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# Toxicological effects of cadmium on deep-sea mussel *Gigantidas platifrons* revealed by a combined proteomic and metabolomic approach

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**Introduction:** Marine metal contamination caused by deep-sea mining activities has elicited great concern from both social and scientific communities. Among the various metals deep-sea organisms might encounter, cadmium (Cd) is a widely detected metal that in very small amounts is nonetheless capable of severe toxicity. Yet due to both remoteness and technical challenges, insights into the effects of metal exposure resulting from mining activities upon deep-sea organisms are limited.

**Methods:** Here, we investigated Cd's toxicological effects on deep-sea mussels of *Gigantidas platifrons* exposed to 100 or 1000 g/L of Cd for 7 days; an integrated approach was used that incorporated proteomics and metabolomics along with traditional approaches (metal concentrations, metal subcellular distribution, and anti-oxidative and immune-related biochemical indexes).

**Results and Discussion:** Results showed that Cd exposure caused significant Cd's accumulation in mussel gills and redistribution of Cd among subcellular compartments, with cellular debris being the primary binding site. Although antioxidative enzymes activities (superoxide dismutase and catalase) were not significantly altered in mussel gills of both exposed groups, the markedly increased level of glutathione S-transferase detected via proteomic technique clearly evinced that deep-sea mussels suffered from oxidative stress under Cd exposure. Besides, altered activities of acid phosphatase and alkaline phosphatase assayed by traditional methods along with the predominant presence of largely altered immune-related proteins detected by Cd. In addition, results of proteomics combined with those of non-targeted metabolomics demonstrated that Cd could exert toxicity by disrupting cytoskeleton structure, ion homeostasis, and primary metabolisms of energy, lipid, and nucleotide in deep-sea mussels. As

demonstrated in this study, proteomics and metabolomics can be used in tandem to provide valuable insights into the molecular mechanisms of deep-sea organisms' response to Cd exposure and for helping to discover potential biomarkers for application during deep-sea mining assessments.

KEYWORDS

metal, toxicology, Mytilidae, deep-sea, omics, environment monitor

# **1** Introduction

The deep-sea floor harbors a wealthy of valuable mineral resources which are necessary components of a variety of lowcarbon energy technologies, such as solar panels, wind turbines, and geothermal capture systems (Carlsson and Gade, 1986). As mineral resources on land are gradually depleted and deep-sea equipment and technology are further developed, deep-sea mineral resources have emerged as a new target for extraction and acquisition by humans, with deep-sea mining now entering a highly active period. The contract numbers for exploring deep-sea mineral resources has soared, from just seven in 2010 to thirty-one in 2021, of which seven mining contracts are for the exploration of hydrothermal areas (ISA).

However, deep-sea mining is a large-scale human activity. In the process of transporting metal minerals from hundreds or even thousands of meters of seabed up to the water surface, a mining operation will inevitably result in the release of toxic substances including copper, cadmium, zinc, lead, and iron. When excessive amounts of these metals are absorbed by local organisms, it can cause acute or chronic damage to deep-sea species, inhibiting their growth and reproduction, affecting the local food chain, and eventually leading to ecosystem shutdown (Boschen et al., 2013; Hauton et al., 2017; Martins et al., 2017). Hindered by both its remoteness and technical challenges, the deep sea is the least explored biome on the planet. As a result, there is very little knowledge of how deep-sea mining impacts deep-sea organisms in terms of their physiology and population dynamics. To establish guidelines and frameworks for sustainable mining practices as soon as possible, it is imperative that metal toxicological research of deep-sea organisms be conducted, which provide a timely empirical basis that can also inform theory for the establishment of sensitive biomarkers for the monitoring of deepsea mining activities.

The deep-sea bivalve *Gigantidas platifrons* is a cosmopolitan mussel that can be found in most deep-sea chemosynthetic ecosystems of the West Pacific Ocean. As stationary filter feeders, individuals of this species live in the benthic environment, where pollutants are deposited and accumulated (Zhou et al., 2020; Zhou et al., 2021). These mussels are very sensitive to environmental fluctuations and can serve as sentinels to monitor the deep-sea environment (Zhou et al., 2020; Zhou et al., 2021). Among the various metals deep-sea life may come into contact with, Cd is a nonessential element in deep-sea mussels' physiology, which can be readily absorbed, transported, and accumulated in their bodies (Company et al., 2010; Zhou et al., 2021). Furthermore, Cd does

not biodegrade and can persist in organisms for a long time, causing serious toxic effects *via* trace amounts on organisms at their physiological, cellular, subcellular and molecular levels. Experimental studies with deep-sea organisms under metal exposure have shown that Cd could cause general cellular injury, impair lysosomal and plasma membrane integrity, induce oxidative stress and overproduction of reactive oxygen species (ROS), disrupt cellular osmoregulation as well as disordering core metabolic pathways of amino acids, carbohydrates, and lipids (Company et al., 2019; Zhou et al., 2021). In deep-sea hydrothermal ecosystems, Cd levels tend to be high, capable of reaching 1.5  $\mu$ M and 3950 ppm in hydrothermal fluids and metallic mineral deposits, respectively (Zeng, 2011).

Previously, researchers have investigated the toxic effects of deepsea mining upon deep-sea fauna by using shallow-water taxa as proxies for deep-sea taxa (Brown et al., 2017b; Mestre et al., 2019; Pinheiro et al., 2021). However, such studies are controversial because they ignore the complexity of the deep-sea environment and species differences between organisms in shallow and deep water habitats (Brown et al., 2017a). A recent study of deep-sea organisms exposed to metals found that the expression of some traditional metal-related biochemical indexes (e.g., metallothioneins) or metabolites (e.g., Dallose) were not significantly induced or exhibited a pattern opposite to that found in shallow-water (Zhou et al., 2021), implying that deepsea organisms might undergo different biochemical and molecular responses when exposed to metals. Besides, when compared with the vast amount of toxicological research in costal organisms, corresponding research on deep-sea organisms far from comprehensive and limited to the responses of a known class of toxicity-responsive genes or particular enzyme activity (Company et al., 2010; Bougerol et al., 2015; Martins et al., 2017). The response mechanisms of deep-sea organisms induced by stress from different metals are not exactly the same, and they depend on a complex genomic regulatory network involving multiple genes and pathways (del Rio, 2016; Wong et al., 2015). Therefore, systematical toxicological work carried out using multiple approaches at once, to unravel these biological effects and responsive mechanisms, can truly reflect the ongoing and expected life processes in those parts of the deep sea incurring mining perturbations.

Recently, with dramatic advances in high-throughput molecular tools, 'omic' technologies including proteomics and metabolomics have been extensively used in toxicological studies of marine environments (Sanchez et al., 2011; Xu et al., 2019). Proteomics theoretically characterizes the entire protein pool at the organelle, cellular or tissue level, and is capable of detecting significantly different expressed proteins (DEPs) caused by exogenous factors in organisms (Tomanek, 2011; Campos et al., 2012). Meanwhile, metabolomics systematically examines the entire range of endogenous low-molecular weight metabolites (< 1000 Da) in biological samples (Ji et al., 2015). It offers the opportunity to compare the metabolites and then detect alterations in their expression profiles in organisms. When used in tandem, proteomics and metabolomics could provide plenty of information concerning the molecular responses of living organisms to contamination. In combination, these two 'omic' approaches would ideally complement each other, offering opportunities to decode novel molecular mechanisms involved in stress response, and being useful for discovering reliable biomarkers of metal exposure and its effects. To our best knowledge, however, no attempt has yet been made to study the toxic effects of metals on deep-sea organisms simultaneously at the protein and metabolite levels.

In this study, to gain a comprehensive understanding of the underlying toxicological effects associated with metal exposure in deep-sea organisms, the deep-sea mussel G. platifrons was selected as the model organism and exposed to cadmium (100 and 1000  $\mu$ g/L) for 7 days, and the response of the treated deep-sea mussels then evaluated using both traditional toxicological methods (i.e., metal accumulation, metal subcellular distribution, biochemical indexes) and 'omic' techniques (proteomics and metabolomics). The objectives of this study were (i) to determine the subcellular distribution and biochemical response of G. platifrons gill tissue under Cd stress; (ii) to identify key proteins/metabolites that are differentially expressed in G. platifrons under Cd stress, and their involved essential pathways; (iii) to uncover the molecular mechanisms by which G. platifrons responds to Cd. This study will contribute to bridging a glaring knowledge gap regarding the potential impacts of deep-sea mining on deep-sea organisms, as well as identifying some important responsive candidate biomarkers associated Cd stress for the monitoring of deep-sea mining.

# 2 Material and methods

### 2.1 Experimental design

Mussels of *G. platifrons* were collected from the seabed of the South China Sea at seawater depths below 1119 m, by using the remotely operated vehicle *Faxian* (ROV Faxian), it mounted to the research vessel *Kexue*. ROV *Faxian* brought the mussels to the sea surface after first placing them in an insulated container. Once aboard, mussels were immediately transferred to a specially designed aquarium system, used for culturing deep-sea organisms. In this system, mussels were cultured in surrounding seawater at 4°C with a pressure of 1 atm. Water was half-renewed daily and any dead mussels were immediately discarded. Mussels received methane gas twice a day for 30 min, as a food and energy source.

Following acclimatization period of 48 hours, thirty mussels (93.2  $\pm$  5.65 mm in shell length) were randomly placed in the three highdensity polypropylene tanks (10 mussels per tank; rearing density: one mussel per liter of seawater). Mussels were exposed to nominal 100 µg/L or 1000 µg/L of Cd for 7 days, with clean filtered seawater serving as a control. The exposure concentrations of Cd chosen in this study reflect those in deep-sea environments (Zeng, 2011). Cadmium was prepared with CdCl<sub>2</sub>•2.5H<sub>2</sub>O (Sigma-Aldrich, analytical grade). The exposed concentrations were environmentally relevant and reported for hydrothermal systems (Zeng, 2011). The filtered seawater and exposed metal solution were renewed thoroughly on a daily basis. During this exposure experiment, other conditions were consistent with those of the acclimatization period, and no mussel mortality was observed.

All of thirty mussels were sacrificed after 7 days of exposure. Immediately after dissecting each with plastic tools, the mussel gills were flash frozen in liquid nitrogen, and then stored at  $-80^{\circ}$ C for further analysis. Mussel gill of each individual was separated into five sets for their metal contents' determination, metal subcellular distribution, biochemical analysis, and metabolic and proteomic assessments.

# 2.2 Metal content and metal subcellular distribution

To quantify the gills' Cd accumulation, inductively coupled plasma mass spectrometry was used as previously described (ICP-MS, Thermo-Scientific iCAPQc, Bremen, Germany) (Zhou et al., 2020). A detailed description of these procedures can be found in the Supplementary Material (S1.1).

Metal subcellular distribution in mussel gills was determined according to a well-established protocol reported in several studies (Wallace and Luoma, 2003; Pan and Wang, 2008; Yu et al., 2013; Ma and Wang, 2020). Briefly, each gill tissue sample was individually homogenized in a freshly prepared 30 mM Tris-NaCl buffer (pH 8.0, 0.15 M NaCl, 5 mM 2-mercaptoehanol, with 0.1 mM of freshly prepared phenylmethylsufonyl fluoride) using an Ultra-Turrax homogenizer (Jing Xin, China) for 30 sec. The homogenized tissue was first fractionated by centrifugation at  $1450 \times g$  for 15 min. The collected pellets were digested in 1 N NaOH. After the mixture had been heated at 80°C for 10 min in a water bath, a second centrifugation at 5000× g for 10 min was carried out. Next, the metal-rich granule fraction (MRG) was obtained from the pellets, while cellular debris was obtained from the supernatant. A third centrifugation at 100  $000 \times g$  for 60 min was performed on the supernatant from the first centrifugation. From this, the resulting pellet resulting was treated as the organelle fraction, and a 10-min water bath at 80°C was used to heat the ensuing supernatant. After doing another centrifugation at 50 000× g for 30 min, the heat-sensitive protein fraction (HSPF) was obtained from the supernatant, while metallothionein-like protein fraction (MTLP) was derived from the pellet. All procedures were conducted at 4°C unless otherwise stated. The experiment was conducted with blank samples lacking any tissue addition. Following the freeze-drying of each fraction, the fractions were digested with 2 mL of 15.6 M HNO<sub>3</sub> at 80°C for 48 h. The obtained digested mixture was then diluted to 5 mL with deionized water and the Cd concentration in each fraction likewise measured according to the above procedure. The percentages of Cd metal in each fraction were defined as the metal subcellular distributions.

### 2.3 Biochemical measurements

Six biochemical indexes—acid phosphatase (ACP), alkaline phosphatase (AKP), superoxide dismutase (SOD), catalas (CAT), reduced glutathione (GSH) and lipid peroxidation (LPO)—were measured in the mussel gill samples. All these biochemical indexes were quantified using commercial kits from the Nanjing Jiancheng Bioengineering Institute (China). AKP and ACP activity were determined according to the method King (1965). SOD was measured by the method of McCord and Fridovich (1969). CAT activity was quantified by the method of Goth (1991). Reduced GSH was determined following the method of Jollow et al. (1974). LPO was assessed by the method of Korobeinikova (1989). A detailed description of these procedures can be found in the Supplementary Material (S1.2).

### 2.4 Proteomic analysis

Three mussel gill tissues were pooled into one biological replicate, and three biological replicates were analyzed for each group. This analysis using 4D Label free proteomics was conducted in collaboration with Jingjie PTM BioLabs (Hangzhou, China) (Ding et al., 2022). The main procedures included protein extraction, protein alkylation, trypsin digestion, LC-MS/MS analysis, and the bioinformatics analysis. Briefly, a lysis buffer containing 8 M urea and 1% protease inhibitor cocktail was used to extract the mussel gill proteins, which were then digested with trypsin. The tryptic peptide was dissolved in mobile phase A of liquid chromatography, and separated by a NanoElute ultrahigh-performance liquid chromatography system (Bruker, Germany). Protein profiles were generated by timsTOF Pro (Bruker Daltonics) mass spectrometry after separation with the ultra-performance liquid system. The resulting MS/MS data was analyzed using the MaxQuant search engine (v1.6.15.0). Tandem mass spectra were searched against deep-sea mussel G. platifrons genome data concatenated with the reverse decoy database. The reference genome of G. platifrons employed in this study was originally published by Sun et al. (2017) and updated by Wang et al. (GenBank accession no. JAOEFJ000000000, unpublished data). The FDR (false discovery rate) was adjusted to < 1%.

Proteins with a fold-change > 1.2 or < 0.83 *and* a P < 0.05 were designated as being significantly differentially expressed. These proteins were classified and grouped using Gene Ontology (GO) and KEGG databases. A detailed description of these procedures can be found in the Supplementary Material (S1.3).

# 2.5 Non-target metabolomic profiling analysis

Metabolites in mussel gills were extracted as previously described (Zhou et al., 2021). After that, an ultrahigh performance liquid chromatography instrument (UHPLC) (1290 Infinity LC, Agilent Technologies, Santa Clara, CA, USA), equipped with a quadruple time-of-flight mass spectrometer (AB Sciex TripleTOF 6600), was used to generate the untargeted metabolic profiles. This analysis was supported by Shanghai Applied Protein Technology Co., Ltd, APTBIO. The procedures involved are described in detail in the Supplementary Material (S1.4).

### 2.6 Data analysis

The statistical analysis was implemented in SPSS v19 software for Windows (IBM, Armonk, NY, USA). For biochemical indices and metal concentrations, their means and standard deviations were calculated. Before their formal statistical analysis, data of fitted residuals were checked for homogeneity of variance and normality. Analyses of variance (ANOVA) and Kruskal–Wallis tests were applied to parametric and nonparametric variables, respectively. When a significant difference among the three groups was detected, or Tukey's range test or Mann–Whitney U tests were respectively applied in pairwise manner to identify which treatment or control groups differed from each other.

# **3** Results

# 3.1 Metal accumulation and subcellular metal distribution in deep-sea mussel gills under Cd exposure

Accumulated Cd concentrations and the corresponding subcellular Cd distribution in mussel gills are depicted in Figure 1. Mussel gills can significantly accumulate Cd from the surrounding environment, with the higher Cd exposure concentration resulting in a higher Cd concentration in mussel gills (Figure 1A). Cd concentrations within the subcellular fractions (i.e., MTLP, organelle, HSPF, MRG, cellular debris) of the mussel gills were also significantly impacted by exposure concentration (Figure 1B). In the control group, HSPF harbored the most Cd. As the exposure concentration increased, Cd was transferred to MTLP, organelle, and cellular debris, the last being where Cd was predominantly stored in deep-sea mussels.

# 3.2 Biochemical responses of deep-sea mussel gills under Cd exposure

Most biochemical parameters analyzed in this study were affected by the exposure concentrations of Cd (Figure 2). The activity of ACP decreased as the concentration of Cd increased. Conversely, AKP activity and the levels of GSH and LPO all increased with an increased Cd concentration. The activity of SOD or CAT did not differ significantly among the treatments and control group. The ACP activity and LPO level were statistically similar in both low- and high-dose group. The GSH level only raised in mussel gills of the high-dose group, being approximately three times higher that of the control group.

# 3.3 Proteomic profiles of deep-sea mussel gills under Cd exposure

A total of 5260 quantifiable proteins were identified in the deepsea mussel gills (Supplementary Materials S2.1). Of those, in the low dose-Cd exposure group, 162 were significantly differentially expressed proteins (DEPs) consisting of 87 up-regulated (53.7%)



protein; MRG: metal-rich granules; HSPF: heat-sensitive protein.

and 75 down-regulated (46.3%) induced DEPs; in the high dose-Cd exposure group 222 DEPs were found induced, these consisting of 144 up-regulated (64.9%) and 78 down-regulated (35.1%) DEPs (Figure 3A). Venn diagram analysis showed that 47 DEPs were altered in common (30 down-regulated and 17 up-regulated) by the two Cd-exposure treatments (Figure 3B). Detailed information on all significantly regulated proteins is provided in Table 1 and Supplementary Materials Table S1.

The DEPs were classified into three categories based on their GO annotations: biological process (BP), molecular function (MF), and cellular component (CC). Under the BP category, most DEPs were involved in cellular process, biological regulation, and metabolic process in both treatment groups. In the MF category, most DEPs detected were involved in catalytic activity and binding in both

treatment groups. In the CC category, most DEPs were located in cell and intracellular, in both treatment groups (Figure 3C). A detailed description of the GO enrichment results for the DEPs is provided in Table S2 of the Supplementary Materials.

The most representative KEGG pathways for the down- and upregulated DEPs are shown in Figure 4. In the low-dose Cd exposed group, O-glycan biosynthesis, apoptosis, the wnt signaling pathway, and protein processing in the endoplasmic reticulum were the most statistically significant pathways for the up-regulated DEPs. For the down-regulated DEPs, the most significant pathway was galactose metabolism. In the high dose-Cd exposure group, both mismatch repair and spliceosome were significantly enriched for the up-regulated DEPs; vitamin digestion and absorption, linoleic acid metabolism, glycolysis/gluconeogenesis and necroptosis pathway were enriched for the down-regulated DEPs.



FIGURE 2

The activity of acid phosphatase (ACP), alkaline phosphatase (AKP), superoxide dismutase (SOD) and catalase (CAT), and the levels of glutathione (GSH), and lipid peroxidation (LPO) in the gills of *Gigantidas platifrons* after 7 days of exposure to 100 or 1000  $\mu$ g/L Cd (n = 10 per group, mean  $\pm$  SD). Values followed by the same letter are not statistically different (*P* > 0.05) between the treatment groups.



TABLE 1 T	The key o	differentially	expressed	proteins	(DEPs)	responsive	to cadmi	ium in	deep-sea	mussel	gills.
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Protein accession	Accession no.	Protein description		Fold change	
			100 µg/L	1000 µg/L	
Detoxification					
yibei_GLEAN_10021621	XP_022337663.1	microsomal glutathione S-transferase 3-like	3.1235		
yibei_GLEAN_10017956	XP_025111853.1	microsomal glutathione S-transferase 1-like isoform X2		1.8265	
yibei_GLEAN_10021632	QBM06404.1	microsomal glutathione S-transferase 3 isoform b		1.6278	
yibei_GLEAN_10005609	XP_011428128.1	cytochrome P450 2C31		0.7681	
Stress response					
yibei_GLEAN_10004579	XP_021349090.1	heat shock 70 kDa protein 12A-like	2.5479	3.909	
yibei_GLEAN_10020804	XP_011432755.1	major vault protein	1.834		
yibei_GLEAN_10012195	XP_021376559.1	heat shock 70 kDa protein 12A-like	0.6002		
yibei_GLEAN_10023865	BAD99026.1	heat shock protein 70		2.2481	
yibei_GLEAN_10020983	CAH04106.1	heat shock protein 70		1.4894	
yibei_GLEAN_10011520	XP_021376559.1	heat shock 70 kDa protein 12A-like		1.2724	
yibei_GLEAN_10005694	XP_005097388.1	10 kDa heat shock protein, mitochondrial-like		0.7902	
yibei_GLEAN_10008543	XP_021349090.1	heat shock 70 kDa protein 12A-like		0.5919	
		1		(Continued)	

### TABLE 1 Continued

Protein accession	Accession no.	Protein description	Fold c	hange
yibei_GLEAN_10005890	ANN45953.1	byssal HSP-like protein 1		1.2349
yibei_GLEAN_10029721	XP_021352707.1	glutamate receptor-like		1.5127
yibei_GLEAN_10025891	XP_021355385.1	putative glutamate synthase [NADPH]		1.3247
yibei_GLEAN_10030644	BAF63637.1	glucose-regulated protein 94		1.2191
Cytoskeleton	1			
yibei_GLEAN_10033077	XP_021377767.1	unconventional myosin-XVI-like isoform X10	0.6599	
yibei_GLEAN_10025956	XP_022295584.1	IST1 homolog	0.8093	
yibei_GLEAN_10020375	XP_022323498.1	kinesin-like protein KIF28P isoform X1	1.4335	
yibei_GLEAN_10015354	XP_019919334.1	dynein heavy chain 6, axonemal-like isoform X3	1.2369	
yibei_GLEAN_10026715	XP_011446273.1	dynein intermediate chain 2, ciliary isoform X2	1.2518	
yibei_GLEAN_10021778	XP_022339665.1	leupaxin-like isoform X2	0.8168	
yibei_GLEAN_10002721	XP_022306398.1	ribosome biogenesis protein WDR12 homolog		1.5925
yibei_GLEAN_10009294	XP_022340059.1	tektin-3-like isoform X1		1.8697
yibei_GLEAN_10012024	XP_025114800.1	tubulin beta chain-like		3.3209
yibei_GLEAN_10012532	NP_001292292.1	tubulin beta chain-like		1.7281
yibei_GLEAN_10018803		tektin-3-like isoform X2		1.7148
yibei_GLEAN_10020493	XP_021370513.1	dynein heavy chain 10, axonemal-like isoform X2		1.3877
yibei_GLEAN_10022711	XP_011416301.1	tektin-1		1.4629
yibei_GLEAN_10025426	XP_022293672.1	dynein heavy chain 5, axonemal-like isoform X2		1.6873
yibei_GLEAN_10032206	XP_022317078.1	WD repeat-containing protein 63-like isoform X3		1.5104
yibei_GLEAN_10027407	XP_011436383.1	tubulin-specific chaperone A		0.8061
yibei_GLEAN_10034641	XP_022338092.1	translationally-controlled tumor protein homolog		1.2166
yibei_GLEAN_10019514	XP_021340611.1	zinc finger ZZ-type and EF-hand domain-containing protein 1-like isoform X2		1.6331
Immune response				
yibei_GLEAN_10034298	XP_018092523.1	baculoviral IAP repeat-containing protein 7-B isoform X3	1.2636	1.2075
yibei_GLEAN_10035199	AJQ21543.1	stimulator of interferon protein 2	1.3678	1.2846
yibei_GLEAN_10011188	XP_021353098.1	AP-2 complex subunit alpha-2-like	1.2202	1.216
yibei_GLEAN_10015328	XP_021359219.1	interleukin enhancer-binding factor 2 homolog	1.5721	1.5939
yibei_GLEAN_10031808	XP_021373459.1	superkiller viralicidic activity 2-like 2	1.2801	1.5276
yibei_GLEAN_10027593	XP_021353442.1	E3 ubiquitin-protein ligase TRIM56-like	0.6072	0.5192
yibei_GLEAN_10017550	XP_022338572.1	nuclear transport factor 2-like	0.7733	0.6733
yibei_GLEAN_10025664	AKQ70855.1	shell fibrous prismatic cathepsin-like protein 1	0.214	0.1018
yibei_GLEAN_10018923	AJQ21549.1	c-Jun N-terminal kinase	1.2152	
yibei_GLEAN_10009572	XP_012866152.1	deoxynucleoside triphosphate triphosphohydrolase SAMHD1	4.0582	
yibei_GLEAN_10017095	AKS48146.1	protease inhibitor-like protein-B1	0.5625	
yibei_GLEAN_10013055	XP_021342211.1	tumor suppressor candidate 3-like	1.2308	
yibei_GLEAN_10019488	OPL21754.1	forkhead k2-like box protein	1.3562	
yibei_GLEAN_10029659	XP_021342748.1	alpha-L-fucosidase-like isoform X4		1.3694
yibei_GLEAN_10027085	XP_021378597.1	poly [ADP-ribose] polymerase 14-like isoform X1		1.3365
yibei_GLEAN_10018578	XP_025114852.1	probable ATP-dependent RNA helicase DDX17		1.2475
				(Continued)

#### TABLE 1 Continued

Protein accession	Accession no.	Protein description	Fold cha	ange
yibei_GLEAN_10034369	AJQ21489.1	galectin 1		0.7879
yibei_GLEAN_10029806	XP_022300022.1	alpha-L-fucosidase-like isoform X1		0.7722
yibei_GLEAN_10026261	XP_022329082.1	osteoclast-stimulating factor 1-like		0.7388
yibei_GLEAN_10028825	XP_021374846.1	pleckstrin homology domain-containing family A member 1-like		0.6247
yibei_GLEAN_10017110	XP_022293912.1	tyrosine-protein kinase JAK2-like isoform X2		0.3699
Energy metabolism			1	
yibei_GLEAN_10032781	XP_011448800.1	aldose 1-epimerase isoform X1	0.6914	0.6916
yibei_GLEAN_10006843	XP_022321339.1	isocitrate dehydrogenase [NAD] subunit beta, mitochondrial-like	0.7722	0.8028
yibei_GLEAN_10027755	XP_021355538.1	enolase 4-like isoform X4	1.2627	
yibei_GLEAN_10031289	OWF54443.1	Phosphoglucomutase-1	0.8085	
yibei_GLEAN_10030873	XP_027899350.1	NADH dehydrogenase [ubiquinone] 1 alpha subcomplex subunit 5	0.6991	
yibei_GLEAN_10011053	XP_021368351.1	cinnamyl alcohol dehydrogenase 8-like isoform X2		0.4288
yibei_GLEAN_10018823	AOR07076.1	mitochondrial ATP synthase subunit d precursor		1.2296
yibei_GLEAN_10011582	XP_021348297.1	V-type proton ATPase 16 kDa proteolipid subunit		1.8389
yibei_GLEAN_10017617	XP_011424185.1	phosphoenolpyruvate carboxykinase, cytosolic [GTP] isoform X1		0.6209
Lipid metabolism			1	
yibei_GLEAN_10017128	XP_011453569.1	phosphoethanolamine N-methyltransferase 3 isoform X1	1.438	
yibei_GLEAN_10030036	XP_021369900.1	monoglyceride lipase-like isoform X1	1.541	
yibei_GLEAN_10022257	XP_022301267.1	acetoacetyl-CoA synthetase-like	1.4266	
yibei_GLEAN_10032607	XP_011425291.1	phospholipase B1, membrane-associated-like		0.4038
Ion homeostasis				
yibei_GLEAN_10022970	XP_011450059.1	EF-hand calcium-binding domain-containing protein 5-like isoform X1	1.7475	1.8024
yibei_GLEAN_10013077	XP_025088321.1	sodium- and chloride-dependent taurine transporter-like	2.1074	
yibei_GLEAN_10002833	PFX23011.1	Calmodulin	0.3819	
yibei_GLEAN_10015019	XP_022310738.1	calmodulin-like protein 4		0.7516
yibei_GLEAN_10018664	XP_022315953.1	EF-hand calcium-binding domain-containing protein 6-like isoform X2		1.7431
yibei_GLEAN_10019514	XP_021340611.1	zinc finger ZZ-type and EF-hand domain-containing protein 1-like isoform X2		1.6331
yibei_GLEAN_10024939	XP_021377425.1	androglobin-like isoform X6		1.5648
yibei_GLEAN_10011582	XP_021348297.1	V-type proton ATPase 16 kDa proteolipid subunit		1.8389
yibei_GLEAN_10034360	XP_021370577.1	sodium-dependent phosphate transport protein 2B-like isoform X1		1.4312
yibei_GLEAN_10028951	XP_023930119.1	phosphate carrier protein, mitochondrial		1.2895
yibei_GLEAN_10006225	XP_021372793.1	calcium-binding mitochondrial carrier protein SCaMC-2-like		0.7521

# 3.4 Metabolic profiles of deep-sea mussel gills under Cd exposure

Compared with the control group, 21 and 68 metabolic features were significantly altered in the low dose- and high dose-Cd exposure groups, respectively (Table 2). These differential metabolic profiles demonstrated that deep-sea mussels relied on differing metabolic strategies when exposed to Cd at different exposure concentrations. Importantly, only eight metabolites in total were commonly altered by both Cd exposure treatments. The metabolite classes identified included amino acids, carbohydrates, nucleotides, and lipids, as well as others. KEGG enrichment analysis demonstrated seven pathways (ABC transporters, protein digestion and absorption, aminoacyl-tRNA biosynthesis, valine, leucine and isoleucine biosynthesis, choline metabolism in cancer, glycine, serine and threonine metabolism, and mineral absorption) were commonly enriched when comparing each of the Cd-exposed group with the control group (Figure 3S). Specially, pathways of lysine degradation, fructose



#### FIGURE 4

KEGG pathway enrichment analyses of up and low-regulated differentially expressed proteins (DEPs) detected in gill mussels of *Gigantidas platifrons* treated with LCd (100 μg/L Cd) or HCd (1000 μg/L Cd). (A) up-regulated DEPs, LCd; (B) up-regulated DEPs, HCd; (C) down-regulated DEPs, LCd; (D) down-regulated DEPs, HCd.

TABLE 2 Significantly expressed metabolites in the studied deep-sea mussel in response to a low and high dose of cadmium exposure (Note: Symbol  $\uparrow$  indicates the metabolite is upregulated;Symbol  $\downarrow$  indicates the metabolite is downregulated).

Metabolites	Lcd vs Ck	Trend	HCd vs Ck	Trend
Amino acid & derivatives	L-Threonine	1	Beta-Alanine	1
	Allocystathionine	1	D-Proline	1
	Betaine	Ļ	L-Threonine	1
	L-Glutamate	Ļ	L-Isoleucine	1
			L-Histidine	1
			L-Tyrosine	1
			L-Alanine	1
			L-Phenylalanine	1
			Glycine	1
			Allocystathionine	1
			Tyramine	1
			Phenyllactic acid	1
			L-Pipecolic acid	1
			Linoleic acid	Ļ
			Betaine	Ļ
Carbohydrates	D-Mannose	Ļ	Beta-D-Fructose 6-phosphate	Ļ
	Erythritol	Ļ	N-Acetylglucosamine 1-phosphate	Ļ
	Alpha-D-Glucose	Ļ	L-Malic acid	Ļ
			((	Continued)

### TABLE 2 Continued

Metabolites	Lcd vs Ck	Trend	HCd vs Ck	Trend
	D-Allose	Ļ	N-Acetylmannosamine	Ļ
			D-Glucose 6-phosphate	Ļ
			UDP-N-acetylglucosamine	Ļ
			D-Mannose-6-phosphate	Ļ
			UDP-D-Galactose	Ļ
			N-Acetyl-D-Glucosamine 6-Phosphate	Ļ
			Argininosuccinic acid	Ļ
			N-Acetyl-D-glucosamine	Ļ
			Ribulose 5-phosphate	1
			D-Lyxose	1
Nucleotides & derivatives	Inosine 5'-monophosphate	Ļ	Allopurinol riboside	1
	Xanthine	↑ (	Deoxyguanosine	1
			Adenosine 2',3'-cyclic monophosphate	1
			2-Hydroxyadenine	↑ (
			Guanosine	↑ (
			Adenosine 3'-monophosphate	<u></u>
			Hypoxanthine	1
			Xanthine	1
			Thymine	1
			Thymidine	1
			Inosine	↑ (
			2'-Deoxyuridine	↑ (
			2'-Deoxyinosine	↑ (
			Deoxyinosine	↑ (
			Adenine	↑ (
			1-Methyladenosine	↑ (
			Deoxyguanosine	1
			2-Deoxyribose 5-phosphate	<b>†</b>
			Adenosine monophosphate	Ļ
			Guanosine 5'-monophosphate	Ļ
			S-Methyl-5'-thioadenosine	Ļ
Lipids	PGD2	↑ (	PGD2	↑ (
	Glycerophosphocholine	1	Glycerol 3-phosphate	1
	21-hydroxypregnenolone	↑ (	Valeric acid	↑ (
	15-Deoxy-delta-12,14-PGJ2	1	1-Palmitoyl-sn-glycero-3-phosphocholine	↑ (
			1-O-Octadecyl-sn-glyceryl-3-phosphorylcholine	1
			1-O-(cis-9-Octadecenyl)-2-O-acetyl-sn-glycero-3-phosphocholine	1
			1-Stearoyl-2-hydroxy-sn-glycero-3-phosphocholine	1
	I	1	(0	 Continued)

#### TABLE 2 Continued

Metabolites	Lcd vs Ck	Trend	HCd vs Ck	Trend
			Phthalic acid Mono-2-ethylhexyl Ester	1
			Adrenic Acid	Ļ
			gamma-L-Glutamyl-L-glutamic acid	Ļ
			cis-9-Palmitoleic acid	Ļ
Others	L-Pipecolic acid	1	Choline	1
	Stearidonic Acid	1	Stearidonic Acid	1
	Diacetyl	1	5-L-Glutamyl-L-alanine	1
	Diethyltoluamide	1	Propionic acid	1
	2-Ethoxyethanol	1	Acamprosate	1
	Glutaric acid	Ļ	Citraconic acid	Ļ
	2-Oxoadipic acid	Ļ	Muramic acid	Ļ
			Glutaric acid	Ļ

and mannose metabolism, proximal tubule bicarbonate reclamation were uniquely enriched in low-dose group, and pathways of purine metabolism, glucosinolate biosynthesis, mTOR signaling pathway, amino sugar and nucleotide sugar metabolism were enriched in high-dose group (Figure 3S).

# **4** Discussion

# 4.1 Metal subcellular distribution and detoxification

The accumulation of Cd in deep-sea mussels increased with increasing exposure concentrations, indicating that deep-sea mussels are robust indicator organisms of environmental changes caused by metal pollution (Martins et al., 2017; Zhou et al., 2021). The subcellular compartmentalization of metals reflects dynamic processes that occur during metal accumulation and provides insights into temporary metal-binding ligands and the detoxification process in organisms (Wallace et al., 2003; Pan and Wang, 2008). As the exposure concentrations increased, the proportion of Cd in its major binding site (HSPF) significantly decreased, whereas Cd's proportions in other fractions (mainly cellular debris) increased correspondingly. A similar pattern was also found in the scallop Chlamys nobilis under Cd or Zn exposure (Pan and Wang, 2008). This phenomenon is called the spillover effect, namely because once the major binding pool is saturated and its binding capacity now exceeded, metal(s) may subsequently become associated with other fractions (Pan and Wang, 2008). Cellular debris contains tissue fragments, cell membranes, and other cellular components without any known function. Under Cd exposure, newly accumulated Cd in deep-sea mussel gills tended to bind to less sensitive fractions in cellular debris, preventing it from binding to sensitive fractions in the cell and thus decreasing Cd's toxicity (Lucu and Obersnel, 1996). Interestingly, there was a very small proportion of Cd (< 9%) eliminated into MTLP, whether in the control or either exposure group, suggesting that MTLP could play a minor role in Cd detoxification. This result is in line with previous findings that MTLP is probably not relied upon for a major metal detoxification strategy in deep-sea bivalves (Hardivillier et al., 2004; Zhou et al., 2021).

Apart from detoxification by binding with cellular debris other than MTLPs, deep-sea mussels also can express detoxification enzymes for adaption and survival under Cd stress. In this study, exposure to Cd resulted in the significant up-regulation of glutathione transferases (GSTs) in deep-sea mussel gills. The GSTs are phase II detoxification enzymes responsible for catalyzing the conjugation of glutathione into electrophilic compounds that eventually lead to their excretion (McDonagh and Sheehan, 2006). GSTs also play a role in defense against oxidative stress by metabolizing reactive products formed during lipid peroxidation (Canesi et al., 1999). Furthermore, another detoxification enzyme, cytochrome P450 2C31 (CYP450 [2C31]), was found uniquely down-regulated in deep-sea mussels of high dose-Cd exposure group. In a toxicological study with clam Mactra chinensis exposed to Cd, CYP450 (2C31) was inhibited by a low Cd concentration (0.69 mg/L), yet induced by high Cd exposure concentrations (1.38 and 2.76 mg/L) (Zhang et al., 2016). The current results bring forth new and relevant information regarding the metal detoxification response of deep-sea mussels.

# 4.2 Stress response associated with Cd exposure

Oxidative stress is a major mechanism underpinning Cd toxicity, because its overload could disrupt the redox balance at a cellular scale and stimulate the generation of ROS (Valko et al., 2005). To counter oxidative stress, an antioxidant cascade that includes SOD, CAT, and GSH could be invoked. Among these antioxidants, SOD is responsible for the disproportionation of superoxide anions into hydrogen peroxide and dioxygen, meanwhile CAT can act as a scavenger of hydrogen peroxide. This SOD-CAT system provides the first line of defense against Cd toxicity (Li et al., 2011). However, in the current study, both SOD and CAT activity were inactivated in either Cd exposure treatment, implying that they might not be suitable for use as sensitive biomarkers in deep-sea mussels during the monitoring of Cd pollution. Altered GSH levels could also be serve as a biomarker for oxidative stress (Valko et al., 2005). Actually, the patterns of change in GSH tend to differ under different experimental conditions: GSH was reportedly depleted substantially during acute Cd exposure while being markedly augmented following chronic Cd exposure (Liu et al., 2009). Our results show that exposure to a low dose of Cd slightly raised the GSH in deep-sea mussels whereas a high dose of it significantly induced GSH in them. Besides, as a major consequence of Cd-induced oxidative stress (Liu et al., 2009), LPO was increased exclusively in deep-sea mussels belonging to the high-dose group. The altered GSH and increased LPO together suggest oxidative stress in mussels of the high-dose group. According to the biochemical data, a low dose of Cd exposure seems unlikely to have caused overt changes to the mussel's cellular oxidative status and lipid peroxidation. Nevertheless, the markedly augmented level of GST detected via the proteomic technique clearly evidenced that these deep-sea mussels in the low dose-Cd exposure group also incurred oxidative stress. Therefore, the proteome can convey a more complete picture of the biochemical changes that deep-sea mussels undergo when faced with Cd stress, and it is more sensitive than traditional approaches for discovering novel biomarkers.

From the proteomic profiles, Cd exposure altered 11 proteins that are linked to stress responses. Among these proteins, major vault protein (MVP) was uniquely induced in deep-sea mussel gills of lowdose group. MVP is the main component of the ubiquitous, vault particles that exceed the size of ribosomes. Although the exact role of MVP remains unclear, research has shown that vault complexes are involved in cellular responses to environmental toxins by directly binding and excreting conjugated metabolites, thereby precluding a metal's accumulation and toxic effects (Berger et al., 2009). In addition, because MVP can play a strong regulatory role in apoptosis by binding to several stress-induced effectors of signaling cascades, it has been proposed as a candidate target for modulating apoptosis resistance in aging modulation and cancer treatments (Lara et al., 2011). The up-regulated expression of MVP in mussels exposed to Cd's low dose suggested an anti-apoptotic response to offset potential toxic effects of Cd. Expression of seven stress-related proteins belonging to the class of heat shock protein 70 (HSP70) family were significantly altered in deep-sea mussel gills of both Cdexposure groups. HSP70 is a group of highly conserved and abundant proteins that help organisms deal with abiotic stressors (heat shock, salinity change, acidification and contaminants) by refolding, sorting, translocating, and degrading deformed proteins, as well as being cochaperones and regulating the biologically active form of other proteins (Kim and Schöffl, 2002; Chaudhary et al., 2019; Yusof et al., 2022). A field investigation of deep-sea mussel Bathymodiolus azoricus from Atlantic hydrothermal vents confirmed that this species could express and induce a mixture of constitutive heat-shock protein isoforms, ranging from 64 and 76 kDa, enabling it to survive despite toxic vent fluids (Pruski and Dixon, 2007). Accordingly, HSP70 can be set up as a general stress biomarker in deep-sea mussels to monitor deep-sea environmental changes caused by Cd leakage during mining activities.

# 4.3 Cytoskeleton changes associated with Cd exposure

Eighteen proteins associated with cytoskeleton were altered by Cd exposure. In constituting the system of fibrillar structures in the cytoplasm of eukaryotic cells, the cytoskeleton is primarily composed of three types of fibrils: microtubules, actin filaments, and intermediate filaments. These fibrils could serve as 'weak' elements during episodes of metal stress, inducing the expression of several molecular chaperones, including small HSP and HSP70, to protect the cytoskeleton (Liang and MacRae, 1997). Besides, there is compelling evidence that cytoskeletal proteins could be among the first targets of oxidative stress and that they participate in the homeostasis of oxidative stress (Dalle-Donne et al., 2001). Given that cytoskeletal proteins are crucial for a cell's structure and intracellular organization, any disruption in the architecture of one of the three cytoskeletal networks is liable to cause adverse effects to cells' integrity and functioning (Fletcher and Mullins, 2010). Cytoskeletal injuries in bivalves induced by metal pollution have been repeatedly confirmed (Gomes et al., 2013; Xu et al., 2016). And these cytoskeletal-related DEPs are generally contain actins, tubulins, and myosins in marine bivalves (Xu et al., 2019), a finding consist with our study's results. Consequently, the significantly altered cytoskeletal proteins demonstrate that oxidative stress and subsequent disturbed cytoskeleton and cell structures can be induced by Cd in deepsea mussels.

# 4.4 Immune response associated with Cd exposure

The immune system of deep-sea mussels was also affected by their exposure to Cd, as evinced by the altered ACP and AKP activity as well as the thirty regulated proteins (17 DEPs in the low-dose group, 20 DEPs in the high-dose group) involved in the GO term "immune system process". Both ACP and AKP are ubiquitous enzymes in organisms that hydrolyze phosphate-containing compounds and therefore can serve as effective biomarkers of immune responses to environmental contamination (Mazorra et al., 2002). AKP activity is reliably indicative of cell membrane's integrity and is involved in the osmoregulation (Lovett et al., 1994), immune defense (Revel et al., 2019), and calcification process of bivalves (Hohagen and Jackson, 2013). Any perturbation (i.e., altered temperature, increased exposure to metals, microplastics exposure) to the membrane properties could result in AKP's altered activity (Jiang et al., 2012; Huang et al., 2021; Sun et al., 2021). In the present study, a significant increase in AKP activity resulted from increasing the Cd exposure concentrations, this implying a strong immune response of mussels under Cd exposure to resist metal toxicity, or disturbance in osmoregulation from disrupted membrane integrity. ACP has been proven to be associated with lysosomal functionality (Pampanin et al., 2002). Upon entering body, toxic chemicals such as metals can be sequestered within the lysosome, causing increased toxicity and eventually affecting the permeability, integrity or structure of the lysosome (Moore, 1994). Under certain circumstances when lysosomes become overburdened, hydrolases burst out of their leaky membranes and are released into

the extracellular environment, with the possibility of cause cellular injury or even death (Kohler et al., 2002). Hence, the decreased ACP activity in mussels of the high-dose group might due to enzyme escape from damaged lysosomes (Jiang et al., 2012).

When studying the effects of metals on the immune functioning of marine bivalves, researchers have mainly examined total and differential hemocyte counts and phagocytic activity (Hoher et al., 2012). By using proteomics, we discerned the predominant presence of largely altered immune-related proteins in deep-sea mussel gills under Cd stress. These significantly expressed proteins have never been reported before in deep-sea invertebrates in response to metal exposure. For instance, interleukin enhancer-binding factor 2 (ILF2) homolog, stimulator of interferon genes (STING) protein 2, and E3 ubiquitin-protein ligase TRIM56-like were the three of the most prominently regulated proteins in deep-sea mussels when exposed to either Cd treatment, indicating that Cd can disturb the immune homeostasis in deep-sea mussels. As one of the six categories of cytokines that are inducible immune regulatory proteins, interleukin could play pivotal roles in modulating cell communication, growth, development, and differentiation to maintain immune system homeostasis (Yang et al., 2022). Although there are few reports of interleukin and its analogues in invertebrates, ILF has been reported to exist in invertebrates where it functions as regulator in some innate immune responses by affecting the Toll signaling pathway and cell apoptosis (Venier et al., 2011; Yang et al., 2022). As for STING, it is well known for inducing the generation of interferon and proinflammatory cytokines. Being an endoplasmic reticulum membrane protein anchored by several transmembrane domains at its N-terminal region, STING can function as a cGMP-AMP synthase downstream signal adapter in the innate immune cytosolic DNAsensing pathway (McKnight et al., 2020; Zhang et al., 2020). Concerning E3 ubiquitin-protein ligase TRIM56, it can promote the function of STING by binding to the latter's C-terminal domain and serve as a powerful modulator of innate immunity (Tsuchida et al., 2010). Overall, the significant alterations of the aforementioned biochemical indexes and molecules strongly suggest that multiple mechanisms are involved in mobilizing the immune response against Cd stress in the studied mussel species.

# 4.5 Alteration of energy metabolism associated with Cd exposure

Metals can impair energy homeostasis in marine organisms because the latter require more energy for mounting one or more stress responses, detoxification steps, or damage repair mechanisms (Sokolova et al., 2012). Glycolysis is the process of decomposing glucose into pyruvate and lactate, leading to the generation of ATP (Satyanarayana, 2021). Herein, five glycolysis-related proteins (phosphoglucomutase-1, enolase 4-like isoform X4, aldose 1epimerase isoform X1, cinnamyl alcohol dehydrogenase 8-like isoform X2, and phosphoenolpyruvate carboxykinase) were found to be significantly changed in both Cd-exposure groups. Of them, four proteins were down-regulated significantly, suggesting inhibition of glycolysis in deep-sea mussels from exposure to low or high doses of Cd. Concurrently, metabolomics data revealed that all metabolites ( $\alpha$ - D-glucose,  $\beta$ -D-fructose 6-phosphate, D-glucose 6-phosphate) associated with glycolysis were down-regulated, which further confirmed that Cd exposure inhibited glycolysis in deep-sea mussels, as noted above. Previous research with the sea cucumber *Apostichopus japonicas* revealed that its glycolysis could be promoted in the initial phase of exposure to Cd, but then became inhibited as the duration of metal stress was prolonged. Inhibition effects tend to occur earlier if the exposure concentrations are higher (Li et al., 2016). In addition, we also detected isocitrate dehydrogenase (IDH) as commonly down-regulated in both Cd-exposure groups. IDH is an indispensable enzyme that regulates the citric acid cycle (TCA), catalyzing the conversion of isocitrate to oxalosuccinate and then to  $\alpha$ -ketoglutarate (Satyanarayana, 2021). Thus the decreased IDH and glycolysis-associated proteins together suggests that Cd toxicity is able to impair the energy metabolism in deep-sea mussels.

Notably, we found disparate effects of Cd on the oxidative phosphorylation of deep-sea mussels in the different treatment groups. Linked to the mitochondrial electron transport chain (ETC), oxidative phosphorylation is the major source of ATP production that takes place in the mitochondria of aerobic organisms. In the low-dose group, the NADH dehydrogenase [ubiquinone] 1 alpha subcomplex subunit 5 (NDUFA5), which is responsible for converting NAD from reduced form (NADH) into oxidized form (NAD<sup>+</sup>) during oxidative phosphorylation, was downregulated relative to the control group. This reduction in NDUFA5 might have been caused by the reduced IDH activity in the TCA cycle. Metals can reportedly influence key enzymes such as IDH in the TCA cycle, leading to less generated NADPH and NADH, which can then negatively affect the normal function of ETC and ultimately lead to lower ATP production (Sun et al., 2022). In the high-dose group, however, we found that NDUFA5 was unaltered, but all the proteins associated with oxidative phosphorylation (mitochondrial ATP synthase subunit d precursor and V-type proton ATPase 16 kDa proteolipid subunit) were up-regulated and related to ATP synthase, implying an increased demand for ATP production in deep-sea mussels. Increased oxidative phosphorylation was reported in the rock fish Sebastes schlegelii under arsenic exposure and the mussel Mytilus galloprovincialis under nano-Ag exposure, both revealing a process of energetic re-adjustment to counteract the effects of arsenic toxicity (Gomes et al., 2013; Xu et al., 2019). In particular, to compensate for the diminished energy supply caused by the inhibition of glycolysis and the TCA, deep-sea mussels bolstered amino acid levels in their body to protect themselves from more damage caused by high-dose Cd exposure (Table 2). By serving as precursors for acetyl CoA and glucose synthesis, amino acids can also be degraded and utilized as energy sources in the body (Satyanarayana, 2021). Using metabolomics, a similar phenomenon has been extensively documented in marine bivalves when they are exposed to toxic chemicals (Viant et al., 2003; Zhou et al., 2021). Overall, our findings clearly show that Cd could alter mitochondrial metabolism and thereby risk impairment of ATP generation in deepsea mussel gills. Collectively, these changes in energy metabolism coupled with a disordered cytoskeleton and lipid peroxidation would lead to stark physical changes in the membrane, affecting the respiratory, oxidative and calcium buffer capacity of mitochondria in gills of deep-sea mussels (Gomes et al., 2013).

# 4.6 Alteration of lipid and nucleotide metabolisms associated with Cd exposure

As the principal concentrated fuel reserve of organisms, lipids play important roles in maintaining cellular structure and regulating cellular metabolism (Satyanarayana, 2021). Alterations to lipid metabolism reportedly serve as protective strategies of organisms in response to environmental stress (Liu et al., 2016; Meng et al., 2018; Zhou et al., 2021). A metal exposure study using blue mussel Mytilus edulis demonstrated that Cd could significantly disturb its lipid metabolism by changing lipid structures and the fatty acid composition in mussels (Fokina et al., 2013). Furthermore, a metal exposure experiment with the crab Chiromantes dehaani showed that Cd could disturb lipid metabolism by weakening the ability of lipid digestion, transportation and synthase, thereby affecting the normal growth and development of that organism (Liu et al., 2016). Here, two proteins associated with lipid metabolism, phosphoethanolamine Nmethyltransferase 3 isoform X1 and monoglyceride lipase-like isoform X1, were found uniquely up-regulated in low-dose group. Monoglyceride lipase catalyzes the final step of phospholipid and triglyceride breakdown by hydrolyzing monoacylglycerols into fatty acids and glycerol molecules (Aschauer et al., 2016). These products can be used for energy production or synthetic reactions. Hence, the up-regulated monoglyceride lipase-like isoform X1 is further evidence of a disturbed energy homeostasis in deep-sea mussels caused by Cd exposure. In plants, phosphoethanolamine N-methyltransferase has been suggested as one key bottleneck for the de novo synthesis of phosphatidylcholine, the chief component of plasma membrane phospholipids in most eukaryotes, which has been shown to be upregulated under abiotic stress conditions (Jost et al., 2009). We also observed augmented levels of glycerophosphocholine in the metabolic profiles of mussels in the low-dose group. Glycerophosphocholines are critical components of biological membranes and contribute to lipid bilayer asymmetry. Therefore, those up-regulated proteins and metabolites in mussels of the low dose-Cd exposure group might together reflect an alteration of biological membranes. In the high dose-Cd exposure group, the phospholipase B1 membraneassociated-like was down-regulated exclusively. Phospholipase B1 (PLB1) can hydrolyze ester linkages on membrane phospholipids and liberate fatty acids (Chayakulkeeree et al., 2011). It is involved in membrane homeostasis and remodeling in fungi (Siafakas et al., 2007) and could be modulated under various environmental and physiological conditions (Mukherjee et al., 2003). As gleaned from the metabolic profiles, fatty acids including adrenic acid, cis-9palmitoleic acid, and valeric acid were up- or down-regulated. One explanation for those altered proteins and metabolites is that deep-sea mussels may modify their cell membrane's composition in order to maintain/rebuild its stability and integrity. Another possibility is that Cd exposure resulted in damaged membranes of deep-sea mussels due to lipid peroxidation (He et al., 2020). Therefore, our results demonstrate that Cd has the potential to alter the membrane state of mussel gills, directly or indirectly, by disrupting membrane lipid biosynthesis and fatty acid composition.

Perturbed nucleotide metabolism is often reported in bivalves exposed to pollutants (Jones et al., 2008; Fasulo et al., 2012; Zhou et al., 2021). In the current study, a total of 21 nucleotides or their derivatives (Table 2) were found altered in mussels of high-dose group, with only two

nucleotides changed in low-dose group, indicating dose-dependent toxicity effects of Cd. As structural components of DNA, RNA as well as many coenzymes, the availability of nucleotides underpins diverse metabolic reactions (Satyanarayana, 2021). Changed nucleotides profiles could be linked to neurotransmitter processes and cell signaling given that nucleosides (e.g., guanosine and adenosine), bases (e.g., guanine and adenine), and their metabolic products (e.g., xanthine and hypoxanthine) can be discharged into the extracellular space as intercellular signaling molecules (Rathbone et al., 1999). Apart from their function as signaling molecules, adenine and its analogues and inosine can also act as immunomodulatory molecules, exerting cellular protective and multiple anti-inflammatory effects in organisms (Haskó et al., 2000). That these purine metabolites increased in deep-sea mussels suggests their immune and inflammatory responses were enhanced under Cd exposure. Changing nucleotide profiles may also have contributed to peroxisome proliferation in rats (Ringeissen et al., 2003). As common oxidative organelles in eukaryotes, peroxisomes can function to protect the cell from toxic substances, contributing to the detoxification process of organisms. Thus, the increased levels of nucleotides identified in this study may point to elevated antioxidant enzyme activities and other catalytic activities during the metal detoxification process (Leonard et al., 2014).

# 4.7 Ion homeostasis associated with Cd exposure

Calcium binding proteins (CaBPs) have been identified as molecular targets for cadmium binding and cadmium toxicity. Due to the similarity between Cd<sup>2+</sup> and Ca<sup>2+</sup> in both structure and size, Cd<sup>2+</sup> can not only enter cells via the Ca<sup>2+</sup> uptake pathway but also compete with Ca<sup>2+</sup> for the binding sites in CaBPs, in this way inhibiting the calcium influx of organisms (Choong et al., 2014), which can eventually affect cellular Ca<sup>2+</sup> homeostasis. Here, CaBPs of the EF-Hand superfamily (EF-CaBPs) including calmodulin, calmodulin-like protein 4, EF-hand calciumbinding domain-containing protein 5-like isoform X1, EF-hand calcium-binding domain-containing protein 6-like isoform X2, and zinc finger ZZ-type and EF-hand domain-containing protein 1-like isoform X2, were all significantly altered in both Cd exposure groups. EF-CaBPs regulate all aspects of cellular calcium signaling pathways by binding to Ca<sup>2+</sup>, including Ca<sup>2+</sup> gating control, modulation of the amplitude and duration of Ca2+ signals, and transduction of Ca2+ signals into biochemical responses (Girard et al., 2015). Therefore, the abnormal expression of EF-CaBPs indicated that Ca<sup>2+</sup>-triggered signaling could be involved in how deep-sea mussels overcome a metal challenge.

Deep-sea mussel gills are in direct contact with the ambient environment, and serve as an important organ for regulating cellular osmoregulation. We found that certain membrane transport proteins belonging to the solute carrier (SLC) family, such as sodium- and chloride-dependent taurine transporter-like in low-dose Cd-exposed group and sodium-dependent phosphate transport protein 2B-like isoform X1, phosphate carrier protein, and calcium-binding mitochondrial carrier protein SCaMC-2-like were significantly altered in high dose-Cd exposure group. Specifically, for these treated deep-sea mussels of high-dose group, the 16 kDa proteolipid (subunit c) of V-ATPase involved in translocating protons across lipid bilayers was significantly up-regulated. V-ATPase is capable of ensuring



#### FIGURE 5

Potential mechanism schematic of Cd toxicology on deep-sea mussel G. platifrons according to KEGG (100µg/L Cd, A and 1000µg/L Cd, B). The altered proteins and metabolites were shown by marking the names in red (up-regulated) or blue (down-regulated) color. Abbreviations: ACP, acid phosphatase, AKP, alkaline phosphatase; IAP, baculoviral IAP repeat-containing protein 7-B isoform X3; STING, stimulator of interferon protein 2; AP-2, AP-2 complex subunit alpha-2-like; IEB, interleukin enhancer-binding factor 2 homolog; SVA, superkiller viralicidic activity 2-like 2; TRIM56, E3 ubiquitin-protein ligase TRIM56-like; NTF, nuclear transport factor 2-like; CTSL: shell fibrous prismatic cathepsin-like protein 1; Myosin, unconventional myosin-XVI-like isoform X10; IST1, IST1 homolog; PXN, leupaxin-like isoform X2; DNAH, dynein heavy chain 6, axonemal-like isoform X3; DNAI1, dynein intermediate chain 2, ciliary isoform X2; TUBB, tubulin beta chain-like; DNAH10, dynein heavy chain 10, axonemal-like isoform X2; DNAH5, dynein heavy chain 5, axonemallike isoform X2; TCTP, translationally-controlled tumor protein homolog; ZZEF, zinc finger ZZ-type and EF-hand domain-containing protein 1-like isoform X2; TSC, tubulin-specific chaperone A; 3'-AMP, Adenosine 3'-AMP monophosphate; 2',3'-cyclic AMP, Adenosine 2',3'-cyclic monophosphate; D-Ribulose-5P, Ribulose 5'-phosphate; PRPP, phosphoribosyl diphosphate; ATP5H, mitochondrial ATP synthase subunit d precursor; ATPeV0C, V-type proton ATPase 16 kDa proteolipid subunit; IDH3, isocitrate dehydrogenase [NAD] subunit beta, mitochondrial-like; NDUFA5, NADH dehydrogenase [ubiquinone] 1 alpha subcomplex subunit 5; ADH, cinnamyl alcohol dehydrogenase 8-like isoform X2; GPC, Glycerophosphocholine; PEA, phosphoethanolamine; Cyt P450, cytochrome P450 2C31; GST, glutathione S-transferase; HSP70, heat-shock protein 70 family; CALM. Calmodulin; pgm, Phosphoglucomutase-1; GALM, aldose 1-epimerase isoform X1; ENO, enolase 4-like isoform X4; AACS, acetoacetyl-CoA synthetase-like; NMT, phosphoethanolamine N-methyltransferase 3 isoform X1; MGLL, monoglyceride lipase-like isoform X1; PCK, phosphoenolpyruvate carboxykinase, cytosolic [GTP] isoform X1; NMDA, glutamate receptor ionotropic; NPT2b, sodium-dependent phosphate transport protein 2B-like isoform X1; GPCR, 5hydroxytryptamine receptor 7; PSMD4, 26S proteasome non-ATPase regulatory subunit 4-like; PSMF1, Proteasome inhibitor PI31 subunit; PSME3, proteasome activator complex subunit 3-like.

intracellular pH homeostasis, protecting cells against endogenous oxidative stress induced by divalent heavy metals, and inhibiting ROS production (Techo et al., 2020). Similar findings were found for the shrimp *Litopenaeus vannamei* under copper exposure (Guo et al., 2021). Meanwhile, the metabolic profiles of deep-sea mussels exposed to a high dose of Cd revealed that large amounts of amino acids were significantly induced. Previous research has shown that marine invertebrates such as deep-sea mussels can utilize intracellular amino acids to maintain their osmolality in balance with their surroundings (Zhou et al., 2021). Collectively then, the accumulated amino acids, dysregulated SLCs, and V-ATPase (subunit c) further suggest that Cd disturbed the ion exchange function of gills in the studied deep-sea mussel.

# 5 Conclusion

This study took a holistic perspective, incorporating both traditional approaches and 'omic' techniques to study the toxicological effects of Cd upon a keystone deep-sea bivalve

Gigantidas platifrons. To do this, G. platifrons was exposed for 7 days to Cd at concentrations of 100 or 1000  $\mu$ g/L, and metal accumulation, metal distribution in the subcellular compartments, anti-oxidative and immune-related biochemical parameters, and protein and metabolic profiles were evaluated as response variables. Under Cd exposure, the deep-sea mussel G. platifrons heavily accumulated Cd in its gills, where this Cd tended to bind to cellular debris other than MTLP, suggesting that MTLP might not be a major detoxification strategy in deep-sea bivalves. Parallel proteomic and metabolomics analyses revealed that the two doses of Cd evoked similar cellular responses, namely the elicitation of stress and immune responses and the disruption of the cytoskeleton structure, ion homeostasis, and primary metabolism processes in deep-sea mussels. A potential toxicological mechanism of Cd on deep-sea organisms is schematically illustrated in Figure 5. Overall, with deep-sea mining activities predicted to significantly impact marine ecosystems, the current results provide much-needed critical insights into how deep-sea organisms respond to metals as well as contributing to global information for identifying early responses of deep-sea organisms to metal pollution.

### Data availability statement

The datasets presented in this study can be found in online repositories. The names of the repository/repositories and accession number(s) can be found in the article/Supplementary Material.

### Author contributions

LZ contributed to conception, methodology, software, writing of the study. ML, ZZ, HC, and CL carried out the experiments. MW, HW, HZ, and LC organized the database. CLL directed the manuscriptrevision. All authors contributed to the article and approved the submitted version.

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# **Conflict of interest**

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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### Supplementary material

The Supplementary Material for this article can be found online at: https://www.frontiersin.org/articles/10.3389/fmars.2023.1087411/ full#supplementary-material

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