

1 **What can genetics tell us about the history of a human-mediated introduction of the Golden-striped**
2 **salamander south of its native range**

3

4 Sequeira F¹, Aguilar FF^{1,4}, Madeira FM⁴, Teixeira J³, Crespo E², Ferrand N^{1,5,6}, Rebelo R⁴

5

6 ¹CIBIO, Centro de Investigação em Biodiversidade e Recursos Genéticos, InBIO Laboratório Associado,
7 BIOPOLIS Program in Genomics, Biodiversity and Land Planning, Campus de Vairão, Universidade do
8 Porto, 4485-661 Vairão, Portugal

9 ²Departamento de Biologia Animal, Faculdade de Ciências da Universidade de Lisboa, Campo Grande,
10 1749-016 Lisboa, Portugal

11 ³CIIMAR - Interdisciplinary Centre of Marine and Environmental Research, U. Porto – University of Porto,
12 Terminal de Cruzeiros do Porto de Leixões, Av. General Norton de Matos s/n, 4450-208 Matosinhos,
13 Portugal

14 ⁴cE3c centre for Ecology, Evolution and Environmental Changes & CHANGE - Global Change and
15 Sustainability Institute, Faculdade de Ciências da Universidade de Lisboa, Campo Grande, 1749-016
16 Lisboa, Portugal

17 ⁵Departamento de Biologia, Faculdade de Ciências, Universidade do Porto, Rua do Campo Alegre s/n,
18 Porto, 4169-007, Portugal

19

20 ⁶Department of Zoology, University of Johannesburg, PO Box 524, Johannesburg, Auckland Park, 2006,
21 South Africa

22 *Corresponding author – Fernando Sequeira, CIBIO, Centro de Investigação em Biodiversidade e

23 Recursos Genéticos, InBIO Laboratório Associado, Campus de Vairão, Universidade do Porto, 4485-661

24 Vairão, Portugal. Telephone: 00351 252660400; E-mail: fsequeira@cibio.up.pt

25 **Keywords:** Amphibians, *Chioglossa lusitanica*, founder effect, genetic drift, isolation, Sintra, Portugal

26 **Running title:** Human-mediated introduction of an Iberian salamander

27 **Abstract**

28 The golden-striped salamander is a streamside species endemic to the northwestern corner of the Iberian
29 Peninsula. In the first half of the twentieth century, an undisclosed number of individuals of this species
30 were reportedly captured in Buçaco, central Portugal, and deliberately introduced in Serra de Sintra, 170
31 km south of its native distribution range. The discovery of a breeding population of this salamander in
32 Sintra during 2015 prompted this work: we used neutral genetic markers, the mitochondrial DNA
33 Cytochrome b (cyt b) and seven microsatellite loci to elucidate on the relic/human-introduced nature of
34 Sintra population, identify the potential source population and infer the severity of founder effect. Our
35 results support a human-mediated introduction. First, sequencing analysis of cyt b showed the presence of
36 a unique haplotype (h31) in Sintra, which was detected only in Buçaco and in two additional populations
37 located close to Mondego river. Second, microsatellite analysis showed that Sintra is more closely related
38 with populations in between Douro and Mondego rivers (Central Portugal), instead of its geographically
39 closest populations (southernmost), as would be expected if Sintra were a relic population isolated in an
40 interglacial refuge. Third, Sintra presents both reduced levels of genetic variability and effective population
41 size when compared to native populations, particularly to those of Central Portugal. Consistent with an
42 isolated population funded by a small number of individuals (inferred herein to be ca. 10 salamanders),
43 Sintra forms a geographically coherent genetic unit that is significantly differentiated from the extant native
44 *C. lusitanica* populations. Although changes in the genetic makeup of Sintra do not allow to track
45 unequivocally the origin of the introduced individuals, genetic signs from both nuclear and mtDNA data
46 provide supporting evidence for Buçaco as the most likely source population, which coincides with the
47 documented history of the introduced population in Sintra.

48 **Introduction**

49 Elucidating the history of species introductions requires a multidisciplinary endeavour in that molecular
50 approaches appear as powerful tools. For instance, crossing historical records with patterns of genetic

51 variation has contributed to identify introduced populations that were previously considered native ones
52 (Gipoliti and Amori, 2006; Clavero et al., 2016; Tiberti and Splendiani, 2019; Kehlmaier et al., 2020).
53 Since introduced populations are expected to have a genetic makeup that reflects their source population
54 (Sakai et al., 2001; Antzen et al., 2010), genetic analyses can also give important insights on the
55 reconstruction of the history of the introduction, including the identification of the source population as
56 well as the number of founders that formed the initial introduced population (“propagule size”), the severity
57 of the founder effect and bottleneck, and the number of generations over which the population has been
58 isolated from the source population (e.g. Nei et al., 1975; Frankham, 1995; Ficetola et al., 2008; Simberloff,
59 2009).

60 Amphibians are currently the most endangered tetrapod group (IUCN, 2022). This is often explained by
61 their exquisite requirements, such as of humid or unpolluted freshwater microhabitats, that makes them
62 vulnerable to even slight changes in their quality or microclimate (Button and Borzée, 2021). Dependence
63 on microclimate has led several species to cling on relic, small extent habitats, increasing their vulnerability
64 to extinction (Ceballos et al. 2020). As a response to range reductions or to a threatened status, a possible
65 conservation tool is population translocations and/or reintroductions to climatically suitable areas, but this
66 has rarely been attempted with amphibians (Kraaijeveld-Smit et al. 2006).

67 The golden-striped salamander, *Chioglossa lusitanica* (Bocage 1864), is a monotypic endemic caudate of
68 Northwestern Iberian Peninsula, where it inhabits the banks of swift running streams with dense
69 surrounding vegetation in fairly mountainous areas, characterized by high rainfall, high topographical relief
70 and mild summer and winter temperatures (Arntzen 1981; Sequeira and Alexandrino 2008). This
71 salamander has unique morpho-physiological traits that make displacements outside moist habitats severely
72 restricted (Arntzen 1995). In an undisclosed date before 1943, a Portuguese zoologist (Anthero Seabra)
73 ordered that “a few specimens” (unknown number) of the golden-striped salamander should be collected at
74 Mata do Buçaco (a 400-ha State Forest in Buçaco Mountains, Coimbra, central Portugal; Figure 1) and
75 released in the Sintra Mountains (located 20 km NW of Lisbon and about 170 km of the currently southern
76 distribution limit of the species; Figure 1) (Seabra 1943). At the time, the association of this salamander

77 with the mountain streams in the North of the country was already clear to Seabra, as well as its probable
78 absence south of the region of Coimbra-Buçaco (Seabra 1943). Seabra mentions that environmental
79 conditions of Sintra and other mountains in the south of the country were similar to those of the northern
80 mountains where *C. lusitanica* was common, and that the species should thrive there if it could somehow
81 cross the inhospitably hot and dry habitats in between (Seabra 1943). However, the species presence had
82 been cited for the Sintra Mountains at the end the XIX century by Vieira (1886) and, based on a distribution
83 ecological modelling approach, this area was indicated as a potentially suitable area for *C. lusitanica*.
84 (Teixeira et al. 2001), leading to the possibility that the species already existed in the region in the past or
85 some relict rear edge population may still subsisted till the present (Arntzen and Teixeira 2006). The success
86 of Seabra's "re-stocking essay" (as he referred to it) was never monitored, and in fact Seabra mentions it
87 only in a footnote, and specifically "so that in the future, if by chance the species is found at Sintra, its
88 origin will be explained" (Seabra 1943). Indeed, only 60 years after its putative introduction, one single
89 individual of the species was observed by the naturalist Gaston-Denis Guex in Sintra (Arntzen 1999). The
90 occurrence of a reproducing population of *C. lusitanica* in Sintra was only confirmed in 2015 (Aguilar et
91 al. 2018). Through a 2 years monitoring program, these authors found a relatively small population
92 (estimated at *ca.* 340 individuals) confined to a small area along *ca.* 110 m stretch of a single stream (Aguilar
93 et al. 2018).

94 Here we use neutral genetic markers, mitochondrial DNA cytochrome b (mtDNA cyt b) sequences and
95 seven microsatellite loci, to address whether the genetic composition of Sintra population is compatible
96 with the introduction history described by Seabra (1943) or the possibility of a relict population. By
97 comparing genetic variation between Sintra population with that reported for populations across the entire
98 species' native range based on the same genetic markers (Alexandrino et al. 2000, 2002; Sequeira et al.
99 2008), we aimed to identify the history of this population, the possible occurrence and source of introduced
100 individuals and gain insights about the severity of founder effects by determining the reduction in levels of
101 genetic diversity and the potential number of founder individuals of the introduced population.

102 **Materials and methods**

103 *Sampling and data collection*

104 A total of 97 salamander tail-tip tissue samples were collected from three localities: Sintra (47), two
105 additional sites close to Mondego river: Misarela (25) and Riba de Cima (25); and 5 individuals from
106 Buçaco (Figure 1, Table 1). Tissue samples were preserved in 70% ethanol. Whole genomic DNA was
107 extracted using QIA Quick DNEasy columns (Qiagen, Inc., Valencia, CA, USA) following standard DNA
108 extraction protocols. Seven microsatellite polymorphic loci (*CL5*, *CL6*, *CL17*, *CL19*, *CL39*, *CL136*, and
109 *CL145*) described by Sequeira et al. (2005) and sequences of the mitochondrial cytochrome b (cyt b), 700-
110 bp long fragment (Alexandrino et al. 2002), were chosen for analysis.

111 PCR amplification of cyt b and sequencing protocols are as in Alexandrino et al (2000). PCR products were
112 sequenced in both directions by using the PCR primers with the BigDye Terminator v3.1 Cycle sequencing
113 protocol (Applied Biosystems), an ABI Prism 3130XL Genetic Analyzer automated sequencer (Applied
114 Biosystems). Sequences were edited and aligned using BioEdit, version 7.2 (Hall, 1999). DNA samples
115 were amplified at seven microsatellite markers, accomplished with fluorescently labelled primers, using
116 multiplexed PCR and published protocols optimized by Sequeira et al (2005, 2008) with slight
117 modifications. PCR products were separated by capillary electrophoresis on an automatic sequencer
118 ABI3130xl Genetic Analyzer (AB Applied Biosystems). Fragments were scored against the GeneScan-500
119 LIZ Size Standard using the GENEMAPPER 4.1 (Applied Biosystems). To ensure no bias in allele sizing,
120 several samples previously analysed by Sequeira et al. (2008) were genotyped as control.

121 *Genetic diversity and population assignment*

122 For the mtDNA cyt b gene dataset, we used DnaSP 6.0 software (Rozas et al. 2017) to estimate diversity
123 parameters, including nucleotide diversity and haplotype diversity (h). For microsatellites, MICRO-
124 CHECKER 2.2.1 (van Oosterhout et al. 2004) was used to check amplified microsatellite genotypes for
125 large allele dropout, scoring errors due to stuttering and the presence of null alleles. Measures of genetic
126 diversity, including the mean number of alleles, the expected heterozygosity (H_e) and f estimator of FIS
127 per population, was estimated using GENETIX v.4.05 (Belkhir et al. 2000). Allelic Richness was estimated

128 using a rarefaction procedure implemented in HP-RARE 1.0 (Kalinowski 2005). The minimum number of
129 genes in analysed populations (Table 1, but see details below for total dataset and Sequeira et al. 2008) was
130 32 (16 genotypes), so this was used as a basis for rarefaction. To test for linkage disequilibrium (LD) and
131 departures from Hardy-Weinberg equilibrium (HWE) among all pairs of loci in each population, we used
132 GENEPOP 3.3 (Raymond and Rousset 1995). All probability tests were based on MCMC simulations (Guo
133 and Thompson 1992; Raymond and Rousset 1995) using default values, with significance levels adjusted
134 for multiple tests using sequential Bonferroni corrections to minimize type I errors ($\alpha = 0.05$; Rice 1989).
135 BOTTLENECK 1.2.02 (Piry et al. 1999) was used to investigate signatures of a recent reduction in effective
136 population size of the (introduced) Sintra population by using the mode-shift test and one-tailed Wilcoxon
137 signed-rank test (Cornuet and Luikart 1996), under a two-phase model (TPM; Di Rienzo et al. 1994) with
138 90% stepwise mutation and 10000 iterations.

139 To accomplish the main goals of this study (elucidate the origin and genetic diversity of the introduced
140 population – Sintra), for downstream analyses we used mtDNA (Alexandrino et al. 2002) and microsatellite
141 data (Sequeira et al. 2008) that had been analysed previously (Figure 1; Table 1). Microsatellite data
142 consisted of allele frequencies (total of 96 alleles) at the same seven polymorphic loci scored for 286 (16
143 to 27 individuals/location) salamanders (Sequeira et al. 2008). MtDNA data consisted of 120 sequences (2
144 to 12 individuals/location) of a cyt b fragment (30 distinct haplotypes; GenBank accession numbers:
145 AF329285-AF329314), sampled from the same 13 locations as used for microsatellite analyses by
146 Alexandrino et al. (2002). Altogether, a total dataset composed of 140 cyt b sequences and 375 individuals
147 genotyped at seven microsatellite loci, from 16 populations distributed across the entire species' range,
148 were analysed in this study (Table 1; Supplementary Table S1). Intraspecific relationships of the cyt b
149 haplotypes were inferred by a median-joining network using the NETWORK software v. 5.0.0.1 (Bandelt
150 et al. 1999). Pairwise genetic differentiation between populations based on microsatellite data was evaluated
151 using Weir and Cockerham's unbiased F-statistics (F_{ST}) (Weir and Cockerham 1984) and Shared Allele
152 Distance (DAS; Chakraborty and Jin 1993) using Populations 1.2.31 (Langella 1999). To infer relationship
153 among individuals and populations, a phylogenetic tree was reconstructed by the Neighbor-Joining method

154 using the Populations software. The Tree was based on the DAS with 1000 bootstraps, and was viewed and
155 edited in Mega 11.0 (Kumar et al. 2018). We further assigned individuals to genetic groups using the
156 clustering-based approach STRUCTURE v.2.3.4 (Pritchard et al. 2000). We ran STRUCTURE with 5
157 replicates for each K value ranging from $K=1$ to $K=10$, with a burnin period of 100,000 and 500,000 steps
158 under the admixture model and uncorrelated allele frequencies. The optimal number of ancestral
159 populations was determined using Evanno's ΔK method (Evanno et al. 2005) as implemented in
160 STRUCTURE HARVESTER (Earl and von Holdt 2012), and to account for label switching between results
161 of different runs with same K , replicate runs were merged using the CLUMPAK (Kopelman et al. 2015) on
162 the webserver <http://clumpak.tau.ac.il>. Additionally, a Factorial Correspondence Analysis on the allelic
163 frequencies, as implemented in GENETIX 4.05, was used as a model-independent approach (i.e., free of
164 assumptions on the underlying population genetics model) to identify and describe clusters of genetically
165 related individuals.

166 *Effective population size and minimum number of founders*

167 We estimated contemporary effective population size (N_e) with a single temporal sample through the
168 linkage disequilibrium (LD) method with Jackknifing implemented in NeEstimator 2.1 (Do et al. 2014). We
169 also estimated N_e based on theoretical predictions of the relationship between genetic drift and population
170 size (see Hendrick 2000 and references therein). Based on the assumption that heterozygosity decreases
171 approximately at a rate of $1/(2N_e)$ per generation, the effective size of a population over time can be
172 calculated using the equation (Hedrick 2000): $N_e = 1 / [2 * (1 - H^{1/t})]$, where H is the ratio (H_T/H_E) of the
173 expected heterozygosity found in Sintra (H_T) to that expected (H_E) in the source population, and t is the
174 number of generations. For calculations we used a generation time estimate of four years (Lima et al. 2001),
175 and thus, a T of 18 generations, assuming that population was founded 72 years ago (Seabra 1943). For the
176 calculations we used Buçaco as the most likely source population, but because populations around Buçaco
177 and Mondego river (SA, BU, VA, MI and RC) group together, we also used H_E averaged across those
178 populations (see Results).

179 To obtain a minimum estimate of the founder population, we used an approach based on Rasner et
180 al. (2004), which requires empirical information from microsatellite alleles observed in the present
181 population. We simulated in R 4.0.2 (R Core Team 2020), via a custom written script (see Fisher et al.
182 2015), genetic profiles of the founder individuals by randomly sampling alleles (10,000 replicates)
183 independently for each locus, and without replacement from the genetic profiles. The smallest number of
184 founders consistent with the observed data (i.e. smallest number of individuals containing all the alleles
185 identified at the seven microsatellites), was taken to be the number that gave a p-value > 0.05 of capturing
186 the observed alleles.

187 **Results**

188 Sequencing of the mtDNA cyt b fragment (700 bp long) performed on 25 samples from the three newly
189 sampled populations and additional five individuals from Buçaco, uncovered a total of four haplotypes.
190 Considering the published data (Alexandrino et al. 2002), three haplotypes (h31-33; XXX-XXX GenBank
191 accession no. upon acceptance) were newly described (Fig. 2; Supplementary Table S3). Samples from
192 Sintra were fixed for the newly described haplotype h31. This haplotype was only detected among
193 populations close to Mondego river (RC and MI) and Buçaco (Fig. 2; Supplementary Table S3). The
194 haplotype network showed that the newly described haplotypes, including the one observed in Sintra,
195 belong to *C. lusitanica longipes*, the subspecies that occurs in southern part of *C. lusitanica* range, south of
196 Buçaco population (Fig. 2).

197 We successfully genotyped 85 individuals at all seven microsatellite markers. No evidence of
198 scoring errors due to stuttering or large allele dropout was found. For the analysed populations and across
199 all loci, there were no significant deviations from Hardy-Weinberg or linkage equilibrium. STRUCTURE
200 analysis revealed that the most likely number of genetically distinct clusters of *C. lusitanica* is four ($K = 4$;
201 Fig. 3 and Supplementary Fig. S1), albeit a second clear peak at $K = 5$ was also apparent (Supplementary
202 Fig. S1). At $K = 3$, Sintra population clustered with populations located in between Douro and Mondego
203 rivers, while populations south of Mondego River and the ones north of Douro river represent the other two

204 clusters. At $K \geq 4$, Sintra formed a separate population cluster (Supplementary Fig. S1). At $K = 4-5$, BU is
205 the population with higher average levels of proportion of assignment to Sintra cluster (12.4 -14.0%,
206 respectively; Supplementary Fig. S1 and Table S4). In addition, individuals partly admixed (membership
207 proportion $\geq 30\%$) with the Sintra cluster were only found in SA (2) and BU (5) populations. The FCA
208 analysis is in line with the STRUCTURE results, supporting the population allocation of four spatial groups
209 (Fig. 3). According to FCA, Sintra appears clearly separated from the extant groups, being closely related
210 with populations of Central Portugal between Douro and Mondego rivers (in between rivers), while
211 populations north of Douro and south of Mondego river, form independent groups. The microsatellite-based
212 phylogenetic tree showed similar results to those obtained with FCA and STRUCTURE. However, the
213 separation of Sintra from the other populations of Central Portugal (in between rivers) is not statistically
214 supported (Fig. 3). Pairwise F_{ST} and DAS genetic distance are in agreement with the genetic differentiation
215 of the introduced population (Sintra). Sintra is strongly differentiated ($F_{ST} > 0.25$; $DAS > 0.4$) from most
216 populations, except those close to Mondego river ($F_{ST} = 0.19-0.21$; $DAS = 0.29-0.31$) and, in particular
217 with Buçaco ($F_{ST} = 0.14$ and $DAS = 0.20$), from which Sintra appears as only moderately differentiated
218 (Supplementary Table S5). A summary of all pairwise F_{ST} s and DASs values is provided in Supplementary
219 Table S5.

220 For the Sintra population, the seven microsatellite loci yielded a total of 19 alleles, with a N_a of 2.7,
221 an A_R of 2.5, and H_E of 0.452 (Table 1, Supplementary Table S2). With exception of the four populations
222 north of Douro river, genetic diversity statistics in Sintra are consistently lower than those of the extant
223 populations. When compared (One sample t-test) to populations from Central Portugal (populations
224 phylogenetically most closely allied with Sintra, and therefore harboring the potential source population
225 from which founder individuals have been taken), all measures of genetic diversity in Sintra are
226 significantly lower (N_a : $t=17.269$, $df=6$, $p<0.001$; A_R : $t= 16.352$, $df=6$, $p<0.001$; H_E : $t=12.205$, $df=6$,
227 $p<0.001$).

228 Bottleneck tests indicated (using Wilcoxon tests of significance, and mode-shift of allele frequency
229 distributions) that Sintra population has undergone a significant size reduction (TPM model, $p = 0.023$;

230 shifted mode distribution of allele frequencies). Results from NeEstimator returned several infinite
231 estimates, both for point estimates and upper confidence limits (Supplementary Table S6). Despite of this,
232 our results showed that estimates for Sintra were relatively lower (≈ 24 ; 95% CI 10.5-77.8) than most of
233 the other populations from its native range, especially those close to Mondego river ($Ne \approx 182-997$, 95%
234 CI 25.3- ∞). Based on the average decline of the expected heterozygosity in Sintra relative to that in Buçaco
235 after 18 generations ($\approx 18\%$), we estimated a current $Ne \approx 25$, which is similar to that returned by LD
236 method implemented in NeEstimator. Using the decline of heterozygosity relative to the averaged value
237 (15%) of populations from Central Portugal around Buçaco and Mondego river, our estimate was slightly
238 higher ($Ne \approx 34$), albeit within the confidence interval returned by NeEstimator (Table S6). Based on
239 resampling technique of empirical data set, the effective number of founders was estimated to be of around
240 10-11 individuals (Supplementary Fig. S3).

241 **Discussion**

242 According to scarce historical available information, there was an old and dubious citation of the species
243 presence in the area at the end of the XIX century (Vieira 1886) and the report of relatively few individuals
244 of *C. lusitanica* collected at Buçaco mountains being deliberated introduced in Sintra at mid XX century
245 (Seabra, 1943). Our genetic analyses provide several evidences that support this documented human-
246 mediated introduction as the source of the present population. First, individuals from Sintra form a
247 geographically coherent genetic unit that are differentiated from the extant *C. lusitanica* populations.
248 Second, Sintra appears more closely related with populations in between Douro and Mondego rivers
249 (Central Portugal), instead of its geographically closest populations (southernmost), as would be expected
250 if Sintra were a natural population that has been kept isolated in an interglacial refuge (see example of the
251 Iberian north-western lacertid, *Lacerta schreiberi*, in Brito et al. 1986). Third, in line with theoretical
252 predictions and empirical studies of introduced populations (e.g. Nei et al. 1975; Allendorf and Lundquist
253 2003), Sintra presents reduced levels of genetic diversity when compared to native populations, particularly
254 to those of Central Portugal, which is consistent with the significant signs of population reduction returned

255 by the TPM model, and shifted mode distribution of allele frequencies in the bottleneck analysis. Finally,
256 in agreement with the hypothesis that Sintra population experienced a bottleneck at the founding event, the
257 reduction of genetic diversity is much more pronounced in number of alleles than in levels of heterozygosity
258 (Nei et al. 1975; Luikart et al. 1998; Dlugosch and Parker 2008). Sintra has 60% less of the allelic richness
259 and 15-18% less of heterozygosity when compared to its closely related populations (Central Portugal),
260 which agrees with the proportional losses of allelic richness *versus* heterozygosity (on average 5.1% higher
261 for allelic richness) reported by several studies of introduced populations (see review in Dlugosch and
262 Parker 2008). Besides, when compared to the northernmost populations (GE, PO, SAL), which correspond
263 to a recent range expansion and postglacial colonization of northern Iberia (Alexandrino et al., 2000;
264 Sequeira et al., 2008), Sintra has similar number of alleles but higher levels of heterozygosity. The rationale
265 is that allelic richness does not consider the frequency of the alleles but only their presence, being thus
266 particularly sensitive to the loss of rare alleles (as expected in founder events) that contribute little to
267 heterozygosity (Allendorf 1986; Spencer 2000, Leberg 2002; Greenbaum et al. 2014). This is particularly
268 evident when using microsatellite markers because a higher proportion of their alleles are normally at low
269 frequency (Dlugosch and Parker 2008).

270 Amongst the genetic cluster composed by the populations in between Douro and Mondego rivers,
271 which is the most closely related with Sintra, Buçaco is the population that presents the lower levels of
272 genetic differentiation (F_{ST} and DAS) and higher levels of admixture proportion with Sintra cluster (12.4 -
273 14.0%, as revealed by Structure analyses at $K = 4-5$, respectively; Supplementary Table S4). Despite these
274 signatures from nuclear data together with the presence of the mtDNA haplotype h31 (that is fixed in Sintra)
275 provide supporting evidence for Buçaco as the source population, overall we cannot definitively exclude
276 other neighboring closely related populations as alternative source. Indeed, the single mtDNA haplotype
277 observed in Sintra (h31) also occurs in two populations close to Mondego river (MI and RC; Fig. 1) and at
278 higher frequency (0.8-0.6, respectively) than observed in Buçaco (0.07). Furthermore, some individuals
279 from SA appeared partly assigned and clustered to Sintra according to STRUCTURE and FCA analyses,
280 respectively (Fig. 3). Finally, SA and other populations close to Mondego river (VA and MI) are also only

281 moderately differentiated from Sintra ($F_{ST}= 0.19-0.21$). It is worth mentioning that substantial loss of
282 genetic diversity and shifts in allele/haplotype frequencies are expected in small isolated populations
283 through the process of random genetic drift (Chakraborty and Nei 1977; Keller et al. 2012). The effects of
284 this process could even be more exacerbated in the case of mtDNA because its effective population size is
285 one-fourth of the one calculated from nuclear markers (e.g. Ballard and Whitlock 2004). In any case, the
286 expected changes in the genetic makeup of an introduced population after several generations of isolation
287 make difficult the identification of its source population, especially if founder individuals represented only
288 a fraction of the total amount of genetic variation (Keller et al. 2012). A visual inspection of alleles profile
289 among loci showed that Sintra has an overall lower number of alleles and a substantial shift in their
290 frequency. For instances, many alleles, independently of its frequency in between rivers cluster (putative
291 source populations), were not found in Sintra, and some low-frequency alleles present on that cluster
292 reached relatively high frequencies in Sintra (e.g. allele 119; locus C15), as expected under the effect of
293 random drift acting on a small founding population (e.g. Nei et al. 1975; Bartlett 1985). Other factor that
294 may affect the ability of identifying the source population is the level of genetic differentiation within
295 species' native populations (source area). Although significant, the level of differentiation between Buçaco
296 and its closely related neighboring populations, especially among SA, VA and MIS is relatively low ($F_{ST}=$
297 $0.014-0.034$). In spite of the limitations aforementioned, further work increasing the number of loci
298 involved and the intensity of sampling in the potential source area, especially in Buçaco population could
299 improve the accuracy in tracking the origin of Sintra.

300 The high proportion of heterozygosity retained in Sintra ($\sim 82\%$) is likely to reflect this population's
301 relatively recent isolation (Furlan et al. 2012), supposedly around 72 years ago (Seabra, 1943). Based on
302 the assumption that heterozygosity decreases approximately at a rate of $1/(2Ne)$ per generation (Hedrick
303 2000), a decline in heterozygosity of 18% after 18 generations (72 years) of genetic drift would correspond
304 to a current effective population size (Ne) of ≈ 25 , which is similar to the estimates returned by LD method
305 implemented in NeEstimator (24.3; 95% CI 8.5-125.7). Besides, using our Ne estimate together with the
306 previous estimate of population size ($N \approx 340$) from a capture-mark-recapture study in Sintra (Aguilar et

307 al. 2018), we found a ratio of $Ne/N \approx 0.1$, which is within estimates found for other amphibians (Frankham
308 1995; Jehle et al. 2001; Álvarez et al. 2015).

309 Depending on the life-history traits of the organism and on other biotic and abiotic factors, patterns
310 of genetic variation may be strongly affected by different selective regimes and/or demographic
311 stochasticity (e.g. Roderick and Navajas 2003). So, estimating the number of founders could be difficult
312 when information on demographic history of the population is missing. For example, when introduced to
313 new areas, some species' populations often grow to a large population size after only few generations or
314 remains small for several generations before sudden, rapid growth. In other cases, the introduced population
315 size may be continuously small and stable (Monnet et al. 1993; Crooks and Soulé 1999; Ficetola et al. 2008;
316 Sendell-Price et al. 2020). Although the number of individuals of *C. lusitanica* that were introduced in
317 Sintra is unknown, based on the Seabra's writings (Seabra 1943) and the pronounced reduction of genetic
318 diversity observed in the present study, there are reasons to expect that the founding population was
319 relatively small. Actually, our estimate based on genotypes simulated by resampling all alleles across the
320 entire microsatellite dataset, indicated that Sintra population may have resulted from the introduction of about
321 10-11 individuals (Supplementary Fig. S2). This estimate, however, should be treated as a proxy of the
322 effective number of founders, which may be lower than the size of the founding population.

323 Estimates of current population size ($N \approx 340$) and density (3.2 individuals per m of brook) of the
324 introduced population by 2015/2016 (Aguilar et al. 2018) are lower than estimates (size $\approx 1250-2200$ and
325 density 11-17 per m of brook) for local populations among the native range (Arntzen 1981; Arntzen et al.
326 2015; Teixeira et al. 1998; Lima 1995), suggesting that present-day Sintra population is facing ecological
327 limitations. According to Aguilar et al. (2018), salamanders were restricted to a very small site, being
328 confined to a stretch of ≈ 100 m along the margins of one small stream and its tiny tributaries. Although
329 the exact location of the introduction was not referred by Seabra (1943), it is possible that the current
330 population is limited to the vicinity of the release site and may have reached carrying capacity. Despite
331 documented movements over distances of 700 m along the wet banks of streams (Arntzen 1981, 1984) and
332 the species' propensity for dispersal by larval drift (Arntzen 1995; Thiesmeier 1994), displacements of this

333 species outside moist habitats are severely restricted (Arntzen 1995). According to Aguilar et al. (2018),
334 considering both microclimate and vegetation cover, the habitat surrounding the stream occupied by this
335 population is not suitable for *C. lusitanica*. This, together with the fact that *C. lusitanica* presents relatively
336 “slow” life history traits (*sensu* Allen et al. 2017), including low fecundity (average clutch size = 18;
337 Sequeira et al., 2003) and long reproductive lifespan (age of sexual maturity = 4 years; generation time = 4
338 years; and, longevity = 10 years; Lima et al., 2001), may hamper the species to expand its range into other
339 waterways and thus increase the total population size. Together with possible scenarios of climate change
340 or stochastic local extinctions due to e.g. increasing forest fires, the long-term survival or expansion
341 prospects of this introduced rear edge population of an ecologically demanding species may be in risk;
342 however, its maintenance for more than 70 years without any human-assisted management highlights the
343 potential of amphibian reintroductions or assisted migrations as effective conservation tools.

344 **Acknowledgements**

345 The permission of capture and manipulation were provided by the Institute for Nature Conservation and
346 Forests (ICNF). This work was funded by FEDER funds through the Operational Programme for
347 Competitiveness Factors-COMPETE and by National Funds through FCT-Foundation for Science and
348 Technology under the UIDB/50027/2020 and PEst-OE/BIA/UI0329/2014. F. Aguilar was supported by a
349 PhD grant from FCT (2020. 06142.BD). J Teixeira is financed by national funds through FCT.

350 **Data availability**

351 Data generated and analysed in this study are included in the Supplementary Information file and
352 mitochondrial cytochrome b sequences are available in the GenBank repository (Accession Nos.
353 XXXXX– XXXXX).

354 **REFERENCES:**

- 355 Aguilar, F.F.; Madeira, F.M.; Crespo, E.; Rebelo, R (2018). Rediscovery of the Golden-Striped Salamander
356 *Chioglossa lusitanica* of Sintra, Portugal. *Herpetolog J* 2018, 28, 148–154.
- 357 Alexandrino, J., Arntzen, J. W., Ferrand, N. (2002). Nested clade analysis and the genetic evidence for
358 population expansion in the phylogeography of the golden-striped salamander, *Chioglossa lusitanica*
359 (Amphibia: Urodela). *Heredity*, 88, 66–74.

360 Alexandrino, J., Froufe, E., Arntzen, J. W., Ferrand, N. (2000). Genetic subdivision, glacial refugia and
361 postglacial recolonization in the golden-striped salamander, *Chioglossa lusitanica* (Amphibia: Urodela).
362 Mol Ecol, 9, 771–781.

363 Allen, W.L., Street, S.E. and Capellini, I. (2017), Fast life history traits promote invasion success in
364 amphibians and reptiles. Ecol Lett, 20: 222-230.

365 Allendorf FW (1986) Genetic drift and the loss of alleles versus heterozygosity. Zoo Biol 5: 181–190.

366 Allendorf, F. W. and Lundquist, L. L. (2003). Introduction: Population biology, evolution, and control of
367 invasive species. Conserv Biol 17, 24–30.

368 Alvarez D, Lourenco A, Oro D, Velo-Anton G (2015) Assessment of census (N) and effective population
369 size (Ne) reveals consistency of Ne single-sample estimators and a high Ne/N ratio in an urban and isolated
370 population of fire salamanders. Conservation Genetics Resources, 7, 705–712.

371 Arntzen, J. W. (1981). Ecological observations on *Chioglossa lusitanica* (Caudata, Salamandridae).
372 Amphibia-Reptilia, 1, 187–203.

373 Arntzen, J. W. (1984). Speedy salamanders: Sedentariness and migration of *Chioglossa lusitanica*. Revista
374 Española de Herpetología, 8, 81–86.

375 Arntzen, J. W. (1995). Temporal and spatial distribution of the golden-striped salamander (*Chioglossa*
376 *lusitanica*) along two mountain brooks in northern Portugal. Herpetolog J, 5, 213.

377 Arntzen, J. W. (1999). *Chioglossa lusitanica* Bocage 1864 - der Goldstreifensalamander. In: Handbuch der
378 Reptilien und Amphibien Europas. Pp. 301-321, Grossenbacher, K. and Thiesmeier, B. (eds.). Germany:
379 Akademische Verlagsgesellschaft.

380 Arntzen, J.W. (2015). Drastic population size change in two populations of the Golden Striped Salamander
381 over a forty year period – Are Eucalypt plantations to blame? Diversity 7, 270–294.

382 Arntzen J.W., Teixeira J. (2006). History and new developments in the mapping and modelling of the
383 distribution of the golden-striped salamander, *Chioglossa lusitanica*. Zeitschrift für Feldherpetologie, 10:
384 113–126.

385 Ballard JWO, Whitlock M (2004) The incomplete natural history of mitochondria. Mol Ecol, 13, 729– 744.

386 Bandelt, H.J., Forster, P., Röhl, A., 1999. Median-joining networks for inferring intraspecific phylogenies.
387 Mol Biol Evol 16, 37– 48.

388 Belkhir K, Borsa P, Chikhi L, Bonhomme F, Raufast N. 2000. GENETIX 4.01 logiciel sous Windows™
389 pour la génétique des populations. Laboratoire Génome, populations, Interactions, CNRS UMR 5000
390 Université de Montpellier II, Montpellier, France

391 Bocage, B. (1864). Notice sur un Batracien nouveau du Portugal (*Chioglossa lusitanica*, Nob.). Proceedings
392 of the Zoological of London 1864, 264–265.

- 393 Brito, J.C., Abreu, F.E., Paulo, O.S., Rosa, H.D., et al. (1996). Distribution of Schreiber's green lizard
394 (*Lacerta schreiberi*) in Portugal: a predictive model. *Herpetolog J* 6, 43–47.
- 395 Button, S., & Borzée, A. (2021). An integrative synthesis to global amphibian conservation priorities.
396 *Global Change Biology*, 27, 4516–4529.
- 397 Ceballos, G., Ehrlich, P. R., & Raven, P. H. (2020). Vertebrates on the brink as indicators of biological
398 annihilation and the sixth mass extinction. *Proceedings of the National Academy of Sciences of the United*
399 *States of America*, 117(24), 13596–13602.
- 400 Chakraborty R, Jin L (1993) A unified approach to study hypervariable polymorphisms: Statistical
401 considerations of determining relatedness and population distances. In: *DNA Fingerprinting: State of the*
402 *Science* (eds Pena SDJ, Chakraborty R, Epplen JT, Jeffreys AJ), pp. 163–175. Birkhauser Verlag Basel,
403 Switzerland.
- 404 Chakraborty, R. and Nei, M. (1977). Bottleneck effects on average heterozygosity and genetic distance with
405 the stepwise mutation model. *Evolution*, 31: 347-356.
- 406 Clavero, M., Nores, C., Kubersky-Piredda, S., & Centeno-Cuadros, A. (2016). Interdisciplinarity to
407 reconstruct historical introductions: solving the status of cryptogenic crayfish. *Biological Reviews*, 91,
408 1036–1049.
- 409 Cornuet JM, Luikart G (1996) Description and power analysis of two tests for detecting recent population
410 bottlenecks from allele frequency data. *Genetics* 144:2001–2014.
- 411 Crooks, J. A., and M. E. Soule. (1999). Lag times in population explosions of invasive species: causes and
412 implications. Pages 103–126 in O. T. Sandlund et al., editors. *Invasive species and biodiversity*
413 *management*. Kluwer, New York.
- 414 Do, C., Waples, R. S., Peel, D., Macbeth, G. M., Tillett, B. J. and Ovenden, J. R. (2014) NeEstimator V2:
415 re-implementation of software for the estimation of contemporary effective population size (Ne) from
416 genetic data. *Molecular Ecology Resources*, 14, 209-214.
- 417 Dlugosch KM, Parker IM (2008) Founding events in species invasions: genetic variation, adaptive
418 evolution, and the role of multiple introductions. *Molecular Ecology*, 17, 431–449.
- 419 Di Rienzo A, Peterson AC, Garza JC et al (1994) Mutational processes of simple-sequence repeat loci in
420 human populations. *Proc Natl Acad Sci U S A* 91(8):3166–3170
- 421 Earl, D. A., & vonHoldt, B. M. (2012). STRUCTURE HARVESTER: A website and program for
422 visualizing STRUCTURE output and implementing the Evanno method. *Conservation Genetic Resources*,
423 4, 359– 361
- 424 Evanno, G., Regnaut, S., & Goudet, J. (2005). Detecting the number of clusters of individuals using the
425 software STRUCTURE: A simulation study. *Mol Ecol*, 14, 2611– 2620.
- 426 Ficetola, G.F., Bonin, A. and Miaud, C. (2008), Population genetics reveals origin and number of founders
427 in a biological invasion. *Mol Ecol*, 17: 773-782.

428 Fischer ML, Hochkirch A, Heddergott M, Schulze C, Anheyer-Behmenburg HE, Lang J, et al. (2015).
429 Historical Invasion Records Can Be Misleading: Genetic Evidence for Multiple Introductions of Invasive
430 Raccoons (*Procyon lotor*) in Germany. PLoS ONE 10(5): e0125441.

431 Furlan E, Stoklosa J, Griffiths J et al. (2012) Small population size and extremely low levels of genetic
432 diversity in island populations of the platypus, *Ornithorhynchus anatinus*. Ecol Evol, 2, 844–857.

433 Gippoliti, S. & Amori, G. (2006). Ancient introductions of mammals in the Mediterranean Basin and their
434 implications for conservation. Mammal Review 36, 37-48.

435 Greenbaum G, Templeton AR, Zarmi Y, Bar-David S (2014) Allelic Richness following Population
436 Founding Events – A Stochastic Modeling Framework Incorporating Gene Flow and Genetic Drift. PLoS
437 ONE 9(12): e115203.

438 Guo SW, Thompson EA. (1992). Performing the exact test of Hardy-Weinberg proportions for multiple
439 alleles. Biometrics 48: 361–372.

440 Hall TA. (1999). BioEdit: a user-friendly biological sequence alignment editor and analysis program for
441 Windows 95/98/ NT. Nucleic Acids Symposium Series 41: 95–98.

442 Hedrick, P.W. (2000) Genetics of Populations. 2nd Edition, Jones and Bartlett Publishers, Sudbury, MA.

443 IUCN. 2022. The IUCN Red List of Threatened Species. Version 2022-1. <https://www.iucnredlist.org>.
444 Accessed on 8/9/2022.

445 Jehle, R., Arntzen, J.W., Burke, T., Krupa, A. and Hödl, W. (2001). The annual number of breeding adults
446 and the effective population size of syntopic newts (*Triturus cristatus*, *T. marmoratus*). Mol Ecol 10: 839–
447 850.

448 Leberg, P.L. (2002), Estimating allelic richness: Effects of sample size and bottlenecks. Molecular Ecology,
449 11: 2445-2449.

450 Kalinowski, S. T. (2005). HP-Rare 1.0: A computer program for performing rarefaction on measures of
451 allelic richness. Molecular Ecology Notes, 5, 187– 189.

452 Kehlmaier, C., Zinenko, O., & Fritz, U. (2020). The enigmatic Crimean green lizard (*Lacerta viridis*
453 *magnifica*) is extinct but not valid: Mitogenomics of a 120-year-old museum specimen reveals historical
454 introduction. Journal of Zoological Systematics and Evolutionary Research, 58(1), 303–307.
455 <https://doi.org/10.1111/jzs.12345>

456 Kopelman, N M, Mayzel J, Jakobsson M, Rosenberg N A., and Mayrose I. (2015). "Clumpak: a program
457 for identifying clustering modes and packaging population structure inferences across K. Molecular ecology
458 resources 15, 1179-1191.

459 Kraaijeveld-Smit, F. J. L., Griffiths, R. A., Moore, R. D., & Beebee, T. J. C. (2006). Captive breeding and
460 the fitness of reintroduced species: A test of the responses to predators in a threatened amphibian. J Appl
461 Ecol, 43(2), 360–365.

462 Kumar S, Stecher G, Li M, Knyaz C, and Tamura K (2018) MEGA X: Molecular Evolutionary Genetics
463 Analysis across computing platforms. *Molecular Biology and Evolution* 35:1547-1549.

464 Keller SR, Gilbert KJ, Fields PD, Taylor DR (2012). Bayesian inference of a complex invasion history
465 revealed by nuclear and chloroplast genetic diversity in the colonizing plant, *Silene latifolia*. *Mol Ecol* 21:
466 4721–4734.

467 Roderick GK, Navajas M (2003) Genes in new environments: genetics and evolution in biological control.
468 *Nature Reviews Genetics*, 4, 889–899.

469 Langella O. (1999). Populations 1.2.31. <http://bioinformatics.org/~tryphon/populations/>.

470 Lima, V. (1995). Estudo Comparativo de Alguns Aspectos da Biologia de *Chioglossa lusitanica* em Duas
471 Populações do Noroeste de Portugal. Master's Thesis, Faculty of Sciences, University of Porto, Porto,
472 Portugal. (In Portuguese)

473 Lima, V., Arntzen, J. W., and Ferrand, N. (2001). Age structure and growth pattern in two populations of
474 the golden-striped salamander *Chioglossa lusitanica* (Caudata, Salamandridae). *Amphibia-Reptilia*, 22,
475 55–68.

476 Luikart G, Allendorf F, Cornuet J, Sherwin WB (1998) Distortion of allele frequency distributions
477 provides a test for recent population bottlenecks. *J Hered* 89: 238–247.

478 Monnet C, Thibault JC, Varney A (1993) Stability and changes during the twentieth century in the breeding
479 landbirds of Tahiti (Polynesia). *Bird Conserv Int* 3:261–280

480 Nei, M., Maruyama, T. and Chakraborty, R. (1975) The bottleneck effect and genetic variability in
481 populations. *Evolution* 29, 1–10 .

482 Prichard, J. K., Stephens, M., & Donnelly, P. (2000). Inference of population structure using multilocus
483 genotype data. *Genetics*, 155, 945– 959.

484 Piry S, Luikart G, Cornuet JM (1999) BOTTLENECK: A computer program for detecting recent reductions
485 in the effective population size using allele frequency data. *J Hered* 90(4):502–503.

486 Rasner CA, Yeh P, Eggert LS, Hunt KE, Woodruff DS, Price TD. (2004) Genetic and morphological
487 evolution following a founder event in the dark-eyed junco, *Junco hyemalis thurberi*. *Mol Ecol* 13:
488 671–681.

489 Raymond, M., & Rousset, F. (1995). GENEPOP (version 1.2): Population genetics software for exact tests
490 and ecumenicism. *Journal of Heredity*, 86, 248– 249.

491 R Core Team (2020). R: A language and environment for statistical computing. R Foundation for Statistical
492 Computing, Vienna, Austria. URL <https://www.R-project.org/>.

493 Rice WR. 1989. Analyzing tables of statistical tests. *Evolution* 43: 223–225

494 Rozas J, Ferrer-Mata A, Sánchez-DelBarrio JC, Guirao-Rico S, Librado P, Ramos-Onsins SE, Sánchez-
495 Gracia A. (2017). DnaSP 6: DNA sequence polymorphism analysis of large data sets. *Molecular Biology*
496 *and Evolution*, 34(12):3299–3302

- 497 Sakai, A. K., Allendorf, F. W., Holt, J. S., Lodge, D. M., Molofsky, J., With, K. A., ... Weller, S. G. (2001).
498 The population biology of invasive species. *Annual Review of Ecology and Systematics*, 32, 305– 332.
- 499 Sequeira F, Alexandrino J, Weiss S, Ferrand N (2008) Documenting the advantages and limitations of
500 different classes of molecular markers in a well-established phylogeographic context: lessons from the
501 Iberian Endemic Golden-striped salamander, *Chioglossa lusitanica* (Caudata: Salamandridae). *Biological*
502 *Journal of the Linnean Society*, 95, 371–387.
- 503 Sequeira, F. and Alexandrino, J. (2008). *Chioglossa lusitanica* Barbosa du Bocage, 1864. Pp. 92-93 in:
504 Loureiro, A., Ferrand de Almeida, N., Carretero, M. A. and Paulo, O. S. (eds.), *Atlas dos Anfíbios e Répteis*
505 *de Portugal*. Instituto da Conservação da Natureza e da Biodiversidade. Lisboa, Portugal.
- 506 Sequeira, F., Ferrand, N. and Crespo, E.G. (2003). Reproductive cycle of the golden-striped salamander
507 *Chioglossa lusitanica* (Caudata, Salamandridae) in NW Portugal. *Amphibia-Reptilia* 24, 1–12.
- 508 Sequeira F, Rocha S, Ferrand N, Weiss S. (2005b). Isolation and characterization of seven microsatellite
509 loci in *Chioglossa lusitanica* (Urodela: Salamandridae). *Molecular Ecology Notes* 5: 212–214.
- 510 Seabra, A.F. (1943). Apontamentos sobre a fauna do Algarve Vertebrados. *Memórias e Estudos do Museu*
511 *Zoológico da Universidade de Coimbra* 147, 1–18.
- 512 Spencer CC, Neigel JE, Leberg PL (2000) Experimental evaluation of the usefulness of microsatellite DNA
513 for detecting demographic bottlenecks. *Mol Ecol*, 9, 1517– 1528.
- 514 Sendell-Price, A.T., Ruegg, K.C. and Clegg, S.M. (2020). Rapid morphological divergence following a
515 human-mediated introduction: the role of drift and directional selection. *Heredity* 124, 535–549.
- 516 Simberloff, D. 2009. The role of propagule pressure in biological invasions. *Annual Reviews in Ecology,*
517 *Evolution, and Systematics*, 40:81–102.
- 518 Teixeira J, Sequeira F, Alexandrino J and Ferrand N (1998). Bases para a Conservação da Salamandra
519 lusitânica *Chioglossa lusitanica*. *Estudos de Biologia e Conservação da Natureza* nº 24. Instituto da
520 Conservação da Natureza, Lisboa.
- 521 Teixeira J, Ferrand N, Arntzen JW. 2001. Biogeography of the golden-striped salamander *Chioglossa*
522 *lusitanica*: a field survey and spatial modelling approach. *Ecography* 24: 618– 624.
- 523 Tiberti, R., & Splendiani, A. (2019). Management of a highly unlikely native fish: The case of arctic charr
524 *Salvelinus alpinus* from the Southern Alps. *Aquatic Conservation: Marine and Freshwater Ecosystems*,
525 29(2), 312–320. <https://doi.org/10.1002/aqc.3027>
- 526 Van Oosterhout, C., Hutchinson, W.F., Wills, D.P.M. and Shipley, P. (2004), micro-checker: software for
527 identifying and correcting genotyping errors in microsatellite data. *Molecular Ecology Notes*, 4: 535-538.
- 528 Vieira, A.X.L. (1886): *Catalogo dos Amphibios e Reptis de Portugal existentes actualmente no Museo*
529 *Zoologico da Universidade de Coimbra. Relatório do Professor de Zoologia, 1885–1886. Anuario da*
530 *Universidade de Coimbra, 1886–1887. – Imprensa da Universidade, Coimbra.*

531 Weir BS, Cockerham CC. (1984). Estimating F-statistics for the analysis of population structure. *Evolution*
532 38: 1358–1370.

Table 1 - Genetic variation at seven microsatellite loci and mitochondrial Cyt b in 16 *Chioglossa lusitanica* populations. Total number of analyzed samples (N), number of alleles per locus (N_a), allelic richness (A_R), expected heterozygosity (H_E), fixation index (F_{IS}), number of mitochondrial Cyt b haplotypes (N_H), haplotype diversity (h) and nucleotide diversity (π) for each sampled population. Population codes are as in Fig.1.

Population ID/code	Lat.	Long.	Microsatellites					Mitochondrial cytb				Reference
			N	H_E	N_a	A_R	F_{IS}	N	N_H	h	π	
Salas (SAL)	43.394	-6.256	24	0.286	2.9	2.7	-0.039	6	1	0	0.0000	1, 2
Pontevedra (PO)	42.502	-8.482	16	0.273	3.6	3.6	-0.048	9	2	0.286	0.0004	1, 2
Gerês (GE)	41.757	-8.146	20	0.286	2.9	2.8	0.073	10	2	0.222	0.0003	1, 2
Valongo (VAL)	41.179	-8.49	22	0.474	4.7	4.5	0.051	9	3	0.556	0.0020	1, 2
Montemuro (MO)	41.043	-8.066	20	0.553	5.6	5.2	-0.033	10	4	0.533	0.0324	1, 2
Covelo (CO)	40.777	-8.213	25	0.569	6.6	5.9	0.015	12	4	0.561	0.0012	1, 2
Saíde (SA)	40.446	-8.324	21	0.578	7.3	6.6	0.1	4	2	0.5	0.0007	1, 2
Buçaco (BU)	40.377	-8.367	22	0.634	6.9	6.5	0.053	14	5	0.758	0.0088	1, 2, this study
Várzea (VA)	40.248	-8.375	22	0.594	7.0	6.5	0.028	10	5	0.756	0.0026	1, 2
Misarela (MI)	40.218	-8.358	26	0.619	7.6	7.0	0.028	5	2	0.4	0.0011	this study
Riba de Cima (RC)	40.259	-8.236	24	0.587	6.9	6.9	-0.055	5	3	0.7	0.0017	this study
Lousã (LO)	40.114	-8.224	26	0.756	8.4	7.7	0.043	11	3	0.655	0.0027	1, 2
Castanheira Pêra (CP)	40.091	-8.201	22	0.712	7.4	6.9	0.012	2	1	0	0.0000	1, 2
Açor (AC)	40.221	-7.919	18	0.640	6.9	6.6	-0.053	10	5	0.756	0.0023	1, 2
Muradal (MU)	40.007	-7.697	21	0.579	5.1	4.8	-0.068	10	3	0.378	0.0006	1, 2
Sintra (SI)	38.796	-9.424	47	0.452	2.7	2.5	0.049	10	1	0	0.0000	this study

1 – Alexandrino et al. (2002)

2 – Sequeira et al. (2008)

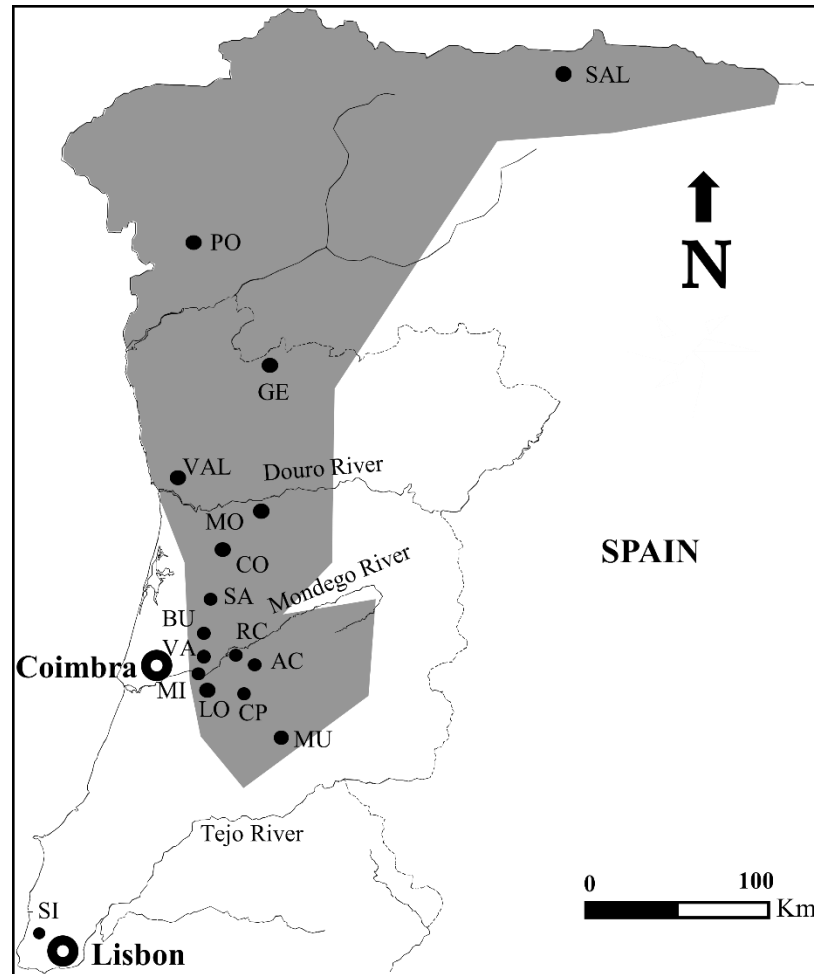


Figure 1. Map showing the distribution of *Chioglossa lusitanica* in the Iberian Peninsula (grey shading; Arntzen 1999) and sampling localities as identified in Table 1.

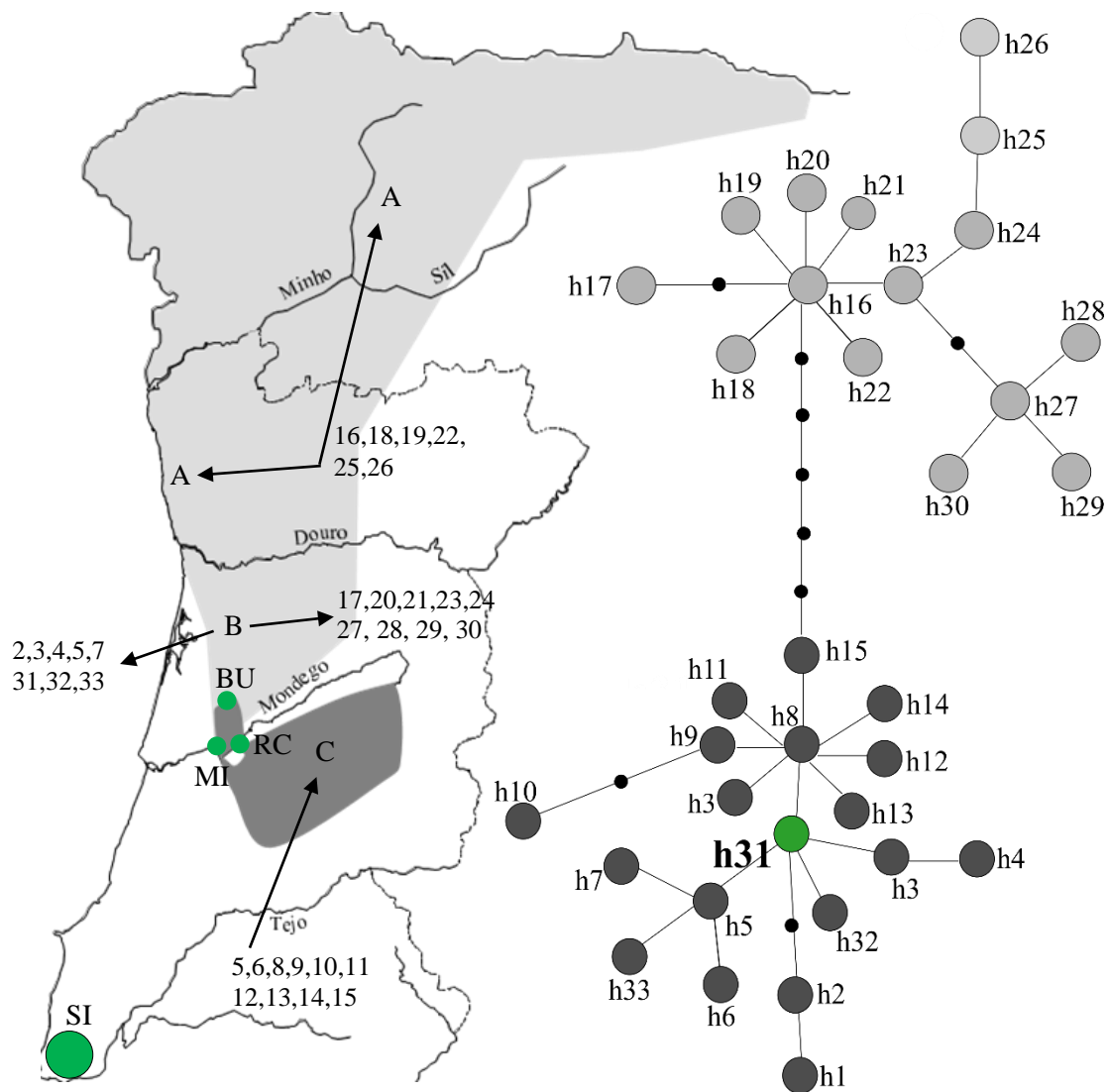


Figure 2. Median-joining network and geographic distribution of cytochrome b haplotypes observed in *Chioglossa lusitanica*. Each circle represents a specific haplotype: h1-h30, described by Alexandrino et al. (2002); and, haplotypes h31-h33, newly described in the present study. Black dots represent hypothetical undetected haplotypes, and each line represents one nucleotide substitution. The light gray haplotypes correspond to the subspecies *C. lusitanica longipes*, and the dark gray haplotypes are from *C. l. lusitanica* (Alexandrino et al., 2002; Arntzen et al. (2007)). The green circles in the map correspond to populations that share the haplotype h31 (the haplotype found in Sintra). Numbers in the map correspond to the geographic distribution of haplotype as shown in the network. Letters across *C. lusitanica* range distribution, corresponds to: A, north of Douro river; B, between Douro and Mondego rivers; and; C, south of Mondego river (see Supplementary Table S3 for detailed information on haplotype frequency across populations and identification of haplotype numbers).

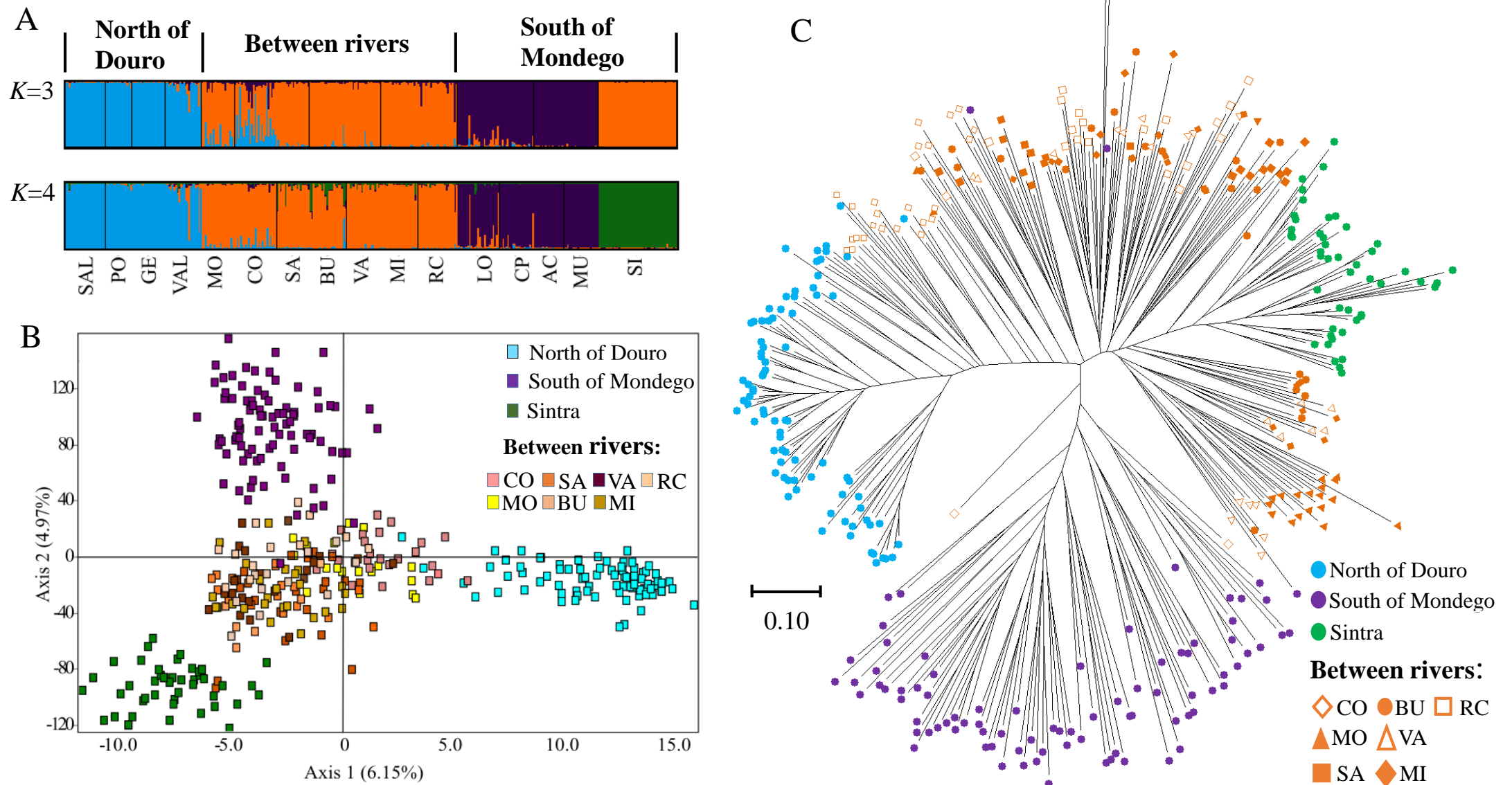
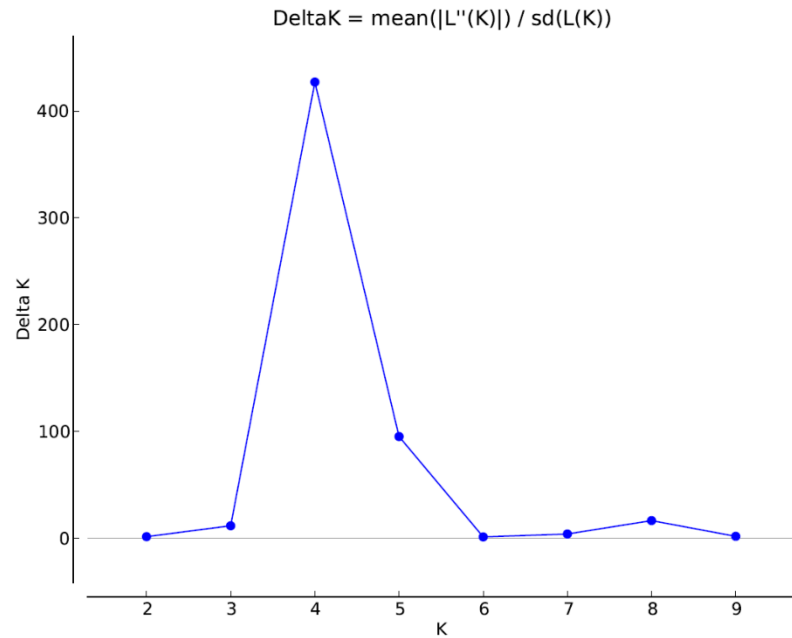


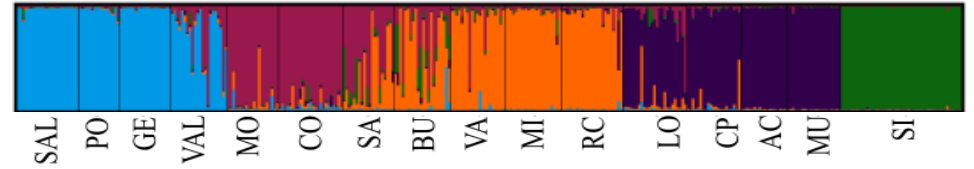
Figure 3. Analyses based on microsatellite genotypes at seven loci for 286 *Chioglossa lusitanica* individuals: A) Bayesian clustering results of STRUCTURE. Each vertical bar represents one individual and its assignment proportion into one of the three ($K=3$) or four ($K=4$) clusters. A black line separates individuals of different populations. These are labelled below the figure (as identified in the Table 1) and are sorted from north to south (from the left to the right of the figure). Population localisation according to geographical region is indicated on the top of the figure. B) Factorial Correspondence Analysis (FCA) of population multilocus scores computed using GENETIX 4.0543. Multilocus scores are computed in the bivariate space defined by the first two factorial components.. C) Neighbor-Joining tree based on allele-sharing distance (DAS) using Populations 1.2.31 software.

A

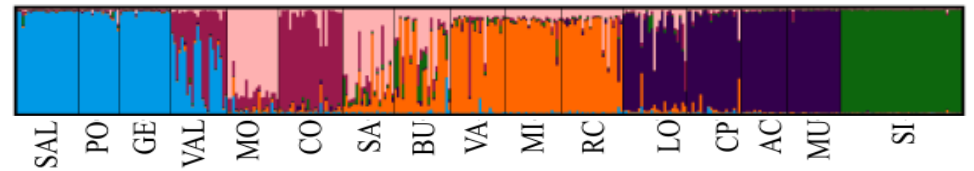


B

K=5



K=6



K=7

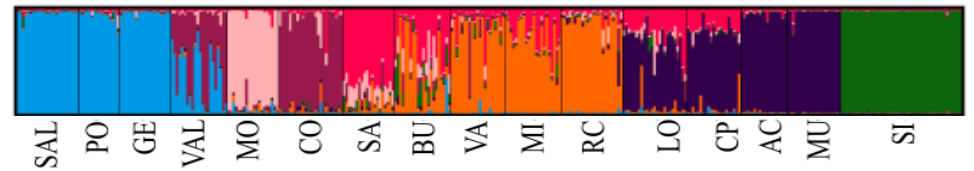


Figure S1. Analyses based on microsatellite genotypes at seven loci for 375 *Chioglossa lusitanica* individuals: A) Delta K values estimated according to Evanno et al. 2005 Method. B) Bayesian clustering results of STRUCTURE for ($K=5-7$). Each vertical bar represents one individual and its assignment proportion into one of the clusters. A black line separates individuals of different populations. These are labelled below the figure (as identified in the Table 1) and are sorted from north to south (from the left to the right of the figure).

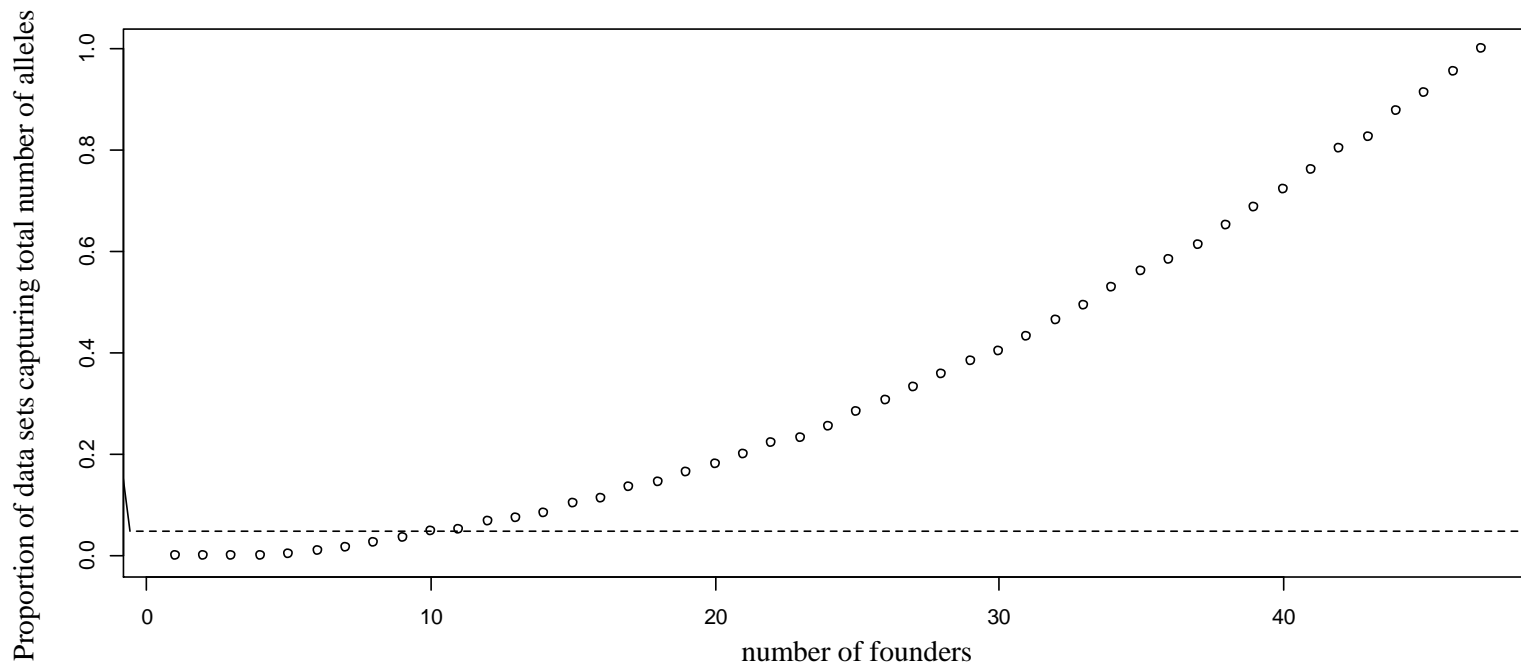


Figure S2. Minimum number of founders consistent with the observed data (smallest number of individuals containing all the alleles identified at the seven microsatellites), taken to be the number that gave a probability > 0.05 of capturing the observed alleles (dotted line).

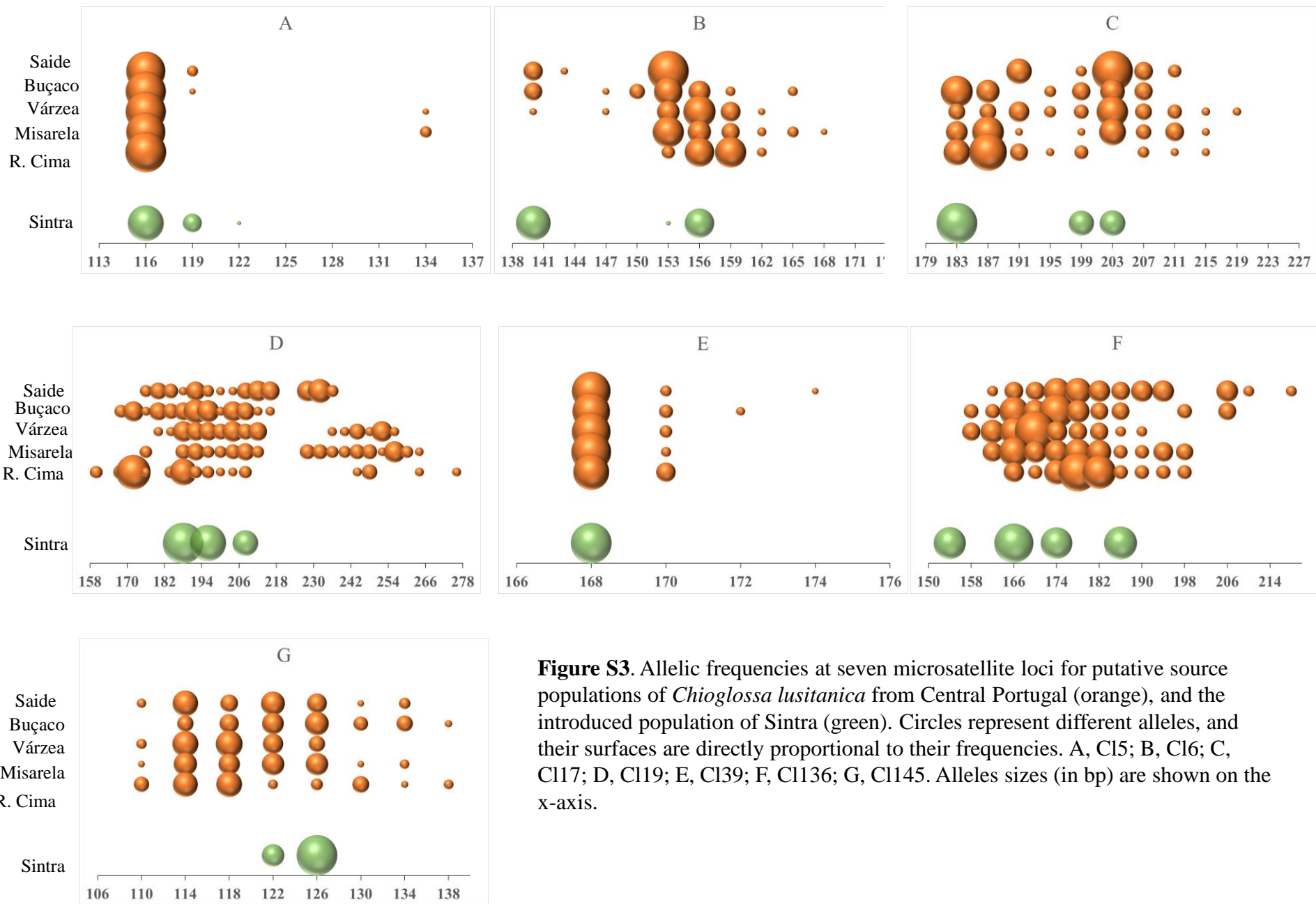


Figure S3. Allelic frequencies at seven microsatellite loci for putative source populations of *Chioglossa lusitanica* from Central Portugal (orange), and the introduced population of Sintra (green). Circles represent different alleles, and their surfaces are directly proportional to their frequencies. A, Cl15; B, Cl16; C, Cl17; D, Cl119; E, Cl139; F, Cl136; G, Cl145. Alleles sizes (in bp) are shown on the x-axis.