Biology, 51(2), 274–286. https://doi.org/10.1111/ j.1365-2427.2005.01496.x
- Rinaldi, M., Gurnell, A. M., del Tanago, M. G., Bussettini, M., & Hendriks, D. (2016). Classification of river morphology and hydrology to support management and restoration. *Aquatic Sciences*, 78(1), 17–33. https://doi.org/10.1007/s00027-015-0438-z
- Risse-Buhl, U., Felsmann, K., & Mutz, M. (2014). Colonization dynamics of ciliate morphotypes modified by shifting sandy sediments. European Journal of Protistology, 50, 345–355. https://doi.org/10.1016/j.ejop.2014.03.006
- Risse-Buhl, U., Schlief, J., & Mutz, M. (2015). Phagotrophic protists are a key component of microbial communities processing leaf litter under contrasting oxic conditions. *Freshwater Biology*, 60(11), 2310–2322. https://doi.org/10.1111/fwb.12657
- Risse-Buhl, U., Trefzger, N. M., Seifert, A.-G., Schönborn, W., Gleixner, G., & Küsel, K. (2012). Tracking the authochthonous carbon flow in stream biofilm food webs. *FEMS Microbiology Ecology*, 79, 118–131. https://doi.org/10.1111/j.1574-6941.2011.01202.x
- Rivera-Yoshida, N., Del Angel, J. A. A., & Benitez, M. (2018). Microbial multicellular development: Mechanical forces in action. Current Opinion in Genetics & Development, 51, 37–45. https://doi.org/10.1016/j.gde.2018.05.006
- Robertson, A. L. (2000). Lotic meiofaunal community dynamics: Colonisation, resilience and persistence in a spatially and temporally heterogeneous environment. *Freshwater Biology*, 44(1), 135–147. https://doi.org/10.1046/j.1365-2761.2000.00595.x
- Robertson, A. L., Lancaster, J., & Hildrew, A. G. (1995). Stream hydraulics and the distribution of microcrustacea: A role for refugia? Freshwater Biology, 33(3), 469–484. https://doi.org/10.1111/j.1365-2427.1995.tb00407.x
- Robertson, A. L., & Wood, P. J. (2010). Ecology of the hyporheic zone: Origins, current knowledge and future directions. Fundamental and Applied Limnology, 176(4), 279–289. https://doi.org/10.1127/1863-9135/2010/0176-0279
- Romaní, A. M., Fund, K., Artigas, J., Schwartz, T., Sabater, S., & Obst, U. (2008). Relevance of polymeric matrix enzymes during biofilm formation. *Microbial Ecology*, 56(3), 427–436. https://doi.org/10.1007/s00248-007-9361-8
- Romaní, A. M., Guasch, H., Muñoz, I., Ruana, J., Vilalta, E., Schwartz, T., Emtiazi, F., & Sabater, S. (2004). Biofilm structure and function and possible implications for riverine DOC dynamics. *Microbial Ecology*, 47(4), 316–328. https://doi.org/10.1007/s00248-003-2019-2
- Ryo, M., Aguilar-Trigueros, C. A., Pinek, L., Muller, L. A. H., & Rillig, M. C. (2019). Basic principles of temporal dynamics. Trends in Ecology & Evolution, 34(8), 723–733. https://doi.org/10.1016/j.tree.2019.03.007
- Scheidweiler, D., Mendoza Lera, C., Mutz, M., & Risse-Buhl, U. (2021). Overlooked implication of sediment transport at low flow: Migrating ripples modulate streambed photo- and heterotrophic microbial activity. *Water Resources Research*, 57(3), e2020WR027988. https://doi. org/10.1029/2020WR027988

- Schratzberger, M., & Larcombe, P. (2014). The role of the sedimentary regime in shaping the distribution of subtidal sandbank environments and the associated meiofaunal nematode communities: An example from the southern North Sea. *PLoS One*, *9*, e109445. https://doi.org/ 10.1371/journal.pone.0109445
- Segura, C., McCutchan, J. H., Lewis, W. M., Jr., & Pitlick, J. (2011). The influence of channel bed disturbance on algal biomass in a Colorado mountain stream. *Ecohydrology*, *4*, 411–421. https://doi.org/10.1002/eco.142
- Shimeta, J., Amos, C. L., Beaulieu, S. E., & Katz, S. L. (2003). Resuspension of benthic protists at subtidal coastal sites with differing sediment composition. *Marine Ecology-Progress Series*, 259, 103–115. https://doi.org/10.3354/meps259103
- Sinsabaugh, R. L., Shah, J. J. F., Findlay, S. G., Kuehn, K. A., & Moorhead, D. L. (2015). Scaling microbial biomass, metabolism and resource supply. *Biogeochemistry*, 122(2–3), 175–190. https://doi.org/10.1007/s10533-014-0058-z
- Soininen, J., & Teittinen, A. (2019). Fifteen important questions in the spatial ecology of diatoms. *Freshwater Biology*, 64(11), 2071–2083. https://doi.org/10.1111/fwb.13384
- Stoodley, P., Cargo, R., Rupp, C. J., & Klapper, I. (2002). Biofilm material properties as related to shear-induced deformation and detachment phenomena. *Journal of Industrial Microbiology & Biotechnology*, *29*, 361–367. https://doi.org/10.1038/sj.jim.7000282
- Swan, C. M., & Palmer, M. A. (2000). What drives small-scale spatial patterns in lotic meiofauna communities? *Freshwater Biology*, 44(1), 109–121. https://doi.org/10.1046/j.1365-2427.2000.00587.x
- Teitelbaum, Y., Dallmann, J., Phillips, C. B., Packman, A. I., Schumer, R., Sund, N. L., Hansen, S. K., & Arnon, S. (2021). Dynamics of hyporheic exchange flux and fine particle deposition under moving bedforms. *Water Resources Research*, 57(4), e2020WR028541. https://doi. org/10.1029/2020WR028541
- Teitelbaum, Y., Shimony, T., Cifuentes, E. S., Dallmann, J., Phillips, C. B., Packman, A. I., Hansen, S. K., & Arnon, S. (2022). A novel framework for simulating particle deposition with moving bedforms. *Geophysical Research Letters*, 49, e2021GL097223. https://doi.org/10.1029/2021gl097223
- Tockner, K., Ward, J. V., Arscott, D. B., Edwards, P. J., Kollmann, J., Gurnell, A. M., Petts, G. E., & Maiolini, B. (2003). The Tagliamento River: A model ecosystem of European importance. *Aquatic Sciences*, 65(3), 239–253. https://doi.org/10.1007/s00027-003-0699-9
- Tolonen, K. E., Leinonen, K., Erkinaro, J., & Heino, J. (2018). Ecological uniqueness of macroinvertebrate communities in high-latitude streams is a consequence of deterministic environmental filtering processes. *Aquatic Ecology*, *52*(1), 17–33. https://doi.org/10.1007/s10452-017-9642-3
- Utz, R. M., & Hilderbrand, R. H. (2011). Interregional variation in urbanization-induced geomorphic change and macroinvertebrate habitat colonization in headwater streams. *Journal of the North American Benthological Society*, 30(1), 25–37. https://doi.org/10.1899/10-007.1
- Venditti, J. G., Nelson, P. A., Bradley, R. W., Haught, D., & Gitto, A. B. (2017). Bedforms, structures, patches, and sediment supply in gravelbed rivers. In D. Tsutsumi & J. B. Laronne (Eds.), *Gravel-bed rivers: Processes and disasters* (pp. 439–466). John Wiley & Sons.
- Verzano, K., Bärlund, I., Flörke, M., Lehner, B., Kynast, E., Voss, F. & Alcamo, J. (2012). Modeling variable river flow velocity on continental scale: Current situation and climate change impacts in Europe. *Journal of Hydrology*, 424, 238–251. https://doi.org/10.1016/j.jhydrol. 2012.01.005
- Weise, W., & Rheinheimer, G. (1978). Scanning electron microscopy and epifluorescence investigation of bacterial colonization of marine sand sediments. *Microbial Ecology*, 4(3), 175–188. https://doi.org/10.1007/Bf02015075
- Weitere, M., Erken, M., Majdi, N., Arndt, H., Norf, H., Reinshagen, M., Traunspurger, W., Walterscheid, A., & Wey, J. K. (2018). The food web perspective on aquatic biofilms. *Ecological Monographs*, 88(4), 543–559. https://doi.org/10.1002/ecm.1315
- Wilkes, M. A., Gittins, J. R., Mathers, K. L., Mason, R., Casas-Mulet, R., Vanzo, D., Mckenzie, M., Murray-Bligh, J., England, J., Gurnell, A., & Jones, J. I. (2019). Physical and biological controls on fine sediment transport and storage in rivers. WIREs Water, 6(2), e1331. https://doi.org/10.1002/wat2.1331
- Wolke, P., Teitelbaum, Y., Deng, C., Lewandowski, J., & Arnon, S. (2020). Impact of bed form celerity on oxygen dynamics in the hyporheic zone. Water, 12(1), 62. https://doi.org/10.3390/w12010062
- Wood, P. J., Boulton, A. J., Little, S., & Stubbington, R. (2010). Is the hyporheic zone a refugium for aquatic macroinvertebrates during severe low flow conditions? *Fundamental and Applied Limnology*, 176(4), 377–390. https://doi.org/10.1127/1863-9135/2010/0176-0377
- Xuehua, D., Zhaoyin, W., Mengzhen, X., & Zhang, K. (2009). Effect of streambed sediment on benthic ecology. International Journal of Sediment Research, 24(3), 325–338.
- Zhang, Y., Collins, A. L., Murdoch, N., Lee, D., & Naden, P. S. (2014). Cross sector contributions to river pollution in England and Wales: Updating waterbody scale information to support policy delivery for the water framework directive. *Environmental Science & Policy*, 42, 16–32. https://doi.org/10.1016/j.envsci.2014.04.010
- Zhang, Y., Meerschaert, M. M., & Packman, A. I. (2012). Linking fluvial bed sediment transport across scales. Geophysical Research Letters, 39, L20404. https://doi.org/10.1029/2012gl053476
- Zheng, L. Z., Cardenas, M. B., Wang, L. C., & Mohrig, D. (2019). Ripple effects: Bed form morphodynamics cascading into hyporheic zone biogeochemistry. Water Resources Research, 55(8), 7320–7342. https://doi.org/10.1029/2018wr023517
- Zlatanović, S., Fabian, J., Mendoza-Lera, C., Woodward, K. B., Premke, K., & Mutz, M. (2017). Periodic sediment shift in migrating ripples influences benthic microbial activity. *Water Resources Research*, 53, 4741–4755. https://doi.org/10.1002/2017WR020656

**How to cite this article:** Risse-Buhl, U., Arnon, S., Bar-Zeev, E., Oprei, A., Packman, A. I., Peralta-Maraver, I., Robertson, A., Teitelbaum, Y., & Mutz, M. (2023). Streambed migration frequency drives ecology and biogeochemistry across spatial scales. *WIREs Water*, e1632. https://doi.org/10.1002/wat2.1632