Impact of plant growth promoting rhizobacteria (PGPR) on stress resistance of winter wheat (*Triticum aestivum* L.)

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PREFACE

This thesis is written as a cumulative PhD thesis. After a general introduction, Chapter 1-3 are independent papers that were accepted for publication in international peer-reviewed journals. Thereby, the differential formats of the chapters meet the requirements of the respective journals, where the manuscripts have been published. The thesis is finalized by a synpopsis including an integrating discussion over all Chapters and critical review of limitations and derived possible perspectives for future studies. A CV, list of publications and presentations, as well as my contribution to each are provided at the end of this thesis.

BIBLIOGRAPHIC DESCRIPTION

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Impact of plant growth promoting rhizobacteria (PGPR) on stress resistance of winter wheat

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The here presented cumulative thesis provides important insights into the complex interplay of plant-soil-interactions in agricultural systems under future expected drought as a consequence of climate change. Thereby, not only drought-induced impacts, but also the single and interacting effects of different abiotic (soil type and land use management) and biotic (cultivar type and plant development stages) factors on functional and structural changes of wheat rhizosphere bacterial communities, but also individual bacterial species, were evaluated by cultivation independent and dependent approaches, respectively. The experiments were conducted either as a pot experiment in the cold greenhouse or under field conditions using the platform of the Global Change Experimental Facility (GCEF). The GCEF is a long-term experiment investigating the effect of climate change on different land-use managements.

In *Chapter 1* a pot experiment was conducted to study the impact of multiple factors, i.e. soil type, different agricultural systems and weat cultivar, on wheat rhizosphere bacterial communities under extreme drought. The results indicated a strong soil-type specific response of the rhizosphere bacterial communities under drought. The manuscript highlight the importance of studying multiple factors to reveal common or specific adaptation processes in the rhizosphere of wheat under drought.

Chapter 2 aimed to study wheat rhizosphere bacterial communities' adaptation to different climate scenarios, in different agricultural management systems and at two different plant development stages under field conditions using the GCEF. In addition, the study was used to evaluate the accuracy of computational tools to predict functional profiles based on gene abundances by comparing these to real measured activities. We observed an overall strong effect of plant development stage, followed by agricultural management system and climate treatment on community compostion and function, which was also qualitatively confirmed by the computational approach.

Bibliographic description

Chapter 3 is a follow-up study of Chapter 2, using the same experimental system, but focusing on individual bacterial species in the rhizospere of wheat. Specifically screening for phosphate-solubilizing bacteria, we identified three dominant genera in the rhizosphere of wheat, i.e. Streptomyces, Pseudomonas and Phyllobacterium, whereby dominance of Pseudomonas and Phyllobacterium species shifted according to plant development stage. Since Phyllobacterium species have not been reported as effective P-solubilizer in the rhizosphere of wheat before and further indicated a strong drought-tolerance in vitro, we introduced Phyllobacterium species as promising candidate for use as biofertilizers.

ZUSAMMENFASSUNG

Weizen ist eine der weltweit am häufigsten angebauten Kulturpflanzen und trägt zur Sicherung der Nahrungsmittelproduktion in verschiedenen Regionen der Welt bei. Obwohl er fast überall angebaut werden kann, ist die Produktion durch Trockenheit limitiert. Daher rücken mehr und mehr die mikrobiellen Gemeinschaften im Boden und in der Rhizosphäre in den Mittelpunkt der modernen agrarbiologischen Forschung, um die Produktivität bei Trockenheit aufrechtzuerhalten und eine nachhaltige Produktion zu fördern. Während bereits zahlreiche Studien über die Weizenproduktion und den positiven Einfluss des Bodenmikrobioms in ariden und semiariden Regionen der Welt durchgeführt wurden, sind vergleichbare Studien in Mitteleuropa selten. Dies könnte sich aufgrund der anhaltenden Klimakrise und der zu erwartenden ausbleibenden Sommerniederschläge ändern. Dabei haben die meisten Studien, die sich mit der Akklimatisierung des Weizenrhizobioms an Wasserdefizite befasst haben, bestenfalls den Einfluss von Trockenheit und ein oder zwei weiteren biotischen oder abiotischen Einflussfaktoren, die zudem miteinander interagieren können, auf die Struktur und Funktion der mikrobiellen Gemeinschaften in der Rhizosphäre untersucht. Ziel dieser Arbeit war es daher, verschiedene komplementäre Analysemethoden zu kombinieren, um trockenheitsbedingte strukturelle und funktionelle Veränderungen in den bakteriellen Gemeinschaften und auch einzelner Arten in der Weizenrhizosphäre, in Abhängigkeit von Landnutzungssystem, Weizensorte und Pflanzenentwicklungsstadium untersuchen, und zu ermitteln, wie sich diese Veränderungen auf die Produktivität des Weizens als Folge möglicher Szenarien des Klimawandels in Mitteldeutschland auswirken.

Die vorliegende Arbeit leitet mit einer allgemeinen Einführung und Vorstellung des Projekts ein, gefolgt von drei aufeinanderfolgenden Kapiteln, die die wichtigsten Ergebnisse enthalten, die in von Fachleuten begutachteten Artikeln veröffentlicht wurden. Beginnend mit einem Experiment im Gewächshaus (Kapitel 1) und weiterführend zu einem realistischen Klimaszenario unter Feldbedingungen (Kapitel 2 und 3), beschreiben die drei Kapitel die

alleinigen und interagierenden Auswirkungen von Trockenheit und Anbausystem (Kapitel 1-3), Bodentyp und Weizensorte (Kapitel 1), sowie Pflanzenwachstumsstadien (Kapitel 2 und 3) auf Bakteriengemeinschaften und einzelne Taxa des Weizenrhizobioms. Die verwendeten Methoden reichen dabei von der traditionellen Kultivierung und In-vitro-Bioassays (Kapitel 3), über extrazelluläre Enzymaktivitätspotenziale (Kapitel 1 und 2), bis hin zu fortschrittlicheren Technologien, wie Metabarcoding (Kapitel 1 und 2) und computergestützten Vorhersagen (Kapitel 1 und 2). Zum Abschluss der Arbeit werden in einer abschließenden Synopsis die gewonnenen Ergebnisse zusammengetragen und kritisch betrachtet, sowie Ideen für zukünftige Studien formuliert.

In Kapitel 1 untersuchten wir die Auswirkungen des Bodentyps (lehmig vs. sandig), der Bewirtschaftung (konventionell vs. ökologisch), der Weizensorte (anspruchslos anspruchsvoll) und die Wechselwirkungen zwischen diesen Faktoren die Zusammensetzung und Funktion der Bakteriengemeinschaft in der Rhizosphäre von Weizen unter extremen Trockenheitsbedingungen. Das Wasserdefizit übte einen starken Druck auf die Rhizosphärenbakteriengemeinschaften aus und stand in Wechselwirkung mit dem Bodentyp und der Bewirtschaftung, nicht aber mit den Weizensorten. In den Sandböden beobachteten wir eine starke trockenheitsbedingte Veränderung der Zusammensetzung der Gemeinschaft mit einem Rückgang der Artenvielfalt und der extrazellulären Enzymproduktion, während die Veränderungen durch die Trockenheit in den fruchtbaren Lehmböden weniger stark ausgeprägt waren. Eine besondere Ausnahme von diesem Muster wurde für Enzymaktivitäten gefunden, die am Kohlenstoffkreislauf im Sandboden beteiligt sind, was auf eine positive Rückkopplung zwischen Pflanze und Bodengemeinschaften unter Trockenheit hindeutet.

In Kapitel 2 wurden zwei einzelne, jedoch miteinander verknüpfte Ziele verfolgt. Erstens nutzten wir die Plattform der Global Change Experimental Facility (GCEF), um die Auswirkungen von zwei Anbaupraktiken (konventionell vs. ökologisch) und zwei Klimabehandlungen (ambient vs. zukünftig) auf die Zusammensetzung der Bakteriengemeinschaft und die Aktivitätsprofile extrazellulärer Enzyme, die an den C-, N- und P-Zyklen in der Rhizosphäre von Weizen beteiligt sind, in zwei verschiedenen

Pflanzenwachstumsstadien zu untersuchen. Die Klimabehandlung in der GCEF hatte keinen Einfluss auf die Rhizosphärenbakteriengemeinschaften. Die Zusammensetzung und die Funktionen der Rhizosphärenbakteriengemeinschaften unterschieden sich signifikant zwischen dem vegetativen und dem generativen Wachstumsstadium der Pflanzen, sowohl im konventionellen als auch im ökologischen Landbau. In einem zweiten Schritt nutzten wir die gewonnenen Daten, um die Genauigkeit rechnerischer Ansätze wie Tax4Fun und PanFP zur Vorhersage funktioneller Profile von Bakteriengemeinschaften auf der Grundlage von 16S rDNA-Daten zu überprüfen. Zu diesem Zweck verglichen wir die gemessenen Enzymaktivitäten mit den jeweiligen Genhäufigkeiten in der Gemeinschaft unter verschiedenen Klima- und Anbaubedingungen und in den beiden Entwicklungsstadien der Pflanzen. Diese Analyse ergab qualitative, aber nicht unbedingt quantitative Übereinstimmungen, d. h. wir fanden Auswirkungen der verschiedenen Behandlungen auf die gemessenen Enzymaktivitäten, die sich auch in den Genhäufigkeiten widerspiegeln.

Kapitel 3 stellt einen ergänzenden Ansatz zu Kapitel 2 dar, wobei der Schwerpunkt auf einzelnen Bakterienarten liegt. Mit kulturabhängigen Methoden wurden gezielt stark Phosphatsolubilisierende Bakterien aus der Rhizosphäre von Weizen isoliert und auf ihre In-vitro-Trockenheitstoleranz getestet. Unter den mehr als 800 isolierten Arten dominierten Phyllobacterium-, Pseudomonas- und Streptomyces-Arten. Während Anbaumanagement und Klimabehandlung nur geringe Auswirkungen hatten, wirkten sich die Wachstumsstadien des Weizens signifikant auf die Zusammensetzung und Funktionen der Isolate aus, wobei eine Dominanz von Pseudomonas-Arten in der vegetativen Wachstumsphase durch eine Dominanz von Phyllobacterium-Arten in der generativen Wachstumsphase ersetzt wurde. Da das Potenzial zur P-Solubilisierung mit einer hohen in vitro-Trockenheitstoleranz einherging, wurden Phyllobacterium-Arten als vielversprechende pflanzenwachstumsfördernde Rhizobakterien (PGPR) für Weizen unter zukünftigen Trockenheitsbedingungen charakterisiert.

Zusammenfassung

In der Synopsis dieser Arbeit bewerteten wir die multifaktoriellen und multidisziplinären Ansätze, und untersuchten, inwieweit die Anpassungen der Bakteriengemeinschaften in Feldund Topfversuchen übereinstimmen oder sich unterscheiden.

Insgesamt fanden wir allgemeine, aber auch differenzielle Anpassungsprozesse von Bakteriengemeinschaften und einzelnen Arten in der Rhizosphäre von Weizen an die Trockenheit, wobei einzelne Faktoren, aber auch interagierende Effekte einen starken Einfluss auf diese Prozesse ausübten. Diese Studie unterstreicht damit die Bedeutung multifaktorieller Ansätze, um gemeinschafts- oder artspezifische Rückkopplungen zwischen Pflanze und Boden zu untersuchen.

SUMMARY

Wheat is one of the worldwide most cultivated crop and highly contribute to secure food production in different world regions. Although, it grows almost ubiquitous, its production is severely vulnerable to drought. Soil and rhizosphere microbial communities associated to plants come more and more into the focus of modern agrobiology research, as a solution to maintain productivity under drought, and reinforce sustainable production. Whereas numerous studies on wheat production and the beneficial influence of the soil microbiome under drought have been performed in arid and semiarid regions of the world, comparable studies in Central Europe are rare. This might change due to the ongoing climate crisis and expected less frequent precipitations during the vegetation season. So far, most studies that focus on acclimatization of the wheat rhizobiome to water deficit mostly consider, at best, two interacting factors, and lack to consider other biotic or abiotic drivers of rhizosphere microbial communities structure and function. Therefore, the aim of this thesis was to combine complementary analytical approaches to investigate drought-induced structural and functional changes in wheat rhizosphere bacterial communities and individual species in dependency of soil type, farming system, wheat cultivar and plant development stage, and to determine how these changes affect wheat performance as a consequence of possible climate change scenarios in Central Germany.

The presented thesis starts with a general introduction and presentation of the project, followed by three consecutive chapters containing the main findings published in peer-reviewed articles. Starting with an experiment performed in the greenhouse (Chapter 1) and then moving to a realistic climate scenario under field conditions (Chapter 2 and 3), the three chapters demonstrate the sole and interacting effects of drought and farming system (Chapter 1-3), soil type and wheat cultivar (Chapter 1), as well as plant growth stages (Chapter 2 and 3) on bacterial communities and individual taxa of the wheat rhizobiome. The methods used reach from traditional cultivation and *in-vitro* bioassays (Chapter 3), over extracellular enzyme activity

potentials (Chapter 1 and 2) to more advanced technologies such as metabarcoding (Chapter 1 and 2) and computational tools (Chapter 1 and 2), addressing single bacterial taxa as well as community level. Finalizing the thesis, a concluding synopsis compiles and critically reviews the gained results and formulates future study perspectives.

In Chapter 1, we evaluated the impact of soil type (loamy vs. sandy), farming management (conventional vs. organic), wheat cultivar (non-demanding vs. demanding), and the interacting effects of these factors on wheat rhizobacterial community composition and function under extreme drought conditions. Water deficit exerted a strong pressure on rhizobacterial communities, and interacted with soil type and farming management, but not with the wheat cultivar types. In the sandy soil, we observed a strong drought-induced shift in community composition, with a decrease in species diversity and extracellulare enzyme production, while changes by drought were less prominent in the fertile loamy soil. A particular exception from this pattern was found for enzyme activities involved in carbon cycling in the sandy soil suggesting a positive plant-soil-feedback on enzyme activities by drought conditioning.

In Chapter 2, two individual, but interrelated aims were pursued. First, we used the platform of the Global Change Experimental Facility (GCEF) to explore the impact of two farming practices (conventional vs. organic) and two climate treatments (ambient vs. future) on bacterial community composition and activity profiles of extracellulare enzymes involved in C,N and P cycles in the wheat rhizosphere at two different plant growth stages. The climate treatment in the GCEF had no effect on the rhizobacterial communities. Rhizobacterial community composition and functions significantly differed between vegetative and mature growth stages of the plants, in both conventional and organic farming. In a second step, we reused the data to explore further the accuracy of computational approaches, like Tax4Fun and PanFP, to predict functional profiles of bacterial communities based on 16S rDNA abundance data. To this end, we compared the measured enzyme activities with respective gene abundances in the community under different climate and farming treatments, and at the two plant development stages. This analysis revealed qualitative, but not necessarily quantitative

concordances, i.e. we found effects of the different treatments on the measured enzyme activities reflected in the gene abundances.

Chapter 3 is a complementary approach to Chapter 2 with a focus on individual bacterial species level. Culture-dependent methods were used to specifically isolate strong P-solubilizing bacteria from the rhizosphere of wheat, which were tested for their *in-vitro* drought tolerance. Among the more than 800 isolated species, *Phyllobacterium*, *Pseudomonas* and *Streptomyces* species dominated. While farming management and climate treatment had only minor effects on composition and functions of the isolates, the wheat growth stages had an impact, whereby a dominance of *Pseudomonas* species at the vegetative growth phase was replaced by dominance of *Phyllobacterium* species at the mature growth phase. Since P-solubilizing potential was paralleled by a high *in-vitro* drought tolerance, *Phyllobacterium* species were characterized as promising plant growth promoting rhizobacteria (PGPR) of wheat under future drought conditions.

In the synopsis part, we evaluated the multifactorial and multidisciplinary approaches and investigated to what extent the adaptations of bacterial communities in field and pot experiments coincided or differed.

Overall, we found common and distinct adaptation processes of bacterial communities and individual species in the rhizosphere of wheat to drought, whereby single factors, but also interacting effects exerted a strong impact on these processes. This study underlines the importance of multifactorial approaches to reveal community- or species-specific plant-soil-feedbacks.

INTRODUCTION

When extreme events become the new normal

The greenhouse effect is a natural process that stabilizes temperature on Earth and makes life possible. Under clear skies, circa 60 % of this greenhouse effect arises from water vapor in the atmosphere, 26 % from carbon dioxide, 8 % from ozone and ca. 6 % from trace gases (Ch₄, N₂O and others), and these values correspond to a balance between their contribution to heat absorption and re-radiation from the sun and the Earth's surface, respectively (Kiehl and Trenberth 1997). In the last century, anthropogenic activities, e.g. the burning of fossil fuels to produce energy, caused a misbalance of these gases in the atmosphere. For instance, since the industrial revolution the atmospheric concentration of carbon dioxide has increased from 280 parts per million (ppm) to 370 ppm in 2003 (increase of 31 %, CDIAC) and 409.8 ppm in 2019 (increase of 42 %, CDIAC) (Figure 1). The consequences of increasing and accumulating anthropogenic produced greenhouses gases in the atmosphere over decades as well as the effects of land use changes on the global climate have long been underestimated.

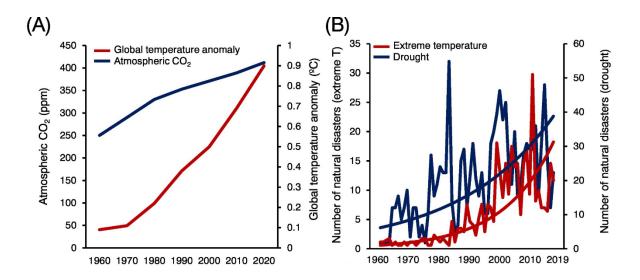


Figure 1 Exceeding bounds of natural variability in (A) CO₂ concentrations and global temperature, which led to (B) higher number of natural disasters in drought and extreme temperatures between 1960 and 2019 (Zandalinas et al. 2021)

Figure 1 shows that since 1960, atmospheric CO₂ concentrations steadily increased, which caused overall changes in global mean temperatures exceeding the bounds of natural

variability (Zandalinas et al. 2021). Long-term changes in climate are accompanied by an increasing number of extreme and disastrous events in local climates around the globe, e.g. floods (Pall et al. 2011, Min et al. 2011), extended periods of drought (Ciais et al 2003, Hari et al 2020, Zandalinas et al. 2021) and storms (Demski et al 2017), which severely endangers human life standards and food production. While extreme events like floods and storms are highly destructive, they occur more occasionally, and their spatial and temporal influence is restricted to certain regions, e.g. coastal regions and islands. In contrast, heat events and related rainfall deficits put extreme pressure on vegetation over larger areas and for longer periods of time, especially during active growing seasons of plants (Hari et al. 2020, Chung et al 2014). Becoming more frequent in the last decades, heat events and related drought periods are regarded as the most challenging outcome of global warming.

Figure 2 shows the distribution of weather anomalies and the vegetation health index (VHI) across Europe in the summers (June-August) 2003, 2018 and 2019 (Hari et al. 2020), which were extremely hot and dry. Especially Central Europe experienced high mean summer temperature and precipitation anomalies pressing the VHI below 30, which is the critical threshold for a healthy vegetation (Figure 2). Investigating the origin of 2003 heat event, Liu et al. (2020) identified anomalies in the atmospheric circulation and the strong humiditytemperature coupling as main reason for the extreme drought in Central Europe. The heatwaves of 2018 and 2019 differed from the one in 2003, not only regarding the geographical extension, which included Northern Scandinavian countries (Figure 2, Hari et al 2020), and the occurrence in two consecutive year, but mainly due to their origin. Changes in land cover with higher percentage of agricultural sites compared to forests caused a higher net surface radiation, which was mainly responsible for the heat event in 2018 (Liu et al. 2020), and possibly 2019. Since land cover defines the amount of soil water retention and carbon sequestration - a closed vegetation cover can take more carbon dioxide from the atmosphere and reduce evaporation - agricultural sites are highly vulnerable to future effects of climate change.

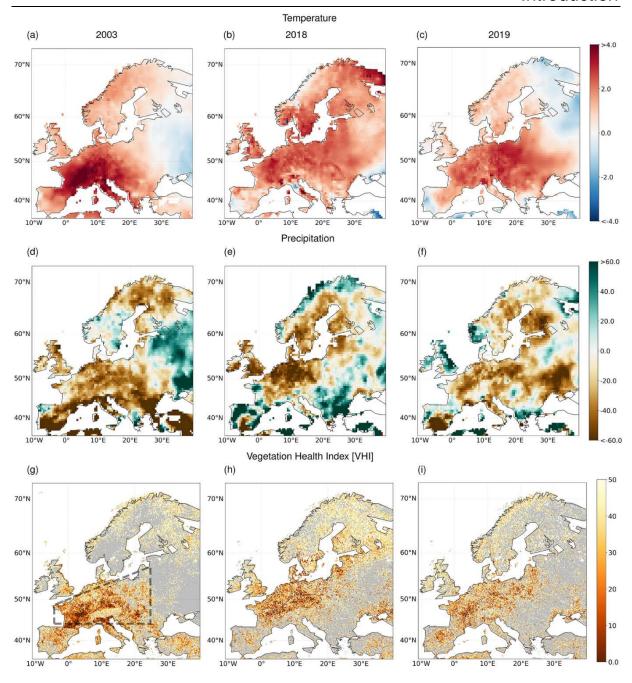


Figure 2 Weather anomalies and vegetation health index (VHI) across Europe during extreme events in 2003, 2018 and 2019. For each year (a-c) mean summer (June-August) temperature (°C), (d-f) precipitation anomalies (%) and (g-i) consequences for vegetation (poor vegetation health, VHI≤30) are presented (Hari et al. 2020).

The temperate zone of Central Germany where this study was conducted is characterized by such an agriculture-dominated landscape. Established climate models predict for this region long-term increases in temperatures and a reduction in summer precipitation up to 20 % by 2080 (mean values of: COSMO-CLM (Rockel et al. 2008), REMO (Jacob and Podzun 1997), and RCAO (Doscher et al. 2002)). Additionally, extreme events, such as the two consecutive dry years in 2018 and 2019, which were exceptional so far, but exerted extreme pressure on

vegetation health in Central Germany (Figure 2), are expected to be more frequent and severe in the future (Hari et al. 2020). Mitigating and adapting to climate change and derived weather anomalies has thus become a central focus for food production (Berrang-Ford et al 2011).

Feedback to agricultural production and need for management adaptation

Between 1979 and1989, worldwide areas for cropland and permanent pastures increased about 2.2 % and 0.1 % respectively, while forest area decreased by 1.8 % (Dale 1997). Today, half of the habitable land is used for agricultural production, and 37 % are covered by forests (Ritchie 2019). From the anthropogenic point of view, the increasing need for agricultural products to the feed growing world population justifies intensive agricultural production. By 2050, the world's population is expecting to grow to 9.8 billion people (UN 2017), which requires an increase in agricultural production by 70 % compared to 2005 (ELD, 2015). Intensification of agricultural production has become a major constraint to fight global warming, but concurrently the agricultural lands are highly sensitive to climate variations limiting their productivity. For instance, global yield variation between 15% and 35% for wheat, oilseeds and coarse grains can be explained by El Nino Southern Oscillation phenomenon that causes alternating flood and drought events (Ferris 1999). For comparison, extreme drought events, such as in 2003 caused a 30% reduction in European gross primary production (Ciais et al. 2005), and the one in 2018, a drop of 17 % in wheat production compared to previous five year average (FAOSTAT).

The "intensive agriculture" paradigm states that increased input of mineral fertilizers and pesticides combined with intensive mechanical management leads to an increased yield. Under conditions of water scarcity this paradigm is however challenged, as the applied fertilizers cannot be taken up by the plant, accumulate in the soil and may create a toxic environment (Magid et al. 2020), or, in case of nitrogen fertilizers, be leached to deeper soil layers and groundwater (Wang and Li 2019). To minimize the ecological footprint of agriculture, following changes in agricultural management have been proposed to reach sustainability (summarized in Bodner et al. 2015 and Howden et al. 2007):

- The choice of appropriate plant varieties/species with increased resistance to heat shock, drought and pests, as well as adaptation of fertilization, irrigation and pest control regimes according to prevailing climate.
- Efficient water management to retain water under water deficit and prevent water logging, erosion, and nutrient leaching during extreme rainfalls by e.g. crop residue retention.
- Adapting cropping activities to precipitation patterns in timing and location.
- Broader spectrum of activities, e.g. agriculture and livestock raising.

Nevertheless, these improvements are still based on traditional approaches of high input systems, which induce changes in biogeochemical and hydrological cycles of the soils, threaten soil organisms and functions (State of knowledge of soil biodiversity, 2020), and often leave agricultural lands vulnerable and with limited self-regulating properties to be able to cope with disturbances (Orgiazzi et al. 2016). Therefore, there is an urgent need to bioremediate the degraded areas of long-term agricultural sites while maintaining productivity. In line with this objective, the concept of biofertilization has raised increased attention (Al Abboud et al. 2014, Bhat et al. 2015, Hernández-Fernández et al. 2021). Biofertilizers are beneficial microbes often mixed with a stabilizing carrier material, such as organic manure or clay. It was shown that the use of biofertilizers stimulates various soil processes, i.e. fixation of nutrients, soil stability, biological control of diseases or bioremediation of contaminated soils, and thus indirectly or directly promotes plant growth (reviewed in Sneha et al. 2018). However, biofertilizer application does not necessarily guarantee better plant performance (Mitter et al, 2021), since site-specific abiotic parameters, and especially the on-site soil microbial communities can restrict or even eliminate biofertilizer efficiency under changing environmental conditions (Debnath et al. 2019).

Difficulties in exploring the soil microbiome and identification of plant beneficial microbial taxa

The soil microbiome is a highly diverse community, which comprises microbial taxa from all three domains of life, Bacteria, Archaea and Eukarya. Bacteria and fungi contribute 102-104 times more to soil biomass than other groups of microorganisms (Fierer, 2017). As such, most studies on soil functioning relate to bacteria- and fungi-mediated mechanisms (Kuzyakov and Blagodatskaya, 2015; Verstraete and Mertens, 2004, Fierer, 2017). Through their activities, both groups provide important ecosystem services and, thus, are crucial for soil health and productivity. Ecosystem services include the production and consumption of atmospheric trace gases, the regulation of soil acidity, carbon dynamics and nutrient cycles balancing the pools of available and unavailable nutrients in the soil matrix (Fierer, 2017, reviewed in Verstraete and Mertens, 2004). Plants take advantage and even foster these services on the narrow rootsoil interface, i.e. the rhizosphere, to promote their growth (reviewed in Morgan et al. 2005). Releasing carbon-rich root exudates into the rhizosphere, plants influence the rhizosphere microbiome and seek to selectively attract beneficial microbial taxa from the surrounding pool of microorganisms (Yuan et al 2018). Depending on the involved microorganisms such mutual symbioses are termed either mycorrhizal associations, plant growth-promoting fungi (plantfungal symbiosis, Rayner 1927, Hossain et al. 2017) or plant growth promoting rhizobacteria (PGPR) (plant-bacterial symbiosis, Kloepper and Schroth 1978).

While mycorrhizal and other plant-fungal associations dominate in undisturbed forest and grassland ecosystems (Banerjee et al. 2019), in intensively managed agricultural systems beneficial plant-microbe interactions are mainly attributed to bacteria (Miller and Lodge, 1997, de Vries et al. 2013). Thereby, the pool of soil bacterial taxa from which beneficial ones can be recruited by plants is highly diverse, and their community composition strongly depends on surrounding conditions. The characteristics of different soil types, which vary in terms of edaphic parameters such as water storage and transport capacities, nutrient status and degree of mineral complexation, in combination with changes in climatic conditions, are main drivers for shifts in the presence and abundance of certain bacterial taxa (Singh et al 2007, Ramirez

et al 2020). Furthermore, plant identity, plant diversity and developmental stage exert a controlling influence on the selection process in the rhizosphere microbiome based on the changes in the quantity and quality of rhizodeposits (Garbeva et al. 2008, Francioli et al. 2018). The magnitude of oscillations and changes of all these influential factors makes it difficult to identify key species or beneficial species-interactions that drive ecosystem functions (reviewed in Schloter et al. 2018). Of note, the functional redundancy, a measure of the number of different species that contribute to one and the same function (Lawton and Brown, 1994), is high among bacteria (Allison et al. 2008). Microbial taxa, which might be functionally redundant in one system or at a certain time point, might become of greater unique importance in another system with different environmental conditions and species composition (Schloter et al. 2018). In recent years, community composition in particular has become the focus of research, and techniques in this area such as next-generation amplicon sequencing have greatly improved (reviewed in Slatko et al. 2018). However, changes in composition are not necessarily associated with functional changes. Standardized protocols to measure community functionality, as e.g. extracellular enzyme measurements, fail to trace back activities to single organisms (German et al. 2011). Appropriate methods of metatranscriptomics and metabolomics are still cost-intensive, require strong experience in sample preparation and data analysis, plus, sequence data interpretation is limited by existing numbers of sequenced genomes (Aguiar-Pulido et al. 2016, Sandhu et al. 2019). All this makes broad analyses of rhizosphere microbiota challenging. Computational approaches, which are based on functional analyses of barcode data, may be a more cost-saving alternative, but their accuracy strongly depends on the completeness of the used databases (Aßhauer et al. 2015: Tax4Fun, Jun et al. 2015). Consequently, a much more meaningful characterization of microbial diversity might be achieved by not only combining the different analytical methods to resolve structural and functional diversity on community and individual species level, but also by implementing environmental data as important drivers of diversity.

Our approach with wheat

To be able to detect key species and sub-communities of soil microorganisms enriched in response to certain environmental conditions, at least one common factor that exerts a measurable impact on the microbial community composition has to be selected. In agricultural systems, the reduction of biodiversity to a low number of cultivated plants grown in monocultures predestines the plant itself as such a factor. Winter wheat is one of the five most widely grown crops in the world and the most widely grown crop in Germany (FAOSTAT). Winter wheat requires favorable conditions in terms of nutrient supply and precipitation, and just these conditions are susceptible to changing climate endangering productivity and yield of wheat monocultures. Numerous studies have so far described the positive effect of PGPR on wheat growth under conditions of salinity (Naili et al. 2018, Desoky et al. 2020) and drought (Zafar-ul-Hye et al. 2019, Ansari et al. 2021), and also more direct effects for nutrient acquisition (Liu et al. 2019, Wang et al. 2020). While these studies indicate that taxonomically different PGPR act on wheat, Kavamura et al (2021) suggest the existence of a wheat core microbiome with a full set of beneficial functional properties. This core microbiome may persist even under changing climatic and edaphic conditions (Kavamura et al. 2021, Simonin et al. 2020). For instance, Schlatter et al. (2020) described a core rhizosphere microbiome with bacterial taxa from the genera Bradyrhizobium, Sphingomonadaceae, Massilia, Variovorax, Oxalobacteraceae, and Caulobacteraceae for dryland wheat. The identification of such a wheat core microbiome or possible key species may not only be of ecological interest, but also a crucial element toward a sustainable wheat production under future climate change.

Although several studies already investigated the impact of drought on the wheat microbiome, the combination of multifactorial and/or multidisciplinary approaches are still underrepresented. Therefore, the central aim of this study was to determine how the response of wheat rhizosphere bacterial community to water deficit is influenced by further biotic and abiotic factors, and how these factor interactions drive changes in rhizosphere microbiome that influence plant drought tolerance. For this purpose, state of the art Illumina MiSeq amplicon sequencing, computational tools and classic cultivation techniques were combined to examine

the impact of drought on wheat rhizobacterial communities and individual species in dependency of soil type, farming system, wheat cultivar and plant growth stage. This integrative pathway was realized in pot and field experiments co-manipulating different numbers of environmental factors, whereby some overlapped (Figure 3).

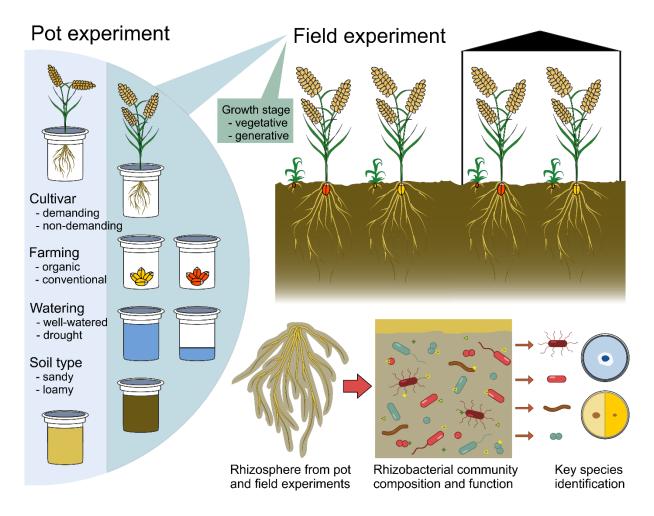


Figure 3 Conceptual overview of the thesis. The objective was to reveal the relative impacts of multiple interacting factors on wheat rhizosphere microbial community composition and function, both at the prokaryote DNA and bacterial isolate levels. Pot experiment examined how the factors wheat cultivar, farming system, watering and soil type affected community composition and function (presented in Chapter 1 of the thesis). Field experiment surveyed how the factors farming system, watering, and the plant growth stage affected rhizosphere community composition and function (presented in Chapter 2), as well as the functional traits of bacterial isolates (presented in Chapter 3).

Chapter 1 evaluates how rhizosphere prokaryote community responses to drought are modulated by treatments of farming system (conventional vs. organic), soil type (sandy vs. loamy soil), wheat cultivar (demanding vs. non-demanding) and water availability (drought vs. well-watered). The multifactorial experiment was performed as pot experiment in the green house. In order to trace back changes due to applied treatments over a whole growing season, rhizosphere samples were collected at harvest and compared to those of the starting

conditions. We expected a decrease in diversity under drought with particular enrichment of drought-tolerant taxa. The effect of drought was further expected to be stronger in the sandy soil, than in the fertile loamy soil, as well as in conventional than in organic farming.

Chapter 2 follows two goals. First, the impacts of three experimental factors, farming system (CF vs. OF), climate manipulation (ambient vs. future) and plant growth stage (vegetative vs. mature) on wheat rhizosphere prokaryote community composition and function were evaluated under field conditions in the Global Change Experimental Facility (GCEF, 20 plots with a size of 24 x 16 m, detailed study design in Schädler et al. 2019). And second, the performances of the 16S sequencing datasets based, functional annotation tools Tax4Fun and PanFP were assessed by evaluating their accuracy to predict the results obtained in relation to the effects of the three tested experimental factors land use, climate and plant development stage. For this, the functional tool predicted extracellular enzyme gene numbers were cross-compared with the corresponding measured enzyme activities in the rhizosphere of wheat. Expectations in this chapter were mainly related to the method comparison part assuming concordances and deviations between predicted and measured functions to be a consequence of community' response to the applied treatments. Thereby, we expected that the deviations should be more pronounced in treatments that induce a strong dynamic in the bacteria community.

Chapter 3 aims to identify plant beneficial taxa in wheat rhizosphere. In contrast to the other two studies, cultivation-based approaches were applied. They focused on two important traits, phosphate cycling and drought tolerance, using the same experimental system as described in Chapter 2. In response to the demand of plants for nutrients at different growth stages, we expected the abundance and activity of P-solubilizing bacteria to be increased at the active, vegetative growth stage. We further expected the climate manipulation to trigger a less diverse community, which however maintains activity under future climate conditions, while activity potentials should be increased under organic farming with lower nutrient availability.

Overall, the three studies represent two directions of scaling. Upscaling from pot (Chapter 1) to field experiment (Chapters 2 and 3), and down-scaling from the community level (Chapters 1 and 2) to the functions of individual bacterial isolates (Chapter 3). The relative impacts of

drought and interactions with the other experimental factors, as well as possible implementations for future work are discussed in the synopsis part of the thesis.

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@ CHAPTER 1

Interactions Between Soil Properties, Agricultural Management and Cultivar Type Drive Structural and Functional Adaptations of the Wheat Rhizosphere Microbiome To Drought

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Special Issue Article

Interactions between soil properties, agricultural management and cultivar type drive structural and functional adaptations of the wheat rhizosphere microbiome to drought

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Summary

Rhizosphere microbial communities adapt their structural and functional compositions to water scarcity and have the potential to substantially mitigate drought stress of crops. To unlock this potential, it is crucial to understand community responses to drought in the complex interplay between soil properties, agricultural management and crop species. Two winter wheat cultivars, demanding and non-demanding, were exposed to drought stress in loamy Chernozem and sandy Luvisol soils under conventional or organic farming management. Structural and functional adaptations of the rhizosphere bacteria were assessed by 16S amplicon sequencing, the predicted abundance of drought-related functional genes in the bacterial community based on 16S amplicon sequences (Tax4Fun) and the activity potentials of extracellular enzymes involved in the carbon cycle. Bacterial community composition was strongly driven by drought and soil type. Under drought conditions, Gram-positive phyla became relatively more abundant, but either less or more diverse in Luvisol and Chernozem soil respectively. Enzyme activities and functional gene abundances related to carbon

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degradation were increased under drought in the rhizosphere of the demanding wheat cultivar in organic farming. We demonstrate that soil type, farming system and wheat cultivar each constitute important factors during the structural and/or functional adaptation of rhizobacterial communities in response to drought.

Introduction

Mitigating negative impacts of climate change on crop production is one of the most challenging tasks of modern agriculture. In temperate ecosystems of Central Germany, climate change has been characterized by an increase in temperatures and a changed intra-annual precipitation pattern moving main rain events out of the growing season (Schädler et al., 2019). Winter wheat is the second most-produced crop worldwide (FAO, 2018) and covers more than half of the cereal production area in Germany (DESTATIS, 2021). Significant yield losses in the first decade of the 21st century were strongly correlated with heat and drought-related weather indices (Lüttger and Feike, 2018). Also during the extreme drought in 2018, wheat production in Germany significantly dropped on average by 12.6% (DESTATIS, 2021). Thereby, a strong gradient from the Northeastern to the Southwestern part of Germany emerged, ranging from -21.2% in Brandenburg, -20.8% in Saxony-Anhalt to almost no losses in Baden-Württemberg (-2.6%) (DESTATIS, 2021). Since drought events are going to further aggravate in the future, sharp declines in agricultural productivity are expected (Hari et al., 2020). Thus, if novel cultivars and agricultural systems are not applied, immense losses in crop yields can be expected.

Besides management options, the negative impacts of drought on agricultural yields can be mitigated by certain bacterial taxa colonizing the crop rhizosphere (Marasco et al., 2012; Coleman-Derr and Tringe, 2014; Azarbad et al., 2020). These plant-beneficial bacteria express a wide range of functional traits to cope with drought

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themselves and promote crop adaptation to drought conditions. To maintain their functionality soil bacteria are able to produce biofilms and protect themselves from dehydration under water scarcity (Kim et al., 2012). Mechanisms for plant adaptation include biofertilization and reduction of stress hormone production, as well as induction of systemic resistance to various biotic and abiotic stress factors (reviewed in Goswami et al., 2016). The most commonly investigated marker related to drought-adaptation of plants is the bacterially produced 1aminocyclopropane-1-carboxylate (ACC) which suppresses production of the stress-related phytohormone ethylene (Glick, 2005). Bacterial enzymes that are secreted into the rhizosphere can also reduce plants' drought stress responses aboveground by affecting glutathione, butanediol and spermidine metabolisms (Cho et al., 2008; Kasim et al., 2013; Zhou et al., 2016). However, the essential factor by which plant growth is impaired is the drought-related reduction in available soil carbon and nutrients (reviewed in van der Molen et al., 2011). To cope with the imbalance of need and availability of C sources, microbes degrade more complex carbohydrates in dry soils. This is indicated by the enrichment of microbial functional genes and enzyme activities involved in degradation of celluloses and hemicelluloses (Bouskill et al., 2016; Martiny et al., 2017).

The extent of plant growth support by rhizobacteria is strongly dependent on their trait composition, which is in turn linked to community structure and diversity. Less diverse communities with single functional key players are considered to be more vulnerable to disturbances than more diverse communities due to the high functional redundancy potential of these (Nautiyal and Dion, 2008; Wagg et al., 2021). Both community structure and its trait composition are steered by various environmental factors. Vegetation cover (Francioli et al., 2018) and plant growth stage (Wang et al., 2016; Breitkreuz et al., 2020), as well as soil moisture dynamics, alter composition and traits of soil and plant rhizosphere communities prominently (De Vries et al., 2012; Santos-Medellín et al., 2017; Ochoa-Hueso et al., 2018). Moreover, Lauber et al. (2008) found that the microbial community structure is strongly affected by soil type, a finding that was further supported by a study comparing soil bacteria as well as fungi-dominated food webs across various European field sites (de Vries et al., 2013). Besides, intensive cropland management decreases soil biodiversity and favours a bacteria-dominated food web due to the application of agrochemicals, high nutrient inputs and mechanical treatments as tillage (reviewed by Thiele-Bruhn et al., 2012, de Vries et al., 2013). Kavamura et al. (2021) recently reviewed the findings of over a decade of research on the wheat microbiome and outlined the intricate complexity of interactions under different abiotic and biotic

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conditions. So far, dependence of drought-induced community adaptation of rhizobacterial communities in dependency on soil type, agricultural management and crop cultivar has been poorly investigated, despite its critical importance for both, soil health and maintenance of agricultural yields.

For this purpose, we conducted a pot experiment to address this knowledge gap. It builds upon recent findings made in the Global Change Experimental Facility (GCEF, Schädler et al., 2019) that bacterial community adaptations in the rhizosphere of winter wheat to climate change are modified by agricultural management intensity (Breitkreuz et al., unpublished). For the pot experiment, we used loamy Haplic Chernozem and sandy Albic Luvisol, which strongly differ in their physical and chemical properties. The nutrient and humus-rich Chernozem (Altermann et al., 2005) offers favourable conditions for wheat growth. Contrary, the nutrient and humus-poor sandy Luvisol is characterized by a much lower waterholding capacity and thus higher vulnerability to drought (Schweitzer, 2010). To consider agricultural management as a driver of microbial community response, we collected topsoil from organic (OF) and conventional (CF) farming plots of well-established experimental field platforms. Two winter wheat cultivars with different site suitability, i.e. different tolerance to drought, were sown in October 2016 and exposed to drought and control conditions from tillering stage until harvest in July 2017. The composition of rhizosphere bacterial communities was determined using next-generation Illumina amplicon sequencing. Community traits were estimated by prediction of functional gene abundances, based on the relative abundances of amplicon sequence variants using Tax4Fun (Aßhauer et al., 2015), as well as by measuring the potential of cellulose and hemicellulose decomposing enzyme activities.

Our work was directed by three working hypotheses. We hypothesized (i) that drought reduces diversity of bacteria and leads to an increased relative abundance of drought-tolerant bacteria and (ii) that these drought impacts are more pronounced in the sandy Luvisol than in the loamy Chernozem. We further expected, that (iii) the communities in OF soils are more resistant to drought than in CF soils, since the lower management intensity in OF provides a more diverse community with a broader functional repertoire that can cope with drought stress.

Results

Soil type and farming system-specific initial soil parameters and watering treatment effects

Chernozem and Luvisol soils collected from conventional and organic farming field plots were analysed for their chemical properties (Table 1). As expected, total nitrogen (N) and carbon (C) contents, as well as available N

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Table 1. Initial properties of the used soils.

	HaplicChernozem				Albic Luvisol			
	Conventional farming		Organic farming		Conventional farming		Organic farming	
Total organic carbon in %	1.95	(0.04)a	1.95	(0.10)a	0.81	(0.02)b	0.69	(0.04)
Total nitrogen in %	0.16	(0.01)a	0.15	(0.01)b	0.06	(0.01)c	0.06	(0.01)
C/N ratio	12.2	(0.70)ab	12.9	(0.50)a	12.6	(0.40)ab	11.9	(0.50)
Available mineral nitrogen (mg kg ⁻¹)	14.8	(1.10)a	13.5	(0.90)b	9.22	(0.43)c	8.16	(0.21)
NO ₃ -N (mg kg ⁻¹)	12.7	(0.60)a	10.8	(0.60)b	7.22	(0.27)c	6.82	(0.09)
NH ₄ -N (mg kg ⁻¹)	2.11	(0.50)b	2.68	(0.27)a	2.00	(0.16)b	1.34	(0.12)
Available phosphorus (mg/100 g dry soil)	3.42	(0.18)c	3.49	(0.28)c	5.74	(0.32)a	4.75	(0.21)
Available potassium (mg/100 g dry soil)	14.29	(0.35)b	14.19	(1.25)b	17.26	(0.58)a	7.42	(0.37)
pH Y	6.38	(0.05)a	6.37	(0.03)a	5.87	(0.05)c	6.08	(0.05)
Maximum water holding capacity (WHC _{max})	35%	,		,,-	28%	, -/-		/

Standard deviations are provided in brackets. Different letters within each line indicate significant differences between treatments according to ANOVA and Tukev's HSD test.

concentration and water holding capacities were higher for Chernozem than for Luvisol. However, we measured higher available phosphorus concentrations in the Luvisol (Table 1). The effect of farming system was more pronounced in the Luvisol, with higher concentrations of all measured parameters, except for total N, and higher pH in the conventional compared to the organic farming system. In contrast, only in Chernozem mineral N concentrations were higher in conventional than in organic farming system (Table 1).

At the beginning of the experiment, all pots were adjusted to 60% of the soil type-specific maximum water holding capacity (WHC_{max}). Due to the lower water holding capacity, the gravimetric soil moisture content of Luvisol (12.4 \pm 0.7%) was lower compared to that of Chernozem (16.4 \pm 0.4%). At wheat harvest, well-watered pots showed comparable soil moisture contents with averages of 16.2 \pm 0.7% for the Chernozem and 11.1 \pm 0.7% for the Luvisol. The drought treatment caused strong decreases in moisture in both soil types, with averages of 9.1 \pm 1.2% for Chernozem and 5.8 \pm 0.7% for the Luvisol (Table S1).

Wheat performance strongly affected by experimental factors

Wheat biomass production was strongly impaired by drought (analysis of variance (ANOVA), p < 0.001, F = 512.36). The reduction of wheat dry biomass by drought was comparable between the treatments ranging from 37% to 45% (ANOVA, p > 0.05 for soil, farming system and wheat cultivar). Exclusively for organic farming treatment on Chernozem, the reduction of wheat dry biomass was larger, 56% for 'RGT-Reform' and 57% for 'Dichter'.

Besides the expected reduction in plant biomass, watering regime significantly influenced the C:N ratio of grain (ANOVA, p < 0.001, F = 31.99) and straw (ANOVA, p < 0.001, F = 26.51), with higher ratios in

control than under drought conditions (Fig. S1). While the farming system significantly influenced grain C:N ratios (CF > OF, ANOVA, p < 0.001, F = 68.68), wheat cultivars differed in their straw C:N ratios ('RGT-Reform' > 'Dichter', ANOVA, p = 0.02, F = 5.62) (Fig. S1). Interestingly, the soil type had no effect on grain and straw C:N ratios. In contrast, the water use per gram biomass (ANOVA, p < 0.001, F = 172.87, Fig. S1) significantly differed between the soil types and were, in line with the biomass itself, strongly influenced by the watering regime (ANOVA, p < 0.001, F = 786.56). Thereby, higher water use (Fig. S1) was observed in Luvisol than in Chernozem, and under well-watered conditions compared to drought conditions for both soil types.

Treatment effects on bacterial community composition

Among the set of experimental factors, soil type had by far the largest impact on bacterial community assemblage (PERMANOVA, p < 0.001, $R^2 = 0.41$). In contrast, the bacterial community compositions were comparable between the two wheat cultivars (Table 2). We therefore subsequently focused on the wheat cultivar 'Dichter', while corresponding results for 'RGT-Reform' are provided in the Supplementary Material (Tables S2 and S3; Fig. S2).

Non-metric multidimensional scaling (NMDS) revealed for both soil types a separation of the original ('start') bacterial communities from the ones present at wheat harvest in July 2017 (Fig. 1A and C). At harvest, the bacterial communities were additionally clearly separated by farming system and watering regime, as well as by compartment (rhizosphere vs. bulk soil) (Fig. 1A and C; Table 2). In the Chernozem (Fig. 1A and B), the bulk soil communities were similarly affected by watering (PERMANOVA, $R^2 = 0.15$) and farming system ($R^2 = 0.13$).

In the rhizosphere of the same soil type the effect of watering was stronger ($R^2 = 0.21$) than that of the

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Table 2. Effect of experimental factors on bacterial community composition in bulk soil and wheat rhizosphere for Chernozem and Luvisol soil at wheat harvest.

	Haplic Ch	ernozem			uvisol			
	Bulk		Rhizosphere		Bulk		Rhizosphere	
	<i>p</i> -value	R^2	<i>p</i> -value	R^2	p-value	R^2	<i>p</i> -value	R^2
Watering	<0.001	0.15	<0.001	0.21	0.002	0.07	<0.001	0.25
Farmingsystem	< 0.001	0.13	< 0.001	0.11	< 0.001	0.48	< 0.001	0.38
Wheatcultivar	0.43	0.02	0.11	0.02	0.37	0.01	0.08	0.02
Watering: Farming	0.19	0.02	0.02	0.03	0.003	0.05	<0.001	0.09

PERMANOVA was applied to test for significant differences (p-value) in community composition and the respective explanatory value of each treatment (R^2). Bold values highlight significant impacts of applied treatments on community composition, p-values <0.05.

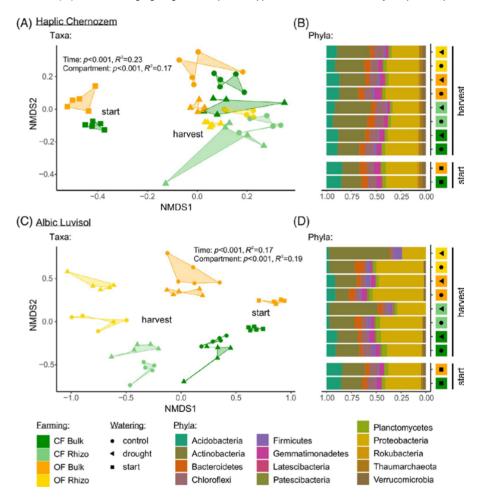


Fig 1. Impact of experimental factors on bacterial community composition in Chernozem and Luvisol for wheat cultivar 'Dichter'. NMDS for (A) Chernozem and for (C) Luvisol represent separation of communities from start conditions to harvest. Significance (p) and explanatory values (R^2) for time point and compartment according to PERMANOVA analyses are provided for each soil (refer to Table 2 for further PERMANOVA results). Relative abundances of predominant phyla within (B) Chernozem and (D) Luvisol are given for start conditions and at harvest.

farming system ($R^2 = 0.12$). For Luvisol, a different pattern was found (Fig. 1C and D), since the farming system was identified as the main driving factor of bacterial

community composition (Table 2). In bulk soil, the effect of the farming system ($R^2 = 0.47$) explained almost half of the variance, but only little explanation was provided

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by watering regime ($R^2 = 0.07$). In the rhizosphere, the impact of the water regime on community composition increased ($R^2 = 0.25$), but that of the farming system remained larger ($R^2 = 0.38$). The effects of watering and farming interacted for bulk soil and rhizosphere in Luvisol but only for rhizosphere in Chernozem. However, these interactions explained only little of the total variance (Table 2).

Changes in the distribution of main bacterial phyla across the treatments

In the Chernozem soil in all treatments and with both wheat varieties, the abundance of Acidobacteria,

Gemmatimonadetes, Rokubacteria and Verrucomicrobia decreased, while Actinobacteria and Thaumarchaeota increased in rhizosphere compared to bulk soil (Table S2). Moreover, relative abundances of rhizosphere Bacteroidetes, Gemmatimonadetes, Patescibacteria and Proteobacteria decreased (Table S2), while Acidobacteria, Chloroflexi and Firmicutes for 'Dichter' (Fig. 2), and Rokubacteria and Thaumarchaeota for 'RGT-Reform' (Fig. S2) were increased under drought. The effect of farming system was weak for Chernozem soil and differently affected phyla abundances in the rhizospheres of the two wheat varieties (Fig. 2 and Fig. S2).

In Luvisol, farming system and watering-related shifts in relative abundances of phyla were more pronounced than in

Haplic Chernozem	Rhizos	phere		
	CF	OF	control	drought
Acidobacteria	***		***	
Actinobacteria	*			
Bacteroidetes			***	
Chloroflexi	*		***	
Firmicutes			***	
Gemmatimonadetes			***	
Latescibacteria	*		*	
Patescibacteria			***	
Planctomycetes				
Proteobacteria	*		**	
Rokubacteria				
Thaumarchaeota				
Verrucomicrobia	**			
Albic Luvisol				
Acidobacteria			*	
Actinobacteria	**		***	
Bacteroidetes			***	
Chloroflexi	***			
Firmicutes	*		*	
Gemmatimonadetes	***		***	
Patescibacteria	***			
Planctomycetes	***		***	
Proteobacteria			***	
Thaumarchaeota	***			
Verrucomicrobia	**		***	

Fig 2. Bars showing the relative abundances of phyla in the rhizosphere of 'Dichter' at the harvest (separated by the two soil types). ANOVA was applied to test for differences between farming systems – conventional farming (CF) versus organic farming (OF) – and watering regimes – control versus drought – in the rhizosphere: ***p < 0.001, **p < 0.01 and *p < 0.05

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Chernozem (Fig. 1D; Table S3). Thereby, relative abundances of Acidobacteria. Bacteroidetes. Gemmatimonadetes. Planctomycetes, Proteobacteria and Verrucomicrobia declined, but Actinobacteria and Firmicutes increased under drought. Interestingly, some of the phyla whose abundance increased - Actinobacteria and Firmicutes or decreased - Gemmatimonadetes, Planctomycetes and Verrucomicrobia - under drought, also increased or decreased in the same manner in organic compared to conventional farming (Figs 2 and S2). Moreover, the relative abundances of drought nonresponsive phyla Chloroflexi, Patescibacteria and Thaumarchaeota were lower in the organic than in the conventional treatment (Fig. 2). Independent of watering regime and farming system, relative abundances of phyla Acidobacteria, Gemmatimonadetes and partly of Chloroflexi were lower in rhizosphere samples compared to bulk soil (Table S3). We further compared treatment effects on diversity indices of the two most dominant phyla. Proteobacteria and Actinobacteria in the rhizosphere (Figs 3 and S3).

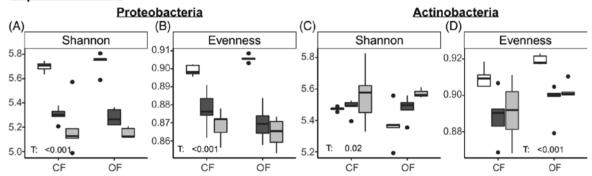
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Overall, we observed a decline of their relative Shannon diversity and evenness values under drought, as well as from start to harvest, in both soil types. These effects were more pronounced in Luvisol than Chernozem. One exception of this pattern was found for Actinobacteria diversity in the Chernozem, which increased by trend under drought in the rhizospheres of both wheat cultivars (Figs 3 and S3). The effect of farming on diversity indices was prominent in Luvisol (ANOVA, p < 0.001, CF > OF) but not in Chernozem.

Functional adaptation of the bacterial communities to drought

Functional adaptation of the microbial community was estimated by measuring activity potentials of microbial enzymes involved in carbon cycling, and the abundances of drought stress-responsive functional genes in the rhizosphere using functional predictions by Tax4Fun.

Haplic Chernozem:



Albic Luvisol:

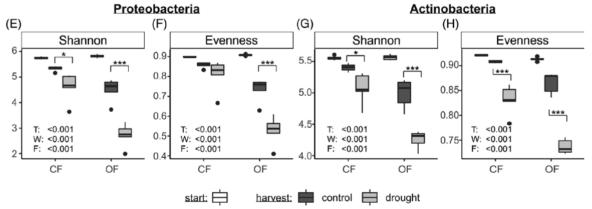


Fig 3. Alpha-diversity indices for Proteobacteria (subfigures A, B, E, F) and Actinobacteria (subfigures C, D, G, H) in the rhizosphere of 'Dichter' in dependency of sampling time (T), watering regime (W) and farming system (F) (each treatment, n = 5). Panels A–D and E–H illustrate differences in Chernozem and Luvisol respectively. Significant differences induced by the treatments are given in each panel as p-values (ANOVA). Significant differences within treatments are marked according to TukeyHSD as following: ***p < 0.001, **p < 0.01 and *p < 0.05.

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Enzyme activities of cellulases, glucosidases and xylosidases were higher in Chernozem than in Luvisol at harvest (ANOVA, p < 0.001 for all three enzymes). Besides, we determined higher enzyme activities at the harvest compared to starting conditions in Chernozem soil, while the opposite pattern was observed in Luvisol soil (Figs 4 and S4). At harvest, glucosidase and cellulase activities were significantly higher in conventional farming treatments than in organic farming treatments for Chernozem (Figs 4 and S4). In the Luvisol, we detected the same farming system effect on all three measured enzyme activities (Figs 4 and S4). The effect of drought on enzyme activities was stronger for 'Dichter' (Fig. 4) than for 'RGT-Reform' (Fig. S4). For 'Dichter', enzyme activities were lower under drought conditions in Chernozem and for cellulase activity in Luvisol. In contrast glucosidase and xylosidase activities were increased by drought in the latter soil type (Fig. 4). Interaction effects between watering and farming system were significant for glucosidase, cellulase and xylosidase activities in the Luvisol. Thereby, enzyme activities were increased under drought conditions in the organic farming soil and did not differ between watering treatments in the conventional soil (Fig. 4). In pots with 'RGT-Reform', drought negatively affected xylosidase and glucosidase

activities in the Chernozem but promoted glucosidase activities in Luvisol (Fig. S4). Instead of a general drought effect on cellulase activity in Chernozem, a significant interaction between watering and farming system occurred: cellulase activity reduction was confined to organic farming soil (Fig. S4B).

The Tax4Fun-predicted abundances of drought-related functional genes in the rhizobacterial community at harvest were related to the abundances of genes in the starting community. Increases and decreases showed a soil type and farming-specific pattern and were compara- ble for both wheat cultivars (Table 3). Among the func-tional genes that are supposed to be more abundant under drought, only genes of dextransucrase and fructan betafructosidase showed increased abundances in the communities of both soil types. The increase of these genes was most prominent in the Luvisol with organic farming management. In contrast, genes related to bio-film production, as well as ACC-deaminase production were surprisingly more abundant in communities of well-watered soils, irrespective of soil type and farming system (Table 3). Functional genes related to glutathione, spermidine and butanediol degradation were more abun- dant under control conditions, which was in line with our expectations (Table 3).

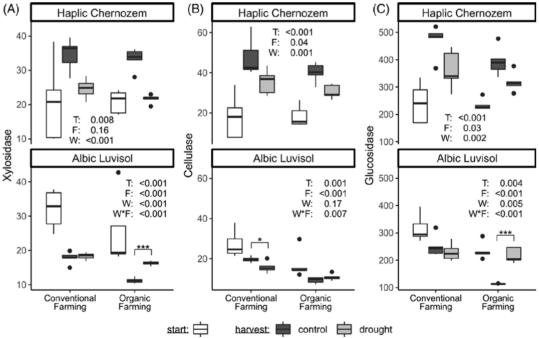


Fig 4. Activity potentials of C-cycling enzymes in the rhizosphere of wheat compared to starting conditions without plants. (A) Xylosidase, (B) cellulase and (C) glucosidase activities are presented for the wheat cultivar 'Dichter'. Significant impacts of experimental factors were calculated using ANOVA and are given as p-values for each enzyme and soil type. Thereby, sampling time effect (T) was tested between start and harvest, while watering (W) and farming (F) effects, as well as interaction of the two factors, were evaluated at harvest. Significant interaction effects are given as follows: ***p < 0.001, **p < 0.05.

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Table 3. Abundances of bacterial functional genes related to drought stress response of plants.

			Haplic C	hernozem			Albic	Luvisol	
		Convent	ional	Organic		Convent	ional	<u>Organic</u>	
		control	drought	control	drought	control	drought	control	drought
	Dext1	b	ab	ab	ab	ab	ab	ab	a
	Fruct1	b	ab	ab	ab	b	b	ab	а
er"	PGA ¹	c	bc	ab	bc	a	С	bc	С
cht	ACC1	b	b	a	b	b	b	b	b
"Dichter"	Buta ²	b	b	a	b	ab	b	ab	b
-	Sper ²	b	ab	а	ab	ab	ab	ab	ab
	Glut ²	bc	bc	a	bc	b	С	bc	С
	Dext ¹	ab	ab	ab	ab	b	a	ab	a
<u>"</u> E	Fruct1	ab	a	ab	a	b	a	ab	a
ģ	PGA ¹	a	ab	ab	ab	a	ab	ab	b
Re	ACC1	a	ab	ab	ab	ab	а	ab	b
Ť.	Buta²	а	ab	ab	ab	ab	ab	a	b
"RGT-Reform"	Sper ²	ab	ab	ab	ab	ab	a	ab	b
	Glut ²	a	а	а	ab	ab	а	ab	b

Blue bars indicate an increase and red bars a decrease of functional genes at harvest compared to starting community. Significant differences of gene abundances between treatments (i.e. in each row) are indicated by different small letters (Tukey's HSD). Dext: dextransucrase, Fruc: fructan beta-fructosidase, PGA: biofilm PGA synthesis lipoprotein, ACC: 1-aminocyclopropane-1-carboxylate deaminase, Buta: (R,R)-butanediol dehydrogenase, Sper: spermidine dehydrogenase, Glut: glutathione reductase (NADPH).

conditions.

Discussion

Performance of wheat is impaired by drought

Even though we selected wheat cultivars with different site requirements, growth performance was equally impaired by drought for both cultivars. This may be related to comparable cultivar properties that are relevant for the specific treatments that were used in our study. In contrast, Shao et al. (2005) found strong differences in the capability of 10 different wheat cultivars to cope with drought, which were related to the origin of cultivaand individual water stress thresholds of the wheat cultivars. We observed a decrease in water use per biomass production under drought conditions. It was shown that the reduction in transpiration and also turgor leads to a decrease in nitrate uptake and induces N deficiency (Ferrario-Méry et al., 1998; Xu and Zhou, 2006). In parallel, sugars accumulate in leaves to avoid dehydrationinduced tissue damage (Nio et al., 2011; Reguera et al., 2013). In line with this, we found significantly higher C to N ratios in wheat straw under drought conditions. Interestingly, we detected the same pattern for the C to N ratio of grain. In wheat grain, carbon is stored in form of

starch making up to 75% of the final dry weight (Hurkman et al., 2003). Severe drought conditions were found to decrease not only grain weight but also starch accumulation in wheat endosperm (Yu et al., 2016; Zhang et al., 2017), which would suggest lower grain C to N values under drought conditions. The contrasting pattern, observed in our study, may be related to impaired protein storage in grains under drought. The grain C to N ratio was at its lowest in OF wheat, suggesting severe N limitation. Overall, our results indicate that lower fertility of organic farming soil enhances negative drought effects on wheat performance. However, since most of the currently available wheat cultivars were bred with the focus on high yields, these cultivars have a lack of functional traits to perform well in low-input agroecosystems (Van Bueren et al., 2011).

The lower water use per gram biomass may suggest a higher water use efficiency of wheat plants in Chernozem soil than in Luvisol soil. Under field conditions, the high water retention capacity of the Chernozem soil offers a more favourable environment for wheat growth in the field (Altermann *et al.*, 2005) than sandy soils that allow a fast run-off and transpiration of water (Schweitzer, 2010). However, by using pots watered from below, we

X¹ supposed to be enriched under drought conditions.

X² supposed to be decreased under drought

prevented the loss of water and nutrients and stimulated the growth of winter wheat in Luvisol. Higher water use per gram biomass in Luvisol than Chernozem might be related to faster root growth (lannucci et al., 2017), enabling wheat to reach water and mobile nutrients in Luvisol sooner (Liu et al., 2017). Moreover, low clay content of Luvisol (3%) leads to less frequent bonding of nutrients to soil matrix than in Chernozem (20%) (Altermann et al., 2005; Schweitzer, 2010). The exorbitant use of water to produce biomass indicates a much higher vulnerability to drought in Luvisol compared to Chernozem.

Changes in bacterial community assemblage in response to soil type and farming system under drought

In both soil types, drought-induced shifts in the bacterial composition, which were more pronounced in the rhizosphere than in bulk soil. So far, drought is considered to have only minor direct impacts on diversity but larger impacts on composition of the soil microbiome (Acosta-Martinez et al., 2014, discussed in Naylor and Coleman-Derr, 2018). However, it is well known that plants closely interact with microbes competing for water and nutrients in the rhizosphere (Schimel and Bennett, 2004; Mendes et al., 2013; Philippot et al., 2013) and, thereby, foster specific selection of bacteria from the surrounding diverse bulk soil communities (Bell et al., 2014; Fan et al., 2017). Moreover, drought induces both, direct changes of the soil microbial community and indirect changes that are caused by alterations in plant physiology and biochemistry (reviewed in Naylor and Coleman-Derr, 2018). The additive microbial response to direct and indirect drought effects may explain the stronger impact of drought on the rhizosphere community, compared to that of bulk soil in our study.

Generally, it is described that drought reduces the relative abundances of fast-growing Gram-negative bacterial phyla, such as Proteobacteria, Verrucomicrobia, Gemmatimonadetes and Bacteroidetes, which predominantly exhibit a copiotrophic lifestyle. Instead, relative abundances of slow-growing Gram-positive bacterial phyla, e.g. Actinobacteria and Firmicutes, will be fostered, which are predominantly related to an oligotrophic lifestyle (as discussed in Naylor and Coleman-Derr, 2018). Correspondingly, we found that the relative abundance of Gram-negative Proteobacteria in wheat rhizosphere decreased upon drought, while Actinobacteria prevailed under drought conditions, especially in Luvisol soil. This supports our assumption made in hypothesis one, that the abundances of droughttolerant bacteria will be promoted by water limitation. Actinobacteria account for a large proportion of bacteria identified in the rhizosphere of wheat (Yadav

et al., 2018), and their drought tolerance, as well as their potential to mitigate drought stress, has already been proven (Yandigeri et al., 2012; Singh et al., 2018). In addition, we found an enrichment of the phylum Firmicutes under drought in the rhizosphere, which may support plant performance under these unfavourable conditions (discussed in Naylor and Coleman-Derr, 2018, Xu et al., 2018).

However, Gram-negative phyla were also shown to play a role in plant drought tolerance. Two studies investigated the effects of drought on the rice and cotton rhizobiome and revealed that Chloroflexi species dominated under drought conditions (Santos-Medellín et al., 2017; Ullah et al., 2019). We also observed a drought-induced increase of Chloroflexi in Chernozem and no effect of drought in Luvisol, suggesting a general drought response by this phylum. For Acidobacteria species, the response to drought was dependent on the soil type, with a relative enrichment of species in Chernozem and a decrease in Luvisol. Acidobacteria species have been described as oligotrophs, favoured in soils with lower carbon availability (Fierer et al., 2007). Extreme drought significantly increased the relative abundance of Acidobacteria in a subtropical evergreen forest soil (Bu et al., 2018) but reduced their abundance in rice rhizobiome (Santos-Medellín et al., 2017). Differences in drought responses between different studies could be related to different levels of drought stress, other plant species, whether bulk soil or rhizosphere was studied, and varying soil properties. In line with our study, the importance of soil-specific properties was supported by a study along a precipitation gradient of Mediterranean forest sites. It showed that the relative abundance of Acidobacteria correlates positively with soil ammonium and phosphorus and negatively with nitrate and magnesium levels (Bachar et al., 2010). These results underline the central role of environmental parameters for understanding the rhizobiome response to drought.

Effects of drought on bacterial community composition were more prominent in Luvisol than in Chernozem, which is in agreement with our second hypothesis. The relative abundance of rhizosphere Actinobacteria was only minor affected by drought in Chernozem, but their diversity and evenness increased under drought. In contrast, drought responses for Luvisol included a strong increase of the relative abundance of rhizosphere Actinobacteria accompanied by a significant decrease in diversity indices. Especially the decrease in evenness and Shannon diversity indicate a less diverse, actinobacterial community with only few dominating species. A higher diversity and species richness are assumed to be beneficial to cope with disturbances and stress due to higher metabolic capacities (Nautiyal and Dion, 2008). These biological indices are usually strongly associated

with soil organic matter content and composition (Fierer *et al.*, 2003; Delgado-Baquerizo *et al.*, 2017), which implicates that they can be steered by agricultural management. Indeed we found for Luvisol, that lower P, K and C contents in OF soil were accompanied by dramatically lower actino- and proteobacterial diversity. This finding contradicts our third hypothesis, which stated that lower nutrient availability fosters a more diverse community with a possibly broader functional repertoire. In Chernozem, C and nutrient contents were comparable between CF and OF soils, which may explain the comparable diversity patterns observed for these two predominant phyla.

Functional community adaptations to drought are influenced by soil properties and wheat cultivar

Functional community adaptations are important for the capacity of the microbiome to buffer drought stress for the wheat plants. Since the soil sampling was conducted at wheat harvest, the results do not necessarily reflect direct links between the microbial community and wheat performance under drought but provide indications for those.

In Chernozem, the enzyme activity potentials were reduced under drought. In general, prolonged drought stress inhibits microbial carbon turnover by suppressing microbial biomass, thereby reducing the availability of substrates and inhibiting microbial metabolism (Naylor and Coleman-Derr, 2018). To our surprise, drought increased xylosidase and beta-glucosidase activities in the rhizosphere of the cultivar 'Dichter' under organic farming treatment in Luvisol soil, which is against this expected pattern, but partly supports our third hypothesis. In line with this finding, Mganga et al. (2019) and Sanaullah et al. (2011) also observed increased betaglucosidase and cellulase activity under drought in the rhizosphere of grasses and suggested that this might be related to higher root turnover or altered rhizodeposition patterns. Indeed it has been shown that under drought an increased proportion of plant assimilates are allocated to the roots (Palta and Gregory, 1997), which changes root exudation and limits biomass production. Since this altered C allocation is not desirable for agricultural production, great efforts have been made on breeding wheat cultivars that can cope with unfavourable conditions and exhibit higher drought stress resistance (Bodner et al., 2015). The cultivar 'RGT-Reform' thus may comprise properties that allow to buffer drought stress, while 'Dichter' may be more dependent on the support of the rhizobacterial community to cope with drought stress. Since the effect was exclusively observed in organic farming soil, the contrasting responses to drought may also be caused by the use of fungicide-treated seeds in conventional farming systems versus untreated seeds

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in organic farming systems. However, due to the general bacteria-dominated food webs in agricultural systems (de Vries *et al.*, 2013), as well as the high differences in nutrient availability between the two farming systems and soil types, we consider fungicide treatment of seeds as a comparable minor impact.

We further applied Tax4Fun predictions to estimate functional adaptations of the rhizobacterial community. We predicted an enrichment of functional genes encoding for enzymes involved in complex carbohydrate degradation under drought, especially in that of dextran and fructan. Using metagenomic sequencing, Martiny et al. (2017) have shown that changes in precipitation altered the potential for bacterial carbohydrate degradation. This supports our findings and reveals possible adaptation mechanisms of the communities to cope with drought stress conditions. We further investigated a set of functional genes involved in bacterial-mediated drought stress tolerance of plants. Biofilm production in the rhizosphere is often associated with systemic resistance against plant pathogens (Timmusk Nevo, 2011), but it has also been identified as a response to drought conditions (Kim et al., 2012). Besides, ACC deaminase is a well-known marker for direct drought stress response in the rhizosphere of plants, as it controls expression of the plant stress hor-mone ethylene (Glick, 2005). Contrary to our expectation, the predicted abundances of functional genes involved in biofilm and ACC deaminase production were reduced under drought. Genes encoding for spermidine and (R,R)butanediol dehydrogenase, as well as glutathione reductase that lower the spermidine, butanediol and glutathione levels respectively, decreased, even though not significantly, under drought. This was in line with our expectations, since higher levels of these genes and thus enzyme production have been associated with plant drought tolerance (Cho et al., 2008; Kasim et al., 2013; Zhou et al., 2016). These results suggest that among the members of the rhizobacterial communities in our study we found specific functional adaptations as key for drought tolerance.

Interactions of factors drive microbial community responses to drought

Several studies highlight the overall negative effect of drought on soil microbial community structure and function. However, such studies often deny the interplay of different factors contributing to the community response pattern. For instance, Hueso *et al.* (2012) observed that the amendment of organic manure confers a higher resistance on the microbial community to deal with drought compared to communities in unamended soils. In our study, we revealed soil type-, farming system- and wheat

cultivar-specific responses of the microbial community to drought. Changes of the relative abundances of several phyla were highly treatment specific. Moreover, we observed under drought conditions a strong decrease of Actinobacteria diversity indices in Luvisol soil, while the same indices were not affected or by trend increased in Chernozem soil. Besides structural parameters, we observed treatment-specific responses of soil enzyme activities upon drought. Only for the combination of Luvisol soil x organic farming system x cultivar 'Dichter' an increase of soil enzyme activities was detected. Song et al. (2012) investigated the response of soil enzymes and root exudates from corn hybrids with different tolerance to drought at different growth stages. Despite the overall higher enzyme activities for the drought-tolerant hybrid along all growth stages, the relative importance of growth stage and/or water stress differed between the hybrids and varied among the enzyme classes measured. Therefore, the interacting effects of different abiotic and biotic factors play a crucial role to understand drought-induced soil microbial community responses. Clearly, more factors and levels of factors would be desirable to catch the broad spectrum of interactions influencing rhizobacterial community responses. On the other hand, a high number of factors also aggravate data interpretation. Thus, our approach comprises a selection of relevant levels of each factor that mainly influence wheat performance in agricultural systems.

Conclusions and outlook

The various interacting effects of the different treatment levels in our multifactorial approach revealed contextdependent responses at structural and functional levels of the bacterial communities in the wheat rhizosphere. Structural and partly functional adaptations within a vegetation period. While drought-induced shifts in community structure were strongly pronounced and differed mainly between different soil types and farming systems, the functional potential of the respective rhizobiomes to mitigate drought stress was more dependent on the wheat cultivar and the interacting effects of soil type and farming system. These results suggest that a specific pool of functional traits in the bacterial community under drought and with respect to different surrounding conditions may be more relevant, than the diversity and structure of the microbial community itself.

While Zia et al. (2021) provide general implications of rhizosphere management strategies to mitigate drought stress in agricultural systems, the review of Kavamura et al. (2021) even goes a step further being more specific on winter wheat. Using multifactorial and multidisciplinary approaches, the identification of a specific wheat core microbiome with a certain pool of functions may be the

key towards the development of microbiome-facilitated sustainable wheat production. With our work, we contributed to these aims with a particular focus on drought selection of rhizosphere microbiota – an ever-increasing process in Central European ecosystems.

Material and methods

Design of pot experiment and sampling

We selected two winter wheat cultivars, which differed in their site suitability (Landesamt für Umwelt, Landwirtschaft und Geologie, Sachsen-Anhalt, Sortenempfehlung 2016, Winterweizen). Cultivar 'RGT-Reform' (RGT, A-quality. approved 13th of March 2014, Société RAGT 2n, France) is recommended for a wide variety of agricultural production sites, whereas cultivation of 'Dichter' (D, A-quality, approved 10th of March 2014, Saatzucht Josef Breun GmbH & CO. KG, Germany) is recommended only for sites with favourable soil properties and climatic conditions. We thus classified the two cultivars as either nondemanding ('RGT-Reform') or demanding ('Dichter'), also with respect to water availability. The cultivars were grown on two soil types, which fundamentally differed in their properties: a loamy Haplic Chernozem with high humus and nutrient contents (Altermann et al., 2005), and a sandy Albic Luvisol with low water-holding capacity and nutrient contents (Schweitzer, 2010). Chernozem Luvisol soils were collected in September 2016 from experimental field research stations in Bad Lauchstädt (Saxony-Anhalt, Germany; 51°23° N 11°52° E, 118 m a.s.l.) and in Thyrow (Brandenburg, Germany, 52°16° N, 13°12° E, 40 m a.s.l.) respectively. At both sites, the upper 15 cm of topsoil were collected from conventional and organic farming plots of experimental field platforms (Bad Lauchstädt: GCEF, established in 2013, (Schädler et al., 2019); Thyrow: demonstration system for arable farming systems, established in 2005).

To define initial properties for the four soils (two soils x two farming managements, further referred as 'original soil' in the text or as 'start' in graphs and tables), pH as well as the contents of C, N, P and K were determined. Additionally, the maximum water holding capacities (WHC $_{max}$) were determined experimentally for both soil types. For this purpose, soil was soaked with water until saturation. The starting weight and the soil moisture at saturation were used to calculate the amount of water taken up by the specific soil type in percentage.

For the pot experiment, 9.3 kg for Chernozem or 11.3 kg of Luvisol were weighed in bags, mixed with fertilizers and water (60% of soil type-specific WHC_{max}) and filled in 7 L Kick-Brauckmann pots (STOMA GmbH, Siegburg, Germany). The applied fertilizers were selected in accordance with the guidelines for conventional and

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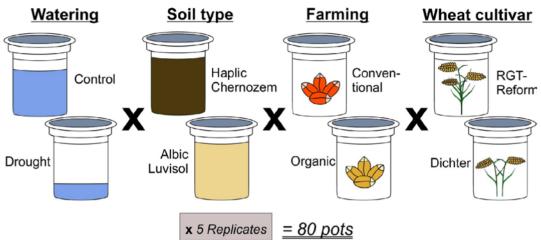


Fig 5. Experimental design including four factors: watering, soil type, farming system and wheat cultivar. Watering was identical for all pots from sowing (October 2016) until the tillering stage (March 2017). Drought was applied during the growing season (March 2017–July 2017).

organic farming (Table S4). Both conventional and organic farming pots were fertilized with 2 g N, 1 g P and 2 g K. The addition of 0.5 g Mg (MgSO₄), 0.15 g FeCl₃ and a mixture of micronutrients (A–Z solution by Hoagland and Snyder, 1933) was exclusive for conventional farming pots. For each treatment combination five replicates were prepared, which ended up in a total number of 80 pots (Fig. 5).

In October 2016, winter wheat was sown with either 16 fungicide-treated seeds per pot for the conventional farming treatment or 16 untreated seeds per pot for the organic farming treatment. The pots were randomly placed on trolleys in a cold greenhouse and kept at 60% of WHC_{max} over the winter 2016/17. In March 2017, plants were thinned and equally adjusted to a number of 12 per pot. After that, watering was stopped for half of the pots until they reached 25% of soil type-specific WHC_{max}. Subsequently, the watering regimes - 25% and 60% of WHC_{max}, reflecting limited and ample watering conditions - were maintained. The water contents of the pots were checked daily by weighing and adjusted to the targeted water content. Water was added to the cache pot in order to avoid leaching of nutrients by top irrigation and the amount of added water was recorded for each pot. During harvest in July 2017, we collected soil and wheat rhizosphere samples. For the measurement of soil abiotic parameters, five soil cores were taken from each pot, pooled, sieved to 2 mm and frozen at -20 °C. Subsequently, the roots of all plants were harvested and loose soil was removed by shaking. The remaining soil closely attached to the roots (defined as rhizosphere soil) was collected for functional and structural community analysis and stored at -80 °C.

Plant and abiotic soil parameters

Fresh and dry (60 °C, 24 h) weights of wheat were recorded for each pot at harvest. Water use by wheat plants (ml water/g biomass) was estimated for each pot, based on the total amount of added water and the final aboveground biomass. Dried straw and corn were grained to powder and total carbon and nitrogen contents were measured using an elemental analyser (Elementar Vario EL III, Elementar, Hanau, Germany).

Soil mineral nitrogen (nitrate and ammonium) was measured per flow injection analysis (FIAstar 5000, Foss GmbH, Rellingen, Germany). Briefly, 5 g of fresh soil was suspended in 20 ml of 1 M KCl solution, shaken for 1 h on a horizontal shaker and filtered through 0.45 µm cellulose nitrate filter (Sartorius Biolab Products, Göttingen, Germany). According to Schulz (2002), hot water extractable carbon and nitrogen from air-dried soil samples were measured using an elemental analyser for liquid samples (Multi N/C, Analytik Jena, Germany). Available phosphate in the soil was extracted with double lactate solution (1:50 wt./vol.) at pH 3.6 and phosphate concentrations were determined colorimetrically using the molybdenum blue method (Murphy and Riley, 1962). A halogen moisture analyser (Mettler Toledo, Gießen, Germany) was used to determine gravimetric soil water contents. For pH measurements, 12 g of air-dried soil was suspended in 30 ml of 0.01 M CaCl₂, shaken for 1 h and measured with pH electrode.

Soil enzyme activities

A modified fluorometric assay of Sinsabaugh et al. (2003) was used to determine potential bulk soil and rhizosphere

extracellular enzyme activities. The activity potentials of three hydrolytic soil enzymes involved in carbon acquisition (β -glucosidases, xylosidases and cellobiohydrolases) were measured as turnover rate of 4-methylumbelliferon (MUF)-coupled substrates (Table S5). The amount of released, fluorescent MUF can directly be related to enzymatic activity potentials. The substrate concentration (300 μ M for all substrates) applied has been optimized for the Chernozem soil of Bad Lauchstädt and was also used for the Luvisol for proper comparison of the enzymatic activity between soil types. The activity potentials were measured for bulk soil samples at germination and harvest, as well as for rhizosphere soil at harvest.

For each sample, a separate black 96-well microplate was prepared. The plates contained all three substrates. MUF dilutions to calculate quench and extinction coefficients (1.25 and 2.5 µM), as well as substrate and soil suspension controls. Approximately 250 mg of fresh soil sample was suspended in 50 ml of 50 µM acetate buffer (pH 5) for analysis. To break up soil aggregates, the soil suspensions were sonicated for 5 min, then transferred to the prepared microplates and incubated at 25 °C for 60 min. The addition of 30 ul 1 M NaOH solution stopped the enzymatic reactions. Subsequently, fluorescence was measured for eight replicates after another 3 min using an Infinite 200 PRO instrument (Tecan Group, Männedorf, Switzerland) with 360 nm excitation and 465 nm emission filters. Enzyme activities are provided as turnover rate of substrate in nmol per gram dry soil and hour (nmol g soil-1 h-1) (German et al., 2011).

DNA extraction and next-generation sequencing (Illumina MiSeq)

DNA was extracted from original soil samples from September 2016 as well as from bulk and rhizosphere soil samples taken at the harvest of wheat in July 2017. The DNeasy PowerSoil kit (QIAGEN, Hilden, Germany) was used to extract DNA from 400 mg rhizosphere and bulk soil following the manufacturer's instructions. The approximate concentrations and quality of extracted DNA were analysed with a NanoDrop ND-8000 spectrophotometer (Thermo Fischer Scientific, Dreieich, Germany). Until amplification DNA was stored at -20 °C. Using the universal primer pair 515f and 806r (Caporaso et al., 2011), which was equipped with Illumina adapter sequences, the bacterial and archaeal 16S rRNA gene V4 region was amplified. Proofreading KAPA HiFi polymerase (KAPA Biosystems, Boston, MA, United States) ensured correct amplification of the sequences and running amplification in three repetitions to minimized amplification biases. The following program was used to run PCR: initial denaturation at 95 °C for 3 min, followed by 25 cycles of denaturation at 98 °C for 20 s, annealing at 55 °C for 15 s and elongation at 72 °C for 15 s and final extension at 72 °C for 5 min.

PCR products were examined by gel electrophoresis, the three repetitions pooled and a first clean up performed using the Agencourt AMPure XP kit (Beckmann Coulter, Krefeld, Germany). In a second PCR, Illumina Nextera XT indices were attached to both ends of the bacterial fragments. The combination of indices is unique and assigns the sequences to the respective samples. The following program was used to run index PCR: initial denaturation at 95 °C for 3 min, followed by 8 cycles of denaturation at 98 °C for 30 s, annealing at 55 °C for 30 s and elongation at 72 °C for 30 s and final extension at 72 °C for 5 min. In a second clean-up, PCR products were purified using AMPure beads. The final concentration of amplified DNA was quantified with the PicoGreen assay (Molecular Probes, Eugene, OR, United States). Bacterial amplicon libraries were set to defined volumes (corresponding to 60 ng DNA for each sample) to reach an equimolar representation of each sample in the sequencing approach. The libraries were pooled in one tube and again checked for their quality with an Agilent 2100 bioanalyzer (Agilent Technologies, Palo Alto, CA, United States). Following the protocol of the MiSeq v3 reagent kit, sample libraries and PhiX control libraries were denatured and diluted to a final concentration of 10 pM, combined to a volume of 600 µl (30 µl of PhiX control library and 570 µl of bacterial amplicon library) and loaded onto MiSeq v3 reagent cartridge for sequencing. Paired-end sequencing of 2 x 300 bp was implemented on an Illumina MiSeq platform (Illumina, San Diego, CA, United States) at the Department of Soil Ecology of the Helmholtz-Centre for Environmental Research (UFZ, Halle/Saale, Germany). Demultiplexed sequences are accessible in the Sequence Read Archive under the Bioproject PRJNA680908.

Bioinformatics workflow

Demultiplexed sequences were processed using the house-intern pipeline dadasnake (Weißbecker et al., 2020), which uses the open-source program R's (v. 3.6.1; R Core Team 2017) DADA2 package (Callahan et al., 2016) in snakemake (Köster and Rahmann, 2012). 16S rDNA amplicon reads were cut and filtered using default settings of the pipeline with exception of truncation quality (set to 9) and maximum expected error (set to 0.5). Read pairs were merged with a minimum overlap of 12 bp and zero mismatches, and chimeric reads removed using the consensus algorithm. For taxonomical classification of the 16S rDNA gene amplicon sequences, the mothur implementation of the Bayesian Classifier (Schloss et al., 2009) and, as a follow up in the case of a

missing classification, BLASTn were applied, referring to the SILVA database (version 132, non-redundant at 99%; Quast *et al.*, 2012). The final output was comprised of an OTU table with taxonomic classifications for all samples.

Functional gene abundances and normalization

The package Tax4Fun (R, Aßhauer et al., 2015) was used to predict functional capabilities of the bacterial communities. The approach is based on 16S rDNA datasets generated by next-generation sequencing and the cross-comparison with KEGG orthology (KO) database (Kanehisa and Goto, 2000). Running the Tax4Fun command line in R, the latest supported version of SILVA database was used (SILVA123, released July 2015) (Quast et al., 2012). The output table contained all functional gene abundances found in the community and provided KO numbers for gene annotations and Enzyme Commission (EC) number as object identifier for enzymes. The enzymes of interest involved in direct and indirect drought stress resistance were identified and selected by KO numbers and EC numbers (Table S6). Gene abundances of the selected enzymes were corrected with bacterial biomass, which was obtained for each sample using qRT-PCR. Briefly, a standard curve of pure bacterial genomic DNA was prepared from a Phyllobacterium isolate (ranging from 0.5 to 4 ng µl⁻¹ genomic DNA). Environmental samples were diluted to 2 ng μ l⁻¹. Measured C_{Γ} values of environmental samples were related to the standard curve to estimate the amount of bacterial DNA in the environmental sample. Obtained bacterial DNA concentrations, including the dilution factor, were used for normalization of functional gene abundances.

Statistics

All statistical analyses and visualizations were performed in R (R Core Team, 2019). The output table of the dadasnake pipeline containing OTUs and taxonomy was loaded into R. In preparation, the dataset was cleaned from sequences assigned to mitochondria and chloroplasts, as well as from unassigned sequences. Furthermore, singletons were removed. Using the phyloseg package in R (McMurdie and Holmes, 2013), the OTU and taxonomy table, as well as the sample file with descriptions of the sample sites, were merged to a phyloseq object and used for further analysis. To correct different read numbers between the samples, the dataset was rarefied to the sample with minimum read number, sample B08 with 45 925 reads, and this resulted in a total number of 23 848 OTUs. PerMANOVA (vegan package, Oksanen et al., 2013) was used to indicate impacts of experimental factors watering regime, farming system, soil type, sampling time point, soil compartment and Adaptations of the wheat rhizobiome to drought 5879

wheat cultivar on community composition. Diversity indices of observed species number, Shannon diversity and evenness of the community were calculated using phyloseq package. Influence of experimental factors on plant and soil parameters, diversity indices, enzyme activities and functional genes were examined using ANOVA followed by *post hoc* tests of Tukey's HSD. Single-sided *t*-test was applied to identify significant decreases in plant biomass production under drought compared to control conditions.

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Supporting Information

Additional Supporting Information may be found in the online version of this article at the publisher's web-site:

Appendix S1. Supplementary Information

Supplemental Material:

Interactions between soil properties, agricultural management and cultivar type drive structural and functional adaptations of the wheat rhizosphere microbiome under drought

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Supplemental Tables

Table S 1 Gravimetric soil moistures (%) at wheat harvest. Standard deviations are provided in brackets. Significant differences (ANOVA and Tukey's HSD test; p<0.05) were calculated for each soil type separately and indicated by different letters.

	71		,	
System	Cultivar	Watering	Haplic Chernozem	Albic Luvisol
Conventional	"Dichter"	control	16.27 (1.00) a	11.28 (0.41) bc
farming		drought	9.26 (1.12) bc	5.85 (0.68) de
	"RGT"	control	16.13 (0.68) a	11.69 (0.96) b
		drought	9.79 (1.15) b	6.54 (0.56) d
Organic	"Dichter"	control	16.49 (0.42) a	10.94 (0.19) bc
farming		drought	8.07 (0.80) c	5.38 (0.23) e
	"RGT"	control	15.80 (0.44) a	10.49 (0.45) c
		drought	9.47 (1.18) bc	5.24 (0.57) e

Table S 2 Relative abundances of phyla in Haplic Chernozem. Abundances have been calculated for each sample and are given as average for each treatment combination. Empty spots mark relative abundances of phyla <1% in the respective treatments.

Haplic Chernozem																		
			Harvest	_														
			Conven	tional fa	Conventional farming (CF)	Ξ					Organi	Organic farming (OF)	(OF)					
			Bulk soil	=			Rhizosphere	phere			Bulk soil	<u>=</u> :			Rhizosphere	phere		
	Start		Dichter		RGT		Dichter	,	RGT		Dichter	,	RGT		Dichter	,	RGT	
	CF	OF	control	drought	control drought control	drought	control	drought	control	control drought control drought	control	drought	control	control drought control drought	control	control drought control drought	control	drought
Acidobacteria	0.14	0.16	0.11	0.12	0.10	0.11	0.06	0.08	0.06	0.09	0.13	0.13	0.11	0.13	0.08	0.10	0.08	0.11
Actinobacteria	0.20	0.17	0.21	0.26	0.20	0.24	0.33	0.36	0.30	0.33	0.19	0.26	0.23	0.25	0.29	0.32	0.29	0.30
Bacteroidetes	0.06	0.07	0.06	0.04	0.07	0.04	0.07	0.04	0.09	0.05	0.06	0.04	0.05	0.04		0.04	0.08	0.05
Chloroflexi	0.08	0.07	0.09	0.09	0.08	0.09	0.09	0.10	0.08	0.11	0.07	0.08	0.08	0.09		0.09	0.07	0.09
Firmicutes	0.01	0.01	0.02	0.01	0.02	0.01	0.01	0.01	0.01	0.01	0.02	0.02	0.02	0.02	0.01	0.02		0.02
Gemmatimonadetes	0.05	0.05	0.05	0.05	0.06	0.05	0.04	0.03	0.04	0.03	0.06	0.04	0.05	0.04	0.04	0.03	0.04	0.03
Latescibacteria	0.01	0.02	0.01	0.01	0.01	0.01		0.01			0.02	0.01	0.01	0.01		0.01	0.01	0.01
Patescibacteria			0.01		0.01		0.01		0.01		0.01				0.01		0.01	
Planctomycetes	0.04	0.05	0.03	0.03	0.03	0.03	0.03	0.03	0.03	0.03	0.03	0.03	0.03	0.03	0.03	0.03	0.03	0.03
Proteobacteria	0.30	0.30	0.31	0.29	0.34	0.30	0.28	0.25	0.29	0.25	0.32	0.29	0.32	0.29	0.31	0.27	0.31	0.27
Rokubacteria	0.02	0.02	0.01	0.01	0.01	0.01		0.01		0.01	0.01	0.01	0.01	0.02	0.01	0.01		0.01
Thaumarchaeota	0.04	0.03	0.03	0.03	0.02	0.03	0.03	0.03	0.03	0.03	0.02	0.03	0.03	0.03	0.03	0.03	0.02	0.03
Verrucomicrobia	0.04	0.04	0.04	0.04	0.04	0.04	0.03	0.03	0.03	0.03	0.04	0.04	0.04	0.04	0.03	0.03	0.03	0.03
Others	0.02	0.02	0.02	0.02	0.01	0.02	0.02	0.01	0.02	0.02	0.02	0.02	0.02	0.02	0.02	0.02	0.02	0.02

Table S 3 Relative abundances of phyla in Albic Luvisol. Abundances have been calculated for each sample and are given as average for each treatment combination. Empty spots mark relative abundances of phyla <1% in the respective treatments.

Albic Luvisol																		
			Harvest	st														
			Conve	ntional fa	Conventional farming (CF)	[]					Organic	Organic farming (OF)	(OF)					
			Bulk soil	oil Oil			Rhizosphere	here			Bulk soil	_			Rhizosphere	here		
	Start		Dichter	بر	RGT		Dichter		RGT		Dichter		RGT		Dichter		RGT	
	5	OF.	control	control drought control	control	drought	contro/	drought	control	drought	control	drought	control	drought	control	drought	control drought	drought
Acidobacteria	0.14	0.15	60:0	0.11	60.0	0.10	0.03	0.02	0.03	0.02	60.0	0.08	80.0	0.07	0.03	0.02	0.03	0.02
Actinobacteria	0.21	0.22	0.24	0.20	0.20	0.24	0.25	0.48	0.21	0.46	0.14	0.21	0.14	0.20	0.25	09.0	0.19	0.61
Armatimonadetes																		
Bacteroidetes	0.05	0.04	0.03	0.04	0.04	0.03	0.08	0.05	0.12	0.07	90.0	0.04	90.0	0.04	0.10	0.01	0.10	0.01
Chloroflexi	0.08	0.08	0.08	90.0	0.08	0.07	0.05	0.05	0.04	0.04	0.04	0.04	0.04	0.03	0.03	0.02	0.02	0.02
Firmicutes	0.01	0.01	0.04	0.01	0.03	0.03	0.04	0.03	0.01	0.02	0.02	0.05	0.04	0.04	0.03	60.0	0.02	90.0
Gemmatimonadetes	0.05	0.05	0.05	90.0	90'0	90.0	0.03	0.02	0.03	0.02	0.05	0.05	90.0	0.05	0.02		0.02	0.01
Patescibacteria	0.01		0.02	0.01	0.02	0.01	0.01	0.02	0.02	0.02	0.02	0.01	0.02	0.01	0.02		0.01	
Planctomycetes	0.03	0.04	0.05	0.05	90'0	0.05	0.05	0.02	0.04	0.02	0.04	0.03	0.04	0.03	0.03		0.03	
Proteobacteria	0.35	0.31	0.34	0.40	0.35	0.35	0.42	0.28	0.46	0.30	0.50	0.44	0.48	0.50	0.47	0.23	0.55	0.25
Thaumarchaeota	0.02	0.02	0.01		0.01	0.01	0.01	0.01										
Verrucomicrobia	0.03	0.03	0.03	0.03	0.03	0.03	0.02	0.01	0.02	0.01	0.02	0.02	0.02	0.02	0.02		0.02	
Others	0.02	0.04	0.03	0.03	0.03	0.05	0.02	0.01	0.03	0.01	0.02	0.02	0.02	0.01	0.02	0.03	0.02	0.02

Table S 4 Applie	ed fertilizers.	
	Conventional Farming	Organic Farming
Phosphorus	Triple Super Phosphate	Granulated raw phosphate (Physalg 25)
Potassium	60 % K₂O (60er Kali)	Muriate of potash (Patentkali)
Nitrogen	Ammonium nitrate	Urea

Table S 5 Measured enzyme activities involved in carbon acquisition.

Enzyme	Substrate
β-glucosidase (EC 3.2.1.21)	4-MUF-β-p-glucopyranoside
cellobiohydrolase (EC 3.2.1.91)	4-MUF-β-D-cellobioside
xylosidase (EC 3.2.1.37)	4-MUF-β-D-xylopyranoside

Table S 6 Bacterial enzyme categories related to direct and direct drought stress mitigation for

plants. For each enzy	yme, the respective KEGG ortholo	gy (KO) identifiers and Enzyme
Commission (EC) num KEGG/EC-numbers	bers were extracted from the output Description	of the Tax4Fun prediction tool. Activity
Direct response to d		y
K00689 [EC:2.4.1.5]	dextransucrase	degradation of complex plant polysaccharides under drought
K03332 [EC:3.2.1.80]	fructan beta-fructosidase	(Martiny et al. 2016)
K11931 K11935 K11937	biofilm PGA synthesis lipoprotein	biofilm production (Timmusk & Nevo, 2011; Kim et al., 2013)
Drought-induced, in	direct stress response:	
K00383 [EC:1.8.1.7]	glutathione reductase (NADPH)	involved in the plant's ascorbate—glutathione cycle, higher resilience against drought for wheat (Kasim et al. 2013)
K01505 [EC:3.5.99.7]	1-aminocyclopropane-1- carboxylate deaminase	control of stress hormone ethylene (Glick, 2005)
K03366 [EC:1.1.1.4 1.1.1.303]	(R,R)-butanediol dehydrogenase / diacetyl reductase	Butanediol is involved in stomata closure, SA signaling pathway (Cho et al., 2008)
K00316 [EC:1.5.99.6]	spermidine dehydrogenase	spermidine scavenges ROS, upregulates ABA biosynthesis and response genes, and by extension augments photosynthesis and root system architecture (Zhou et al., 2016)

Supplemental Figures

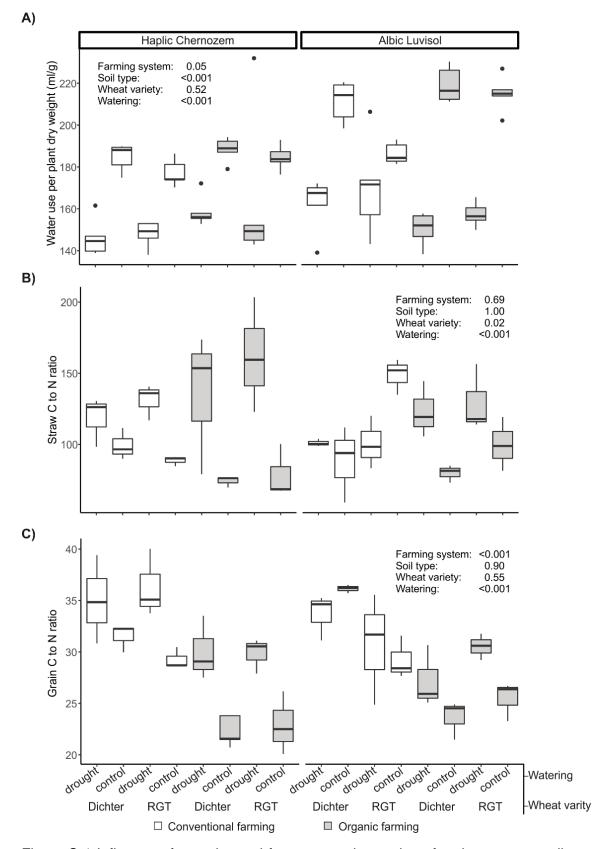
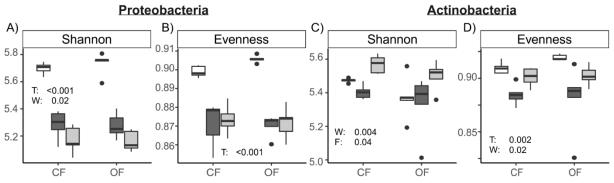


Figure S 1 Influence of experimental factors watering regime, farming system, soil type and wheat variety on C to N ratios of straw (A) and corn (B). ANOVA was applied to test for significant differences between the treatments and calculated p-values are given in each subfigure.

Haplic Chernozem	Rhizos	phere		
•	CF	OF	control	drought
Acidobacteria	*		***	
Actinobacteria				
Bacteroidetes			***	
Chloroflexi	***		***	
Firmicutes	**		***	
Gemmatimonadetes			**	
Latescibacteria	***		*	
Patescibacteria			***	
Planctomycetes				
Proteobacteria			***	
Rokubacteria			*	
Thaumarchaeota			**	
Verrucomicrobia				
Albic Luvisol				
Acidobacteria			*	
Actinobacteria	*		** *	
Bacteroidetes	***		***	
Chloroflexi	***			
Firmicutes	** *		**	
Gemmatimonadetes	***		***	
Patescibacteria	***			
Planctomycetes	***		***	
Proteobacteria			***	
Thaumarchaeota	***			
Verrucomicrobia	**		***	

Figure S 2 Bars showing the relative abundances of phyla in the rhizosphere of "RGT-Reform" at the harvest (separated by the two soil types). ANOVA was applied to test for differences between farming systems - conventional farming (CF) vs. organic farming (OF) - and watering regimes - control vs. drought - in the rhizosphere: p<0.001***, p<0.01** and p<0.05*.

Haplic Chernozem:



Albic Luvisol:

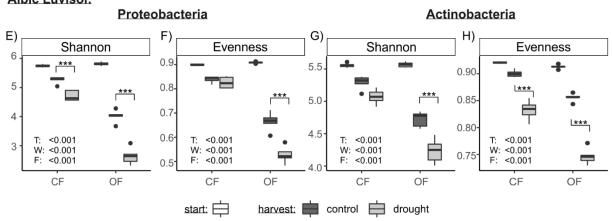


Figure S 3 Alpha-diversity indices for Proteobacteria (subfigures A,B,E,F) and Actinobacteria (subfigures C,D,G,H) in the rhizosphere of "RGT-Reform" in dependency of sampling time (T), watering regime (W) and farming system (F) (each treatment, n=5). Panels A-D and E-H illustrate differences in Chernozem and Luvisol, respectively. Significant differences induced by the treatments are given in each panel as p-values (ANOVA). Significant differences within treatments are marked according to TukeyHSD as following: p<0.001***, p<0.01** and p<0.05*.

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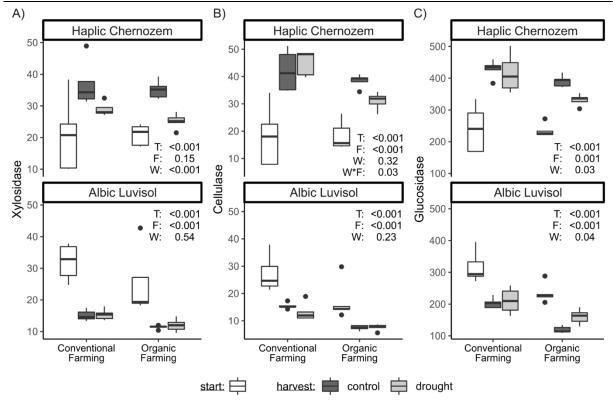


Figure S 4 Activity potentials of C-cycling enzymes in the rhizosphere of wheat. A) xylosidase, B) cellulase and C) glucosidase activities are presented for the wheat cultivar "RGT-Reform". Significant impacts of experimental factors were calculated using ANOVA and are given as p-values for each enzyme and soil type. Thereby, sampling time effect (T) was tested between start and harvest, while watering (W) and farming (F) effects as well as interaction of the two factors were evaluated at harvest.

CHAPTER 2

Can We Estimate Functionality of Soil Microbial Communities from Structure-Derived Predictions? A Reality Test in Agricultural Soils

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Can We Estimate Functionality of Soil Microbial Communities from Structure-Derived Predictions? A Reality Test in Agricultural Soils

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ABSTRACT Computational approaches that link bacterial 16S rRNA gene amplicon data to functional genes based on prokaryotic reference genomes have emerged. This study aims to validate or refute the applicability of the functional gene predic- tion tools for assessment and comparison of community functionality among experi- mental treatments, inducing either fast or slow responses in rhizosphere microbial community composition and function. Rhizosphere samples of wheat and barley were collected in two consecutive years at active and mature growth phases from organic and conventional farming plots with ambient or future-climate treatments of the Global Change Experimental Facility. Bacterial community composition was determined by 16S rRNA gene amplicon sequencing, and the activities of five extrac- ellular enzymes involved in carbon (b-glucosidases, cellobiohydrolase, and xylosi- dase), nitrogen (Nacetylglucosaminidase), and phosphorus (acid phosphatase) cycles were determined. Structural community data were used to predict functional pat-terns of the rhizosphere communities using Tax4Fun and PanFP. Subsequently, the predictions were compared with the measured activities. Despite the fact that differ- ent treatments mainly drove either community composition (plant growth phase) or measured enzyme activities (farming system), the predictions mirrored patterns in the treatments in a qualitative but not quantitative way. Most of the discrepancies between measured and predicted values resulted from plant growth stages (fast community response), followed by farming management and climate (slower com- munity response). Thus, our results suggest the applicability of the prediction tools for comparative investigations of soil community functionality in less-dynamic envi- ronmental systems.

IMPORTANCE Linking soil microbial community structure to its functionality, which is important for maintaining health and services of an ecosystem, is still challenging. Besides great advances in structural community analysis, functional equivalents, such as metagenomics and metatranscriptomics, are still time and cost intensive. Recent computational approaches (Tax4Fun and PanFP) aim to predict functions from structural community data based on reference genomes. Although the usability of these tools has been confirmed with metagenomic data, a comparison between predicted and measured functions is so far missing. Thus, this study comprises an expansive reality test on the performance of these tools under different environmental conditions, including relevant global change factors (land use and climate). The work provides a valuable validation of the applicability of the prediction tools for comparison of soil community functions across different sufficiently established soil ecosystems and suggest their usability to unravel the broad spectrum of functions provided by a given community structure.

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KEYWORDS GCEF, PanFP, Tax4Fun, agriculture, bacterial communities, barley, climate change, enzymes, wheat

ver the last decades, we experienced a rapid advancement of molecular approaches to explore structural diversity of soil microbial communities. The use of next-generation amplicon sequencing allows for high-resolution analyses of microbial community structure, e.g., on its temporal dynamics and adaptation to different environmental conditions (1, 2). Corresponding studies revealed that soil microbial communities change over the growing season (3, 4) and are dependent on the plant species (5–8) as well as on the plant development stage (8–11). Moreover, soil type as well as land-use- and management-related variations in pH and available nutrient concentrations shape soil microbial communities (12–15). However, it remains challenging to determine the functional traits of a given microbial community in order to estimate resultant soil processes and ecosystem services (16). This is because soil processes and functions can be maintained in spite of community shifts by functional redundancy (reviewed in reference 17), while others may be lost by losing individual, possibly even low-abundant, key species (18). Thus, it is crucial to have information on the traits of all present taxa to derive the functionality of the whole community (19–21).

Cultivation has been traditionally used to cross-examine the taxonomic and functional properties of bacteria. Even though the longstanding "1% cultivability paradigm" has been questioned in recent discussions (22, 23), cultivation-based approaches are hardly meaningful for functional trait assessment in environmental samples, since trait variation is strongly reduced by studying only few isolates (24). When measuring aggregated functional properties of the microbial community, e.g., by analyzing community enzyme activities (25–27) or gene expression profiles (28, 29), it is often difficult to assign activities to certain taxa. Available methods that link structural and functional information of bacteria include stable isotope labeling of substrates and subsequent amplicon sequencing of isotope-enriched DNA or RNA (28, 30) or using genome-resolved metagenomics or metatranscriptomics (31, 32). Nevertheless, capturing the functional diversity of whole microbial communities in depth and breadth with these methods remains cost and time intensive (32).

Computational prediction tools in microbial ecology, such as Tax4Fun and PanFP, offer the possibility to translate structural community data into ecosystem functions in a cost-effective way (33–35). These approaches use the link between bacterial 16S rRNA gene amplicon sequencing and functional gene annotations of prokaryotic reference genomes. As output, the programs provide abundance estimates of functional genes. The applicability of both tools has been validated by comparison of the predicted functional gene abundance with the number of detected genes in the respective metagenome (36–38). Median Spearman rank correlation coefficients range up to

0.87 for Tax4Fun (36) and 0.80 for PanFP (37), suggesting good approximations of functional profiles. At the same time, assessment of whether and how well such predicted functional profiles mirror microbial community trait expression and thus allow estimation of ecosystem processes is still missing.

In this study, the activity potentials of five extracellular microbial enzymes (*b*-gluco-sidase, cellobiohydrolase, xylosidase, *N*-acetylglucosaminidase, and acid phosphatase) were measured and compared to the abundances of the respective genes predicted with PanFP and Tax4Fun based on Illumina MiSeq amplicon sequencing data. These enzymes were chosen because (i) they play a crucial role in soil C, N, and P cycling, (ii) their activities are commonly measured in environmental studies as representative of soil function, and (iii) the protocols and assays for activity determination are standar-dized and well established.

We expected that a linear link might not be conceivable, since a direct correlation of gene abundance and its related function would require that (i) the genes of interest are constitutively transcribed to mRNA, (ii) the mRNAs are translated into proteins, (iii) all proteins responsible for the same reaction have the same kinetics and optimal conditions for

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activity, and (iv) all enzymes have the same life span (28). In reality, gene expression and enzyme secretion are not consistent but are regulated in response to soil conditions. Moreover, the life span of extracellular enzymes in soil can range from hours to months, depending on local biotic and abiotic soil parameters (28, 39). Thus, enzyme activity measurements depict the situation in soil at a certain point in time but do not necessarily reflect short-term changes in microbial community composition (39, 40). Nevertheless, we assumed that patterns of measured enzyme activities follow those of the corresponding functional gene abundances in the microbial communities along treatments or environmental gradients that exert a continuous and steady impact.

This study aimed to evaluate this assumption and to validate or refute the applicability of the functional gene prediction tools for assessment and comparison of soil processes. Since the functional predictions by Tax4Fun and PanFP refer exclusively to bacterial genomes, we selected croplands as study systems. Agricultural soils are usually dominated by bacteria, whereby the fungal contribution to enzyme profiles is minimized (41, 42). We collected rhizosphere soils, i.e., the hot spot for abundance, activity, and turnover of soil bacteria (43), of wheat (and barley in the subsequent year) from agricultural plots of the Global Change Experimental Facility (GCEF) (44). This experimental field platform cross-manipulates climatic conditions (ambient versus future) and farming management (conventional versus organic farming). Both experimental treatments are known to steer structure and function of bacterial communities (12-14). The small but continuous impact of changed climatic conditions induces a slow response of the soil community, whereas the adaptation to different management measures induces quicker community responses. To account for very rapid responses, we collected rhizosphere samples at two different plant growth phases; active biomass production and mature phase. The dynamics of roots from active to mature growth stages (45, 46) are known to cause rapid temporal changes in rhizobac- terial community structures (reviewed in reference 47). From all collected samples, we determined the rhizobacterial community composition using 16S rRNA gene amplicon sequencing, estimated the functional gene abundances by the prediction tools Tax4Fun and PanFP, and measured the enzyme activity potentials.

We hypothesized that (i) deviations between the predicted traits and the measured enzyme activities show a positive correlation with the speed of the community's response to the treatments. Thus, the strongest deviations should be related to plant growth phases (strong dynamics, rapid adaptation) followed by the impact of the farming management, while the most concurrent patterns should be observed along the climate treatments (slow, consistent community adaptation). We further hypothesized that (ii) across growth phases, the deviations are more pronounced during the mature growth phase. Plants stimulate rhizobacterial growth and activity by a gradually increasing release of rhizodeposits during active growth but strongly reduce rhizodeposition when reaching maturity, inducing a reduction of bacterial biomass (9). Accordingly, functional gene abundance drops quickly, while there is a delay for enzyme activity. We also hypothesized that (iii) deviations in measured and predicted values are more pronounced in conventional farming soil, as these systems experience more disturbances by, e.g., pesticide application, compared to that of organic soil. Finally, we hypothesized that (iv) under future climatic conditions, with larger variability of annual precipitation, the deviations between measured and predicted functions are more pronounced than under ambient climatic conditions.

RESULTS

Experimental treatments drive rhizosphere community composition and enzyme activities. (i) Impact on the bacterial community composition. The effect of the experimental treatments and the related differences in abiotic soil parameters (for more information on edaphic parameters and impact on community composition refer to Material S1 and S2, respectively, in the supplemental material) on the rhizobacterial community composition was studied for both crops in the two consecutive years of cultivation (Fig. 1). In the first year, when wheat was cultivated, growth phase was the main driver for bacterial community composition (permutational multivariate analysis

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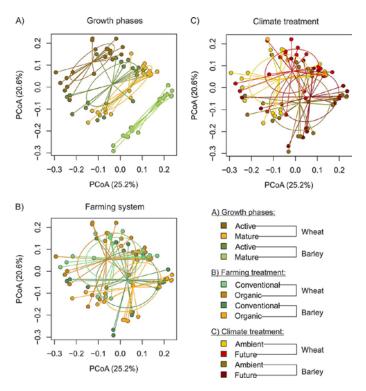


FIG 1 Principal coordinates analysis for beta-diversity of bacterial rhizosphere communities. The points are colored and circled according to growth phases (A), farming system (B), and climate treatment (C).

of variance [PERMANOVA], R2 = 0.26, P , 0.001) (Fig. 1A), followed by farming system (PERMANOVA, R2 = 0.08, P = 0.003) (Fig. 1B) and climate treatment (PERMANOVA, R2 = 0.03, P = 0.18) (Fig. 1C). In a comparable way, rhizobacterial community composition of barley in the subsequent year was affected in decreasing order by growth phase (PERMANOVA, R2 = 0.22, P , 0.001) (Fig. 1A), farming system (PERMANOVA, R2 = 0.04, P = 0.08) (Fig. 1B), and climate treatment (PERMANOVA, R2 = 0.03, P = 0.33) (Fig. 1C). In line with these results, analysis and visualization of indicator species in a bipartite net- work indicated a strong grouping of species according to the growth phase and farm- ing system in the wheat rhizosphere as well as according to the growth phase in the barley rhizosphere (Material S3). Besides, wheat and barley strongly differed in their rhizobacterial community composition (PERMANOVA, R2 = 0.13, P , 0.001).

(ii) Impact on rhizosphere enzyme activities. Farming system and the related differences in edaphic parameters (Material S1) were the main drivers of enzyme activities (Table 1; Material S2). Thereby, higher enzyme activities were found in rhizosphere soil from conventional farming than in the ones from organic farming, which was evident for wheat at both growth phases, while for barley, it was mainly observed in the active growth phase (see the blue boxes in Fig. 2). The effects of the growth phase and of the climate treatment on rhizosphere enzyme activities were comparably weak, with signif- icant impacts of individual extracellular enzymes and in a crop-specific manner (Table 1). The growth phase affected chitinase activity in wheat (active , mature) (Fig. 2C) and acid phosphatase activity in barley rhizosphere (active . mature) (Fig. 2D) (Table 1). Climate treatment effects were found for the activities of xylosidases and acid phosphatases in the rhizosphere of wheat and for the activity of cellulases in the rhizosphere of barley (Table 1). Besides, all enzyme activities strongly differed (P , 0.001) between wheat and barley rhizospheres, with higher enzyme activities in the wheat rhizosphere (Fig. 2).

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	P value ^b			
Enzyme	Farming	Growth phase	Climate	Growth phase $ imes$ farming
Wheat				
Glucosidases	< 0.001	0.76	0.11	0.57
Xylosidases	0.007	0.09	0.04	0.38
Chitinases	<0.001	0.002	0.41	0.19
Phosphatases	<0.001	0.19	0.04	0.21
Cellulases	<0.001	0.84	0.77	0.91
Barley				
Glucosidases	< 0.001	0.71	0.32	0.007
Xylosidases	< 0.001	0.15	0.67	0.03
Chitinases	0.02	0.24	0.19	0.14
Phosphatases	< 0.001	,0.001	0.79	<0.001
Cellulases	< 0.001	0.85	0.04	0.05

^aActivities of *b*-glucosidases, xylosidases, *N*-acetylglucosaminidases (chitinases), acid phosphatases, and cellobiohydrolases (cellulases) were tested against the factors farming system, growth phase, climate, and interaction of farming system and growth phase.

Patterns of measured enzyme activities compared to predicted enzyme gene abundances in the rhizosphere. (i) Correlations along and relative changes between factors of growth phase, farming system, and climate. Spearman rank correlations were tested to identify common and specific patterns of predicted functional gene abundances and measured enzyme activities over both growth phases (n=40) (Table S1). In the wheat rhizosphere, a positive correlation between measured and Tax4Fun- as well as PanFP-predicted values was found for xylosidases and with a trend observed for acid phosphatases. Contrary to that, in the rhizosphere of barley, functional gene abundances predicted by Tax4Fun were found to be positively correlated with the measured enzyme activities of glucosidases, xylosidases, chitinases, and cellulases. Regarding PanFP predictions, significant correlations with the measured activities were indicated for xylosidases and chitinases and with a trend observed also for glucosidases and cellulases.

Assessing the two growth phases separately, the significance level of correlations was commonly reduced, mainly due to the lower number of samples (n = 20). Nevertheless, we

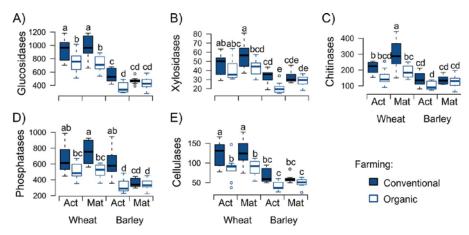


FIG 2 Impacts of farming system, crop species, and crop growth phase on measured enzyme activities (nmol g soil 21 h 21). b-glucosidases (A), xylosidases (B), N-acetylglucosaminidases (chitinases) (C), acid phosphatases (D), and cellobiohydrolases (cellulases) (E). Measured enzyme activities at the active (Act) and mature (Mat) growth phases in conventional and organic farming soils are given. Different lowercase letters within each panel indicate significant differences between the treatments (P, 0.05) according to Tukey's HSD.

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^bSignificant impacts according to ANOVA are indicated by italic font.

TABLE 2 Relative changes between factors of growth phase (active versus mature), farming system (conventional versus organic), and climate (ambient versus future) treatment^o

	Relative	change							
	Active	smature		Conventi	onal vs organic		Ambient	vs future	
Enzyme	Tax	Pan	Enzymes	Tax	Pan	Enzymes	Tax	Pan	Enzymes
Wheat									
Glucosidase	0.27	0.53	0.02	-0.23	-0.23	-0.23	-0.20	-0.18	-0.08
Xylosidase	0.52	0.55	0.14	-0.25	-0.24	-0.19	-0.24	-0.18	-0.14
Chitinase	0.15	0.36	0.31	-0.27	-0.23	-0.32	-0.17	-0.21	-0.07
Phosphatase	0.89	0.59	0.09	-0.26	-0.25	-0.28	-0.27	-0.22	-0.13
Cellulase	0.84	0.40	0.02	-0.22	0.02	-0.31	-0.30	0.01	-0.02
Barley									
Glucosidase	0.60	1.15	-0.02	-0.28	-0.22	-0.22	0.00	0.02	0.06
Xylosidase	0.73	1.30	0.12	-0.26	-0.19	-0.25	-0.03	0.01	0.03
Chitinase	0.16	1.12	0.11	-0.33	-0.20	-0.19	-0.06	-0.02	0.12
Phosphatase	1.27	1.19	-0.26	-0.25	-0.22	-0.31	0.00	0.02	0.02
Cellulase	1.03	0.77	0.01	-0.26	-0.23	-0.28	0.00	-0.02	0.16

"Relative changes are given for predicted gene abundances of Tax4Fun (Tax) and PanFP (Pan), as well as for measured enzyme activities in the rhizosphere of wheat and barley.

observed stronger positive correlations between the predictions and the measured activity for xylosidases and acid phosphatases in the wheat rhizosphere at maturity than at the active growth phase. In contrast, for barley, these correlations were stronger at the active growth phase than at crop maturity. Furthermore, the predictions by Tax4Fun at the active growth phase of barley, as well as the predictions by PanFP at both growth phases of barley were highly positively correlated with the measured phosphatase activities (Table S1). Relative differences between active and mature growth phases, conventional and organic farming, and ambient and future climate for predicted and measured values are presented in Table 2. Thereby, relative differences of predicted functional gene abundances mostly mirrored the measured enzyme activities in a qualitative way, i.e., in terms of the direction (positive, negative, or no difference). Exceptions were found for acid phosphatases and, to a lower extent, also for glucosidases in the rhizosphere of barley. While predictions indicated higher gene abundances of the respective functional genes at the active phase than at the mature growth phase, measured enzyme activities showed an opposing pattern (Table 2).

(ii) Concordance and discordance between the measured and predicted values. To be able to quantitatively compare the patterns of predicted gene abundances and measured enzyme activities, z-transformed data were used.

The degrees of over- or underestimations of functions varied between the tested experimental conditions (Fig. 3 and Fig. S1). Regarding the wheat rhizosphere, measured enzyme activities were mostly underestimated by the predictions at the active growth phase (Fig. S1), whereby the strongest deviations were found for conventional farming (CF) under ambient-climate conditions (Fig. 3) (Tax4Fun, standard deviation [SD] = 0.51; PanFP, SD = 0.63) and organic farming (OF) under future climatic conditions (Fig. 3) (Tax4Fun, SD = 0.65; PanFP, SD = 0.54). At crop maturity, we found strong concordances between activities and predictions (Fig. S1). They were particularly strong in OF under ambient-climate conditions (Fig. 3) (SD , 0.20 for both prediction tools) as well as in CF in an ambient climate (Fig. 3) (Tax4Fun, SD = 0.23; PanFP, SD = 0.36). Under future-climate conditions, the activities were underestimated in CF and overestimated in OF (Fig. 3). Overall, we found a better fit between measured activities and predictions in OF than in CF (Fig. S1) and an overall good fit for future- and ambient-climate treatments (Fig. S1).

For the barley rhizosphere, a clear pattern emerged with almost perfect fits of measured activities and predicted gene abundances at the active growth phase (Fig. 3) (SD , 0.2 for both prediction tools) (Fig. S1, red lines indicate zero deviations). The strongest deviations occurred in CF under future-climate conditions, when mainly phosphatase

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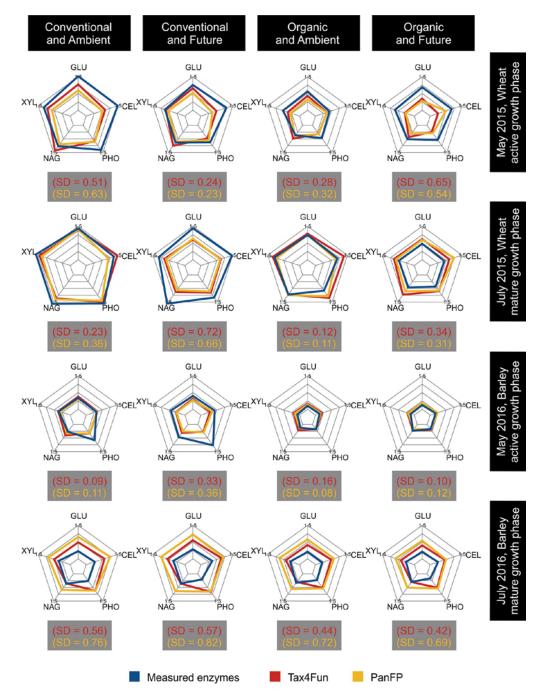


FIG 3 Measured enzyme activities and predicted functional gene abundances of Tax4Fun and PanFP, arranged by growth phases of the two crops (horizontal) and by the experimental treatment (vertical). Data were normalized by z-transformation. The spider charts represent the measured enzyme activity levels (blue), and gene abundance levels estimated by Tax4Fun (red) and PanFP (yellow). Higher values are more distant from the center of the web. The median of the standard deviations between measured and predicted values for all enzymes is given for both prediction tools separately in brackets (SD). GLU, b-glucosidases; XYL, xylosidases; NAG, N- acetylglucosaminidases (chitinases); PHO, acid phosphatases; CEL, cellobiohydrolases (cellulases).



TABLE 3 Significance of deviations in measured to predicted values with respect to the experimental factors^a

	P value ^b								
	Growth pha	ise		Farming	system		Climate		
Enzyme	Both	Wheat	Barley	Both	Wheat	Barley	Both	Wheat	Barley
Tax4Fun vs enzymes									
Glucose	0.01	0.25	< 0.001	0.44	0.41	0.87	0.52	0.63	0.52
Xylosidase	0.06	0.31	0.02	0.39	0.71	0.20	0.79	0.96	0.58
Chitinase	0.21	0.24	0.67	0.22	0.25	0.72	0.40	0.73	0.13
Phosphatase	< 0.001	0.04	< 0.001	0.05	0.21	0.05	0.71	0.73	0.84
Cellulase	< 0.001	0.02	< 0.001	0.15	0.16	0.54	0.10	0.18	0.20
PanFP vs enzymes									
Glucose	< 0.001	0.03	< 0.001	0.40	0.34	0.76	0.71	0.77	0.70
Xylosidase	0.003	0.26	< 0.001	0.39	0.64	0.28	0.87	0.72	0.81
Chitinase	0.18	0.72	< 0.001	0.24	0.20	0.88	0.32	0.52	0.27
Phosphatase	< 0.001	0.15	< 0.001	0.06	0.24	0.05	0.90	0.92	0.92
Cellulase	0.001	0.05	< 0.001	0.05	0.00	0.90	0.74	0.80	0.37

The deviation between z-transformed values of measured and predicted indices for the activity of five enzymes was calculated and tested for significance across growth phases, farming systems, and climate treatments.

activities were heavily underestimated by the predictions (Tax4Fun, SD = 0.33; PanFP, SD = 0.36). At the mature growth phase of barley, enzyme activities were overestimated by the predictions (Fig. 3 and Fig. S1), except for chitinases, where Tax4Fun predictions perfectly matched measured activities (Fig. 3). Overall, we observed less deviations of the predicted from the measured values in OF than in CF (Fig. S1), while the deviations were similar across the two climate treatments (Fig. S1).

Testing the effect of different treatments on deviations, growth phase definitely exerted the strongest impact on the concordance between predicted and measured enzyme indices for individual enzymes in the wheat rhizosphere and for all of them in the barley rhizo-sphere (Table 3). Thereby, variance partitioning revealed that 22% of the total variations in the deviations can be explained by this treatment. In contrast, the experimental factors farming system and climate did not significantly affect the accuracy of the predictions (Table 3) (analysis of variance [ANOVA], P = 0.05; variance partitioning, 0.9% and 0.1%), and P values were smaller for farming system than for climate (Table 3). The decreasing differences in deviations of predictions from measured values, growth phase = farming system = climate treatment, was more obvious in the barley than in the wheat rhizosphere (Table 3).

DISCUSSION

Structure and function of microbial communities in agroecosystems are affected by different drivers. Plant growth phase was the most prominent driver of community composition in the rhizospheres of wheat and barley. This result agrees with findings of Houlden et al. (48) and Francioli et al. (3), who demonstrated a strong shift in rhizobacterial community composition according to plant growth stages in crop plants such as wheat, pea, and sugar beet (48) but also in grassland species (3). Such differences can be explained by quantitative and qualitative changes in rhizodeposition that are related to the different plant development phases (46, 49).

In contrast, measured enzyme activities were mainly driven by farming practice and associated differences in mineral nitrogen and total C and N contents (see results in Material S1 in the supplemental material). Farming practice is an important driver of soil enzyme activity, which drastically changes soil structure (50, 51) and soil chemical parameters (25). We found higher enzyme activities in CF than in OF. In line with that, Arcand and colleagues (52) observed increased activity and production of enzymes in conventional farming soil compared to that in organic farming soil. The pattern may be caused by a higher availability of nitrogen in CF, which is known to foster the production and activity of polysaccharide-degrading enzymes (53–55).

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^bThe P values are given according to ANOVA for the total data set (both wheat and barley, n = 80) as well as separately for wheat (n = 40) and barley (n = 40). Significant impacts according to ANOVA are indicated by italic font.

Our results revealed that shifts in composition and functions of the rhizobacterial com- munity are caused by different drivers, indicating a decoupling of community composition and function. In concordance, Francioli et al. (56) found that mineral and organic fertilizers mainly affect either activity or composition of the microbiome in an agricultural soil. Additionally, Bowles et al. (57) indicated that structurally highly similar bacterial communities can show very contrasting enzyme activities in differently managed organic fields.

Measured enzyme activities partly confirmed by predictions. The predicted gene abundances in our study responded to the drivers of both community composition and activity and were thus affected by the growth stage and by the farming system (Table S2, Fig. S2). In line with the first hypothesis, the performance of the prediction tools was driver dependent, whereby the strongest deviations could be related to the crop growth phase, followed by farming system and climate treatment. Furthermore, a remarkable impact of the crop species on the level of concordance between the predicted and measured activities was indicated.

For barley, in accordance with hypothesis two, strong correlations and concordances between measured activities and predicted gene abundances of both tools were observed at the active growth phase. In contrast, the enzyme activities were overestimated by both prediction tools at the mature growth phase. This finding is likely based on a faster response of the community composition than of enzyme activity. To promote plant growth, plants exudate carbon compounds into their rhizosphere that stimulate growth and activity of soil microorganisms (58 and reviewed in reference 59). Depending on plant development stage, root exudation patterns differ and thus strongly influence the rhizobacterial community (9, 46). Root exudation of carbon-rich compounds (sugar) is at its strongest in the juvenile growth phase, represented by the active growth phase of this experiment, and decreases thereafter (46). The overestimation of activity by the prediction tools at the mature growth phase of barley may therefore be a result of accumulated, mostly inactive rhizobacterial genes.

Deviation patterns of measured and predicted activities were more heterogeneous in the wheat rhizosphere. Three possible reasons may explain the discordance between predictions and measured activities:

- (i) A major impact on enzyme activities was attributed to soil mineral N concentrations, which were positively correlated (Material S2) and significantly higher in the rhi- zosphere of wheat than in that of barley (Material S1). This relation has already been demonstrated for activities of cellulase and *b*-glucosidase (54, 60) as well as of acid phosphatase (61) and chitinase (62). The high mineral N concentration in the wheat rhizosphere likely fostered enzyme production without microbial growth, especially at the early active growth phase, and may explain deviations between measured activ- ities and predictions.
- (ii) For wheat, we observed an interaction effect of the experimental climate and farming system treatment on enzyme activities, which was not reflected in the predictions. Supporting our fourth hypothesis, we observed strong deviations under a future-climate condition at the mature growth phase and under organic farming conditions also at the active growth phase of wheat. With conventional farming at the active growth phase of wheat, the pattern of deviations was inverse. As such a pattern was not observed in the barley rhizosphere, this may indicate a plant-specific drought effect on the enzyme activity that could not be mirrored by the prediction tools. When Kosová et al. (63) summarized the knowledge about wheat and barley responses to drought, they stated no strong advantages or disadvantages for either of the two but variations along different genotypes. It is therefore likely that the cultivated wheat ge- notype expressed a different adaption capacity to drought than the barley genotype, which may feedback to structure and functionality of the rhizosphere community and thus functional predictions.
- (iii) Indicator species analysis and a bipartite network indicated a clear effect of the growth phase for barley, which was also the main driver for community composition (Material S3). In contrast, the impact of wheat growth phase on indicator species was

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surpassed by farming system, as we found high numbers of shared indicator species between active and mature growth phases under conventional farming conditions. Thus, the contradicting drivers for overall community and indicator species in the wheat rhizosphere may contribute to deviations between predictions and measured activities.

Interestingly, and against our assumptions of hypothesis three, farming system-related deviations between measured and predicted values were not obvious either in the wheat or in the barley rhizosphere. An explanation could be the normalization of the predicted functional gene abundances with 16S rRNA gene abundances, which were strongly biased by farming system. Another possibility would be that the effect of farming system was outcompeted by the overall stronger effect of growth phases. Besides the deviations between measured and predicted values, the correlations along the treatments and relative changes between factors of treatments were not affected, suggesting that the prediction tools mirrored the impacts of the experimental factors in a qualitative way.

Limitations. (i) Prediction tools. Although the applied tools were created for universal use, their predictive power depends on the quality of the databases. Another commonly used prediction tool, PiCrust (phylogenetic investigation of communities by reconstruction of unobserved states [64]), is tailored to functional predictions in the human microbiome by using the Integrated Microbial Genomes database (65) containing genomes from the Human Microbiome Project (The Human Microbiome Jumpstart Reference Strains Consortium [66]) and Greengenes database (67). For soil microbiomes, Tax4Fun (36) and PanFP (37) outperform predictions of PiCrust (36, 37). These tools rely on the bacterial sequences of the SILVA database (68) which comprises quality-controlled aligned rRNA gene sequences. Nevertheless, all tools are subject to some restrictions which have to be considered for analysis.

First, the contribution of other organisms, including plants and fungi, to the extracellular enzyme production cannot be estimated by the tools. While extracellular enzymes in the soil are mainly attributed to origination from edaphic microorganisms and the contribution of plants may be thus negligible (39, 69), the contribution of fungal communities is of considerable importance, especially in extensively managed grassland and forest ecosystems (70). For comparison of predicted and measured activities, we therefore performed our study in agricultural systems which are known to be dominated by bacterial communities (41, 42).

Second, the predictive power of the tools relies on the integrity of the databases. SILVA (68) and KEGG Orthology (KO) (71) databases facilitate annotation of bacteria preferably to the genus level (36), thereby loosing information about functional differences on species level. Aßhauer et al. (36) further noticed that the members of the highly diverse soil community are not sufficiently represented in the KEGG database. Since publication of the prediction tools in 2015, the SILVA and KO databases have been updated frequently. Notwithstanding, the tools do not implement the latest versions of KO and SILVA databases. While PanFP is based on SILVA v128 (released in 2017), Tax4Fun supports only SILVA v123 (released in 2015). The KO database has been licensed and, thus, allows only free access to version 64.0 released in October 2012. In our analysis, we used the latest applicable versions for Tax4Fun and PanFP analysis. For Tax4Fun predictions, we were able to trace back the percentage of bacterial sequences used for predictions and the distribution of KO identifiers representing the five enzymatic classes among samples (Fig. S3). Only 37% and 32% of bacterial sequences could be used for the functional predictions in the rhizospheres of wheat and barley, respectively. Regarding indicator species, the numbers improved to at least 43% for wheat and 42% for barley. It would, therefore, be desirable to integrate updated databases and more recent metagenomic data to improve predictions.

Third, the expected discrepancy between gene presence and expression is a major constraint. Next-generation sequencing does not discriminate between the fractions of living and dead cells. Furthermore, in a given soil, only a certain proportion of

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microorganisms are active at a certain time point (reviewed in references 72). The identity and number of active or inactive taxa depend on external conditions and stimuli (73). In our study, we specifically investigated the impact of land use/management, climate, and plant growth. Our results imply that the various proportions of active taxa and the contributions of dead organisms may partly explain discrepancy between measured and predicted activities and may be the main reason for the missing quantitative concordance.

(ii) Evaluation of enzyme activities as functional indicators. Besides database-related deficiencies, discrepancies in the comparisons may also emerge due to the approach used to determine enzyme activities. The measurement of soil enzyme activ- ities cannot distinguish between recently secreted enzymes and enzymes that were produced earlier by taxa whose relative abundance may have declined. These so-called abiontic enzymes are protected against their degradation in clay complexes (39). The persistence and accumulation of enzymes in the soil matrix may strongly influence overall enzyme activity measurement. The amount of immobilized enzymes strongly depends on the soil type and respective clay and organic matter content (74). We found a strong relationship between 16S rRNA gene abundances and enzyme activities at the active growth phase of both crops but a decrease of enzyme activity accompa- nied by an increase of gene abundances in the mature growth phase. This indicates a minor role of abiontic enzymes for the measured activities and, rather, suggests differ- ent root exudation rates at active and mature growth phases as the main driver of enzyme secretion.

A second methodological aspect that should be considered is that we determined enzyme activities under standardized conditions which are close to the optima of the different enzymes (pH 5, 25°C). Moreover, the high substrate concentration in the assay (300 mM) ensured no limitation due to substrate availability (75). While the used temperature (25°C) represents reasonable daytime temperatures at the sampling dates in late May and July, the used pH was much lower than the average from all samples (pH 6.3). Since we did not compare absolute values but only differential expression of enzymes, and the pH was comparable between all treatments, differences along the treatments should be maintained. In contrast, farming system-specific substrate availability may result in different enzyme patterns than those obtained under saturating substrate concentrations. These methodological issues plead for further studies measuring enzyme activities at realistic temperature and pH as well as particularly under substrate concentrations reflecting the availability in the respective system.

Conclusions and perspectives. Our results demonstrate that Tax4Fun and PanFP provide cost-effective tools to estimate functional patterns of rhizobacterial communities in a qualitative (i.e., direction of response) but not in a quantitative (i.e., extent of response) way. The response of the studied activities to experimental drivers was predominantly predicted correctly by both tools in terms of direction, i.e., increase or decrease. This is particularly noteworthy, since drivers of community structure and measured activities differed from each other. Moreover, we observed a gradual decrease in predictability the faster the treatments acted on community structure. The tools do not provide a one-size-fits-all solution, and interpretation of predicted functions has to be performed thoroughly. To trace mechanisms behind concordances and discordances, a deeper understanding of the underlying drivers of functions and structure is necessary. We emphasize the importance of more studies on predicted and measured functional traits to explore relevant drivers for functions and structure in different environments. These studies should be combined with transcriptomics data to explore whether the link between predicted and real soil processes will be strengthened, as assumed for this study, or, rather, diminished.

Nevertheless, our finding provides a valuable validation of the applicability and robustness of these prediction tools for comparison of soil community functions across stable soil ecosystems. While the enzymes used in this approach address solely activities related to nutrient cycling, the positive validation would plead for further research

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on the possibility to predict other, more intractable functions of microbial communities in a simple and quick way.

MATERIALS AND METHODS

Soil sampling and sample preparation. Samples were obtained from the Global Change Experimental Facility (GCEF) situated at the research station in Bad Lauchstädt. Central Germany (51°23'35"N 11°52'55"E. 118 m above sea level [a.s.l.]). The site is characterized by a temperate climate with an average temperature of 9.7°C (1993 to 2013) and a mean annual precipitation of 525 mm (1993 to 2013). The soil type is a fertile loamy soil (haplic chernozem) (76). The experimental platform of the GCEF was established in 2013 and combines land use and climate treatments as described by Schädler et al. (44). In our study, we focused on the cultivated cropland systems, organic farming (OF; 10 plots) and conventional farming (CF; 10 plots). The grown crop was identical for CF and OF in 2015 (winter wheat) and 2016 (winter barley) but differed in 2014 (CF, winter rape seed; OF, field bean). In CF, synthetic fertilizers (N, P, and K), growth regulators, and pesticides are applied. The use of pesticides in OF is restricted, and fertilization is realized by including legumes in the crop rotation as well as by the application of rock phosphate (P-Ca-Mg) and patent kali (K-Mg-S) every 3 years. Half of the plots experience ambient climate (A), while the other half is exerted to simulated future-climate conditions (F) comprising a warming (+0.55°C on average) and a changed precipitation pattern (-20% in summer, +10% in spring and fall) (for details refer to Schädler et al. [44]). Cereal roots with closely adherent soil were sampled in the active cereal growth phase in May and at the mature state in July in 2015 (3 wheat plants per plot, 20 plots) and 2016 (3 barley plants, 20 plots). Samples were transported in cooling boxes to the field station and immediately frozen. Simultaneous to root sampling, surrounding bulk soil was sampled for the analysis of soil parameters. For this, six soil cores (diameter [Ø] 15 mm, 0- to 15-cm depth) were taken from each plot, pooled, sieved to 2 mm, manually cleaned from organic material, and frozen at -20°C.

To separate rhizosphere soil from roots, the roots with adherent soil were crushed and put in 50-ml Falcon tubes with 40 ml of 0.5% NaCl solution. Tubes were vortexed for 1 min to loosen adherent soil from the roots. Subsequently, roots were transferred to a second set of 50-ml Falcon tubes. Soil suspensions without roots were centrifuged at $12,851 \times g$ for 10 min. Then, the supernatants were filled into the tubes with the roots and used for a second washing step. After vortexing, soil suspensions were transferred into the Falcon tubes with the pellets from the first washing step. The procedure of washing and centrifuging was repeated three times. Rhizosphere soil pellets were frozen at -20° C.

Soil parameters. Since the amount of rhizosphere soil was limited to 2 to 3 g per sample, basic soil parameters were determined using respective bulk soil. For pH analysis, 12 g of air-dried soil was suspended in 30 ml of 0.01 M CaCl2 solution (1:2.5 [wt/vol]). The soil suspension was equilibrated at room temperature and thoroughly mixed every 20 min. After 1 h, the pH was measured with a pH electrode. Total carbon and nitrogen contents were determined from air-dried soil using an elemental analyser (Elementar Vario EL III; Elementar, Hanau, Germany). For analysis of mineral nitrogen, 5 g of fresh soil was suspended in 20 ml of 1 M KCl solution and measured via flow injection analysis (FlAstar 5000; Foss GmbH, Rellingen, Germany). Available phosphorus was extracted from fresh soil with double lactate (1:50 [wt/vol], pH 3.6) and quantified using the colorimetrical molybdenum blue method (77).

Soil enzymes. The activity potentials of hydrolytic soil enzymes were measured using a modified fluorometric assay introduced by Sinsabaugh et al. (78). The analyzed enzymes are involved in phosphorus acquisition (phosphatases), nitrogen acquisition (*N*-acetylglucosaminidases), and carbon acquisition (*b*-glucosidases, xylosidases, and cellobiohydrolases). Enzymatic activities were determined as turnover rate of 4-methylumbelliferon (MUF)-coupled substrates (Table S3 in the supplemental material), where the amount of released fluorescent MUF was directly related to enzymatic activity potentials. To avoid underestimation of enzyme activities (79), the substrate concentration was optimized for the haplic chernozem soil and set to 300 mM for all substrates.

For each sample, a black 96-well microplate was prepared containing all five substrates, MUF dilutions to calculate quench and extinction coefficients (1.25 mM and 2.5 mM), and the substrate and soil suspension controls. For analysis, approximately 250 mg of fresh rhizosphere soil was suspended in 50 ml of 50 mM acetate buffer (pH 5) and sonicated for 5 min to break up soil aggregates. Subsequently, the soil suspension was added to the substrates and incubated at 25°C for 60 min. The enzyme reaction was stopped by the addition of 1 M NaOH solution. After 3 min, fluorescence was measured for eight replicates using an Infinite 200 PRO instrument (Tecan Group Ltd., Männedorf, Switzerland) with 360-nm excitation and 465-nm emission filters. Enzyme activity was calculated as turnover rate of substrate in nanomoles per gram dry soil per hour (nmol g soil 1 h 1) (80).

DNA extraction and next-generation sequencing (Illumina MiSeq). The extraction of soil bacterial genomic DNA was performed using the PowerSoil DNA isolation kit (MO BIO Laboratories Inc., Carlsbad, CA, USA). The protocol was slightly modified by increasing the soil amount from 250 to 400 mg. The concentration of extracted DNA was examined with a NanoDrop ND-8000 spectrophotometer (Thermo Fischer Scientific, Dreieich, Germany), and the DNA was then stored at 220°C. Before running the PCR, the concentrations of DNA extracts were adjusted to 10 to 15 ng/ml. The amplification of the bacterial 16S rRNA gene V4 region was performed with the universal primers 515f and 806r (81), which were equipped with Illumina adapter sequences. To ensure correct amplification of the sequences, all PCRs were performed using proofreading KAPA HiFi polymerase (KAPA Biosystems, Boston, MA, USA). The conditions of the PCR are summarized in Table S4 (PCR 1).

PCR products were tested by gel electrophoresis and purified using the Agencourt AMPure XP kit (Beckmann Coulter, Krefeld, Germany). To assign the sequences to the respective samples, Illumina Nextera XT indices were attached to both ends of the bacterial fragments in a second PCR. The

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conditions of the index PCR are presented in Table S4 (PCR 2). PCR products were purified using AMPure beads, and DNA was quantified with the PicoGreen assay (Molecular Probes, Eugene, OR, USA). For an equimolar representation of each sample, defined volumes of prepared bacterial amplicon libraries (corresponding to 80 ng DNA for each sample) were pooled in one tube. The fragment sizes and the quality of DNA sequencing libraries were again checked with the Agilent 2100 bioanalyzer (Agilent Technologies, Palo Alto, CA, USA). Sample libraries and PhiX control libraries were denatured according to the protocol of the MiSeq v3 reagent kit and diluted to a final concentration of 10 pM. Denatured and diluted libraries were combined to a volume of 600 ml (30 ml of PhiX control library, and 570 ml of bacte- rial amplicon library) and loaded onto MiSeq v3 reagent cartridge for sequencing. Finally, paired-end sequencing of 2 by 300 bp was implemented on an Illumina MiSeq platform (Illumina Inc., San Diego, CA, USA) at the Department of Soil Ecology of the Helmholtz Centre for Environmental Research (UFZ, Halle, Germany).

Bioinformatics workflow, functional predictions, and normalization. In total, 13,912,979 demultiplexed sequencing reads were processed using an in-house pipeline (described in reference 82, with modifications) based on mothur (83) and OBITools (84). In brief, reads without the 515f and 806r primers were discarded, and the primers were clipped from the remaining sequences using cutadapt (85). Read pairs were assembled using PANDAseq (86) and quality trimmed to an average Phred score of 26, retaining 59% of the reads. After preclustering at 99% identity using CD-HIT-454 (87), chimeric reads were removed using the UCHIME algorithm (88), and the remaining reads were clustered into operational taxonomic units (OTUs) at 97% identity using vsearch (89). To safeguard against artifacts, singletons were removed corresponding to 2.5% of the reads. The representative reads of all OTUs were examined for chimeric reads, which were removed using the UCHIME algorithm. Subsequently, the representative reads were taxonomically assigned based on the reference sequences from the SILVA database (version 128, nonredundant at 99% [90]) using the mothur implementation of the naive Bayesian classifier (91). OTUs of chloroplasts and mitochondria and those not assigned to the kingdoms *Bacteria* or *Archaea* were removed. The final sequencing depth per sample (± standard deviation) was 96,000 ± 14,000 reads.

Functional predictions of the bacterial communities were performed with two programs, PanFP (37) and Tax4Fun (36), working on the basis of the OTU abundance and taxonomy data. Tax4Fun (36) and PanFP (37) create functional profiles of bacterial communities using two related approaches for the analvses. Tax4Fun assigns operational taxonomic units (OTUs) to reference sequences in the SILVA database (SILVA database [68]) and converts the counts of SILVA-labeled OTUs to a taxonomic profile of organisms in the KEGG database (71). PanFP creates pangenomes of taxonomically related genomes with their identity also obtained via the SILVA database and subsequently weighs the pangenome's functional profile with OTU abundances. The authors' instructions were followed to run Tax4Fun program line in R (version 3.4.0 [92]) using the latest supported version of SILVA database (SILVA123, released July 2015). Taxonomic assignment was adapted for Tax4Fun analysis to SILVA123. PanFP was executed on a suitably formatted OTU table. The output tables provided KEGG orthology (KO) numbers for gene annotations and Enzyme Commission (EC) number as object identifier for enzymes. Gene abundances of the enzymes of interest were extracted from the output tables of Tax4Fun and PanFP predictions (Table S5). Gene abundances of the three acid phosphatases and the two b-glucosidases, which belong to the same enzymatic class according to EC numbers, were summed up from the Tax4Fun and PanFP output tables, respectively, for further analysis.

Both programs provide abundance estimates of functional genes which are compromised by methodical restraints of 16S rRNA gene sequencing. To allow a balanced reading, the input samples had to be adjusted to a certain DNA concentration, vanishing actual differences between samples. To correct predictions of gene abundances for biomass differences in the samples (36, 37), we estimated bacterial DNA concentrations of rhizosphere samples by quantitative real-time PCR (qPCR) analysis. In conformity with the Illumina sequencing, the reactions were performed with the primer pair 515f and 806r (81) tar- geting the 16S rRNA gene V4 region of the bacterial genomes. All samples were diluted to 2-ng/ml DNA input concentrations, as measured by the PicoGreen assay (Molecular Probes, Eugene, OR, USA), and the dilution factor was recorded. As reference, a dilution series of 0.05, 0.125, 0.5, 1, and 4 ng/ml genomic DNA of a Phyllobacterium isolate from the same soil was prepared to generate a standard curve. Quantitative PCR was run with the Bio-Rad iCycler (Bio-Rad Laboratories GmbH. Munich, Germany) under the conditions listed in Table S4 (qPCR). Measured threshold cycle (CT) values of the rhizosphere samples were related to the standard curve to calculate the mass of bacterial DNA. Subsequently, the obtained DNA concentrations were multiplied with the dilution factors to yield the relative bacterial DNA concen-trations for each sample. The obtained values were used as factors to normalize gene abundances in each sample and are further given as DNA concentrations in micrograms per gram dry soil in Table S6.

Statistics. All analyses were performed with the open-source software R (version 3.4.0, R Core Team). The impacts of farming system, climate, and growth phases were tested separately for OTU data to identify the main drivers of the bacterial community. The factors were then ordered by decreasing impact. PERMANOVA ("adonis" R package vegan) was run, separately for crop species, using the following model: Bray-Curtis dissimilarity (OTU table) ::: growth phase X farming system X climate. PERMANOVA ("adonis," R package vegan) was also performed to analyze the influence of soil abiotic parameters on community structure. The stratification by crop species ensured permutations only within groups of samples belonging to wheat or barley rhizosphere. To visualize significant grouping factors of bacterial community composition, principal-coordinate analysis (PCoA) was performed. For this, absolute abundances of each OTU were normalized to the total read counts in the samples, a Bray-Curtis dissimi- larity matrix was calculated, and the first two axes of the PCoA were plotted.

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Indicator species analysis was performed to identify OTUs that were either specific or shared between wheat and barley among the two growth stages and the two farming systems. According to Hartman et al. (93), we applied two different approaches to test for indicator species using R (version 3.4.0 [92]). The correlation-based approach calculates point-biserial correlation coefficients (R package indicspecies [94]) indicating positive associations of OTUs to one or various conditions of farming system and plant growth phases. Associations were considered significant at a P value of ,0.05. A likelihood ra- tio test evaluated differences in abundances of OTUs between plant growth phases in the two different farming systems (R package edgeR [95]). Differences in abundances were considered significant by a false-discovery rate (FDR)-corrected P value of ,0.05. OTUs, when confirmed by both tests to be significant, were regarded as indicator species and further implemented in bipartite network analysis. The net- work was constructed with the R package igraph (93, 96).

Subsequently, Fisher tests were performed identifying enrichments of phyla within the indicator species compared to the overall community composition in wheat and barley rhizospheres to examine if indicator species were a random subset of the overall community (R package rcompanion [97]).

The measured enzyme activities of glucosidases, xylosidases, chitinases, phosphatases, and cellulases were evaluated using the following linear model: test variable ::: (crop species) x farming system x growth phase x climate.

The crop species was included when analyzing effects across both years/crop species. To test for significant impacts of single factors and for interaction effects, but also for the influence of abiotic soil parameters, an ANOVA was run with the respective models followed by Tukey's honestly significant difference (HSD) $post\ hoc$ tests. Significance levels were classified as highly significant (P, 0.001), strongly significant (P, 0.01), significant (P, 0.05), and tendency (P, 0.1.).

Spearman rank correlation tests were applied to assess correlations of measured enzyme activities with the respective predicted gene abundances within climate and farming system treatments and also within and among the different growth phases. To compare expression levels of measured and predicted enzyme activities, values were z-transformed. The differences between measured and predicted Tax4Fun and PanFP, as well as relative changes between factors of the treatment, growth phase, farming system, and climate, were calculated for each enzyme. Significant deviations within and among experimental factors tested with ANOVA and variance partitioning are given (R package *vegan*).

Data availability Demultiplexed sequences are accessible in the Sequence Read Archive under BioProject accession PRJNA605022. The pipeline used for analysis of metabarcoding raw read libraries is available at https://github.com/lentendu/DeltaMP. Tax4Fun and PanFP are open access, author's descriptions can be found for Tax4Fun under https://tax4fun.gobics.de/ and for PanFP under https://github.com/srjun/PanFP.

SUPPLEMENTAL MATERIAL

Supplemental material is available online only. SUPPLEMENTAL FILE 1, PDF file, 2.1 MB.

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F.B., M.T., T.R., A.H.-B., and C.B. conceived and designed the experiment. T.R. and C.B. performed the field experiments. C.B., T.R., and S.F.M.W. performed the laboratory work. A.H.-B. ran bioinformatics. C.B. and A.H.-B. analyzed data. Results were interpreted by

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M.T., T.R., F.B., A.H.-B., and C.B. M.T., T.R., A.H.-B., and C.B. wrote the manuscript with input from F.B. and S.F.M.W.

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Chapter 2

Supplemental Material:

Can we estimate functionality of soil microbial communities from structure-derived predictions? A reality test in agricultural soils

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Supplementary Tables

Table S 1 Correlations between functional gene abundances predicted by Tax4Fun and PanFP and measured enzyme activities of β-glucosidases, xylosidases, N-acetylglucosaminidases (chitinases), acid phosphatases and cellobiohydrolases (cellulases). The correlation coefficients (Spearman's rho) and respective significance levels were calculated for wheat and barley rhizospheres for combined and separated growth phases. Significant impacts according to ANOVA are indicated by italic p values and significance levels according to Spearman rank correlation test are given as follows: <0.001(***), <0.01(**), <0.05(*) and p<0.1(.).

Spearman rank	rank				Wheat						Barley		
COLLEGATION		Tota	Total (n=40)	activ	active (n=20)	matur	mature (n=20)	Tota	al (n=40)	activ	active (n=20)	matu	mature (n=20)
		Q	p value	Ф	p value	٥	p value	٥	p value	Ф	p value	٥	p value
	Glucosidases	0.22	0.15	0.35	0.12	0.18	0.44	0.44	0.003**	0.55	0.01*	0.41	0.06.
Tax4Fun	Xylosidases	0.45	0.003**	0.41	0.06.	0.44	0.04*	0.41	0.008**	0.57	0.008**	0.16	0.49
VS.	Chitinases	0.19	0.23	0.32	0.16	0.009	0.97	0.41	0.007**	0.40	0.08.	0.39	0.08.
Enzymes	Phosphatases	0.30	0.05.	0.17	0.46	0.50	0.02*	0.08	0.58	0.59	0.006**	0.36	0.11
	Cellulases	0.17	0.28	0.15	0.51	0.28	0.22	0.33	0.03*	0.38	0.09.	0.46	0.04*
	Glucosidases	0.23	0.13	0.34	0.14	0.29	0.20	0.29	0.06.	0.51	0.02*	0.46	0.04*
	Xylosidases	0.45	0.003**	0.34	0.13	0.53	0.01*	0.33	0.03*	0.42	0.06.	0.24	0.29
Enzymes	Chitinases	0.20	0.21	0.18	0.43	0.09	0.67	0.41	0.007**	0.38	0.09.	0.38	0.09.
,	Phosphatases	0.27	0.08.	0.17	0.44	0.41	0.07.	0.14	0.36	0.60	0.005**	0.45	0.04*
	Cellulases	0.05	0.71	0.33	0.15	-0.08	0.70	0.27	0.08.	0.33	0.15	0.27	0.24

Table S 2 Drivers of Tax4Fun- and PanFP-predicted functional gene abundances. Abundances of genes encoding for β -glucosidases, xylosidases, N-acetylglucosaminidases (chitinases), acid phosphatases and cellobiohydrolases (cellulases) were tested against the factors farming system, growth phase, climate and interaction of farming system and growth phase. Significant impacts according to ANOVA are indicated by italic p values and significance levels are given as follows: p <0.001(***), p <0.01(**), p < 0.05(*) and p<0.1(.).

Predicted: 1	Гах4Fun	Farming	Growth Phase	Climate	Growth Phase: Farming	Crop Species
	Glucosidases	0.12	0.15	0.18	0.25	
	Xylosidases	0.10	0.02*	0.13	0.27	
Wheat	Chitinases	0.09.	0.45	0.31	0.12	
	Phosphatases	0.10	0.001**	0.08.	0.47	
	Cellulases	0.20	0.003**	0.06.	0.42	
	Glucosidases	0.01*	<0.001***	0.98	0.64	
	Xylosidases	0.02*	<0.001***	0.81	0.74	
Barley	Chitinases	0.006**	0.29	0.62	0.25	
	Phosphatases	0.02*	<0.001***	0.97	0.84	
	Cellulases	0.02*	<0.001***	0.99	0.88	
	Glucosidases					<0.001***
	Xylosidases					<0.001***
Wheat &	Chitinases					<0.001***
Barley	Phosphatases					0.004**
	Cellulases					0.002**

Predicted: F	PanFP		Growth		Growth	Crop
		Farming	Phase	Climate	Phase:Farming	Species
	Glucosidases	0.10	0.01*	0.21	0.46	
	Xylosidases	0.09.	0.01*	0.22	0.41	
Wheat	Chitinases	0.13	0.07.	0.16	0.30	
	Phosphatases	0.10	0.01*	0.16	0.42	
	Cellulases	0.87	0.01*	0.94	0.80	
	Glucosidases	0.05.	<0.001***	0.89	0.97	
	Xylosidases	0.10	<0.001***	0.96	0.91	
Barley	Chitinases	0.08.	<0.001***	0.86	0.66	
	Phosphatases	0.06.	<0.001***	0.88	0.70	
·	Cellulases	0.08.	<0.001***	0.88	0.81	
	Glucosidases					0.02*
	Xylosidases					0.09.
Wheat &	Chitinases					0.003**
Barley	Phosphatases					0.008**
-	Cellulases					0.17

Table S3 Measured enzyme categories and respective 4-methylumbelliferon (MUF)-coupled substrates

Enzyme	Substrate
β-glucosidase	4-MUF-β-D-glucopyranoside
(EC 3.2.1.21)	
cellobiohydrolase	4-MUF-β-D-cellobioside
(EC 3.2.1.91)	
xylosidase	4-MUF-β-D-xylopyranoside
(EC 3.2.1.37)	
N-acetylglucosaminidase	4-MUF-N-acetyl-β-p-glucosaminide
(EC 3.2.1.14)	
acid phosphatase	4-MUF-phosphate
(EC 3.1.3)	

Table S4 Settings of PCRs used for next generation sequencing with Illumina and qPCR. Conditions for amplification of 16S rRNA gene region (PCR 1), index PCR (PCR 2) and qPCR are listed.

	Step	Temperature	Time (min:sec)
		(°C)	
PCR 1			
	Initial denaturation	95	3:00
25 cycles	Denaturation	98	0:20
	Annealing	55	0:15
	Elongation	72	0:15
	Final extension	72	5:00
PCR 2			
	Initial denaturation	95	3:00
8 cycles	Denaturation	98	0:30
	Annealing	55	0:30
	Elongation	72	0:30
	Final extension	72	5:00
qPCR			
	Initial denaturation	95	5:00
40 cycles	Denaturation	95	0:10
	Annealing	55	0.30

Table S5 Enzyme categories extracted from output of Tax4Fun and PanFP prediction tools. The respective KEGG orthology (KO) identifiers and Enzyme Commission (EC) identifiers for each enzyme are given.

КО	Description (EC number)
K05349	beta-glucosidase [EC:3.2.1.21]
K05350	beta-glucosidase [EC:3.2.1.21]
K01198	1,4-beta-xylosidase [EC:3.2.1.37]
K01205	alpha-N-acetylglucosaminidase [EC:3.2.1.50]
K01078	acid phosphatase [EC:3.1.3.2]
K03788	acid phosphatase (class B) [EC:3.1.3.2]
K09474	acid phosphatase (class A) [EC:3.1.3.2]
K01225	1,4-beta-cellobiosidase [EC:3.2.1.91]

Table S6 Treatment-specific 16S rDNA gene abundances (given as μg DNA per g dry soil) in the rhizosphere. Treatment-specific means (\pm standard deviation in brackets) of DNA concentrations were calculated for both crops separately. Different small letters indicate significant differences between the treatments (ANOVA and Tukey HSD).

Crop/ Year	Growth Phase	Farming system [†]	Concentrations (µg DNA per g dry soil)
	A ative	CF	3.13 ab (1.22)
Wheat	Active	OF	1.94 bc (1.20)
2015	Mature	CF	3.63 a (1.33)
	Mature	OF	3.26 ab (1.86)
	CF Active		2.01 bc (0.90)
Barley	Active	OF	1.25 c (0.49)
2016	Moturo	CF	3.28 ab (1.29)
	Mature	OF	2.69 abc (0.60)

[†] CF = conventional farming, OF = organic farming

Supplemental Figures

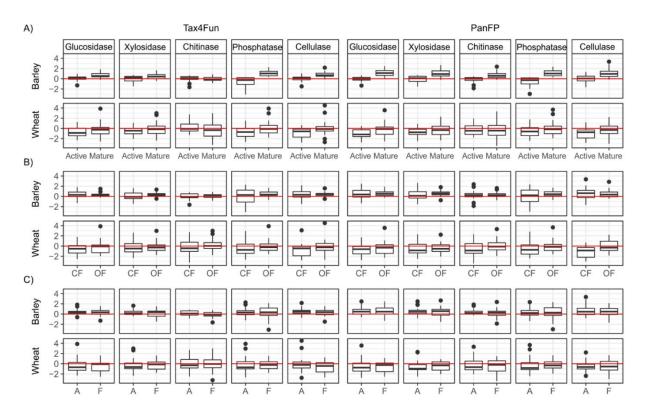


Figure S 1 Difference between measured and Tax4Fun and PanFP predicted z-transformed values. Differences are given for barley and wheat relative to A) active and mature growth phases, B) conventional (CF) and organic (OF) farming systems and C) ambient (A) and future (F) climate treatments. The red lines indicate zero deviations in levels of measured and predicted values, while positive values indicate overestimation and negative values underestimation of predictions compared to measured enzyme activities.

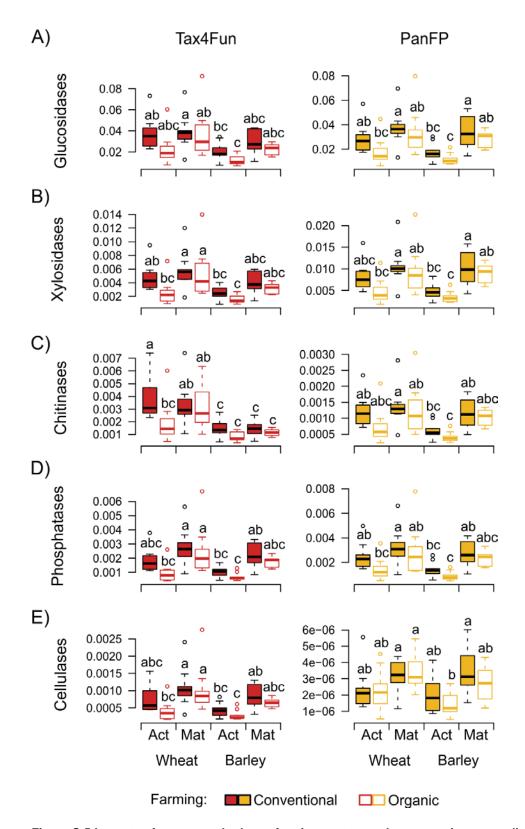


Figure S 5 Impacts of crop growth phase, farming system and crop species on predicted functional gene abundances. A) β -glucosidases, B) xylosidases, C) N-acetylglucosaminidases (chitinases), D) acid phosphatases, and E) cellobiohydrolases (cellulases). Predicted gene abundances by Tax4Fun (red) and PanFP (yellow) at the active (Act) and Mature (Mat) growth phases in conventional and organic farming soils are given. Different small letters within each subfigure indicate significant differences between the treatments (p<0.05) according to Tukey HSD.

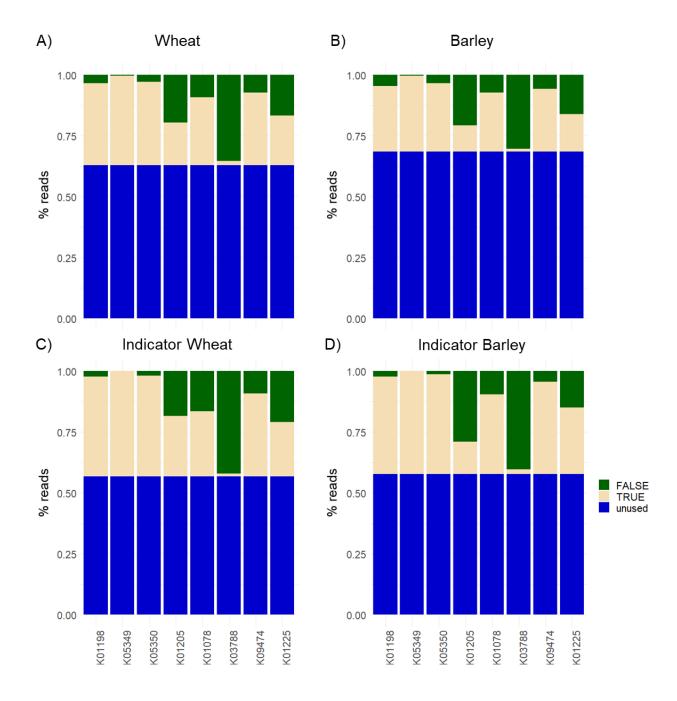


Figure S 6 Proportion of bacterial sequences used for functional predictions with Tax4Fun and containing the enzyme gens. Kegg orthology identifiers represent the five enzymatic classes referred to within this study (see Table S3 for more information on enzymes). Analysis was performed at OTU level, which were weighted by average number of reads in rhizospheres of A) wheat and B) barley, as well as by the respective indicator species abundance in C) wheat and D) barley. FALSE: gene not represented, TRUE: gene represented, unused: no information about presence or absence of gene.

Supplemental Material 1:

Variation in edaphic parameters according to experimental factors

Among the tested abiotic soil parameters, mineral nitrogen concentration (ANOVA, F = 62.74, p < 0.001) as well as total carbon (ANOVA, F = 11.52, p = 0.001) and total nitrogen content (ANOVA, F = 7.47, p = 0.008) in the soil were influenced by farming practice. Thereby, the concentrations were higher in conventional than in organic farming systems (Table SM 1). Moreover, mineral nitrogen concentration (ANOVA, F = 51.96, p < 0.001) and soil moisture (ANOVA, F = 1894.94, p < 0.001) differed significantly between the two growth phases. Mineral nitrogen concentration and soil moisture were higher at the mature than the active growth phase (Table SM 1). Measured soil moisture rather displays short-term changes due to rainfall events at sampling dates than overall climate treatment. An effect of the crop species in each year was found for total nitrogen content (ANOVA, F = 8.20, p = 0.006) and mineral nitrogen concentrations (ANOVA, F = 43.33, p < 0.001). For pH and P neither an effect of the farming system nor of crop species or growth phase could be detected.

Table SM 1 Abiotic soil parameters. Treatment-specific means (\pm standard deviation in brackets) are given for mineral nitrogen (N_{min}), total carbon (TC), total nitrogen (TN), pH, soil moisture, available phosphor (P_{av}). Different small letters within each column indicate significant differences between the treatments.

Crop/ Year	Growth Phase	Farming system [†]	N _{min} (mg/kg)	TC (%)	TN (%)	рН	Moisture (%)	P _{av} (mg/kg)
		CF	10.79 ab	2.05 a	0.17 a	6.79	10.45 a	82.4
	Active -	CF	(5.15)	(0.07)	(0.01)	(0.50)	(0.91)	(40.6)
	Active	OF	4.35 def	1.94 ab	0.16 ab	6.64	10.57 a	73.4
Wheat		OF	(0.89)	(0.07)	(0.01)	(0.56)	(0.33)	(39.7)
2015		CF	13.45 a	1.95 ab	0.16 ab	6.82	15.94 b	83.1
	Mature -	CF	(2.31)	(0.20)	(0.02)	(0.52)	(0.39)	(40.6)
	Mature	OF	7.49 bcd	1.85 b	0.14 b	6.71	15.74 b	72.4
		OF	(1.27)	(0.28)	(0.03)	(0.59)	(0.56)	(36.4)
		CF	3.96 ef	2.01 ab	0.17 a	6.79	9.85 a	75.0
	Active -	CF	(1.02)	(0.09)	(0.01)	(0.48)	(0.59)	(37.3)
	Active	OF	2.71 f	1.92 ab	0.16 ab	6.66	10.58 a	71.2
Barley		OF	(0.41)	(0.09)	(0.01)	(0.55)	(0.75)	(40.9)
2016		CF	9.26 bc	2.00 ab	0.18 a	6.81	16.08 b	73.6
	Moturo —	CF	(2.78)	(0.07)	(0.01)	(0.52)	(0.73)	(36.9)
	Mature -	OF	6.53 cde	1.86 ab	0.16 ab	6.64	15.79 b	70.0
		OF	(2.26)	(0.09)	(0.01)	(0.58)	(0.68)	(37.7)

[†] CF = conventional farming, OF = organic farming

Supplemental Material 2

Effect of abiotic soil parameters on bacterial community structure and function

Plot-specific soil pH was found to be the most prominent edaphic factor shaping rhizosphere bacterial community composition (Figure SM 1, PERMANOVA, R²=0.20, p<0.001), followed by soil moisture (PERMANOVA, R²=0.16, p<0.001), available phosphate concentration (PERMANOVA, R²=0.14, p<0.001), mineral nitrogen concentration (PERMANOVA, R²=0.05, p=0.003) and total carbon content (PERMANOVA, R²=0.05, p=0.003). No effect of total nitrogen content on community composition was found (PERMANOVA, R²=0.02, p=0.33).

Contrary to the community composition, enzyme activities in the rhizosphere of wheat and barley were mainly driven by mineral nitrogen and TC (Table SM 1). While the effect of mineral nitrogen was mainly pronounced in the rhizosphere of wheat, the effect of TC was mainly driving enzyme activities in the barley rhizosphere (Table SM 1). TN influenced enzyme activities of glucosidases and cellulases in the rhizospheres of wheat and barley. Soil moisture had only effects on phosphatases in the barley rhizosphere and on chitinases in the wheat rhizosphere. Available phosphate concentration and pH did not affect enzyme activities (Table SM 1).

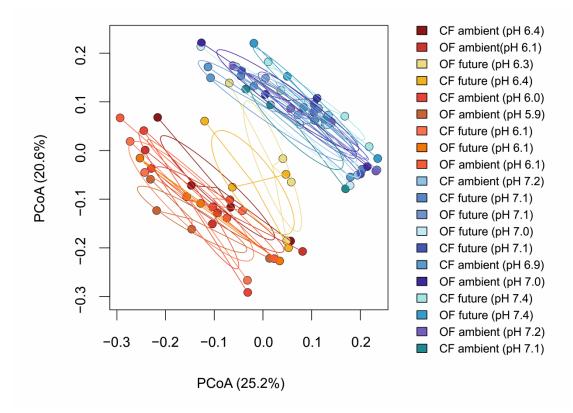


Figure SM 1 Comparison of the beta diversity of the rhizosphere bacterial communities between all plots (n = 80). Principal component analysis based on OTU abundances. Points are colored and connected according to origin of samples from the plots of the GCEF and respective treatments of farming conventional farming (CF) and organic farming (OF) - and climate - ambient and future climate. Means of the respective pH values at the four time points sampled are given in brackets.

Table SM 1 Impact of abiotic soil parameters on rhizosphere enzyme activities. Activities of beta-glucosidases, xylosidases, N-acetylglucosaminidases (chitinases), acid phosphatases and cellobiohydrolases (cellulases) were tested against soil pH, soil moisture, mineral nitrogen (N_{min}), available phosphor (P), total carbon (TC) and total nitrogen (TN). Significant impacts according to ANOVA are indicated by italic p values and significance levels according to ANOVA are given as follows: p < 0.001(***), p < 0.05(*) and p < 0.1(.).

Measured	ı <u> </u>						
Micasarco		рН	Moisture	N_{min}	Р	TC	TN
	Glucosidases	0.50	0.60	0.01*	0.20	0.02*	0.005**
	Xylosidases	0.35	0.09.	0.11	0.67	0.17	0.05.
Wheat	Chitinases	0.60	0.01*	0.002**	0.95	0.17	0.17
	Phosphatases	0.31	0.23	0.005**	0.09.	0.11	0.06.
	Cellulases	0.53	0.80	0.01*	0.28	0.01*	0.01*
	Glucosidases	0.36	0.44	0.44	0.40	0.002**	0.03*
	Xylosidases	0.46	0.40	0.18	0.43	0.01*	0.16
Barley	Chitinases	0.58	0.51	0.21	0.92	0.05.	0.58
	Phosphatases	0.41	0.001**	0.28	0.86	0.08.	0.92
·	Cellulases	0.39	0.72	0.32	0.36	<0.001***	0.03*

Supplemental Material 3

Indicator species analysis

To identify OTUs that drove the observed separation according to experimental factors, indicator species analysis was performed (Figure SM 1). Over the four sampling times and in relation to the farming system treatment 950 and 564 indicator species were identified for wheat and barley, respectively. The phylum-level composition of the set of indicator species was comparable to the total community (Fisher test; Table SM 1 and SM 2). For barley, we found a clear separation between growth phases and across the two farming systems with 122 (21.6%) at active and 143 (25.4%) common indicator species at mature growth phases (Venn diagram, Figure SM 1). Only 24 (4.3%) and 14 (2.5%) indicator species were shared within CF and OF at different growth phases. In contrast, indicator species distribution for wheat showed a partially different pattern (Figure SM 1). While we observed 75 (7.9%) and 84 (8.8%) growth phase-specific shared indicator species, we identified 225 (23.7%) and 72 (7.8%) shared indicator species at different growth phases in CF and OF, respectively (Venn diagram, Figure SM 1).

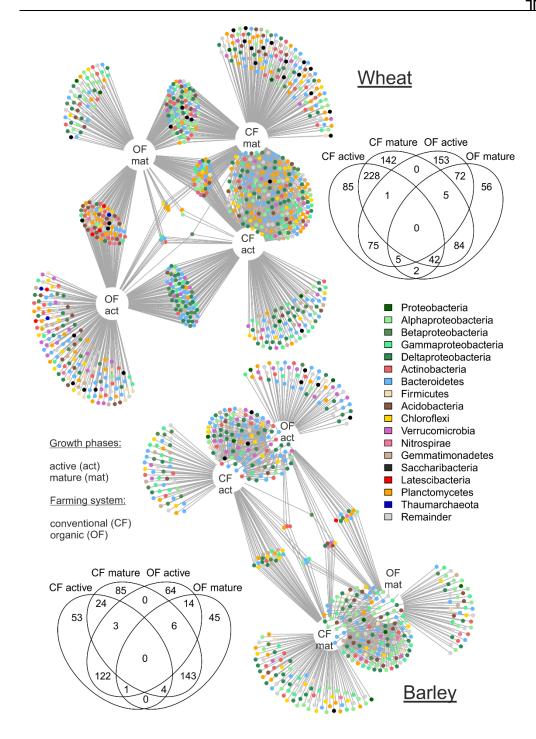


Figure SM 1 Indicator species analysis visualized as bipartite networks and Venn diagrams summarizing indicator species number and their phylogenetic assignment at different growth phases of wheat and barley in conventional and organic farming soils. Circles represent bacterial OTUs significantly associated with single or combined treatments according to consistent results of correlation and likelihood ratio tests (p<0.05). Circles are colored according to the phylum. Remainder comprises low abundant species (cumulative cut-off was set to 98 %).

非 Chapter 2

Table SM 2 Distribution of indicator species among treatments, given at the phylum level for wheat. Enrichments in the occurrence of indicator species within treatments compared to overall community in the rhizosphere of wheat (Total OTU) were estimated by Fisher test. Significant differences are given as different letters and marked by grey background.

Wheat		Indicator species	•	al Farming	Organia	Forming	Charad	Shared
	Total OTU	Total	Convention	mature	Organic active	mature	Shared two	three
Acidobacteria	7917	55 ab	2 a	10 ab	7 a	3 a	31	2
Actinobacteria	9965	64 a	2 a	11 a	13 a	5 a	29	4
Alphaproteobacteria	2916	85 cdefg	4 abc	1 a	14 b-f	11 a	49	6
ARKDMS-49	2	1 a-m	1 abc	0 abcd	0 a-f	0 a	0	0
Armatimonadetes	311	14 cdehikl	1 abc	2 abcd	2 a-f	0 a	8	1
Bacteria_unclassified	1167	12 abj	1 abc	1 abc	0 abc	5 a	5	0
Bacteroidetes	3155	144 hikl	16 b	25 cd	21 bdef	7 a	66	9
Bathyarchaeota	2	0 a-m	0 abc	0 abcd	0 a-f	0 a	0	0
Betaproteobacteria	5458	39 ab	5 ac	8 ab	3 a	2 a	20	1
BJ-169	11	0 a-m	0 abc	0 abcd	0 a-f	0 a	0	0
BRC1	40	5 chkl	1 abc	1 abcd	3 d	0 a	0	0
Candidatus_Berkelbacteria	81	0 a-jm	0 abc	0 abcd	0 a-f	0 a	0	0
Chlamydiae	105	0 abd-gijm	0 abc	0 abcd	0 a-f	0 a	0	0
Chlorobi	44	7 k	0 abc	0 abcd	1 a-f	1 a	4	1
Chloroflexi	2539	40 jm	4 abc	9 abcd	4 abc	2 a	19	2
Cyanobacteria	109	11 kl	1 abc	0 abcd	4 de	1 a	5	0
Deferribacteres	1	0 a-m	0 abc	0 abcd	0 a-f	0 a	0	0
Deinococcus-Thermus	17 2275	1 a-m	0 abc	0 abcd	0 a-f	0 a 9 a	1	0 3
Deltaproteobacteria Elusimicrobia	130	118 hikl 5 c-m	17 b 0 abc	12 bcd 3 bcd	12 b-f 0 a-f	9 a 0 a	65 2	0
Euryarchaeota	32	0 a-m	0 abc	0 abcd	0 a-i 0 a-f	0 a	0	0
FBP	58	2 a-m	0 abc	0 abcd	1 a-f	0 a	0	1
FCPU426	4	2 a-111 1 a-m	0 abc	1 abcd	0 a-f	0 a	0	0
Fibrobacteres	91	4 a-m	1 abc	1 abcd	1 a-f	0 a	1	0
Firmicutes	705	20 cdefghijm	1 abc	11 d	4 a-f	0 a	3	1
Fusobacteria	2	0 a-m	0 abc	0 abcd	0 a-f	0 a	0	0
GAL15	4	0 a-m	0 abc	0 abcd	0 a-f	0 a	0	0
Gammaproteobacteria	2886	53 dfjm	9 bc	6 abc	2 ac	3 a	29	4
Gemmatimonadetes	1486	54 ceghil	9 bc	8 abcd	4 abcf	1 a	30	2
Gracilibacteria	13	1 a-m	0 abc	0 abcd	0 a-f	0 a	1	0
Hydrogenedentes	15	2 a-m	0 abc	0 abcd	1 a-f	0 a	1	0
Ignavibacteriae	1	1 a-m	0 abc	1 abcd	0 a-f	0 a	0	0
JTB23	1	0 a-m	0 abc	0 abcd	0 a-f	0 a	0	0
Latescibacteria	115	5 c-m	0 abc	2 abcd	0 a-f	0 a	3	0
Lentisphaerae	1	0 a-m	0 abc	0 abcd	0 a-f	0 a	0	0
Microgenomates	30	0 a-m	0 abc	0 abcd	0 a-f	0 a	0	0
Nitrospirae	545	4 abjm	0 abc	1 abcd	0 abcf	0 a	3	0
Omnitrophica	48	0 a-m	0 abc	0 abcd	0 a-f	0 a	0	0
Parcubacteria	173	0 abfgjm	0 abc	0 abcd	0 a-f	0 a	0	0
Peregrinibacteria	40	0 a-m	0 abc	0 abcd	0 a-f	0 a	0	0
Planctomycetes	2056	112 hikl	3 abc	16 cd	22 def	3 a	55	13
Proteobacteria_unclassified	193	5 a-m	1 abc	0 abcd	2 a-f	2 a	0	0
RBG-1_(Zixibacteria)	1	0 a-m	0 abc	0 abcd	0 a-f	0 a	0	0
Saccharibacteria	1307	30 cdefgjm	1 abc	5 abcd	15 def	1 a	8	0
Spirochaetae SR1 (Absconditabacteria)	3	0 a-m	0 abc	0 abcd	0 a-f	0 a	0	0
Synergistetes	11 1	1 a-m 0 a-m	0 abc 0 abc	0 abcd 0 abcd	0 a-f 0 a-f	0 a 0 a	1 0	0 0
Tectomicrobia	53		0 abc	0 abcd				0
Tenericutes	55 15	2 a-m 0 a-m	0 abc	0 abcd	0 a-f 0 a-f	0 a 0 a	2 0	0
Thaumarchaeota	361	3 abdefgjm	0 abc	2 abcd	0 a-i	0 a	1	0
TM6_(Dependentiae)	121	2 a-m	0 abc	1 abcd	0 abcei	0 a	1	0
Verrucomicrobia	1811	47 cdefgm	5 abc	15 cd	6 abcef	0 a	18	3
Woesearchaeota_(DHVEG-6)	14	0 a-m	0 abc	0 abcd	0 abcer	0 a	0	0
WS2	20	0 a-m	0 abc	0 abcd	0 a-f	0 a	0	0
WWE3	8	0 a-m	0 abc	0 abcd	0 a-f	0 a	0	0
Total	48462	950	85	153	142	56	461	53

Table SM 3 Distribution of indicator species among treatments, given at the phylum level for barley. Enrichments in occurrence of indicator species within treatments compared to overall community in the rhizosphere of barley (Total OTU) were estimated by Fisher test. Significant differences are given as letters and additionally marked by grey background.

Barley		Indicator specie		al Farmina	Organia	Earmin a	Charad	Charad
	Total OTU	Total	Convention active	mature	Organic active	mature	Shared two	Shared three
Acidobacteria	8463	23 a	1 a	1 a	0 a	3 a	16	2
Actinobacteria	9218	35 a	3 ab	6 abc	4 ab	3 a	17	2
Alphaproteobacteria	3085	63 b-g	1 abc	14 de	2 abcd	10 a	35	1
ARKDMS-49	2	1 a-j	0 abc	0 a-f	1 bcd	0 a	0	0
Armatimonadetes	346	8 b-j	1 abc	0 a-i 0 a-f	3 bcd	2 a	2	0
Bacteria_unclassified	1281	20 bcefhi	1 abc	4 abcde	6 cd	2 a 2 a	7	0
Bacteroidetes	2881		11 c	5 abcde	14 c	2 a 5 a	, 56	2
Bathyarchaeota	1	93 dgj 0 a-j	0 abc	0 a-f	0 abcd	0	0	0
Betaproteobacteria	5369	26 a	9 abc	1 ab	1 ab	1 a	13	1
ВЈ-169	12	0 a-j	0 abc	0 a-f	0 abcd	0 a	0	0
BRC1	43	0 a-j 1 a-j	0 abc	1 a-f	0 abcd	0 a	0	0
Candidatus_Berkelbacteria	111	•	0 abc	0 a-f	0 abcd	1 a	0	0
_	122	1 a-j 0 abcdhi	0 abc		0 abcd	0 a	0	0
Chlamydiae Chlorobi	48	3 b-hj	0 abc	0 a-f 3 ef	0 abcd	0 a	0	0
	2683	•				0 a 2 a	16	1
Chloroflexi Cyanobacteria	2683 125	30 bhi 11 j	3 abc 1 abc	8 cd 5 f	0 abd 0 abcd	∠ a 1 a	4	0
Cyanobacteria Deforribactores		•	0 abc			та 0	0	0
Deferribacteres Deinococcus-Thermus	1 18	0 a-j 0 a-j	0 abc	0 a-f 0 a-f	0 abcd 0 abcd	0 0 a	0	0
				14 def				
Deltaproteobacteria Diapharatritas	2723 2	79 cdeg	2 abc		8 bcd	5 a 0	49	1 0
Diapherotrites Elusimicrobia		0 a-j	0 abc	0 a-f	0 abcd		0	0
	137	2 a-j	0 abc	0 a-f	0 abcd	0 a	2	
Epsilonproteobacteria	1	0 a-j	0 abc	0 a-f	0 abcd	0	0	0
Euryarchaeota	36	0 a-j	0 abc	0 a-f	0 abcd	0 a	0	0
FBP	59	2 a-j	2 abc	0 a-f	0 abcd	0 a	0	0
FCPU426	4	1 a-j	0 abc	0 a-f	1 abcd	0 a	0	0
Fibrobacteres	118	3 a-j	0 abc	0 a-f	0 abcd	1 a	2	0
Firmicutes	698	12 b-i	0 abc	2 a-f	1 abcd	1 a	8	0
Fusobacteria	2	0 a-j	0 abc	0 a-f	0 abcd	0 a	0	0
GAL15	6	0 a-j	0 abc	0 a-f	0 abcd	0 a	0	0
Gammaproteobacteria	3009	46 bfhi	9 bc	7 bcd	3 abcd	4 a	22	1
Gemmatimonadetes	1600	20 bhi	4 abc	0 abcd	0 abcd	2 a	14	0
Gracilibacteria	26	1 a-j	0 abc	0 a-f	0 abcd	0 a	1	0
Hydrogenedentes	16	0 a-j	0 abc	0 a-f	0 abcd	0 a	0	0
Ignavibacteriae	1	0 a-j	0 abc	0 a-f	0 abcd	0 a	0	0
JTB23	1	0 a-j	0 abc	0 a-f	0 abcd	0 a	0	0
Latescibacteria	126	1 a-j	0 abc	0 a-f	0 abcd	0 a	1	0
Lentisphaerae	1	0 a-j	0 abc	0 a-f	0 abcd	0 a	0	0
Microgenomates	36	0 a-j	0 abc	0 a-f	0 abcd	0 a	0	0
Nitrospirae	620	5 abhi	0 abc	1 abcde	0 abcd	1 a	2	1
Omnitrophica	65	0 a-j	0 abc	0 a-f	0 abcd	0 a	0	0
Parcubacteria Parcubacteria	329	2 abcdhi	0 abc	0 a-f	0 abcd	0 a	2	0
Peregrinibacteria	72	0 a-j	0 abc	0 a-f	0 abcd	0 a	0	0
Planctomycetes	2165	28 bhi	3 abc	9 de	6 bcd	0 a	9	1
Proteobacteria_unclassified	242	7 b-hj	0 abc	1 a-f	1 abcd	1 a	4	0
RBG-1_(Zixibacteria)	1	0 a-j	0 abc	0 a-f	0 abcd	0 a	0	0
Saccharibacteria	1256	8 ahi	0 abc	1 abcd	4 bcd	0 a	3	0
Spirochaetae	5 46	0 a-j	0 abc	0 a-f	0 abcd	0 a	0	0
SR1_(Absconditabacteria)	16	3 efgj	0 abc	1 a-f	1 abcd	0 a	1	0
Synergistetes	1	0 a-j	0 abc	0 a-f	0 abcd	0 a	0	0
Tectomicrobia	59	0 a-j	0 abc	0 a-f	0 abcd	0 a	0	0
Tenericutes	10	0 a-j	0 abc	0 a-f	0 abcd	0 a	0	0
Thaumarchaeota	376	0 ai	0 abc	0 abcde	0 abcd	0 a	0	0
TM6_(Dependentiae)	138	1 a-i	0 abc	0 a-f	0 abcd	0 a	1	0
Verrucomicrobia	1935	28 bfhi	2 abc	1 abcd	8 cd	0 a	16	1
Woesearchaeota_(DHVEG-6)	22	0 a-j	0 abc	0 a-f	0 abcd	0 a	0	0
WS2	24	0 a-j	0 abc	0 a-f	0 abcd	0 a	0	0
WWE3	14	0 a-j	0 abc	0 a-f	0 abcd	0 a	0	0
Total	49747	564	53	85	64	45	303	14

	Chapter	3
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CHAPTER 3

Shifts Between and Among Populations of Wheat Rhizosphere Pseudomonas, Streptomyces and Phyllobacterium Suggest Consistent Phosphate Mobilization at Different Wheat Growth Stages Under Abiotic Stress

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Shifts Between and Among Populations of Wheat Rhizosphere Pseudomonas, Streptomyces and Phyllobacterium Suggest Consistent Phosphate Mobilization at Different Wheat Growth Stages Under Abiotic Stress

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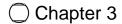
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Climate change models predict more frequent and prolonged drought events in Central Europe, which will exert extraordinary pressure on agroecosystems. One of the consequences is drought-related nutrient limitations for crops negatively affecting agricultural productivity. These effects can be mitigated by beneficial plant growth promoting rhizobacteria. In this study, we investigated the potential of cultivable bacterial species for phosphate solubilization in the rhizosphere of winter wheat at two relevant growth stages - stem elongation and grain filling stages. Rhizosphere samples were collected in the Global Change Experimental Facility in Central Germany, which comprises plots with conventional and organic farming systems under ambient and future climate. Phosphate-solubilizing bacteria were selectively isolated on Pikovskaya medium, phylogenetically classified by 16S rRNA sequencing, and tested for in vitro mineral phosphate solubilization and drought tolerance using plate assays. The culture isolates were dominated by members of the genera Phyllobacterium, Pseudomonas and Streptomyces, Cultivation-derived species richness and abundance of dominant taxa, especially within the genera Phyllobacterium and Pseudomonas, as well as composition of Pseudomonas species were affected by wheat growth stage. Pseudomonas was found to be more abundant at stem elongation than at grain filling, while for Phyllobacterium the opposite pattern was observed. The abundance of Streptomyces isolates remained stable throughout the studied growth stages. The temporal shifts in the cultivable fraction of the community along with considerable P solubilization potentials of Phyllobacterium and Pseudomonas species suggest functional redundancy between and among genera at different wheat growth stages. Phosphate-solubilizing Phyllobacterium species were assigned to Phyllobacterium



ifriqiyense and *Phyllobacterium sophorae*. It is the first time that phosphate solubilization potential is described for these species. Since *Phyllobacterium* species showed the highest drought tolerance along all isolates, they may play an increasingly important role in phosphate solubilization in a future dryer climate.

Keywords: climate change, agriculture, wheat, PGPR, phosphate solubilization, drought tolerance

INTRODUCTION

Winter wheat (Triticum aestivum L.) is the most commonly grown crop in Germany accounting for 54% of the total agricultural area in 2017 (Schmeling, 2017). At the same time wheat is the second most produced crop worldwide after maize (Statista, 2018). The production of winter wheat depends on frequent precipitation and sufficient nutrient supply (Fischer, 2003). Its cultivation is hence restricted to appropriate regions and is commonly supported by sufficient fertilization. Besides nitrogen and potassium, phosphorus (P) has to be applied, since the plant-available P in the rooted topsoil often represents a growth limiting factor. However, P fertilization does not necessarily increase P availability. Up to 90% of fertilized mineral P is quickly immobilized by reactive cations, such as Al3+ or Fe³⁺ in acidic or Ca²⁺ in calcareous or neutral soils (Oehl et al., 2002). At a global scale, models suggest an increase in P input into agricultural systems of 51-86% by the year 2050 (Mogollón et al., 2018) to maintain productivity. In parallel, the source for P fertilizer - natural rock phosphate - is expected to deplete in 50-100 years (Cordell et al., 2009). Regarding the application of organic fertilizers phosphorus is bound into complex compounds and first has to be released before plant uptake.

Processes that release P from mineral and organic compounds in the soil are mainly driven by microorganisms. Therefore, to counteract P limitation, plants enrich a variety of beneficial microbes in their rhizospheres by secreting nourishing, carbon-containing rhizodeposits. Besides mycorrhizal fungi, plant growth promoting rhizobacteria (PGPR) can enhance plant P uptake. PGPR thereby stimulate root growth, release P from immobilized inorganic and organic P pools in the soil, and transport solved P compounds to the plant's roots (Artursson et al., 2006). However, these mutual interactions between the plant and its rhizosphere community are highly dynamic and related to the plant growth stage as well as to environmental conditions.

Seasonal shifts in rhizosphere microbiomes related to plant growth stages are well documented for canola and grasses (Dunfield and Germida, 2003; Wolsing and Priemé, 2004; Francioli et al., 2018). The changes are mainly caused by variations in the quantity and quality of carbon-derived rhizodeposits serving as energy sources for soil microbes (reviewed in Bais et al., 2006). These rhizodeposits are predominantly released into the rhizosphere when plants are actively growing (Badri and Vivanco, 2009). Thus, flowering plants (Aulakh et al., 2001; Lucas García et al., 2001) mainly secrete root exudates during stem elongation, booting and flowering stages, and much less during fruit development and ripening stages. Malhi et al. (2006) also found stem elongation

and tillering to be the stages with maximum nutrient uptake of cereals. Investigations of the wheat rhizobiome often focus on the stages of tillering and flowering, but also on grain filling (predominantly in field experiments, e.g., Juhnke et al., 1987; Creus et al., 2004; Naiman et al., 2009; Ilyas and Bano, 2010) or are independent of certain growth stages throughout the first weeks of wheat development (predominantly in pot experiments, e.g., Kasim et al., 2013; Timmusk et al., 2014). Clearly, more information on the interrelationships between plant growth stages and the PGPR are needed.

Besides the seasonal dynamics, microbial communities in agricultural soils have to cope with frequent disturbances. Agricultural management practices such as fertilization, tillage, pesticide application and crop rotation including brown fallow periods cause dramatic changes of C supply patterns within the soil rooting zone. It was shown that tillage (Frey et al., 1999; Kladivko, 2001) and crop rotation (Tiemann et al., 2015) lead to a decrease in fungal biomass (Smith and Read, 2010) and diversity (Verbruggen et al., 2010). A recent study of Banerjee et al. (2019) also demonstrated that land use intensification reduces fungal network complexity in conventional and even in no tillage land use systems. In contrary, bacterial biomass was nearly unaffected (Elliott et al., 1988), leading to establishment of a primarily bacteria-based food web in managed agroecosystems (Hendrix et al., 1986; de Vries and Shade, 2013). Therefore, we focused on the bacterial rather than the fungal community in this study.

The high adaptation potential of soil bacteria to changing conditions is related to their high growth and mutation rates, as well as rapid recombination by lateral gene transfer (Badri and Vivanco, 2009). Moreover, to maintain crucial soil processes after disturbances and under changing conditions, the huge bacterial diversity allows quick and diverse reordering of the active community (Kennedy, 1999). According to Sheik et al. (2011), stress conditions such as the combination of warming and drought effects lead to less diverse microbial populations. In contrast, functions that have to be provided for plants by microbes may be even more essential under stress conditions and a more complex population structure may arise, since more complex co-occurrence networks improve complementarity and efficiency (Tardy et al., 2014; Karimi et al., 2017). Thus, soil bacterial communities play a key role in stress resistance and recovery after disturbances of agroecosystems (Shade et al., 2012).

Most field studies on beneficial wheat rhizosphere microbes and their plant growth promoting properties have been performed in arid and semiarid regions of India (Rana et al., 2011; Jog et al., 2014; Kumar et al., 2014; Singh and Lal, 2016; Verma et al., 2016), Pakistan (Ilyas and Bano, 2010; Naveed et al., 2014)

and Argentina (Creus et al., 2004; Fischer et al., 2007). As these regions are strongly limited in water supply and therefore in the amount of available nutrients, which are solved in the soil water, the potential of PGPR to support wheat growth under drought conditions by different functional properties is of high interest. In contrast, studies in temperate zones with moderate precipitation are underrepresented, although climate change already causes obvious negative effects on agroecosystem productivity (Ciais et al., 2005; Trnka et al., 2014). For instance in 2018 East and Middle Germany experienced a severe drought period, which led to premature ripening and an early harvest of crops (BMEL, 2018). Drought along with nutrient limitation caused yield losses will likely increase in future as climate change models predict altered precipitation patterns with extended drought periods combined with higher temperatures in the vegetation period (Petersen and Weigel, 2015). Therefore, the role of soil bacteria and their potentials for tolerance and resilience of agroecosystems in the temperate zones is of increasing importance.

To understand the role of the rhizosphere bacterial community in providing P to wheat plants over the growing season, it is indispensable to identify P-solubilizing key species and their functionality in dependency on wheat growth stage while simultaneously considering the impacts of the two most important global change drivers, agricultural management and climate (Sala et al., 2000). In order to address this issue, we investigated P-solubilizing rhizosphere bacteria at two relevant wheat growth stages, namely stem elongation (vegetative biomass production) and grain filling stage (generative biomass production). Wheat samples were collected in the Global Change Experimental Facility (GCEF, Schädler et al., 2019). The design of the experimental field platform allows the comparison of two different farming systems (conventional vs. organic agriculture) exposed to either ambient climate, at present, or future expected climate scenario with warming and increased summer drought in the latter. Within this experimental setup we used a cultivation-dependent approach based on Pikovskaya medium (Pikovskaya, 1948) to isolate exclusively for P-solubilizing bacteria. We decided for tri-calcium phosphate (TCP) as insoluble phosphate source in the neutral Pikovskaya media since the pH in the plots of the GCEF is mainly neutral and reactive anion Ca²⁺ is predominantly binding mineral P in these soils forming stable complexes (Oehl et al., 2002, Bashan et al., 2013). Solubility of TCP increases with decreasing pH. Acidification by production and release of organic acids is thus one possibility to release phosphate from insoluble TCP source and is therefore an indicator for the potential of phosphate solubilization in vitro (Bashan et al., 2013). We compared the in vitro P solubilization potential and the drought tolerance of the predominantly isolated P-solubilizing taxa at the stem elongation stage (vegetative biomass production) with those present at the grain filling stage (generative biomass production) in different land use systems and under different climatic conditions. We aimed to distinguish the responses in terms of structural resistance and/or functional redundancy between and among cultivable P-solubilizing bacterial taxa. Under changing conditions, identical taxa and functions would suggest adaptation and resistance. In contrast, shifts in the cultivable fraction of the community would either reflect changes in functionality or reveal functional redundancy among P-solubilizing bacterial species.

We hypothesized that (i) the abundances of the predominant P-solubilizing taxa in the wheat rhizosphere vary between stages of wheat development, displaying a higher abundance, diversity and activity of P-solubilizing bacteria at the stem elongation stage due to higher resource demand by plants than at the grain filling stage (Malhi et al., 2006). Furthermore, we hypothesized that (ii) increased drought decreases diversity but induces targeted structural changes within the bacterial community, in a way that P solubilization is maintained. Finally, due to a lower availability of nutrients, we expected (iii) a lower abundance, but higher activity potentials of P-solubilizing bacteria in the organic farming system compared to the conventional farming system.

MATERIALS AND METHODS

Experimental Platform

The Global Change Experimental Facility (GCEF) (see http: //www.ufz.de/index.php?en=40038 for general information and visualizations) was established in 2013 to investigate climate change effects on managed terrestrial ecosystems (Schädler et al., 2019). The GCEF is located at the field research station of the Helmholtz-Centre for Environmental Research (UFZ) in Bad Lauchstädt, Germany (51°23'35"N 11°52'55"E, 118 m a.s.l.). The site is characterized by a temperate climate with an average temperature of 9.7°C (1993-2013) and a mean annual precipitation of 525 mm (1993-2013). The soil type is a humusand nutrient-rich Haplic Chernozem (Altermann et al., 2005). Besides three grassland management types, a conventional and an organic farming system are part of the GCEF design. In total, the facility includes 50 large plots (24 \times 16 m) that allow realistic field management. The 50 plots are arranged in ten blocks, whereby each block comprises all five land use types that are randomly distributed (Supplementary Figure S1). Five blocks are subjected to ambient climatic conditions and the other five to future climatic conditions, resulting in a split-plot experimental design (Supplementary Figure S1). Ambient climate treatment refers to the actual and non-manipulated climate in terms of precipitation and temperature at the field site. According to the projected climate change within the next 50 years for Central Germany, the future climate treatment of the GCEF consists in increasing the mean daily temperature and shifting the annual precipitation pattern. Temperature is increased passively by roofing the plots during the night. Precipitation is modulated based on the actual weather conditions, with a 20% decrease of precipitation during summer (Jun-Aug) and a 10% increase of precipitation in spring (March-May) and autumn (September-November). In this study we focused on the conventional farming (CF) and organic farming (OF) systems. The CF includes a typical regional crop rotation consisting of a sequence of winter rape, winter wheat and winter barley. In two out of 3 years, the crop cycle for OF includes winter wheat and winter barley just as for the CF. In the first and fourth year of this bipartite crop sequence, the nitrogen-fixating legumes alfalfa and white clover are included in the crop cycle, respectively. Herbicides, fungicides, insecticides and plant growth regulators are applied as usual in CF practice (Schädler et al., 2019). For OF only mechanical weed control and a restricted use of pesticides are allowed. Whereas N, P, and K in CF is applied in form of mineral fertilizers, N fertilization in OF is done by the inclusion of legumes in the crop rotation and further fertilization is restricted to the application of rock phosphate (P-Ca-Mg) and patent kali (K-Mg-S) in the first year of the crop cycle.

Sampling and Characterization of Soil Chemical Parameters

Rhizosphere from root systems of winter wheat was collected in organic (OF) and conventional farming (CF) plots of the GCEF exposed either to ambient (A) or simulated future (F) climate conditions. From each of these plots of the GCEF (20 plots) - CF-A (5 plots), CF-F (5 plots), OF-A (5 plots) and OF-F (5 plots) - three plants were taken with root system finalizing in 15 samples per treatment. To cover relevant growth stages for vegetative and generative biomass production over the growing season, sampling campaigns took place at the 26th of May 2015 (BBCH growth stages 37-39, representing stem elongation phase) and at the 8th of July 2015 (BBCH growth stages 75-77, representing the grain filling stage), respectively. Growth stages were identified according to BBCH-scale ("Biologische Bundesanstalt, Bundessortenamt und CHemische Industrie"). Wheat plants were carefully removed from the soil, their root systems cut and shaken to discard unattached soil, and transferred in plastic bags on ice to the laboratory. Subsequently, rhizosphere soil was collected by gently, manually loosening the root-attached soil to avoid damages to the roots. To measure soil chemical properties, bulk soil samples were collected in parallel to rhizosphere samples. Six soil cores per plot were taken to a depth of 15 cm, pooled, sieved to 2 mm and manually cleaned from organic material. The obtained rhizosphere and bulk soil were stored frozen until usage.

Soil chemical properties of the bulk soil samples were analyzed as follows: gravimetric water content of fresh soil was quantified with a halogen moisture analysator (Mettler Toledo, Gießen, Deutschland). pH was measured with an electrode after shaking the soil for 1 h in 0.01 M CaCl2 (1:2.5 w/v). Amounts of total carbon and nitrogen content were determined from airdried soil using an elemental analyzer (Elementar Vario EL III, Elementar, Hanau, Germany). For analysis of mineral nitrogen 5 g of fresh soil were suspended in 20 ml of 1 M KCl solution and measured per flow injection analysis (FlAstar 5000, Foss GmbH, Rellingen, Germany). Extraction of labile phosphorus in the soil was performed with double lactate solution (1:50 w/v) at pH 3.6 and determined using the colorimetrical molybdenum blue method (Murphy and Riley, 1962).

Selective Isolation and Phylogenetic Classification of Phosphate Solubilizing Bacteria From Wheat Rhizosphere

For cultivation and isolation of bacteria from wheat rhizosphere, 0.5 g of fresh rhizosphere soil was suspended in 50 ml of distilled

water and stirred for 5 min to break up soil particles and bring bacteria in suspension. Subsequently, the samples were sonicated for 10 s and stirred for another 2 min. Mineral tricalcium phosphate solubilizing bacteria were isolated from a 1:200 (v/v) dilution of this rhizosphere soil suspension by plating 50 μl of it in triplicate on Pikovskaya agar (Pikovskaya, 1948), supplemented with the fungicide cycloheximide (10 mg/ml). After 2 weeks of incubation at 25°C, colony forming units (CFU) were counted. Bacterial strains produce and release organic acids to solubilize P from insoluble tri-calcium phosphate source, therefore, forming a clear zone in the otherwise turbid agar around the colony - from heron referred to as halo. Up to 25 colonies (Supplementary Table S4) with the largest halos were collected for each sample from the Pikovskaya plates and transferred to yeast malt extract (YME) agar plates. To identify and store these isolates, polyethylene glycol (PEG) and glycerol stocks were prepared and frozen. Cells in the PEG stock were destroyed mechanically with glass beads by vortexing to release their DNA into the liquid. Polymerase chain reaction (PCR) was performed on the DNA obtained from the PEG stocks with the primers 27f (10 μ M - 5'-AGAGTTTGATCMTGGCTCAG-3': Lane, 1991) and 1492r (10 μM - 5'-GGTTACCTTGTTACGACTT-3': Lane, 1991) and Promega Green (Promega, Madison, WI, United States). Partial 16S rRNA sequencing was performed with the primer BAC 341f primer (10 µM - 5'-CCTACGGGAGGCAGCAG-3': Muyzer et al., 1993) using Big Dye Termination Mix (GeneCust Europe, Dudelange, Luxemburg). Quality control of the sequences was done manually using Sequencher 5.4.5. Quality-checked sequences were compared with type strain reference sequences of the National Center of Biotechnology Information (NCBI), and subsequently clustered according to a level of 99.5% identity. Representative sequences of each cluster were aligned using neighbor-joining algorithms of the MAFFT server (Katoh and Standley, 2013). Phylogenetic trees were constructed and visualized along with the trait and abundance data using Evolview (Zhang H. et al., 2012). All 16S rRNA gene sequences were deposited in the NCBI database with continuous accession numbers from MK637853 to MK638668 (Supplementary Table S1).

Estimating the Functional Properties of Isolated Bacteria

Pure cultures of the dominant P-solubilizing taxa were characterized by functional tests using agar plate bioassays. For each isolate, organic acid production to solubilize tri-calcium phosphate was quantitatively and qualitatively determined on Pikovskaya agar (Pikovskaya, 1948) by the formation of a clear zone in the otherwise turbid agar around the colony. Drought tolerance of the isolates was determined by comparing their growth on YME to surface polyethylene glycol 6000 (PEG) infiltrated YME. Thus, in this study drought tolerance potential of the strains was not related to the formation of dormant spores, but to their potential to maintain metabolisms and growth under water limiting conditions. To simulate severe drought stress, PEG concentration was set to 500 g/L, which

corresponds to an osmotic potential level of around -1.1 MPa (Verslues et al., 2006).

For both bioassays, the strains were grown to the end of the exponential growth phase in liquid YME and 1 μ l of each liquid pre-culture was applied on the test media in triplicates. P solubilization was assessed after 14 and drought tolerance after 3 days of incubation in the dark at 25°C. Phosphate solubilization was quantified as P release derived from phosphate concentration per area of plate and valued using two complementary indices. On the one hand, P release was calculated based on the whole area including colony plus halo, representing the total amount of P released by the colony (Phosphate Solubilization Index 1, PSI 1). Since primarily the outer cells of a colony determine the distance between colony rim and halo, we further calculated P release (Phosphate Solubilization Index 2, PSI 2) based solely on the halo around the colony (halo diameter = twice the distance from colony rim to the end of the halo), representing cellspecific, rather than colony-specific P solubilization potential. Drought stress tolerance was quantified based on the percentage differences of colony diameters on PEG-YME agar, compared to control YME agar. For statistical analysis only data from bacterial strains growing on the functional media were used.

Statistical Analyses

All statistical analyses were performed in the open source program R version 3.4.0 (2017-04-21), GNU project (R Core Team, 2017). An analysis of variance (ANOVA) was run to evaluate, if the abundance and traits of P-solubilizing bacteria were influenced by climate, land use, wheat growth stage, or interaction of these factors. The ANOVA was followed by a multiple comparison analysis with Tukey HSD to identify respective differences in the means of groups. To test for differences in phosphate solubilization potentials and drought tolerances between dominant bacterial taxa, the factor taxa was added to the formula resulting in a four factorial ANOVA. Significance levels were defined according to ***p < 0.001, **p < 0.01, and *p < 0.05.

To test for homogeneity of multivariate dispersion of *Phyllobacterium*, *Pseudomonas* and *Streptomyces* phylogenetic clusters among and between test factors climate, land use and wheat growth stage, betadisper function (R, vegan package) was run based on a dissimilarity matrix using relative abundance (Bray Curtis) and incidence data (Jaccard). Biodiversity indices, i.e., species richness and Shannon-Weaver-index were calculated to test for impacts of climate, land use and wheat growth stage on structure and composition of cultured *Phyllobacterium*, *Pseudomonas* and *Streptomyces* species. ANOVA was applied to test for differences in cluster dispersion regarding structure and composition of *Phyllobacterium*, *Pseudomonas* and *Streptomyces* species between and among test factors.

RESULTS

Soil Chemical Parameters

Soil chemical properties of the studied samples are summarized in **Supplementary Table S2**. Since our study was performed only 2 years after establishment of the GCEF, basic soil parameters did not differ between treatments and showed comparable values of pH (6.7), total organic carbon (1.9%) and total nitrogen (0.16%). Available P and mineral N were higher in CF than in OF, but the difference was only significant for mineral N (Supplementary Table S2). The soil moisture differed between the ambient and future plots at the days of sampling and showed average values of 10.6 and 15.9% at stem elongation stage in May and at grain filling stage in July, respectively (Supplementary Table S2).

Effects of Wheat Growth Stage, Land Use and Climate on P-Solubilizing Rhizosphere Bacteria

Colony counts on Pikovskaya agar indicated higher numbers of P-solubilizing rhizobacteria at the stage of grain filling in July compared to the stage of stem elongation in May (p < 0.001). At both growth stages more bacteria were detected in CF than in OF (May: p = 0.008, July: p = 0.03) and under ambient climate than under future climate at grain filling stage in July (p = 0.02) (Supplementary Figure S2).

In total 410 bacterial isolates were obtained from samples collected at the stage of stem elongation and another 407 strains from samples at the grain filling stage (Supplementary Table S4). Within the isolates collected from the May samples at stem elongation stage, the dominant taxa (relative abundance ≥2%) belonged to the genera Pseudomonas (150 isolates = 37%), Streptomyces (81 isolates = 20%), Phyllobacterium (37 isolates = 9%), Rhizobium (20 isolates = 5%), Mesorhizobium (16 isolates = 4%), Bacillus (12 isolates = 3%) and Agrobacterium (10 isolates = 2%), while in the collection from July samples at the grain filling stage Phyllobacterium (147 isolates = 36%), Streptomyces (116 isolates = 29%), Rhizobium (24 isolates = 6%), Agrobacterium (20 isolates = 5%), Pseudomonas (17 isolates = 4%), Mesorhizobium (16 isolates = 4%), Ensifer (10 isolates = 2%) and Bacillus (9 isolates = 2%) (Supplementary Table S5) were the dominant genera. Thus, along both wheat growth stages, the taxa belonging to the genera Phyllobacterium, Pseudomonas and Streptomyces were found to dominate the bacterial collections, jointly accounting for 65% and 69% of the total isolates at stem elongation in May and grain filling stage in July, respectively. However, between the two plant growth stages, shifts in abundances of these dominant taxa were observed (Figures 1A,B and Supplementary Table S5). At the stage of stem elongation in May, Pseudomonas were predominant with 150 isolates accounting for more than one third of the bacterial isolates, whereas Phyllobacterium species only made up a small fraction with 37 isolates (9%). Conversely, at the grain filling stage in July, a much higher abundance of P-solubilizing Phyllobacterium species (147 isolates = 36%) than Pseudomonas species (17 isolates = 4%) was detected. Contrary to these, the abundance of Streptomyces isolates was less affected by the wheat growth stage, when 81 isolates (20%) and 116 isolates (29%) were obtained from stem elongation stage and grain filling stage, respectively.

Comparable abundances along the two growth stages were also found for the less abundant taxa of the genus

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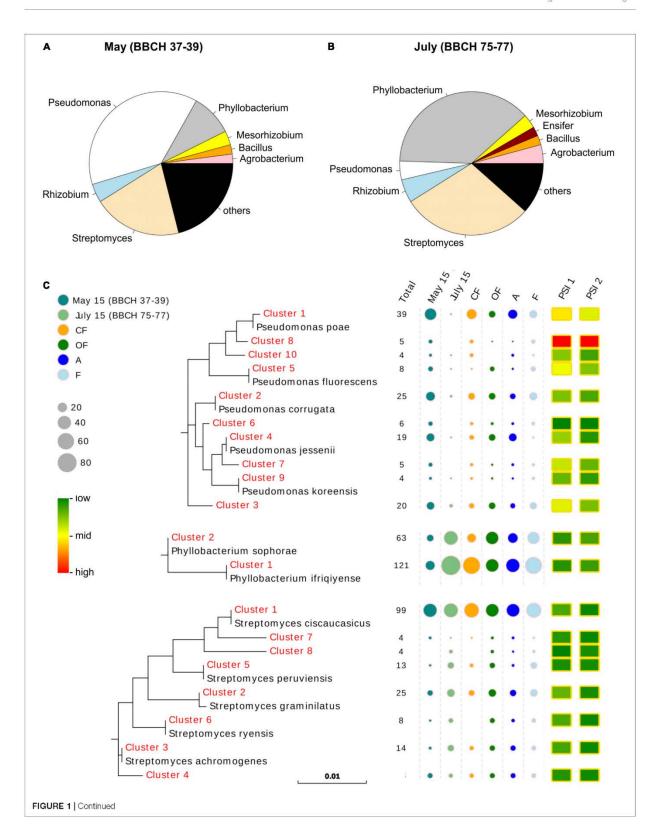


FIGURE 1 | Identity, abundance and activity levels of cultivable mineral phosphate-solubilizing bacteria in the wheat rhizospheres. Taxa have been summarized on genus level and the most dominant ones, with a relative abundance >2% within the collection, are shown in the pie charts at (A) stem elongation stage in May 2015 and (B) grain filling stage in July 2015. The predominantly isolated taxa of the genus Phyllobacterium, Pseudomonas and Streptomyces were clustered (identity level of 99.5%) and (C) phylogenetic trees were created for dominant clusters (number of isolates > 3) with reference sequences from NCBI. Information on abundance of the dominant clusters within treatment levels of growth stage (stem elongation in May, grain filling in July), climate (A-ambient, F-future) and land use (CF-conventional farming, OF-organic farming) are given as circles, whereas circle size is related to the number of isolates. Moreover, the average phosphate solubilizing potentials for each bacterial 16S rPMA gene cluster were calculated for PSI 1 and PSI 2 and differences in expression levels are indicated using a heatmap.

Agrobacterium, Bacillus and Rhizobium, while the genus Ensifer was predominantly detected in July (Figure 1B and Supplementary Table S5). Some of the less abundant genera also showed indications for land use and climate-specific patterns. Agrobacterium and Buttiauxiella were more frequently detected in CF than in OF, whereby Agrobacterium spp. were mainly isolated from ambient climate plots, while Buttiauxiella spp. were exclusively detected in future climate plots (Supplementary Table S5). A comparable enrichment under future climatic conditions was also observed for Bacillus, and species were more often isolated from OF than from CF (Supplementary Table S5).

To assess species identity and trait distribution within the dominant genera, we clustered the Phyllobacterium, Pseudomonas and Streptomyces isolates according to their 16S rRNA gene sequences at a similarity level of 99.5%. Phylogenetic trees were constructed with representative sequences of clusters comprising more than three isolates (Figure 1C), resulting in two 16S rRNA gene clusters for Phyllobacterium, ten for Pseudomonas and eight for Streptomyces. The clusters of Phyllobacterium species were related to Phyllobacterium ifriqiyense (cluster 1) and Phyllobacterium sophorae (cluster 2). Besides the strong impact of growth stage and the fact that more isolates were assigned to cluster one than to cluster two, the abundances of the Phyllobacterium clusters showed divergent patterns between the two farming systems. Cluster one was more abundant in CF than in OF, while cluster two showed a contrary trend (Figure 1C and Supplementary Table S6). Phylogenetic cluster distributions of the taxa within the genus Pseudomonas and Streptomyces were more heterogeneous. Pseudomonas clusters grouped with Pseudomonas poae (clusters 1, 8, and 10), Pseudomonas corrugate (cluster 2), Pseudomonas jessenii (cluster 4 and 7), Pseudomonas fluorescens (cluster 5), and Pseudomonas koreensis (cluster 9) (Figure 1C). Streptomyces cluster 1 contained the highest number of isolates and grouped with clusters 7 and 8, and reference sequence Streptomyces ciscaucasicus (Figure 1C). The other less abundant Streptomyces clusters grouped with reference sequences of Streptomyces graminilatus (cluster 2), Streptomyces achromogenes (clusters 3 and 4), Streptomyces peruviensis (cluster 5), and Streptomyces ryensis (cluster 6) (Figure 1C). Effects of climate and land use treatments on Streptomyces clusters could not be detected, but for Pseudomonas cluster 1. Representatives of this cluster were approximately twice as often isolated from CF and plots with ambient climatic conditions compared to OF and plots with future climatic conditions (Supplementary Table S6).

Dispersion patterns for *Phyllobacterium*, *Pseudomonas* and *Streptomyces* clusters significantly changed over the growing season (p = 0.001) (**Figure 2A**). At the genus level, compositional shifts were observed among the taxa of the genus *Pseudomonas*

(p = 0.006) and *Phyllobacterium* (p = 0.04), but not for *Streptomyces* (**Supplementary Figure S3**). At cluster level, a shift in the cultivable fraction of the community was detected only for clusters of genus *Pseudomonas* (p = 0.02). Species richness (**Figure 2B**) and Shannon-Weaver diversity index (**Figure 2C**) were found to be higher at stem elongation stage of May samples compared to grain filling stage of July samples with p = 0.02 and p = 0.04, respectively. Effects of climate or land use on dispersion patterns and diversity indices could not be observed.

P Solubilization Potentials of Key Taxa

Phosphate-solubilizing potentials of dominant taxa within the genera Phyllobacterium, Pseudomonas, and Streptomyces isolated from the wheat rhizospheres at the stem elongation stage in May, as well as at the grain filling stage in July are presented for PSI 1 and PSI 2 in Figures 3A,B, respectively. Among the taxa of the three genera, Pseudomonas species expressed highest P solubilization potentials irrespective of employed indices, PSI 1 (Figure 3A) and PSI 2 (Figure 3B), and at the two growth stages studied. Phyllobacterium species showed approximately half of the P solubilization potential of the Pseudomonas species at both growth stages independently of the used P solubilization index. In contrast, the assessment of the P solubilization potentials of Streptomyces species depended on the index used. Excluding the colony size for the assessment of P solubilization, as done in PSI 2, clearly revealed significantly lower P solubilization potentials of Streptomyces species compared to taxa of the two other genera (p < 0.001) (Figure 3B).

Regarding the effect of studied impact factors within the three dominant genera, neither plant growth stage (Figure 3), nor climate or land use affected the P solubilization potential of *Phyllobacterium* and *Streptomyces*. However, interaction effects of climate and land use were found for *Pseudomonas* species, which showed higher P solubilization potentials under future climatic conditions compared to ambient climatic conditions in CF (p < 0.001 for PSI 2) (Supplementary Figure S4). Furthermore, under future climatic conditions, P solubilization potentials of *Pseudomonas* species were significantly higher in CF compared to OF (p < 0.001 for both indices, Supplementary Figure S4).

The phylogenetic cluster-specific, mean P solubilization potentials of the genera *Phyllobacterium* and *Streptomyces* revealed consistent P solubilization potentials in either employed index (Figure 1C and Supplementary Table S6). In contrast to that, the average P solubilization potentials of the individual clusters assigned to the genus *Pseudomonas* showed a huge variance (Figure 1C and Supplementary Table S6). While, *Pseudomonas* cluster 6 showed a low P solubilization potential, the isolates assigned to cluster 8 expressed the highest average P

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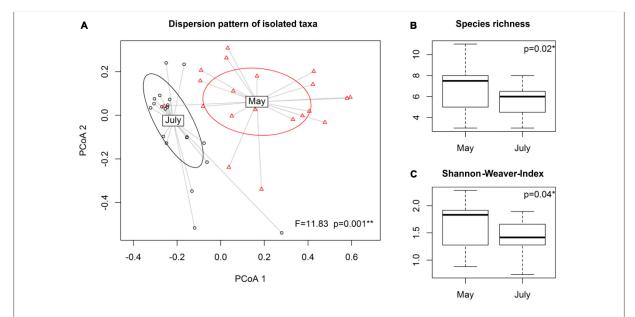


FIGURE 2 | Effect of time on composition of isolated rhizobacterial taxa among and within the dominating genera *Phyllobacterium*, *Pseudomonas* and *Streptomy*ces. Running betadisper function based on Bray-Curtis dissimilarity (A) dispersion patterns of the isolated taxa are presented in an ordination plot. Diversity indices (B) species richness and (C) Shannon-Weaver-Index were calculated, followed by ANOVA, testing for differences in isolated species composition of the three dominant genera between the two sampling points, May (stem elongation stage, BBCH 37-39) and July (grain filling stage, BBCH 75-77) 2015, respectively.

solubilization potential among all clusters of all three dominant genera. Even more important, the most abundant *Pseudomonas* cluster (cluster 1, represented by 39 isolates = 23% of all *Pseudomonas* isolates) showed a high average P solubilization potential which exceeded mean P solubilization of all genera and clusters, besides the aforementioned *Pseudomonas* cluster 8 (Figure 1C and Supplementary Table S6).

Drought Tolerance of Key Taxa

Drought stress tolerance was quantified as relative difference in colony size on PEG medium compared to control medium. Thus, higher drought tolerance was reflected by higher maintenance of colony size (maximum 100%) and vice versa. Pseudomonas species (35.1% at stem elongation in May and 35.7% at grain filling stage in July) and Streptomyces species (43.3% in May and 46.7% in July) were less resistant to drought compared to species assigned to the genus Phyllobacterium (54.1% in May and 63.6% in July) (Table 1). Although the low number of isolates at stem elongation in May samples prevents from drawing significant conclusions, species drought tolerance of genus Phyllobacterium was improved in the course of the growing season (Table 1). Total drought tolerance of taxa belonging to the genera Pseudomonas and Streptomyces did not change throughout the observed growth stages, but show different behavior under ambient and future climatic conditions. Under ambient climate an increase and under future climate conditions a decrease of drought tolerance could be observed for Pseudomonas species, while for Streptomyces species the opposite pattern showed up (Table 1). Land use did not impact key genera's drought tolerance, but the climate treatment affected average drought tolerance of the genus Streptomyces. Strains isolated from plots with future climatic conditions showed higher drought tolerance (52.4%) as compared to those isolated from plots with ambient climate (38.6%) (p=0.01). This trend was also observed at the cluster level, even though it was not significant due to the low number of isolates per cluster (**Supplementary Figure S5**). In contrast to streptomycetes, climate treatment showed no effect on the drought tolerance of taxa within the genera *Phyllobacterium* and *Pseudomonas*.

At the cluster level, a marginal higher drought tolerance of the species isolated from plots with future climatic conditions, compared to those isolated from ambient plots was observed for *Phyllobacterium* cluster 2 (**Supplementary Figure S5**). Within the genus *Pseudomonas*, we observed no increase in drought tolerance of species from future climate plots except for the low abundant cluster 10 and, to a lower extent, for cluster 3 and cluster 8 (**Supplementary Figure S5**). Overall, species of both *Phyllobacterium* clusters exhibit a significantly higher drought tolerance as compared to those of the predominant clusters of *Pseudomonas* (clusters 1 to 4) to and *Streptomyces* species (clusters 1 to 4) (**Supplementary Figure S5** and **Supplementary Table S6**).

DISCUSSION

In this study we have shown that members of the genera *Phyllobacterium*, *Pseudomonas* and *Streptomyces* are predominant within the cultivable fraction of the P-solubilizing

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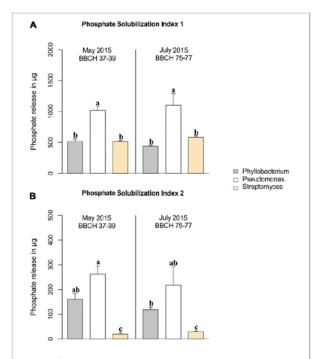


FIGURE 3 | Phosphate-solubilizing potentials of the three dominant genera. Comparison of the two indices, (A) PSI 1 and (B) PSI 2 for Phyllobacterium, Pseudomonas und Streptomyces isolates at stem elongation in May and at grain filling in July 2015 sampled from wheat rhizosphere. Standard errors are given in the bars. Significant differences were calculated for the both indices separately and are marked by different letters (ANOVA and Tukey HSD Test).

TABLE 1 | Drought resistance of the members of three dominant P-solubilizing genera.

	Phyllobacterium		Pseudomonas		Streptomyces	
	Мау ВВСН 37-39	July BBCH 75–77	May BBCH 37–39	July BBCH 75–77	Мау ВВСН 37–39	July BBCH 75–77
Ambient	55.9% ab (6.4)	62.5% a (3.1)	34.2% b (3.9)	55.2% ab (15.2)	39.6% b (4.8)	37.8% b
Future	53.2% ab (4.0)	64.4% a (2.5)	36.2% b (3.9)	20.2% b (16.3)	47.8% ab (6.1)	54.9% a (4.7)
Total	54.1% de (3.4)	63.6% d (2.0)	35.1% f (2.7)	35.7% ef (12.3)	43.3% ef (3.8)	46.7% e (3.2)

Mean levels of drought resistance exhibited by the bacterial isolates from grain filling stage (BBCH 37-39) in May and stem elongation stage (BBCH 75-77) in July under ambient and future climatic conditions. Standard error is given in parenthesis. Significant differences according to Tukey post hoc test are marked by different letters for ambient/future (a-b) and total (d-f), respectively.

community in the wheat rhizosphere along both growth stages studied, namely vegetative biomass production (BBCH 37–39, stem elongation) and generative biomass production stages (BBCH 75–77, grain filling). All taxa within the three dominant genera showed high resilience against differences

in studied agricultural management practices and climatic conditions, but structure shifted with wheat growth stages. This shift was mainly driven by changes among P-solubilizing taxa of the genera Pseudomonas and Phyllobacterium, as well as on single taxa level within the genus Pseudomonas. The considerable in vitro P solubilization potential of both genera Pseudomonas and Phyllobacterium pleads for functional redundancy upon structural changes within the cultivable fraction of the P-solubilizing community. While the genera Pseudomonas and Streptomyces comprise well-known P solubilizers, Phyllobacterium species have rarely been associated with P solubilization. Considering their medium-borne P solubilization potential in combination with an accentuated tolerance of actively growing colonies to drought conditions, Phyllobacterium species may play an important role in P supply for plants under drought.

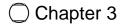
Wheat Growth Stage and Farming System Drive Abundance and Identity of Phosphate-Solubilizing Rhizobacteria

The abundance of cultivable P-solubilizing rhizobacteria (PSR), mirrored by CFU per gram dry soil, was higher at the grain filling stage of wheat in July than at the stem elongation stage in May and contradicts our assumptions made in the first hypothesis. This discrepancy might be related to higher soil moisture at the day of sampling in July, since increased water availability supports microbial growth and enhances metabolic activities (Williams and Rice, 2007). Due to the realistic scenario-based approach of the GCEF, the soil moisture is mainly dependent on the actual weather. However, the impact of the climate treatment was clearly evident for the annual soil moisture dynamics of 2015 (Schädler et al., 2019) and led to significantly reduced winter wheat yields under future climatic conditions compared to ambient ones (Supplementary Table S3).

According to our third hypothesis, the conventional farming (CF) showed higher PSR numbers than the organic farming (OF), which is likely related to the higher nutrient availability in CF. Physicochemical soil parameters are well known as important drivers of microbial biomass (Widmer et al., 2006) as well as of the community structure and composition (Cleveland and Liptzin, 2007; Lauber et al., 2008; Francioli et al., 2016). Nevertheless, we consistently identified Phyllobacterium, Pseudomonas and Streptomyces species as the predominant cultivable PSR in both farming systems. In general, plants recruit beneficial PGPR from the surrounding soil to their rhizospheres (Badri and Vivanco, 2009) by plant specific rhizodeposition patterns (Paterson et al., 2007). This selection effect may mask environmental impacts on the community structure and composition in the rhizosphere and, thus, explain the low differences in the identity of predominant PSR isolated across both farming systems.

Phosphate-Solubilizing Potential First Time Reported for *P. ifriqiyense* and *P. sophorae*

Multiple studies report Streptomyces (Molla et al., 1984; Jog et al., 2014) and Pseudomonas species to be strong P



solubilizers (e.g., Richardson et al., 2009; Marasco et al., 2012; Sharma et al., 2013) and successful colonizers of the wheat rhizosphere (Mohammadi, 2012; Jog et al., 2014). Conversely, members of the genus Phyllobacterium, taxonomically classified to the order Rhizobiales, are hardly described for their P solubilization potential, but rather for their biocontrol activities (Lambert et al., 1990; Aisyah et al., 2017) or the capacity to fix nitrogen (Rojas et al., 2001). Our study now provides evidence for the potential of two Phyllobacterium species, namely P. ifriqiyense and P. sophorae, to in vitro solubilize P from its bound mineral form in the rhizosphere of winter wheat in the temperate climatic zone. The two strains have first been isolated from root nodules of Astragalus algerianus and Lathyrus numidicus in South Tunisia - P. ifriqiyense - (Mantelin et al., 2006), and of Sophora flavescens in China - P. sophorae - (Jiao et al., 2015). Only two further previous studies indicate the potential of Phyllobacterium myrsinacearum for P solubilization in subtropical soils (Chen et al., 2006) and in metal-polluted soils (Ma et al., 2013). Genome shotgun sequencing of Phyllobacterium sp. (National Center of Biotechnology Information, NCBI) revealed quinoprotein glucose dehydrogenase coding genes. These genes are involved in gluconic acid production and, thus, suggest a potential for mineral phosphate solubilization (Rodríguez and Fraga, 1999). This coding gene has also been indicated for Pseudomonas species (Miller et al., 2010) and Streptomyces species (Jog et al., 2014) and may, therefore, serve as indicator for P solubilization potential of Phyllobacteria in vivo. Therefore, the contribution of Phyllobacterium species to phosphate solubilization in soils and rhizospheres might be still underestimated.

Complementarity Between Phyllobacterium and Pseudomonas at Different Wheat Growth Stages

Abundances of P-solubilizing Phyllobacterium and Pseudomonas isolates were complementary between the two wheat growth stages. Since an isolation-based approach was used, this complementarity mirrors different proportions of active microorganisms. Blagodatskaya and Kuzyakov (2013) suggested that more than 95% of the total microbial biomass are inactive at a given time point. Accordingly, active growth and dormancy phases of rhizosphere species are dependent on nutrient levels, interspecific competition as well as on the plant growth stage and its related rhizodeposition (Blagodatskaya and Kuzyakov, 2013). Plant nutrient uptake and root exudation were reported to be high at stem elongation stage and considerably lower at grain filling (Aulakh et al., 2001; Malhi et al., 2006). Of note, Pseudomonas species were predominantly isolated from the wheat rhizosphere at the stage of stem elongation in May. Members of this genus are well characterized for their rapid growth, high chemotactic activity and motility toward root exudates (Scher et al., 1988; Walsh et al., 2001), as well as the production of secondary metabolites (Haas and Defago, 2005) and secretion of siderophores (Kloepper et al., 1980), to metabolize rhizodeposits and outcompete other microorganisms (Paulsen et al., 2005). Indeed, root colonization of active Phyllobacterium species may be inhibited by pseudomonads competing for root exudates at the stage of high biomass production in May. Even though *Phyllobacterium* species are described to live in a large variety of habitats and to "communicate" with plant tissues, the group is also known to be non-pathogenic (Mantelin et al., 2006). When plants reach their mature state, the release of root exudates and competition between microorganisms in the rhizosphere decreases drastically (Hamlen et al., 1972), likely explaining the decline in abundance of *Pseudomonas* and the enrichment of *Phyllobacterium* species at the grain filling stage. Antagonistic effects have not been tested within this study, but may provide important information on colonization success in plant rhizospheres at different growth stages.

Resistance and Functional Redundancy in Terms of P Solubilization Between and Among Dominant Phosphate Solubilizing Genera

Besides abundance shifts, the cluster dispersion patterns of the cultivated dominant taxa varied significantly between the two time points with increased species richness at the stage of stem elongation in May samples. This partly supports our first hypothesis, since the diversity decrease in July is in line with reduced carbon and nutrient availabilities (Hamlen et al., 1972). The shifts within the cultivable fraction of the PSR community were based on two different responses: changes in the dominant PSR taxa, which are mainly driven by shifts within the genera Phyllobacterium and Pseudomonas, as well as altered species composition of Pseudomonas species. Changes in the structure and identity of Pseudomonas species at different wheat growth stages indicate that upon a change in the environmental conditions, the culture derived Pseudomonas community is dynamic. Concurrent changes in community structure and composition according to plants growth stages are widely observed in ecology (Dunfield and Germida, 2003; Wolsing and Priemé, 2004; Francioli et al., 2018). As we found moderate to high average P solubilization potentials of taxa classified to the genera Pseudomonas and Phyllobacterium at the two growth stages studied, the observed shifts plead for functional redundancy (Allison and Martiny, 2008) between and among genera at different wheat growth stages.

Besides the effect of different growth stages over the growing season, an effect of farming system was found. Increased P solubilization potentials for *Pseudomonas* isolates from CF, compared to OF, were detected under future climatic conditions. The mechanisms behind this observation cannot be revealed within this study, but we assume soil nutrient availability to play a role in the activity of PSR under drought. Interestingly, this finding is, however, supported by a previous study performed in grasslands, where we found higher P solubilization potentials of *Pseudomonas* species under low soil moisture conditions (Breitkreuz, unpublished). Within this study, we found comparable P solubilization potentials expressed by the predominant taxa along both time points, which is in contrast to our assumptions made in hypothesis one that P solubilization potential is higher at the stem elongation stage. Higher activities

Rhizobacteria Along Wheat Growth Stages

of *Pseudomonas* species in CF under future climate contradicts hypothesis three expecting higher P solubilization potential in OF than in CF.

The abundances of isolated Streptomyces species, but also the structure and composition within the genus were only slightly affected by the wheat growth stage and not at all by the climate or land use treatments. Despite their high abundance, P solubilization potential expressed according to PSI 2 were significantly less pronounced compared to those of the isolated *Phyllobacterium* and *Pseudomonas* species. However, Battini et al. (2017) recently emphasized the capability of Streptomyces species to facilitate P uptake into fungal hyphae in root free compartments, which is subsequently transported via the mycorrhizosphere to maize plants. In fact, Tarkka et al. (in preparation) isolated phosphate solubilizing bacteria from bulk soils of conventional and organic farming plots of the GCEF and identified Streptomyces species among the most abundant phosphate solubilizes. Their lower phosphate solubilizing potentials may be caused by slower development rates and delayed activity expression, compared to the mainly fast growing Phyllobacterium and Pseudomonas species (Garrity et al., 2007). The slow growth of streptomycetes is related to the formation of complex structures, e.g., branching hyphal filaments and spores (Chater et al., 2010). Dormant spores in the soil are beneficial to endure severe drought events waiting for more favorable conditions (Viaene et al., 2016). In line with this, recently published studies observed accumulation of actinobacteria in drought-treated soils and rhizospheres of different plants (Bouskill et al., 2016; Taketani et al., 2017).

For estimating the strain-specific P solubilization potentials, two complementary indices were calculated. PSI 1, which is based on the total colony and halo size, is directly related to the total amount of released P, but the values are strongly biased by the respective colony size. Consequently, identical distances between halo and colony rim along with larger colony sizes may overrate the activity potential since bigger colonies have higher area values and vice versa. The same problem occurs when using the commonly published ratios, halo to colony size (e.g., Nguyen et al., 1992; Kumar and Narula, 1999; Pérez et al., 2007), However, the diffusion expanse of organic acids in the surrounding of bacterial colonies to solubilize bound P is mainly determined by the outermost cells of the colony. PSI 2 remedies the strong impact of the colony size that, in our case for Streptomyces species, may impact appraisal of the P solubilization potential. We thus suggest to focus on approaches based on diffusion expanse (i.e., the distance between halo and colony) for estimation of phosphate solubilizing potential when using plate bioassays.

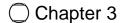
Drought Tolerance of P-Solubilizing Bacteria

The tolerance of the predominant taxa against severe drought stress was quantified in this study by water deficit bioassays with polyethylene glycol 6000 (PEG). Using this approach, we focused on whether isolated strains remained physiologically active under low moisture conditions. In this respect, evaluating growth on nutrient agar with PEG may provide information on active PSR under drought conditions.

As assumed in our second hypothesis, we consistently observed a (by trend) increased drought tolerance of the dominant taxa isolated at stage of grain filling from July as compared to those at stem elongation from May. This indicates enhanced adaptive bacterial responses against dehydration stress in summer. While we detected no land use effect on drought tolerance, we found a higher drought tolerance for Streptomyces species, when isolated from plots under future climatic conditions, compared to those from ambient climate plots. Overall, we observed the highest drought tolerance for Phyllobacterium species, followed by Streptomyces and consecutively Pseudomonas species. Soil microorganisms tolerate water stress in a variable extent (Manzoni et al., 2012). As a most common mitigation strategy, bacteria as well as eukaryotes accumulate low molecular weight compounds, so called osmolytes, once water potentials decrease (Kempf and Bremer, 1998). These compounds do not interfere with cellular functions, but stabilize cellular structures and prevent desiccation. Their production and accumulation in the abundant clusters observed within this study is established, for Phyllobacterium (Fujihara, 2008; Sagot et al., 2010; Zaprasis et al., 2015), but also for Pseudomonas species, when isolated from arid or semi-arid regions of India (Sandhya et al., 2009). However, Lennon et al. (2012) performed a study in (MI, United States, humid region) and found with decreasing water potentials that respiration of Pseudomonas species drastically dropped, which supports the findings for Pseudomonas in our study.

Field Application Potential of P-Solubilizing Bacteria

Besides rhizosphere competence, the trait combination of high P solubilization potential and drought tolerance is a prerequisite for field applications. Inoculation with microbial species of the genera Bacillus and Paenibacillus (Zhang J. et al., 2012; Kasim et al., 2013; Kumar et al., 2014; Timmusk et al., 2014), as well as Azospirillum and Pseudomonas (Creus et al., 2004; Naiman et al., 2009; Ilyas and Bano, 2010; Kasim et al., 2013) and Streptomyces (Jog et al., 2014) was found to increase wheat yields in field and pot experiments. In particular, Pseudomonas species are well investigated, as the genus comprises species with promising plant growth promoting properties (Burr et al., 1978; Naiman et al., 2009; Rajkumar et al., 2010). In contrast, prior to this study little was known about P-solubilizing Phyllobacterium species, or their drought tolerance in wheat rhizosphere. They do act as PGPR e.g., in canola plants (Bertrand et al., 2001), but underlying mechanisms remain unclear. Model studies with Arabidopsis thaliana and Phyllobacterium brassicacearum indicated higher drought tolerance (Bresson et al., 2014). In the same context, Kechid et al. (2013) observed a delay in reproduction and lowered transpiration rate of Arabidopsis thaliana upon inoculation with Phyllobacterium brassicacearum accompanied by stimulated lateral root and root hair growth. Within this study we could prove in vitro potential for P-solubilization of two taxa in the genus Phyllobacterium, which has to be further justified in vivo under different biotic and abiotic conditions. To check applicability of these species as



phosphate delivering biofertilizers, the timing of inoculation and bacterial level has to be defined in pretests to ensure survival of the inoculated strains, but also to keep viability of the seeds. Bashan (1986) found inoculation at early growth stages of wheat with 105-106 colony forming units per ml to be optimal for Azospirillum and Pseudomonas strains to increase productivity. The success of colonization depends on different factors and is favored by high mobility of the bacteria along the growing root, rapid multiplication and the ability to colonize the rhizoplane and inner root tissues (Höflich et al., 1995). Therefore, the colonization potential for different plant root systems, the effect of different soil parameters and antagonistic effects within the native microbial community has to be quantified in inoculation experiments under natural (non-sterile) and controlled (sterile) conditions (Gaskins et al., 1985; Höflich et al., 1995). Finally, the strains have to be tested for their plant growth promoting activities exploiting various sources of insoluble P in the soil, besides tri-calcium phosphate also other insoluble inorganic and organic phosphate sources (Bashan et al., 2013), and their drought stress tolerance under natural conditions in pot and field experiments. Nevertheless, the non-pathogenic status and high plant-interaction potential of Phyllobacterium species, along with the considerable P solubilization potential in vitro and high drought tolerance revealed in this study, make this group attractive for possible future application as plant growth promoting inoculants for wheat production.

CONCLUSION

Our findings suggest that the bacterial taxa involved in the provision of P in the rhizosphere of wheat plants depend on plant's growth stage and only minor on farming system and climate under in vitro conditions. This observation clearly pleads for more studies at defined and yield-determining stages of plant growth under in vitro and in vivo conditions. We found that Pseudomonas species were dominant at the stem elongation stage, i.e., the peak of vegetative biomass production, while Phyllobacterium species dominated at the grain filling stage (generative stage). This is the first time that the two isolated Phyllobacterium species are reported for their high P solubilization potential. Drought tolerance potentials of the three dominant genera were found to be increased at the second growth stage studied in July compared to the first one in May indicating an adaptation to drought over time and a higher tolerance in summer months. Phyllobacterium species expressed highest drought tolerance potential among isolated dominant phosphate solubilizing rhizobacteria. Therefore, further work should concentrate on functionality and abundance of this genus in rhizospheres of plants. Especially in the context of climate change related summer drought, the observed trait combination of Phyllobacterium species in this study may be of particular importance to adapt agriculture to dryer conditions in the future.

DATA AVAILABILITY STATEMENT

The raw data supporting the conclusions of this manuscript will be made available by the authors, without undue reservation, to any qualified researcher.

AUTHOR CONTRIBUTIONS

All authors conceived and designed the experiments, obtained funding, interpreted the results, and contributed to revisions and approved submission of the manuscript. TR and CB performed the field experiments. CB performed the laboratory works and data analysis. MT, TR, and CB wrote the manuscript with input from FB.

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SUPPLEMENTARY MATERIAL

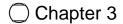
The Supplementary Material for this article can be found online at: https://www.frontiersin.org/articles/10.3389/fmicb. 2019.03109/full#supplementary-material

Chapter 3

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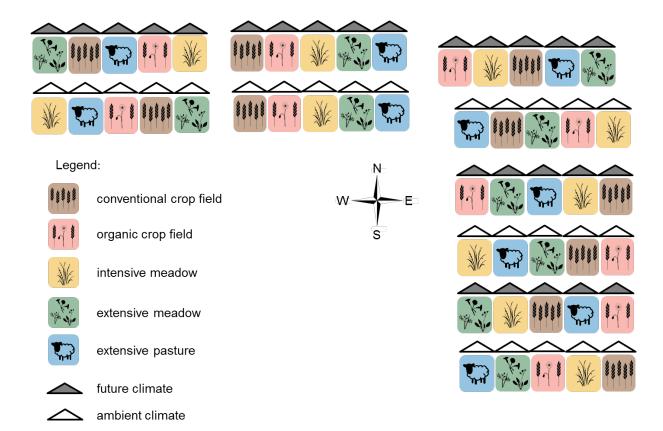
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Conflict of Interest: The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

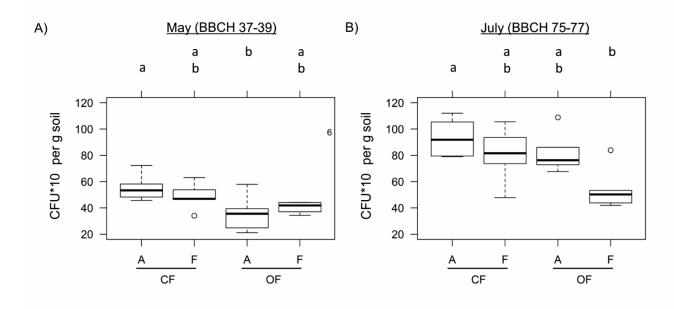
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Supplementary Material:

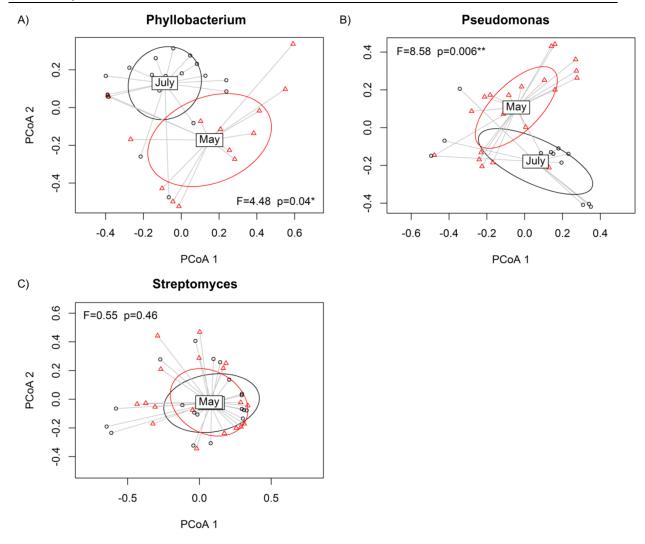
Supplementary Figures



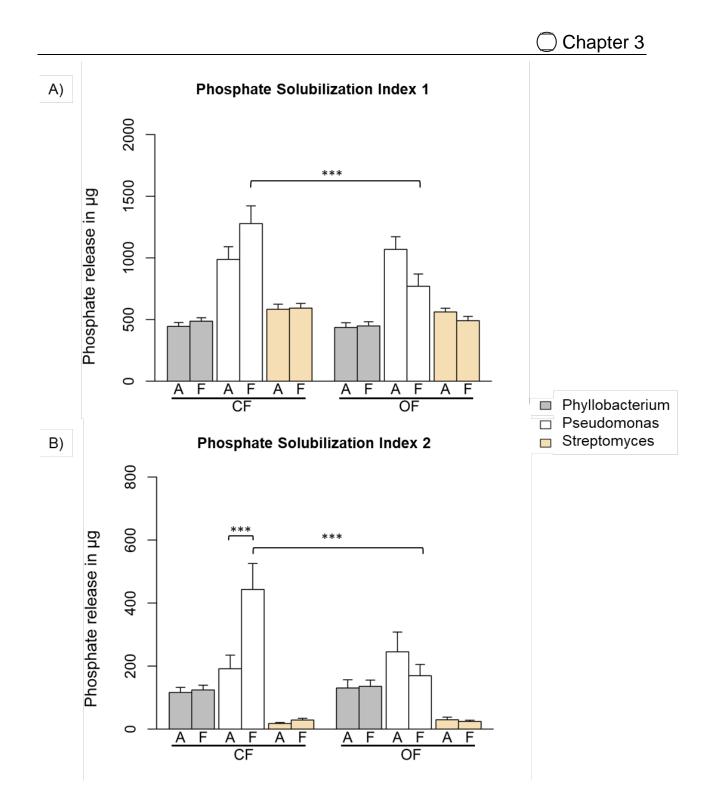
Supplementary Figure 1: Experimental design of the Global Change Experimental Facility (GCEF). The facility includes 50 plots (subplots) arranged in ten blocks (mainplots). Five blocks are attributed to ambient and future climate, respectively, whereby each block comprises all five land use types. Copyrights of land use icons are held by Gottschall/Siebert (Schädler et al. 2019).



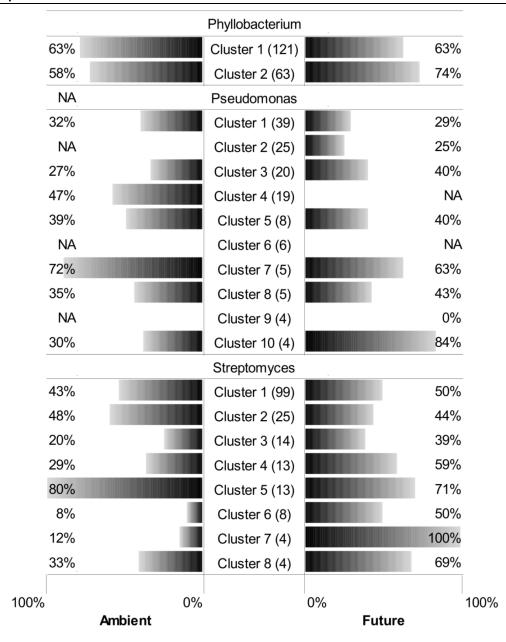
Supplementary Figure 2: Abundance of phosphate solubilizing rhizobacteria. Colony forming units of wheat rhizobacteria per g of soil, grown on Pikovskaya agar plates, in A) May 2015 sampling at BBCH stage 37-39 and in B) July 2015 sampling, at BBCH stage 75-77. Significant differences between treatments - ambient (A) and future (F) climate; conventional (CF) and organic (OF) farming system - for each time point are marked by different small letters.



Supplementary Figure 3: Effect of wheat growth stage on A) *Phyllobacterium*, B) *Pseudomonas* and C) *Streptomyces* isolates structure. Dispersion patterns running betadisper function were calculated on distance matrix using method of Bray Curtis, followed by ANOVA testing for differences in composition of isolated species between the two sampling points, May (stem elongation, BBCH 37-39, red triangles) and July (grain filling, BBCH 75-77, black circles) respectively.



Supplementary Figure 4: Impact of climate and farming system on phosphate solubilization potentials of the three most dominant genera. Potentials are given for *Phyllobacterium*, *Pseudomonas* and *Streptomyces* calculated with A) PSI 1 and B) PSI 2, and isolated from conventional farming (CF) and organic farming (OF) under ambient (A) and future (F) climate conditions. Significant interactions within each genus are marked with p<0.001***.



Supplementary Figure 5: Drought resistance under ambient and future climate conditions among the predominant phylogenetic clusters of *Phyllobacterium*, *Pseudomonas* and *Streptomyces*. Number of isolates for each cluster is given in parenthesis. Respective clusters marked with NA contain isolates that did not grow, neither on control (YME) nor on test media (YME+PEG).

Supplementary Tables

Supplementary Table 1: Total number, origin and description of isolates cultivated on Pikovskaya medium. Isolates were gained at two different sampling points (May and July 2015), from two different farming systems (CF-conventional farming, OF-organic farming) and climate treatments (A-ambient climate, F-future climate). Number of base pairs (bp), assignment to genus and NCBI (National Center for Biotechnology Information-gene data bank) number is given.

Isolates	Sampling	Genus	Size	Land	Climate	Acc. Number
			(bp)	use		(NCBI)
PN2-B01P2-1	May 2015	Pseudomonas sp.	1086	CF	Α	MK637864
PN2-B01P2-2	May 2015	Pseudomonas sp.	1057	CF	Α	MK637859
PN2-B01P2-3	May 2015	Pseudomonas sp.	1057	CF	Α	MK637861
PN2-B01P2-4	May 2015	Pseudomonas sp.	1050	CF	Α	MK637884
PN2-B01P2-5	May 2015	Streptomyces sp.	1053	CF	Α	MK638553
PN2-B01P2-6	May 2015	Pseudomonas sp.	1045	CF	Α	MK637871
PN2-B01P2-7	May 2015	Variovorax sp.	1051	CF	Α	MK638652
PN2-B01P2-8	May 2015	Agrobacterium sp.	1078	CF	Α	MK638040
PN2-B01P2-9	May 2015	Pseudomonas sp.	1057	CF	Α	MK637870
PN2-B01P2-	May 2015	Pseudomonas sp.	1057	CF	Α	MK637882
10						
PN2-B01P2-	May 2015	Pseudomonas sp.	1057	CF	Α	MK637867
11						
PN2-B01P2-	May 2015	Pseudomonas sp.	1057	CF	Α	MK637869
12						
PN2-B01P2-	May 2015	Dyella sp.	1063	CF	Α	MK638116
13						
PN2-B01P2-	May 2015	Dyella sp.	1059	CF	Α	MK638115
14					_	
PN2-B01P2-	May 2015	Mesorhizobium sp.	1011	CF	Α	MK638184
15		- '	4055	0=	_	
PN2-B01P2-	May 2015	Pseudomonas sp.	1057	CF	Α	MK637865
16	NA- 2045		4026	C.F.	•	N4V620206
PN2-B01P2-	May 2015	Rhizobium sp.	1036	CF	Α	MK638386
17	N40., 2015	N 4 a a a shi = a hi a a a	1027	CE	۸	N4VC204E4
PN2-B01P2- 18	May 2015	Mesorhizobium sp.	1037	CF	Α	MK638154
PN2-B01P2-	May 2015	Ctrantamycas sn	1023	CF	۸	MK638528
19	May 2015	Streptomyces sp.	1023	CF	Α	IVINUSOSZO
PN2-B01P2-	May 2015	Mesorhizobium sp.	1036	CF	Α	MK638159
20	IVIAY 2013	wiesornizobium sp.	1030	Ci	^	WIK038133
PN2-B01P4-1	May 2015	Pseudomonas sp.	1057	OF	Α	MK637877
PN2-B01P4-2	May 2015	Massilia sp.	1062	OF	A	MK638153
PN2-B01P4-3	May 2015	Pseudomonas sp.	1056	OF	A	MK637855
PN2-B01P4-4	May 2015	Rhizobium sp.	1033	OF	A	MK638383
PN2-B01P4-5	May 2015	Pseudomonas sp.	1057	OF	A	MK637858
PN2-B01P4-6	May 2015	Streptomyces sp.	1018	OF	A	MK638619
PN2-B01P4-7	May 2015	Streptomyces sp.	1018	OF	A	MK638454
FINZ-DUIP4-/	iviay 2013	streptomytes sp.	1023	UF	A	IVINU30434

PN2-B01P4-8	May 2015	Streptomyces sp.	1019	OF	Α	MK638482
PN2-B01P4-9	May 2015	Streptomyces sp.	1086	OF	Α	MK638574
PN2-B01P4-	May 2015	Streptomyces sp.	997	OF	Α	MK638597
10						
PN2-B01P4-	May 2015	Streptomyces sp.	1018	OF	Α	MK638618
12						
PN2-B01P4-	May 2015	Mesorhizobium sp.	1037	OF	Α	MK638157
13						
PN2-B01P4-	May 2015	Streptomyces sp.	1049	OF	Α	MK638545
14						
PN2-B01P4-	May 2015	Mesorhizobium sp.	1037	OF	Α	MK638158
15						
PN2-B01P4-	May 2015	Burkholderia sp.	1043	OF	Α	MK638086
16		_			_	
PN2-B01P4-	May 2015	Streptomyces sp.	924	OF	Α	MK638646
17		5 1 1 1 1 1 1		0.5	_	
PN2-B01P4-	May 2015	Bradyrhizobium	600	OF	Α	MK638085
18	N4 204 F	sp.	662	0.5	۸	NAVC204.05
PN2-B01P4-	May 2015	Mesorhizobium sp.	662	OF	Α	MK638185
19 PN2-B01P4-	May 2015	Dhyllohactorium	716	OF	Α	MK638374
20	May 2015	Phyllobacterium	746	UF	А	IVIK038374
PN2-B01P4-	May 2015	sp. Rhizobium sp.	1033	OF	Α	MK638385
21	Way 2015	niiizobiuiii sp.	1055	OF	А	IVINUSOSOS
PN2-B01P4-	May 2015	Streptomyces sp.	1080	OF	Α	MK638468
22	Way 2013	streptomyces sp.	1000	Oi	^	WIN030408
PN2-B01P4-	May 2015	Phyllobacterium	1037	OF	Α	MK638203
24	,	sp.		Ο.		
PN2-B02P3-1	May 2015	Pseudomonas sp.	1061	OF	F	MK637881
PN2-B02P3-2	May 2015	Stenotrophomonas	1034	OF	F	MK638444
	•	sp.				
PN2-B02P3-3	May 2015	Pseudomonas sp.	1049	OF	F	MK637868
PN2-B02P3-4	May 2015	Pseudomonas sp.	1060	OF	F	MK637872
PN2-B02P3-5	May 2015	Pseudomonas sp.	1053	OF	F	MK637862
PN2-B02P3-6	May 2015	Streptomyces sp.	1085	OF	F	MK638470
PN2-B02P3-7	May 2015	Pseudomonas sp.	1057	OF	F	MK637874
PN2-B02P3-8	May 2015	Streptomyces sp.	1049	OF	F	MK638476
PN2-B02P3-9	May 2015	Arthrobacter sp.	1044	OF	F	MK638053
PN2-B02P3-	May 2015	Pseudomonas sp.	1057	OF	F	MK637879
10		•				
PN2-B02P3-	May 2015	Tardiphaga sp.	1068	OF	F	MK638648
11						
PN2-B02P3-	May 2015	Stenotrophomonas	1087	OF	F	MK638445
12		sp.				
PN2-B02P3-	May 2015	Pseudomonas sp.	1069	OF	F	MK637856
13						
PN2-B02P3-	May 2015	Rhizobium sp.	1077	OF	F	MK638384
14						

PN2-B02P3-							Unapier 3
PNZ-B02P3-	PN2-B02P3-	May 2015	Mesorhizobium sp.	1037	OF	F	MK638156
16	15						
PNZ-B02P3-	PN2-B02P3-	May 2015	Pseudomonas sp.	1057	OF	F	MK637873
17	16						
PN2-B02P3-	PN2-B02P3-	May 2015	Streptomyces sp.	1042	OF	F	MK638575
Name	17						
PN2-B02P3-	PN2-B02P3-	May 2015	Dyadobacter sp.	1015	OF	F	MK638112
19	18						
PNZ-B02P3-		May 2015	Streptomyces sp.	1049	OF	F	MK638452
20 Sp. PN2-B02P3- May 2015 Mesorhizobium sp. 1037 OF F MK638155 22 PN2-B02P4-1 May 2015 Pseudomonas sp. 1069 CF F MK637863 PN2-B02P4-2 May 2015 Pseudomonas sp. 1057 CF F MK637857 PN2-B02P4-3 May 2015 Pseudomonas sp. 1057 CF F MK637880 PN2-B02P4-4 May 2015 Buttiauxella sp. 1060 CF F MK638100 PN2-B02P4-5 May 2015 Buttiauxella sp. 1060 CF F MK638101 PN2-B02P4-6 May 2015 Pseudomonas sp. 1077 CF F MK638102 PN2-B02P4-7 May 2015 Pseudomonas sp. 1057 CF F MK637853 PN2-B02P4-8 May 2015 Pseudomonas sp. 1057 CF F MK637860 PN2-B02P4-1 May 2015 Pseudomonas sp. 1057 CF F MK637860 PN2-B02P4-1							
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22 PN2-B02P4-1 May 2015 Pseudomonas sp. 1069 CF F MK637863 PN2-B02P4-2 May 2015 Pseudomonas sp. 1057 CF F MK637885 PN2-B02P4-3 May 2015 Pseudomonas sp. 1057 CF F MK637880 PN2-B02P4-4 May 2015 Buttiauxella sp. 1060 CF F MK638100 PN2-B02P4-5 May 2015 Buttiauxella sp. 1060 CF F MK638101 PN2-B02P4-6 May 2015 Pseudomonas sp. 1077 CF F MK637853 PN2-B02P4-7 May 2015 Pseudomonas sp. 1057 CF F MK637853 PN2-B02P4-8 May 2015 Pseudomonas sp. 1057 CF F MK637876 PN2-B02P4-9 May 2015 Pseudomonas sp. 1057 CF F MK637878 PN2-B02P4-1 May 2015 Buttiauxella sp. 1060 CF F MK638099 PN2-B02P4-1 May 2015			=		_		
PN2-B02P4-1 May 2015 Pseudomonas sp. 1069 CF F MK637863 PN2-B02P4-2 May 2015 Pseudomonas sp. 1057 CF F MK637875 PN2-B02P4-3 May 2015 Pseudomonas sp. 1057 CF F MK637880 PN2-B02P4-4 May 2015 Buttiauxella sp. 1060 CF F MK637801 PN2-B02P4-5 May 2015 Buttiauxella sp. 1060 CF F MK638101 PN2-B02P4-6 May 2015 Pseudomonas sp. 1077 CF F MK637853 PN2-B02P4-7 May 2015 Pseudomonas sp. 1057 CF F MK637861 PN2-B02P4-8 May 2015 Pseudomonas sp. 1057 CF F MK637860 PN2-B02P4-9 May 2015 Pseudomonas sp. 1063 CF F MK637878 PN2-B02P4-1 May 2015 Pseudomonas sp. 1060 CF F MK638099 PN2-B02P4-1 May 2015 Buttiauxella sp.		May 2015	Mesorhizobium sp.	1037	OF	F	MK638155
PN2-B02P4-2 May 2015 Pseudomonas sp. 1057 CF F MK637875 PN2-B02P4-3 May 2015 Pseudomonas sp. 1057 CF F MK637880 PN2-B02P4-4 May 2015 Buttiauxella sp. 1060 CF F MK638100 PN2-B02P4-5 May 2015 Buttiauxella sp. 1077 CF F MK638101 PN2-B02P4-6 May 2015 Buttiauxella sp. 614 CF F MK637853 PN2-B02P4-7 May 2015 Pseudomonas sp. 1057 CF F MK637856 PN2-B02P4-8 May 2015 Pseudomonas sp. 1057 CF F MK637876 PN2-B02P4-9 May 2015 Pseudomonas sp. 1057 CF F MK637860 10 PN2-B02P4-1 May 2015 Pseudomonas sp. 1063 CF F MK638096 12 PN2-B02P4-1 May 2015 Buttiauxella sp. 1068 CF F MK638099 14 PN2-B02P4-1				1000	0=	_	
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PN2-B02P4-4 May 2015 Buttiauxella sp. 1060 CF F MK638100 PN2-B02P4-5 May 2015 Buttiauxella sp. 1060 CF F MK638101 PN2-B02P4-6 May 2015 Pseudomonas sp. 1077 CF F MK637853 PN2-B02P4-7 May 2015 Buttiauxella sp. 614 CF F MK637876 PN2-B02P4-8 May 2015 Pseudomonas sp. 1057 CF F MK637876 PN2-B02P4-9 May 2015 Pseudomonas sp. 1057 CF F MK637854 PN2-B02P4-10 May 2015 Pseudomonas sp. 1063 CF F MK637860 10 PN2-B02P4-11 May 2015 Buttiauxella sp. 1083 CF F MK638096 12 PN2-B02P4-14 May 2015 Pseudomonas sp. 1060 CF F MK638097 14 PN2-B02P4-14 May 2015 Buttiauxella sp. 1068 CF F MK638097 PN2-B02P4-15 <th></th> <td>•</td> <td>·</td> <td></td> <td></td> <td></td> <td></td>		•	·				
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PN2-B02P4-10 May 2015 Pseudomonas sp. 1057 CF F MK637860 PN2-B02P4-11 May 2015 Pseudomonas sp. 1063 CF F MK637878 11 PN2-B02P4-12 May 2015 Buttiauxella sp. 1083 CF F MK638096 12 PN2-B02P4-13 May 2015 Pseudomonas sp. 1060 CF F MK638009 13 May 2015 Buttiauxella sp. 1068 CF F MK638097 14 PN2-B02P4-14 May 2015 Pantoea sp. 1080 CF F MK638201 15 PN2-B02P4-14 May 2015 Buttiauxella sp. 1068 CF F MK638098 16 PN2-B02P4-14 May 2015 Pseudomonas sp. 1057 CF F MK637866 PN2-B02P4-15 May 2015 Buttiauxella sp. 1073 CF F MK638099 19 PN2-B02P4-14 May 2015 Buttiauxella sp. 1073 CF F MK63809		•	·				
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PN2-B02P4- 11 May 2015 Pseudomonas sp. 1063 CF F MK637878 PN2-B02P4- 12 May 2015 Buttiauxella sp. 1083 CF F MK638096 PN2-B02P4- 13 May 2015 Pseudomonas sp. 1060 CF F MK638009 PN2-B02P4- 14 May 2015 Buttiauxella sp. 1068 CF F MK638097 PN2-B02P4- 15 May 2015 Pantoea sp. 1080 CF F MK638201 PN2-B02P4- 15 May 2015 Buttiauxella sp. 1068 CF F MK638098 16 PN2-B02P4- May 2015 Pseudomonas sp. 1057 CF F MK637866 17 PN2-B02P4- May 2015 Buttiauxella sp. 1069 CF F MK637857 18 PN2-B02P4- May 2015 Buttiauxella sp. 1073 CF F MK638099 19 PN2-B03P1-1 May 2015 Pseudomonas sp. 1038 CF F MK638103 20 PN2-B03P1-1		May 2015	Pseudomonas sp.	1057	CF	F	MK637860
11		NA 2045	Daardamaaaaaa	1002	C.F.	-	N4VC27070
PN2-B02P4- 12 May 2015 Buttiauxella sp. 1083 CF F MK638096 PN2-B02P4- 13 May 2015 Pseudomonas sp. 1060 CF F MK638009 PN2-B02P4- 14 May 2015 Buttiauxella sp. 1068 CF F MK638097 PN2-B02P4- 15 May 2015 Pantoea sp. 1080 CF F MK638201 PN2-B02P4- 15 May 2015 Buttiauxella sp. 1068 CF F MK638098 16 PN2-B02P4- May 2015 Pseudomonas sp. 1057 CF F MK637866 17 PN2-B02P4- May 2015 Buttiauxella sp. 1073 CF F MK637857 18 PN2-B02P4- May 2015 Buttiauxella sp. 1073 CF F MK638099 19 PN2-B02P4- May 2015 Buttiauxella sp. 1038 CF F MK638093 19 PN2-B03P1-1 May 2015 Pseudomonas sp. 1036 CF F MK638103 20 PN2-B03P1-1		May 2015	Pseudomonas sp.	1063	CF	F	IVIK03/8/8
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PN2-B02P4- 14 May 2015 Buttiauxella sp. 1068 CF F MK638097 PN2-B02P4- 15 May 2015 Pantoea sp. 1080 CF F MK638201 PN2-B02P4- 16 May 2015 Buttiauxella sp. 1068 CF F MK638098 16 PN2-B02P4- PN2-B02P4- May 2015 Pseudomonas sp. 1057 CF F MK637866 17 PN2-B02P4- PN2-B02P4- May 2015 Pseudomonas sp. 1069 CF F MK637857 18 PN2-B02P4- PN2-B03P1-4 May 2015 Buttiauxella sp. 1073 CF F MK638099 19 PN2-B03P1-1 May 2015 Buttiauxella sp. 1038 CF F MK638103 20 PN2-B03P1-1 May 2015 Pseudomonas sp. 1057 CF A MK637883 PN2-B03P1-2 May 2015 Sphingomonas sp. 1034 CF A MK638443 PN2-B03P1-3 May 2015 Agrobacterium sp. 604 CF A		Way 2013	r scadomonas sp.	1000	Ci	•	WINOSOOOS
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PN2-B03P1-4 May 2015 <i>Agrobacterium sp.</i> 1032 CF A MK638045		•					
	PN2-B03P1-4	May 2015	Agrobacterium sp.	1032	CF	A	MK638045

PN2-B03P1-5	May 2015	Bacillus sp.	1068	CF	Α	MK638061
PN2-B03P1-6	May 2015	Bacillus sp.	883	CF	Α	MK638063
PN2-B03P1-7	May 2015	Streptomyces sp.	1063	CF	Α	MK638469
PN2-B03P1-	May 2015	Pedobacter sp.	1013	CF	Α	MK638129
10	Widy 2013	readbacter sp.	1013	C.	,,	1411(030123
	May 2015	Duggmanasas	1050	CE	٨	MAKCOOOL
PN2-B03P1-	May 2015	Rugamonas sp.	1059	CF	Α	MK638095
11	2045	C	1000	6 5		N 41/620 452
PN2-B03P1-	May 2015	Streptomyces sp.	1080	CF	Α	MK638453
12						
PN2-B03P1-	May 2015	Rhodococcus sp.	1037	CF	Α	MK638438
13						
PN2-B03P1-	May 2015	Mesorhizobium sp.	1077	CF	Α	MK638160
14						
PN2-B03P1-	May 2015	Phyllobacterium	841	CF	Α	MK638207
15		sp.				
PN2-B03P1-	May 2015	Caulobacter sp.	1069	CF	Α	MK638104
16						
PN2-B03P1-	May 2015	Phyllobacterium	1070	CF	Α	MK638208
17	•	sp.				
PN2-B03P1-	May 2015	Streptomyces sp.	1072	CF	Α	MK638530
18	,	,				
PN2-B03P1-	May 2015	Streptomyces sp.	1087	CF	Α	MK638596
19		ou epioniy des op.	1007	O.	, ,	······································
PN2-B03P1-	May 2015	Rhizobium sp.	1076	CF	Α	MK638389
20	Widy 2015	mizobiam sp.	1070	C.	,,	1411(030303
PN2-B03P1-	May 2015	Rhizobium sp.	1066	CF	Α	MK638391
21	Way 2013	Kilizobiaili sp.	1000	Ci	^	WIK038331
PN2-B03P1-	May 2015	Ctrontomusos en	1018	CF	Α	MK638620
	iviay 2015	Streptomyces sp.	1010	CF	A	WIK036020
22	N4 204F	A sure le sustanti una sur	1054	C.F.	Δ.	N4VC2004C
PN2-B03P1-	May 2015	Agrobacterium sp.	1054	CF	Α	MK638046
23		54: 4:	1000	0=	_	
PN2-B03P1-	May 2015	Rhizobium sp.	1066	CF	Α	MK638394
24						
PN2-B03P1-	May 2015	Streptomyces sp.	1066	CF	Α	MK638529
25						
PN2-B03P2-1	May 2015	Pseudomonas sp.	1072	OF	Α	MK637910
PN2-B03P2-2	May 2015	Pseudomonas sp.	1045	OF	Α	MK637908
PN2-B03P2-3	May 2015	Variovorax sp.	1079	OF	Α	MK638653
PN2-B03P2-4	May 2015	Arthrobacter sp.	1071	OF	Α	MK638057
PN2-B03P2-5	May 2015	Pseudomonas sp.	1056	OF	Α	MK637897
PN2-B03P2-6	, May 2015	Arthrobacter sp.	1060	OF	Α	MK638054
PN2-B03P2-7	May 2015	Arthrobacter sp.	1071	OF	Α	MK638058
PN2-B03P2-8	May 2015	Pseudomonas sp.	1057	OF	Α	MK637902
PN2-B03P2-9	May 2015	Bacillus sp.	1068	OF	A	MK638067
PN2-B03P2-	May 2015	Bacillus sp.	1082	OF	A	MK638067
10	WIGY ZUID	υατιπας σμ.	1002	O1	^	MINOSOOO
PN2-B03P2-	May 2015	Strantomucos en	1004	OF	٨	MARSOESO
	May 2015	Streptomyces sp.	1084	UF	Α	MK638539
11						

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PN2-B03P2-	May 2015	Pseudomonas sp.	1069	OF	А	MK637904
12						
PN2-B03P2-	May 2015	Streptomyces sp.	1090	OF	Α	MK638570
13						
PN2-B03P2-	May 2015	Streptomyces sp.	1090	OF	Α	MK638611
14						
PN2-B03P2-	May 2015	Streptomyces sp.	853	OF	Α	MK638531
15	•	. , ,				
PN2-B03P2-	May 2015	Mesorhizobium sp.	1070	OF	Α	MK638162
16	-	·				
PN2-B03P2-	May 2015	Streptomyces sp.	1066	OF	Α	MK638475
17						
PN2-B03P2-	May 2015	Inquilinus sp.	1043	OF	Α	MK638134
18						
PN2-B03P2-	May 2015	Mesorhizobium sp.	596	OF	Α	MK638182
19						
PN2-B03P2-	May 2015	Streptomyces sp.	1066	OF	Α	MK638474
20	•	. , ,				
PN2-B03P2-	May 2015	Mesorhizobium sp.	1083	OF	Α	MK638161
21	•	·				
PN2-B03P2-	May 2015	Streptomyces sp.	1084	OF	Α	MK638536
22	•	. , ,				
PN2-B03P2-	May 2015	Mesorhizobium sp.	1070	OF	Α	MK638163
23	•	·				
PN2-B04P1-1	May 2015	Pseudomonas sp.	1072	CF	F	MK637891
PN2-B04P1-2	May 2015	Mucilaginibacter	1083	CF	F	MK638193
	,	sp.				
PN2-B04P1-3	May 2015	Streptomyces sp.	1072	CF	F	MK638557
PN2-B04P1-4	May 2015	Streptomyces sp.	1085	CF	F	MK638472
PN2-B04P1-5	, May 2015	Rugamonas sp.	1061	CF	F	MK638090
PN2-B04P1-6	, May 2015	Chitinophaga sp.	1087	CF	F	MK638106
PN2-B04P1-7	, May 2015	Rhizobium sp.	1079	CF	F	MK638388
PN2-B04P1-8	May 2015	Arthrobacter sp.	1057	CF	F	MK638055
PN2-B04P1-9	May 2015	Chitinophaga sp.	1088	CF	F	MK638107
PN2-B04P1-	May 2015	Arthrobacter sp.	1084	CF	F	MK638060
10	ay 2013	memobacce sp.	1001	C.	•	William
PN2-B04P1-	May 2015	Streptomyces sp.	1018	CF	F	MK638621
11	, _5_5			- .	•	
PN2-B04P1-	May 2015	Rhizobium sp.	1061	CF	F	MK638387
12	., ====				•	
PN2-B04P1-	May 2015	Phyllobacterium	1079	CF	F	MK638205
13	,	sp.	· · •		•	
PN2-B04P1-	May 2015	Streptomyces sp.	1090	CF	F	MK638555
14	,				•	
PN2-B04P1-	May 2015	Bradyrhizobium	1071	CF	F	MK638082
15	., ====	sp.	- · -		•	
PN2-B04P1-	May 2015	Rhizobium sp.	1066	CF	F	MK638393
16	., ====				•	

PN2-B04P1-	May 2015	Phyllobacterium sp.	1043	CF	F	MK638206
PN2-B04P1-	May 2015	Streptomyces sp.	1055	CF	F	MK638569
18 PN2-B04P1-	May 2015	Chryseobacterium	1049	CF	F	MK638108
19 PN2-B04P1-	May 2015	sp. Rhizobium sp.	1073	CF	F	MK638392
20	•					
PN2-B04P2-1	May 2015	Pseudomonas sp.	1072	OF	F	MK637894
PN2-B04P2-2	May 2015	Pseudomonas sp.	1085	OF	F	MK637888
PN2-B04P2-3	May 2015	Bacillus sp.	1076	OF	F	MK638065
PN2-B04P2-4	May 2015	Burkholderia sp.	1051	OF	F	MK638094
PN2-B04P2-5	May 2015	Bacillus sp.	1089	OF	F	MK638064
PN2-B04P2-6	May 2015	Streptomyces sp.	1018	OF	F	MK638622
PN2-B04P2-7	May 2015	Pseudomonas sp.	1066	OF	F	MK637901
PN2-B04P2-8	May 2015	Streptomyces sp.	1018	OF	F -	MK638623
PN2-B04P2-9	May 2015	Streptomyces sp.	962	OF	F -	MK638537
PN2-B04P2-	May 2015	Streptomyces sp.	1055	OF	F	MK638573
10	NA- 2045	0	4000	0.5	-	N41/627002
PN2-B04P2-	May 2015	Pseudomonas sp.	1083	OF	F	MK637893
11 DN2 B04B2	May 2015	Decudomonas en	1057	OΓ	_	N4V627902
PN2-B04P2- 12	May 2015	Pseudomonas sp.	1057	OF	F	MK637892
PN2-B04P2-	May 2015	Arthrobacter sp.	1078	OF	F	MK638059
13	iviay 2015	Artinobucter sp.	1076	OF	Г	MINOSOUSS
PN2-B04P2-	May 2015	Mesorhizobium sp.	1054	OF	F	MK638164
14	Widy 2015	wcsomizobiam sp.	1054	O1	•	WIN030104
PN2-B04P2-	May 2015	Streptomyces sp.	1055	OF	F	MK638571
15	,	от оргонну осо орг		•	•	
PN2-B04P2-	May 2015	Streptomyces sp.	1055	OF	F	MK638572
16	•	. , .				
PN2-B04P2-	May 2015	Ochrobactrum sp.	1062	OF	F	MK638197
17						
PN2-B04P2-	May 2015	Streptomyces sp.	1076	OF	F	MK638471
18						
PN2-B04P2-	May 2015	Phyllobacterium	1078	OF	F	MK638204
19		sp.				
PN2-B04P2-	May 2015	Rhizobium sp.	1074	OF	F	MK638390
20		_ , ,, ,, ,			_	
PN2-B04P2-	May 2015	Paraburkholderia	1061	OF	F	MK638087
21		sp.	4050	0.5	_	N.41/62.04.00
PN2-B04P2-	May 2015	Ochrobactrum sp.	1053	OF	F	MK638198
22 DN2 B0ED1 1	May 2015	Decudomonas en	1057	OΓ	٨	N4V627906
PN2-B05P1-1	May 2015	Pseudomonas sp.	1057	OF OF	A	MK637896
PN2-B05P1-2 PN2-B05P1-3	May 2015	Pseudomonas sp.	1056 1086	OF OF	A	MK637900 MK637889
	May 2015 May 2015	Pseudomonas sp.		OF OF	A	
PN2-B05P1-4	•	Pseudomonas sp.	1085	OF OF	A	MK637886
PN2-B05P1-5	May 2015	Pseudomonas sp.	1086	OF OF	A	MK637887
PN2-B05P1-6	May 2015	Pseudomonas sp.	1057	OF	Α	MK637905

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PN2-B05P1-7	May 2015	Pseudomonas sp.	1064	OF	А	MK637906
PN2-B05P1-8	May 2015	Pseudomonas sp.	1069	OF	Α	MK637911
PN2-B05P1-9	May 2015	Pseudomonas sp.	1069	OF	Α	MK637907
PN2-B05P1-	May 2015	Pseudomonas sp.	1075	OF	Α	MK637898
10	•	·				
PN2-B05P1-	May 2015	Pseudomonas sp.	1055	OF	Α	MK637890
11	•	•				
PN2-B05P1-	May 2015	Pseudomonas sp.	1071	OF	Α	MK637885
12	,	•				
PN2-B05P1-	May 2015	Pseudomonas sp.	1073	OF	Α	MK637899
13	•	•				
PN2-B05P1-	May 2015	Pseudomonas sp.	1086	OF	Α	MK637903
14	•	•				
PN2-B05P1-	May 2015	Pseudomonas sp.	1072	OF	Α	MK637895
15	•	•				
PN2-B05P1-	May 2015	Pseudomonas sp.	1066	OF	Α	MK637909
16	,	,				
PN2-B05P1-	May 2015	Streptomyces sp.	1090	OF	Α	MK638556
17	,	, , ,				
PN2-B05P1-	May 2015	Plantibacter sp.	1084	OF	Α	MK638377
18	,	,				
PN2-B05P1-	May 2015	Rhizobium sp.	1066	OF	Α	MK638395
20	,	,				
PN2-B05P3-1	May 2015	Pseudomonas sp.	1085	CF	Α	MK637917
PN2-B05P3-2	, May 2015	Pseudomonas sp.	1086	CF	Α	MK637924
PN2-B05P3-3	, May 2015	Pseudomonas sp.	1076	CF	Α	MK637952
PN2-B05P3-4	May 2015	Pseudomonas sp.	1085	CF	Α	MK637937
PN2-B05P3-5	May 2015	Pseudomonas sp.	1072	CF	Α	MK637945
PN2-B05P3-6	May 2015	Pseudomonas sp.	1081	CF	Α	MK637923
PN2-B05P3-7	May 2015	Pseudomonas sp.	1085	CF	Α	MK637935
PN2-B05P3-8	May 2015	Pseudomonas sp.	1077	CF	Α	MK637936
PN2-B05P3-9	May 2015	Pseudomonas sp.	1085	CF	Α	MK637944
PN2-B05P3-	May 2015	Agrobacterium sp.	1068	CF	A	MK638047
10	Way 2015	rigrobacteriam sp.	1000	Ci	/\	1VII(050047
PN2-B05P3-	May 2015	Streptomyces sp.	1012	CF	Α	MK638463
11	Way 2015	streptomyces sp.	1012	Ci	/\	WIN030403
PN2-B05P3-	May 2015	Streptomyces sp.	1085	CF	Α	MK638458
12	Way 2015	streptomyces sp.	1005	Ci	/\	WIN030430
PN2-B05P3-	May 2015	Phyllobacterium	1071	CF	Α	MK638211
13	Way 2013	sp.	1071	C.	,,	WW.030211
PN2-B05P3-	May 2015	Phyllobacterium	1070	CF	Α	MK638213
14	2013	sp.	10,0	Ç.	, ,	1111030213
PN2-B05P3-	May 2015	Pseudomonas sp.	1076	CF	Α	MK637919
15	2013	. seadonnondo op.	_5,0	٥.	, ,	
PN2-B05P3-	May 2015	Rugamonas sp.	1047	CF	Α	MK638091
16	, 2010			U .		
PN2-B05P3-	May 2015	Achromobacter sp.	1017	CF	Α	MK638022
17	- ,			- -		

PN2-B05P3-	May 2015	Phyllobacterium	1068	CF	Α	MK638399
18		sp.				
PN2-B05P3-	May 2015	Streptomyces sp.	1087	CF	Α	MK638568
19		51 11 1	1001	0=	_	
PN2-B05P3-	May 2015	Phyllobacterium	1021	CF	Α	MK638216
20	NA - 2045	sp.	4072	C.F.		N41/C200E2
PN2-B05P3- 21	May 2015	Agromyces sp.	1072	CF	Α	MK638052
PN2-B06P2-1	May 2015	Deaudomonae en	1076	CF	F	MK637941
	May 2015	Pseudomonas sp.				
PN2-B06P2-2	May 2015	Pseudomonas sp.	1066	CF	F	MK637939
PN2-B06P2-3	May 2015	Pseudomonas sp.	1072	CF	F -	MK637930
PN2-B06P2-4	May 2015	Pseudomonas sp.	1077	CF	F	MK637942
PN2-B06P2-5	May 2015	Pseudomonas sp.	1053	CF	F	MK637940
PN2-B06P2-6	May 2015	Pseudomonas sp.	1085	CF	F	MK637926
PN2-B06P2-7	May 2015	Pseudomonas sp.	1085	CF	F	MK637949
PN2-B06P2-8	May 2015	Bacillus sp.	280	CF	F	MK638081
PN2-B06P2-9	May 2015	Pseudomonas sp.	1085	CF	F	MK637914
PN2-B06P2-	May 2015	Pseudomonas sp.	1082	CF	F	MK637951
10						
PN2-B06P2-	May 2015	Pseudomonas sp.	1073	CF	F	MK637913
11						
PN2-B06P2-	May 2015	Pseudomonas sp.	1081	CF	F	MK637922
12						
PN2-B06P2-	May 2015	Pseudomonas sp.	1077	CF	F	MK637928
14					_	
PN2-B06P2-	May 2015	Pseudomonas sp.	1085	CF	F	MK637931
15		_, , , , ,			_	
PN2-B06P2-	May 2015	Phyllobacterium	1070	CF	F	MK638212
16		sp.			_	
PN2-B06P2-	May 2015	Streptomyces sp.	1086	CF	F	MK638567
17				_		
PN2-B06P2-	May 2015	Pseudomonas sp.	1080	CF	F	MK637920
19		_, , , , ,			_	
PN2-B06P2-	May 2015	Phyllobacterium	1079	CF	F	MK638210
20		sp.			_	
PN2-B06P4-1	May 2015	Pseudomonas sp.	1072	OF	F	MK637925
PN2-B06P4-2	May 2015	Pseudomonas sp.	1084	OF	F	MK637933
PN2-B06P4-3	May 2015	Pseudomonas sp.	1077	OF	F	MK637929
PN2-B06P4-4	May 2015	Pseudomonas sp.	1057	OF	F	MK637934
PN2-B06P4-5	May 2015	Serratia sp.	1039	OF	F	MK638441
PN2-B06P4-6	May 2015	Pseudomonas sp.	1072	OF	F	MK637921
PN2-B06P4-7	May 2015	Streptomyces sp.	1070	OF	F	MK638459
PN2-B06P4-8	May 2015	Streptomyces sp.	1080	OF	F	MK638461
PN2-B06P4-9	May 2015	Bacillus sp.	1094	OF	F	MK638068
PN2-B06P4-	May 2015	Streptomyces sp.	1070	OF	F	MK638595
10						
PN2-B06P4-	May 2015	Pseudomonas sp.	1072	OF	F	MK637918
_11						

PN2-B06P4-	May 2015	Streptomyces sp.	1085	OF	F	MK638457
12						
PN2-B06P4-	May 2015	Pseudomonas sp.	1085	OF	F	MK637915
13						
PN2-B06P4-	May 2015	Phyllobacterium	1066	OF	F	MK638396
14	•	sp.				
PN2-B06P4-	May 2015	Phyllobacterium	1070	OF	F	MK638214
16	,	sp.				
PN2-B06P4-	May 2015	Streptomyces sp.	1085	OF	F	MK638473
17		on epioniyoes sp.	1005	0.	•	
PN2-B06P4-	May 2015	Streptomyces sp.	1064	OF	F	MK638593
18	Way 2015	streptomyees sp.	1001	O.	•	WINGSGSSS
PN2-B06P4-	May 2015	Streptomyces sp.	1079	OF	F	MK638612
19	Widy 2015	streptomyees sp.	1075	O1	•	1411(030012
PN2-B06P4-	May 2015	Streptomyces sp.	1010	OF	F	MK638462
20	Way 2015	Streptomyces sp.	1010	O1	•	WIK030402
PN2-B06P4-	May 2015	Phyllobacterium	1068	OF	F	MK638397
22	IVIAY 2013	•	1008	Oi	1	WIK030337
PN2-B07P1-1	May 2015	sp. Pseudomonas sp.	1067	OF	F	MK637954
PN2-B07P1-2	May 2015	Pseudomonas sp.	1007	OF	, F	MK637953
PN2-B07P1-2 PN2-B07P1-3	May 2015	•	1076	OF	F	MK637956
	•	Pseudomonas sp.				
PN2-B07P1-4	May 2015	Pseudomonas sp.	1029	OF	F	MK637946
PN2-B07P1-5	May 2015	Pseudomonas sp.	1070	OF	F	MK637943
PN2-B07P1-6	May 2015	Pseudomonas sp.	1069	OF	F	MK637947
PN2-B07P1-7	May 2015	Rugamonas sp.	1068	OF	F	MK638092
PN2-B07P1-8	May 2015	Pseudomonas sp.	1073	OF	F	MK637912
PN2-B07P1-9	May 2015	Rhizobium sp.	1025	OF	F	MK638400
PN2-B07P1-	May 2015	Rhodococcus sp.	1078	OF	F	MK638436
10					_	
PN2-B07P1-	May 2015	Streptomyces sp.	1085	OF	F	MK638565
11				_		
PN2-B07P1-	May 2015	Pseudomonas sp.	1072	OF	F	MK637916
12						
PN2-B07P1-	May 2015	Streptomyces sp.	1085	OF	F	MK638566
13						
PN2-B07P1-	May 2015	Pseudomonas sp.	466	OF	F	MK638010
14						
PN2-B07P1-	May 2015	Ochrobactrum sp.	1049	OF	F	MK638199
15						
PN2-B07P1-	May 2015	Pseudomonas sp.	1077	OF	F	MK637927
16						
PN2-B07P1-	May 2015	Rhizobium sp.	1062	OF	F	MK638401
17						
PN2-B07P1-	May 2015	Plantibacter sp.	1076	OF	F	MK638378
18						
PN2-B07P1-	May 2015	Streptomyces sp.	1060	OF	F	MK638460
19						
PN2-B07P1-	May 2015	Variovorax sp.	1031	OF	F	MK638654
20						

PN2-B07P5-1	May 2015	Tsukamurella sp.	1058	CF	F	MK638650
PN2-B07P5-3	May 2015	Streptomyces sp.	1085	CF	F	MK638456
PN2-B07P5-4	May 2015	Pseudomonas sp.	1067	CF	F	MK637938
PN2-B07P5-5	May 2015	Pseudomonas sp.	1085	CF	F	MK637950
PN2-B07P5-6	May 2015	Tsukamurella sp.	1076	CF	F	MK638651
PN2-B07P5-7	May 2015	Pseudomonas sp.	1077	CF	F	MK637932
PN2-B07P5-8	May 2015	Pseudomonas sp.	1078	CF	F	MK637948
PN2-B07P5-9	May 2015	Streptomyces sp.	1085	CF	F	MK638467
PN2-B07P5-	May 2015	Dyadobacter sp.	1071	CF	F	MK638114
10	•	,				
PN2-B07P5-	May 2015	Pseudomonas sp.	1076	CF	F	MK637955
11						
PN2-B07P5-	May 2015	Dyadobacter sp.	1040	CF	F	MK638113
12						
PN2-B07P5-	May 2015	Phyllobacterium	1065	CF	F	MK638215
13		sp.				
PN2-B07P5-	May 2015	Streptomyces sp.	1018	CF	F	MK638624
14						
PN2-B07P5-	May 2015	Phyllobacterium	1070	CF	F	MK638219
15		sp.				
PN2-B07P5-	May 2015	Mesorhizobium sp.	1080	CF	F	MK638165
16						
PN2-B07P5-	May 2015	Streptomyces sp.	673	CF	F	MK638647
17						
PN2-B07P5-	May 2015	Phyllobacterium	1077	CF	F	MK638224
18		sp.				
PN2-B07P5-	May 2015	Phyllobacterium	1066	CF	F	MK638225
20		sp.				
PN2-B07P5-	May 2015	Streptomyces sp.	1076	CF	F	MK638532
21						
PN2-B07P5-	May 2015	Phyllobacterium	1070	CF	F	MK638221
22		sp.				
PN2-B07P5-	May 2015	Phyllobacterium	1070	CF	F	MK638223
23		sp.				
PN2-B07P5-	May 2015	Phyllobacterium	1077	CF	F	MK638217
24		sp.			_	
PN2-B08P4-1	May 2015	Pseudomonas sp.	1074	CF	Α	MK637962
PN2-B08P4-2	May 2015	Pseudomonas sp.	1072	CF	Α	MK637967
PN2-B08P4-3	May 2015	Pseudomonas sp.	1070	CF	Α	MK637976
PN2-B08P4-4	May 2015	Pseudomonas sp.	1076	CF	Α	MK637960
PN2-B08P4-5	May 2015	Pseudomonas sp.	1073	CF	Α	MK637963
PN2-B08P4-6	May 2015	Pseudomonas sp.	1077	CF	Α	MK637971
PN2-B08P4-7	May 2015	Pseudomonas sp.	1077	CF	Α	MK637973
PN2-B08P4-8	May 2015	Pseudomonas sp.	1072	CF	Α	MK637984
PN2-B08P4-9	May 2015	Pseudomonas sp.	766	CF	Α	MK637990
PN2-B08P4-	May 2015	Pseudomonas sp.	1074	CF	Α	MK637977
10						
PN2-B08P4-	May 2015	Pseudomonas sp.	1072	CF	Α	MK637969
11						

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PN2-B08P4-	May 2015	Pseudomonas sp.	1072	CF	А	MK637983
12						
PN2-B08P4-	May 2015	Pseudomonas sp.	1077	CF	Α	MK637975
13						
PN2-B08P4-	May 2015	Pseudomonas sp.	1073	CF	Α	MK637961
14						
PN2-B08P4-	May 2015	Serratia sp.	1081	CF	Α	MK638442
15						
PN2-B08P4-	May 2015	Pseudomonas sp.	1073	CF	Α	MK637982
16						
PN2-B08P4-	May 2015	Pseudomonas sp.	1074	CF	Α	MK637957
17	•	•				
PN2-B08P4-	May 2015	Microbacterium	303	CF	Α	MK638190
18	•	sp.				
PN2-B08P4-	May 2015	Pseudomonas sp.	1085	CF	Α	MK637972
19	•	•				
PN2-B08P4-	May 2015	Pseudomonas sp.	1077	CF	Α	MK637965
20	•	•				
PN2-B08P5-1	May 2015	Pseudomonas sp.	1072	OF	Α	MK637986
PN2-B08P5-2	May 2015	Pseudomonas sp.	1086	OF	Α	MK637970
PN2-B08P5-3	May 2015	Pseudomonas sp.	1079	OF	Α	MK637966
PN2-B08P5-4	May 2015	Variovorax sp.	1067	OF	Α	MK638655
PN2-B08P5-5	May 2015	Pseudomonas sp.	1074	OF	Α	MK637974
PN2-B08P5-6	May 2015	Pseudomonas sp.	1072	OF	Α	MK637985
PN2-B08P5-7	May 2015	Pseudomonas sp.	1072	OF	Α	MK637980
PN2-B08P5-8	May 2015	Pseudomonas sp.	1072	OF	Α	MK637981
PN2-B08P5-9	May 2015	Rhodococcus sp.	721	OF	Α	MK638440
PN2-B08P5-	May 2015	Pseudomonas sp.	1072	OF	Α	MK637979
10	•	•				
PN2-B08P5-	May 2015	Rhizobium sp.	1062	OF	Α	MK638404
11						
PN2-B08P5-	May 2015	Rhodococcus sp.	1030	OF	Α	MK638437
12						
PN2-B08P5-	May 2015	Flavobacterium sp.	1053	OF	Α	MK638130
13						
PN2-B08P5-	May 2015	Phyllobacterium	1070	OF	Α	MK638232
14		sp.				
PN2-B08P5-	May 2015	Streptomyces sp.	1073	OF	Α	MK638466
15						
PN2-B08P5-	May 2015	Ensifer sp.	1070	OF	Α	MK638118
16						
PN2-B08P5-	May 2015	Phyllobacterium	1070	OF	Α	MK638218
17		sp.				
PN2-B08P5-	May 2015	Phyllobacterium	1066	OF	Α	MK638402
18		sp.				
PN2-B08P5-	May 2015	Rhizobium sp.	1066	OF	Α	MK638403
19						
PN2-B08P5-	May 2015	Streptomyces sp.	1056	OF	Α	MK638576
20						

PN2-B08P5-	May 2015	Phyllobacterium	1070	OF	Α	MK638230
21		sp.				
PN2-B09P3-1	May 2015	Streptomyces sp.	1076	CF	F	MK638479
PN2-B09P3-2	May 2015	Pseudomonas sp.	1072	CF	F	MK637964
PN2-B09P3-3	May 2015	Streptomyces sp.	1066	CF	F	MK638522
PN2-B09P3-4	May 2015	Rugamonas sp.	1071	CF	F	MK638093
PN2-B09P3-5	May 2015	Streptomyces sp.	1079	CF	F	MK638465
PN2-B09P3-7	May 2015	Arthrobacter sp.	1093	CF	F	MK638056
PN2-B09P3-8	May 2015	Streptomyces sp.	1070	CF	F	MK638480
PN2-B09P3-9	May 2015	Rhodococcus sp.	1069	CF	F	MK638439
PN2-B09P3-	May 2015	Streptomyces sp.	1056	CF	F	MK638592
10						
PN2-B09P3-	May 2015	Streptomyces sp.	1073	CF	F	MK638598
11						
PN2-B09P3-	May 2015	Flavobacterium sp.	1040	CF	F	MK638131
13						
PN2-B09P3-	May 2015	Phyllobacterium	1070	CF	F	MK638227
14		sp.				
PN2-B09P3-	May 2015	Streptomyces sp.	1070	CF	F	MK638521
15	-					
PN2-B09P3-	May 2015	Phyllobacterium	1068	CF	F	MK638228
16	•	sp.				
PN2-B09P3-	May 2015	Phyllobacterium	1068	CF	F	MK638229
17	•	sp.				
PN2-B09P3-	May 2015	Phyllobacterium	1068	CF	F	MK638226
18	,	sp.				
PN2-B09P3-	May 2015	Microbacterium	1077	CF	F	MK638186
19	,	sp.				
PN2-B09P3-	May 2015	Phyllobacterium	1065	CF	F	MK638231
21	•	sp.				
PN2-B09P3-	May 2015	Streptomyces sp.	1076	CF	F	MK638481
22	,	, , ,				
PN2-B09P3-	May 2015	Phyllobacterium	1070	CF	F	MK638222
24	,	sp.				
PN2-B09P5-1	May 2015	Pseudomonas sp.	1073	OF	F	MK637978
PN2-B09P5-2	May 2015	Plantibacter sp.	1071	OF	F	MK638380
PN2-B09P5-3	May 2015	Pseudomonas sp.	1074	OF	F	MK637968
PN2-B09P5-4	May 2015	Pseudomonas sp.	1074	OF	F	MK637958
PN2-B09P5-5	May 2015	Agrobacterium sp.	1073	OF	F	MK638044
PN2-B09P5-6	May 2015	Streptomyces sp.	1087	OF	F	MK638464
PN2-B09P5-7	May 2015	Bacillus sp.	1073	OF	F	MK638071
PN2-B09P5-8	May 2015	Pseudomonas sp.	1049	OF	, F	MK637989
PN2-B09P5-9	May 2015	Flavobacterium sp.	907	OF	, F	MK637383
PN2-B09P5-	May 2015	Pseudomonas sp.	1045	OF	r F	MK637988
10	iviay ZUIS	i seudomonus sp.	1043	Oi	ı	IVINUS/300
PN2-B09P5-	May 2015	Bacillus sp.	1072	OF	F	MK638070
11	iviay ZUIJ	υατιίας ομ.	10/2	O1	ı	IVINUJOU/U
PN2-B09P5-	May 2015	Agrobacterium sp.	1029	OF	F	MK638043
12	iviay ZUIS	Agrobacterium sp.	1023	OI.	ı	1411/030043
14						

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PN2-B09P5-	May 2015	Pseudomonas sp.	806	OF	F	MK637987
13						
PN2-B09P5-	May 2015	Streptomyces sp.	1085	OF	F	MK638519
14						
PN2-B09P5-	May 2015	Phyllobacterium	1066	OF	F	MK638220
15		sp.				
PN2-B09P5-	May 2015	Streptomyces sp.	1066	OF	F	MK638477
16						
PN2-B09P5-	May 2015	Plantibacter sp.	1057	OF	F	MK638379
17						
PN2-B09P5-	May 2015	Bacillus sp.	1076	OF	F	MK638069
18						
PN2-B09P5-	May 2015	Plantibacter sp.	1015	OF	F	MK638381
21		-, , , ,			_	
PN2-B09P5-	May 2015	Phyllobacterium	1026	OF	F	MK638359
22		sp.	1010	0.5		
PN2-B10P3-1	May 2015	Pseudomonas sp.	1043	OF	Α	MK638012
PN2-B10P3-2	May 2015	Stenotrophomonas sp.	1032	OF	Α	MK638450
PN2-B10P3-3	May 2015	Stenotrophomonas	1034	OF	Α	MK638451
	•	sp.				
PN2-B10P3-5	May 2015	Pseudomonas sp.	1032	OF	Α	MK638013
PN2-B10P3-6	May 2015	Pseudomonas sp.	1032	OF	Α	MK638014
PN2-B10P3-7	May 2015	Pseudomonas sp.	1032	OF	Α	MK638015
PN2-B10P3-8	May 2015	Pseudomonas sp.	1032	OF	Α	MK638016
PN2-B10P3-9	May 2015	Pseudomonas sp.	1032	OF	Α	MK638017
PN2-B10P3-	May 2015	Variovorax sp.	464	OF	Α	MK638661
10						
PN2-B10P3-	May 2015	Variovorax sp.	470	OF	Α	MK638662
11						
PN2-B10P3-	May 2015	Variovorax sp.	751	OF	Α	MK638663
12						
PN2-B10P3-	May 2015	Stenotrophomonas	1034	OF	Α	MK638448
13		sp.		_		
PN2-B10P3-	May 2015	Variovorax sp.	506	OF	Α	MK638664
14	2045		4007	0.5		NAVC20402
PN2-B10P3-	May 2015	Mesorhizobium sp.	1037	OF	Α	MK638183
15 DN2 B10B2	May 2015	Ctrontomycocco	1025	OΓ	۸	N4V629612
PN2-B10P3- 17	May 2015	Streptomyces sp.	1025	OF	Α	MK638613
17 PN2-B10P3-	May 2015	Streptomyces sp.	957	OF	Α	MK638614
18	Way 2013	streptorryces sp.	337	Oi	A	WIK038014
PN2-B10P3-	May 2015	Rhizobium sp.	1004	OF	Α	MK638432
20	1414 ZUIJ	mizobiani sp.	1004	01	Γ	14111030432
PN2-B10P3-	May 2015	Stenotrophomonas	1024	OF	Α	MK638449
21	111dy 2013	sp.	1027	<u>J</u> .	, ,	14111050-775
PN2-B10P5-1	May 2015	Pseudomonas sp.	1032	CF	Α	MK638018
PN2-B10P5-2	May 2015	Flavobacterium sp.	1027	CF	Α	MK638133
PN2-B10P5-3	May 2015	Rhizobium sp.	1017	CF	A	MK638433
	, _5_5	сэлангор.		<u> </u>	- ' '	

PN2-B10P5-4	May 2015	Pseudomonas sp.	1032	CF	А	MK638021
PN2-B10P5-5	May 2015	Streptomyces sp.	1071	CF	Α	MK638515
PN2-B10P5-6	May 2015	Clavibacter sp.	1003	CF	Α	MK638110
PN2-B10P5-8	May 2015	Agrobacterium sp.	1008	CF	Α	MK638048
PN2-B10P5-	May 2015	Pseudomonas sp.	1032	CF	Α	MK638019
10	·	·				
PN2-B10P5-	May 2015	Pseudomonas sp.	1032	CF	Α	MK638020
11	,	,				
PN2-B10P5-	May 2015	Bacillus sp.	1027	CF	Α	MK638062
12	·	·				
PN2-B10P5-	May 2015	Streptomyces sp.	1066	CF	Α	MK638455
13						
PN2-B10P5-	May 2015	Pseudomonas sp.	1077	CF	Α	MK637959
14						
PN2-B10P5-	May 2015	Streptomyces sp.	1001	CF	Α	MK638594
15						
PN2-B10P5-	May 2015	Agrobacterium sp.	1068	CF	Α	MK638042
16						
PN2-B10P5-	May 2015	Agrobacterium sp.	1070	CF	Α	MK638041
17						
PN2-B10P5-	May 2015	Streptomyces sp.	1073	CF	Α	MK638478
18						
PN2-B10P5-	May 2015	Streptomyces sp.	1069	CF	Α	MK638560
19						
PN2-B10P5-	May 2015	Rhizobium sp.	1075	CF	Α	MK638398
20						
PN2-B10P5-	May 2015	Phyllobacterium	1070	CF	Α	MK638209
21		sp.				
PN2-B10P5-	May 2015	Leifsonia sp.	1082	CF	Α	MK638143
22						
PN3-B01P2-1	July 2015	Phyllobacterium	998	CF	Α	MK638236
		sp.				
PN3-B01P2-2	July 2015	Streptomyces sp.	1018	CF	Α	MK638483
PN3-B01P2-3	July 2015	Rhizobium sp.	1023	CF	Α	MK638410
PN3-B01P2-4	July 2015	Mesorhizobium sp.	1025	CF	Α	MK638169
PN3-B01P2-5	July 2015	Streptomyces sp.	1047	CF	Α	MK638559
PN3-B01P2-6	July 2015	Rhizobium sp.	1026	CF	Α	MK638412
PN3-B01P2-7	July 2015	Streptomyces sp.	1008	CF	Α	MK638514
PN3-B01P2-9	July 2015	Rhizobium sp.	1009	CF	Α	MK638408
PN3-B01P2-	July 2015	Mesorhizobium sp.	1034	CF	Α	MK638167
10						
PN3-B01P2-	July 2015	Rhizobium sp.	1029	CF	Α	MK638411
12						
PN3-B01P2-	July 2015	Mesorhizobium sp.	1026	CF	Α	MK638168
14						
PN3-B01P2-	July 2015	Leifsonia sp.	1043	CF	Α	MK638144
15						
PN3-B01P2-	July 2015	Rhizobium sp.	1020	CF	Α	MK638407
17						

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PN3-B01P2-	July 2015	Inquilinus sp.	952	CF	Α	MK638135
18						
PN3-B01P2-	July 2015	Streptomyces sp.	1018	CF	Α	MK638512
19						
PN3-B01P2-	July 2015	Inquilinus sp.	999	CF	Α	MK638136
20	•	, ,				
PN3-B01P2-	July 2015	Phyllobacterium	997	CF	Α	MK638241
21	ou, _o_o	sp.		О.		
PN3-B01P2-	July 2015	Phyllobacterium	1025	CF	Α	MK638239
22	July 2013	sp.	1025	C.	,,	WIN030233
PN3-B01P2-	July 2015	Leifsonia sp.	1032	CF	Α	MK638145
23	July 2013	Leijsoilla sp.	1032	Ci	^	1411/030143
PN3-B01P4-1	July 2015	Streptomyces sp.	1026	OF	Α	MK638563
PN3-B01P4-3	July 2015 July 2015	Streptomyces sp.	1010	OF	A	MK638558
	•					
PN3-B01P4-5	July 2015	Streptomyces sp.	1035	OF	A	MK638562
PN3-B01P4-6	July 2015	Burkholderia sp.	926	OF	Α	MK638088
PN3-B01P4-7	July 2015	Streptomyces sp.	1015	OF	Α	MK638540
PN3-B01P4-8	July 2015	Phyllobacterium	1026	OF	Α	MK638242
		sp.				
PN3-B01P4-9	July 2015	Agrobacterium sp.	943	OF	Α	MK638023
PN3-B01P4-	July 2015	Ensifer sp.	1016	OF	Α	MK638119
10						
PN3-B01P4-	July 2015	Streptomyces sp.	1041	OF	Α	MK638518
11						
PN3-B01P4-	July 2015	Streptomyces sp.	1064	OF	Α	MK638509
12						
PN3-B01P4-	July 2015	Agrobacterium sp.	1042	OF	Α	MK638039
13						
PN3-B01P4-	July 2015	Streptomyces sp.	1038	OF	Α	MK638534
14						
PN3-B01P4-	July 2015	Phyllobacterium	1026	OF	Α	MK638237
15		sp.				
PN3-B01P4-	July 2015	Streptomyces sp.	1028	OF	Α	MK638516
16	-					
PN3-B01P4-	July 2015	Agrobacterium sp.	948	OF	Α	MK638024
17	•					
PN3-B01P4-	July 2015	Rhizobium sp.	1028	OF	Α	MK638409
18	•	•				
PN3-B01P4-	July 2015	Rhizobium sp.	818	OF	Α	MK638413
19	•	•				
PN3-B01P4-	July 2015	Rhizobium sp.	895	OF	Α	MK638414
20	, -	r-	-			
PN3-B01P4-	July 2015	Ensifer sp.	1029	OF	Α	MK638120
21	,	, -r		-		
PN3-B01P4-	July 2015	Ensifer sp.	1023	OF	Α	MK638122
22	, – . 	- <i>J</i> - · - - F -				
PN3-B01P4-	July 2015	Rhizobium sp.	1039	OF	Α	MK638405
23	, – . 	- 2 3/2.				

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PN3-B01P4-	July 2015	Streptomyces sp.	1025	OF	Α	MK638517
24						
PN3-B02P3-1	July 2015	Phyllobacterium sp.	1026	OF	F	MK638246
PN3-B02P3-2	July 2015	Variovorax sp.	266	OF	F	MK638660
PN3-B02P3-3	July 2015	Pseudomonas sp.	951	OF	F	MK637991
PN3-B02P3-4	, July 2015	Phyllobacterium	950	OF	F	MK638238
	·	sp.				
PN3-B02P3-5	July 2015	Streptomyces sp.	1013	OF	F	MK638600
PN3-B02P3-6	July 2015	Bacillus sp.	841	OF	F	MK638072
PN3-B02P3-7	July 2015	Bacillus sp.	982	OF	F	MK638074
PN3-B02P3-8	July 2015	Bacillus sp.	1024	OF	F	MK638075
PN3-B02P3-9	July 2015	Bacillus sp.	1009	OF	F	MK638079
PN3-B02P3-	July 2015	Bacillus sp.	1011	OF	F	MK638077
10						
PN3-B02P3-	July 2015	Bacillus sp.	967	OF	F	MK638073
11						
PN3-B02P3-	July 2015	Bacillus sp.	1024	OF	F	MK638076
12						
PN3-B02P3-	July 2015	Streptomyces sp.	1010	OF	F	MK638526
13						
PN3-B02P3-	July 2015	Streptomyces sp.	1031	OF	F	MK638510
14						
PN3-B02P3-	July 2015	Bacillus sp.	977	OF	F	MK638078
15						
PN3-B02P3-	July 2015	Streptomyces sp.	1018	OF	F	MK638523
17						
PN3-B02P3-	July 2015	Phyllobacterium	999	OF	F	MK638243
18		sp.				
PN3-B02P3-	July 2015	Phyllobacterium	1026	OF	F	MK638244
19		sp.			_	
PN3-B02P3-	July 2015	Phyllobacterium	1024	OF	F	MK638245
20		sp.	1010	0.5	_	
PN3-B02P3-	July 2015	Streptomyces sp.	1018	OF	F	MK638535
21		Description of the second	4027	05	-	N41/C20002
PN3-B02P3-	July 2015	Bradyrhizobium	1027	OF	F	MK638083
22	I. d. 2045	Sp.	1024	C.F.	-	N41/C2022E
PN3-B02P4-1	July 2015	Phyllobacterium sp.	1034	CF	F	MK638235
PN3-B02P4-3	July 2015	Mesorhizobium sp.	1026	CF	F	MK638166
PN3-B02P4-4	July 2015	Phyllobacterium	907	CF	F	MK638247
	=	sp.				
PN3-B02P4-5	July 2015	Herbiconiux sp.	293	CF	F	MK638668
PN3-B02P4-6	July 2015	Phyllobacterium	663	CF	F	MK638248
		sp.				
PN3-B02P4-7	July 2015	Phyllobacterium	1030	CF	F	MK638240
		sp.				
PN3-B02P4-	July 2015	Streptomyces sp.	1057	CF	F	MK638561
15						
				_		

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PN3-B02P4-	July 2015	Streptomyces sp.	1012	CF	F	MK638601
17						
PN3-B02P4-	July 2015	Streptomyces sp.	914	CF	F	MK638603
18						
PN3-B02P4-	July 2015	Phyllobacterium	1024	CF	F	MK638234
19		sp.				
PN3-B02P4-	July 2015	Pseudaminobacter	1012	CF	F	MK638382
20		sp.				
PN3-B02P4-	July 2015	Streptomyces sp.	1018	CF	F	MK638513
21					_	
PN3-B02P4-	July 2015	Rhizobium sp.	1023	CF	F	MK638406
22		A	1020	C.F.		N4V620020
PN3-B03P1-1	July 2015	Agrobacterium sp.	1029	CF	A	MK638038
PN3-B03P1-2	July 2015	Variovorax sp.	624	CF	A	MK638659
PN3-B03P1-3	July 2015	Mesorhizobium sp.	1001	CF	A	MK638171
PN3-B03P1-4	July 2015	Streptomyces sp.	814	CF	A	MK638602
PN3-B03P1-5	July 2015	Mesorhizobium sp.	1052	CF	A	MK638170
PN3-B03P1-6	July 2015 July 2015	Ensifer sp.	1021	CF	A	MK638121
PN3-B03P1-7 PN3-B03P1-	July 2015 July 2015	Streptomyces sp.	1022 908	CF CF	A A	MK638564 MK638604
10	July 2015	Streptomyces sp.	908	CF	А	IVIN036004
PN3-B03P1-	July 2015	Inquilinus sp.	1019	CF	А	MK638137
12	July 2013	mquiimus sp.	1019	Ci	^	WINOSOIS7
PN3-B03P1-	July 2015	Phyllobacterium	1011	CF	Α	MK638233
13	July 2010	sp.	1011	C.	, ,	WW.050255
PN3-B03P1-	July 2015	Leifsonia sp.	1019	CF	Α	MK638146
14	,	, ,				
PN3-B03P1-	July 2015	Streptomyces sp.	1010	CF	Α	MK638599
16						
PN3-B03P1-	July 2015	Streptomyces sp.	1025	CF	Α	MK638511
17						
PN3-B03P1-	July 2015	Dyella sp.	962	CF	Α	MK638117
18		_		_		
PN3-B03P1-	July 2015	Streptomyces sp.	1010	CF	Α	MK638533
19	I. d. 2015	Mania	004	C.E.	Δ.	NAVCORCEC
PN3-B03P1- 20	July 2015	Variovorax sp.	991	CF	Α	MK638656
PN3-B03P1-	July 2015	Streptomyces sp.	1060	CF	А	MK638550
21	July 2013	Streptomyces sp.	1000	Ci	^	WINOSOSSO
PN3-B03P1-	July 2015	Streptomyces sp.	1034	CF	Α	MK638520
22	ou.,	от орготу осо орг		C.		
PN3-B03P1-	July 2015	Streptomyces sp.	1024	CF	Α	MK638525
23	•	. , .				
PN3-B03P2-1	July 2015	Dyella sp.	1051	OF	Α	MK638665
PN3-B03P2-2	July 2015	Dyella sp.	1039	OF	Α	MK638666
PN3-B03P2-3	July 2015	Streptomyces sp.	1070	OF	Α	MK638581
PN3-B03P2-4	July 2015	Streptomyces sp.	1019	OF	Α	MK638548
PN3-B03P2-5	July 2015	Streptomyces sp.	1059	OF	Α	MK638547
PN3-B03P2-6	July 2015	Rhizobium sp.	1050	OF	Α	MK638420

PN3-B03P2-7	July 2015	Streptomyces sp.	1070	OF	Α	MK638546
PN3-B03P2-8	July 2015	Rhizobium sp.	1062	OF	Α	MK638415
PN3-B03P2-9	July 2015	Streptomyces sp.	1066	OF	Α	MK638544
PN3-B03P2-	, July 2015	Rhizobium sp.	1058	OF	Α	MK638418
10	,					
PN3-B03P2-	July 2015	Leifsonia sp.	1058	OF	Α	MK638147
12	July 2013	Leijsoma sp.	1030	O1	, ,	WIN030147
PN3-B03P2-	July 2015	Streptomyces sp.	1041	OF	Α	MK638579
13	July 2013	streptomyces sp.	1041	O1	А	14111030373
PN3-B03P2-	July 2015	Mesorhizobium sp.	1061	OF	Α	MK638172
14	July 2013	wicsornizobiani sp.	1001	Oi	Α	WINOSOT72
PN3-B03P2-	July 2015	Mesorhizobium sp.	1073	OF	Α	MK638173
15	July 2015	iviesornizobium sp.	10/5	OF	A	101020172
PN3-B03P2-	July 2015	Ctrantamucas sn	1002	OF	Α	MK638586
16	July 2015	Streptomyces sp.	1002	UF	А	IVINOSOSOO
	July 2015	Ctrontonovoos en	1022	OΓ	۸	NAME 20E 42
PN3-B03P2-	July 2015	Streptomyces sp.	1022	OF	Α	MK638543
17		C11	1010	0.5		N4VC20C2E
PN3-B03P2-	July 2015	Streptomyces sp.	1018	OF	Α	MK638625
18		- '	1000	0-		
PN3-B03P2-	July 2015	Pseudomonas sp.	1069	OF	Α	MK637995
19					_	
PN3-B03P2-	July 2015	Luteibacter sp.	1039	OF	Α	MK638152
20						
PN3-B03P2-	July 2015	Streptomyces sp.	1070	OF	Α	MK638552
21						
PN3-B03P2-	July 2015	Streptomyces sp.	1024	OF	Α	MK638609
22						
PN3-B04P1-1	July 2015	Pseudomonas sp.	1067	CF	F	MK637996
PN3-B04P1-2	July 2015	Streptomyces sp.	1055	CF	F	MK638577
PN3-B04P1-3	July 2015	Rhizobium sp.	1069	CF	F	MK638421
PN3-B04P1-4	July 2015	Phyllobacterium	1052	CF	F	MK638259
		sp.				
PN3-B04P1-6	July 2015	Rhizobium sp.	1024	CF	F	MK638422
PN3-B04P1-7	July 2015	Phyllobacterium	1061	CF	F	MK638249
		sp.				
PN3-B04P1-8	July 2015	Microbacterium	1048	CF	F	MK638188
		sp.				
PN3-B04P1-9	July 2015	Stenotrophomonas	1041	CF	F	MK638446
	•	sp.				
PN3-B04P1-	July 2015	Mesorhizobium sp.	1042	CF	F	MK638175
10	•	•				
PN3-B04P1-	July 2015	Streptomyces sp.	1055	CF	F	MK638578
11	,	, , ,				
PN3-B04P1-	July 2015	Phyllobacterium	1049	CF	F	MK638250
12	,	sp.		-		
PN3-B04P1-	July 2015	Phyllobacterium	1068	CF	F	MK638256
14	, _0_0	sp.			•	
PN3-B04P1-	July 2015	Streptomyces sp.	1073	CF	F	MK638551
15	, 2010	cp to , cco op.	20,0	.	•	

PN3-B04P1-	July 2015	Microbacterium	1047	CF	F	MK638187
17		sp.				
PN3-B04P1-	July 2015	Tardiphaga sp.	1035	CF	F	MK638649
18						
PN3-B04P1-	July 2015	Caulobacter sp.	541	CF	F	MK638105
20						
PN3-B04P1-	July 2015	Streptomyces sp.	1024	CF	F	MK638589
21		_			_	
PN3-B04P1-	July 2015	Streptomyces sp.	1007	CF	F	MK638554
22			4000	0.5	-	N4VC20454
PN3-B04P2-1	July 2015	Luteibacter sp.	1029	OF	F	MK638151
PN3-B04P2-2	July 2015	Pseudomonas sp.	1055	OF	F	MK637993
PN3-B04P2-3	July 2015	Streptomyces sp.	1061	OF	F	MK638580
PN3-B04P2-4	July 2015	Pseudomonas sp.	1055	OF	F	MK637997
PN3-B04P2-5	July 2015	Streptomyces sp.	1061	OF	F	MK638582
PN3-B04P2-6	July 2015	Rhizobium sp.	1026	OF	F	MK638419
PN3-B04P2-7	July 2015	Phyllobacterium sp.	1038	OF	F	MK638260
PN3-B04P2-8	July 2015	Ensifer sp.	1029	OF	F	MK638123
PN3-B04P2-9	July 2015	Mycobacterium sp.	1067	OF	F	MK638195
PN3-B04P2-	July 2015	Luteibacter sp.	1051	OF	F	MK638150
10						
PN3-B04P2-	July 2015	Agrobacterium sp.	1065	OF	F	MK637994
11						
PN3-B04P2-	July 2015	Mycobacterium sp.	1047	OF	F	MK638194
12						
PN3-B04P2-	July 2015	Phyllobacterium	541	OF	F	MK638367
13		sp.				
PN3-B04P2-	July 2015	Streptomyces sp.	1018	OF	F	MK638626
14						
PN3-B04P2-	July 2015	Streptomyces sp.	955	OF	F	MK638607
15			1001	0.5	_	
PN3-B04P2-	July 2015	Rhizobium sp.	1064	OF	F	MK638416
16		0.0	4057	0.5	-	N4V620474
PN3-B04P2-	July 2015	Mesorhizobium sp.	1057	OF	F	MK638174
17 PN3-B04P2-	July 2015	Mesorhizobium sp.	1068	OF	F	MK638177
19	July 2015	wiesornizobium sp.	1006	OF	Г	IVINUSOT//
PN3-B04P2-	July 2015	Streptomyces sp.	1055	OF	F	MK638605
20	July 2013	streptomyces sp.	1033	O1	•	WINOSOOOS
PN3-B04P2-	July 2015	Rhizobium sp.	1062	OF	F	MK638417
21	34.y 2013	mizobiam sp.	1002	O.	•	WIN030117
PN3-B04P2-	July 2015	Streptomyces sp.	1067	OF	F	MK638584
22	, -			- -	-	
PN3-B04P2-	July 2015	Phyllobacterium	541	OF	F	MK638366
23	•	sp.				
PN3-B05P1-2	July 2015	Phyllobacterium	1051	OF	Α	MK638255
		sp.				
PN3-B05P1-3	July 2015	Inquilinus sp.	1053	OF	Α	MK638138

PN3-B05P1-4	July 2015	Microbacterium sp.	507	OF	А	MK638191
PN3-B05P1-5	July 2015	Phyllobacterium sp.	1073	OF	Α	MK638251
PN3-B05P1-6	July 2015	Streptomyces sp.	1025	OF	Α	MK638588
PN3-B05P1-7	, July 2015	Paraburkholderia	1036	OF	Α	MK638089
	•	sp.				
PN3-B05P1-8	July 2015	Inquilinus sp.	1047	OF	Α	MK638139
PN3-B05P1-9	July 2015	Streptomyces sp.	1073	OF	Α	MK638583
PN3-B05P1-	July 2015	Streptomyces sp.	891	OF	Α	MK638608
10	-					
PN3-B05P1-	July 2015	Phyllobacterium	1029	OF	Α	MK638258
11		sp.				
PN3-B05P1-	July 2015	Phyllobacterium	1070	OF	Α	MK638257
12		sp.				
PN3-B05P1-	July 2015	Pseudomonas sp.	1030	OF	Α	MK637998
13						
PN3-B05P1-	July 2015	Streptomyces sp.	1022	OF	Α	MK638585
14						
PN3-B05P1-	July 2015	Mesorhizobium sp.	1074	OF	Α	MK638176
15						
PN3-B05P1-	July 2015	Streptomyces sp.	864	OF	Α	MK638606
16		51 11 1	1010	0.5		
PN3-B05P1-	July 2015	Phyllobacterium	1043	OF	Α	MK638254
17	1 2015	sp.	1055	0.5	Δ.	N4VC20E07
PN3-B05P1- 18	July 2015	Streptomyces sp.	1055	OF	Α	MK638587
PN3-B05P1-	July 2015	Rhizobium sp.	1026	OF	Α	MK638423
19	July 2013	nnizobiani sp.	1020	Oi	^	WIN030423
PN3-B05P1-	July 2015	Agrobacterium sp.	1062	OF	Α	MK638025
21	3diy 2013	rigiobacterium sp.	1002	O.	, ,	WIN030023
PN3-B05P3-1	July 2015	Phyllobacterium sp.	1052	CF	Α	MK638253
PN3-B05P3-2	July 2015	Streptomyces sp.	1060	CF	Α	MK638524
PN3-B05P3-3	July 2015	Agromyces sp.	1023	CF	Α	MK638051
PN3-B05P3-4	July 2015	Streptomyces sp.	1054	CF	Α	MK638527
PN3-B05P3-5	July 2015	Pseudomonas sp.	1067	CF	Α	MK637992
PN3-B05P3-6	July 2015	Phyllobacterium	1065	CF	Α	MK638252
	,	sp.				
PN3-B05P3-7	July 2015	Phyllobacterium sp.	1031	CF	Α	MK638261
PN3-B05P3-8	July 2015	Phyllobacterium	1055	CF	А	MK638262
1113 2031 3 3	July 2013	sp.	1033	Ci	,,	WW.030202
PN3-B05P3-9	July 2015	Phyllobacterium	1051	CF	Α	MK638272
		sp.				
PN3-B05P3-	July 2015	Streptomyces sp.	931	CF	Α	MK638484
10						
PN3-B05P3-	July 2015	Inquilinus sp.	1054	CF	Α	MK638140
11						

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PN3-B05P3-	July 2015	Phyllobacterium	1053	CF	Α	MK638285
PN3-B05P3-	July 2015	sp. Pseudomonas sp.	1074	CF	Α	MK638000
13 PN3-B05P3-	July 2015	Phyllobacterium	1052	CF	Α	MK638264
14 PN3-B05P3-	July 2015	sp. Phyllobacterium	1053	CF	Α	MK638288
15 PN3-B05P3-	July 2015	sp. Streptomyces sp.	1060	CF	А	MK638490
17 PN3-B05P3-	July 2015	Phyllobacterium	1062	CF	А	MK638269
18	·	sp.				
PN3-B05P3- 19	July 2015	Phyllobacterium sp.	1034	CF	Α	MK638277
PN3-B05P3- 20	July 2015	Phyllobacterium sp.	1045	CF	Α	MK638270
PN3-B05P3- 21	July 2015	Streptomyces sp.	1064	CF	Α	MK638491
PN3-B05P3- 23	July 2015	Streptomyces sp.	1060	CF	Α	MK638508
PN3-B05P3- 24	July 2015	Streptomyces sp.	1060	CF	Α	MK638610
PN3-B06P2-1	July 2015	Phyllobacterium	588	CF	F	MK638336
PN3-B06P2-2	July 2015	sp. Phyllobacterium	1029	CF	F	MK638296
PN3-B06P2-3	July 2015	sp. Phyllobacterium	1004	CF	F	MK638309
PN3-B06P2-4	July 2015	sp. Phyllobacterium	1034	CF	F	MK638298
PN3-B06P2-5	July 2015	sp. Phyllobacterium	1065	CF	F	MK638276
PN3-B06P2-6	July 2015	sp. Phyllobacterium sp.	1035	CF	F	MK638291
PN3-B06P2-7	July 2015	Devosia sp.	1030	CF	F	MK638111
PN3-B06P2-8	July 2015	Phyllobacterium sp.	1062	CF	F	MK638275
PN3-B06P2-9	July 2015	Phyllobacterium sp.	1062	CF	F	MK638273
PN3-B06P2- 10	July 2015	Streptomyces sp.	1060	CF	F	MK638505
PN3-B06P2- 11	July 2015	Phyllobacterium sp.	1053	CF	F	MK638265
PN3-B06P2- 12	July 2015	Phyllobacterium	837	CF	F	MK638297
PN3-B06P2-	July 2015	sp. Bacillus sp.	527	CF	F	MK638080
13 PN3-B06P2-	July 2015	Phyllobacterium	1065	CF	F	MK638284
14		sp.				

PN3-B06P2- July 2015 Phyllobacterium 1053 CF F MK638 15 sp. PN3-B06P2- July 2015 Streptomyces sp. 1068 CF F MK638 16 PN3-B06P2- July 2015 Streptomyces sp. 605 CF F MK638 17 PN3-B06P2- July 2015 Rhizobium sp. 939 CF F MK638 18 PN3-B06P2- July 2015 Phyllobacterium 1030 CF F MK638 19 sp. PN3-B06P2- July 2015 Streptomyces sp. 1064 CF F MK638 20 PN3-B06P2- July 2015 Phyllobacterium 1073 CF F MK638 21 sp. PN3-B06P2- July 2015 Chryseobacterium 1045 CF F MK638 22 sp. PN3-B06P2- July 2015 Streptomyces sp. 1062 CF F MK638 <th>3507 3627 3424 3274 3489 3271 3109</th>	3507 3627 3424 3274 3489 3271 3109
PN3-B06P2- July 2015 Streptomyces sp. 1068 CF F MK638 16 PN3-B06P2- July 2015 Streptomyces sp. 605 CF F MK638 17 PN3-B06P2- July 2015 Rhizobium sp. 939 CF F MK638 18 PN3-B06P2- July 2015 Phyllobacterium 1030 CF F MK638 19 sp. PN3-B06P2- July 2015 Streptomyces sp. 1064 CF F MK638 20 PN3-B06P2- July 2015 Phyllobacterium 1073 CF F MK638 21 sp. PN3-B06P2- July 2015 Chryseobacterium 1045 CF F MK638 22 sp.	3627 3424 3274 3489 3271 3109
16 PN3-B06P2- July 2015 Streptomyces sp. 605 CF F MK638 17 PN3-B06P2- July 2015 Rhizobium sp. 939 CF F MK638 18 PN3-B06P2- July 2015 Phyllobacterium 1030 CF F MK638 19 sp. PN3-B06P2- July 2015 Streptomyces sp. 1064 CF F MK638 20 PN3-B06P2- July 2015 Phyllobacterium 1073 CF F MK638 21 sp. PN3-B06P2- July 2015 Chryseobacterium 1045 CF F MK638 22 sp.	3627 3424 3274 3489 3271 3109
PN3-B06P2- July 2015 Streptomyces sp. 605 CF F MK638 17 PN3-B06P2- July 2015 Rhizobium sp. 939 CF F MK638 18 PN3-B06P2- July 2015 Phyllobacterium 1030 CF F MK638 19 sp. PN3-B06P2- July 2015 Streptomyces sp. 1064 CF F MK638 20 PN3-B06P2- July 2015 Phyllobacterium 1073 CF F MK638 21 sp. PN3-B06P2- July 2015 Chryseobacterium 1045 CF F MK638 22 sp.	3424 3274 3489 3271 3109
17 PN3-B06P2- July 2015 Rhizobium sp. 939 CF F MK638 18 PN3-B06P2- July 2015 Phyllobacterium 1030 CF F MK638 19 sp. PN3-B06P2- July 2015 Streptomyces sp. 1064 CF F MK638 20 PN3-B06P2- July 2015 Phyllobacterium 1073 CF F MK638 21 sp. PN3-B06P2- July 2015 Chryseobacterium 1045 CF F MK638 22 sp.	3424 3274 3489 3271 3109
PN3-B06P2- July 2015 Rhizobium sp. 939 CF F MK638 18 PN3-B06P2- July 2015 Phyllobacterium 1030 CF F MK638 19 sp. PN3-B06P2- July 2015 Streptomyces sp. 1064 CF F MK638 20 PN3-B06P2- July 2015 Phyllobacterium 1073 CF F MK638 21 sp. PN3-B06P2- July 2015 Chryseobacterium 1045 CF F MK638 22 sp.	3274 3489 3271 3109
18 PN3-B06P2- July 2015 Phyllobacterium 1030 CF F MK638 19 sp. PN3-B06P2- July 2015 Streptomyces sp. 1064 CF F MK638 20 PN3-B06P2- July 2015 Phyllobacterium 1073 CF F MK638 21 sp. PN3-B06P2- July 2015 Chryseobacterium 1045 CF F MK638 22 sp.	3274 3489 3271 3109
sp. PN3-B06P2- July 2015 Streptomyces sp. 1064 CF F MK638 20 PN3-B06P2- July 2015 Phyllobacterium 1073 CF F MK638 21 sp. PN3-B06P2- July 2015 Chryseobacterium 1045 CF F MK638 22 sp.	3489 3271 3109
PN3-B06P2- July 2015 Streptomyces sp. 1064 CF F MK638 20 PN3-B06P2- July 2015 Phyllobacterium 1073 CF F MK638 21 sp. PN3-B06P2- July 2015 Chryseobacterium 1045 CF F MK638 22 sp.	3271 3109
20 PN3-B06P2- July 2015 Phyllobacterium 1073 CF F MK638 21 sp. PN3-B06P2- July 2015 Chryseobacterium 1045 CF F MK638 22 sp.	3271 3109
PN3-B06P2- July 2015 Phyllobacterium 1073 CF F MK638 21 sp. PN3-B06P2- July 2015 Chryseobacterium 1045 CF F MK638 22 sp.	3109
21 sp. PN3-B06P2- July 2015 Chryseobacterium 1045 CF F MK638 22 sp.	3109
PN3-B06P2- July 2015 <i>Chryseobacterium</i> 1045 CF F MK638 <i>sp.</i>	
22	
•	3492
= = =	
23	
PN3-B06P4-1 July 2015 <i>Pseudomonas sp.</i> 1030 OF F MK638	3002
PN3-B06P4-2 July 2015 Pseudomonas sp. 1068 OF F MK637	
PN3-B06P4-3 July 2015 <i>Phyllobacterium</i> 1055 OF F MK638	3279
sp.	
PN3-B06P4-4 July 2015 <i>Variovorax sp.</i> 1031 OF F MK638	
PN3-B06P4-5 July 2015 <i>Streptomyces sp.</i> 1060 OF F MK638	
PN3-B06P4-6 July 2015 <i>Phyllobacterium</i> 1070 OF F MK638 <i>sp.</i>	3263
PN3-B06P4-7 July 2015 <i>Phyllobacterium</i> 1050 OF F MK638 <i>sp.</i>	3287
PN3-B06P4-8 July 2015 <i>Phyllobacterium</i> 1065 OF F MK638 <i>sp.</i>	3282
PN3-B06P4-9 July 2015 <i>Phyllobacterium</i> 1029 OF F MK638	3278
sp. PN3-B06P4- July 2015 Streptomyces sp. 1060 OF F MK638	3496
10	
PN3-B06P4- July 2015 <i>Streptomyces sp.</i> 1059 OF F MK638	3506
11	
PN3-B06P4- July 2015 <i>Phyllobacterium</i> 1029 OF F MK638	3302
12	
PN3-B06P4- July 2015 <i>Phyllobacterium</i> 1031 OF F MK638	3294
Sp. 1064 OF F. MK638	2402
PN3-B06P4- July 2015 <i>Streptomyces sp.</i> 1064 OF F MK638	5493
PN3-B06P4- July 2015 <i>Streptomyces sp.</i> 1055 OF F MK638	8494
15	, 1,5,7
PN3-B06P4- July 2015 <i>Phyllobacterium</i> 1055 OF F MK638	3268
16	
PN3-B06P4- July 2015 Streptomyces sp. 1056 OF F MK638	3590
17	

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PN3-B06P4-	July 2015	Streptomyces sp.	1025	OF	F	MK638498
18						
PN3-B06P4-	July 2015	Streptomyces sp.	1062	OF	F	MK638499
19						
PN3-B06P4-	July 2015	Phyllobacterium	1055	OF	F	MK638290
20	•	sp.				
PN3-B06P4-	July 2015	Phyllobacterium	1061	OF	F	MK638267
21	,	sp.		•	·	
PN3-B06P4-	July 2015	Phyllobacterium	1037	OF	F	MK638303
22	July 2013	•	1037	O1	'	WINOSOSOS
	I.J. 2015	Sp.	1056	OF	F	MK638500
PN3-B06P4-	July 2015	Streptomyces sp.	1056	UF	Г	IVINOSOSOU
23		C	4070	0.5	_	NAV620407
PN3-B07P1-1	July 2015	Streptomyces sp.	1073	OF	F	MK638497
PN3-B07P1-3	July 2015	Phyllobacterium	1061	OF	F	MK638280
		sp.				
PN3-B07P1-4	July 2015	Phyllobacterium	1027	OF	F	MK638293
		sp.				
PN3-B07P1-5	July 2015	Phyllobacterium	1034	OF	F	MK638306
	-	sp.				
PN3-B07P1-6	July 2015	Phyllobacterium	1062	OF	F	MK638286
	,	sp.				
PN3-B07P1-7	July 2015	Phyllobacterium	1006	OF	F	MK638310
1110 0071 17	34.y 2013	sp.	1000	O1	•	1411(030310
PN3-B07P1-8	July 2015	Phyllobacterium	1033	OF	F	MK638301
FN3-D0/F1-0	July 2013	•	1033	Oi	ı	MIK026201
DN2 D07D1 0	I.J. 2015	Sp.	1064	OΓ	F	N4VC20E40
PN3-B07P1-9	July 2015	Streptomyces sp.	1064	OF	F	MK638549
PN3-B07P1-	July 2015	Pseudomonas sp.	900	OF	F	MK638011
11		_			_	
PN3-B07P1-	July 2015	Streptomyces sp.	1056	OF	F	MK638591
12						
PN3-B07P1-	July 2015	Phyllobacterium	1066	OF	F	MK638289
13		sp.				
PN3-B07P1-	July 2015	Phyllobacterium	1035	OF	F	MK638307
14		sp.				
PN3-B07P1-	July 2015	Streptomyces sp.	1053	OF	F	MK638542
15	•					
PN3-B07P1-	July 2015	Streptomyces sp.	1046	OF	F	MK638487
16	,	, , ,				
PN3-B07P1-	July 2015	Phyllobacterium	1046	OF	F	MK638292
17	34.y 2013	sp.	1010	O.	•	1411(030232
PN3-B07P1-	July 2015	Streptomyces sp.	1056	OF	F	MK638541
18	July 2013	Streptomyces sp.	1030	Oi	'	1411020241
	I.J. 2015	Dhullahaatarium	000	OΓ	F	NAVCOROL
PN3-B07P1-	July 2015	Phyllobacterium	999	OF	F	MK638295
19		sp.	405.	0.5	-	NAVC22224
PN3-B07P1-	July 2015	Phyllobacterium	1054	OF	F	MK638281
20		sp.		_		
PN3-B07P1-	July 2015	Phyllobacterium	999	OF	F	MK638300
21		sp.				

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PN3-B07P1-	July 2015	Phyllobacterium	1035	OF	F	MK638299
22		sp.				
PN3-B07P5-1	July 2015	Phyllobacterium sp.	1030	CF	F	MK638283
PN3-B07P5-2	July 2015	Phyllobacterium	1021	CF	F	MK638376
PN3-B07P5-3	July 2015	sp. Phyllobacterium	1006	CF	F	MK638311
PN3-B07P5-4	July 2015	sp. Phyllobacterium	1010	CF	F	MK638305
PN3-B07P5-5	July 2015	sp. Agrobacterium sp.	946	CF	F	MK638037
PN3-B07P5-6	July 2015	Mesorhizobium sp.	1027	CF	F	MK638178
PN3-B07P5-7	July 2015	Mesorhizobium sp.	1030	CF	F	MK638179
PN3-B07P5-8	July 2015	Streptomyces sp.	1064	CF	F	MK638495
	•					
PN3-B07P5-9	July 2015	Streptomyces sp.	1018	CF	F	MK638630
PN3-B07P5- 10	July 2015	Agrobacterium sp.	1065	CF	F	MK638001
PN3-B07P5- 11	July 2015	Streptomyces sp.	670	CF	F	MK638628
PN3-B07P5- 12	July 2015	Streptomyces sp.	1059	CF	F	MK638488
PN3-B07P5- 13	July 2015	Streptomyces sp.	1018	CF	F	MK638629
PN3-B07P5- 14	July 2015	Streptomyces sp.	1070	CF	F	MK638501
PN3-B07P5- 15	July 2015	Ensifer sp.	1072	CF	F	MK638124
PN3-B07P5- 16	July 2015	Mesorhizobium sp.	1070	CF	F	MK638180
PN3-B07P5- 18	July 2015	Inquilinus sp.	1035	CF	F	MK638141
PN3-B07P5- 19	July 2015	Agrobacterium sp.	1030	CF	F	MK638036
PN3-B07P5- 20	July 2015	Phyllobacterium sp.	925	CF	F	MK638314
PN3-B07P5- 22	July 2015	Mesorhizobium sp.	1072	CF	F	MK638181
PN3-B07P5- 23	July 2015	Pseudomonas sp.	1030	CF	F	MK638006
PN3-B08P4-1	July 2015	Leifsonia sp.	1059	CF	Α	MK638148
PN3-B08P4-2	July 2015 July 2015	Agrobacterium sp.	1051	CF	A	MK638032
PN3-B08P4-2	•		1070	CF		MK638030
	July 2015	Agrobacterium sp.			A	
PN3-B08P4-4	July 2015	Agrobacterium sp.	1032	CF	A	MK638035
PN3-B08P4-5	July 2015	Agrobacterium sp.	1071	CF	Α	MK638031
PN3-B08P4-6	July 2015	Agrobacterium sp.	1029	CF	Α	MK638033
PN3-B08P4-7	July 2015	Phyllobacterium sp.	1070	CF	Α	MK638327
PN3-B08P4-9	July 2015	Agrobacterium sp.	1068	CF	Α	MK638027

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PN3-B08P4-	July 2015	Agrobacterium sp.	1029	CF	А	MK638029
10						
PN3-B08P4-	July 2015	Agrobacterium sp.	1022	CF	Α	MK638034
11	-					
PN3-B08P4-	July 2015	Phyllobacterium	1042	CF	Α	MK638338
12	,	sp.				
PN3-B08P4-	July 2015	Pseudomonas sp.	1057	CF	Α	MK638004
14	July 2010	, seadomendo sp.	2007	O.	, ,	
PN3-B08P4-	July 2015	Phyllobacterium	1027	CF	Α	MK638370
15	July 2013	sp.	1027	C.	, ,	1411(030370
PN3-B08P4-	July 2015	Agrobacterium sp.	1061	CF	Α	MK638028
16	July 2013	Agrobacteriam sp.	1001	Ci	Α	1411030020
PN3-B08P4-	July 2015	Pseudomonas sp.	1085	CF	Α	MK638003
18	July 2013	r seudomonus sp.	1005	Ci	^	MIKOSOOS
PN3-B08P4-	July 2015	Phyllobacterium	1068	CF	Α	MK638319
19	July 2013	•	1000	CF	A	INIKOSOSIS
	Luby 2015	Sp.	1072	CE	۸	NAVC20247
PN3-B08P4-	July 2015	Phyllobacterium	1072	CF	Α	MK638347
21 DN2 D00D4	I.J. 2015	sp.	1046	CE	۸	NAVC20140
PN3-B08P4-	July 2015	Leifsonia sp.	1046	CF	Α	MK638149
22		2 /	4040	0.5		. 41/620000
PN3-B08P5-1	July 2015	Pseudomonas sp.	1043	OF	Α	MK638008
PN3-B08P5-2	July 2015	Phyllobacterium	1030	OF	Α	MK638334
		sp.				
PN3-B08P5-3	July 2015	Pseudomonas sp.	1056	OF	Α	MK638005
PN3-B08P5-4	July 2015	Phyllobacterium	1073	OF	Α	MK638340
		sp.				
PN3-B08P5-5	July 2015	Phyllobacterium	1068	OF	Α	MK638320
		sp.				
PN3-B08P5-6	July 2015	Streptomyces sp.	1070	OF	Α	MK638502
PN3-B08P5-7	July 2015	Pseudomonas sp.	1043	OF	Α	MK638007
PN3-B08P5-8	July 2015	Phyllobacterium	1068	OF	Α	MK638329
		sp.				
PN3-B08P5-9	July 2015	Agrobacterium sp.	1072	OF	Α	MK638026
PN3-B08P5-	July 2015	Phyllobacterium	1034	OF	Α	MK638352
10		sp.				
PN3-B08P5-	July 2015	Microbacterium	855	OF	Α	MK638192
11		sp.				
PN3-B08P5-	July 2015	Streptomyces sp.	1018	OF	Α	MK638631
12						
PN3-B08P5-	July 2015	Phyllobacterium	1054	OF	Α	MK638331
13		sp.				
PN3-B08P5-	July 2015	Streptomyces sp.	955	OF	Α	MK638632
14		·				
PN3-B08P5-	July 2015	Phyllobacterium	1070	OF	Α	MK638426
15	=	sp.				
PN3-B08P5-	July 2015	Streptomyces sp.	561	OF	Α	MK638633
16	•	,				
PN3-B08P5-	July 2015	Phyllobacterium	1042	OF	Α	MK638346
18	, -	sp.				
		- F-				

PN3-B08P5- 19	July 2015	Phyllobacterium	1070	OF	Α	MK638428
		sp.	4070	0.5		N 41/C20420
PN3-B08P5-	July 2015	Rhizobium sp.	1070	OF	Α	MK638430
20		51 11 1	4000	0.5		N.41/C20224
PN3-B08P5-	July 2015	Phyllobacterium	1068	OF	Α	MK638321
21		sp.	4074	0.5		
PN3-B08P5-	July 2015	Streptomyces sp.	1071	OF	Α	MK638504
22		Class to the second	4020	0.5	Δ.	NAVC20447
PN3-B08P5-	July 2015	Stenotrophomonas	1039	OF	Α	MK638447
23	I. d. 2015	sp.	055	0.5	۸	N4VC20C24
PN3-B08P5-	July 2015	Streptomyces sp.	955	OF	Α	MK638634
24 PN3-B09P3-2	July 2015	Encifor on	1070	CF	F	MK638127
PN3-B09P3-2	July 2015 July 2015	Ensifer sp. Afipia sp.	834	CF	F	MK638667
PN3-B09P3-4	July 2015 July 2015		1070	CF	F	MK638503
PN3-B09P3-4 PN3-B09P3-5	July 2015 July 2015	Streptomyces sp. Streptomyces sp.	1070	CF	F F	MK638638
PN3-B09P3-6	July 2015 July 2015	Streptomyces sp. Streptomyces sp.	1019	CF	F F	MK638638
PN3-B09P3-8	July 2015 July 2015	Phyllobacterium	1018	CF	F	MK638317
PN3-D03P3-0	July 2015	•	1051	CF	Г	IVINUSOSTI
PN3-B09P3-9	July 2015	sp. Bosea sp.	1072	CF	F	MK638084
PN3-B09P3-	July 2015 July 2015	Phyllobacterium	1072	CF	r F	MK638425
10	July 2013	sp.	1075	Ci		WIN030423
PN3-B09P3-	July 2015	Ensifer sp.	1070	CF	F	MK638126
11	July 2013	Ensiger sp.	1070	Ci	•	WIN030120
PN3-B09P3-	July 2015	Streptomyces sp.	660	CF	F	MK638635
12	July 2010	ou epioniy des spi		O.	•	······································
PN3-B09P3-	July 2015	Phyllobacterium	1070	CF	F	MK638335
14	•	sp.				
PN3-B09P3-	July 2015	Phyllobacterium	1029	CF	F	MK638343
15		sp.				
PN3-B09P3-	July 2015	Phyllobacterium	1070	CF	F	MK638322
16		sp.				
PN3-B09P3-	July 2015	Ensifer sp.	1029	CF	F	MK638128
17						
PN3-B09P3-	July 2015	Phyllobacterium	1052	CF	F	MK638351
18		sp.				
PN3-B09P3-	July 2015	Streptomyces sp.	1018	CF	F	MK638636
19						
PN3-B09P3-	July 2015	Streptomyces sp.	1018	CF	F	MK638637
20						
PN3-B09P3-	July 2015	Phyllobacterium	1055	CF	F	MK638353
21		sp.				
PN3-B09P3-	July 2015	Phyllobacterium	1052	CF	F	MK638350
22		sp.			_	
PN3-B09P5-1	July 2015	Phyllobacterium	1068	OF	F	MK638318
DNI2 D2277 C		sp.	4070	0.5	-	N 41/ C 2 2 2 2 2
PN3-B09P5-2	July 2015	Phyllobacterium	1070	OF	F	MK638330
		sp.				

						Unapier 3
PN3-B09P5-3	July 2015	Phyllobacterium sp.	1062	OF	F	MK638341
PN3-B09P5-4	July 2015	Phyllobacterium sp.	1071	OF	F	MK638315
PN3-B09P5-5	July 2015	Phyllobacterium sp.	1073	OF	F	MK638324
PN3-B09P5-6	July 2015	Phyllobacterium sp.	1025	OF	F	MK638371
PN3-B09P5-7	July 2015	Phyllobacterium sp.	1011	OF	F	MK638373
PN3-B09P5-8	July 2015	Streptomyces sp.	393	OF	F	MK638643
PN3-B09P5-9	July 2015	Streptomyces sp.	1018	OF	F	MK638644
PN3-B09P5-	July 2015	Paenarthrobacter	1040	OF	F	MK638200
10	34., 2013	sp.	20.0	0.	•	
PN3-B09P5- 11	July 2015	Phyllobacterium sp.	944	OF	F	MK638357
PN3-B09P5- 12	July 2015	Phyllobacterium sp.	1070	OF	F	MK638328
PN3-B09P5- 13	July 2015	Phyllobacterium sp.	1070	OF	F	MK638427
PN3-B09P5- 14	July 2015	Streptomyces sp.	1018	OF	F	MK638640
PN3-B09P5- 15	July 2015	Streptomyces sp.	1018	OF	F	MK638641
PN3-B09P5- 16	July 2015	Phyllobacterium sp.	1034	OF	F	MK638356
PN3-B09P5- 17	July 2015	Phyllobacterium sp.	1070	OF	F	MK638431
PN3-B09P5- 18	July 2015	Streptomyces sp.	601	OF	F	MK638642
PN3-B09P5- 19	July 2015	Phyllobacterium sp.	1034	OF	F	MK638354
PN3-B09P5- 20	July 2015	Phyllobacterium sp.	1026	OF	F	MK638358
PN3-B09P5- 22	July 2015	, Phyllobacterium sp.	1052	OF	F	MK638349
PN3-B10P3-1	July 2015	Microbacterium sp.	1020	OF	Α	MK638189
PN3-B10P3-2	July 2015	Phyllobacterium sp.	1060	OF	Α	MK638344
PN3-B10P3-4	July 2015	Phyllobacterium sp.	1070	OF	Α	MK638316
PN3-B10P3-5	July 2015	Agrobacterium sp.	1008	OF	Α	MK638049
PN3-B10P3-6	July 2015	Streptomyces sp.	813	OF	Α	MK638617
PN3-B10P3-7	July 2015	Phyllobacterium	1023	OF	Α	MK638372
PN3-B10P3-8	July 2015	sp. Phyllobacterium sp.	1008	OF	Α	MK638365
PN3-B10P3-9	July 2015	Rhizobium sp.	718	OF	Α	MK638434

<u> </u>						
PN3-B10P3-	July 2015	Phyllobacterium	1025	OF	Α	MK638369
PN3-B10P3-	July 2015	sp. Kaistia sp.	1008	OF	Α	MK638142
12 PN3-B10P3-	July 2015	Phyllobacterium	1008	OF	Α	MK638360
13 PN3-B10P3-	July 2015	sp. Streptomyces sp.	1018	OF	Α	MK638615
14 PN3-B10P3-	July 2015	Phyllobacterium	1008	OF	Α	MK638361
15 PN3-B10P3-	July 2015	sp. Phyllobacterium	1008	OF	Α	MK638362
16 PN3-B10P3-	July 2015	sp. Phyllobacterium	1008	OF	Α	MK638363
17 PN3-B10P3-	July 2015	sp. Streptomyces sp.	750	OF	Α	MK638616
18 PN3-B10P3-	July 2015	Phyllobacterium	1008	OF	А	MK638364
19 PN3-B10P3-	July 2015	sp. Phyllobacterium	1053	OF	А	MK638355
20 PN3-B10P3-	July 2015	sp. Rhizobium sp.	1028	OF	A	MK638429
21 PN3-B10P3-	July 2015	Rhizobium sp.	1021	OF	A	MK638435
22	•	·				
PN3-B10P5-1	July 2015	Phyllobacterium sp.	686	CF	Α	MK638313
PN3-B10P5-2	July 2015	Phyllobacterium sp.	1073	CF	Α	MK638339
PN3-B10P5-3	July 2015	Phyllobacterium sp.	1062	CF	Α	MK638345
PN3-B10P5-4	July 2015	Mycobacterium sp.	442	CF	Α	MK638196
PN3-B10P5-5	July 2015	Phyllobacterium sp.	1052	CF	Α	MK638337
PN3-B10P5-7	July 2015	Phyllobacterium sp.	1048	CF	Α	MK638333
PN3-B10P5-8	July 2015	Phyllobacterium sp.	1037	CF	Α	MK638368
PN3-B10P5-9	July 2015	Phyllobacterium sp.	1067	CF	Α	MK638325
PN3-B10P5-	July 2015	Phyllobacterium	1071	CF	Α	MK638326
10 PN3-B10P5-	July 2015	sp. Phyllobacterium	1055	CF	Α	MK638342
11 PN3-B10P5-	July 2015	sp. Phyllobacterium	1071	CF	Α	MK638323
12 PN3-B10P5-	July 2015	sp. Variovorax sp.	1050	CF	Α	MK638658
13 PN3-B10P5-	July 2015	Phyllobacterium	1061	CF	Α	MK638348
14		sp.				

PN3-B10P5-	July 2015	Phyllobacterium	1027	CF	Α	MK638375
15		sp.				
PN3-B10P5-	July 2015	Ensifer sp.	1073	CF	Α	MK638125
16						
PN3-B10P5-	July 2015	Phyllobacterium	1070	CF	Α	MK638332
17		sp.				
PN3-B10P5-	July 2015	Streptomyces sp.	1018	CF	Α	MK638645
18						
PN3-B10P5-	July 2015	Streptomyces sp.	970	CF	Α	MK638486
19						
PN3-B10P5-	July 2015	Phyllobacterium	1029	CF	Α	MK638308
20		sp.	_	_		
PN3-B10P5-	July 2015	Streptomyces sp.	955	CF	Α	MK638485
21		-1 11 1			_	
PN3-B10P5-	July 2015	Phyllobacterium	1053	CF	Α	MK638304
22		sp.	700	0=		
PN3-B10P5-	July 2015	Phyllobacterium	708	CF	Α	MK638312
23		sp.				

Supplementary Table 2: Properties of the studied soils in the treatments of the GCEF. Values represent means and standard deviations (in parenthesis) for soil moisture, pH, total carbon content (TOC), total nitrogen content (TN), mineral nitrogen (N_{min}) and available phosphorus (P_{DL}) of five replicates for each treatment. Treatments include conventional (CF) and organic farming (OF) under ambient (A) and future (F) climatic conditions implemented in the GCEF. Samples were taken in May and July 2015.

	Ma	y 2015 (F	BBCH 37-	39)	Jul	ly 2015 (E	BBCH 75-	77)
Treatment	CF-A	CF-F	OF-A	OF-F	CF-A	CF-F	OF-A	OF-F
Moisture	10.9a	10.1a	10.6a	10.6a	15.9a	15.9a	16.0a	15.8a
[% w/w]	(0.6)	(0.8)	(0.6)	(0.4)	(0.2)	(0.5)	(0.2)	(0.5)
pН	6.8a	6.8a	6.5a	6.6a	6.8a	6.8a	6.6a	6.6a
þп	(0.4)	(0.4)	(0.6)	(0.5)	(0.4)	(0.5)	(0.6)	(0.5)
TOC	2.0a	2.0a	2.0a	1.9a	1.9a	1.9a	1.9a	1.9a
[% w/w]	(0.1)	(0.1)	(0.1)	(0.1)	(0.2)	(0.2)	(0.1)	(0.2)
TN	0.17a	0.17a	0.16a	0.16a	0.16a	0.15a	0.15a	0.15a
[% w/w]	(0.01)	(0.01)	(0.01)	(0.01)	(0.02)	(0.02)	(0.01)	(0.02)
$\mathbf{N}_{\mathbf{min}}$	13.4ab	9.5b	5.0c	4.8c	12.3ab	13.7a	9.6ab	7.7bc
[mg/kg]	(5.1)	(2.6)	(1.4)	(1.0)	(2.1)	(1.6)	(4.4)	(1.2)
$\mathbf{P}_{\mathbf{DL}}$	7.6a	8.4a	6.8a	6.0a	7.3a	8.9a	7.2a	6.0a
[mg/100g]	(4.1)	(3.7)	(3.7)	(2.5)	(3.2)	(4.3)	(4.2)	(2.2)

Chapter 3

Supplementary Table 3: Winter wheat yields in the GCEF in 2015. Corn and straw yields are presented for conventional and organic farming systems under ambient and future climatic conditions. Significant differences in yield between the four treatments are marked by different letters according to analysis of variance and Tukey post hoc test.

Land use	Climate	Crop	Corn yield	Straw yield	Ratio
			in dt/ha	in dt/ha	corn/straw
Conventional	Ambient	Winter	83.4 a	69.0	0.82
Farming		wheat	(3.2)	(4.3)	(0.04)
Conventional	Future	Winter	69.0 b	58.3	0.85
Farming		wheat	(10.4)	(9.7)	(0.02)
Organic Farming	Ambient	Winter	82.9 a	71.0	0.86
		wheat	(4.9)	(4.5)	(0.06)
Organic Farming	Future	Winter	74.4 b	69.2	0.93
		wheat	(5.0)	(6.1)	(0.03)

Supplementary Table 4: Number of isolated colonies cultivated on Pikovskaya medium. Plot code refers to the coding of the GCEF plots, and land use and climate treatment of each plot is indicated. The land use treatment includes conventional and organic farming, and the climate ambient and future climatic conditions. Samples were taken in May and July 2015.

Plot	Land use	Climate	May 2015 (BBCH 37- 39)	July 2015 (BBCH 75- 77)
B01P2	Conventional Farming	Ambient	20	19
B01P4	Organic Farming	Ambient	22	22
B02P3	Organic Farming	Future	21	21
B02P4	Conventional Farming	Future	20	14
B03P1	Conventional Farming	Ambient	23	19
B03P2	Organic Farming	Ambient	23	21
B04P1	Conventional Farming	Future	20	18
B04P2	Organic Farming	Future	22	22
B05P1	Organic Farming	Ambient	19	19
B05P3	Conventional Farming	Ambient	21	22
B06P2	Conventional Farming	Future	18	23
B06P4	Organic Farming	Future	20	23
B07P1	Organic Farming	Future	20	20
B07P5	Conventional Farming	Future	22	21
B08P4	Conventional Farming	Ambient	20	18
B08P5	Organic Farming	Ambient	21	23
B09P3	Conventional Farming	Future	20	19
B09P5	Organic Farming	Future	20	21
B10P3	Organic Farming	Ambient	18	20
B10P5	Conventional Farming	Ambient	20	22

Supplementary Table 5: Bacteria cultivated on Pikovskaya medium from rhizosphere soil of winter wheat. Bacteria were identified by partial *16S rRNA* sequencing. Distribution of genera among conventional (CF) and organic farming system (OF), ambient (A) and future climate conditions (F), as well as sampling time, May and July 2015.

Genus	Total	May 2015 (BBCH 37- 39)	July 2015 (BBCH 75-77)	CF	OF	A	F
Achromobacter	1	1	0	1	0	1	0
Afipia	1	0	1	1	0	0	1
Agrobacterium	30	10	20	21	9	24	6
Agromyces	2	1	1	2	0	2	0
Arthrobacter	8	8	0	3	5	3	5
Bacillus	21	12	9	5	16	5	16
Bosea	1	0	1	1	0	0	1
Bradyrhizobium	3	2	1	1	2	1	2
Burkholderia	3	2	1	0	3	2	1
Buttiauxella	8	8	0	8	0	0	8
Caulobacter	2	1	1	2	0	1	1
Chitinophaga	2	2	0	2	0	0	2
Chryseobacterium	2	1	1	2	0	0	2
Clavibacter	1	1	0	1	0	1	0
Devosia	1	0	1	1	0	0	1
Dyadobacter	3	3	0	2	1	0	3
Dyella	5	2	3	3	2	5	0
Ensifer	11	1	10	6	5	6	5
Flavobacterium	4	4	0	2	2	2	2
Herbiconiux	1	0	1	1	0	0	1
Inquilinus	8	1	7	5	3	7	1
Kaistia	1	0	1	0	1	1	0
Leifsonia	7	1	6	6	1	7	0
Luteibacter	3	0	3	0	3	1	2
Massilia	1	1	0	0	1	1	0
Mesorhizobium	32	16	16	16	16	20	12
Microbacterium	7	2	5	4	3	4	3
Mucilaginibacter	1	1	0	1	0	0	1
Mycobacterium	3	0	3	1	2	1	2
Ochrobactrum	3	3	0	0	3	0	3
Paenarthrobacter	1	0	1	0	1	0	1
Pantoea	1	1	0	1	0	0	1
Paraburkholderia	2	1	1	0	2	1	1
Pedobacter	1	1	0	1	0	1	0
Phyllobacterium	184	37	147	96	88	77	107
Plantibacter	5	5	0	0	5	1	4
Pseudaminobacte	1	0	1	1	0	0	1
r Doggan	1.7	150	17	02	0.5	01	76
Pseudomonas Di : 1:	167	150	17	82	85 25	91	76
Rhizobium	44	20	24	19	25	29	15
Rhodococcus	5	5	0	2	3	3	2

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D				4	4		
Rugamonas	5	5	0	4	1	2	3
Serratia	2	2	0	1	1	1	1
Sphingomonas	1	1	0	1	0	1	0
Stenotrophomona	8	6	2	1	7	5	3
S							
Streptomyces	197	81	116	85	112	95	102
Tardiphaga	2	1	1	1	1	0	2
Tsukamurella	2	2	0	2	0	0	2
Variovorax	13	8	5	4	9	10	3
Total	817	410	407	399	418	412	405

Supplementary Table 6: Abundances and activity levels of dominant genera and their respective phylogenetic clusters. Numbers of isolates of each cluster are listed for May (BBCH 37-39) and July 2015 (BBCH 75-77) samples, in conventional (CF) and organic farming systems (OF), as well as under ambient (A) and future (F) climatic conditions. Mean activities ± standard deviations for phosphate solubilization, PSI 1 and PSI 2, and mean drought resistance are given for each cluster. Significant differences in phosphate solubilizing and drought resistance, between clusters are marked by different letters according to analysis of variance and Tukey post hoc test.

Cluster	May	July	CF	OF	A	F	PSI 1 in μg PO ₄ ³⁻	PSI 2 in µg PO ₄ ³⁻	Resistance in %
Phyllobacterium Phyllobacterium									
	33	142	94	81	73	102	456.1 ± 16.7	126.6 ± 9.7	61.7 ± 1.8
C 1	24	97	76	45	50	71	$452.3 \pm 20.2^{\circ}$	117.8 ± 10.8^{b}	58.8 ± 1.7^{ac}
C 2	9	45	18	36	23	31	$463.3 \pm 27.9^{\circ}$	143.5 ± 18.4^{b}	67.1 ± 3.7^{a}
							Pseudomonas		
	150	19	83	86	91	78	1027.3 ± 51.2	258.8 ± 26.8	37.3 ± 2.6
C 1	38	1	26	13	24	15	1378.7 ± 104.5^{a}	433.8 ± 69.5^{a}	31.0 ± 2.5^{b}
C 2	23	2	12	13	10	15	637.81 ± 95.6^{bc}	142.3 ± 32.7^{bc}	25.0 ± 0^{bc}
C 3	17	3	7	13	8	12	1076.2 ± 149.7^{ab}	236.7 ± 41.5^{ab}	35.6 ± 5^{abc}
C 4	18	1	6	13	18	1	758.89 ± 72^{b}	68.1 ± 37.8^{bc}	39.5 ± 9.9^{bc}
C 5	7	1	1	7	2	6	1151.5 ± 178.8	251.0 ± 91.1	40.1 ± 2.7
C 6	6	0	3	3	3	3	100.87 ± 56.4	0.7 ± 0.5	NA
C 7	5	0	3	2	3	2	951.77 ± 284.4	196.7 ± 109.6	56.7 ± 19.7
C 8	5	0	4	1	1	4	2386.8 ± 75.4	1028.4 ± 113	41.4 ± 7.1
C 9	3	1	2	2	2	2	538.43 ± 110.9	88.2 ± 19.2	0.0
C 10	3	1	4	0	3	1	688.32 ± 218.2	126.5 ± 126	43.9 ± 14.2
C 11	0	3	0	3	0	3	290.92 ± 105.9	0.0 ± 0	0.0
C 12	1	2	1	2	3	0	1450.8 ± 0	0.0 ± 0	100.0 ± 0
C 13	3	0	1	2	1	2	692.48 ± 402.9	71.5 ± 39.8	30.0 ± 0
C 14	2	0	2	0	1	1	1276.7 ± 174.1	105.4 ± 10	46.5 ± 20.2
C 15	2	0	0	2	2	0	431.25 ± 30.4	1.5 ± 0.6	75.4 ± 6.7
C 16	1	1	2	0	1	1	1682.5 ± 0	504.6 ± 0	0.0
C 17	2	0	0	2	1	1	508.39 ± 87.8	11.9 ± 11.9	$28.6\ \pm0$
C 18	2	0	0	2	2	0	1451.7 ± 163.6	483.1 ± 21.5	25.4 ± 16.3
C 19	1	1	1	1	1	1	584.22 ± 0	85.8 ± 51.5	16.7 ± 0
C 20	1	0	1	0	0	1	344.33 ± 0	61.0 ± 0	33.3 ± 0
C 21	1	0	1	0	1	0	NA	NA	NA
C 22	1	0	1	0	1	0	NA	NA	NA
C 23	1	0	0	1	0	1	1236.2 ± 0	$420.64 \ \pm 0$	NA
C 24	1	0	1	0	0	1	1763.6 ± 0	420.6 ± 0	30.0 ± 0
C 25	1	0	0	1	0	1	NA	NA	NA
C 26	1	0	0	1	0	1	1038.7 ± 0	105.16 ± 0	42.9 ± 0
C 27	1	0	1	0	1	0	802.16 ± 0	15.3 ± 0	0.0
C 28	1	0	1	0	1	0	NA	NA	NA
C 29	1	0	0	1	0	1	86.082 ± 0	$0.2385 \ \pm 0$	46.7 ± 0
C 30	1	0	1	0	1	0	137.35 ± 0	34.3 ± 0	NA
							Streptomyces		
	81	116	85	112	95	102	553.7 ± 16.9	25.3 ± 2.8	45.4 ± 0.1
C 1	48	51	55	44	45	54	580.4 ± 30.6^{bc}	$20.8\ \pm 4.5^{c}$	46.3 ± 3^{b}

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C 2	9	16	9	16	12	13	687.9 ± 62.9^{b}	41.8 ± 5.5^{bc}	45.8 ± 6.1^{bc}
C 3	3	11	5	9	9	5	562.8 ± 70.9^{bc}	24.6 ± 6.7^{bc}	26.4 ± 9.3^{b}
C 4	8	5	5	8	8	5	431.9 ± 55^{bc}	31.2 ± 8.8^{bc}	41.1 ± 9.7^{ab}
C 5	2	11	4	9	3	10	384.8 ± 86.5^{bc}	40.3 ± 11.8^{bc}	71.6 ± 11.8^{ac}
C 6	2	6	0	8	4	4	586.9 ± 77.6	22.0 ± 8.3	31.8 ± 11.8
C 7	3	1	1	3	3	1	441.2 ± 158	19.7 ± 19.2	33.7 ± 23.6
C 8	0	4	0	4	2	2	291.5 ± 65.2	44.4 ± 25.6	50.7 ± 15.9
C 9	0	3	2	1	1	2	118.4 ± 10.8	3.2 ± 2.2	59.5 ± 29.8
C 10	1	1	0	2	1	1	858.4 ± 0	27.7 ± 19.1	46.5 ± 9
C 11	1	0	1	0	1	0	236.6 ± 0	0.5 ± 0	66.7 ± 0
C 12	1	0	0	1	1	0	NA	NA	NA
C 13	1	0	0	1	1	0	504.6 ± 0	0.0 ± 0	50.0 ± 0
C 14	1	0	1	0	1	0	34.3 ± 0	0.0 ± 0	0.0
C 15	0	1	0	1	0	1	292.1 ± 0	2.1 ± 0	81.8 ± 0
C 16	1	0	0	1	1	0	596.1 ± 0	3.8 ± 0	61.1 ± 0
C 17	0	1	0	1	1	0	774.7 ± 0	0.2 ± 0	65.6 ± 0
C 18	0	1	0	1	1	0	309.04 ± 0	8.6 ± 0	$100.0\ \pm 0$
C 19	0	1	1	0	0	1	504.6 ± 0	8.6 ± 0	0.0
C 20	0	1	0	1	0	1	68.9 ± 0	6.0 ± 0	66.7 ± 0
C 21	0	1	1	0	0	1	596.1 ± 0	46.7 ± 0	44.1 ± 0
C 22	0	1	0	1	0	1	381.5 ± 0	95.4 ± 0	30.3 ± 0

SYNOPSIS

The present thesis aims to contribute unravelling structural and functional adaptation processes in the wheat rhizobacterial communities to drought, and how these changes induced by drought are further influenced by abiotic and biotic factors. The three studies that constitute the thesis use cultivation-based and -independent methods to gain a broad perspective on the complex interplay between plants and soil in agricultural systems under semi-controlled conditions in the cold greenhouse and under natural field conditions (Global Change Experimental Facility, Schädler et al. 2019). In particular, the impact of farming system and water availability/climate (all experiments, Chapter 1 to 3), soil type and wheat cultivar (pot experiment, Chapter 1), as well as plant growth stage (field experiment, Chapter 2 and 3) were investigated. The main findings of single and joint effects of the three studies are summarized under the following bullet points:

- The drought treatment in the pot experiment dramatically shaped rhizobacterial community composition of wheat, with a strong potential for degradation of complex carbon compounds within the obtained drought-adapted community (Chapter 1). In contrast, the climate treatment in the field experiment had only a minor effect on community composition, but led to an overall decrease in extracellular enzyme activities (Chapter 2). Since the GCEF was established in 2013 and samples were taken in May and July 2015, the moderate climate treatment will probably only result in changes in soil conditions that will accumulate over time. Therefore, the specific adaptations are likely to become more evident in subsequent years. However, an effect was detected in the fraction of cultivated bacteria already indicating a higher drought-stress resistance potential among dominant bacterial species under future climate conditions (Chapter 3). This result suggest adaption of single species to drought, which are drowning in the community.
- When screening for effective phosphate solubilizing bacteria in the rhizosphere of wheat grown under field conditions, we found, besides already known representatives of the phyla Pseudomonas and Streptomyces, Phyllobacterium species to be dominant

- (Chapter 3). Since our findings proved for the first time *Phyllobacterium* species as effective phosphate-solubilzers in the rhizosphere of wheat grown in temperate climate zone, and since they further expressed a strong resistance to drought, species of this phylum might be potential candidates for future application as biofertilizer.
- Apart from the drought treatment, the cultivated soil type was the strongest driver of community composition and function in the pot experiment, followed by agricultural practice and no significant effect of wheat cultivar was detected (Chapter 1), which is in concordance with previous results (Schlatter et al 2020). The effect of wheat cultivar was specific to increased production of carbon cycling enzymes under drought in dependency of soil type and farming system in the rhizosphere of the demanding cultivar (Chapter 1). In the field experiment, growth stage mainly shaped rhizobacterial community composition, whereby extracellular enzyme production was mainly affected by agricultural practice (Chapter 2). These results demonstrate the complexity of the studied topic and plead for more studies on the interconnecting effects of multiple experimental factors to unravel common and more specific adaptations of bacterial communities in structure and function, especially with respect to the impact of climate change.
- The evaluation of computational prediction tools Tax4Fun and PanFP in relation to the different treatments applied, i.e. farming system, climate and plant growth stages in Chapter 2, proved qualitative concordances between predicted abundances of functional genes and the respective corresponding measured enzyme activities. This suggests that we can use this method to conduct time- and cost-saving research by pre-analyzing soil bacterial communities from various environments for differences in their functional profiles. Building on that more advanced analysis with a higher resolution may be applied specifically to further quantitatively assess functions of interest.

Multidisciplinary approaches combine advantages of cultivationbased and high throughput community-based methods

For a long time, methodical approaches to study diversity and functions of soil bacterial communities were limited to cultivation-based approaches. Microbes from environmental samples were isolated by plating, purified, classified and characterized (reviewed in Hugerth and Andersson, 2017). By the use of this approach, detailed physiological studies could be performed and a comprehensive functional profile of a given organism obtained, which is a prerequisite for the identification of potential plant growth promoting rhizobacteria (PGPR) (term first introduced by Kloepper and Schroth, 1978). In Chapter 3, we applied this method in a modified form. The traditional approach was to cultivate bacterial colonies on different media to obtain diverse collections with potentially wide range of functions from the environmental sample. Whereas such cultivation strategy may yield a diverse collection of bacteria, it does not necessarily lead to the isolation of potentially plant beneficial ones (reviewed in Alaylar et al 2020). Avoiding the random selection we steered the composition of the culture collection towards a specific functional trait by screening for phosphate-solubilizing bacteria on Pikovskaya medium (Pikovskaya 1948), and subsequently tested for further functional traits (Chapter 3). To capture both fast and slowly growing bacterial species we prepared soil solutions with a high dilution factor (1:200) and let the colonies grow for two weeks. The large culture collection we obtained in that way comprised around 800 functionally classified bacterial isolates from different environmental samples.

Since one gram of soil can harbor millions of microorganisms representing thousands of species (Torsvik et al. 1990), and since each species may show strong variations in its trait spectrum (Li et al 2021), the obtained collection still represents only a small proportion of the community diversity. The development of Sanger sequencing (Sanger 1975) and subsequent next generation sequencing methods (NGS, e.g. 454, Illumina MiSeq) overcame restrictions of structural resolution by targeting multiple organisms at the same time. Currently, 16S rRNA gene MiSeq amplicon sequencing with relatively moderate costs and moderately long read lengths is the standard procedure to analyze prokaryote community composition in various

environments (Brandt et al. 2018, Kameoka et al 2021, Paulraj et al. 2021, Schöler et al. 2017). In this thesis, we used Illumina MiSeq sequencing to investigate structural adaptations to drought in the rhizobacterial community of wheat (Chapter 1 and 2). The high resolution of this approach revealed differences in the bacterial community compositions in different samples up to species level. Unfortunately, 16S amplicon sequencing does not differentiate between the contribution of either actively growing, passive (e.g. spores) or dead organisms within the studied community (Emerson et al. 2017). Comparing the cultivation-based phosphate-solubilizing community isolated from the rhizosphere of wheat, which was dominated by species of the phyla *Pseudomonas*, *Streptomyces* and *Phyllobacterium*, with 16S rDNA gene abundances in the overall community, we found poor correlations between the high numbers on plate and the relative abundancies in the amplicon datasets (Figure 4). Although the two methods are hardly comparable for the reasons mentioned above, this discrepancy raises the question which of the cultivation-based or NGS approach is more meaningful to reveal structural and functional adaptation in rhizobacterial communities.

To bridge the gap between structural and functional diversity analyses, recent computational approaches link structural data obtained from NGS with available functional data as strategy to derive a functional profile of bacterial communities. In this line, we evaluated the tools of Tax4Fun (Aßhauer 2015) and PanFP (Jun 2015) to test the concordances between functional gene abundances and measured enzyme activity values in different agricultural systems, climate treatments and at different plant growth stages. As concluded in Chapter 2 we could find qualitative, but not necessarily quantitative concordances. This quantitative discrepancy might reflect that the passive part of the community introduced a bias to predict the extent but not the direction of expressed activity levels. The bias by dormant sub-communities can be minimized by barcoding the active part of bacterial communities based on RT-PCR of RNA and compare it to DNA extracts, as demonstrated by Zhu et al. (2020). However, at the present state, a combination of different methods may be the best option to gain a deeper and more precise understanding of plant-microbe interactions under different environmental conditions.

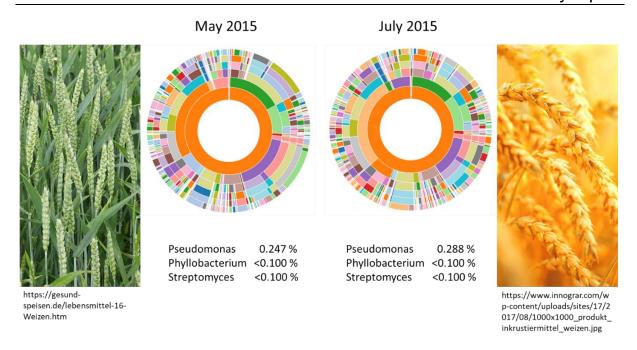


Figure 4 Rhizobacterial community composition based on 16S rDNA analysis at two growth phases in May 2015 and July 2015. From inner to outer circle the abundances and diversities of species at different taxonomy levels are presented, starting with the domain (inner circle) and ending with genus (outer circle). Percentage of total community for dominant P-solubilizing genera *Pseudomonas*, *Phyllobacterium* and *Streptomyces* are given.

Multifactorial approaches to gain a more holistic understanding of plant-microbe interactions in pot experiments

As pointed out in the introduction, different environmental parameters shape bacterial and rhizobacterial community composition, which was also confirmed in our studies (Chapter 1 to 3). In comparison to structural changes, ecosystem functions are assumed to be less affected and more resilient due to functional redundancy within soil microbial communities. Recently, Rossmann et al. (2020) described that as a consequence of domestication and breeding, highly productive cultivars form less intricate microbial co-occurrence networks compared to landraces and thus are more vulnerable to disturbances. On the other hand, de Vries et al. (2018) indicated that the recovery of such bacterial networks after drought stress is linked to soil functions creating a resilient and adapted community. Thereby, the question arises which mechanisms of microbial communities respond to changed conditions and are these common or unique for different environments. By simultaneously manipulating multiple experimental factors in field and pot experiments, we in fact revealed different, but also common adaptation

processes of bacterial communities, which can contribute to maintain productivity in agricultural systems under drought stress.

Of all the experimental factors studied in the pot experiment, one factor predominantly determined the response of the bacterial communities under drought: the soil type. As described in the first chapter, we decided for comparing two soil types that fundamentally differed in their physical and chemical properties (Altermann et al 2005, Schweitzer 2010). After a whole vegetation period, the fertile loamy soil showed tendencies for buffering unfavorable conditions of less water and nutrient supply independently of the plant genotype, and this, in case of extracellular enzyme activities, maintained at least a status quo compared to initial conditions without drought treatment. In contrast, unfavorable conditions in the sandy soil completely altered the composition and functioning of the rhizosphere communities by decreasing diversity and depleting available resources, respectively. These effects might simply reflect the differences in water holding capacities of the two soil types, since watering levels were applied according to maximum water holding capacities instead of soil water availability. However, comparing wheat performance in both soil types under drought and wellwatered conditions in the potting system, we observed maximum difference in water use for the loamy soil in early May (phase 3: stem elongation, Hack et al. 1992, data not published), while in the sandy soil the maximum was delayed until early June (phase 6: flowering, Hack et al. 1992, data not published). This suggested a much earlier effect of drought on wheat plants and rhizosphere communities at sensitive development stages in loamy than in sandy soil, and led to an overall stronger reduction in plant biomass production (Chapter 1). Thus, changes in rhizosphere communities and functions probably result from an interplay of soil type and plant presence, whereby above- and belowground responses to drought were found to be opposite in the two soil types.

A very specific interaction effect was observed in the sandy soil when the impacts of cultivar (demanding vs. non-demanding) and farming system (conventional vs. organic) related changes were considered as well. Although mechanisms could not be fully resolved, the organic manure application in organic farming treatments may have ameliorate water retention

capacities of the sandy soil (Rawls et al 2003) and possibly fostered increased enzyme activities under drought conditions in the rhizosphere of the demanding cultivar. These findings do not only emphasize the importance of carbon concentrations to increase drought tolerance. We could also conclude from Chapter 1 that the interacting effects of different factors offer a realistic and holistic picture of different adaptation processes than solely regarding single factors.

Transferability of findings gained in the pot experiment to field conditions

While pot experiments allow the simultaneous investigation of multiple factors under similar environmental conditions, the knowledge gained in a controlled environment does not necessarily apply under field conditions, where soil properties and plant growth rates are more heterogeneous, and uncontrolled changes in the environment affect the bacterial diversity and functions (Passioura et al. 2012, McKersie et al. 1999). In the Chapters 2 and 3, we used the platform of the Global Change Experimental Facility (GCEF) as the field scale experiment. The platform includes a model-based climate treatment (ambient vs. future) and realistic agricultural treatments (conventional vs. organic) with crop rotation (wheat-barley-rapeseed/clover) (Schädler et al 2019). Moreover, the fertile loamy soil (Haplic Chernozem, Altermann et al. 2005) used for the pot experiment originated from organic and conventional farming plots of the GCEF, that allowed to test the extent to which the results of the pot experiment could be transferred to field conditions.

Unlike the pot experiment, in which water supply was controlled to either 25 % or 60 % of the soil's WHC_{max}, representing drought or well-watered conditions respectively, the climate treatment in the GCEF controls for precipitation, reducing the amount of natural precipitation by 20 % in the summer to simulate future climate (Schädler et al. 2019). Gravimetric soil moistures thus ranged between 6 % to 22 %, which equals 17 % to 63 % of the soil's WHC_{max}, even on future climate plots. Modeling long-term rather than immediate changes in the GCEF, bacterial communities will adjust over years as a result of accumulative water reduction,

whereby individual groups or species may respond more rapidly due to specification on certain ecosystem functions (reviewed in Fierer et al. 2017). Accordingly, climate treatment did not affect community composition or function in our study two years after GCEF establishment (Chapter 2), but *Streptomyces* species, however, expressed higher drought tolerance when isolated from future climate plots in the same year (Chapter 3). Creating extreme conditions of prolonged drought, as in the pot experiment, will probably led to a strong community response within one growing season. Hence, the intended transferability of findings from the pot to the field would fail if only the results of one year without extremes is considered. Given the severe heat waves in 2018 and 2019 in Central Germany (Hari et al. 2020) with hardly any precipitation from spring to fall, drought conditions in the field resembled those in the pot experiment and likely had respective consequences for community composition and function. Even though it was not part of this study, it may be of future interest, whether GCEF's climate manipulations over the years have a more additive or neutral effect on such extreme events, and where the limits of adaptation lie.

Like in the pot experiment, we observed farming management-related effects on rhizobacterial community composition and function in the field (Chapter 1 and 2). Since pre-adapted soil from the GCEF farming plots, as well as fertilization according to the GCEF management were applied in the pot experiment, the implementation of more realistic scenarios in pot experiments might be the key for such, at least directional, concordances. With respect to community composition, direct comparisons between the pot and field experiments on species level may be not applicable due to the use of different pipelines for post-processing of sequencing data ("dadasnake" by Weißbecker et al. 2020 in Chapter 1 and "DeltaMP" by Lentendu (https://github.com/lentendu/DeltaMP) in Chapter 2). At the phylum level, we found a general dominance of *Acidobacteria*, *Actinobacteria*, *Bacteroidetes*, *Chloroflexi*, *Proteobacteria* and *Verrucomicrobia* for both experiments, with abundances differing between the pot and field experiments, but not necessarily between the farming treatments. The extracellular enzyme activity potentials increased in the conventional farming system, with readily available nutrients, compared to the organic farming system, irrespective of water

treatment (results for fertile loamy Haplic Chernozem in Chapters 1 and 2). Thereby, nutrient availability might have influenced bacterial extracellular enzyme production directly, or indirectly. The indirect effect is directed by different mechanisms of plant-soil-interactions, i.e. differences in root exudation patterns (reviewed in Lambers et al. 2009), and is particularly strong in the rhizosphere due to the close spatial proximity. Accordingly, in the pot experiment, community composition differed between compartments of rhizosphere and bulk soil, whereas cultivar-specific effects on composition were missing (Chapter 1).

Former studies either support (Donn et al. 2015) or neglect (Simonin et al. 2020) the effect of wheat cultivar on surrounding conditions, but indicate a common effect of plant growth stage. At the early growth stages, the investment in plant tissue requires high amount of nutrients particularly fostering the exchange with soil communities, thereby shaping composition and function by differential production rate and spectrum of rhizodeposits. At the mature growth phases the plant mainly allocates compounds from the vegetative to generative parts, this reduces interactions on the root-soil-surface (Malhi et al. 2006, Francioli et al. 2018). Although the data were not presented in Chapter 1, sampling in the pot experiment at three different time points (March: early vegetative phase -April: intermediate vegetative phase - July: mature phase) also indicated a strong effect of the plant development stage on rhizobacterial community composition and function, which was confirmed in the field for the overall community (Chapter 2), and also at individual species level at two different growth stages (vegetative phase and mature phase) (Chapter 3). Changes in community composition in the pot and field experiments were accompanied by reduction in enzyme activities from vegetative to mature growth phases (Chapter 1 and 2). At the individual species level, community composition of dominant groups rather than activity potentials were responsible for differences between vegetative and mature growth phases (Chapter 3).

Towards a wheat core microbiome?

At the beginning of this thesis, we introduced the idea by Kavamura et al. (2021) of using multifactorial and multidisciplinary approaches to identify a possible wheat core microbiome or

key species, which may persist under different environmental conditions and promote plant growth. As such, it needs to be first clarified, how the wheat core microbiome is defined, either as the starting microbiome in the seeds or of adult plants. Second, it must be answered whether the taxa of the core microbiome play an active role or are just present in the community. Coming to the end of this thesis, some answers to these two points can be provided.

Analysis of seed microbial communities were not part of this study, but recent findings indicated the existence of a core seed microbiome. Investigating 50 different plant species, Simonin et al. (2021) found 30 bacterial and fungal taxa present in seeds of most plant species from all over the world. In our study, the rhizospheres of adult plants were investigated, and rhizobacterial communities had already experienced different biotic and abiotic impacts shaping their composition. As discussed before, abundances of phyla might differ, while especially dominant groups are present in each treatment. However, going into deeper taxonomic levels, indicator species analysis revealed that no OTUs were shared between the rhizospheres under treatments of different farming systems (conventional vs. organic) and plant growth phases (vegetative vs. mature growth phase) (Chapter 2). This does not necessarily imply, that no species at all are shared between the different rhizospheres, since the amplicon sequencing data are compositional, i.e. transformed to relative abundancies instead of absolute numbers, and rare sequences are often removed from datasets (Fernandes et al. 2014). As such, singletons were removed during the bioinformatics workflow in Chapter 2, which corresponded to 2.5 % of the total reads. It is also possible that the presence of a certain taxon might not be system-relevant, which would suggest a decoupling of high structural variability from stable functional structures (Louca et al. 2016). For instance, Bastida et al. (2021) identified the most common microbial proteins involved in different edaphic processes across biomes and concluded a higher resolution of environmental impacts using protein-based data compared to 16S rRNA gene sequencing. Most recently, Fernandes et al. (2022) identified a core functional soil mycobiome across ecosystems by metaproteomics analysis. These findings are not that surprising given the concept of functional redundancy (Banerjee et al. 2016). However, metaproteome analysis offers new perspectives in biofertilizer creation, as such that besides living inoculants, factor-relevant bacterial metabolites may be identified and genomic information transferred to or deleted in organisms. Something comparable has already been tested for *Pseudomonas syringae*. To enhance frost tolerance of strawberries and potatoes, the ice nucleation protein has been removed from the bacterial genome (Skirvin et al. 2000). There also exist ideas for plant beneficial microbe breeding, which aim to transmit the endophytic microbiota of one plant to the next generation (Wei & Jousset 2017) and thus may be of particular interest under changing climatic conditions. In this study, functional trait approaches at individual species level revealed *Phyllobacterium* species as promising P-solubilizers under future climate conditions in the rhizosphere of wheat, which has not been described before and may be useful to create new biofertilizer (Chapter 3).

Study limitations and outlook

The aim of this study was to investigate different adaptation processes of wheat rhizobacterial communities to drought using multifactorial and multidisciplinary approaches in pot and field experiments. While the broad spectrum of factors included in this study fulfilled the intention of a multifactorial approach, the choice of analyzing methods in the three chapters was mainly driven by considerations of indirect impacts on nutrient cycling as response to drought. Certainly, direct indicators of stress in the plants, such as phytohormone regulation, oxidative enzyme production and differential gene expression levels (reviewed in Chandra et al. 2021), might provide important mechanistic insights into microbial activation and regulation of stress responses; and how they may differ under changing surrounding conditions. For instance, Gebauer et al. (submitted to Frontiers in Microbiology) studied the community composition of aminocyclopropane-1-carboxylate (ACC) deaminase carrying bacterial communities in the rhizosphere of wheat, which regulate concentrations of stress phytohormone ethylene, and found a strong farming related effect.

As already coined, for the analysis of sequencing data, we used different pipelines in the pot and field experiments, which did not allow direct comparisons of communities and, thus, the

Synopsis

identification of possible key players. Therefore, it would be interesting to re-analyze these datasets, also in the light of the novel bioinformatics approach determining. Amplicon Sequence variants (ASVs), which offers a more precise identification of microbes and, thus of their diversity within samples, compared to conventional use of operational taxonomic units (OTUs) (Callahan et al. 2017). Despite this limitation, the use of next generation amplicon sequencing provided a high structural resolution of the rhizobacterial communities in dependency of the different factors used in this thesis. Accompanied changes in community function could not be directly linked to structural changes with the methods used in this study, but the assignment of functional capabilities of individual taxa is a prerequisite for identification of key species and a possible core microbiome (Kavamura et al. 2021). The integration of metatranscriptomics and proteomics data should overcome these restrictions, and the fast development in these techniques may introduce these methods as standard procedures in the future. As such, metatranscriptomics may be also relevant to improve existing cultivation standards in soil ecology, as it has already been described for gut microbiomes (Bomar et al. 2011).

One of the greatest potentials, which has not been explored at the time of writing this thesis, is the use of the pre-adapted communities and single species for feedback analysis in inoculation experiments under controlled conditions, but also in the field. Priming with beneficial bacteria can be performed as reinoculation experiments using pre-adapted soils (Wubs et al. 2016), as well as creation of synthetic communities (Tsolakidou et al. 2019) or single species inoculation experiments (Wehner et al. 2019). Overall, all these studies on priming indicated a significant improvement in plant performance under stress conditions, and might thus be the key for sustainable agriculture.

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- Siles, J. A., Díaz-López, M., Vera, A., Eisenhauer, N., Guerra, C. A., Smith, L. C., ... Breitkreuz,
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STATUTORY DECLARATION

Hereby, I, Claudia Breitkreuz, affirm that I take note and accept the doctorate regulations of the Faculty of Life Science at the University of Leipzig from September, 30th of 2019. I further affirm that the presented thesis was prepared autonomously without inadmissible help. All aids used in this thesis as well as scientific ideas which are quoted from or based on other sources were cited at the respective point.

All people who helped me to prepare the conception, to select and analyze the materials of this thesis as well as to improve the manuscripts are namely cited in the acknowledgments. With exception of the namely mentioned people no other persons were involved in the intellectual work. No PhD consultant service was employed. Third parties did not get money's worth for benefits that were conjunction with the content of this dissertation.

I declare that this dissertation has been neither presented nationally nor internationally in its entirety or in parts to any institution for the purpose of dissertation or other official or scientific examination and/or publishing.

Previously unsuccessful dissertations had no taken place.

The original document of the verification of the co-author parts are deposited in the office of the dean.

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EIDESSTATTLICHE ERKLÄRUNG

Hiermit erkläre ich, Claudia Breitkreuz, eidesstattlich, dass mir die Promotionsordnung der Fakultät für Lebenswissenschaften der Universität Leipzig vom 30.09.2019 bekannt ist und von mir anerkannt wird.

Zudem versichere ich, dass die vorliegende Dissertation von mir selbstständig und ohne unzulässige Hilfsmittel angefertigt worden ist. Sämtliche von mir verwendetet Hilfsmittel, sowie die aus fremden Quellen direkt oder indirekt übernommenen wissenschaftlichen Gedanken, sind als solche in der vorliegenden Arbeit gekennzeichnet. Alle Personen, von denen ich bei der Auswahl und Auswertung des Materials, sowie bei der Herstellung der Manuskripte Unterstützungsleistungen erhalten habe, sind an entsprechender Stelle namentlich in der/den Danksagungen genannt. Außer den genannten Personen waren keine weiteren Personen an der geistigen Herstellung der vorliegenden Arbeit beteiligt. Die Hilfe eines Promotionsberaters wurde nicht in Anspruch genommen. Dritte haben von mir für Arbeiten, die im Zusammenhang mit dem Inhalt der vorliegenden Dissertation weder im Inland noch im Ausland in gleicher oder ähnlicher Form einer anderen Prüfungsbehörde zum Zwecke der Promotion oder andere staatlicher oder wissenschaftlicher Prüfungsverfahren vorgelegt und/oder veröffentlicht wurde.

Frühere erfolglose Promotionsversuche haben nicht stattgefunden.

Der Nachweis über die Anteile der (Co-)Autorenschaften sind im Original im Dekanat der Universität Leipzig hinterlegt.

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