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# Modelling spatio-temporal soil moisture dynamics in mountain tundra

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## Funding information

Soil moisture has a fundamental influence on the processes and functions of tundra ecosystems. Yet, the local dynamics of soil moisture are often ignored, due to the lack of fine resolution, spatially extensive data. In this study, we modelled soil moisture with two mechanistic models, SpaFH<sub>y</sub> (a catchment-scale hydrological model) and JSBACH (a global land surface model), and examined the results in comparison with extensive growing-season field measurements over a mountain tundra area in northwestern Finland. Our results show that soil moisture varies considerably in the study area and this variation creates a mosaic of moisture conditions, ranging from dry ridges (growing season average 12 VWC%, Volumetric Water Content) to water-logged mires (65 VWC%). The models, particularly SpaFH<sub>y</sub>, simulated temporal soil moisture dynamics reasonably well in parts of the landscape, but both underestimated the range of variation spatially and temporally. Soil properties and topography were important drivers of spatial variation in soil moisture dynamics. By testing the applicability of two mechanistic models to predict fine-scale spatial and temporal variability in soil moisture, this study paves the way towards understanding the functioning of tundra ecosystems under climate change.

## KEYWORDS

soil moisture, Arctic, tundra, hydrological modelling

## 1 | INTRODUCTION

Soil moisture is a crucial part of the hydrological cycle, influencing interactions between the land surface and atmosphere (Robock et al., 2000; Koster et al., 2004; Seneviratne et al., 2010). In mountain tundra, the importance of soil moisture is highlighted in its connection with vegetation patterns and Earth surface processes (Aalto et al., 2013; le Roux et al., 2013). It is also strongly linked to plant performance and ecosystem functionality, emphasising its ecological relevance under contemporary climate change (Bjorkman et al., 2018). Tundra ecosystems are characterised by a short but intensive growing season and a prolonged snowmelt period, which is strongly correlated with local topography (Niittynen et al., 2018). In cold climates, topography also influences the distribution of a spatially uneven organic layer (Seibert et al., 2007; Zhu et al., 2019). Processes linked to climate, varying topography and vegetation characteristics interact with soil moisture, causing spatial and temporal variation in its fine-scale patterns (Kemppinen et al., 2018; Penna et al., 2009).

At large scales, the spatio-temporal variation of soil moisture follows general climatic conditions (Seneviratne et al., 2010). However, at finer scales, its patterns are controlled by various landscape characteristics as well as local climate. Water flows within and above ground are controlled partly by soil hydraulic properties, topography and intensity of precipitation, while evapotranspiration is influenced by vegetation characteristics and radiation (Western et al., 2002; Seneviratne et al., 2010). These fine-scale variations of soil moisture can be considerable, particularly in heterogeneous landscapes, such as in mountain tundra, and need to be understood when considering area-averaged soil moisture variations at larger scales (Western et al., 2002). They also have important local ecosystem impacts. Spatio-temporal variations of soil moisture are an important driver of greenhouse gas fluxes (Lohila et al., 2016; Virkkala, 2020) as well as fine-scale patterns of vegetation properties (le Roux et al., 2013; Kemppinen, 2020). Therefore, through various feedback mechanisms, soil moisture in the tundra plays an important role in global change and its accurate predictions are fundamental to our ability to understand tundra ecosystems now and in the future.

Mechanistic models are useful tools in examining dynamic and complex processes, such the hydrological cycle (Abbott et al., 1986; Fatichi et al., 2016). Models depicting soil moisture dynamics have been developed for various applications, such as estimating global wetland areas, improving catchment scale flood forecasts and simulating fine-scale species distribution patterns (Berthet et al., 2009; Maclean et al., 2012; Zhang et al., 2016). Therefore, the level of detail in how and which hydrological processes are described varies even amidst similar models such as land surface models (Dirmeyer et al., 2006; Koster et al., 2009; Romano, 2014). As a result, model comparison and evaluation studies have found considerable differences when simulating soil moisture and its dynamics (Dirmeyer et al., 2004; Koch et al., 2016; Yuan and Quiring, 2017). In mountain tundra, where landscape heterogeneity is an important aspect of soil hydrology, evaluating soil moisture model performances requires spatially detailed measurements. Recent developments of *in situ* measurement techniques have improved the spatio-temporal resolution in which soil moisture can be measured and in turn provide more a comprehensive understanding of soil moisture dynamics and model performances (Kopecký et al., 2021; Vereecken et al., 2014; Wild et al., 2019).

The objectives of this study are to 1) quantify the spatio-temporal variability of soil moisture and its drivers in mountain tundra and 2) evaluate soil moisture simulations of two mechanistic models using high-resolution soil moisture field measurements. To the best of our knowledge, this is the first time that extensive, high-resolution observation data are used in detailed model-based analysis to unravel the mountain tundra soil moisture variability and

64 drivers. One of the models, JSBACH, has been developed for use with large-scale climate models and concentrates on  
65 interactions between the land surface and atmosphere (Reick et al., 2013). Understanding how variation in landscape  
66 characteristics realises itself in model results is important in order to better understand what uncertainties relate  
67 to large-scale simulations of the land surface. The other one, Spatial Forest Hydrology Model (SpaFHy), has been  
68 developed for simulating catchment-level hydrology in boreal forests (Launiainen et al., 2019). The soil moisture  
69 estimations in sloping terrain with patchy soil and vegetation require specific capabilities from models, some of which  
70 are found in the more soil-vegetation oriented models and some in hydrology oriented models. Here we assess the  
71 impact of soil type, layers and implementation of topographical redistribution on soil moisture variation and test the  
72 models' ability to characterise the spatial and temporal variation of soil moisture with literature-based parameters.  
73 We also use a statistical model to examine which environmental variables, namely soil, topography and vegetation,  
contribute most to the soil moisture variation, and whether these controls are well addressed in the models.

## 75 2 | MATERIALS AND METHODS

### 76 2.1 | Study area

77 The study area is located in a valley between Mount Saana and Mount Jehkas in northwestern Fennoscandia (69°03'  
78 N 20°51' E, Fig. 1). The region experiences a subarctic climate with monthly average temperatures ranging from  
79 -12.9 °C in January to 11.2 °C in July (averages during 1981–2010; Pirinen et al., 2012). The total annual precipitation  
80 is 487 mm and snow covers the ground largely from October to May, although late-lying snowpacks can persist far  
81 into the summer. The landscape is characterised by varying vegetation (Riihimäki et al., 2019), soil type, geomorphol-  
82 ogy (le Roux and Luoto, 2014) and topography (Kemppinen et al., 2018). Vegetation consists mainly of dwarf-shrub  
dominated mountain heath with sporadic meadows and mires (Kemppinen et al., 2018; Riihimäki et al., 2019). The  
84 ground surface consists of thin mineral and organic soil layers that are partly covered by eroded boulders and exposed  
85 bedrock. Tundra mires with thicker layers of organic soil have formed mainly in the valley and flat upland areas in the  
86 west. The environmental variation is driven by fine-scale variation of topography, with relative elevation difference  
reaching nearly 250 meters (Aalto et al., 2013; le Roux et al., 2013).

88 [Insert Figure 1]

### 89 2.2 | Study setting

90 In this study, we measured the local variation of top soil moisture using 50 soil moisture loggers (TMS-4 datalogger;  
91 TOMST s.r.o., Prague, Czech Republic). They were installed in June 2018 and their locations recorded with an accuracy  
92  $\leq 6$  cm using a hand-held Global Navigation Satellite System (GeoExplorer GeoXH 6000 Series; Trimble Inc., Sunnyvale,  
93 CA, USA). The loggers were situated to represent the entire soil moisture gradient of the landscape (Fig. 1) based on  
previous field studies in the area, with particular attention paid to the extremes: the water-logged peatlands (average  
95 soil moisture level > 60 VWC% (Volumetric Water Content); 10 loggers) and the dry ridges and mountain tops (< 15  
96 VWC%; 10 loggers) (Happonen et al., 2019; Kemppinen et al., 2018, 2021). Some of the loggers are situated close to  
97 each other rather than evenly around the study area in order to describe the very fine-scale patterns of soil moisture  
98 variation caused by the spatially heterogeneous soil properties and topography of mountain tundra. They measure  
99 moisture to a depth of c. 14 cm below ground at 15 minute intervals (Wild et al., 2019). In the data processing, the  
100 raw time-domain transmission data were calibrated into VWC using a conversion tool provided by the manufacturer.  
101 The measurement uncertainties related to these loggers and their calibration have been discussed in Wild et al. (2019).

102 Calibration curves were chosen based on field-quantified soil moisture measurements recorded with a hand-held time-  
 103 domain reflectometry sensor (FieldScout TDR 300; Spectrum Technologies Inc., Plainfield, IL, USA) during summer  
 104 2018 (four measurement campaigns in July–August). Of the 50 loggers, two were excluded from the analyses since  
 105 they had dislocated during the study period. Thus the final data consisted of 48 loggers.

## 106 2.3 | Model input data

107 [Insert Table 1]

108 Environmental data required by SpaFH<sub>y</sub> and JSBACH were obtained from remote sensing techniques and field  
 109 surveys (Table 1). The soil data, consisting of a rasterized classification of the surficial deposits (Fig. 1) and point  
 110 measurements of the mineral and organic layer depths, have been described in detail by Kemppinen et al. (2018). In  
 111 the brief, the surficial deposit map was created using field surveys and aerial images (0.5 m \* 0.5 m resolution) provided  
 112 by the National Land Survey of Finland. Three layer depth measurements in a 1 m \* 1 m plot were taken every 50  
 113 meters from the whole study area. Three soil types were then defined for both models. Glacial till, fluvial deposits  
 114 and boulders from the surficial deposits map were classified as mineral soils and peat deposits as peat soils. A third  
 115 soil type was defined as a mixture of organic and mineral soil based on the average proportion of each layer. This soil  
 116 type was classified as a combination from the surficial deposits map and vegetation type map as meadow and mire  
 117 vegetation overlaying mineral soil (glacial till, fluvial deposits or boulders). Soil parameters for SpaFH<sub>y</sub> were kept close  
 118 to those used in Launiainen et al. (2019). While a full sensitivity analysis for soil parameters was outside the scope  
 119 of this study, we adjusted the field capacity in peat soils following a sensitivity analysis (Figure A1). This was done  
 120 because the original parameters led to noticeably drier VWC% which is likely due to differences in peat soil properties  
 121 in this study area. Soil type specific parameters for JSBACH were taken from Hagemann and Stacke (2015) (Table A1).  
 122 Both models describe vegetation by type and coverage. To create a raster of vegetation types, we utilised a  
 123 Random Forest (RF) model trained by vegetation observations and five PlanetScope images (resolution 3 m \* 3 m)  
 124 from growing season 2018 (Breiman, 2001; Planet Team, 2017). The RF model was run 100 times by bootstrapping  
 125 the training data. The final pixel values were determined as the most common class value from a five-class vegetation  
 126 classification including meadows, deciduous shrubland, evergreen shrubland, barren tundra and wetlands. In SpaFH<sub>y</sub>,  
 127 the parameters (Table A1) for these classes were obtained from the literature (Launiainen et al., 2019; Lin et al., 2015;  
 128 Pop et al., 2000; Starr et al., 2008). In JSBACH, we used the plant functional types of peatland (wetland class), C3  
 129 meadow class) and tundra (deciduous and evergreen shrubland) with their default parameter values (Kattge  
 130 et al., 2009; Knorr et al., 2010). To estimate vegetation cover, we calculated the Normalized Difference Vegetation  
 131 Index (NDVI) from a Sentinel-2 image taken in August 2019 (ESA, 2021) using Eq. (1)

$$NDVI = \frac{NIR - red}{NIR + red} \quad (1)$$

132 where NIR and red refer to the near infrared and red bands (Huete et al., 2002). For SpaFH<sub>y</sub>, the maximum leaf area  
 133 index (LAI) was then calculated from NDVI and the vegetation type map based on an approach by Street et al. (2007).  
 134 Topography variables were calculated from a LiDAR-based (light detection and ranging) Digital Elevation Model  
 135 (DEM; horizontal resolution 2 m \* 2 m, vertical resolution 30 cm; NLS, 2020). SAGA Wetness Index (SWI) (Böhner and  
 136 Selige, 2006) can be used as a proxy for soil moisture similarly as the original Topographic Wetness Index (TWI) (Beven  
 137 and Kirkby, 1979). However, SWI is an algorithm specific to SAGA GIS (Conrad et al., 2015) and is a modified version  
 138 of the Multiple-flow Freeman algorithm (FD8f) (Freeman, 1991). Different from FD8f, SWI uses a modified catchment  
 139 area. Thus, SWI produces a spatially smoothed TWI distribution, that is, a smooth stream network (Kopecký et al.,

140 2021). This means that compared to the original TWI, SWI allows low-lying flat areas close to flow channels to have  
 141 higher index values (Böhner and Selige, 2006). We calculated SWI using t-value 10 (that is, the suction effect) and a  
 142 filled DEM following Wang and Liu (2006) in SAGA GIS hydrology module (Böhner and Antonić, 2009) with specific  
 143 catchment area and local slope methods following Eq. (2)

$$SCA_M = SCA_{max} \frac{1 - \beta \exp(20\beta)}{20} \quad \text{for } SCA < SCA_{max} \frac{1 - \beta \exp(20\beta)}{20}$$

$$SWI = \ln \frac{SCA_M}{\tan(\beta)} \quad (2)$$

144 where  $SCA$  and  $SCA_M$  are the specific and modified specific catchment areas,  $\beta$  is the slope angle and  $\tan(\beta)$  is the  
 145 local slope (Böhner and Selige, 2006).

The shadowing influence of topography on incoming solar radiation was calculated for each month of the year  
 147 from the DEM using the potential incoming solar radiation module with a sky view factor option and lumped atmo-  
 148 spheric transmittance in the RSAGA package (Böhner and Antonić, 2009). The monthly values were then divided  
 149 by the potential radiation received by a flat surface in the same latitude, and interpolated to obtain daily correction  
 150 factors for incoming solar radiation at each grid-cell for both models.

The meteorological data for January 2015–September 2019 were obtained from the Finnish Meteorological In-  
 152 stitute's Kilpisjärvi kyläkeskus meteorological station (69°02' N 20°47' E, 480 m a.s.l.; Finnish Meteorological Insitute,  
 153 2020) ca. 1.5 km southwest from the centre of the study area. The daily variables used were air temperature (°C),  
 154 precipitation (mm d<sup>-1</sup>), relative humidity (%), wind speed (m s<sup>-1</sup>), wind direction (°) and air pressure (hPa). Global  
 radiation (W m<sup>-2</sup>) was extracted from the 10 km \* 10 km gridded dataset provided by Finnish Meteorological Insitute  
 156 (2009).

## 157 2 | Models

### 158 2.4.1 | SpaFH<sub>y</sub>

159 The Spatial Forest Hydrology Model (SpaFH<sub>y</sub>) is a semi-distributed hydrological model developed to simulate evapo-  
 160 transpiration and water balance in a boreal forest landscape (Launiainen et al., 2019). It has been tested both at the  
 161 grid and at the catchment level at various sites in Finland, including a catchment similar to this study area. SpaFH<sub>y</sub>  
 162 consists of three submodules that simulate water balance above ground, within topsoil and within the catchment.  
 163 Above-ground processes are included in the canopy module, which describes the processes related to vegetation,  
 164 ground surface and snowpack. Vegetation is divided into classes, which in this study include deciduous and conif-  
 165 erous shrubland and mire vegetation. These differ mainly in their seasonal cycle, water usage and photosynthetic  
 166 capacities. Soil moisture is depicted as a two-layer bucket model which consists of an organic top layer and a root  
 167 layer. The organic top layer is a shallow layer (4 cm) with soil properties similar to peat soils. The root layer depth  
 168 is set to 20 cm to keep it close to the measured soil moisture. Finally, the TOPMODEL submodule links the grid  
 169 cell water balance conceptually with the catchment-scale water balance through the subsurface storage bucket. The  
 170 saturation deficit of each grid cell is linked to the average saturation deficit of the whole catchment subsurface storage  
 171 so that grid cells with higher index values, in this study SWI values, are more likely to be saturated (Launiainen et al.,  
 172 2019). This allows accumulation of soil water in lowland areas with high SWI values and dynamic formation of water-  
 173 saturated areas. Using SWI instead of TWI means that soil water should accumulate more evenly in flatland areas with  
 174 high index values. In general, this modelling approach allows recognising and describing the landscape-level hetero-

175 geneity in biogeophysical conditions through geospatial data and remote sensing methods and linking this variability  
176 to a spatially explicit mechanistic model in order to better understand landscape-level hydrological processes.

177 SpaFHy was run in daily time steps as a catchment-scale version. For input data, it requires raster files of maximum  
178 LAI for each vegetation class, as well as canopy height, soil type, SWI and masks of the catchment area and water  
179 bodies. Spatial resolution was set to 10 m \* 10 m. Canopy height was set to 0.5 m as its influence is negligible in low  
180 vegetation (Launiainen et al., 2019).

## 181 2.4.2 | JSBACH

182 JSBACH is the land surface model of the Max Planck Institute for Meteorology (MPI-M) Earth System Model (Gior-  
183 nani et al., 2013; Reick et al., 2013). It describes processes involved in the interactions between the lower level of the  
184 atmosphere and land surface and has been used in several studies simulating biogeophysical and -chemical processes,  
185 including hydrological research (Gößling and Reick, 2011; Gao et al., 2016; Heidkamp et al., 2018). Structurally, JS-  
186 BACH consists of several submodules that describe the terrestrial energy balance, heat transfer and water budget,  
187 vegetation dynamics and phenology, carbon cycle over land, land cover change and surface albedo (Böttcher et al.,  
188 2016; Groner et al., 2018; Hagemann and Stacke, 2015; Heidkamp et al., 2018; Raddatz et al., 2007; Thum et al.,  
189 2011). Vegetation is described through plant functional types, which are included in each grid cell as overlapping tiles.  
190 Each grid cell, of user-defined resolution, can thus have several vegetation types.

191 In JSBACH, the vertical movement of soil moisture is depicted through one-dimensional Richard's equation which  
192 is typically used in soil moisture modelling to study processes related to interactions between land surface and atmo-  
193 sphere (Romano, 2014). In the new hydrology scheme developed by Hagemann and Stacke (2015), the soil profile  
194 consists of five layers with increasing depths up to 10 m, improving descriptions of bare soil evaporation and soil  
195 moisture buffering. Soil properties in each layer are kept constant. The actual soil depth is controlled through a soil  
196 depth variable and a root depth variable controls the depth from which transpiration may occur. Water flow between  
197 grid cells is not accounted for and each grid cell acts as a separate hydrological unit.

198 Here, JSBACH was run as an offline version with user-generated meteorological forcing data with modules bethy,  
199 phenology, albedo and yasso turned on. The model was run over 210 independent grid cells to allow for spatial  
200 variation in the input data, namely in soil properties, vegetation characteristics and topographical shading of solar  
201 radiation. Unlike in global simulations with spatially averaged soil properties, specific soil classes were used to describe  
202 conditions in the landscape. Surface parameters were taken from Hagemann (2002) using parameters for fens and  
203 bogs, upland tundra and polar deserts (Table A1). Minimum soil and rooting depths were set to 0.5 m as soil depths  
204 less than 0.5 m led to negligible transpiration and canopy conductance rates. A spin-up run of three years prior to  
205 the study period was performed for both models in order to equilibrate slowly changing variables. To visualize spatial  
206 variation in JSBACH, the point-based results were mapped to matching environmental conditions in the study area.

## 207 2.4.3 | Statistical model

208 Field data was used to address the drivers of soil moisture variation using a Generalised Additive Model (GAM). It  
209 allows for non-linearity in the relationship between response and predictor variables by splitting the regression line  
210 into segments to which the regression line is fitted using a user-controlled smoothing function (Hastie and Tibshirani,  
211 1987). We modelled the spatial variation of growing season average soil moisture and the temporal range of variation  
212 (growing season maximum VWC% - growing season minimum VWC%) as a function of organic soil depth (up to 80 cm),  
213 vegetation cover (%), topographical shading of incoming solar radiation, elevation and SWI. The variables describing

soil and topography are commonly used in soil moisture research (Kemppinen et al., 2018; Williams et al., 2009). Vegetation cover has a more complex relationship with soil moisture but was used as a predictor variable due to its influence on temporal soil moisture dynamics in the mechanistic models used in this study.

The model was fitted using the `mgcv` package in R (version 3.4.4; Wood, 2011; R Development Core Team, 2020), with maximum degrees of smoothing restricted to three. The response variables were log-transformed to approximate normal distribution and then transformed back before plotting them. Effect sizes for each predictor were calculated based on the predicted minimum and maximum VWC% and range of VWC% over the field data while other terms were held constant at their mean values. To quantify observation-related uncertainty in model estimates, we used bootstrap sampling with 200 repetitions. A similar model was developed for the results of SpaFH<sub>y</sub> and JSBACH using soil porosity, vegetation cover, SWI (in SpaFH<sub>y</sub>) and solar radiation as the predictor variables in order to estimate the influence of these variables in the mechanistic models.

## 2.5 | Analysis of results

Field measurements and model simulations were grouped into three regimes in order to reduce uncertainty related to the accuracy of single logger data and to examine temporal variation of soil moisture. As soil conditions influenced logger and model average VWC% considerably (Fig. A2), the field measurements were classified based on organic layer depth to xeric (organic depth < 5 cm, 20 loggers), mesic (organic depth 5–25 cm, 18 loggers) and hydric (organic depth > 25 cm, 10 loggers) regimes. Model results were classified similarly based on soil type. Growing-season months with no extensive snow cover (July–September) in 2018 and 2019 were selected for further analysis. For JSBACH, the weighted average VWC% of the top two soil layers (6 and 25 cm thick) was calculated (center depth 16 cm), and the same was done for SpaFH<sub>y</sub> for the top and root zone layers (center depth 12 cm)(Table A1). In order to estimate soil moisture variation in the landscape, we calculated growing season averages from July–September 2018–2019 and used range of variation VWC% of the same time period as a measure of temporal variation. To show the temporal correlation between the timeseries of modelled and measured VWC%, measurements and model outputs were averaged over the regimes. Then the growing season average of each timeseries was deducted from each averaged timeseries for simpler plots and the amount of explained variance ( $R^2$ ) by a linear regression model was calculated in each regime.

## 3 | RESULTS

### 3.1 | Temporal variation of soil moisture

The temporal patterns of soil moisture are distinctively different among the moisture regimes (Fig. 2, Fig. A3 and Table 1). According to field measurements, the xeric regime has on average 14 VWC% throughout the growing season. There is little variation in the VWC% but clear short-term responses to precipitation events. In the mesic regime, average VWC% is 26 %, and in the hydric regime 61 %. Variation between growing season months is higher in the hydric regime, with August being the driest month (10 VWC% variation). In the xeric and mesic regimes, monthly averages are nearly constant (1–2 VWC% variation) and range of variation over the whole growing season is low (15 VWC%). Apart from one logger, the time series measured in the xeric regime are closer to each other compared to the hydric and mesic regimes.

[Insert Figure 2]

The modelled temporal variation of soil moisture followed the characteristics of these regimes reasonably closely



(Fig. 2 and Fig. 3). Average growing season soil moisture ranged between 17–60 VWC% in JSBACH and 15–48 VWC% in SpaFH<sub>y</sub> (Table 2). Modelled range of variation was lower than the measured range in all regimes, with JSBACH closer to measurements in the mesic (JSBACH 12 VWC%, measured 16 VWC%) and hydric (12 VWC% and 44 VWC%) regimes and SpaFH<sub>y</sub> in the xeric regime (10 VWC% and 15 VWC%). Variation between monthly averages was low (< 3 VWC%) in all regimes, although variation was higher in the hydric regime than in the xeric and mesic regimes. The minimum values in SpaFH<sub>y</sub>'s hydric regime decreased notably compared to average values, otherwise the minimum and maximum values in both models were close to regime averages (Fig. 2).

[Insert Table 2]

The correlation between modelled and measured timeseries also depended on the moisture regime (Fig. 3). The highest  $R^2$  for both models (0.60 for JSBACH and 0.72 for SpaFH<sub>y</sub>) was in the hydric regime while  $R^2$  values in the xeric regime were around 0.5–0.6. However, in the mesic regime the results were more scattered, and thus,  $R^2$  was lower for both models. The slopes for both models in the hydric regime were fairly large, indicating that while the models are capable of producing the temporal patterns, their magnitudes are smaller compared to measurements. SpaFH<sub>y</sub>'s  $R^2$  was higher than JSBACH in all regimes but it also had higher slopes in all regimes.

[Insert Figure 3]

## 3.2 | Spatial variation in soil moisture

[Insert Figure 4]

Spatial variation of soil moisture in the landscape was considerable (Fig. 4). Both model results and field measurements showed that dry conditions (15 – 20 VWC%) dominate the landscape while wetter regimes (> 20 VWC%) are concentrated mostly in flatter areas in the west and in the valley between the two fells. However, modelled spatial variability of soil moisture across the landscape was smaller than observed, with model results concentrating close to regime averages and measured averages spread more evenly between regime averages.

Temporal variation was generally higher (range > 25 VWC%) in wetter areas in the field measurements. However, range in field measurements was more scattered, with some drier (wetter) loggers also showing high (low) temporal variation (average VWC 15 % (75 %) and range 45 % (15 %)). In the model results however, there was significantly less temporal variation in general, with maximum range values below 20 VWC%. In both models, wetter areas had generally higher temporal variation as well.

The statistical model GAM explained 79 % of the spatial variation in average soil moisture (Fig. 5). Organic layer depth controlled a large part of the variation (effect size 29 VWC%) with thin organic layers resulting in lower VWC. In thick organic layers (> 50 cm) the fitted function was associated with large uncertainty, partly due to fewer measurements. The second most influential variable was SWI (effect size 22 VWC%), which also had a positive relationship with VWC, meaning that high SWI values, found in lowlands and local depressions, had on average higher VWC% than upland areas. Other variables had only a minor or no clear effect on average soil moisture. In SpaFH<sub>y</sub> and JSBACH, only soil porosity had any notable effect on the average VWC% (effect size 36 VWC% and 42 VWC% respectively).

GAM explained 44 % of the temporal variation in the data. Organic layer depth and SWI had the largest effect sizes (42 and 11 VWC% respectively). Other variables had no clear effects. In SpaFH<sub>y</sub>, vegetation cover had the strongest influence on range of variation (effect size 7 VWC%), while in JSBACH soil type played the most important role (effect size 8 VWC%).

[Insert Figure 5]

## 291 4 | DISCUSSION

292 Our study shows that soil moisture exhibits considerable spatial and some temporal variation in the study area, cre-  
293 ating distinct moisture regimes (Fig. 2 and Fig. 4). Most of the study area is characterised by dry average moisture  
294 conditions (xeric regime) in which there is little variation between monthly averages during the growing season. Wet-  
295 ter moisture conditions, including mesic and hydic regimes, are found in low-lying and depression areas with thicker  
organic soil layers (> 25 cm) and exhibit higher average moisture conditions. There is greater temporal variation, par-  
297 ticularly in the hydic regime. This is partly due to the lower water retention ability of coarse mineral soils compared  
298 to soils with more organic material.

299 The mechanistic models capture parts of the temporal variation fairly well (Fig. 2 and Fig. 3). Spatial and temporal  
300 variation is similar in all three moisture regimes. In the xeric regime, the responses to precipitation events are similar in  
301 the models and measurements, although JSBACH in particular underestimates the range of variation (Table 2). In the  
302 mesic and hydic regimes, the average time series show little temporal variation compared to measurements, although  
303 SpaFH's minimum values follow a similar monthly pattern as the measurements (Fig. 2). While average temporal and  
304 spatial patterns are similar in the models and measurements, both models underestimated the range of variation.  
305 Particularly temporal variation might benefit from adjusting soil properties in the model simulations. However, we  
306 wanted to retain the existing features and scalability of the regional model. Further, removing the cause of the lack  
307 of variation in a proper way might require more than just parameter tuning, including re-think of the model set-up for  
308 vegetation and soil.

309 Previous studies in the study area have linked the fine-scale spatial variation of soil moisture to the environmental  
310 gradients of the landscape, such as the varying topographical conditions and soil properties (Kemppinen et al., 2018).  
311 Our results indicate that a large part of the spatial variation of soil moisture can be attributed to soil properties (Fig. 5).  
312 Thin mineral soils are not as efficient at retaining water as thick organic soils, and in turn, the former dry quickly after  
313 precipitation events compared to the latter, which stay more stable by retaining soil moisture (Legates et al., 2011;  
314 Migala et al., 2014). Although the importance of soil properties is also evident in the model results, it does demonstrate  
315 a common problem in hydrological process-models and soil properties. Soil properties can vary considerably over short  
distances, particularly in a landscape such as mountain tundra where the soil layer can be thin and the accumulation  
317 of soil organic matter depends on topography (Migala et al., 2014; Seibert et al., 2007). However, measuring this  
318 variability at a high spatial resolution and broad spatial extent is challenging, and consequently the input data in  
319 hydrological models cannot account for real variability in soil hydrological properties. This and the lack of spatial  
320 variation in the organic soil layer are likely to explain a large part of the underestimated spatio-temporal variation of  
321 modeled soil moisture (Fig. 4 and Fig. 5).

322 Vegetation seems to little to no effect on average VWC% or its temporal range (Fig. 5). Previous studies have  
323 shown that woody vegetation cover in particular can decrease soil moisture in the tundra (Kemppinen et al., 2021), for  
324 instance through increased transpiration (Pearson et al., 2013). The estimations on the influence of vegetation cover  
325 on the temporal variation of soil moisture seem to differ in SpaFH and JSBACH, which might explain the lack of spatial  
326 variation in JSBACH (Fig. 4). Figure A4 shows that in JSBACH vegetation cover does influence transpiration. However,  
327 as transpiration extracts water first from a deeper soil depth, it does not instantaneously control soil moisture in the  
328 upper soil layers. This leads to the apparently negligible influence of vegetation to the soil moisture in JSBACH.

329 Topography, here accounted through SWI, was found to influence spatial and temporal variability of soil moisture  
330 (Fig. 5). Mechanistically modelling this variation is possible on catchment level models such as SpaFH, but in this study  
331 the influence of SWI on SpaFH's results was small. This is likely because peatland areas in the study area are strongly  
332 concentrated in areas with high SWI, as organic matter accumulates in local depressions and flatlands, similarly to soil

moisture (Fig. 1). However, high SWI values in particular could explain the lack of temporal variation in the hydric regime (Fig. 2), as the distribution of water in SpaFH<sub>y</sub> might cause these areas to be nearly constantly saturated. In JSBACH, the influence of hillslope level topography on soil hydrology is ignored as global land surface models focus on large-scale processes. However, hillslope level processes are important for land surface – atmosphere interactions and recent studies have focused on incorporating this subgrid variability in global scale models (Fan et al., 2019). Another aspect of topography that might be important in this study area is snow distribution. While typically most of the snow in the study area melts by May, the spatially and temporally uneven snowmelt period creates hydrologically distinct conditions such as meltwater streams and late-lying snowpacks, which in turn influence soil moisture far into the growing season (Niittynen et al., 2018; Sturm et al., 2005; Woo et al., 2006). Incorporating this variation into mechanistic models and examining its influence on soil moisture variation is an important future research question, as snow conditions are predicted to change considerably due to climate change (Bintanja and Andry, 2017; Fountain et al., 2012).

Previous research into soil moisture products has revealed that different outputs may not be intercomparable (Dirmeyer et al., 2006; Koster et al., 2009). In this study, we have compared three data sources for soil moisture and the model parameters and input data have been harmonised in as much detail as the model structures allow, it is important to understand how these results differ from each other. Firstly, the field loggers describe soil moisture conditions in an exact point and are thus considerably influenced by for example fine-scale soil heterogeneity. In comparison, the input data resolution and model configuration in both SpaFH<sub>y</sub> and JSBACH mean that they describe moisture conditions in a larger area. Thus, comparing specific measurements to their matching model grids is not particularly helpful. Here, we have instead examined more generalised results by grouping both models and measurements to moisture regimes to diminish the uncertainty related to specific point measurements. Secondly, the soil moisture loggers describe topsoil conditions which interactions between land surface and atmosphere influence the most. In some areas of the study site, the soils are shallow enough, and thus, we can assume that the loggers describe the whole vertical soil moisture content which might not be the case in areas where soils are thick. The models also describe moisture deeper than what the loggers can reach (> 15 cm depth) and in a larger vertical space than the measurements. In SpaFH<sub>y</sub>, the root zone depth was set to 0.2 in order to make it more comparable with measurements and the average soil conditions in the study area. In areas with shallow soils, the results represent the measured conditions fairly well. However, for example in areas with deep peat deposits, the model does not take the influence of soil depth into account. In JSBACH, which does simulate the vertical moisture profile more explicitly, the model results have been calculated from the top two layers in order to make them more comparable with the measurements. However, the minimum depth of 0.5 m in JSBACH means that soil moisture variation in mineral soils particularly is dampened by buffering from deeper layers. These aspects may explain the relatively large slopes between modelled and measured timeseries (Fig. 3). To conclude, the models and measurements are not entirely comparable with each other from all perspectives. However, while the precise estimates of VWC may vary, the similarities and disparities between spatio-temporal dynamics of each product may still be compared and used to inform soil moisture dynamics (Koster et al., 2009; Saleem and Salvucci, 2002).

Our results show that in order to model the spatio-temporal variation of soil moisture accurately in mountain tundra, soil properties, including the thickness of the organic layer, are important. In spatially distributed models, such as SpaFH<sub>y</sub>, this requires developing methods to depict the soil organic layer and ways to infer soil hydrological properties at high spatial resolution. Recent advances in modelling spatial variation in soil properties based on DEMs (Li et al., 2020) could provide an option which should be further tested in mountain tundra. Another possibility is utilising remotely sensed soil moisture datasets to better understand its spatial variation (Mohanty et al., 2017; Manninen et al., 2021). Although global land surface models, such as JSBACH, cannot capture the fine-scale variation of soil

376 properties or other environmental variables in a similar fashion to catchment-scale models, methods of describing  
377 sub-grid heterogeneity in such models do exist. In JSBACH, vegetation cover is described through tiles to allow for  
378 multiple plant functional types within one grid cell (Reick et al., 2013). Hydrologic response units have been used  
379 similarly to include variation in soil properties (Chaney et al., 2016). Description of the soil organic layer has also been  
380 found to be important in land surface models (Rinke et al., 2008; Ekici et al., 2015).

381 Considering temporal soil moisture dynamics, range of variability as a static measure ignores many aspects such  
382 as temporal resolution (whether the variation is linked solely to short-term variation in, for example, precipitation, or  
383 more seasonal variation) as well as temporal development in the land surface variables that influence soil moisture  
384 such as vegetation phenology and snowmelt dynamics). These aspects can be included in mechanistic models as well  
385 as estimates of future changes in, for example, climate or vegetation patterns. Thus, our findings suggest that more  
386 effort should be made in considering local processes that influence soil moisture dynamics, possibly through a fusion  
387 of remote sensing, in situ data and mechanistic models. In a time of rapid environmental changes in the tundra, such  
388 methods will be fundamental in making dynamical future predictions on the functioning of Arctic ecosystems.

## 389 | CONCLUSIONS

390 To contribute to the understanding soil moisture dynamics in mountain tundra, we modelled its spatial and temporal  
391 variation using extensive field measurements and two mechanistic models, SpaFHy and JSBACH. We found substantial  
392 landscape-scale spatial variation in soil moisture ranging from dry mineral soils to wet peatlands. By investigating the soil  
393 moisture dynamics, we identified distinct hydrological regimes over the landscape. Our results show that mechanistic  
394 models are able to simulate average VWC% conditions within the regimes but underestimate both temporal and spatial  
395 variation compared to measurements. Spatial variation of soil moisture was largely related to soil properties in both  
396 model simulations and measurements. Our results indicate that improving these descriptions as well as simulations  
397 of soil moisture variability in mechanistic models is needed to improve modelling of soil moisture dynamics in tundra  
398 ecosystems. The results are important for understanding uncertainties related to global and regional analyses and  
399 to inform future model developments needed to understand the ecosystem consequences of the Arctic change.

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594 **Tables**

**TABLE 1** Geospatial datasets used in the study. Soil layer data are based on field surveys, whereas vegetation types and surficial deposits were classified based on a combination of field and remote sensing data. NDVI (Normalized Difference Vegetation Index) was calculated from a satellite image. Topographical variables were calculated from a Digital Elevation Model (DEM).

Category	Dataset	Data source	Spatial resolution	Time	Reference
Soil	Surficial deposit map	Field survey	0.5 m * 0.5 m	2016–2019	Kemppinen et al. (2018)
	Organic and mineral layer depths	Field survey	Point measurements every 50 m	2016–2018	Kemppinen et al. (2018)
Vegetation	NDVI	Sentinel-2	10 m * 10 m	2019	(ESA, 2021)
	Vegetation type	Field data and PlanetScope	3 m * 3 m	2018	New data
Topography	SAGA Wetness Index	DEM	2 m * 2 m	2016	NLS (2020)
	potential solar radiation	DEM	2 m * 2 m	2016	NLS (2020)

**TABLE 2** Growing season ( $\mu$ ) and monthly averages, minimum, maximum and range of Volumetric Water Content (VWC%) in 2018–2019. Values were calculated from the regime average timeseries.

Regime	Model	$\mu$	July	August	September	Min	Range	Max
Xeric	Measured	13.5	13.5	12.9	14.5	6.2	15.1	21.3
	JSBACH	16.7	16.1	16.4	17.9	13.0	6.4	19.4
	SpaFH <sub>y</sub>	15.4	15.1	14.7	16.6	11.4	9.9	21.3
Mesic	Measured	25.5	25.0	24.5	27.1	16.6	15.9	32.5
	JSBACH	32.9	32.1	32.1	35.4	28.5	11.9	40.4
	SpaFH <sub>y</sub>	27.3	27.0	27.1	27.9	24.8	10.0	34.8
Hydric	Measured	61.4	66.2	54.9	64.1	36.7	44.4	81.1
	JSBACH	59.7	59.9	58.6	61.1	52.9	12.2	65.1
	SpaFH <sub>y</sub>	47.7	47.9	46.9	49.2	43.1	9.7	52.8

595 **Figure legends**

**FIGURE 1** The study setting consists of 50 soil moisture loggers situated to measure the different soil moisture conditions of the landscape ((a), Digital Elevation Map provided by the National Land Survey of Finland 2020). The study area is situated in northwestern Finland ((b), Digital Elevation Map provided by European Union, Copernicus Land Monitoring Service 2020, European Environment Agency). Elevation (a), SAGA Wetness Index (SWI; (c)), and the sum of potential incoming solar radiation in July ( $\text{kW m}^{-2}$ , (d)) are topography-based variables. High SWI values indicate high wetness values (c). Soil surficial deposits (e) show the distribution of 1) glacial till, 2) peat deposits, 3) fluvial deposits, 4) boulders and 5) rock outcrops. The Normalized Difference Vegetation Index (NDVI; (f)) represents the variation in vegetation cover, with higher values indicating high amounts of photosynthetic plant tissue.

**FIGURE 2** Field-quantified (a) and modelled (b) temporal variation of Volumetric Water Content (VWC%) in xeric (20 loggers), mesic (18 loggers) and hydric (10 loggers) moisture regimes in July–September 2018 as well as daily precipitation sum (Finnish Meteorological Institute, 2020). Logger measurements (a) are shown separately and model results (b) through regime averages and the range within the regime. Variation within JSBACH results was very small and thus, nearly indistinguishable from regime averages.

**FIGURE 3** Correlation between modelled and measured time series of VWC% (Volumetric Water Content) in the soil moisture regimes during July–September 2018–2019. The x-axis shows temporal variation in the modelled average time series with respect to growing season averages (i.e. with the growing season mean deducted from the values to allow showing all regimes in one figure). The y-axis shows the same for measured regime average time series. The slope  $s$  and  $R^2$  of a linear regression are calculated for each regime and the dashed grey line is the 1 : 1 line.

**FIGURE 4** Spatial and temporal variation in soil moisture over the study area during July–September 2018–2019. The spatial variation is quantified as average VWC% (Volumetric Water Content) and temporal variation as range of variation. Field measurements are shown over the model results.

**FIGURE 5** GAM modelling the statistical relationship between environmental variables and average measured (a) and modelled (b) VWC% (Volumetric Water Content) as well as measured (c) and modelled (d) range of variation during July–September 2018–2019. The environmental variables used were thickness of the organic soil layer in the measurements (cm) and soil porosity in the mechanistic models, proportion of vegetation cover, SAGA Wetness Index (SWI), potential incoming solar radiation ( $\text{kW m}^{-2}$ ) and elevation (for measurements). JSBACH does not simulate water flow in the landscape based on topography so SWI was excluded.

**TABLE A1** Soil and vegetation parameters used for JSBACH and SpaFH<sub>y</sub>.

Parameters (unit)		SpaFH <sub>y</sub>	JSBACH	Note
<b>Soil</b>				
field capacity (%)	coarse	0.2	19.4	
	mixture	0.33	41.8	
	peat	0.6	88	
Saturated hydraulic conductivity (m/s)	coarse	0.0001	1.4E-05	
	mixture	1E-05	1.5E-06	
	peat	5E-05	2E-06	
soil porosity (%)	coarse	40	39.7	
	mixture	50	55.3	
	peat	90	88	
willing point (%)	coarse	8	8.7	
	mixture	14	14.2	
	peat	11	25.5	
Clapp & Hornberger parameter	coarse		4.7	
	mixture		4.5	
	peat		4	
pore size index	coarse		0.4	
	mixture		0.5	
	peat		0.7	
beta parameter	coarse	3.1		
	mixture	4		
	peat	6		
soil depth (m)	coarse	0.1	0.5	Assigned
	mixture	0.3	0.5	
	peat	0.6	0.6	
<b>Vegetation</b>				
maximum photosynthetic rate ( $\mu\text{mol}/(\text{m}^2 \text{ s})$ )	deciduous shrubs	11		Starr et al. (2008)
	evergreen shrubs	6		
	sedge	8		
stomatal parameter (kPa <sup>0.5</sup> )	deciduous shrubs	3.9		Lin et al. (2015)
	evergreen shrubs	1.5		
	sedge	1.8		Lin 2015
light response par. (W/m <sup>2</sup> )		50		Launiainen et al. (2019)
days for bud-burst		87		Pop et al. (2000)
duration of leaf development (d)		17		
day length for senescence start (h)		15		
duration of leaf senescence (d)		11		
background surface albedo	tundra		0.17	Hagemann (2002)
	mire		0.12	
surface roughness length due to vegetation (m)			0.03	

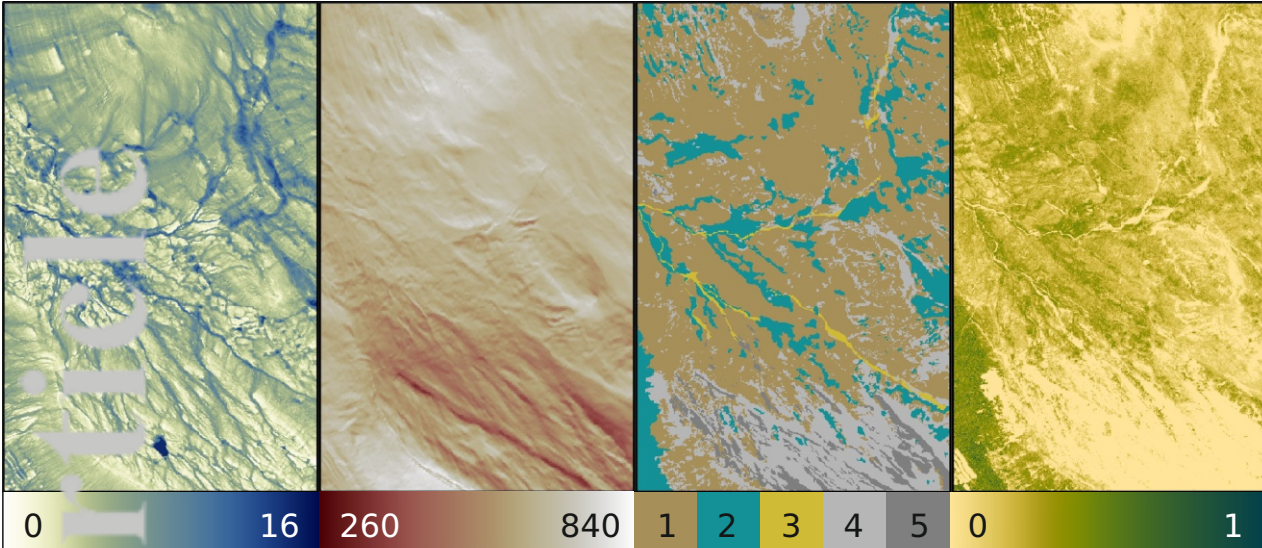
## Appendices

**FIGURE A1** Influence of field capacity on SpaFH<sub>y</sub>'s modelled VWC% (Volumetric Water Content) in peat soil areas.

**FIGURE A2** Relationship between organic layer and growing season average Volumetric Water Content (VWC). Pearson's correlation coefficient between the two variables was 0.79.

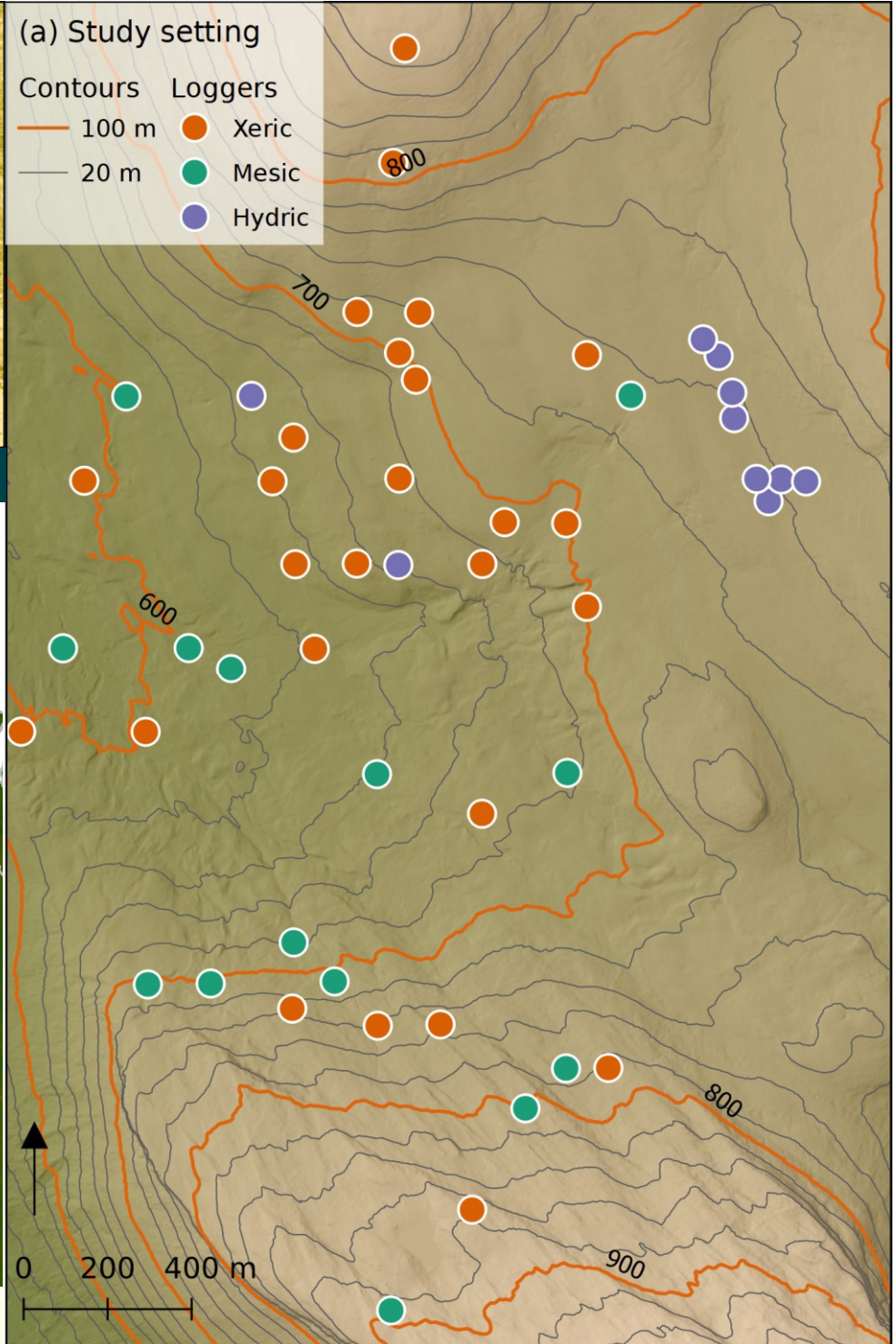
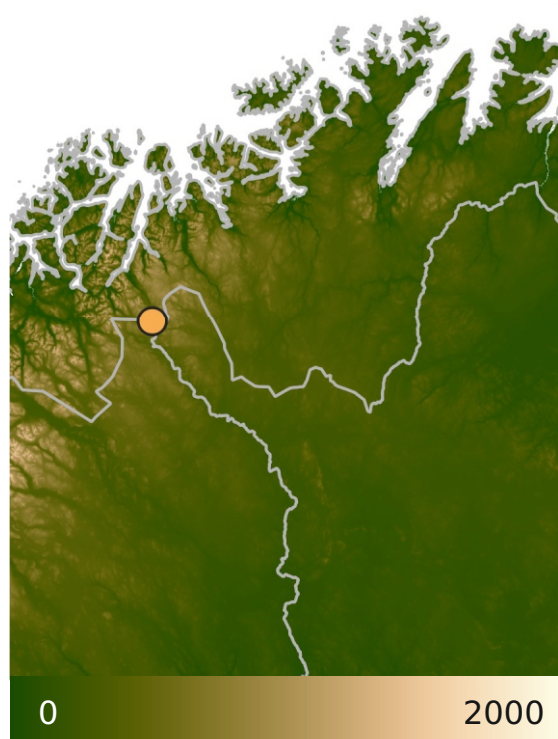
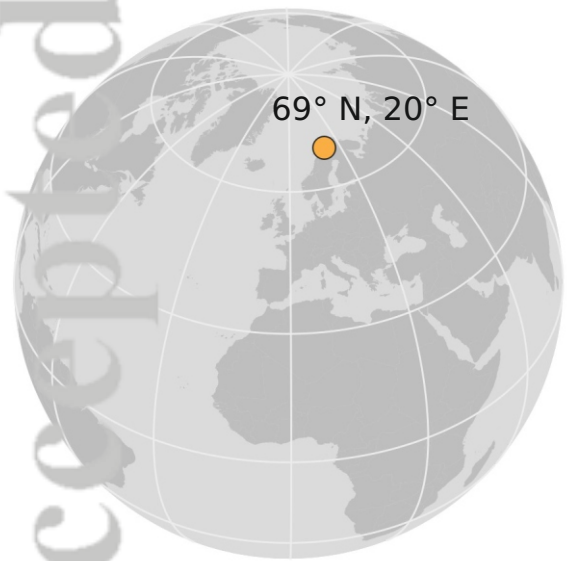
**FIGURE A3** (A) Field-quantified and (b) simulated temporal variation of soil moisture (Volumetric Water Content) in xeric, mesic and hydric moisture regimes in July–September 2019.

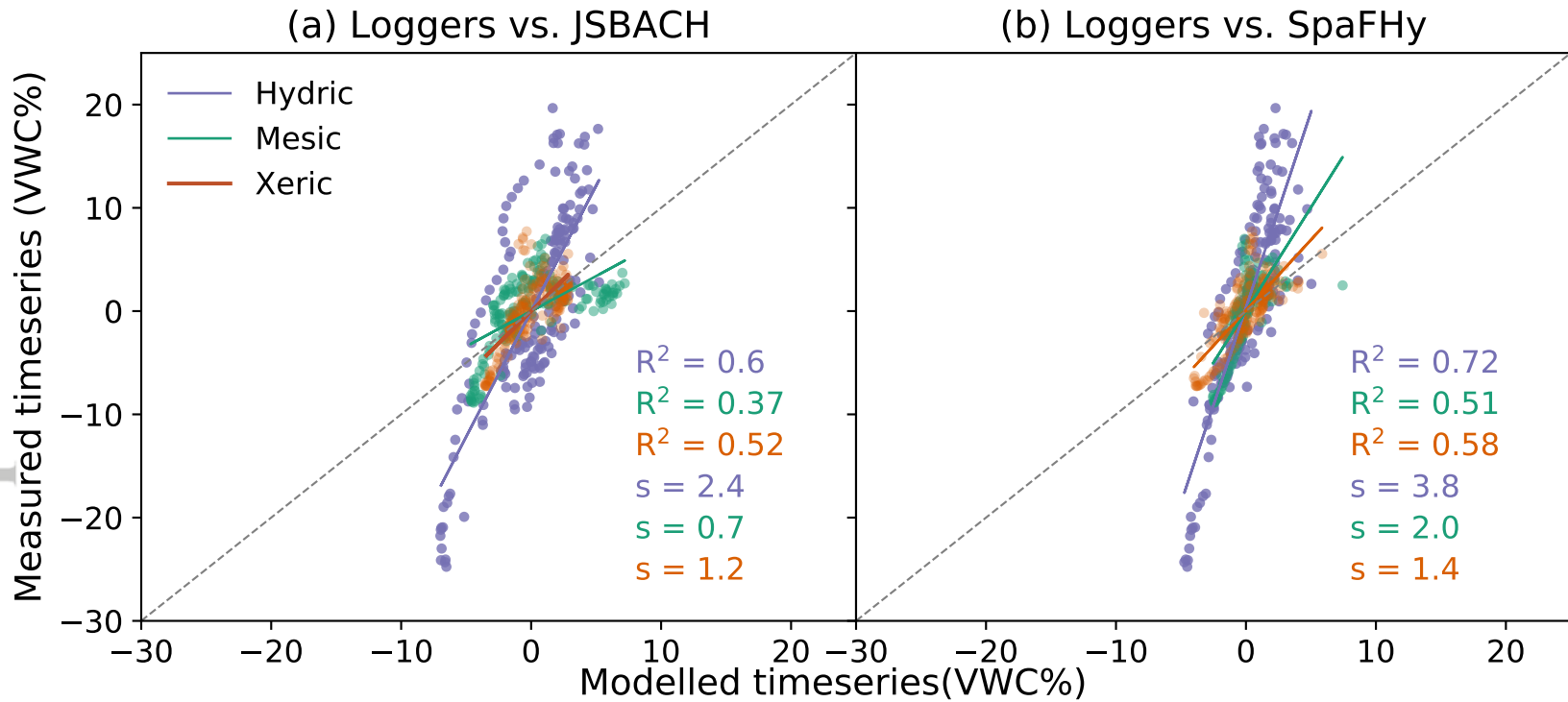
**FIGURE A4** Transpiration in JSBACH based on vegetation cover.



c) SW (d) Radiation (e) Surficial deposits (f) NDVI

(b) Location of the study site

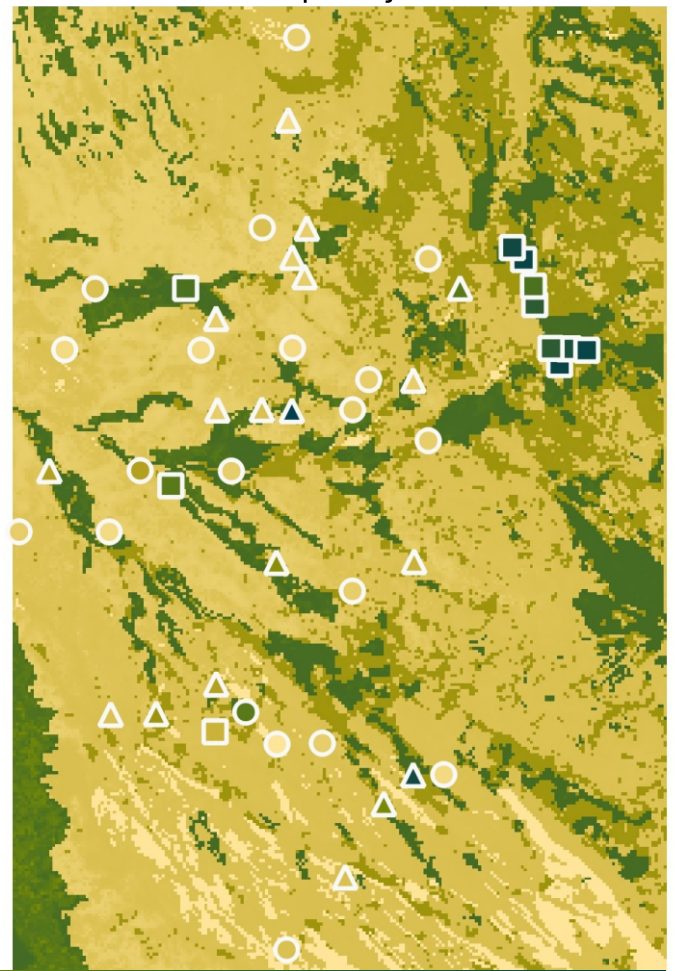
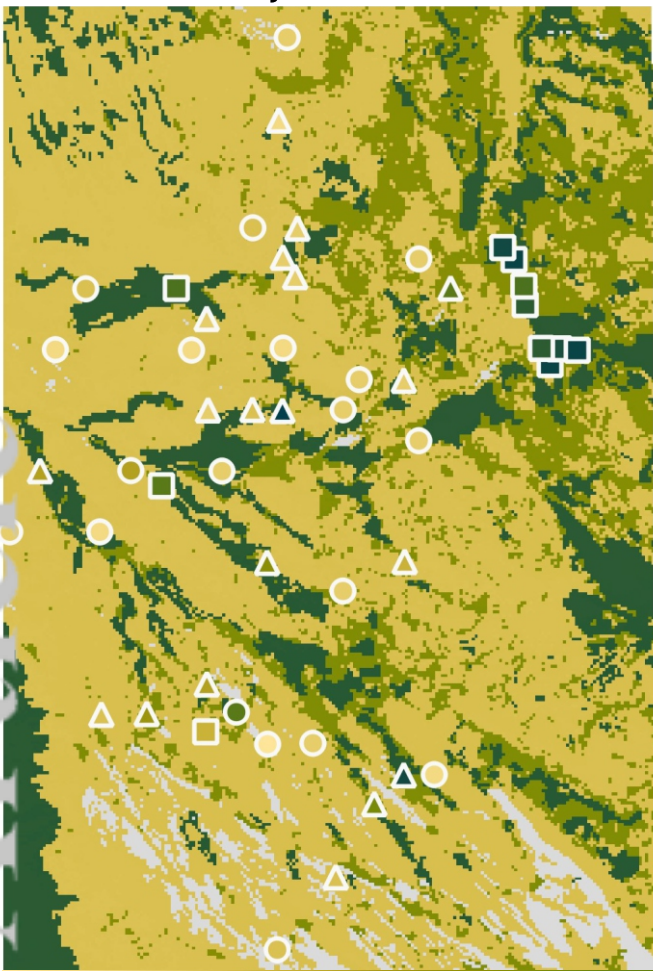




JSBACH

(a) Average VWC%

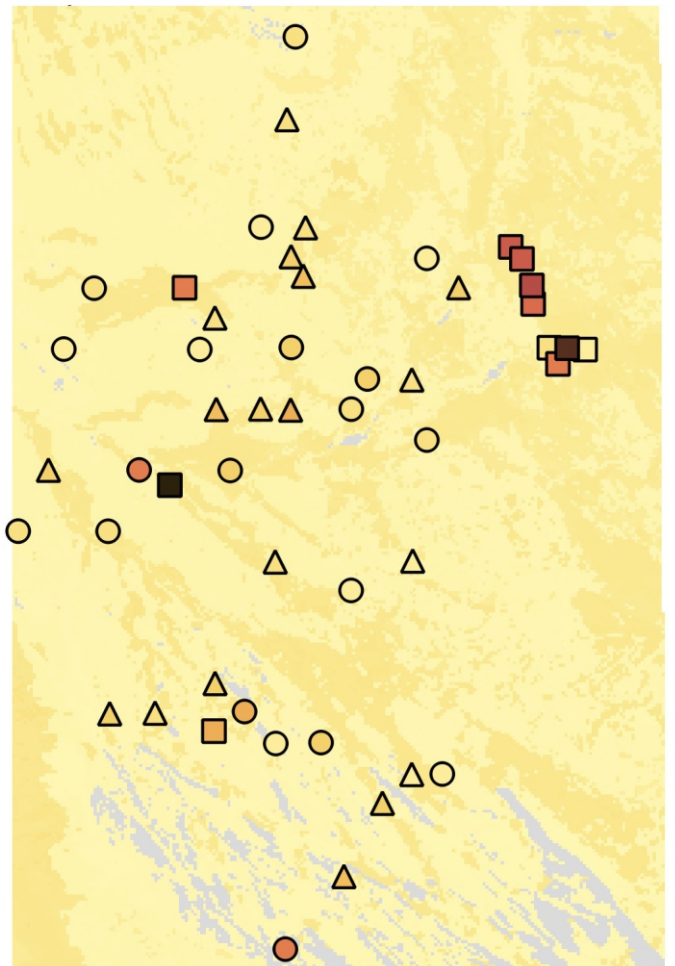
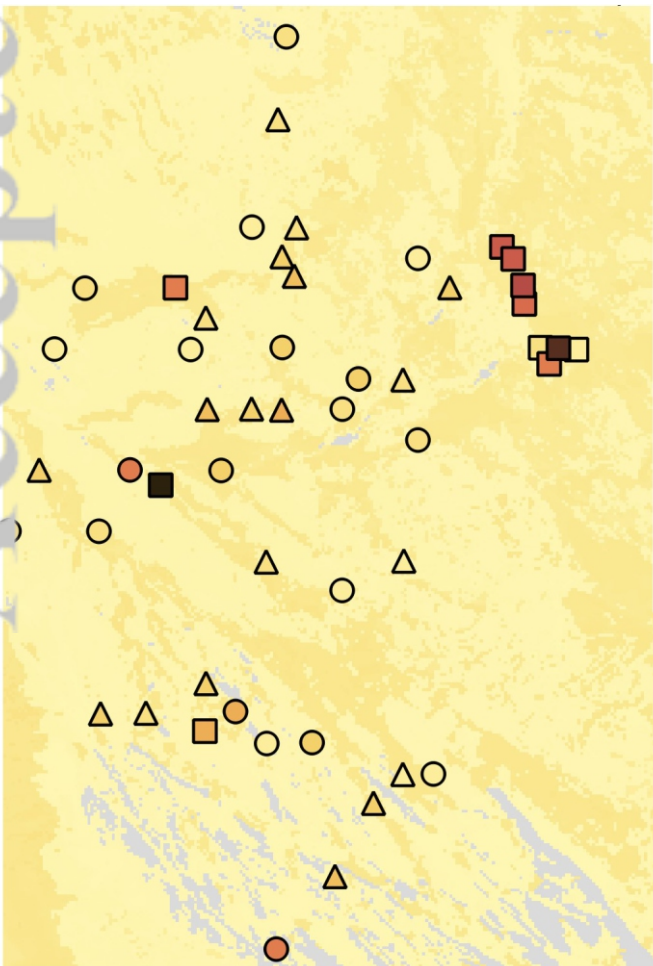
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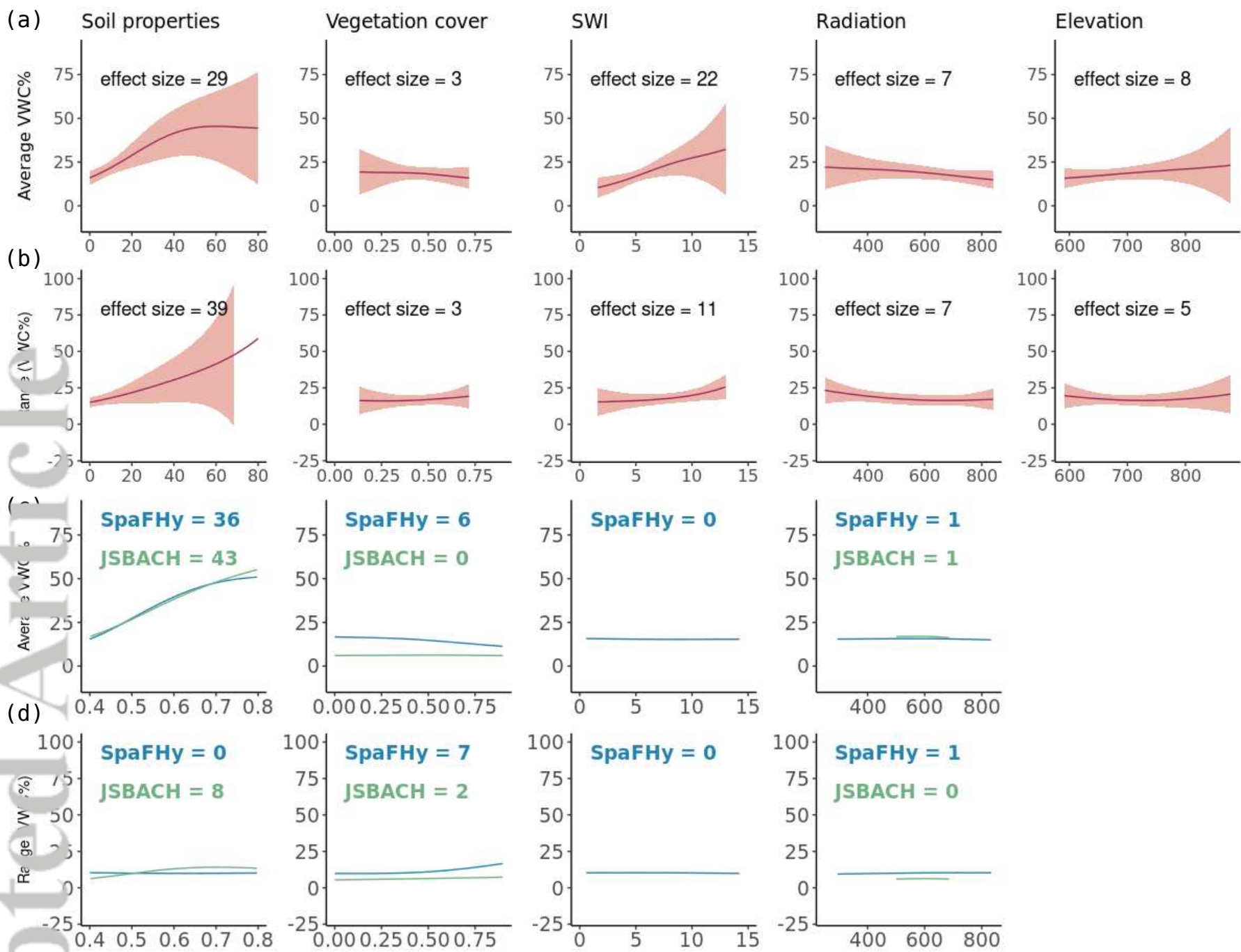
(b) Temporal variation



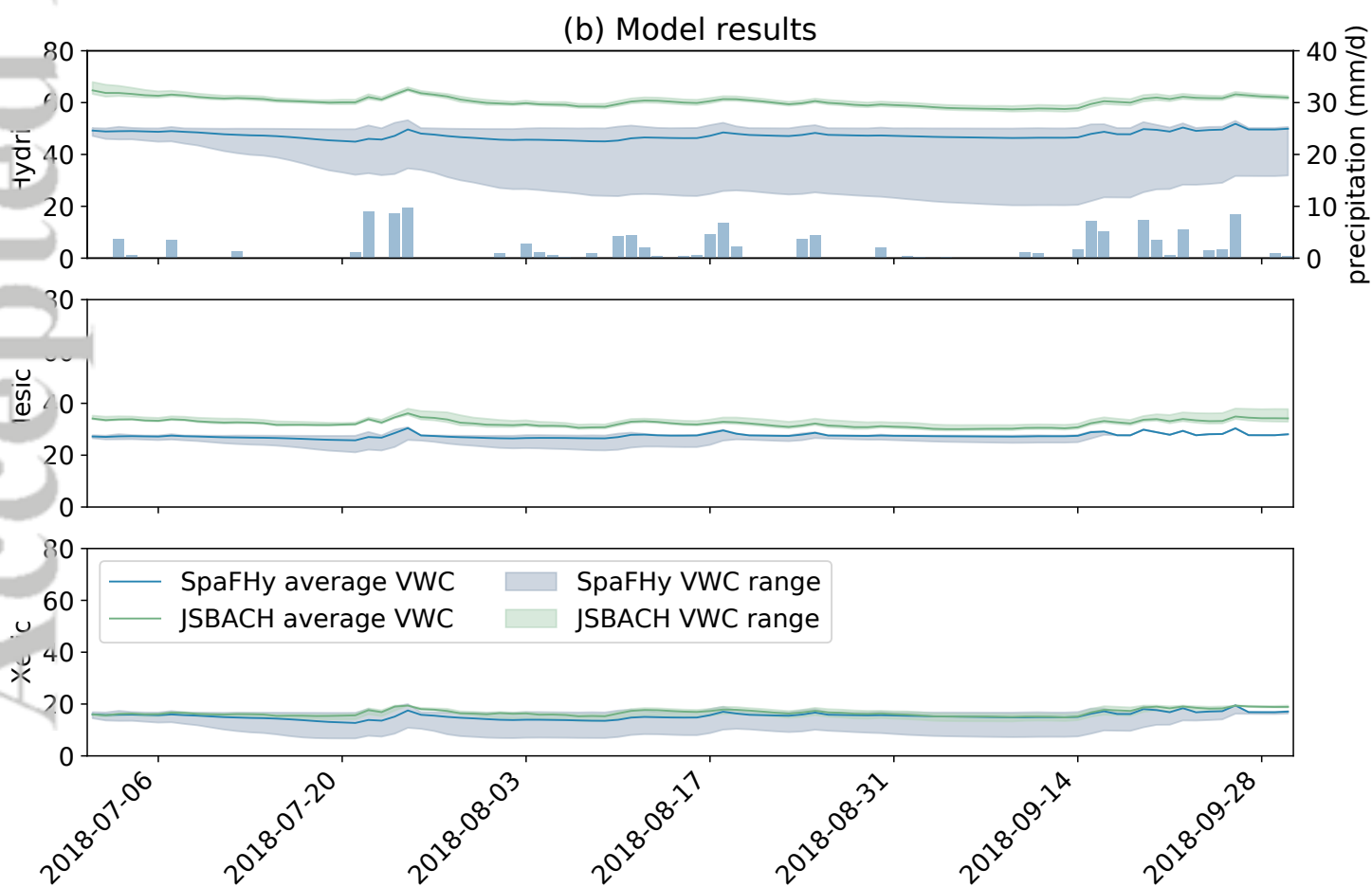
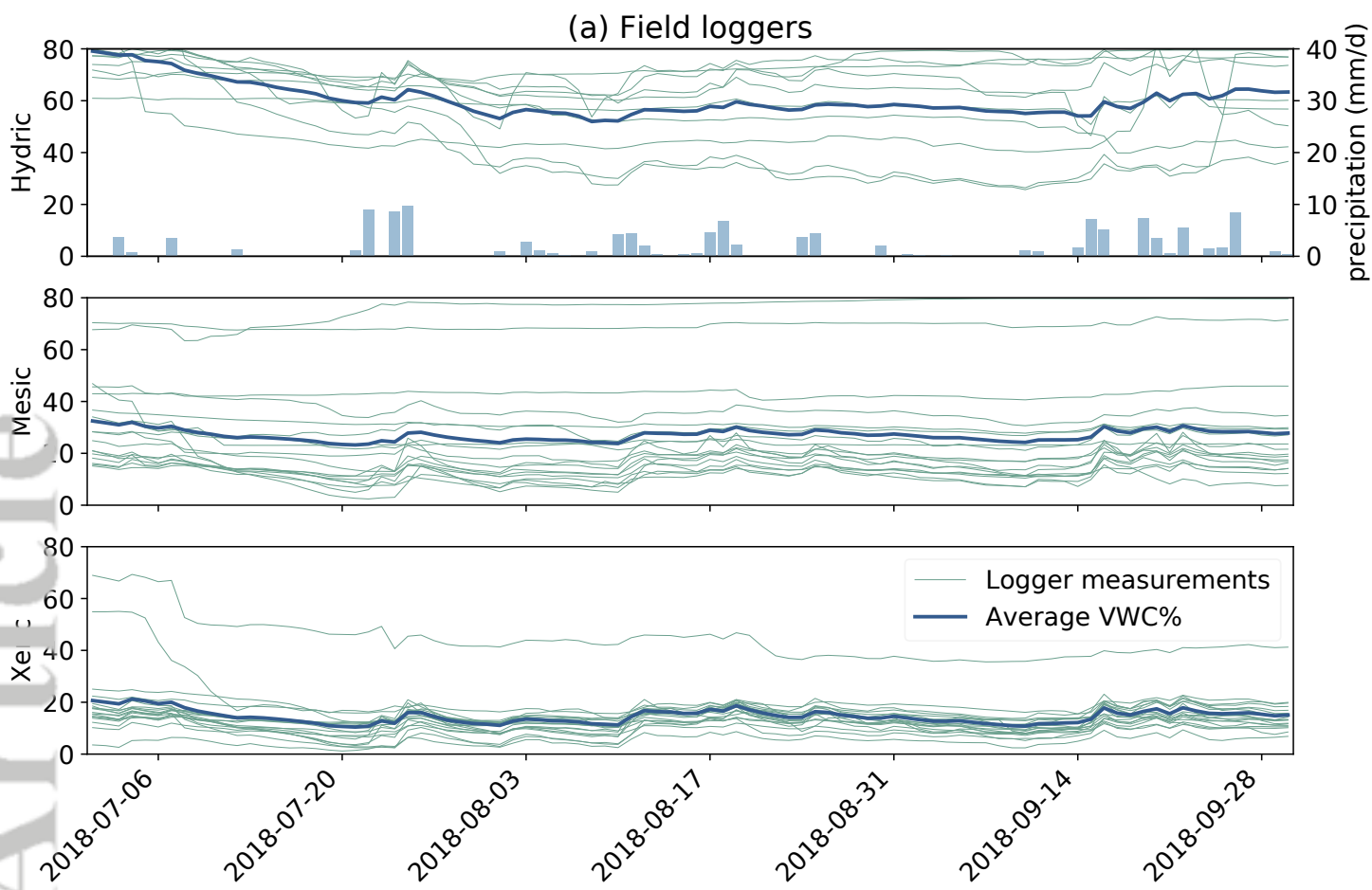
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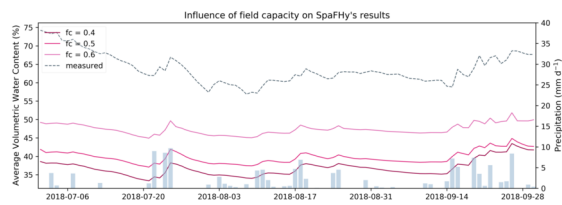
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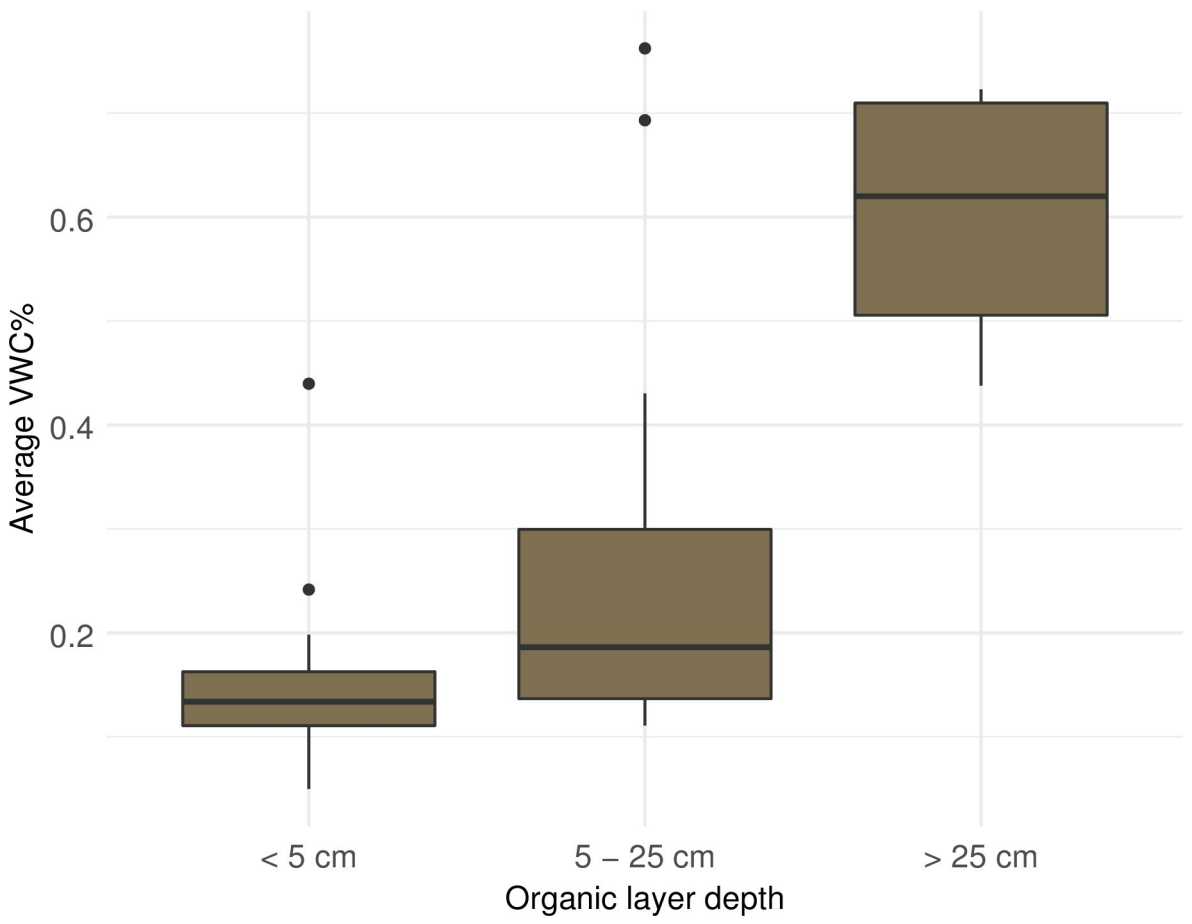


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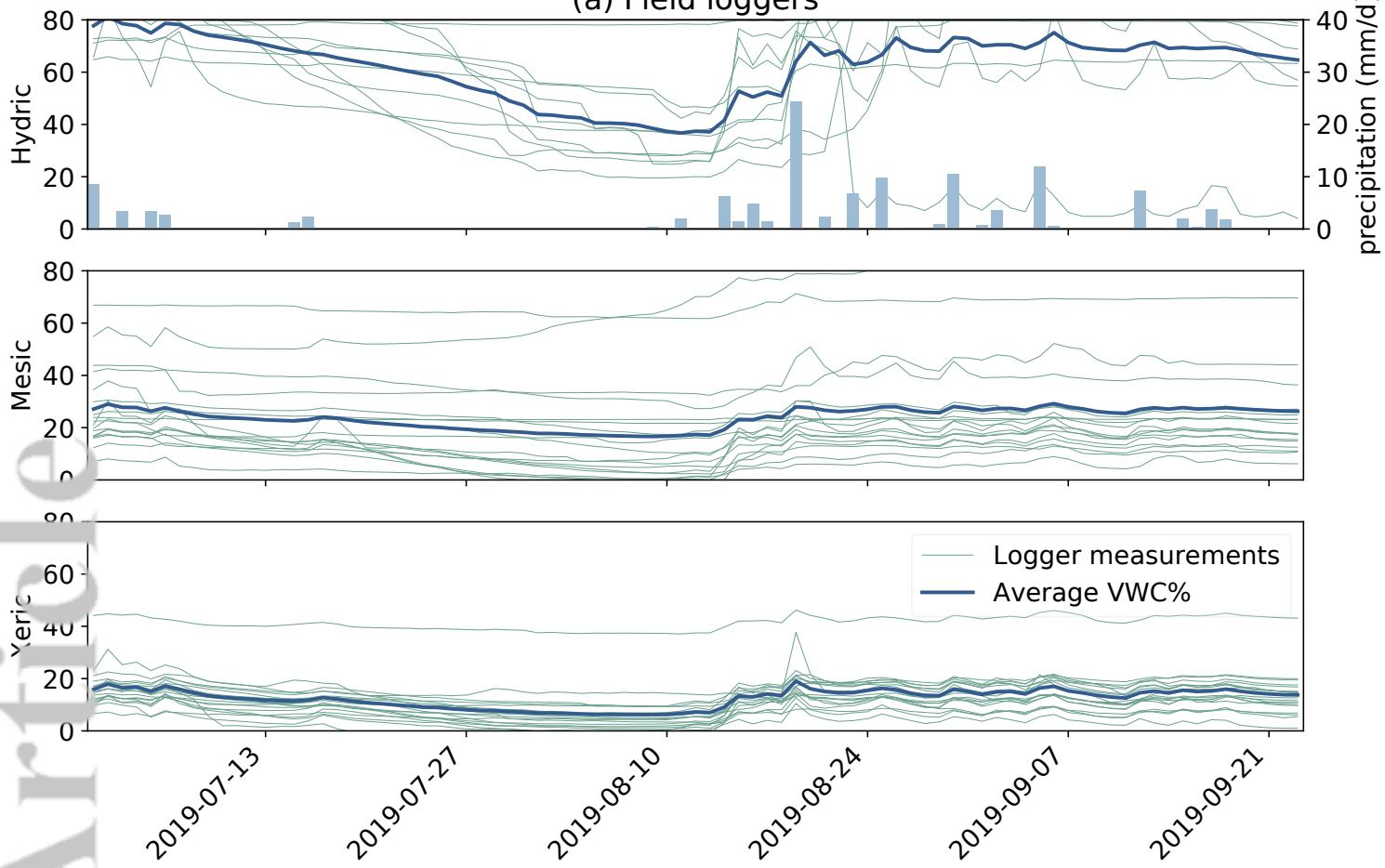


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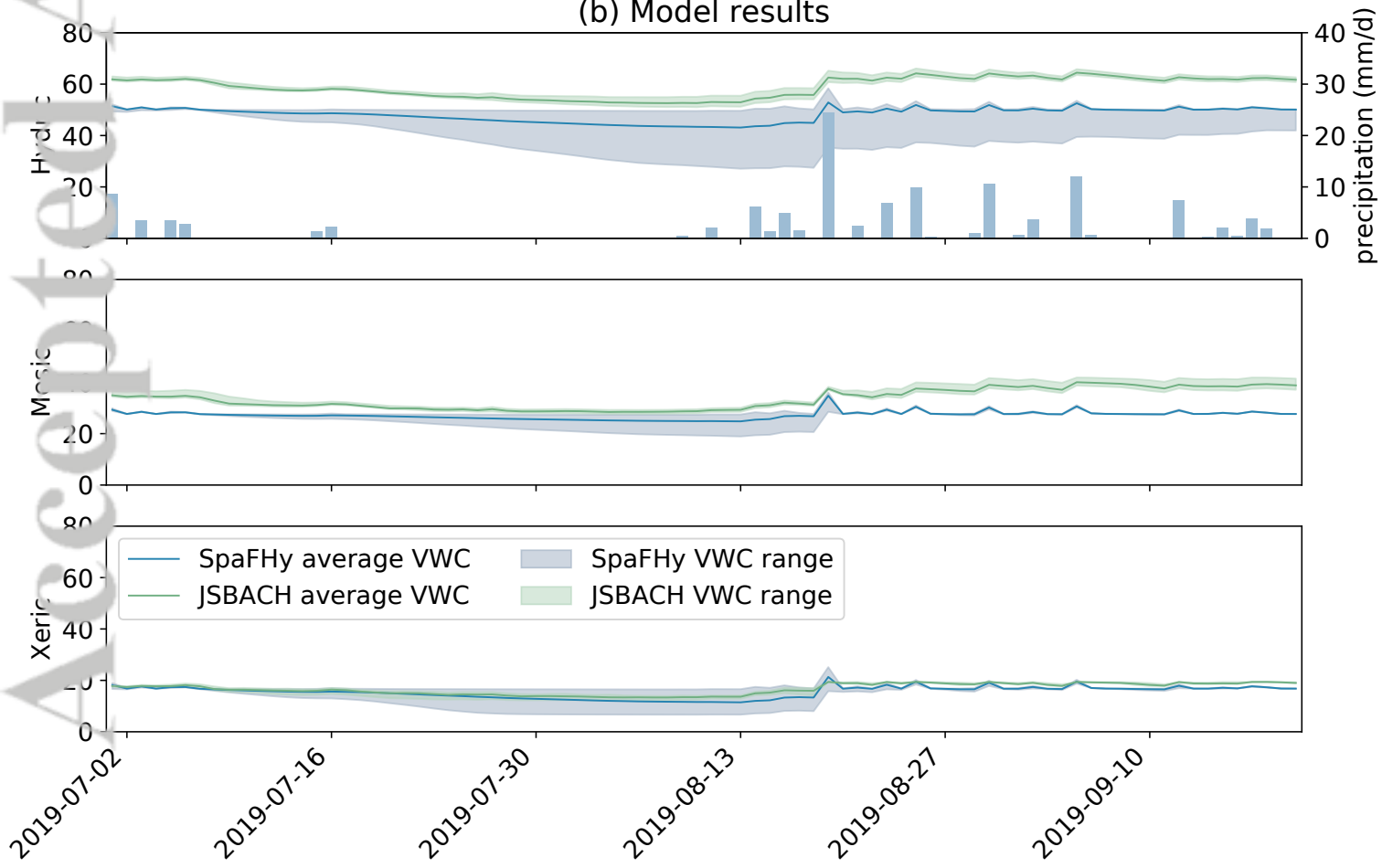


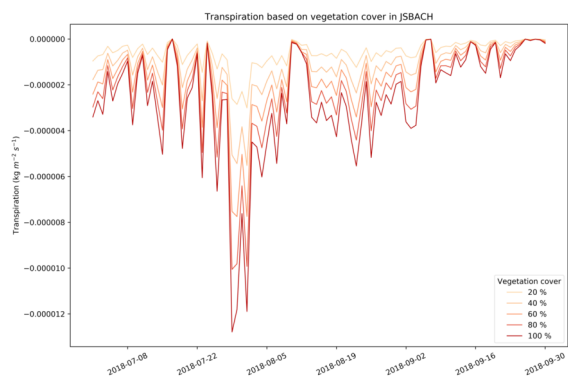
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(a) Field loggers



(b) Model results





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