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Phosphorus Speciation in Long-Term Drained and Rewetted Peatlands of Northern Germany

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Abstract: Previous studies, conducted at the inception of rewetting degraded peatlands, reported that rewetting increased phosphorus (P) mobilization but long-term effects of rewetting on the soil P status are unknown. The objectives of this study were to (i) characterize P in the surface and subsurface horizons of long-term drained and rewetted percolation mires, forest, and coastal peatlands and (ii) examine the influence of drainage and rewetting on P speciation and distributions using wet-chemical and advanced spectroscopic analyses. The total P was significantly ($p < 0.05$) different at the surface horizons. The total concentration of P ranged from 1022 to 2320 mg kg⁻¹ in the surface horizons and decreased by a factor of two to five to the deepest horizons. Results of the chemical, solution ³¹P nuclear magnetic resonance (NMR), and P K-edge X-ray absorption near-edge structure (XANES) indicated that the major proportions of total P were organic P (Po). In the same peatland types, the relative proportions of Po and stable P fractions were lower in the drained than in the rewetted peatland. The results indicate that long-term rewetting not only locks P in organic matter but also transforms labile P to stable P fractions at the surface horizons of the different peatland types.

Keywords: Hedley sequential fractionation; inorganic phosphorus; organic phosphorus; P K XANES; phosphorus fractions; solution ³¹P NMR

1. Introduction

Peatlands can be a sink, source, and transformer of P, other essential plant nutrients, and pollutants that originate from agricultural soils, industrial, and urban waste [1]. Drainage is one of the principal management practices that enabled an agricultural use of peatlands. However, drainage has contributed to peatland degradation in Europe and worldwide [2,3]. Such degraded peatlands can be a source of P that contributes to freshwater eutrophication by enhancing the growth of algae and other aquatic plants [4].

Returning a peatland into carbon (C) accumulating state requires rewetting of degraded peatlands by raising the water table that facilitates the recolonization of peat-forming plant communities [5]. On the other hand, rewetting drained peatlands was reported to mobilize P as compared to pristine peatlands in incubation and leaching experiments [6,7]. Similarly, rewetting agriculturally influenced peatlands increased the concentrations of soluble P as shown in a study across Europe [4]. However, high P mobilization at the inception of rewetting agriculturally influenced peatland perhaps can be transient since soluble P is

dynamic and may be taken up by microbes and plants, lost by leaching and drainage, and form secondary P minerals [8,9]. Thus, a complete account of P species is required to predict the P mobilization or stabilization potential in long-term drained and rewetted peatland types [10,11].

Phosphorus transformations can vary according to management practices, peatland types, topographic positions, historic, and current land use types [12]. For instance, when a peat soil predominated by redox-sensitive elements is rewetted, P mobilization and release of soluble P species can occur by the reductive dissolution of binding partners of phosphate, especially pedogenic Fe-(oxy)hydroxides [6,13,14]. However, there is no information on the influence of long-term drainage and rewetting of percolation mires, forest, and coastal peatlands.

Traditional wet chemical and advanced spectroscopic analytical techniques such as sequential P fractionation, solution ^{31}P nuclear magnetic resonance (NMR), and X-ray absorption near edge structure (XANES) spectroscopy have been used successfully to investigate P species in mineral soils and other environmental samples [15]. However, none of these analytical techniques can decipher P species independently because of the complex chemical and biochemical processes of P such as dissolution-precipitation, sorption-desorption, mineralization-immobilization, and oxidation-reduction [16]. Thus, the combined use of wet chemical and spectroscopic analytical techniques could provide detailed information on the impacts of management practices on various P species distributions and transformations.

About 30,000 ha of degraded peatlands have been rewetted in Mecklenburg-West Pomerania, northern German since the mid-1990s [17,18]. Most of the rewetted peatlands were put under nature conservation, while the usual drainage has been continued to enable grazing, haymaking, and afforestation. Only a few studies investigated different P fractions in recently rewetted percolation mires [13,19,20]. Understanding the influence of such long-term drainage and rewetting of different peatland types on P biogeochemistry can provide an opportunity to improve future practice for sustainable use of a peatland ecosystem. Thus, the objectives of this study were to (i) speciate P in surface and subsurface horizons of long-term drained and rewetted percolation mires, forest, and coastal peatlands with the combined use of wet chemical and advanced spectroscopic analyses (ii) examine the influence of drainage and rewetting on P-distributions and transformations.

2. Materials and Methods

2.1. Site Description and Soil Sampling

Soil samples were collected from pairs of three different peatland types including the drained and rewetted sites of forest, coastal, and percolation mires in Mecklenburg-West Pomerania, northern Germany. Although there was no historical record of the initial drainage date, most intensive drainages occurred in 1960 to 70ies. The present study sites have rewetted since the mid-1990s. The latitude and longitude of the sampling points, depth of water table, peat accumulation depths, altitudes, and dominant plant species, mean annual temperature and rainfall of the study sites have already published [21]. The soil samples were collected from 0 to 80 cm soil depth and the sampling depths depending on the thickness of the accumulated peats, which ranged from 0.3 m to >1.0 m. We collected 72 soil samples from six sites; three drained and three rewetted peatlands at the beginning of June 2017. We transported the collected soil samples to the laboratory in a cooling box and then placed in a freezer at $-20\text{ }^{\circ}\text{C}$ immediately. The types of horizons, sampling depths, degree of humification, water content at sampling, and soil pH have also published elsewhere [21]. The soils of the study sites were characterized following the procedures of FAO [22]. The major soil units were Folic Histosol for the drained forest peatland, drained percolation mire, drained and rewetted coastal peatlands and Sapric Histosol for the rewetted forest peatland and percolation mire according to the International Union of Soil Science Working Group World Reference for Soil Resources [23]. The subsample of the frozen soil samples was freeze-dried and used for wet-chemical, ^{31}P NMR,

and XANES analyses. The concentration of all elements and P fractions were reported per oven-dried weight basis of the soil samples.

2.2. Total Elements and Plant Available P Analyses

The freeze-dried soil samples were ground by a ball mill to <0.5 mm sieve size and used for the determination of total elements. The concentrations of total P, K, Ca, Mg, Al, Fe, Mn, and Na were extracted by microwave-assisted aqua regia digestion of the soil samples and their concentrations were determined by inductively coupled plasma-optical emission spectroscopy (ICP-OES) (PerkinElmer Optima 8300, Waltham, MA, USA). In addition to microwave-assisted aqua regia digestion, we used the ignition method for the determination of total P_o [24]. Briefly, 1.00 g of soil samples (<2 mm) of the surface horizon and deepest subsurface horizons were weighed into the porcelain crucibles, and the crucibles were placed in the cool muffle furnace which was heated gradually to 550 °C and maintained for three hours at this temperature. The ignited samples were transferred to 50 mL centrifuge tubes. Similarly, we weighed 1.00 g of unignited soil samples in a 50 mL centrifuge tube. We added 25 mL of 0.5 M H_2SO_4 to both ignited and unignited samples, and the samples were shaken for 16 h, centrifuged for 30 min at 4500 g and filtered. The P concentration in the aliquot was determined by ICP-OES. The P concentration in the unignited samples was assumed to be P_i and the difference between the total P concentrations of the ignited and unignited peat soil samples were reported as total P_o .

The plant available P was extracted by double lactate (DL) following the procedures outlined in the LUFA Method Book [25]. The field moist soil samples, equivalent to 0.60 g oven dried sample was weighed into 50 mL centrifuge tube and 30 mL DL working solution was added. The samples were shaken for 30 min using the overhead shaker, centrifuged for 15 min and the filtered solution was used for determination of the plant available P concentration by ICP-OES.

2.3. Sequential P Fractionation

About 0.50 g of peat soil sample was weighed into the 50 mL centrifuge tube. The anion exchange membrane (AEM), 0.5 M $NaHCO_3$ at pH 8.5, 0.1 M NaOH, and 1.0 M H_2SO_4 were used to sequentially extract different P fractions [15,26]. Each fractionation step was performed in a 1:60 weight to volume ratio in triplicate. The samples were shaken at 60 rpm using a reciprocal shaker for 16 h and then centrifuged at 4500 g for 30 min. The P_i concentration in all extracts was determined by the molybdenum blue method [27] using a SPECORD 50 PLUS ultraviolet/visible (UV/Vis) spectrophotometer (Analytik Jena AG, Jena, Germany), whereas the concentration of total P in each P fraction determined by ICP-OES. The concentration of P_o was obtained by the difference of the concentration of total P and P_i of each fraction, whereas the residual P was the difference of total P and the sum of all P_i and P_o fractions. The AEM-P, $NaHCO_3$ -P are assumed labile P, NaOH-P is moderately labile, and H_2SO_4 -P and Residual-P are stable P pools of P_i and P_o [28].

2.4. Solution ^{31}P Nuclear Magnetic Resonance (NMR) and P K X-ray Absorption Near Edge Structure (XANES) Analyses

For ^{31}P NMR analysis, the P was extracted by a single step extraction with 0.25 M NaOH –0.05 M Ethylenediaminetetraacetic acid disodium salt dehydrate (Na_2EDTA) ($NaOH$ -EDTA) from the soil samples of the surface and deepest horizons following the procedures outlined by Cade-Menun [29]. The extraction was done at 1:20 soil to solution ratio, by shaking for 16 hr, followed by centrifugation at 4500 g for 30 min, and filtration. The filtered samples were frozen immediately at –20 °C, and then freeze-dried. Before ^{31}P analysis, we dissolved 100 to 200 mg freeze-dried samples in 600 to 1000 μL of 30% NaOD and D_2O mixture. We added 200 to 500 mg of $Na_2S \cdot 9H_2O$ depending on the Fe content of the samples as recommended by Vestergren et al. [30]. The samples were vortexed for 2 min for the complete dissolution of $Na_2S \cdot 9H_2O$ and centrifuged at 45,000 g for 15 min, and the clear solution was used for the ^{31}P NMR analysis.

The solution ^{31}P NMR spectra were recorded using an inverse gated decoupled (IG) pulse sequence on a Bruker Avance 500 spectrometer with a pulse width of 30° ($3.3 \mu\text{s}$), 1.64 s acquisition time, 5.0 s pulse delay, 10,000 scans and a temperature of 298 K. The chemical shifts of signals were determined in parts per million (ppm) relative to the reference P compounds used along with the samples. In this study, we used inositol hexakisphosphate (IHP), glucose 6-P, glucose-1-P, adenosine 5' monophosphate (AMP), deoxyribonucleic acid (DNA), and α - and β -glycerophosphates for the reference P compounds for peak identification as recommended by Cade-Menun [29]. The NMR spectra were processed by Topspin 3.6.1 version. Here we reported the NMR data of surface horizons because we did not get reliable NMR data from the subsurface horizons because of high viscosity due to high organic matter concentrations along with low P and high Fe contents of the samples.

For P K-edge XANES analysis, we followed the detailed procedures described by Prietzel et al. [31]. The spectra were acquired at beamline 8 of the electron storage ring (1.2 GeV; bending magnet; beam current 69–149 mA; 1.1 to 1.7 10^{11} photons s^{-1}) at the Synchrotron Light Research Institute (SLRI) in Nakhon Ratchasima, Thailand. Finely ground homogenized powder of each sample was spread as a thin film ($2 \text{ cm} \times 0.5 \text{ cm}$) and carefully distributed with a spatula to yield a homogeneous particle distribution on Kapton tape (Lanmar Inc., Northbrook, IL, USA) and then mounted on a sample holder. After energy calibration using elemental P powder ($E = 2145.5 \text{ eV}$; precision 0.11 eV), spectra were acquired in the energy range of 2110–2350 eV in 5.0 eV increment (2105 to 2135 eV; 2205 to 2350 eV) and 0.25 eV increment (2135 to 2205 eV). Spectrum acquisition time was 3 sec, and incident X-ray intensity was measured using a N_2 gas-filled ion chamber while the fluorescence emitted from the sample was recorded with a 13-channel Ge detector. For each sample, three spectra were acquired and merged, and the spectra were baseline corrected; pre-edge (energy: 2115–2145 eV) and post edge (energy region: 2190–2215 eV) using the software package ATHENA [32]. We also used the ATHENA software package for the linear combination fitting (LCF) to relatively quantify P species. We used spectra of the following reference compounds for LCF: amorphous AlPO_4 , Ca-phytate, crystalline AlPO_4 , FePO_4 , Ca-hydroxyapatite, IHP, ferrihydrite-IHP, montmorillonite-Al-IHP, ferrihydrite- PO_4 , soil organic matter (SOM)-Al-IHP, montmorillonite-Al- PO_4 , SOM-Al- PO_4 , boehmite- PO_4 , CAHPO_4 , CaH_2PO_4 , crystalline- FePO_4 , and MgHPO_4 [31].

2.5. Statistical Analysis

The analysis of variance of the total concentration of the selected elements and different P fractions was conducted by the SAS software [33] using PROC GLIMMIX. The different soil depths in different peatland types were treated as the fixed effect, whereas the replication and replication \times soil depth interaction were treated as the random effect. The normality was tested before running any statistical analysis, and all the selected total elements and P fractions were normally distributed, and thus we ran the statistical analysis using the original data without data transformation.

3. Results

3.1. Total Elements and Plant-Available P

The total concentrations of elements that potentially form bindings with P compounds were significantly ($p < 0.05$) different among the soil horizons of the studied peatlands (Table S1). Total Ca was the most abundant among the elements in the rewetted forest peatland, drained and rewetted percolation mires but it was significantly lower at the surface horizons of the drained forest peatlands, drained, and rewetted coastal peatlands. However, the concentrations of total P and K were significantly higher in the surface than the subsurface horizons and decreased with depths (Table S1). The concentrations of total Mg in the coastal peatlands were significantly ($p < 0.05$) higher than in the other peatlands. The concentration of total Fe was significantly higher in the rewetted forest peatland than in the other studied peatlands. The concentration of Al was significantly higher in the coastal peatlands and drained

forest peatlands than in the other peatlands. As compared to the other elements, the concentrations of K, Na, and Mn were lower in all horizons and different peatland types.

The concentration of plant available P differed significantly between the surface and subsurface horizons except for the rewetted percolation mire (Figure 1). Similar to the total P (Table S1), the plant available P significantly decreased from the surface to the subsurface horizons. The concentration of plant available P was also significantly higher at the surface than at the subsurface horizons while there were no significant differences ($p > 0.05$) among the subsurface horizons except at the rewetted percolation mire.

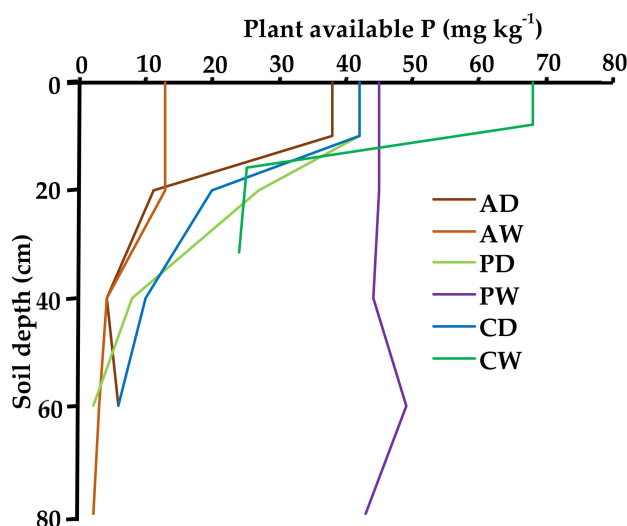


Figure 1. Content of plant available P in the different management and peatland types extracted by double lactate. Standard error of the mean: 12 mg P kg⁻¹. AD: drained forest peatland, AW: rewetted forest peatland, PD: drained percolation mire, PW: rewetted percolation mire, CD: drained coastal influenced peatland, CW: rewetted coastal influenced peatland.

3.2. Phosphorus Fractions Recovered by Different Methods

The concentrations of P_i fractions recovered by sequential P fractionation were lower than the concentrations of their respective P_o and residual-P (Table 1). The concentrations of soluble P_i extracted by anion exchange membrane (AEM-P_i), labile P_i (NaHCO₃-P_i), and moderately labile P_i (NaOH-P_i) were mostly significantly ($p < 0.05$) different among the surface and subsurface horizons. These three P_i fractions also decreased from the surface to the subsurface horizons. In contrast to the AEM-P_i, NaHCO₃-P_i, and NaOH-P_i fractions, the distribution of stable P_i (H₂SO₄-P_i) fraction from the surface to subsurface horizons was different for the different peatland types. The H₂SO₄-P_i fraction increased from the surface to subsurface horizons in the drained forest and coastal peatlands, whereas it decreased from the surface to subsurface horizons in the rewetted forest, rewetted coastal peatland, and percolation mires.

Similar to NaHCO₃-P_i and NaOH-P_i fractions, their respective labile P_o (NaHCO₃-P_o) and moderately labile P_o (NaOH-P_o) fractions decreased from the surface to the subsurface horizons (Table 1). However, the NaHCO₃-P_o was less than the NaOH-P_o at both the surface and subsurface horizons. The concentration of the residual-P, the second to the NaOH-P_o fraction, significantly decreased from the surface to the subsurface horizons. Furthermore, the residual-P was lower in the drained than in its respective rewetted peatland except for the percolation mires (Table 1). The proportion of H₂SO₄-P and residual-P was 57% in the surface horizons of the rewetted forest peatland, whereas these fractions accounted for only 32% in the drained forest peatland (Figure 2a). Similarly, the relative contribution of the H₂SO₄-P and residual-P to the total P of the rewetted and drained coastal peatlands were 48% and 40% (Figure 2b), and that of the rewetted and drained percolation mires were 67% and 59%, respectively (Figure 2c).

Table 1. Contents of phosphorus fractions extracted by sequential fractionation method.

(cm)		(mg P kg ⁻¹ Oven Dried Sample)							
Site	Depth	AEM-P _i	NaHCO ₃ -P _i	NaOH-P _i	H ₂ SO ₄ -P _i	NaHCO ₃ -P _o	NaOH-P _o	H ₂ SO ₄ -P _o	Residual-P
AD	0–10	37 ^{cd}	14 ^{ef}	14 ^{cdefh}	15 ^j	213 ^b	399 ^c	52 ^{efgh}	278 ^{fgh}
	10–20	21 ^{hijk}	11 ^{ef}	16 ^{cdef}	13 ^j	137 ^{ef}	305 ^{de}	57 ^{defg}	114 ^{ijk}
	20–40	24 ^{fghijk}	12 ^{ef}	12 ^{efhi}	19 ^{ij}	143 ^{ed}	307 ^{de}	71 ^d	172 ^{hij}
	40–60	22 ^{ghijk}	7 ^f	8 ^{jk}	56 ^{fg}	47 ^{ij}	95 ^{lkm}	47 ^{fghi}	132 ^{ijk}
AW	0–20	38 ^{bc}	39 ^b	76 ^a	319 ^a	200 ^{bc}	530 ^b	121 ^b	749 ^b
	20–40	28 ^{defgh}	10 ^{ef}	11 ^{fhi}	76 ^e	111 ^{fg}	295 ^{def}	74 ^{cd}	347 ^{ef}
	40–60	35 ^{cde}	8 ^f	10 ^{hij}	51 ^{gh}	84 ^{gh}	193 ^{hij}	62 ^{def}	270 ^{fgh}
	60–80	23 ^{ghijk}	7 ^f	10 ^{hij}	24 ^{ij}	56 ^{hi}	126 ^{jkl}	37 ^{ijk}	169 ^{hij}
CD	0–10	33 ^{cdef}	34 ^{bc}	11 ^{hij}	46 ^{gh}	289 ^a	326 ^d	43 ^{hij}	383 ^{def}
	10–20	27 ^{efghi}	26 ^{cd}	11 ^{hij}	51 ^{gh}	213 ^b	248 ^{efgh}	36 ^k	64 ^{jk}
	20–30	19 ^{hijk}	13 ^{ef}	12 ^{cdefh}	74 ^{ef}	132 ^{ef}	225 ^{ghi}	48 ^{fghi}	106 ^{ijk}
	30–40	20 ^{hijk}	21 ^{ed}	19 ^c	77 ^e	150 ^{ed}	237 ^{fgh}	44 ^{ghij}	36 ^k
	40–60	18 ^{ijk}	14 ^{ef}	7 ^k	9 ^j	95 ^g	159 ^{ijk}	17 ^m	275 ^{fgh}
CW	0–10	47 ^{ab}	50 ^a	16 ^{cdef}	75 ^e	212 ^b	403 ^c	64 ^{de}	581 ^c
	10–20	27 ^{efghi}	28 ^{bcd}	10 ^{hij}	56 ^{fg}	171 ^{cd}	210 ^{ghi}	32 ^{kl}	224 ^{ghi}
	20–30	26 ^{fghij}	28 ^{bcd}	12 ^{efhi}	35 ^{hi}	155 ^{ed}	182 ^{hij}	26 ^l	146 ^{ijk}
PD	0–10	55 ^a	35 ^{bc}	56 ^b	328 ^a	280 ^a	617 ^a	235 ^a	937 ^a
	10–20	31 ^{cdefg}	26 ^{cd}	54 ^b	146 ^b	109 ^{fg}	268 ^{defg}	98 ^{bc}	168 ^{hij}
	20–40	20 ^{hijk}	11 ^{ef}	18 ^{dc}	14 ^j	31 ^{ijk}	82 ^{lm}	42 ^{hijk}	277 ^{fgh}
	40–60	16 ^k	7 ^f	15 ^{cdefc}	14 ^j	13 ^{jk}	43 ^m	39 ^{ijk}	915 ^a
PW	0–20	27 ^{efghi}	14 ^{ef}	13 ^{defh}	114 ^c	59 ^{hi}	234 ^{fgh}	98 ^{bc}	498 ^{cd}
	20–40	21 ^{hijk}	7 ^f	10 ^{hij}	102 ^{cd}	18 ^{jk}	57 ^m	43 ^{hijk}	348 ^{ef}
	40–60	22 ^{ghijk}	8 ^f	10 ^{hij}	100 ^{cd}	40 ^{ij}	661 ^m	59 ^{def}	414 ^{ed}
	60–80	18 ^{jk}	6 ^f	9 ^{hijk}	83 ^{de}	4 ^k	42 ^m	51 ^{efgh}	337 ^{efg}
SE		3	4	3	6.8	11	24	9	42

AD: drained forest peatland, AW: rewetted forest peatland, PD: drained percolation mire, PW: rewetted percolation mire, CD: drained coastal influenced peatland, CW: rewetted coastal influenced peatland. Different letters within the column indicate significant difference at ($p < 0.05$).

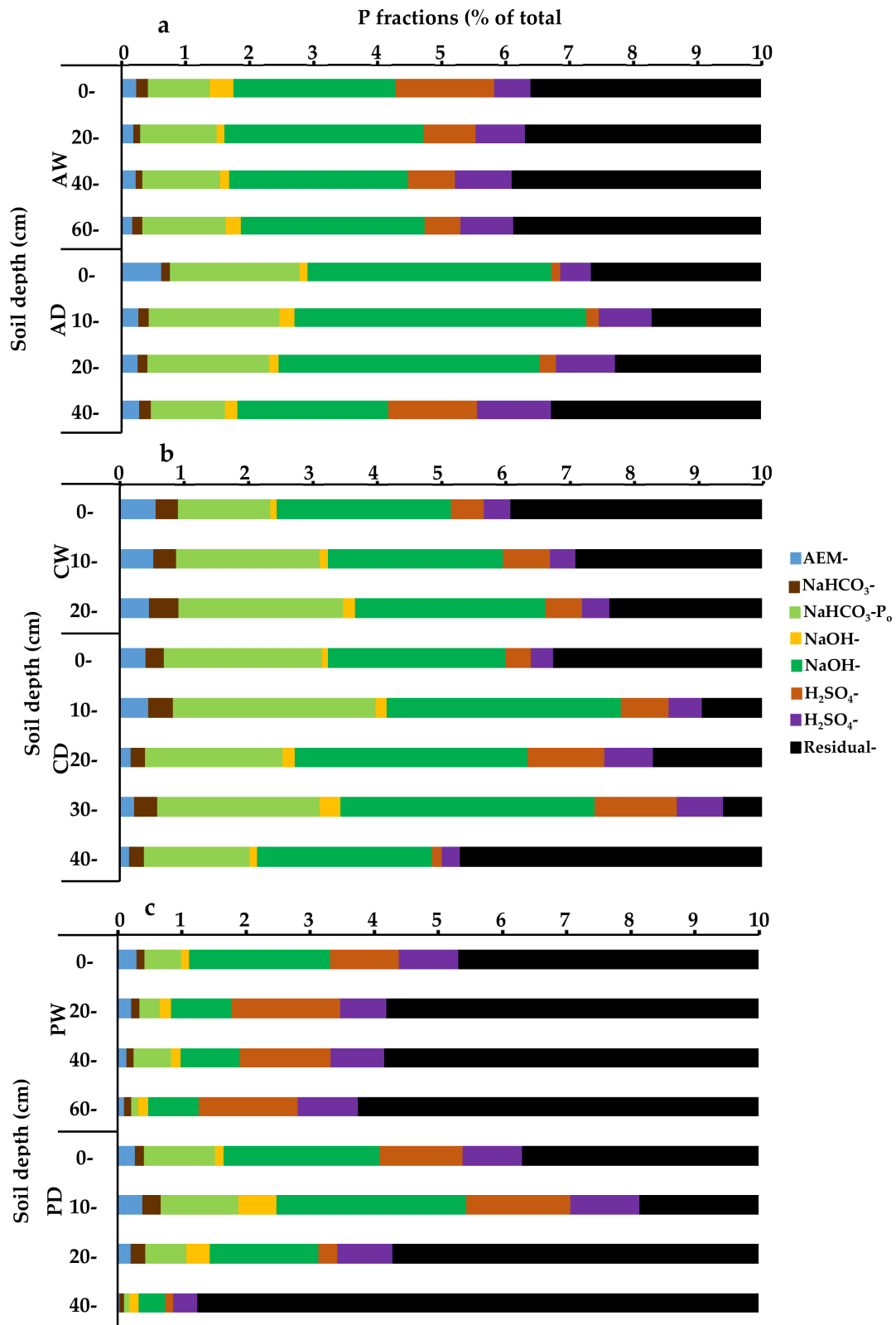


Figure 2. Relative proportions of different P fractions in the drained and rewetted peatlands. (a): drained and rewetted forest peatlands, (b): drained and rewetted coastal peatlands, (c): drained and rewetted peatlands. AEM-P: extracted by anion exchangeable membrane, NaHCO₃-P_i and NaHCO₃-P_o: labile inorganic and organic P, NaOH-P_i and NaOH-P_o: moderately labile organic and inorganic p, H₂SO₄-P and Residual-P: Stable inorganic and organic P fractions.

The sequential P fractionation recovered higher proportions of total P from the drained than the rewetted surface horizons of the same peatland types. Accordingly, the method recovered 73% and 64% from the drained and rewetted forest peatlands, 67 and 60% from the drained and rewetted coastal peatlands, and 60% and 53% of total P from the drained and rewetted percolation mires of the surface horizons, respectively (Table 1 and Figure 2).

The concentrations of total P, P_i , and P_o recovered by the ignition method were significantly ($p < 0.05$) different at the surface horizons, whereas they were not significantly different ($p > 0.05$) at the subsurface horizons (Table 2). Furthermore, the total P determined by the ignition method was also similar to the total P determined by microwave-assisted aqua regia digestion (Table 2 and Table S1). The relative proportion of P_i and P_o determined by the ignition method ranged from 20% to 47% and from 53% to 80% of the total P, respectively. The ignition method recovered the highest proportion of P_i from the surface horizon and the highest P_o from the subsurface horizons of the drained percolation mire. However, the lowest P_i recovered from the surface horizon of the drained forest peatland and subsurface horizons of the rewetted forest peatlands, whereas the lowest P_o was recovered from the surface horizon of drained percolation mire followed by surface horizons of drained coastal peatland. Similar to the sequential P fractionation method (Table 1), the ignition method also recovered higher concentrations of P_o than P_i , although the absolute concentration and relative proportion of P_i in the ignition method were higher than that was recovered by the sequential P fractionation.

Table 2. Phosphorus fractions determined by the ignition method.

Site	(cm)	(mg P kg ⁻¹ Oven Dried Sample)			%	
	Depth	TP	TP _i	TP _o	TP _i	TP _o
AD	0–10	1069 ^{cd}	216 ^{de}	852 ^{bc}	21 ^{fg}	79 ^{ab}
	40–60	471 ^e	128 ^{fg}	343 ^e	27 ^{ef}	73 ^{bc}
AW	0–20	1823 ^b	795 ^b	1027 ^b	43 ^{abc}	57 ^{efg}
	60–80	385 ^e	84 ^g	301 ^e	21 ^{fg}	79 ^{ab}
CD	0–10	1004 ^d	444 ^c	560 ^{de}	44 ^{ab}	56 ^{fg}
	40–60	587 ^e	229 ^{de}	358 ^e	39 ^{bcd}	61 ^{def}
CW	0–10	1291 ^c	480 ^c	811 ^{bcd}	37 ^{dc}	63 ^{de}
	20–30	638 ^e	268 ^d	370 ^e	42 ^{abc}	58 ^{efg}
PD	0–10	2618 ^a	1241 ^a	1377 ^a	47 ^a	53 ^g
	40–60	507 ^e	101 ^g	406 ^e	20 ^g	80 ^a
PW	0–20	966 ^d	273 ^d	692 ^{cd}	28 ^{ef}	72 ^{bc}
	60–80	596 ^e	189 ^{ef}	408 ^e	32 ^{de}	68 ^{cd}
	SE	88	24	88	3	3

AD: drained forest peatland, AW: rewetted forest peatland, PD: drained percolation mire, PW: rewetted percolation mire, CD: drained coastal influenced peatland, CW: rewetted coastal influenced peatland. TP: total P extracted by ignition method, TP_i and TP_o: total inorganic P and organic P fractions, respectively. Different letters within the column indicate significant difference at ($p < 0.05$)

The concentration of total P, P_i , and P_o extracted by NaOH-EDTA is presented in Table 3. The highest concentration of total P was extracted from the surface horizon of the drained percolation mire followed by rewetted forest peatland. There were significant differences in total P and P_i at the surface horizons except for the coastal peatlands. However, the concentration of the total P_o was not significantly (>0.05) different at most of the surface horizons. The relative proportion of P_i recovered by the NaOH-EDTA was in the range of 8% to 40% of the total P, whereas the relative proportion of P_o was in the range of 17% to 67% of the total P determined by microwave-assisted aqua regia digestion (Table 3 and Table S3). Overall, the proportions of the total P recovered by the NaOH-EDTA were in the range of 33% to 85%; the lowest proportion was recovered from the deepest horizon of drained

percolation mire, whereas the highest proportion of the total P was recovered from the surface horizon of the drained forest peatland (Table 3).

Table 3. Phosphorus fractions recovered by single step NaOH-Na₂EDTA (NaOH-EDTA) extraction.

Site	Depth (cm)	(mg P kg ⁻¹ Oven Dried Sample)			(%)		
		TP _f	TP _{fi}	TP _{fo}	TP _{fi}	TP _{fo}	TP _f
AD	0–10	881 ^c	186 ^{ed}	696 ^b	18 ^c	67 ^a	85 ^a
	40–60	263 ^{ef}	51 ^{fg}	213 ^c	12 ^c	51 ^{abc}	62 ^{abcd}
AW	0–20	1500 ^b	804 ^a	697 ^b	40 ^a	50 ^{abc}	74 ^{abc}
	60–80	296 ^e	77 ^{fg}	220 ^c	18 ^c	34 ^{cd}	68 ^{abcd}
CD	0–10	907 ^c	191 ^d	715 ^b	14 ^c	50 ^{abc}	63 ^{abcd}
	40–60	271 ^{ef}	53 ^{fg}	218 ^c	12 ^c	48 ^{abc}	60 ^{bcd}
CW	0–10	1002 ^c	324 ^c	679 ^b	34 ^{ab}	46 ^{bc}	68 ^{abcd}
	20–30	321 ^e	188 ^{de}	133 ^c	22 ^{bc}	17 ^d	51 ^{cde}
PD	0–10	1833 ^a	494 ^b	1340 ^a	21 ^{bc}	57 ^{ab}	79 ^{ab}
	40–60	143 ^f	41 ^g	103 ^c	10 ^c	24 ^d	33 ^e
PW	0–20	663 ^d	115 ^{ef}	581 ^b	21 ^{bc}	54 ^{abc}	62 ^{abcd}
	60–80	254 ^{ef}	82 ^{fg}	140 ^c	8 ^c	25 ^d	45 ^{de}
SE		46	24	47	5	7	8

Full names of the sites are given in Tables 1 and 2. TP_f: total P fractions extracted by NaOH-EDTA, TP_{fi} and TP_{fo}: total inorganic P and organic P fractions, respectively. Different letters within the column indicate significant difference at ($p < 0.05$).

The concentrations of the selected elements recovered by NaOH-EDTA were significantly different among the surface horizons (Table 4). As compared to the total element concentrations (Table S1), the NaOH-EDTA recovered less than 50% of the concentration of total elements except for Ca. Similarly, the method extracted the highest concentration of Fe from the rewetted forest peatland followed by the drained percolation mire.

Table 4. Selected elements recovered by a single step extraction of the NaOH-Na₂EDTA.

Site	Depth (cm)	(g kg ⁻¹ Oven Dried Sample)				
		Al	Fe	Ca	Mg	Mn
AD	0–10	2.72 ^a	2.87 ^d	6.56 ^c	0.25 ^e	0.09 ^c
AW	0–20	1.49 ^c	26.09 ^a	25.40 ^b	0.49 ^c	0.14 ^b
CD	0–10	2.55 ^a	7.42 ^c	5.50 ^c	0.78 ^b	0.04 ^d
CW	0–10	2.23 ^b	2.92 ^d	3.71 ^c	2.21 ^a	0.01 ^e
PD	0–10	0.83 ^d	11.34 ^b	25.25 ^b	0.41 ^d	0.20 ^a
PW	0–20	0.79 ^d	1.09 ^e	31.16 ^a	0.85 ^b	0.14 ^b
SE		0.06	0.01	1.10	0.02	0.01

AD: drained forest peatland, AW: rewetted forest peatland, PD: drained percolation mire, PW: rewetted percolation mire, CD: drained coastal influenced peatland, CW: rewetted coastal influenced peatland. Different letters within the column indicate significant differences at ($p < 0.05$).

3.3. Solution ³¹P NMR

The solution ³¹P NMR analysis revealed the presence of orthophosphate, IHP, α - and β -glycerol phosphates, DNA, AMP, and pyrophosphate (Figure 3 and Table 5). The α - and β -glycerol phosphates were the predominant P species followed by the IHP. Except for the rewetted percolation mire, the proportion of orthophosphate detected by the solution ³¹P NMR analysis was higher than the orthophosphate extracted by the sequential P fractionation (Table 1), ignition method (Table 2),

and NaOH-EDTA (Table 3). The presence of pyrophosphate in some samples slightly overestimated the absolute and relative proportion of P_o by the sequential P fractionation and NaOH-EDTA methods (Tables 1 and 5).

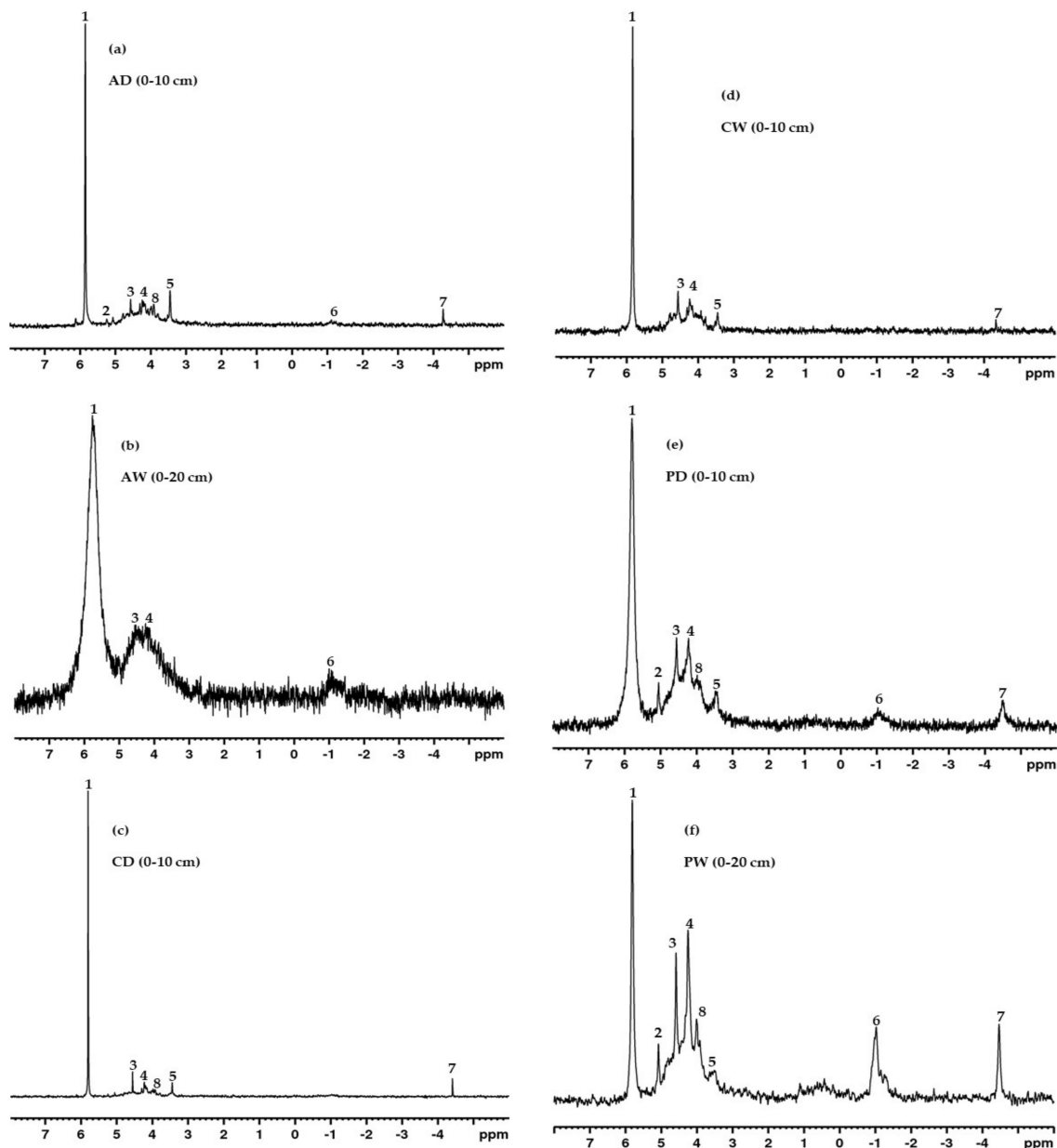


Figure 3. Solution ^{31}P nuclear magnetic resonance (NMR) spectra of the surface soil horizons of the drained and rewetted peatlands. AD: drained forest peatland, AW: rewetted forest peatland, CD: drained coastal peatland, CW: rewetted coastal peatland, PD: drained percolation mire, PW: rewetted percolation mire. 1: orthophosphate, 2: IHP, 3: α -glycerol phosphate, 4: β -glycerol phosphate, 5: IHP, 6: DNA, 7: pyrophosphate, 8: AMP. The line broadening of (b) was due to the highest concentration Fe in the extracted soil solution (Table 4). Numbers in the parenthesis next to the abbreviation of the site names indicate soil-sampling depth.

Table 5. Proportions of phosphorus species, P_i, and P_o determined by the solution ³¹P NMR analysis.

	(cm)	Chemical Shift (ppm)								Total P (%)	
		6.0–5.78	5.09–5.06	4.67–4.58	4.39–4.25	4.20–4.11	4.00–3.48	0.9–1.05	–4.45–(–4.47)		
Site	depth	Integrated area of the spectra (%).								P _i	P _o
		Ortho	IHP	α-glycerol	β-glycerol	AMP	IHP	DNA	Pyro.		
AD	0–10	41	6	10	15	9	10	3	6	47	53
AW	0–20	66			29			5		66	34
CD	0–10	52	2	13	14	10	6		3	55	45
CW	0–10	53	1	19	18		8		1	54	46
PD	0–10	55	4	12	16		5	4	4	59	41
PW	0–20	25	5	15	26		8	15	6	31	69

AD: drained forest peatland, AW: rewetted forest peatland, PD: drained percolation mire, PW: rewetted percolation mire, CD: drained coastal influenced peatland, CW: rewetted coastal influenced peatland. Ortho: orthophosphate, α-glycerol: α-glycerolphosphate, β-glycerol: β-glycerolphosphate, IHP: inositol hexkisphosphate, AMP: adenosine 5' monophosphate, DNA: deoxyribonucleic acid, pyro: pyrophosphate, P_i: ortho + pyro, P_o: IHP + α-glycerol + β-glycerol + AMP + DNA.

3.4. P K-Edge XANES

The P K-edge XANES analysis showed that the major proportion of P species were P_o in the form of IHP and phytate associated with Ca, Al, and boehmite, whereas the P_i associated with the $MgHPO_4$, crystalline $FePO_4$, and amorphous $AlPO_4$ were detected in surface and subsurface horizons of the coastal peatlands and rewetted forest peatlands. Although the relative proportion of IHP was the second abundant of P_o species with the NMR analysis, the XANES analysis accounted for all P_o species to IHP. Considering only the surface horizons, the proportions of P_o , estimated by the XANES analysis, accounted for 86% and 81% in the drained and rewetted forest peatlands, 66% and 100% in the drained and rewetted percolation mires, 73% and 66% in the drained and rewetted coastal peatlands, respectively (Table 6).

Table 6. Phosphorus species, P_i , and P_o determined by P K X-ray absorption near-edge structure (XANES) analysis.

Site	(cm)		P Species	%		R-Factor #
	Depth	Best Fitted P Reference		P_i	P_o	
AD	0–10	Al-phytate	7	14	85	0.0017
		MgHPO ₄	14			
		IHP	37			
	40–60	Ca-phytate	41	0	100	0.0679
		Ca-phytate	3			
		Boehmite_IHP	97			
AW	0–20	Boehmite_IHP	13	19	82	0.0017
		MgHPO ₄	19			
		Ca-phytate	25			
		IHP	44			
	60–80	Boehmite_IHP	22	0	100	0.0093
		Ca-phytate	78			
CD	0–10	Amorphous AlPO ₄	12	25	73	0.0043
		MgHPO ₄	15			
		IHP	28			
	40–60	Ca-phytate	45	10	90	0.0033
		Amorphous AlPO ₄	10			
		Ca-phytate	40			
CW	0–10	MgHPO ₄	17	34	66	0.0045
		Crystalline FePO ₄	17			
		Ca-phytate	66			
	20–30	Crystalline-AlPO ₄	12	44	56	0.0512
		Ca-phytate	15			
		MgHPO ₄	32			
PD	0–10	Crystalline-FePO ₄	11	34	66	0.0031
		MgHPO ₄	23			
		Ca-phytate	66			
	40–60	Boehmite_IHP	28	0	100	0.0053
		Ca-phytate	72			
PW	0–20	IHP	5	0	100	0.0037
		Ca-phytate	95			
	60–80	AlPO ₄ -Crystalline	10	67	32	0.0065
		Ca(H ₂ PO ₄) ₂	17			
		Ca-phytate	32			
		MgHPO ₄	40			

AD: drained forest peatland, AW: rewetted forest peatland, CD: drained coastal peatland, CW: rewetted coastal peatland, PD: drained percolation mire, PW: rewetted percolation mire. P_i and P_o : the proportions of inorganic and organic P estimated by the linear combination fittings. # The R-factor is a measure of the agreement between the combined spectra of P reference compounds and the spectra of the “unknown” soil sample. The combination of compounds yielding the smallest R-factor was chosen as the most likely representation of P species in the soil samples [34].

4. Discussions

4.1. Total Elements and Plant Available Phosphorus

The abundance of Ca, Mg, Fe, and Al (Table S1) indicated that the studied peatlands are minerogenic [35]. The significant differences of these element concentrations among soil horizons and the peatland types can be associated with differences in pedogenesis and degree of peat degradation. As a result, peat degradation and leaching could be attributed to the low concentration of total Ca in relatively low soil organic matter content of the coastal peatlands [36]. However, the higher P and K (Table S1) concentrations at the surface than at the subsurface horizons can be explained by historical fertilizer applications. A mixture of sand and peat layers between the peat horizons explained the slight increase of the concentration of total P at the 20–40 cm soil depth of the drained forest peat (Table S1). However, the sudden increase in the concentration of total P at the 40–60 cm soil depth of the drained percolation mire could be due to P translocation from the P-rich surface horizon. Overall, the total P content of the surface and subsurface horizons was similar to those reported from fen peatlands of Aschersleben, Saxony-Anhalt and the Trebel valley, Mecklenburg-West Pomerania [19,20]. The lower concentrations of K, Na, and Mn than the other elements are explained by their lower concentrations in plant biomass and losses during the partial decomposition of the peat-forming plant materials since these ions of these elements are less strongly bound than ions of the other elements. The high concentrations of Fe and Al in the peatlands can support the formation of metal bridges between humic substances and P_i complexes [37,38]. The high abundance of Al could be beneficial since Al-containing compounds were applied to increase P sorption in peat soils [39]. However, the high abundance of Fe could form redox-sensitive P compounds that can be mobilized in the rewetted peatland [15].

The values of the plant available P of the present study (Figure 1) were higher than those reported from the different European peatlands extracted by the same method [19]. Similar to the total P (Table 2 and Table S1), the plant available P decreased from the surface to subsurface horizons. This is explained by P release from peat decomposition, historical fertilization, excrement from grazing animals, and external inputs with water and atmospheric deposition. The lower concentration of plant available P at the subsurface than surface horizons and lack of significant differences at the subsurface horizons indicated that P translocation from the surface to the subsurface horizons was negligible or the leached P lost from the soil profiles. Other researchers reported the danger of P leaching from the surface to subsurface horizons of peat soils upon rewetting [14,40]. The disagreeing results can be attributed to variations in current land use, soil organic matter content, and predominant P species or experimental approaches, and peatland types. Overall, the plant available P extracted by double lactate in the present study was lower than those extracted by ammonium lactate and sodium bicarbonate since double lactate cannot extract P bound to organic matter.

4.2. Phosphorus Fractions

The predominance of P_o and residual P fractions with the sequential P fractionation method (Table 1 and Figure 2) agreed with the results of similar studies reported from peat soils of different European countries [19,20]. The lower concentration of P_i than P_o in the present study was not surprising since the soil organic matter contents of the studied peatlands were in the range of 42% to 75% at the surface horizons [21]. The distribution of the H_2SO_4 - P_i fraction in the peat horizons followed the distribution of mineral contents in the peat profiles [21], which means the higher mineral content was associated with higher concentrations of H_2SO_4 - P_i .

The concentration of $NaOH$ - P_o was larger by factor two to that of the concentration of $NaHCO_3$ - P_o at their respective horizon indicating that the major proportion of P_o existed in the moderately labile $NaOH$ - P_o form. Rewetting peatlands with such a high P_o content is unlikely to enhance P mobilization unless the P_i is associated with the redox-sensitive elements [41]. However, the XANES analysis indicated the presence of crystalline $FePO_4$, amorphous $AlPO_4$, and $MgHPO_4$ only in a few horizons in the present study (Table 6). The higher proportions of P recovered by the sequential P fractionation

from the drained than from the rewetted peatlands (Figure 2) indicated that the P stability was higher in the rewetted peatlands. Similarly, the lower concentration of stable P fractions ($\text{H}_2\text{SO}_4\text{-P}$ + residual-P) at the surface horizons of the drained than their respective rewetted peatlands indicated that rewetting of former drained sites could transform moderately labile P into stable P fractions.

Similar to the sequential P fractionation method, the higher concentration of P_o than P_i with the ignition (Table 2) and NaOH-EDTA (Table 3) methods further confirmed that P_o was the predominant P fractions in all the studied peatland types. However, the higher P_i in the ignition method (Table 2) than in the sequential P fractionation (Table 1) and NaOH-EDTA (Table 3) can be attributed to differences in P recovering capacity of the different methods. The sequential P fractionation recovered 54% to 94% of total P while the NaOH-EDTA recovered 51% to 85% of total P except the subsurface horizons of percolation mires. The ignition method was assumed to recover 100% of the P_i and P_o , although some P_o compounds can be transformed to P_i by this method [24]. The percent of total P recovered by the NaOH-EDTA was in the range of the results reported from diverse temperate and tropical peatlands [42]. Similar to the sequential P fractionation method, the NaOH-EDTA also recovered less total P from the rewetted and subsurface horizons than drained surface horizons (Tables 1 and 3) which further confirmed that rewetting increased the P stability.

Comparing the present study with nine major soil groups of 13 studies (Table S2 and Figure 4) clearly illustrated the quality of the Hedley sequential P fractionation in distinguishing labile to stable P fractions in weakly to highly weathered soils. The distribution of different P fractions in the rewetted peatlands of the present study (Histosol^w) (Figure 4) related more to that of the Ferralsols indicating that P is more stable in the long-term rewetted peatlands than in the drained peatlands (Histosol^d) and some other major soil groups. Furthermore, the difference in the distribution of P fractions between the mean of three drained and three rewetted peatlands of the present study were remarkable (Figure 4). Accordingly, the proportion of labile (21%) and moderately labile (30%) P fractions in the drained peatlands were higher than the proportions of labile (12%) and moderately labile (22%) P fractions of the rewetted peatlands (Figure 4). This implies the stable P fractions in the drained peatlands (49%) were less than the stable P fractions in the rewetted peatlands (66%). Overall, the Hedley sequential P fractions not only reflect the influence of land uses and soil-management practices [15,43] but also the influence of soil genesis on P distributions and transformations [44,45]. For instance, the predominance of $\text{H}_2\text{SO}_4\text{-P}$ fraction in the Fluvisol (Figure 4) indicated the abundance of primary phosphates like apatite and P associated with Ca (28). Although most researchers have not determined this, the $\text{H}_2\text{SO}_4\text{-P}$ fraction also contained P_o in peat soils (Table 1). The mean proportion of P_o in the $\text{H}_2\text{SO}_4\text{-P}$ fraction in three drained Histosols was 7% of the total P, whereas that of mean for three rewetted peatlands was 11% of total P as P_o (Table 1). However, regardless of the abundance of P_i or P_o , the $\text{H}_2\text{SO}_4\text{-P}$ fraction is stable and contributes less to soluble and mobile P in the short term.

The higher labile and moderately labile P fractions in the drained peatlands than that of the rewetted peatlands of the present study clearly indicated the P fractions could easily be released upon mechanical disturbance of the drained peatlands. Furthermore, not only drainage but also the current land uses and management practices could influence P transformation and distribution. Soils subjected to continuous cultivation and fertilization could be more dangerous than the drained peatlands used as grasslands. For example, the relative proportion of easily exchangeable P (AEM-P) in Luvisols, Alisols, and Histosols (Figure 4) could more easily contribute to freshwater eutrophication than the drained peatlands of the present study. Thus, not only drainage and rewetting but also current land use and management practices, accessibility to a waterbody, and topographic position could determine whether a peatland is a sink or a source of P.

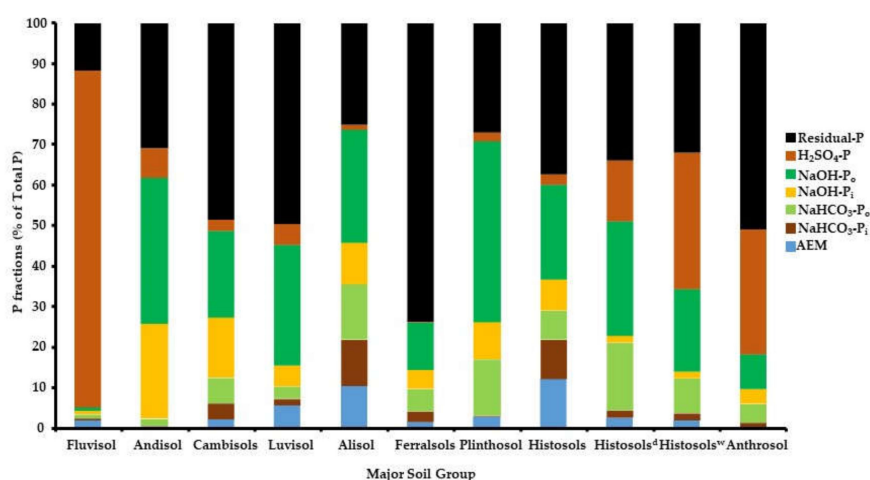


Figure 4. Proportions of different P fractions in some major soil groups and the present study (Histosols^d and Histosols^w). Some of the major groups were the means of more than two studies. Data compiled for this figure derived from the control of each study with no manure and fertilizer application in the reported study if any. We summed to residual any P fraction determined after the fourth P fraction i.e., after 1 M HCl or 1 M H₂SO₄ extraction. Histosols^d and Histosols^w: means of the three drained and three rewetted peatlands of the present study, respectively. Sources: [46–58].

Overall, the results of the present study showed that the response of P to drainage and rewetting depend on the predominant P fractions and biogeochemistry of the studied soils in addition to management practices. For example, rewetting increased the P stability when the predominant P fraction was P_o and drainage increased P_o oxidation regardless of soil types [59,60]. On the other hand, long-term rewetting mineral soils increased P release and decreased the degree of P saturation [61,62]. However, another study reported that drainage increased the maximum P sorption capacity of Cambisol in simulated alternating drainage-rewetting experiment [63]. Such contrasting results can be attributed to variations in the experimental approaches, anthropological factors, and topographic positions. This unequivocally indicates drainage and rewetting cycles can differently influence P dynamics in different soil types. Thus, knowing different P_o and P_i fractions and soil biogeochemistry can help to understand the influence of drainage and rewetting on P dynamics in a given soil.

4.3. Solution ³¹P NMR and P K-XANES

The P species identified by the solution ³¹P NMR analysis agreed with the results of similar studies conducted on agriculturally influenced European peat soils [64]. The predominance of α- and β-glycerol phosphates species in the ³¹P NMR analysis (Figure 3 and Table 5) can be attributed to the hydrolysis of orthophosphate diesters during the alkaline extraction [65]. However, in the ³¹P NMR analysis, the IHP was the second to the glycerol phosphates (Table 5), indicating that P_o in the peat soils was part of structural components of the partially decomposed peats. This is not surprising since the studied peats were partially decomposed with the high C-content [21] that was similar to that of the undecomposed plant materials. A similar study conducted on peatlands of diverse organic matter content also indicated that the IPH abundance was directly proportional to the mineral contents and inversely proportional to soil organic matter content [42].

The lower proportion of P_o recovered by solution ³¹P NMR than P_o recovered by sequential P fractionation, NaOH-EDTA, and ignition methods can be attributed to the high concentration of Fe in the NaOH-EDTA extracts (Table 4). The high Fe concentration (1.09 to 26.09 g kg⁻¹) recovered by the NaOH-EDTA extraction can hinder P speciation by the solution ³¹P NMR since Fe is known to disturb NMR spectra quality in soil and environmental samples due to line broadening and increasing signal relaxation rates [66]. For instance, the concentration of Fe was the highest in the rewetted forest

peatland where the weak signal to noise ratio made it difficult to identify the peaks of orthophosphate monoesters and diesters in this sample (Figure 2b).

In the present study, the P K-XANES analysis was able to distinguish P_o from P_i species; however, it failed to identify different P_o species since all P_o species were identified as if it were IHP (Table 6). This agreed with the recent conclusion derived from a review of many previous studies where only two to five chemical species of P can be identified using one of spectroscopic techniques such as XANES and NMR [67]. The authors also emphasized the matrix complexity of the soil and environmental samples increased the uncertainty in chemical P speciation by most of the spectroscopic techniques. However, the relative proportion of P_i species detected by the XANES analysis at the surface horizons of most sites (Table 6) was similar to the sequential P fractionation and NaOH-EDTA methods (Tables 1 and 3). This is in line with the results of previous studies that indicated XANES is more suitable for P speciation in minerals and crystalline compounds than P_o species [67–69].

4.4. Synthesis

Among the analytical techniques used, the sequential P fractionation method clearly demonstrated the influence of drainage and rewetting on P speciation (Figure 1, Table 1). Although this method has not been used widely for P speciation in peat soils, two global reviews showed that the method can distinguish the impact of soil formation [44] and different land uses and management practices on P dynamics in mineral soils of temperate, tropical, and subtropical climates [15]. Furthermore, rewetting increased the absolute concentration and relative proportions of H_2SO_4 -P and residual-P (Table 1 and Figure 2), which was resistant to extraction by the strong chemicals. Therefore, rewetting degraded peatlands are not an option but a must to lock up P in soil organic matter and transform labile P to stable P forms.

The results of ^{31}P NMR and P K-XANES analyses indicated that P_o was the predominant P species; however, each of the spectroscopic analysis did not decipher the influence of long-term drainage and rewetting on P speciation. Furthermore, degradation of labile P_o species during alkaline extraction, high concentration of Fe and organic matter in the extracts were the challenges during the NMR analysis. The addition of $Na_2S \cdot 9H_2SO_4$, supposed to reduce Fe concentration [30], improved the spectra quality in samples with relatively low Fe content (Table 4 and Figure 2f), but it did not solve the problem for samples with the high Fe content. The high organic matter [21] contributed to high viscosity but the recommended further diluting resulted in the disappearing of peaks from NMR spectra, especially for samples with the low P content. The XANES analysis was able to distinguish P_o and P_i but it was not sensitive enough in distinguishing the impact of drainage and rewetting on P speciation in the studied peat soils.

The abrupt change of total P from the surface (0–20 cm soil depths) to subsurface horizons (20–80 cm soil depths) (Table S1) clearly indicated that P accumulation at the surface horizons originated from relatively recent history since intensive use of chemical P fertilizers have started in the mid-20th century in Germany and elsewhere [70]. The significant differences in total P at the surface horizons of different peatlands can be attributed to variations in P inputs and outputs. However, the significant variations in the distribution of different P fractions could be due to biogeochemical processes in addition to P inputs and outputs to and from the soils. For instance, the distribution of P fractions can change under aerobic-anaerobic conditions [71]. Furthermore, the relative proportion of mineral and organic matter in the horizons influenced the absolute and relative proportion of different P fractions recovered by sequential P fractionation, ignition, and NaOH-EDTA methods (Tables 1–3 and Figure 2). Differences in soil organic matter can explain the higher relative proportion of stable P fractions in the percolation mires than in the costal and drained forest peatlands (Figure 2). The higher proportions of stable P fractions with depths also agreed with the higher soil organic matter content in the subsurface horizons of the percolation mires [21].

Many previous studies reported that rewetting increased P solubility and mobility in soils [13,14,40]. These studies, however, investigated the effect of anaerobic conditions on degraded fen peat that was

previously P-enriched by fertilization. The results of the present study described systems in which new peat has been formed under the influence of longer-term rewetting. In this situation, it is plausible that P was incorporated into stable, mostly organic compounds, in the long-term rewetted peatlands.

5. Conclusions

Both the traditional wet chemical and spectroscopic analyses showed that the predominant P species in the studied peatlands was P_o . However, only the sequential P fractionation revealed the influence of drainage and rewetting on P-distribution and P-transformation in different peatland types. Accordingly, the absolute and relative proportions of P_o and residual-P increased with rewetting, indicating that long-term rewetting increased P sequestration in newly formed peat. Therefore, the key issue for the fate of P under rewetting is the formation of peat that stabilizes P. This view is confirmed by increases in the relative proportions of stable P from the surface to subsurface horizons regardless of differences in absolute P content, management practices, and peatland types. Since the subsurface horizons were water-saturated in most seasons of the year. Any biochemical transformations of P in the subsurface horizons are much lower than in the surface horizons because of fluctuations in the redox status in the later depending on seasons. Therefore, drainage and rewetting influenced P species only in the top 0–20 cm soil depths. Thus, relative proportions of different P fractions, duration of drainage and rewetting, peatland types, and current land systems need to be considered to predict the ecological and economic impact of P due to any peatland management practice.

Supplementary Materials: The following are available online at <http://www.mdpi.com/2571-8789/4/1/11/s1>: Table S1: Concentration of the selected total elements in different peatlands, Table S2: Concentration of different phosphorus fractions in some major soil groups and the present study.

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