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Source identification and temporal evolution of trace ele ments in PM₁₀ collected near to Ny-Ålesund (Norwegian Arctic)

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14 Abstract: This study investigated the elemental composition of PM₁₀ collected in a polar environment (Ny-Ålesund, Norwegian Arctic), to identify its sources and to un-15 derstand the effects of short- and long-range transport processes. Natural (crustal and 16 marine) and anthropogenic Arctic PM sources were identified, and specific emission 17 18 sources were recognised by means of Principal Component Analysis and Hierarchical 19 Cluster Analysis: airborne pollution deriving from ship fuels, local vehicle (non-20 exhaust) and continental emissions (e.g. incinerators or industries) were the main 21 sources of anthropogenic elements. The results obtained so far from samples collected 22 during four successive spring-summer sampling campaigns (2010-2013) show a re-23 markable seasonal trend for most of the investigated elements. For both geogenic and 24 anthropogenic elements, concentrations are generally higher in March and April, when the ground is almost entirely covered by snow and ice, suggesting that long-range 25 26 transport processes might be taking place. On the other hand, the concentrations of an-27 thropogenic metals related to ship emissions (i.e. Co, Ni and V) peak in late spring and 28 summer, when the marine traffic in the fjord is generally higher. For most of the ana-29 lytes, the four campaigns were not significantly different; therefore, in the studied peri-30 od, the composition of PM_{10} in Ny-Ålesund did not vary remarkably. Finally, the behav-31 iour of Rare Earth Elements was discussed in terms of parent material mineralogy.

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12 13 Keywords: Ny-Ålesund (Norwegian Arctic), PM₁₀, elemental composition, seasonal
 trends, enrichment factors, Principal Component Analysis

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36 Declarations of interest: none.

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1. Introduction

39 The ambient concentration and the chemical composition of atmospheric particu-40 late matter (PM) depend on a large number of factors, such as the existence of specific 41 sources (natural or anthropogenic) and the environmental conditions (season, weather 42 and geographical area) (Minguillón et al., 2012). Conversely, PM is able to strongly 43 influence the ongoing climate changes, by taking part in many atmospheric physical and 44 chemical processes. Suspended particles can perturb the radiative balance of the atmos-45 phere by means of both direct and indirect effects, e.g. the scattering and the absorption 46 of the solar radiation and the tendency of particles to act as cloud condensation nuclei 47 and to participate in cloud formation (Giardi et al., 2016; Moroni et al., 2016; Talbi et 48 al., 2018). The link between PM and climate change is particularly important for polar 49 regions since, in these areas, even small changes of PM concentration or composition 50 can have significant effects on climate. In fact, polar regions play a key role in regulat-51 ing the global biogeochemical cycles and hence the Earth's climate system. In the last 52 few decades, global warming has caused a reduction of the polar sea-ice and snow cov-53 er. As a consequence, the capacity of the Earth's surface to reflect the solar radiation 54 (albedo) has decreased, causing a further increase in temperatures. These complex feed-55 back mechanisms cause fast climate changes, which result in poor process understand-56 ing and low-accuracy model predictions (Cappelletti et al., 2016).

57 The study of the chemical composition of polar PM is of great importance in aiding 58 identification of local and global sources and transport mechanisms and potential for 59 deposition in the Arctic. In winter and spring, when polluted air masses are more effi-60 ciently transported from mid-latitudes, various pollutants introduced in temperate zones 61 reach the Arctic, giving rise to the so-called "Arctic haze" (Quinn et al., 2007). The 62 composition of this haze includes sulphates, nitrates, ammonium, organic matter, black 63 carbon and heavy metals (Barrie et al., 1992; Barrie and Hoff, 1985; Bodhaine et al., 64 1989; Clarke, 1989; Gong and Barrie, 2005; Udisti et al., 2016). The latter are among the chemical markers commonly used in source apportionment studies to identify the
origin of PM (Bazzano et al., 2015; Bazzano et al., 2016a; Polissar and Hopke, 1998;
Shaw, 1982; Talbi et al., 2018).

68 In order to gain a better understanding of the effects of short- and long-range 69 transport processes taking place in this remote area, this study investigated the ele-70 mental composition of PM_{10} samples (aerodynamic diameter < 10 μ m). The samples were collected at Gruvebadet Station, near to Ny-Ålesund (Svalbard Islands), during 71 72 four subsequent spring-summer sampling campaigns (2010-2013). The choice of focussing on PM₁₀ fraction was driven by the desire of evaluating the totality of the sources 73 of Arctic PM and by the awareness that soil dust and particles deriving from non-74 75 exhaust vehicle emissions usually have a relatively high aerodynamic diameter. In addi-76 tion, a reduction of the collected particles would probably have resulted in a higher pro-77 portion of results below the LOD, only allowing the determination of major elements. In fact, the total mass concentration of the suspended particles (TSP) previously found 78 over the Arctic Ocean ranged from 0.10 to 3.8 μ g/m³ (Leck and Persson, 1996) and, in 79 particular, an average of 0.61 μ g/m³ was registered in summer 2012 in Ny-Ålesund 80 81 (Zhan et al., 2014). The concentrations of Al, As, Ba, Ca, Cd, Co, Cu, Fe, In, K, Mg, 82 Mn, Mo, Na, Ni, Pb, Ti, V, Zn and most of lanthanides (also indicated as Rare Earth 83 Elements or REEs), namely La, Ce, Pr, Nd, Sm, Eu, Gd, Tb, Dy, Ho, Er, Tm, Yb and 84 Lu, was determined; all these elements are commonly used, with the aid of chemometric 85 treatments and other graphical and statistical tools, as specific chemical markers for identifying the anthropogenic and natural (crustal and marine) sources of atmospheric 86 PM. An adequate analytical procedure to determine the trace element content in PM 87 88 collected in a virtually pristine area was developed and optimized; an Inductively Cou-89 pled Plasma - Optical Emission Spectrometer (ICP-OES) and a High-Resolution Induc-90 tively Coupled Plasma – Mass Spectrometer (HR-ICP-MS) were used for the analysis. 91 The results obtained from samples collected in 2010 campaign have already been pub-92 lished (Bazzano et al., 2016a); in this work, results obtained from samples collected 93 during 2011, 2012 and 2013 campaigns are presented, and comprehensive graphic and 94 statistical treatments are shown.

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96 **2. Experimental**

97 2.1 Study area and sampling

98 Ny-Ålesund (78°55'30"N, 11°55'40" E) is a small settlement acting as a centre for 99 international Arctic scientific research and environmental monitoring. It is located on 100 the shore of the bay of Kongsfjorden, on the western coast of Spitsbergen Island, in the 101 Svalbard archipelago (Figure 1). The aerosol sampling was performed at Gruvebadet 102 research station (40 m a.s.l.), situated at the base of Zeppelin Mountain (474 m a.s.l.), 103 800 m far from Ny-Ålesund. The location was established considering the prevailing 104 wind direction (115° N) (Mazzola et al., 2016), in order to minimize the influence of 105 local pollution during measurements. Moreover, a meteo-trigger system was used for 106 switching off the sampling devices when the wind was absent (< 0.5 m/s) or came from 107 Ny-Ålesund.

108 PM_{10} samples were collected with a 4-days resolution during four subsequent 109 spring-summer sampling campaigns (2010-13), as indicated in Table 1. PTFE hydro-110 philic filters (Advantec, product code: H100A090C, 90 mm diameter, efficiency > 99% 111 for 0.3 µm particles) and an Echo HiVol sampler (TCR Tecora, 200 L/min) were used. 112 After each sampling, filters were placed in polycarbonate Petri dishes, sealed and im-113 mediately frozen; samples were maintained at -20°C during all stages of transportation 114 to Italy and storage.

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2.2 Apparatus and reagents

117 The dissolution of samples was carried out using a Milestone Ethos One micro-118 wave laboratory unit. Analyses were carried out using ICP-OES or HR-ICP-MS, ac-119 cording to the analyte concentration ranges; wavelength, mass resolution and isotope 120 selection were optimized for each element to ensure resolution of spectral interferences 121 and maximization of sensitivity. Table 2 reports model and feature of each instrumental 122 technique used. Operating conditions and experimentally determined limits of detection 123 (LOD) for all the elements determined in PM₁₀ samples, are reported in Table 3; LOD 124 values represent the analyte concentration (ppt or ppb) corresponding to three times the standard deviation of the reagent blank. LOD values reported in Table 3 in fg/m³ and 125 pg/m^3 were obtained by conversion, using the nominal air volume of 1152 m³ and 15 126 127 mL as solution volume.

128 For the digestion of samples, ultrapure hydrogen peroxide (Sigma-Aldrich) and nitric

acid (Fluka), further purified by sub-boiling distillation in a quartz apparatus, were used. Water was purified in a Milli-Q system, resulting in high purity water (HPW) with a resistivity of 18.2 M Ω ·cm. Intermediate element standard solutions were prepared from concentrated (1,000 and 10,000 mg/L) stock solutions (Sigma-Aldrich TraceCERT) and acidified to pH=1.5.

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2.3 Sample pre-treatment and analysis

136 One quarter of each aerosol-loaded filter was subjected to a microwave-assisted di-137 gestion in 30 mL tetrafluoromethoxyl vessels, which were then inserted into 100 mL 138 tetrafluoromethoxyl vessels (vessel-inside-vessel technology (Nóbrega et al., 2015)). 139 Stainless steel scissors were used for cutting the filters. According to the current legisla-140 tion of the European Community in the field of air quality monitoring (UNI EN 141 14902:2005), the digestion mixture was composed of 2 mL HNO₃ and 0.5 mL H₂O₂; a 142 mixture of 10 mL HPW and 1 mL H₂O₂ was introduced in the bigger vessel and the 143 temperature was ramped to 220°C within 20 min, followed by a dwell time of other 20 144 min. Prior to the digestion, samples were left in contact with the digestion mixture for 145 approximately 16 hours, in order to ensure the complete impregnation of the filter. The 146 resulting solutions were filtered with Whatman Grade 5 cellulose filters, previously 147 cleaned with 20 mL HPW, and then diluted to 15 ml with HPW. All the possible steps 148 (filter cuts, digestion mixture preparation and filtration) were performed in a clean envi-149 ronment under a Class-100 laminar flow bench-hood. The reagent blanks, used for pre-150 paring the calibration standard solutions and for LOD calculation, were obtained by 151 microwave-digesting the reagent mixture only. On the other hand, sample blanks were 152 prepared similarly to the samples, by digesting some blank filters which had undergone 153 the transport to and from Ny-Ålesund; sample blank concentrations were subtracted 154 from sample concentrations, in order to eliminate the filter contribution. Sample blank 155 concentrations (ppb or ppt) are reported in Table 3, together with the converted values in pg/m^3 ; the nominal air volume of 1152 m³ and a solution volume of 15 mL were used 156 157 for conversion.

Sets of instrumental blank and calibration checks were run at frequent intervals during the analysis sequence. The relative standard deviation for each element in each sample was always lower than 5%. The Certified Reference Material (CRM) NIST 161 1648a ("Urban Particulate Matter") was used for verifying that a good recovery was 162 obtained for each analyte before proceeding with sample analysis. The recovery rates of 163 certified elements were higher than 80% for most of the analytes. Some of the typically 164 geogenic analytes (i.e. Al and Ti) can be completely extracted only by using HF. Never-165 theless, this is not a major problem since we mainly focussed on the variations of ana-166 lyte concentrations throughout and among the investigated summer seasons.

167 Due to their very low concentrations, it was not possible to obtain reliable results 168 for Tm and Lu. Therefore, these analytes were not included in data processing or graph-169 ical and statistical elaborations.

- 170
- 171 2.4 Enrichment factors

In order to estimate the influence of anthropogenic sources on the concentrations of major, minor and trace elements contained in PM_{10} samples, crustal and marine enrichment factors were calculated. Crustal Enrichment Factors (CEFs) were calculated with respect to the mean values for the Earth's upper crust reported by Wedepohl (Wedepohl, 1995), in order to distinguish elements having geogenic or non-geogenic origin. The equation used is the following:

$$CEF_{i} = \frac{\frac{C_{i PM}}{C_{r PM}}}{\frac{C_{i crust}}{C_{r crust}}}$$

178 where $C_{i PM}/C_{r PM}$ and $C_{i crust}/C_{r crust}$ are the ratios between the concentration of the ele-179 ment i and the concentration of a reference element r in the sample and in the upper 180 crust respectively; in this work, Al was selected as a reference element (Tahri et al., 2017). By convention, CEFs lower than 10 are taken as an indication that an element 181 182 (called "not enriched") has a prevailing geogenic origin, whereas CEFs between 10 and 183 100 indicate a moderate enrichment and CEFs higher than 100 indicate that the element 184 (called "enriched") has a prevailing non-geogenic origin (Lai et al., 2017; Tahri et al., 185 2017; Zajusz-Zubek et al., 2017). Similarly, Marine Enrichment Factors (MEFs) were 186 calculated with respect to the mean abundances of elements in sea water reported by 187 Goldberg (Goldberg, 1965), in order to distinguish elements having marine or nonmarine origin; the abundance of In in sea water, not reported by Goldberg, was taken
from Miyazaki (Miyazaki et al., 2008). The equation used is the following:

$$MEF_{i} = \frac{\frac{C_{i PM}}{C_{r PM}}}{\frac{C_{i sea}}{C_{r sea}}}$$

190 where $C_{i PM}/C_{r PM}$ and $C_{i sea}/C_{r sea}$ are the ratios between the concentration of the element 191 *i* and the concentration of a reference element *r* in the sample and in sea water respec-192 tively; Na was selected as a reference element (Krnavek et al., 2012). Again, it was as-193 sumed that MEFs lower than 10 indicate an analyte not enriched with respect to sea 194 water, whereas MEFs between 10 and 100 indicate a moderate enrichment and MEFs higher than 100 indicate that the analyte has a prevailing non-marine origin. The combi-195 196 nation of both high CEFs and high MEFs allows the identification of elements probably 197 having a strong anthropogenic contribution, as their main source is neither crustal nor 198 marine.

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2.5 Chemometric treatments and seasonal trends

201 Experimental results were processed by Lilliefors normality test, Kruskal-Wallis 202 test, Principal Component Analysis (PCA) and Hierarchical Cluster Analysis (HCA) 203 using XIStat 2017 software package, an add-on of Microsoft Excel. Information on the 204 principles of these techniques can be found elsewhere (Einax et al., 1997; Massart et al., 205 1997; Ruxton and Beauchamp, 2008). The analytes whose concentrations were below 206 the LOD in more than 30% of samples were not included in the treatments (Farnham et 207 al., 2002). Conover-Iman test with a level of confidence of 95% was used for multiple 208 pairwise comparison after the Kruskal-Wallis test. For performing PCA, values below 209 the LOD were considered as equal to it and values below the sample blank were consid-210 ered as equal to zero; the whole dataset was then autoscaled.

In order to gain a better comprehension of the temporal patterns shown by PCA and HCA and to study the behaviour of analytes which cannot be included in PCA (due to the absence of a sufficient amount of data above the LOD), the seasonal trends were evaluated. Metal concentrations were normalized by dividing them by the concentration of the same analyte in the first sampling of each campaign. In this way, the values reported for the first sampling are always equal to one, and the comparison of the temporal trends of metals having different concentration ranges are easier. Values below the
LOD were reported in graph as equal to it.

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2.6 HYSPLIT back-trajectories

Air mass back-trajectories were calculated using the NOAA HYSPLIT 4 transport model (Draxler and Rolph, 2003; Stein et al., 2015) and the GDAS meteorological data supplied by ARL (Air Resources Lab, http://ready.arl.noaa.gov). A propagation time of 120 h was set and Gruvebadet was used as the endpoint (78.92° N, 11.89°E, 0.0 m AGL); a new back-trajectory started every 6 hours.

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3. Results and discussion

228 *3.1 Elemental composition*

229 The concentrations of major, minor and trace elements determined in the PM_{10} 230 samples are shown in Supplementary Table 1 and 2; for converting the LOD from ppt (or ppb) to pg/m^3 (or fg/m^3), the actual solution volume and air volume of each sample 231 232 were used. Table 4 reports the descriptive statistics for PM₁₀ samples collected in Ny-Ålesund in 2010 (Bazzano et al., 2016a) and 2011-2013 (this study) campaigns: mean, 233 median, minimum and maximum values, 5th and 95th percentile levels; for comparison, 234 concentrations of major, minor and trace elements in PM₁₀ samples collected in other 235 236 parts of the world (Dai et al., 2016; Kulmatov and Hojamberdiev, 2010; Moreno et al., 237 2008b; Padoan et al., 2016; Toscano et al., 2005; Valdés et al., 2013) are reported. For 238 calculating mean, median and percentiles, values below the LOD were considered as 239 equal to it. Values below the sample blanks were not included in the calculations.

From Lilliefors normality test, it emerged that the results obtained for all analytes are not normally-distributed; therefore, non-parametric Kruskal-Wallis test and Conover-Iman pairwise comparison test were applied. Taking into account only samples collected in the period common to all the campaigns (May-July), the four campaigns were not significantly different for most of major and minor elements (Supplementary Table 3a). On the other hand, as regards As, Cd, Cu, Fe, Mo, Ti and most of lanthanides, data from one or two sampling campaigns do not belong to the same statistic popula247 tion. For most of lanthanides, this is due to the samples collected in 2010 having higher 248 concentrations, while for each of the other elements a different subdivision of sampling 249 campaigns was found. In addition, it can be seen that 2012 and 2013 sampling cam-250 paigns presented a higher similarity among them than with the other two studied cam-251 paigns; this could be due to a higher similarity of the atmospheric conditions registered 252 in these two years or to a change in the mid-latitudes emissions due to the introduction 253 of some new emission regulations. The same tests were used for checking the presence 254 of significant differences among samples collected in spring and summer (Supplemen-255 tary Table 3b). Again, no significant differences were found for most of the elements, 256 except for Al, Ca, Cd, Cu, Mn, Pb and V: the latter presented significantly higher con-257 centrations in summer than in spring, while for the other analytes it is the opposite. Alt-258 hough for Al, Ca and Mn this variability is likely due to natural phenomena, it is possi-259 ble that the seasonal variability of Cd, Cu, Pb and V is due to anthropogenic activities, 260 as demonstrated for Pb by isotopic analysis (Bazzano et al., 2016b).

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3.2 Ce-La-V ternary diagram

263 For a first identification of the sources of V, a Ce-La-V ternary diagram was used 264 (Becagli et al., 2017; Moreno et al., 2008a; Moreno et al., 2008b). For building this 265 graph, the concentration of Ce and La were multiplied by an adequate coefficient (1.54 266 and 3.1 respectively) so that the centre of the diagram represents the composition of the 267 upper crust (Henderson and Henderson, 2009). In this way, PM samples which are 268 strongly influenced by refinery emissions are generally located near to the lower left 269 corner of the diagram (Moreno et al., 2008a; Moreno et al., 2008b), while PM samples 270 which are strongly influenced by ship emissions or other oil combustion processes are 271 generally located near to the lower right corner (Becagli et al., 2017; Moreno et al., 272 2008a; Moreno et al., 2008b); in addition, Moreno et al. suggested that PM samples 273 strongly influenced by vehicular traffic emissions are likely located close to the triangle 274 centre, slightly toward Ce due to the abrasive loss of Ce-rich PM from catalytic con-275 verters (Moreno et al., 2008b). Figure 2 shows the Ce-La-V ternary diagram for PM₁₀ 276 samples collected near to Ny-Ålesund during 2010-2013 sampling campaigns.

277 Samples are mainly located in an area ranging from the lower right corner to the centre 278 of the diagram, therefore the two most important sources of V are likely weathering of 279 the upper crust and ship emissions; a contribution of refinery emissions can be exclud280 ed, while a contribution of vehicular emissions is conceivable only for very few samples 281 collected in summer 2012. In particular, three samples collected in July, August and 282 September 2012 are located in the upper right part of the diagram, due to their anoma-283 lously high content of Ce and, to a lesser extent, V; as a consequence, a possible contri-284 bution from vehicle exhaust emissions cannot be excluded for these samples. As it re-285 gards the seasonal distribution of samples in the diagram, it is possible to see that sam-286 ples strongly influenced by crustal sources were mainly collected in April; considering 287 that, in spring, the soil of Svalbard Islands is generally still covered by snow and ice, 288 this fact can be explained by the presence of PM deriving from mid-latitudes (Arctic 289 haze) (Yli-Tuomi et al., 2003). On the other hand, samples strongly influenced by oil 290 combustion were mainly collected in June and July, period during which the marine 291 traffic in the fjord is generally high, due to the reopening of research bases and to the 292 presence of cruise ships.

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3.3 Enrichment factors

295 Figure 3 reports the box plots for all the calculated EFs. From CEFs and MEFs, it 296 emerged that the origin of Ba, Co, Fe, Mn, Ti, V and lanthanides in the analysed sam-297 ples is mostly geogenic, the origin of Ca and Mg is primarily marine and the origin of 298 As, Cd, Cu, In, Mo, Ni, Pb and Zn is mainly anthropogenic. Nevertheless, some of these 299 elements seem to have a mixed origin, or a different origin in some of the analysed 300 samples; in particular, the concentrations of Ca, In, Ni, Pb and Zn seem to derive, in 301 some of the samples, from a crustal source, and the concentrations of Co and V might 302 depend, in some samples, on anthropogenic sources. As regards K, both CEFs and 303 MEFs are generally lower than 10, indicating that its concentration is likely influenced 304 by both marine and crustal sources, while an anthropogenic contribution might be ex-305 cluded. The calculation of EFs is a useful way for obtaining a first estimate of the dif-306 ferent sources of the analytes in PM samples; nevertheless, due to the simplicity of the 307 calculation and to the assumptions made in the interpretation of results, the information 308 obtained should be confirmed by other approaches, e.g. the study of the analyte tem-309 poral variations and multivariate analysis.

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311 *3.4 Chemometric treatments and seasonal trends*

312 The complexity of the matrices and of the phenomena under study renders interpre-313 tation of experimental results difficult. PCA and HCA can be a valuable tool for this 314 aim, owing to their multivariate approach, which takes into account the behaviour of 315 multiple variables simultaneously (Einax et al., 1997; Massart et al., 1997). With the aid 316 of PCA and HCA it was possible to obtain a global graphical representation of data and 317 to investigate relationships among variables, similarities and differences among sam-318 pling periods, and to identify the causes of the behaviour of the investigated analytes 319 (e.g. the element sources). In the following treatments, the names used for the samples 320 correspond to the beginning of the sampling.

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322 *3.4.1 2012 campaign*

323 For having information on the seasonal variability of PM₁₀ composition, we per-324 formed PCA including data from one sampling campaign at a time. The sampling cam-325 paign carried out in 2012, which is the longest of the three campaigns presented in this 326 study, is the one in which the seasonal variability of PM₁₀ is more visible. Therefore, it 327 is the only single-campaign PCA reported. Score and loading plots obtained by PCA for 328 2012 samples are shown in Figure 4 (PC1vs. PC2). The first PC, retaining 57.9% of the 329 total variance of the dataset, is mainly associated with analytes having a geogenic 330 origin, such as Al, Ba, Fe, Mn, Ti and lanthanides, except Ce (the label "REEs" in the 331 loading plot cumulatively indicates all lanthanides except the latter). By looking at the 332 score plot, it is possible to distinguish samples collected in April and May, having posi-333 tive scores on PC1, from samples collected in June, August and September, having neg-334 ative scores (except 12 Jun); samples collected in July, though, are spread among the 335 two groups. Considering that, in spring, the soil of Svalbard Islands is still covered by 336 snow and ice, it is likely that long-range transport of PM from mid-latitudes took place 337 in this period, thus influencing the composition of PM₁₀ samples collected near to Ny-338 Alesund; this result is coherent with previous findings, evidencing the so-called "Arctic 339 haze" in spring (Lupi et al., 2016; Udisti et al., 2016). A further proof of the different 340 sources of the PM₁₀ collected in the two seasons can be found in Figure 5, showing 341 HYSPLIT back-trajectories calculated for two samples: 15 May (positive scores on 342 PC1) and 3 Aug (negative scores on PC1). The latter is characterized by a low variabil-343 ity in the direction of the back-trajectories and by a path that remains within the Arctic 344 circle; on the other hand, some of the back-trajectories calculated for the sample collect345 ed in spring indicate a mid-latitude provenance, ranging from Europe to eastern Asia. 346 The fact that some samples collected in July present positive scores on PC1, which 347 means that the concentrations of geogenic analytes in those samples were relatively 348 high, can be explained considering the unusual atmospheric conditions registered in that 349 period; in fact, as shown by HYSPLIT back-trajectories reported in Moroni et al. 350 (2016), air currents coming from eastern Russia arrived in Ny-Ålesund, therefore a PM 351 contribution deriving from Gobi Desert cannot be excluded (Huang et al., 2015). The 352 position of Ce with respect to the other lanthanides suggests that this element might 353 have a non-geogenic, i.e. anthropogenic, origin. This hypothesis is supported by the 354 evidence that, for five samples collected in 2012, the calculated CEFs for Ce are higher 355 than 10.

356 PC2, retaining 14.1% of the total variance, is associated with the marine aerosol source, 357 as Mg and Na present the highest loadings. Other analytes having high loadings on PC2 358 are K, and, to a lesser extent, Ca, Mo and Ni; while Ca, K, Mg and Na are most likely to 359 have a natural origin, this is not true for the other two analytes, as indicated by their 360 MEFs. For Ni, airborne pollution deriving from ship emissions might be the main 361 source during spring, when the air masses coming from mid-latitudes overpass the 362 ocean and become enriched of analytes from marine aerosol. No clear subdivision of 363 samples according to PC2 can be seen.

364 Figure 6 shows the dendrogram obtained by Q-mode HCA, where three main clus-365 ters of analytes can be seen. The first cluster is composed by analytes having high load-366 ings on PC1 (i.e. Al, Ba, Fe, Mn, Ti and lanthanides), which are typically geogenic; as 367 expected, at a lower level of dissimilarity it is possible to see that lanthanides form a 368 distinct cluster from the other five analytes. The second cluster is composed by analytes 369 probably having an anthropogenic origin, i.e. As, Cd, Ce, Cu and Zn; at a lower level of 370 dissimilarity, it is possible to distinguish a small cluster composed of As and Cd, a small 371 cluster composed of Cu and Zn, and Ce as an outlier. As and Cd are commonly emitted 372 during combustion and industrial production processes (Pacyna et al., 2007; Ragazzi, 373 2017), which are likely occurring at mid-latitudes; Cu and Zn are likely associated with 374 non-exhaust vehicle emissions, as these elements are common indicators of brake and 375 tyre wear (Birmili et al., 2006; Councell et al., 2004; Fauser et al., 1999; Lough et al., 376 2005); considering its position in the dendrogram, Ce might also derive from vehicular 377 emissions, as this element is often used in catalytic converters (Angelidis and 378 Sklavounos, 1995; Arul Mozhi Selvan et al., 2014; Moreno et al., 2008b; Sajeevan and 379 Sajith, 2016; Silveston, 1995). The third cluster is composed by analytes deriving from 380 marine aerosol: apart from K, Mg and Na, it is possible to identify a small cluster com-381 posed of Ca and Mo, which can partially have a crustal contribution, and a small cluster 382 composed of Co, Ni and V; therefore, airborne pollution deriving from ship emissions 383 might be the main source not only for Ni but also for Co and V, even though some other individual sources might be present for these analytes. Similar information was obtained 384 385 by PCA, by investigating PC3 and PC4 (Supplementary Figure 1): despite the low por-386 tion of the total variance retained by these PCs, they carry information on PM sources 387 contributing only slightly to the overall PM composition (i.e. anthropogenic sources of 388 PM collected in a pristine environment). In particular, it is evident that the three samples 389 having the highest concentration of Co, Ni and V have been collected at the end of July, 390 when at least two cruise ships having more than 1500 passengers travelled in Kongs-391 fjorden Bay (Zhan et al., 2014).

392 For making hypothesis on the possible sources of Pb, a probably anthropogenic an-393 alyte which could not be included in PCA and HCA, its seasonal trend was used. Figure 394 7 shows that the seasonal trend registered for Pb is analogous to the one obtained for As, 395 Cd and Zn, which presented relatively high concentrations in samples collected in April 396 and in the first part of May, when long-range transport processes from mid-latitudes 397 take place. Therefore, it is likely that the spring origin of Pb coincides with the one of 398 the other analytes, i.e. anthropogenic continental sources such as vehicular traffic, in-399 cinerators or industries. Isotopic analysis indicated that mining and smelting activities in 400 the Rudny Altay region (Central Eurasia) were the main sources of atmospheric lead in 401 spring, whereas industrial emission in north-eastern North America, were the main 402 sources of atmospheric lead in summer (Bazzano et al., 2016b). Zn presents some con-403 centration peaks in June and July and a sharp increase at the beginning of September, 404 not coinciding with the trends registered for As, Cd and Pb; therefore, a second source 405 for Zn might have been introduced from the beginning of June, explaining its different 406 seasonal trend. The other possible source for this element can be non-exhaust vehicle 407 emissions, fact confirmed by the similarity between Cu and Zn shown by HCA (Figure 408 6, Paragraph 3.3.1). In fact, it is likely that the local vehicle contribution became higher 409 in summer, when the population of Ny-Ålesund is greater and the long-range transport 410 processes of PM from mid-latitudes are lower.

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3.4.2 Comprehensive treatment (2010-2013)

413 A comprehensive PCA including data from the four successive sampling cam-414 paigns (2010-2013) was executed. The first PC (46.8% of variance, not shown) is asso-415 ciated with the content of crustal elements, as Al, Ba, Fe, Mn and lanthanides (cumula-416 tively indicated in the graph as "REEs") present the highest loadings; two samples col-417 lected in July 2013 present by far the highest scores on this PC. The score and loading 418 plot for PC2 vs. PC3 (10.8% and 7.6% of variance respectively) are reported in Figure 419 8. Two main groups of analytes can be identified in the loading plot: As, Cd, Pb and Zn, 420 generally representing anthropogenic contributions, and Mg and Na, representing the 421 marine spray input; a cluster made of Co, Ni and V, having small positive loadings, lays 422 in the middle of the two main clusters and might be attributed to ship emissions. Ca and 423 K lay in the same direction as Mg and Na, but their loadings on PC3 are much smaller; considering that they presented quite high loadings on PC1 (not shown), it is possible to 424 425 hypothesize a mixed crustal-marine origin for these elements. Two samples collected in 426 April 2010 and one sample collected in July 2013 have the highest scores on PC2, as a 427 consequence of their relatively high content of the analytes constituting the first cluster; 428 for 2010 samples, this can be attributed to the several eruptions of the Icelandic Ey-429 jafjöll Volcano which took place in that spring. The score plot of Figure 8a shows that 430 samples located at positive values of PC2, having therefore relatively high concentra-431 tions of marine and anthropogenic analytes, were prevalently collected in spring. Apart from a slight prevalence of samples collected in 2010 in the lower right portion of the 432 433 graph (positive values of PC2 and negative values of PC3), which may still be attributed 434 to the Eyjafjöll eruptions, no clear subdivision of samples according to the sampling 435 campaign can be seen on either PC1, PC2 or PC3; consequently, it is possible to state 436 that neither the natural (crustal and marine) nor the anthropogenic contribution signifi-437 cantly changed during the investigated period. This result is a confirmation of what 438 emerged from Kruskal-Wallis and Conover-Iman tests.

439

440 *3.5 Rare Earth Element patterns*

441 As mentioned in paragraph 3.1, REE concentrations determined in the PM_{10} sam-442 ples are shown in Supplementary Table 2 and their descriptive statistics are reported in

443 Table 4; values below the LOD were not included in any of the following calculations. 444 The overall shape of the lanthanides concentration pattern is the naturally occurring 445 saw-tooth distribution, i.e. the odd-even pattern following the Oddo-Harkins rule 446 (Henderson, 1984; Piper and Bau, 2012; Schmidt et al., 1963); Figure 9a shows the 447 mean lanthanide concentrations found for each sampling campaign. Figure 9b shows, 448 for each sampling campaign, the mean lanthanide concentrations normalized to the av-449 erage shale reported by Byrne and Sholkovitz (Byrne and Sholkovitz, 1996), while Fig-450 ure 9c reports the mean normalized lanthanide concentrations for each month of 2012 451 sampling campaign. It is generally agreed that the average shale REEs pattern repre-452 sents that of the upper crust (Taylor and McLennan, 1985); therefore, normalization to 453 the average shale allows the removal of the systematic differences existing between 454 adjacent elements and to highlight significant composition differences arising from the 455 parent material. In this case, the exact value of the normalized concentration has no 456 meaning, hence only the shape of the pattern and the relative values should be taken into 457 account. For each sample, the pattern has a slightly negative slope, indicating a small 458 light REEs enrichment; in fact, the calculated Pr/Yb ratios (Figure 9d) are usually great-459 er than one (Lawrence et al., 2006). A light REEs enrichment is generally a consequence 460 of weathering of rocks with unstable mineralogy under surface conditions, such as ul-461 tramafic and basaltic rocks (Taylor and McLennan, 1985). The mean patterns registered 462 for the investigated campaigns and the mean patterns registered for each month of 2012 463 sampling campaign are approximately the same, indicating that, in the investigated pe-464 riod, the sources of geogenic analytes in the PM₁₀ collected near to Ny-Ålesund do not 465 vary significantly.

466 Ce anomaly was calculated as follows:

$$Ce_{anom.} = \frac{Ce_N}{\sqrt{La_N \cdot Pr_N}}$$

where the subscript "N" indicates that the concentrations are normalized to the average shale. Ce anomaly represents the deviation of Ce concentration from that expected by calculating the geometric average of the concentrations of La and Pr in shale; a deviation from unity of this value is thus a measure of the deviation (fractionation) of Ce from the expected REEs behaviour (Bazzano et al., 2016a; Giardi et al., 2018; Taylor and McLennan, 1985). Ce anomaly in PM samples can arise from an anthropogenic release of Ce in the atmosphere or from the geochemical characteristics of the parent

474 rock, in their turn deriving from the redox conditions registered during the rock formation: under alkaline conditions, Ce^{3+} can be oxidized by atmospheric oxygen to Ce^{4+} , 475 476 and the formation of the insoluble CeO₂ modifies the distribution of this element in the 477 environment (De Baar et al., 1988; Henderson, 1984). Figure 9e shows Ce anomaly for 478 each month of each sampling campaign: its value was higher than 1 in most cases. In 479 particular, high Ce enrichments were registered in August 2012 and 2013 and in Sep-480 tember 2012; considering that, in these months, the input of PM from mid-latitudes was 481 negligible, the high Ce concentrations can be attributed to local pollution. The observa-482 tion of Figure 2 and 6 (Paragraph 3.2 and 3.4.1) drove to the same conclusion.

483 Analogously to Ce, Eu anomaly was calculated as follows:

$$Eu_{anom.} = \frac{Eu_N}{\sqrt{Sm_N \cdot Gd_N}}$$

484 Eu anomaly always reflects the geochemical characteristics of the parent rock, as the additional oxidation state of this element allows the substitution of Eu^{2+} in place of 485 Ca^{2+} , Pb^{2+} or Sr^{2+} in some minerals, particularly in feldspars, determining an Eu en-486 richment with respect to the other REEs; as a consequence, most of the other minerals 487 488 (both igneous and sedimentary) commonly present a complementary Eu depletion with 489 respect to the other REEs. Figure 9f shows Eu anomaly for each month of each sam-490 pling campaign. Eu anomaly was higher than 1 in most cases, suggesting that the geo-491 genic portion of collected PM₁₀ mostly derives from weathering of feldspars. Neverthe-492 less, no specific trends were registered for Eu anomaly; Eu depletion with respect to 493 other REEs was only registered in samples collected in March and April 2011 and in May 2013, probably indicating a prevalence of different geogenic sources for the PM_{10} 494 495 collected during those months.

496

497 *3.6 Comparison with other sites*

For comparison, concentrations of major, minor and trace elements in PM_{10} samples collected in other parts of the world (Kulmatov and Hojamberdiev, 2010; Marx et al., 2014; Moreno et al., 2008b; Öztürk et al., 2012; Padoan et al., 2016; Toscano et al., 2005; Truzzi et al., 2017) are reported in Table 4. As expected, metal concentrations found in PM_{10} samples collected in Terra Nova Bay (Antarctica) are comparable with the ones of Ny-Ålesund, for all the investigated elements. The other areas presented for 504 comparison generally show remarkably higher element concentrations than the median 505 values of this study: this can be explained considering that the atmospheric circulation 506 causes even remote areas located at mid-latitudes (e.g. Abramov Glacier and New Zea-507 land's Southern Alps) to be affected by continental emissions. Nevertheless, the concen-508 tration of anthropogenic analytes registered in samples collected in New Zealand's 509 Southern Alps are comparable with the 95th percentile of the results obtained in this 510 study, which might be considered representative of concentrations registered in spring.

Very few studies have been executed on the REEs content in PM_{10} samples, and most of them regard PM_{10} samples collected near mines, not directly comparable with our study results. Comparison data reported for REEs in Table 4b derive from PM_{10} samples collected in remote and rural areas. The results presented for comparison are remarkably higher; this fact can be attributed to the permanence, in Svalbard Islands, of a snow and ice cover for most of the year, causing weathering phenomena to be markedly reduced with respect to other parts of the world.

518

519 **4.** Conclusions

In this study, the elemental composition of PM₁₀ samples collected near to Ny-520 521 Ålesund (Norwegian Arctic) in the sampling campaigns 2011-2013 was investigated, 522 and compared with results obtained for 2010 sampling campaign, previously published. 523 Natural (crustal and marine) and anthropogenic Arctic PM sources were identified, and 524 a better understanding of the effects of short- and long-range transport processes taking 525 place in this remote area was made possible; these data can be useful for studies on the 526 effect of Arctic PM on the climate change, as well as for a comparison of the composi-527 tion of PM₁₀ collected in other remote and anthropized areas.

528 Crustal Enrichment Factors and Marine Enrichment Factors allowed us the identifica-529 tion of elements having geogenic, marine or other origin. PCA and HCA allowed us a 530 better understanding of the sources of different elements and to lay the basis for an in-531 terpretation of the chemical and physical processes concerning the Arctic atmosphere; 532 seasonal concentration trends were helpful for the source identification of analytes 533 which could not be included in the chemometric treatments. Airborne pollution deriving 534 from ship fuels, local vehicle (non-exhaust) and continental emissions (e.g. incinerators 535 or industries) were the main sources of anthropogenic elements. Most of the elements determined in PM₁₀ samples collected near to Ny-Ålesund present seasonal variations: 536

537 for both geogenic and anthropogenic elements, concentrations are generally higher in 538 spring, when the ground is almost entirely covered by snow and ice, suggesting that 539 long-range transport processes of PM from mid-latitudes might be taking place; on the 540 other hand, the concentrations of anthropogenic metals related to ship emissions (i.e. 541 Co, Ni and V), peak in late spring and summer, when the marine traffic in the fjord is 542 generally high due to the reopening of research bases and to the presence of several 543 cruise ships in Kongsfjorden Bay. For most of the analytes, the four sampling cam-544 paigns were not significantly different; therefore, in the studied period, the composition 545 of PM₁₀ in Ny-Ålesund did not vary remarkably. For all the investigated elements, con-546 centration values registered in PM₁₀ samples collected in Terra Nova Bay (Antarctica) 547 are comparable with the ones of this study; all the other areas considered for comparison 548 exhibit remarkably higher element concentrations.

- 549
- 550 **5.** Acknowledgments

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- 555
- 556 **6. References**
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Campaign	Start date	End date	Samples
2010	16 th March	16 th September	36
2011	29 th March	26 th July	24
2012	17 th April	7 th September	36
2013	1 st May	13 th September	33

Table 1. Sampling dates and number of samples collected in each campaign.

2010 campaign results are reported in Bazzano et al., 2016a.

Table 2. Model and features of the instrumental techniques used for the analysis.

Technique	Model	Features	Analytes
ICP-OES	Perkin Elmer Optima 7000 DV	Mira Mist nebulizer, cyclonic spray chamber, dual Échelle monochromator, dual CCD detector	Na, Mg, K, Ca
HR-ICP-MS	Thermo Finnigan Element 2	Conikal nebulizer, Scott spray chamber, magnetic and electric sector, SEM detector	Al, As, Ba, Cd, Ce, Co, Cu, Dy, Er, Eu, Fe, Gd, Ho, In, La, Lu, Mn, Mo, Nd, Ni, Pb, Pr, Sm, Tb, Ti, Tm, V, Yb, Zn

										HR-IC	P-MS									
			Al	As	Ba	Cd	Со	Cu	Fe	In	Mn	Мо	Ni	P	Pb	Ti	V	Zn	_	
I: r	sotope esolution		27 LR-MR	75 MR	135-137 LR-MR	111-112-114 LR	4 59 LR-MR	63-6 R LR-M	5 56-57 IR MR	/ 115 MR	55 LR-MR	95-98 LR-MI	60-61- R LR-M	62 206-20 R LR-	07-208 -MR	46-47-48 LR-MR	51 LR-MR	66-68 LR-M	8 [R	
L	LOD (ppt)		400	42	62	2.3	1.9	11	400	1.8	8.6	4.2	45	4	40	750	58	190		
I	LOD (fg/m ³)		5200	550	810	30	25	140	5200	23	110	55	590	52	20	9800	760	2500)	
S	ample blank ((ppb)	0.63	< LOD	1.1	< LOD	< LOD	< L0	D 3.2	< LOD	0.040	0.019	0.50	0.	.23	<LOD	< LOD	4.5		
S	ample blank ((pg/m ³)	8.2	< LOD	14	< LOD	< LOD	< L0	D 42	< LOD	0.52	0.25	6.5	3	3.0	< LOD	< LOD	59		
]	HR-ICP-N	MS											ICP	-OES
	7. La	8. Ce	Pr	Nd	Sm	Eu	Gd	Тb	Dy	Но	Er	Tm	Yb	Lu			(Ca	K	Mg
e tion	139 LR	140 LR	141 LR	143-146 LR	147-149 LR	151-153 LR	155-157 LR	159 LR	161-163 LR	165 LR	166-167 LR	169 LR	172 LR	175 LR	Wavel	length (nm)) 31	7.9 7	69.9	285.2
ppt)	0.92	2.5	0.47	1.8	1.0	0.63	3.3	0.20	0.91	0.17	0.70	0.19	1.2	0.37	LOD	(ppb)	3	i.1 ().80	6.9
fg/m ³)	12	33	6.1	23	13	8.2	4.3	2.6	12	2.2	9.1	2.5	16	4.8	LOD	(pg/m ³)	2	40	10	90
hlank (nnt)	LOD	LOD					<100		<10D	<1.00	<10D	<10D	<10D	<10D	Samul	la blank (ni	ab) 2	60	88	< 1.01
e blank (ppt)	<pre><lod< pre=""></lod<></pre>	< LOD	< LOD	< LOD	< LOD	< LOD	< LOD ·		< LOD	< LOD	< LOD	< LOD	< LOD	LOD	Samp	ie blank (pj	pD) 2	.00	0.0	< D01

Table 3. Experimental conditions and limits of detection (LOD) of the analytes of interest with the technique used for their determination.

LR = Low Resolution; MR = Medium Resolution.

Table 4. Descriptive statistics for PM_{10} samples collected in Ny- Ålesund in 2010 (Bazzano et al., 2016a) and 2011-2013 (this study) and average concentrations of major, minor and trace elements in PM_{10} samples collected in other parts of the world. For calculating mean, median and percentiles, values below the LOD were considered as equal to it; therefore, all mean values except the ones reported for Al, Mg and Na should be intended as < N.

		Al (ng/m^3)	As (ng/m^3)	Ba	Ca	Cd (ng/m^3)	Co (ng/m^3)	Cu (ng/m^3)	Fe (ng/m^3)	$\ln (ng/m^3)$	\mathbf{K}	Mg (ng/m^3)	Mn (ng/m^3)	Mo (ng/m^3)	Na (ng/m^3)	Ni (ng/m^3)	Pb (ng/m^3)	Ti (ng/m^3)	V (ng/m^3)	Zn (ng/m^3)
3	Mean	(ng/m) 6.56	(pg/m) 17	(pg/m) 92	20	(pg/m) 6.4	(pg/m) 6.9	(pg/m) 55	(ng/m) 6.6	(pg/III) 1.5	(ng/m) 10	(ng/m) 26.7	150	18	(ng/m) 170	(pg/m) 95	(pg/m) 75	300	40	(pg/m) 600
010-1	Median	4.96	6.76	54.8	17.0	1.82	2.01	43.4	5.03	1.05	7.65	20.8	109	7.59	137	63.6	29.5	183	18.0	276
ND 2	Min	0.290	< 0.029	< 2.2	< 0.56	< 0.0035	< 0.015	< 0.32	< 0.34	< 0.00011	< 0.0017	0.961	< 3.7	< 0.36	2.93	< 0.029	< 0.030	< 1.9	< 0.21	< 2.1
ESU	Max	62.2	242	1020	75.3	125	16.0	608	56.2	24.8	148	143	1320	125	1010	406	1220	4340	585	18000
-ÅL]	5 th percentile	0.541	1.4	13	5.1	0.35	0.053	1.2	0.78	0.048	0.63	7.77	17	2.0	33.6	6.5	0.032	25	0.78	47
λN	95 th percentile	16.8	59.1	193	39.0	28.6	11.9	122	14.3	3.38	26.7	60.2	400	78.7	353	311	381	774	180	1320
z	Abramov Glacier (Kyrgyzstan) ¹	709	5400	6·10 ⁶	-	-	300	11800	1650	-	-	-	30000	-	80.10	10000	-	-	14000	13000
ISO	Antalya (Turkey) ²	354	600	6860	775	170	130	-	193	3000	245	363	6000	-	1296	1750	38000	17000	2140	9400
PAR	New Zealand's Southern Alps ³	-	5.6	1510	-	17.8	30.7	192	-	0.1	-	-	-	14.6	-	371	163	4917	119	-
IMO	Terra Nova Bay (Antarctica)	7.75 ⁴	-	60.9 ⁴	-	4.30 5	5.09 ⁴	410 5	$6.58^{\ 4}$	-	-	-	147 4	-	-	-	33.0 ⁵	-	22^{4}	109 4
Ŭ	Turin (Italy) ⁶	137	380	16700	482	310	180	3900	356	-	232	148	8230	1090	-	6450	6060	18500	1440	37900

		9. La (fg/m ³)	10. Ce (fg/m ³)	Pr (fg/m ³)	Nd (fg/m ³)	Sm (fg/m ³)	Eu (fg/m ³)	Gd (fg/m ³)	Tb (fg/m ³)	Dy (fg/m ³)	Ho (fg/m ³)	Er (fg/m ³)	Yb (fg/m ³)
	Mean	2400	5300	510	1800	400	150	440	82	260	39	160	130
	Median	1150	2330	264	820	209	68.7	199	32.9	139	18.9	80.7	67.7
-13	Min	< 10	< 60	< 3.6	< 15	< 4.1	< 4.7	< 5.9	< 1.4	< 5.9	< 1.0	< 3.0	< 3.4
2010	Max	28500	52900	4430	17200	3510	958	4030	640	1910	484	1030	880
g	5 th percentile	52	100	26	33	14	12	15	1.8	13	1.5	10	9.3
ESU	95 th percentile	8550	19000	2100	8280	1560	648	1940	304	969	123	627	507
ζ-ÅL	Antalya (Turkey) ²	240000	450000	56000	210000	44000	-	47000	6000	35000	7000	20000	18000
Z	Mexico City (Mexico) ⁷	1167000	1402000	205000	596000	81000	22000	79000	6000	42000	9000	30000	40000
	New Zealand's Southern Alps ²	31700	67900	7900	29800	6000	1200	5200	800	4600	900	2600	2400
	Turin (Italy) ⁶	20000	150000	-	-	-	-	-	-	-	-	-	-

¹ Mean values for PM₁₀ samples collected in remote areas in 1996/1997 (Kulmatov and Hojamberdiev, 2010).

² Median values for PM_{10} samples collected in a remote area in 1993/2001 (Öztürk et al., 2012).

³ Mean values for TSP samples collected in a remote area in 2009 (Marx et al., 2014).

⁴ Median values for PM₁₀ samples collected in a remote area in 2001/2002 (Toscano et al., 2005).

 5 Median values for PM_{10} samples collected in a remote area in 2000/2001 (Truzzi et al., 2017).

⁶ Mean values for PM₁₀ samples collected in a rural area ("DR") in 2011 (Padoan et al., 2016).

⁷ Median values for PM₁₀ samples collected in a rural area ("T2") in 2006 (Moreno et al., 2008b).





Figure 2. Ce-La-V ternary diagram for 2010-2013 samples.



Figure 3. Box plots of a-b) crustal enrichment factors calculated on all samples with respect to the mean crust composition reported by Wedepohl, 1995; c-d) marine enrichment factors calculated on all samples with respect to the sea water composition reported by Goldberg (1965) and Miyazaki (2008).





Figure 4. Principal Component Analysis (PC1 vs. PC2) for 2012 samples: a) score plot; b) loading plot.









Figure 7. Seasonal trends of As, Cd, Pb and Zn in 2012.





Figure 8. Comprehensive Principal Component Analysis (PC2 vs. PC3) for 2010-2013 samples: a) score plot; b) loading plot.

Figure 9. REE patterns: a) mean lanthanides concentration; b) mean shale-normalized concentration; c) mean shale-normalized concentration for each month of 2012 sampling campaign; d) Pr/Yb ratio; e) Ce anomaly; f) Eu anomaly.



Highlights

- Natural (geogenic and marine) and anthropogenic PM_{10} sources were identified.
- Anthropic element concentrations show a seasonal trend with maxima in spring.
- Airborne pollution from ship emissions was occasionally registered in summer.
- Anthropogenic element concentrations did not change over the examined years.

SUPPLEMENTARY MATERIAL

Supplementary Table 1. Major and minor analyte concentrations determined in PM₁₀ samples collected in a) 2011; b) 2012; c) 2013.

a)								20	11 CAM	PAIGN								
Start date	Al (ng/m ³)	As (pg/m ³)	Ba (pg/m ³)	Ca (ng/m ³)	Cd (pg/m ³)	Co (pg/m ³)	Cu (pg/m ³)	Fe (ng/m ³)	K (ng/m ³)	Mg (ng/m ³)	Mn (pg/m ³)	Mo (pg/m ³)	Na (ng/m ³)	Ni (pg/m ³)	Pb (pg/m ³)	Ti (pg/m ³)	V (pg/m ³)	Zn (pg/m ³)
29 Mar	5.00 ± 0.04	149 ± 6	53 ± 2	18.4 ± 0.2	19.8 ± 0.2	8.21 ± 0.05	130 ± 3	7.6 ± 0.3	5.1 ± 0.9	16.3 ± 0.1	284 ± 7	4.1 ± 0.4	74 ± 1	193 ± 5	470 ± 10	114 ± 5	40.9 ± 0.8	1140 ± 20
03 Apr	10.1 ± 0.1	50 ± 4	44 ± 2	32.63 ± 0.08	8.8 ± 0.1	10.2 ± 0.2	76 ± 3	11.3 ± 0.2	14 ± 2	51.4 ± 0.3	357 ± 8	11.9 ± 0.8	251 ± 3	350 ± 10	181 ± 6	213 ± 6	61 ± 2	1330 ± 30
07 Apr	7.65 ± 0.04	40.8 ± 0.8	53 ± 2	29.8 ± 0.2	7.9 ± 0.2	6.84 ± 0.06	65 ± 2	10.5 ± 0.2	7.9 ± 0.5	31.5 ± 0.2	336 ± 4	4.5 ± 0.4	138 ± 2	195 ± 7	190 ± 6	242 ± 9	56 ± 1	1080 ± 20
11 Apr	1.82 ± 0.02	5.8 ± 0.2	27.4 ± 0.8	14.1 ± 0.3	< 0.20	< 0.045	30.4 ± 0.2	2.36 ± 0.04	5 ± 1	n.a.	63.7 ± 0.8	3.2 ± 0.3	139 ± 2	105 ± 3	47.4 ± 0.5	49 ± 1	7.4 ± 0.2	61 ± 2
01 May	5.46 ± 0.05	< 6.9	< SB	14.7 ± 0.2	3.7 ± 0.2	7.9 ± 0.2	53 ± 2	5.0 ± 0.1	8.7 ± 0.5	36.2 ± 0.4	140 ± 2	3.6 ± 0.2	227 ± 5	153 ± 9	54 ± 2	98 ± 4	24.0 ± 0.8	< SB
05 May	11.3 ± 0.2	29 ± 2	74 ± 3	39.17 ± 0.09	5.5 ± 0.2	7.6 ± 0.1	64.1 ± 0.9	11.6 ± 0.2	13.9 ± 0.4	46.5 ± 0.8	243 ± 8	3.1 ± 0.2	263 ± 2	< SB	120 ± 4	329 ± 5	30.6 ± 0.8	399 ± 9
09 May	6.09 ± 0.01	24 ± 2	133 ± 2	60.9 ± 0.7	31 ± 1	4.4 ± 0.1	608 ± 8	6.80 ± 0.05	148 ± 7	28.86 ± 0.09	341 ± 6	17.7 ± 0.8	236 ± 2	400 ± 30	161 ± 4	153 ± 4	16.8 ± 0.7	4140 ± 80
13 May	7.37 ± 0.06	16.1 ± 0.4	76 ± 2	35.7 ± 0.2	3.9 ± 0.1	5.7 ± 0.1	91 ± 3	10.2 ± 0.2	11 ± 1	47.3 ± 0.1	238 ± 6	3.0 ± 0.2	298 ± 3	< SB	67 ± 3	168 ± 6	21.9 ± 0.7	890 ± 20
17 May	12.0 ± 0.1	18 ± 1	64 ± 2	22.3 ± 0.1	7.0 ± 0.3	5.7 ± 0.1	59 ± 2	10.9 ± 0.3	7 ± 1	22.6 ± 0.2	191 ± 4	30.1 ± 0.9	117.2 ± 0.8	< SB	66 ± 2	233 ± 7	22.2 ± 0.4	440 ± 10
21 May	7.9 ± 0.1	11.3 ± 0.7	2.16 ± 0.03	23.8 ± 0.3	< 4.5	6.7 ± 0.2	112 ± 2	5.7 ± 0.4	12.8 ± 0.3	22.8 ± 0.1	155 ± 6	6.1 ± 0.3	111 ± 2	6.8 ± 0.4	33.9 ± 0.7	160 ± 3	18 ± 1	< SB
25 May	3.59 ± 0.02	< 6.0	10.4 ± 0.3	11.3 ± 0.2	1.58 ± 0.03	4.49 ± 0.09	57 ± 2	4.8 ± 0.3	3.0 ± 0.6	16.63 ± 0.02	85 ± 3	0.88 ± 0.07	85 ± 2	63 ± 2	23 ± 1	66 ± 3	10.8 ± 0.7	2.08 ± 0.06
29 May	2.82 ± 0.01	< 5.2	46 ± 2	14.0 ± 0.2	1.74 ± 0.06	4.95 ± 0.09	78 ± 1	4.05 ± 0.09	< 0.056	9.73 ± 0.08	78 ± 1	2.0 ± 0.2	34.3 ± 0.8	< SB	17.1 ± 0.5	71 ± 1	12.9 ± 0.2	175 ± 5
05 Jun	2.62 ± 0.02	58 ± 4	17.2 ± 0.5	22.3 ± 0.3	27.3 ± 0.5	3.7 ± 0.1	53 ± 2	2.54 ± 0.03	6.6 ± 0.3	28.3 ± 0.2	59 ± 1	91 ± 3	185 ± 4	< 6.2	1.41 ± 0.05	96 ± 3	7.0 ± 0.5	247 ± 5
09 Jun	1.92 ± 0.02	< 5.8	< SB	22.7 ± 0.4	1.8 ± 0.1	2.7 ± 0.1	30 ± 2	1.98 ± 0.01	8.0 ± 0.5	36.1 ± 0.3	48 ± 1	< 0.64	258 ± 2	< 6.4	< 5.5	34 ± 2	5.4 ± 0.2	252 ± 6
13 Jun	10.7 ± 0.1	16 ± 1	41.4 ± 0.8	53.2 ± 0.5	15.1 ± 0.6	9.4 ± 0.3	51.0 ± 0.5	14.3 ± 0.1	26.7 ± 0.4	36.52 ± 0.09	328 ± 3	10.8 ± 0.7	202.1 ± 0.8	78 ± 4	61 ± 1	380 ± 10	134 ± 5	1200 ± 30
17 Jun	6.40 ± 0.08	138 ± 5	10.0 ± 0.1	28.7 ± 0.7	6.8 ± 0.2	5.1 ± 0.2	34.7 ± 0.6	5.96 ± 0.04	23.8 ± 0.7	21.16 ± 0.07	130 ± 1	6.6 ± 0.6	118 ± 2	90 ± 4	30.1 ± 0.5	182 ± 5	125 ± 3	103 ± 2
21 Jun	2.03 ± 0.02	23 ± 1	< SB	16.0 ± 0.2	< 0.60	1.82 ± 0.08	< 2.5	4.06 ± 0.06	14.3 ± 0.6	4.29 ± 0.07	21.4 ± 0.4	2.9 ± 0.3	21.0 ± 0.4	< SB	< 5.5	36.9 ± 0.8	3.5 ± 0.5	< SB
25 Jun	2.383 ± 0.008	< 7.6	< SB	20.0 ± 0.4	< 0.79	1.88 ± 0.07	56.7 ± 0.8	2.04 ± 0.02	1.9 ± 0.4	10.5 ± 0.2	39 ± 1	8.1 ± 0.6	45.6 ± 0.3	< SB	< SB	40 ± 2	13.3 ± 0.4	< SB
29 Jun	2.85 ± 0.04	< 5.9	< SB	9.5 ± 0.2	< 0.61	2.35 ± 0.03	82 ± 2	2.33 ± 0.02	< SB	12.8 ± 0.1	46 ± 1	3.5 ± 0.2	63.5 ± 0.9	6.3 ± 0.3	< SB	46 ± 3	6.7 ± 0.3	< 55
03 Jul	2.38 ± 0.02	< 5.9	< SB	24.7 ± 0.4	< 0.60	6.2 ± 0.2	125 ± 2	3.50 ± 0.02	15.2 ± 0.3	50.3 ± 0.5	54.4 ± 0.9	6.5 ± 0.4	361 ± 4	108 ± 5	1.19 ± 0.02	69 ± 2	181 ± 2	< SB
07 Jul	2.33 ± 0.01	< 5.8	< SB	38.7 ± 0.2	< 0.60	1.59 ± 0.07	40 ± 1	1.95 ± 0.02	8 ± 3	20.1 ± 0.1	46 ± 1	3.8 ± 0.2	121 ± 2	44 ± 3	32.2 ± 0.7	55 ± 3	247 ± 9	330 ± 10
15 Jul	0.910 ± 0.008	< 5.9	< SB	4.8 ± 0.1	28.5 ± 0.9	0.38 ± 0.02	8.71 ± 0.07	0.82 ± 0.01	< 0.063	3.69 ± 0.06	3.7 ± 0.1	2.6 ± 0.1	14.5 ± 0.4	< 6.5	< SB	< SB	2.3 ± 0.1	< SB
19 Jul	1.51 ± 0.01	< 5.9	< SB	18.5 ± 0.2	< 0.61	1.21 ± 0.06	38.8 ± 0.8	0.890 ± 0.007	10.5 ± 0.6	8.40 ± 0.06	16.9 ± 0.4	2.5 ± 0.2	40.87 ± 0.08	< SB	< SB	1.90 ± 0.08	3.0 ± 0.1	97 ± 3
23 Jul	4.10 ± 0.03	< 5.9	3.15 ± 0.05	30.3 ± 0.7	< 0.61	3.5 ± 0.1	35.6 ± 0.3	3.10 ± 0.08	27 ± 5	32.0 ± 0.2	61 ± 2	5.3 ± 0.2	191 ± 1	7.5 ± 0.2	< SB	105 ± 5	48 ± 3	< 56

SB = *Sample Blank*; *n.a.* = *not analysed. Values below the LOD were obtained by using the real solution volume and air volume of each sample.*

b)									2012	2 CAMPAIO	GN								
Start	Al	As	Ba	Ca	Cd	Co	Cu	Fe	In	K	Mg	Mn	Mo	Na	Ni	Pb	Ti	V .	Zn
date	(ng/m ³)	(pg/m ³)	(pg/m ³)	(ng/m ³)	(pg/m ³)	(pg/m ³)	(pg/m ³)	(ng/m ³)	(fg/m ³)	(ng/m ³)	(ng/m ³)	(pg/m ³)	(pg/m ³)	(ng/m ³)	(pg/m ³)	(pg/m ³)	(pg/m ³)	(pg/m ³)	(pg/m ³)
17 Apr	9.6 ± 0.3	48 ± 3	104 ± 4	21.1 ± 0.2	5.1 ± 0.8	3.2 ± 0.2	106 ± 2	8.7 ± 0.2	< 27	12.6 ± 0.1	38 ± 1	231 ± 5	11.4 ± 0.6	210 ± 20	77 ± 3	168 ± 6	370 ± 10	34 ± 2	560 ± 20
21 Apr	5.8 ± 0.2	25 ± 1	77 ± 2	16.4 ± 0.5	2.7 ± 0.4	1.19 ± 0.01	31.0 ± 0.9	5.6 ± 0.1	< 24	8.1 ± 0.1	22.5 ± 0.7	154 ± 4	5.4 ± 0.4	140 ± 20	28 ± 2	90 ± 1	224 ± 6	21 ± 2	251 ± 6
25 Apr	6.4 ± 0.2	24 ± 2	78 ± 3	17.3 ± 0.1	3.3 ± 0.7	1.23 ± 0.02	43 ± 2	6.32 ± 0.08	< 24	11.79 ± 0.03	35.2 ± 0.3	159 ± 5	5.8 ± 0.4	220 ± 10	40 ± 3	107 ± 2	278 ± 8	26.2 ± 0.8	300 ± 20
29 Apr	7.4 ± 0.4	26 ± 3	105 ± 4	29.2 ± 0.2	3.7 ± 0.4	2.36 ± 0.08	82 ± 2	6.74 ± 0.07	< 24	20.8 ± 0.2	68.1 ± 0.9	191 ± 8	6.9 ± 0.4	398 ± 5	52 ± 2	109 ± 3	280 ± 10	28 ± 1	310 ± 10
03 May	4.5 ± 0.4	16 ± 2	64 ± 2	15.9 ± 0.2	2.6 ± 0.4	3.44 ± 0.07	51.3 ± 0.8	4.72 ± 0.08	420 ± 60	12.6 ± 0.4	34 ± 1	121 ± 4	6.2 ± 0.6	230 ± 10	12.7 ± 0.4	40 ± 1	211 ± 6	20 ± 3	260 ± 10
07 May	8.3 ± 0.2	21 ± 1	89 ± 3	12.3 ± 0.1	2.4 ± 0.3	3.4 ± 0.2	24 ± 1	7.86 ± 0.09	350 ± 40	7.17 ± 0.07	15.0 ± 0.2	213 ± 7	6.4 ± 0.5	75 ± 2	9.2 ± 0.5	49 ± 1	370 ± 20	24 ± 1	250 ± 10
11 May	7.4 ± 0.1	15.9 ± 0.8	91 ± 2	14.0 ± 0.3	2.6 ± 0.5	3.8 ± 0.2	65 ± 3	6.8 ± 0.2	460 ± 50	7.05 ± 0.04	16.6 ± 0.4	194 ± 6	4.8 ± 0.3	89 ± 4	12.0 ± 0.8	73 ± 1	317 ± 6	27.6 ± 0.4	247 ± 5
15 May	14.8 ± 0.2	28 ± 2	146 ± 4	16.76 ± 0.06	4.2 ± 0.4	7.1 ± 0.2	45 ± 1	12.4 ± 0.3	420 ± 30	11.9 ± 0.2	20.4 ± 0.3	410 ± 20	7.6 ± 0.6	112 ± 3	21.3 ± 0.8	126 ± 4	680 ± 20	47 ± 3	380 ± 20
19 May	11.8 ± 0.5	20 ± 1	136 ± 5	22.3 ± 0.3	2.2 ± 0.4	4.8 ± 0.3	52 ± 1	9.7 ± 0.4	340 ± 60	11.6 ± 0.3	28.3 ± 0.5	270 ± 10	6.0 ± 0.5	170 ± 10	5.2 ± 0.4	36 ± 2	510 ± 10	32 ± 2	240 ± 10
23 May	7.49 ± 0.05	14 ± 1	103 ± 2	20.82 ± 0.08	2.2 ± 0.5	1.7 ± 0.1	54 ± 1	7.1 ± 0.3	20 ± 3	9.99 ± 0.09	20 ± 1	219 ± 6	4.1 ± 0.5	132 ± 4	60 ± 1	74 ± 1	330 ± 10	14.2 ± 0.1	270 ± 10
27 May	5.6 ± 0.2	10 ± 2	59 ± 1	30.4 ± 0.5	2.0 ± 0.3	12.2 ± 0.3	97 ± 2	9.9 ± 0.3	360 ± 40	21.4 ± 0.2	74 ± 3	207 ± 7	20.9 ± 0.9	510 ± 20	410 ± 20	< 0.030	260 ± 10	22.6 ± 0.5	220 ± 10
31 May	5.7 ± 0.2	4.4 ± 0.5	48 ± 1	14.62 ± 0.09	1.3 ± 0.4	6.2 ± 0.2	54 ± 1	6.8 ± 0.1	< 27	6.05 ± 0.08	18.2 ± 0.5	148 ± 2	2.9 ± 0.4	110 ± 3	92 ± 2	28.9 ± 0.7	258 ± 3	15.2 ± 0.2	145 ± 9
04 Jun	5.0 ± 0.1	8.2 ± 0.4	68 ± 2	16.0 ± 0.2	0.7 ± 0.2	1.60 ± 0.08	45 ± 2	4.73 ± 0.06	< 26	7.7 ± 0.2	22.7 ± 0.9	125 ± 4	3.2 ± 0.3	158 ± 3	87 ± 3	37.4 ± 0.7	238 ± 6	11.3 ± 0.3	121 ± 6
08 Jun	2.8 ± 0.1	5.5 ± 0.4	37 ± 1	16.3 ± 0.2	1.9 ± 0.5	3.7 ± 0.2	7.4 ± 0.2	2.71 ± 0.09	200 ± 20	10.4 ± 0.1	31 ± 1	64 ± 2	2.8 ± 0.2	250 ± 10	84 ± 2	17.7 ± 0.6	127 ± 5	71 ± 2	68 ± 5
12 Jun	6.1 ± 0.3	9.7 ± 0.7	83 ± 2	26.44 ± 0.08	1.8 ± 0.5	2.55 ± 0.06	60 ± 2	5.7 ± 0.2	410 ± 20	4.5 ± 0.2	12.3 ± 0.3	139 ± 6	7.3 ± 0.9	47 ± 1	31 ± 2	< 0.033	210 ± 10	87 ± 4	450 ± 30
16 Jun	2.37 ± 0.08	5.0 ± 0.5	36.4 ± 0.8	11.6 ± 0.1	1.2 ± 0.3	0.26 ± 0.02	49 ± 1	2.40 ± 0.05	< 27	7.34 ± 0.09	22.7 ± 0.6	65.8 ± 0.8	2.1 ± 0.2	173 ± 3	350 ± 10	24.3 ± 0.5	95 ± 4	13.3 ± 0.2	97 ± 5
20 Jun	1.76 ± 0.05	5.3 ± 0.7	37 ± 1	25.8 ± 0.3	1.2 ± 0.1	1.29 ± 0.04	56 ± 3	2.58 ± 0.06	470 ± 40	12.41 ± 0.09	42.5 ± 0.9	57 ± 2	7.6 ± 0.8	341 ± 7	49 ± 2	< 0.033	92 ± 5	13.4 ± 0.8	145 ± 5
24 Jun	0.290 ± 0.007	3.1 ± 0.8	12.8 ± 0.3	8.83 ± 0.06	2.2 ± 0.9	< 0.016	5.2 ± 0.2	0.34 ± 0.01	160 ± 7	0.39 ± 0.02	0.96 ± 0.02	3.95 ± 0.08	0.36 ± 0.07	2.93 ± 0.05	37.8 ± 0.8	< 0.034	2.9 ± 0.2	< 0.21	128 ± 5
28 Jun	0.94 ± 0.01	6 ± 1	23.2 ± 0.5	5.727 ± 0.009	0.4 ± 0.1	12.5 ± 0.5	5.8 ± 0.1	1.66 ± 0.04	< 27	2.49 ± 0.06	7.9 ± 0.1	25 ± 1	2.2 ± 0.2	58.8 ± 0.7	116 ± 2	10.6 ± 0.6	37 ± 1	128 ± 3	80 ± 3
02 Jul	4.55 ± 0.09	8.3 ± 0.6	46 ± 1	17.64 ± 0.04	0.7 ± 0.1	3.8 ± 0.1	68 ± 2	5.31 ± 0.08	< 28	10.2 ± 0.2	31.9 ± 0.7	110 ± 4	5.8 ± 0.4	237 ± 3	99 ± 3	41 ± 1	300 ± 100	34.6 ± 0.5	340 ± 10
06 Jul	5.3 ± 0.2	6.2 ± 0.8	54 ± 2	20.8 ± 0.1	1.6 ± 0.4	4.6 ± 0.2	56 ± 2	5.88 ± 0.05	330 ± 50	12.5 ± 0.2	39.0 ± 0.7	123 ± 1	11.4 ± 0.7	300 ± 10	200 ± 10	< 0.032	180 ± 10	128 ± 9	116 ± 9
10 Jul	7.7 ± 0.2	7 ± 1	80 ± 2	22.6 ± 0.5	1.1 ± 0.2	2.9 ± 0.1	42.9 ± 0.9	7.7 ± 0.2	300 ± 100	9.2 ± 0.3	28.8 ± 0.5	163 ± 4	4.1 ± 0.4	180 ± 6	9.4 ± 0.5	50 ± 2	260 ± 10	32 ± 1	62 ± 3
14 Jul	3.1 ± 0.2	6 ± 1	43.9 ± 0.8	13.72 ± 0.09	0.5 ± 0.1	0.95 ± 0.07	18.4 ± 0.4	2.95 ± 0.09	< 27	5.312 ± 0.009	14 ± 2	64 ± 2	3.2 ± 0.5	91 ± 4	64 ± 2	2.07 ± 0.05	175 ± 7	64 ± 1	230 ± 10
18 Jul	6.8 ± 0.1	10.3 ± 0.9	51 ± 1	17.0 ± 0.2	0.22 ± 0.04	10.5 ± 0.4	44 ± 1	7.3 ± 0.2	< 28	8.25 ± 0.07	24.4 ± 0.5	141 ± 2	4.9 ± 0.4	161 ± 4	396 ± 7	2.87 ± 0.06	327 ± 5	80.3 ± 0.5	250 ± 10
22 Jul	5.5 ± 0.1	10.9 ± 0.9	64 ± 2	19.9 ± 0.4	1.5 ± 0.2	6.8 ± 0.3	70 ± 2	5.4 ± 0.2	350 ± 30	10.4 ± 0.1	33.6 ± 0.7	110 ± 2	9.8 ± 0.8	250 ± 20	247 ± 9	< 0.032	200 ± 10	580 ± 20	157 ± 9
26 Jul	7.1 ± 0.3	11.6 ± 0.5	84 ± 2	19.6 ± 0.3	1.0 ± 0.4	5.8 ± 0.3	192 ± 3	7.29 ± 0.08	380 ± 30	5.7 ± 0.1	17.0 ± 0.4	144 ± 1	6.7 ± 0.5	80 ± 4	87 ± 3	< 0.032	259 ± 9	177 ± 9	350 ± 20
30 Jul	3.89 ± 0.05	4.6 ± 0.8	35.6 ± 0.8	13.8 ± 0.1	0.7 ± 0.2	2.6 ± 0.1	32.8 ± 0.6	4.19 ± 0.09	< 25	7.01 ± 0.07	22.6 ± 0.3	82 ± 1	5.6 ± 0.5	163 ± 3	134 ± 3	38.9 ± 0.9	169 ± 2	10.2 ± 0.3	78 ± 3
03 Aug	1.40 ± 0.03	5.8 ± 0.4	37.4 ± 0.8	8.84 ± 0.09	0.7 ± 0.1	0.60 ± 0.02	78 ± 2	1.23 ± 0.01	30 ± 2	1.97 ± 0.08	8.5 ± 0.3	23.4 ± 0.5	7.1 ± 0.7	52.3 ± 0.2	110 ± 3	37 ± 1	25 ± 2	18.0 ± 0.3	251 ± 10
07 Aug	2.7 ± 0.1	5.0 ± 0.3	53 ± 1	12.8 ± 0.1	< 0.0035	2.5 ± 0.1	44.7 ± 0.7	3.58 ± 0.09	< 26	6.2 ± 0.3	19.0 ± 0.4	68 ± 2	10.4 ± 0.6	137 ± 4	218 ± 4	66 ± 1	123 ± 2	8.1 ± 0.3	109 ± 5
11 Aug	0.44 ± 0.03	2.8 ± 0.2	21 ± 1	7.81 ± 0.09	1.0 ± 0.3	< 0.017	13.9 ± 0.3	1.41 ± 0.03	300 ± 100	4.08 ± 0.01	13.1 ± 0.3	22.6 ± 0.8	3.7 ± 0.3	105 ± 4	97 ± 2	< 0.036	20 ± 1	< 0.23	7.1 ± 0.5
15 Aug	1.50 ± 0.07	3.1 ± 0.6	16.0 ± 0.6	9.36 ± 0.04	0.9 ± 0.3	0.20 ± 0.01	14.5 ± 0.4	1.93 ± 0.04	290 ± 40	4.4 ± 0.1	15.5 ± 0.4	37 ± 1	2.4 ± 0.2	100 ± 5	< 0.029	< 0.030	85 ± 6	6.7 ± 0.1	59 ± 4
19 Aug	3.51 ± 0.07	2.4 ± 0.7	41 ± 2	14.00 ± 0.06	< 0.0035	0.41 ± 0.02	15.1 ± 0.3	3.81 ± 0.05	< 26	8.09 ± 0.07	24.4 ± 0.9	82.9 ± 0.7	1.2 ± 0.2	191 ± 7	127 ± 4	< 0.031	164 ± 7	7.67 ± 0.07	15.2 ± 0.8
23 Aug	2.65 ± 0.06	1.8 ± 0.3	44 ± 1	9.43 ± 0.08	0.9 ± 0.2	< 0.015	32.3 ± 0.5	2.50 ± 0.03	< 27	4.45 ± 0.09	12.9 ± 0.4	61 ± 2	1.3 ± 0.2	85 ± 2	61 ± 2	1.94 ± 0.05	108 ± 2	3.82 ± 0.07	60 ± 4
27 Aug	1.24 ± 0.03	12 ± 1	24.6 ± 0.7	18.2 ± 0.1	1.5 ± 0.3	0.93 ± 0.03	75 ± 1	2.01 ± 0.05	400 ± 60	11.92 ± 0.06	36.2 ± 0.7	33 ± 2	6.7 ± 0.5	267 ± 9	48 ± 3	< 0.031	34 ± 3	9.2 ± 0.1	49 ± 2
31 Aug	1.17 ± 0.02	4.3 ± 0.3	22.3 ± 0.6	14.7 ± 0.2	0.7 ± 0.2	< 0.014	4.0 ± 0.1	1.48 ± 0.04	< 26	12.7 ± 0.2	34.2 ± 0.7	33.4 ± 0.9	3.1 ± 0.3	321 ± 4	53 ± 4	< 0.031	140 ± 10	7.8 ± 0.3	380 ± 10
04 Sep	0.80 ± 0.02	8 ± 1	21.7 ± 0.5	10.2 ± 0.1	0.4 ± 0.1	< 0.015	372 ± 7	1.09 ± 0.03	< 27	5.9 ± 0.1	20.8 ± 0.4	15.7 ± 0.6	1.7 ± 0.2	151 ± 2	201 ± 6	< 0.031	25 ± 2	3.71 ± 0.07	570 ± 20

Values below the LOD were obtained by using the real solution volume and air volume of each sample.

c)									2013 C	AMPAIGN	-								
Start date	Al (ng/m ³)	As (pg/m ³)	Ba (pg/m ³)	Ca (ng/m ³)	Cd (pg/m ³)	Co (pg/m ³)	Cu (pg/m ³)	Fe (ng/m ³)	In (fg/m ³)	K (ng/m ³)	Mg (ng/m ³)	Mn (pg/m ³)	Mo (pg/m ³)	Na (ng/m ³)	Ni (pg/m ³)	Pb (pg/m ³)	Ti (pg/m ³)	V (pg/m ³)	Zn (pg/m ³)
01 May	11.7 ± 0.2	27 ± 1	170 ± 7	33.9 ± 0.5	2.95 ± 0.2	5.2 ± 0.1	65 ± 1	8.6 ± 0.1	192 ± 5	2.69 ± 0.02	12.9 ± 0.1	209 ± 8	49 ± 2	176 ± 3	10.1 ± 0.2	112 ± 8	500 ± 10	24.0 ± 0.4	430 ± 20
05 May	10.5 ± 0.1	27 ± 1	140 ± 2	17.2 ± 0.3	6.61 ± 0.08	4.5 ± 0.1	117 ± 1	6.35 ± 0.06	178 ± 4	3.88 ± 0.03	36.2 ± 0.1	204 ± 2	34 ± 1	67 ± 1	79 ± 3	206 ± 8	420 ± 10	23.5 ± 0.4	1120 ± 20
09 May	0.97 ± 0.02	1.4 ± 0.2	48.8 ± 0.6	4.57 ± 0.06	< SB	0.88 ± 0.03	91 ± 5	0.63 ± 0.01	< 130	< SB	34.18 ± 0.07	24.6 ± 0.5	4.2 ± 0.2	36.1 ± 0.5	155 ± 5	< 2.8	49.2 ± 0.7	< 0.32	290 ± 10
13 May	9.4 ± 0.1	11.6 ± 0.6	119 ± 2	28.9 ± 0.3	< SB	7.19 ± 0.08	470 ± 20	9.6 ± 0.2	< 150	3.30 ± 0.04	20.8 ± 0.3	144 ± 2	81 ± 1	126 ± 1	360 ± 10	32 ± 2	396 ± 5	7.81 ± 0.08	270 ± 20
17 May	6.9 ± 0.5	16 ± 2	40 ± 2	14.6 ± 0.2	< 0.49	5.7 ± 0.2	< 41	5.9 ± 0.6	1170 ± 10	< SB	29.5 ± 0.3	141 ± 5	22 ± 1	170 ± 1	< 22	19 ± 1	810 ± 40	19.2 ± 0.8	78 ± 1
21 May	5.72 ± 0.09	9.0 ± 0.4	81 ± 2	20.8 ± 0.4	< SB	3.7 ± 0.2	59 ± 3	3.05 ± 0.09	< 140	< SB	13.2 ± 0.2	94 ± 5	34 ± 1	99 ± 1	37 ± 2	17.7 ± 0.9	350 ± 20	17.3 ± 0.7	87 ± 4
25 May	7.3 ± 0.4	4.2 ± 0.2	< 43	11.617 ± 0.009	< SB	2.18 ± 0.05	< 48	1.97 ± 0.04	720 ± 70	< SB	7.8 ± 0.2	40 ± 3	118 ± 4	128 ± 3	21.9 ± 0.5	< 3.5	110 ± 8	3.13 ± 0.05	163 ± 7
29 May	3.35 ± 0.08	6.2 ± 0.3	49 ± 2	21.3 ± 0.5	0.27 ± 0.01	3.3 ± 0.1	75 ± 4	1.51 ± 0.05	< 130	0.453 ± 0.003	21.0 ± 0.2	57.2 ± 0.9	17 ± 1	127.1 ± 0.5	71 ± 3	26.6 ± 0.6	165 ± 9	5.6 ± 0.2	610 ± 20
02 Jun	9.35 ± 0.06	13 ± 1	107 ± 2	28.9 ± 0.4	1.50 ± 0.09	10.4 ± 0.2	80 ± 10	11.06 ± 0.02	< 140	13.8 ± 0.1	25.0 ± 0.5	189 ± 3	22.1 ± 0.8	290 ± 1	154 ± 6	42 ± 2	477 ± 8	19.7 ± 0.1	520 ± 40
06 Jun	25.0 ± 0.9	12.3 ± 0.7	168 ± 3	75.3 ± 0.5	< SB	9.1 ± 0.4	47 ± 2	1.59 ± 0.05	1070 ± 10	30.3 ± 0.4	17 ± 3	300 ± 10	13.8 ± 0.5	1012 ± 7	55.1 ± 0.7	44 ± 1	1110 ± 70	58 ± 2	132 ± 7
10 Jun	5.5 ± 0.3	5.4 ± 0.5	56 ± 2	27.3 ± 0.2	< SB	2.77 ± 0.09	< 39	1.15 ± 0.07	< 130	5.78 ± 0.02	16.5 ± 0.2	71 ± 2	13.8 ± 0.9	306 ± 1	50 ± 2	15.3 ± 0.9	222 ± 9	6.0 ± 0.2	340 ± 10
14 Jun	1.51 ± 0.05	1.9 ± 0.2	33.8 ± 0.9	15.3 ± 0.2	4.2 ± 0.4	1.47 ± 0.04	42 ± 1	0.59 ± 0.02	152 ± 1	23.24 ± 0.05	19.6 ± 0.2	17.7 ± 0.5	8.7 ± 0.3	98.1 ± 0.9	< 22	< 2.9	68 ± 2	0.78 ± 0.02	240 ± 10
18 Jun	3.2 ± 0.2	2.1 ± 0.2	47 ± 2	9.13 ± 0.04	< SB	1.5 ± 0.1	107.8 ± 0.1	2.4 ± 0.2	1100 ± 200	< SB	42.27 ± 0.05	53 ± 3	52 ± 2	50.7 ± 0.7	31.4 ± 0.3	61 ± 2	103 ± 4	6.6 ± 0.3	220 ± 20
22 Jun	3.1 ± 0.1	15 ± 1	43 ± 1	16.7 ± 0.1	< 0.51	2.1 ± 0.1	32 ± 1	0.37 ± 0.01	< 140	< SB	143.0 ± 0.1	34.4 ± 0.8	31 ± 1	127 ± 2	108 ± 4	< 3.0	143 ± 6	70 ± 4	157 ± 4
26 Jun	5.7 ± 0.1	8.7 ± 0.5	66 ± 2	18.9 ± 0.1	0.63 ± 0.04	2.92 ± 0.08	88 ± 2	3.25 ± 0.05	< 140	0.825 ± 0.007	42.7 ± 0.3	90.9 ± 0.4	11.4 ± 0.7	192 ± 2	32 ± 2	11.8 ± 0.6	460 ± 10	18.9 ± 0.6	144 ± 3
30 Jun	4.36 ± 0.04	13.5 ± 0.1	88 ± 1	8.52 ± 0.08	< SB	4.0 ± 0.2	< 51	6.19 ± 0.07	< 170	< SB	15.1 ± 0.3	99.5 ± 0.7	16.1 ± 0.5	111 ± 2	61 ± 2	380 ± 10	419 ± 7	36.4 ± 0.2	710 ± 30
04 Jul	4.40 ± 0.08	13.9 ± 0.6	112 ± 1	17.7 ± 0.1	0.87 ± 0.05	3.1 ± 0.2	42.6 ± 0.8	6.7 ± 0.2	< 140	4.19 ± 0.04	8.0 ± 0.2	108 ± 1	21.9 ± 0.7	93 ± 3	37 ± 2	412 ± 6	200 ± 3	16.1 ± 0.3	500 ± 40
08 Jul	2.1 ± 0.2	2.2 ± 0.4	< 38	13.9 ± 0.1	< 0.51	2.07 ± 0.04	< 42	3.1 ± 0.7	1100 ± 300	1.44 ± 0.01	18.90 ± 0.02	69 ± 2	15.0 ± 0.7	207 ± 2	42 ± 1	2.75 ± 0.02	160 ± 10	25.3 ± 0.2	< 47
12 Jul	62.2 ± 0.9	14.3 ± 0.9	1020 ± 20	68 ± 1	< SB	15.0 ± 0.3	49 ± 3	56.2 ± 0.8	38 ± 3	23.8 ± 0.1	28.1 ± 0.4	1320 ± 20	69 ± 1	233 ± 2	109 ± 4	338 ± 6	2610 ± 60	57 ± 1	1150 ± 30
16 Jul	36.6 ± 0.1	10.7 ± 0.9	701 ± 9	56 ± 1	2.1 ± 0.1	11.5 ± 0.3	88 ± 7	34.6 ± 0.5	63 ± 2	43.4 ± 0.4	14.5 ± 0.5	870 ± 20	101 ± 2	203 ± 4	312 ± 9	460 ± 20	1330 ± 20	30 ± 1	1000 ± 30
20 Jul	2.16 ± 0.02	4.1 ± 0.6	< 36	3.75 ± 0.02	< 0.49	0.96 ± 0.07	< 41	2.57 ± 0.04	< 140	< SB	15.5 ± 0.1	35.1 ± 0.4	75 ± 2	50.3 ± 0.6	< 22	< 2.9	47 ± 1	1.26 ± 0.03	< 45
24 Jul	6.0 ± 0.3	4.4 ± 0.3	64 ± 1	16.42 ± 0.03	< SB	2.62 ± 0.06	42 ± 1	3.26 ± 0.09	< 140	< SB	27.2 ± 0.1	75.0 ± 0.2	87 ± 2	88.3 ± 0.9	96 ± 4	< 3.0	195 ± 5	45 ± 1	255 ± 7
28 Jul	6.37 ± 0.06	43 ± 4	74 ± 2	15.5 ± 0.1	14.5 ± 0.8	4.3 ± 0.2	131 ± 8	4.7 ± 0.2	1100 ± 20	< SB	48.39 ± 0.04	86 ± 9	28.0 ± 0.7	30.6 ± 0.3	106 ± 4	159 ± 5	2200 ± 200	38 ± 2	520 ± 20
05 Aug	3.99 ± 0.09	4.9 ± 0.3	116 ± 2	25.5 ± 0.3	< 0.55	1.37 ± 0.08	< 45	2.92 ± 0.01	< 150	3.08 ± 0.02	44.29 ± 0.04	55 ± 1	85 ± 2	68.5 ± 0.6	17.7 ± 0.9	< 3.3	147 ± 4	4.07 ± 0.02	270 ± 10
09 Aug	6.22 ± 0.05	0.25 ± 0.01	37 ± 1	29.9 ± 0.4	< 0.53	2.58 ± 0.07	< 44	2.6 ± 0.3	1210 ± 50	1.83 ± 0.02	7.79 ± 0.04	47 ± 1	123 ± 6	327 ± 8	90 ± 1	12 ± 1	80 ± 10	2.44 ± 0.03	343 ± 20
13 Aug	18.3 ± 0.8	11.9 ± 0.6	109 ± 2	26.4 ± 0.1	0.8 ± 0.2	4.9 ± 0.1	84 ± 5	9.0 ± 0.4	1200 ± 100	3.71 ± 0.06	13.2 ± 0.5	210 ± 20	32 ± 2	225 ± 2	< 22	62.7 ± 0.8	720 ± 70	25.9 ± 0.2	194 ± 6
17 Aug	3.03 ± 0.07	40 ± 1	77 ± 1	16.6 ± 0.1	1.22 ± 0.09	3.06 ± 0.06	88 ± 6	2.69 ± 0.02	30.8 ± 0.5	< SB	5.6 ± 0.2	53.8 ± 0.5	5.3 ± 0.4	78 ± 2	41 ± 2	124 ± 2	130 ± 3	11.3 ± 0.2	232 ± 8
21 Aug	0.812 ± 0.007	3.5 ± 0.1	< 35	5.73 ± 0.07	< 0.46	0.55 ± 0.04	< 38	0.683 ± 0.008	< 130	< SB	12.1 ± 0.1	15.6 ± 0.2	4.4 ± 0.2	84 ± 2	< 21	< 2.8	27.9 ± 0.7	0.84 ± 0.05	< 43
25 Aug	0.298 ± 0.008	< 0.31	< 35	7.4 ± 0.1	6.3 ± 0.3	2.4 ± 0.4	42 ± 2	0.57 ± 0.06	190 ± 10	0.907 ± 0.007	41.60 ± 0.03	< 16	5.2 ± 0.4	25.1 ± 0.4	< 21	< 2.8	< 30	< 0.32	< 44
29 Aug	0.527 ± 0.006	5.9 ± 0.6	< 39	10.1 ± 0.2	< 0.53	0.60 ± 0.06	< 44	2.09 ± 0.03	< 150	< SB	35.5 ± 0.1	19.2 ± 0.1	4.8 ± 0.2	88 ± 2	57 ± 5	< 3.1	43.1 ± 0.9	0.78 ± 0.01	< 49
02 Sep	4.9 ± 0.1	5.9 ± 0.3	86 ± 3	21.2 ± 0.1	< SB	3.80 ± 0.04	44 ± 4	6.7 ± 0.3	0.11 ± 0.01	2.36 ± 0.03	12.6 ± 0.3	125 ± 2	5.4 ± 0.3	183 ± 1	39 ± 3	380 ± 20	640 ± 30	17.0 ± 0.6	98 ± 4
06 Sep	26.7 ± 0.9	3.8 ± 0.2	177 ± 5	68.0 ± 0.4	< SB	14.1 ± 0.6	110 ± 10	17.9 ± 0.6	1100 ± 100	26.6 ± 0.2	12 ± 2	460 ± 10	8.6 ± 0.3	930 ± 10	159 ± 7	22.1 ± 0.2	1700 ± 100	64 ± 1	380 ± 10
10 Sep	0.315 ± 0.005	< 0.41	< 46	10.76 ± 0.05	< 0.62	4.8 ± 0.4	41 ± 2	0.55 ± 0.09	< 170	0.489 ± 0.005	4.73 ± 0.07	< 21	3.8 ± 0.2	29.0 ± 0.1	< 28	29.6 ± 0.7	< 39	< 0.41	440 ± 30

SB = Sample Blank. Values below the LOD were obtain	iea by using	g the real :	solution vo	ните апа с	ar voiume (of eaci	і затри
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Supplementary Table 2. Lanthanide concentrations determined in PM₁₀ samples collected in a) 2011; b) 2012; c) 2013.

a)					2	011 CAN	MPAIGN					
Start	La	Ce	Pr (for loss ³)	Nd	Sm	Eu (fer (m ³)	Gd	Tb	Dy	Ho (factor ³)	Er	Yb
date	(Ig/m ⁻)	(Ig/m ⁻)	(Ig/m ⁻)	(Ig/m ⁻)	(Ig/m [*])	(Ig/m ⁻)	(ig/m [*])	(Ig/m ⁻)	(Ig/m ⁻)	(Ig/m ⁻)	(Ig/m [*])	(Ig/m [*])
29 Mar	930 ± 70	1180 ± 40	189 ± 5	760 ± 20	157 ± 6	37.8 ± 0.9	200 ± 20	15 ± 1	110 ± 6	3.3 ± 0.1	47 ± 4	38 ± 3
03 Apr	1640 ± 30	4100 ± 200	360 ± 10	1370 ± 20	270 ± 10	62 ± 4	336 ± 6	37 ± 1	220 ± 10	24.8 ± 0.3	104 ± 5	79 ± 2
07 Apr	1740 ± 60	2850 ± 70	383 ± 4	1440 ± 30	300 ± 10	58 ± 2	290 ± 10	33 ± 3	195 ± 8	21.2 ± 0.8	77 ± 6	54 ± 2
11 Apr	< 10	247 ± 9	< 3.6	< 15	< 4.1	< 5.7	< 5.9	3.9 ± 0.2	11.8 ± 0.8	14.9 ± 0.9	21 ± 2	23 ± 2
01 May	620 ± 20	313 ± 6	201 ± 5	542 ± 6	163 ± 7	93 ± 5	176 ± 7	83 ± 6	134 ± 4	69 ± 2	92 ± 5	82 ± 3
05 May	2020 ± 40	3090 ± 60	531 ± 8	1660 ± 40	430 ± 20	190 ± 4	410 ± 10	150 ± 3	320 ± 10	131 ± 6	200 ± 10	182 ± 6
09 May	820 ± 30	3100 ± 100	440 ± 20	610 ± 30	380 ± 20	300 ± 10	370 ± 20	270 ± 20	280 ± 20	260 ± 20	240 ± 20	240 ± 20
13 May	1520 ± 60	11400 ± 400	353 ± 9	1350 ± 20	280 ± 30	77 ± 1	340 ± 10	39.7 ± 0.6	184 ± 2	21.8 ± 0.7	82 ± 2	66 ± 4
17 May	1600 ± 60	2400 ± 70	360 ± 10	1300 ± 20	286 ± 5	71 ± 2	280 ± 20	35.0 ± 0.5	170 ± 10	23 ± 2	73 ± 5	59 ± 4
21 May	250 ± 9	350 ± 10	60 ± 3	230 ± 10	47 ± 3	< 14	48 ± 3	14.7 ± 0.9	30 ± 2	27 ± 2	21 ± 2	8.9 ± 0.7
25 May	430 ± 20	330 ± 10	93 ± 3	400 ± 10	69 ± 5	13 ± 1	43 ± 4	< 2.0	52.8 ± 0.9	< 1.4	13 ± 1	8.0 ± 0.6
29 May	208 ± 3	< 130	61 ± 2	200 ± 10	48 ± 1	39 ± 2	37 ± 4	18.7 ± 0.2	40 ± 3	18 ± 1	27 ± 2	27 ± 3
05 Jun	393 ± 8	3200 ± 100	111 ± 3	370 ± 10	98 ± 5	51 ± 2	105 ± 8	22 ± 1	60 ± 4	19.6 ± 0.6	39 ± 6	33 ± 1
09 Jun	760 ± 20	900 ± 10	175 ± 7	650 ± 50	130 ± 10	73 ± 2	150 ± 10	20 ± 1	86 ± 9	18 ± 2	39 ± 4	37 ± 9
13 Jun	2530 ± 90	4200 ± 200	620 ± 30	2400 ± 100	570 ± 30	186 ± 8	610 ± 40	136 ± 8	390 ± 20	103 ± 6	200 ± 20	150 ± 10
17 Jun	1149 ± 9	4100 ± 100	249 ± 8	920 ± 30	195 ± 9	67 ± 4	220 ± 10	35 ± 3	130 ± 10	24 ± 2	71 ± 6	54 ± 7
21 Jun	161 ± 5	< 140	30 ± 3	88 ± 6	19 ± 2	34.5 ± 0.7	35 ± 4	7.9 ± 0.2	18 ± 2	5.7 ± 0.7	22 ± 1	9.3 ± 0.9
25 Jun	197 ± 6	260 ± 20	72 ± 2	220 ± 10	60 ± 4	45 ± 2	88 ± 8	15 ± 0.7	46 ± 2	16 ± 0.6	30 ± 4	30 ± 2
29 Jun	220 ± 20	298 ± 8	67 ± 4	230 ± 10	45 ± 3	27 ± 2	59 ± 3	8.7 ± 0.6	35 ± 4	15 ± 3	28 ± 2	18 ± 1
03 Jul	350 ± 10	400 ± 10	80 ± 2	240 ± 10	55 ± 5	40 ± 1	67 ± 7	10.9 ± 0.6	35 ± 3	10.8 ± 0.9	24 ± 1	21 ± 1
07 Jul	330 ± 10	450 ± 30	121 ± 7	340 ± 30	94 ± 2	72 ± 6	129 ± 7	53 ± 3	80 ± 7	50 ± 5	67 ± 6	54 ± 3
15 Jul	38 ± 1	1380 ± 80	17 ± 1	49 ± 3	< 12	18.2 ± 0.6	19.9 ± 0.5	3.5 ± 0.3	8 ± 1	1.60 ± 0.01	6.4 ± 0.3	14 ± 1
19 Jul	127 ± 6	170 ± 20	55 ± 7	120 ± 10	38 ± 3	45 ± 2	45.4 ± 0.8	26 ± 1	30 ± 1	26 ± 1	29 ± 3	26 ± 3
23 Jul	440 ± 20	620 ± 20	100 ± 10	380 ± 30	70 ± 4	45 ± 3	92 ± 3	17 ± 3	51 ± 2	9.9 ± 0.2	32 ± 3	19 ± 2

Values below the LOD were obtained by using the real solution volume and air volume of each sample.

b)					20	12 CA	MPAIG	N				
Start date	La (fg/m ³)	Ce (fg/m ³)	Pr (fg/m ³)	Nd (fg/m ³)	Sm (fg/m ³)	Eu (fg/m ³)	Gd (fg/m ³)	Tb (fg/m ³)	Dy (fg/m ³)	Ho (fg/m ³)	Er (fg/m ³)	Yb (fg/m ³)
17 Apr	1680 ± 70	3500 ± 200	360 ± 20	1440 ± 90	271 ± 7	94 ± 7	260 ± 30	42 ± 1	196 ± 9	33 ± 2	101 ± 7	84 ± 6
21 Apr	1150 ± 50	2600 ± 100	270 ± 20	1110 ± 60	219 ± 8	70 ± 3	230 ± 10	34.7 ± 0.9	156 ± 8	27.1 ± 0.4	78 ± 5	61 ± 5
25 Apr	1170 ± 30	8700 ± 100	261 ± 8	1010 ± 50	200 ± 20	93 ± 6	230 ± 7	33 ± 2	144 ± 6	26 ± 1	77 ± 3	67 ± 5
29 Apr	1340 ± 30	3100 ± 100	323 ± 9	1220 ± 30	240 ± 20	79 ± 6	250 ± 10	42 ± 2	183 ± 9	34 ± 3	89 ± 6	76 ± 7
03 May	1040 ± 90	2160 ± 70	230 ± 7	910 ± 40	350 ± 30	58 ± 5	190 ± 10	24 ± 2	130 ± 10	18.9 ± 0.5	66 ± 5	56 ± 5
07 May	1830 ± 70	12200 ± 500	435 ± 3	1640 ± 60	330 ± 10	85 ± 3	390 ± 20	54 ± 2	230 ± 7	37 ± 2	119 ± 3	96 ± 4
11 May	1700 ± 100	3500 ± 100	388 ± 7	1490 ± 50	294 ± 9	81 ± 4	290 ± 10	48 ± 1	189 ± 6	32 ± 1	92 ± 4	87 ± 4
15 May	3400 ± 100	9300 ± 500	806 ± 2	2790 ± 80	570 ± 30	144 ± 7	610 ± 20	94.3 ± 0.8	410 ± 10	66.9 ± 0.4	194 ± 7	160 ± 10
19 May	2160 ± 50	5000 ± 200	550 ± 20	2030 ± 60	430 ± 20	114 ± 5	430 ± 10	69 ± 4	310 ± 20	48.5 ± 0.1	158 ± 4	130 ± 10
23 May	1640 ± 30	3500 ± 100	380 ± 8	1380 ± 20	270 ± 20	76 ± 3	280 ± 20	39 ± 1	216 ± 9	36 ± 2	100 ± 30	100 ± 10
27 May	1040 ± 40	1900 ± 40	190 ± 10	850 ± 30	160 ± 20	48 ± 3	162 ± 5	25 ± 2	110 ± 10	15.8 ± 0.4	52 ± 5	45 ± 3
31 May	700 ± 50	8400 ± 100	162 ± 4	700 ± 100	140 ± 20	51 ± 8	150 ± 20	26 ± 6	97 ± 6	18 ± 4	47 ± 3	51 ± 6
04 Jun	880 ± 90	1980 ± 50	207 ± 7	770 ± 30	170 ± 10	49 ± 3	134 ± 3	21.9 ± 0.2	111 ± 8	16 ± 1	55 ± 5	51 ± 4
08 Jun	730 ± 60	1290 ± 30	131 ± 6	471 ± 9	107 ± 7	40 ± 2	100 ± 20	14.5 ± 0.4	77 ± 4	10 ± 1	33 ± 3	29 ± 2
12 Jun	1800 ± 100	7700 ± 300	373 ± 3	1490 ± 50	310 ± 20	79 ± 3	310 ± 20	47 ± 2	210 ± 7	33 ± 2	102 ± 3	110 ± 30
16 Jun	510 ± 20	1200 ± 100	120.2 ± 0.7	420 ± 20	86 ± 6	34 ± 3	84 ± 5	12.23 ± 0.07	62 ± 3	10 ± 1	32 ± 3	28 ± 2
20 Jun	620 ± 30	1330 ± 50	124 ± 7	490 ± 10	90 ± 10	32 ± 2	107 ± 8	14.3 ± 0.7	65 ± 5	8.3 ± 0.2	37 ± 4	29 ± 2
24 Jun	90 ± 20	350 ± 30	25 ± 3	80 ± 4	13 ± 2	19 ± 1	< 14	< 1.6	< 6.8	< 1.2	5.3 ± 0.4	13 ± 2
28 Jun	277 ± 5	750 ± 30	66 ± 4	227 ± 9	40 ± 10	30 ± 10	43 ± 4	8 ± 3	30 ± 5	4 ± 2	16 ± 3	26 ± 3
02 Jul	840 ± 50	1890 ± 80	200 ± 10	750 ± 20	132 ± 6	44 ± 3	120 ± 10	21 ± 1	104 ± 9	14 ± 1	45 ± 5	40 ± 2
06 Jul	1600 ± 100	3300 ± 100	374 ± 9	1470 ± 50	280 ± 10	66 ± 4	303 ± 5	41 ± 1	187 ± 7	30 ± 2	89 ± 3	74 ± 7
10 Jul	2330 ± 80	5100 ± 100	606 ± 9	2240 ± 70	490 ± 30	108 ± 5	460 ± 40	80 ± 30	340 ± 20	56 ± 6	153 ± 9	120 ± 10
14 Jul	850 ± 40	11200 ± 100	207 ± 8	730 ± 20	140 ± 10	45 ± 2	159 ± 8	20.2 ± 0.8	90 ± 6	12.6 ± 0.8	43 ± 3	50 ± 2
18 Jul	1410 ± 40	2830 ± 90	316 ± 9	1140 ± 40	230 ± 10	60 ± 3	200 ± 20	32 ± 2	150 ± 10	24.7 ± 0.6	79 ± 3	80 ± 7
22 Jul	2300 ± 200	9000 ± 1000	300 ± 20	1240 ± 50	301 ± 6	74 ± 7	270 ± 20	30 ± 10	180 ± 20	25 ± 4	76 ± 7	63 ± 5
26 Jul	2100 ± 100	4870 ± 70	520 ± 20	2020 ± 40	380 ± 10	114 ± 8	400 ± 20	61 ± 8	249 ± 6	40 ± 2	112 ± 5	90 ± 10
30 Jul	500 ± 20	9400 ± 200	121 ± 5	460 ± 20	81 ± 2	27 ± 1	102 ± 7	15 ± 2	68 ± 3	12 ± 2	31 ± 1	37 ± 4
03 Aug	170 ± 20	460 ± 30	36 ± 3	147 ± 7	26 ± 4	22 ± 2	28 ± 2	4.9 ± 0.2	22 ± 6	1.6 ± 0.2	10 ± 1	25 ± 3
07 Aug	480 ± 30	1230 ± 40	110 ± 10	410 ± 20	80 ± 8	37 ± 2	93 ± 7	13.0 ± 0.5	63 ± 6	10.5 ± 0.9	35 ± 3	36 ± 2
11 Aug	170 ± 20	16300 ± 400	43 ± 1	143 ± 9	28 ± 5	19 ± 1	58 ± 3	5.7 ± 0.8	19 ± 2	2.8 ± 0.5	9.9 ± 0.9	9.4 ± 0.9
15 Aug	390 ± 20	930 ± 40	86 ± 4	540 ± 20	77 ± 5	29 ± 2	79 ± 8	10.2 ± 0.6	647	8.01 ± 0.05	31 ± 2	25 ± 3
19 Aug	1030 ± 70	2300 ± 100	248 ± 6	920 ± 30	170 ± 10	48 ± 2	170 ± 20	25 ± 1	121 ± 8	16.5 ± 0.9	53 ± 7	51 ± 3
23 Aug	690 ± 30	1700 ± 60	169 ± 2	650 ± 20	130 ± 10	41 ± 2	117 ± 6	22 ± 1	117 ± 8	19 ± 2	56 ± 5	51 ± 4
27 Aug	390 ± 20	9700 ± 400	85 ± 1	370 ± 10	76 ± 4	26 ± 2	112 ± 5	10.7 ± 0.4	59 ± 5	7.4 ± 0.8	27 ± 3	25 ± 2
31 Aug	450 ± 30	1030 ± 30	109 ± 2	385 ± 9	70 ± 3	27 ± 1	62 ± 2	8.7 ± 0.6	51 ± 4	8 ± 1	24.5 ± 0.6	29 ± 2
04 Sep	240 ± 10	12700 ± 300	63 ± 4	206 ± 7	34 ± 2	24 ± 1	92 ± 7	8.7 ± 0.5	27 ± 3	3.9 ± 0.6	17 ± 2	31 ± 3

c)					20	13 CAM	PAIGN					
Start date	La (fg/m ³)	Ce (fg/m ³)	Pr (fg/m ³)	Nd (fg/m ³)	Sm (fg/m ³)	Eu (fg/m ³)	Gd (fg/m ³)	Tb (fg/m ³)	Dy (fg/m ³)	Ho (fg/m ³)	Er (fg/m ³)	Yb (fg/m ³)
01 May	1600 ± 100	1220 ± 30	450 ± 10	910 ± 30	311 ± 10	77 ± 3	310 ± 20	49 ± 3	250 ± 20	38.7 ± 0.7	110 ± 6	91 ± 6
05 May	2260 ± 30	2000 ± 100	640 ± 10	1320 ± 40	423 ± 30	115 ± 5	410 ± 20	74 ± 1	310 ± 10	61.4 ± 0.8	150 ± 10	118 ± 6
09 May	9100 ± 300	< 100	61 ± 2	< 32	54 ± 7	36 ± 3	67 ± 3	20 ± 2	41 ± 1	21.1 ± 0.9	25.1 ± 0.3	19 ± 3
13 May	1360 ± 20	1090 ± 60	410 ± 10	750 ± 10	256 ± 20	67 ± 5	257 ± 8	39 ± 1	180 ± 10	36 ± 1	85 ± 2	69.0 ± 0.5
17 May	1020 ± 40	2180 ± 70	270 ± 30	1100 ± 20	214 ± 10	49 ± 3	173 ± 6	30 ± 1	160 ± 10	29 ± 2	80 ± 5	65 ± 4
21 May	760 ± 10	500 ± 10	223 ± 7	340 ± 20	157 ± 3	49 ± 2	169 ± 7	24 ± 2	124 ± 7	24 ± 1	67 ± 3	44 ± 2
25 May	280 ± 10	500 ± 20	67 ± 2	290 ± 20	982 ± 40	20 ± 1	< 15	18 ± 2	36 ± 4	8.1 ± 0.7	17 ± 1	31 ± 6
29 May	490 ± 20	195 ± 4	143 ± 8	53 ± 2	89 ± 3	35 ± 1	102 ± 8	11.7 ± 0.8	68 ± 2	14.4 ± 0.6	36 ± 3	27.1 ± 0.6
02 Jun	740 ± 30	510 ± 10	194 ± 8	190 ± 8	138 ± 2	44 ± 2	141 ± 5	22 ± 2	125 ± 3	17 ± 1	54 ± 4	46 ± 2
06 Jun	3010 ± 70	6000 ± 300	730 ± 20	2590 ± 50	508 ± 10	113 ± 3	440 ± 10	76 ± 8	371 ± 8	72 ± 3	176 ± 9	170 ± 10
10 Jun	880 ± 40	730 ± 20	282 ± 5	460 ± 10	188 ± 3	45 ± 2	200 ± 10	26 ± 1	134.9 ± 0.6	21.6 ± 0.4	71 ± 8	54 ± 2
14 Jun	< 22	< 103	17.2 ± 0.7	< 33	14 ± 1	< 10	14 ± 2	< 1.5	20 ± 2	3.4 ± 0.2	16.1 ± 0.4	12 ± 2
18 Jun	< 21	< 99	< 7.6	< 32	< 8.6	< 10	< 12	< 1.4	< 5.9	< 1.0	< 3.0	< 3.5
22 Jun	300 ± 20	< 110	90 ± 2	< 35	57 ± 3	20.1 ± 0.7	64 ± 3	7.9 ± 0.3	49 ± 3	6.8 ± 0.6	27.0 ± 0.8	13.5 ± 0.6
26 Jun	820 ± 70	620 ± 20	266 ± 5	430 ± 20	166 ± 9	57 ± 5	187 ± 5	29 ± 3	145 ± 4	26.5 ± 0.1	72 ± 3	54.7 ± 0.9
30 Jun	1700 ± 200	889 ± 2	350 ± 10	470 ± 20	257 ± 20	153 ± 7	246 ± 9	120 ± 4	217 ± 8	117 ± 5	143 ± 5	150 ± 5
04 Jul	2100 ± 100	1390 ± 20	900 ± 30	1040 ± 20	695 ± 20	630 ± 40	632 ± 9	510 ± 10	620 ± 30	457 ± 8	520 ± 40	460 ± 10
08 Jul	300 ± 30	549 ± 6	70 ± 4	330 ± 20	52 ± 3	18 ± 1	< 13	4.0 ± 0.2	39 ± 1	5.8 ± 0.3	18.8 ± 0.7	9.4 ± 0.4
12 Jul	17300 ± 400	19700 ± 500	3900 ± 100	14600 ± 400	2651 ± 40	960 ± 40	2690 ± 50	595 ± 4	1790 ± 20	480 ± 30	975 ± 4	810 ± 30
16 Jul	9400 ± 500	12100 ± 700	2600 ± 100	9210 400	1734 ± 20	780 ± 20	1820 ± 40	306 ± 3	1190 ± 30	244 ± 8	690 ± 30	580 ± 9
20 Jul	76.8 ± 0.5	< 100	49.5 ± 0.4	< 33	29.2 ± 0.4	< 10	52 ± 3	< 1.5	15.2 ± 0.9	2.5 ± 0.2	10.7 ± 0.8	12.6 ± 0.7
24 Jul	490 ± 20	110 ± 8	121 ± 6	< 34	84 ± 2	25 ± 2	98 ± 3	10.7 ± 0.8	57 ± 5	11.8 ± 0.9	29.8 ± 0.3	15.1 ± 0.8
28 Jul	3300 ± 100	3100 ± 100	360 ± 20	1150 ± 60	241 ± 8	82 ± 6	178 ± 9	61 ± 3	167 ± 9	65 ± 3	110 ± 10	79 ± 4
05 Aug	176 ± 7	< 110	90 ± 8	< 37	64 ± 4	33 ± 1	70 ± 4	9.2 ± 0.8	68 ± 3	14 ± 1	41 ± 1	28.2 ± 0.8
09 Aug	28500 ± 400	53000 ± 2000	118 ± 8	282 ± 7	56 ± 8	43 ± 3	250 ± 30	18.2 ± 0.3	29.0 ± 0.8	14 ± 5	12 ± 1	9 ± 2
13 Aug	1690 ± 60	3480 ± 70	409 ± 8	1530 ± 60	287 ± 10	55 ± 1	200 ± 10	33 ± 2	180 ± 10	29 ± 2	90 ± 6	68 ± 4
17 Aug	300 ± 20	< 100	82 ± 3	< 33	68 ± 3	14.7 ± 0.2	86 ± 9	13.18 ± 0.07	56 ± 8	10.3 ± 0.3	26 ± 2	10.1 ± 0.7
21 Aug	< 20	< 97	< 7.5	< 31	9.4 ± 0.3	< 10	< 12	< 1.4	9.8 ± 0.6	< 1.0	< 3.0	< 3.4
25 Aug	25 ± 1	< 100	28 ± 2	< 32	15 ± 1	< 10	22 ± 1	< 1.4	11.1 ± 0.1	3.1 ± 0.2	7.5 ± 0.8	15 ± 2
29 Aug	< 23	< 110	< 9	< 36	< 10	< 11	< 14	< 1.6	6.1 ± 0.3	< 1.1	< 3.4	< 3.9
02 Sep	400 ± 20	350 ± 10	163 ± 6	166 ± 8	165 ± 2	57.6 ± 0.9	145 ± 6	30 ± 2	160 ± 20	30 ± 2	86 ± 4	73 ± 7
06 Sep	3150 ± 20	6670 ± 80	760 ± 30	2870 ± 60	556 ± 30	125 ± 4	490 ± 20	82.8 ± 0.6	450 ± 20	80 ± 2	250 ± 4	203 ± 8
10 Sep	118 ± 4	< 130	112 ± 2	< 42	72 ± 5	19 ± 1	79 ± 5	6.8 ± 0.2	51 ± 3	12.0 ± 0.8	26 ± 3	22 ± 2

Values below the LOD were obtained by using the real solution volume and air volume of each sample.

Supplementary Table 3. Summary of multiple pairwise comparisons obtained with Kruskal-Wallis and Conover-Iman tests: a) comparison of sampling campaigns; b) comparison of sampling seasons. Letter B identifies campaigns (or seasons) presenting mean of ranks significantly higher than the ones obtained for campaigns (or seasons) identified by letter A; the same concept applies to letter C with respect to letter B.

Group

А

А

В

А

В

Elements

As, Ba, Ce, Co, Dy, Er, Eu, Fe, Gd, Ho, K, La, Mg, Mo, Na, Nd, Ni, Pr,

Sm, Tb, Ti, Yb, Zn

Al, Ca, Cd, Cu, Mn, Pb

V

a)	Campaign	Group	Elements	b) Season
	2010, 2011, 2012, 2013	А	Al, Ba, Ca, Co, Ho*, K, Mg*, Mn, Na, Nd, Ni, Pb, V, Zn	Summer, spring
	2010	А	Cu	Summer
	2011, 2012, 2013	В		Spring
	2011	А	Ti	Spring
	2010, 2012, 2013	В		Summer
	2011	А	La	
	2013	AB		
	2012	В		
_	2010	С		
	2013	А	Fe	
	2011, 2012	AB		
	2010	В		
	2011, 2012	А	Мо	
_	2010, 2013	В		
	2011, 2013	А	Ce	
_	2010, 2012	В		
	2012, 2013	А	Cd	
	2011	AB		
	2010	В		
	2012, 2013	А	As	
	2010	AB		
	2011	В		
	2011, 2012, 2013	А	Dy, Er, Eu, Gd, Pr, Sm, Tb, Yb	
	2010	В		
	2010	D		

* Mg and Ho were not analysed in samples collected during 2010 campaign.



Supplementary Figure 1. Principal Component Analysis (PC3 vs. PC4) for 2012 samples: a) score plot; b) loading plot.