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# Conversion of NH<sub>3</sub> and NH<sub>3</sub>-NO mixtures in a CO<sub>2</sub> atmosphere. A parametric study

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#### ARTICLE INFO ABSTRACT The present work addresses the oxidation of ammonia and ammonia-nitric oxide mixtures in a CO<sub>2</sub> atmosphere, characteristic of oxy-fuel processes and/or biogas combustion, from both experimental and kinetic modelling CO<sub>2</sub> atmosphere points of view. A parametric study of NH<sub>3</sub> and NH<sub>3</sub>/NO mixtures oxidation is carried out, evaluating the influence of the temperature (700–1500 K), stoichiometry (from pyrolysis, $\lambda = 0$ , to significantly oxidizing con-Oxy-fuel combustion ditions, $\lambda = 3.3$ ), gas residence time (low values, 195/T(K) s and high values, 3100/T(K) s) and NH<sub>3</sub>/NO ratio Nitrogen oxides (0.5-2.2), at atmospheric pressure under well-controlled laboratory conditions using two tubular flow reactor Kinetic modeling setups. Experimental results have been simulated with an updated literature reaction mechanism, which has been used to interpret the experimental observations.

#### 1. Introduction

Keywords:

Ammonia

Biogas

The problems derived from CO<sub>2</sub> emissions are very well known. Different strategies have been proposed to minimize these emissions, such as the use of biofuels or carbon free fuels. Among the latter ones, ammonia is believed to be a competitive candidate [1]. NH<sub>3</sub> exhibits a relatively high energy density and a low vapor pressure, which facilitates the transport and storage compared to the use of  $H_2$  [2]. In order to facilitate combustion and even the transition to a pure NH<sub>3</sub> economy, ammonia may also be mixed with other fuels, such as H<sub>2</sub> and/or CH<sub>4</sub> [3-5].

An interesting possibility is the use of NH<sub>3</sub>-biofuel mixtures, being biogas an appropriate biofuel. Biogas consists mainly of CH<sub>4</sub> and CO<sub>2</sub>, and the presence of CH<sub>4</sub> decreases the temperature of NH<sub>3</sub> ignition. Logically, the combustion of biogas also produces  $CO_2$  in the exhaust gas, although because biogas is considered as biofuel, neutral emissions of CO<sub>2</sub> can be assumed. Another additional solution to reduce CO<sub>2</sub> emission is oxy-fuel combustion. This process uses oxygen instead of air in order to significantly increase the  $\mathrm{CO}_2$  concentration of the exhaust gas, due to the absence of N<sub>2</sub>, and this is beneficial for carbon capture strategies. It is true that ammonia combustion will produce N2, although CO<sub>2</sub> concentration would still be high. Therefore, biogas combustion and/or the use of oxy-fuel combustion causes the process to occur in an overall CO<sub>2</sub> atmosphere.

The main drawback of using ammonia is related to the possible

fuel combustion scenarios where the only source of NO is the oxidation of ammonia. It is worthwhile to note that the joint presence of NH<sub>2</sub> and NO in a fuel-lean atmosphere is representative of conditions occurring during the selective non catalytic reduction of NO, process which is extensively used as a NO<sub>x</sub> reduction technology [e.g. 6]. In this sense, NO may be minimized in-situ during the ammonia combustion process. There are a number of studies in the literature addressing the conversion of ammonia [e.g. 7, 8] and the interaction of ammonia and nitric oxide [9–13], but most of them are carried out in air combustion conditions. Results indicate that the conversion of ammonia can produce both molecular nitrogen or nitric oxide, in different amounts depending on the specific operating conditions, and that the interaction of NH<sub>3</sub> and NO does occur, leading to very low NO emissions if the selected operating conditions are adequate [7-13]. Few studies have addressed the conversion of ammonia in the presence of significant amounts of CO2. In particular, the works of Mendiara and Glarborg [14,15] address the conversion of ammonia and the reduction of NO by methane-ammonia mixtures under oxy-fuel combustion conditions in a combined experimental and kinetic modelling study. They observed that using CH<sub>4</sub>, the reduction of NO attained was similar in N2 and CO2 under reducing conditions, but increased in a CO2 atmosphere for stoichiometric and lean conditions, compared to the conventional air environment. The conversion of ammonia and the formation of NO in a CO<sub>2</sub> atmosphere have also been recently studied by Ding et al. [16] and Li et al. [17] in a

formation of nitrogen oxides. This fact is of particular importance in oxy-

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jet-stirred reactor. The authors performed both experimental and simulation works and brought interesting conclusions related to the conversion regime of ammonia in a high  $CO_2$  concentration environment. Other works related to oxy-fuel combustion include the analysis of the effect of  $CO_2$  [18] or steam [19] for the conversion of ammonia to NO, or determinations of laminar speed flames [4], as well as the use of different combustion technologies as cyclone combustors [20].

In an oxy-fuel combustion scenario, the recirculation of the flue gas may cause an accumulation of pollutants, nitrogen oxides in particular or the unreacted ammonia if any. The high  $CO_2$  levels that may be present if biogas-ammonia mixtures are burned would not be expected to directly affect significantly the interaction between NH<sub>3</sub> and NO. However, high concentrations of  $CO_2$  are known to affect the O/H/OH radical pool, through the  $CO_2 + H \Rightarrow CO + OH$  reaction [21], and thus may affect the combustion chemistry of NH<sub>3</sub> and NH<sub>3</sub>-NO mixtures.

In this context, the present work aims to study the oxidation of  $NH_3$ and the interaction of  $NH_3$  and NO in a  $CO_2$  atmosphere, characteristic of biogas or oxy-fuel combustion processes. The study is carried out from both experimental and kinetic modeling points of view using tubular flow reactor experimental installations and a literature detailed kinetic mechanism.

# 2. Experimental methodology

Experiments have been carried out in two different experimental installations (A and B), both consisting basically of a gas feeding system, a reaction system, and a gas analysis system. Gas feeding and analysis is common to both reaction systems. These systems allow us to perform experiments for very different gas residence times and temperature ranges.

Reaction system A includes a quartz flow reactor with a reaction zone of 8.7 mm internal diameter and 200 mm in length, which is placed in a three-zone electrically heated oven, which ensures an isothermal temperature throughout the reaction zone within  $\pm$  10 K. Detailed description of this system can be found in Alzueta et al. [22]. The temperature in the reaction zone was measured with a type-K fine-wire thermocouple. Gases (N50) are led to the reactor from cylinders through mass flow controllers by four separate injectors, a main stream and three additional side streams. The main stream containing O2 diluted in CO2 is preheated in the reactor and mixed with the other feed gases (NH<sub>3</sub>, NO, and Ar) at the inlet of the reaction zone. An additional experiment in a N2 atmosphere was also perfomed in order to compare the results wiht CO2 and N2 as bath gases. A total gas flow rate of 1000 mL (STP)/min is kept constant during the experiments, leading to a gas residence time, in the reaction zone, as a function of the reaction temperature,  $t_r(s) = 195/$ T [K]. The mixture is highly diluted, low heat is released during the reaction and therefore no influence of the reaction on the temperatures can be assumed. At the outlet of the reaction zone, the product gas is cooled by means of external refrigeration with air and conducted to the analysis system.

Reaction system B includes a non-porous alumina tube reactor with 40 mm internal diameter and 800 mm in length. The reactor is placed in an electrically heated oven which allows temperatures up to 1800 K. The longitudinal temperature profile in the reactor was measured with a type-S platinum ceramic-covered thermocouple placed inside the reactor and results in an almost isothermal temperature (within  $\pm$  20 K) which is attained throughout the main reaction zone of 150 mm approximately, with steep temperature gradients toward both the inlet and outlet of the reactor. A total gas flow rate of 1000 mL (STP)/min is kept constant during the experiments, leading to a gas residence time as a function of the reaction system and experimental procedure can be found for example in Giménez-López et al. [23].

measure CO, CO<sub>2</sub>, N<sub>2</sub> and H<sub>2</sub>. NO concentration is measured by means of a continuous IR analyzer. The gas stream is also analyzed by a Fourier Transform Infrared (FTIR) spectrometer, where NH<sub>3</sub>, N<sub>2</sub>O and NO<sub>2</sub> are measured. The uncertainty of the measurements is estimated as  $\pm$  5%, except for the values of the compounds detected with the FTIR spectrometer, which is  $\pm$  10%. In all the experiments, very low, negligible, concentrations of N<sub>2</sub>O and NO<sub>2</sub> are detected, and consequently are not shown.

Table 1 shows the conditions of the different experiments.

With the exception of experiment 1, that has been performed in a N<sub>2</sub> atmosphere, representative of conventional combustion conditions, the rest of experiments (2 to 17) have been made in a CO<sub>2</sub> atmosphere, representative of oxy-fuel combustion conditions. The CO<sub>2</sub> concentration values around 75 % have been taken as representative of the oxyfuel combustion atmosphere according to the work of Giménez-López et al. [21]. Most of experiments contain water, for two reasons. On the one hand, water is a product of the combustion reaction of ammonia and will be present as a reaction product. On the other hand, the presence of water assures a high concentration of the O/H radical pool and thus will minimize, if any, potential surface effects, by minimizing the effect of radical recombination in the reactor walls. In the experiments in the presence of CO<sub>2</sub>, Ar is used to balance the flow rate up to 1000 mL/min (STP). Comparison of experiment 1, with N<sub>2</sub> as bath gas, and experiment 4 with CO<sub>2</sub> as bath gas allows us to compare the results obtained under the two reaction atmospheres for similar operating conditions.

While most of the experiments have been made in installation A with a gas residence time of 195/T(K) seconds, sets 1 to 12 in Table 1, certain experiments, sets 13 to 17, in the presence of NO have been performed with a higher residence time of 3100/T(K) seconds, in installation B.

The stoichiometry  $(\lambda)$  is defined as the quotient between the oxygen/ fuel relation available for reaction and the stoichiometric oxygen/fuel relation, in the following reaction:

$$NH_3 + 3/4 O_2 \rightarrow 1/2 N_2 + 3/2 H_2O$$
 (r1)

The influence of  $\lambda$  (0.33, 1.17, 1.67 and 3.33) on NH<sub>3</sub> oxidation in the CO<sub>2</sub> atmosphere has been studied in installation A for a given residence time, experiments 3 to 6 in Table 1. Also, influence of  $\lambda$  has been faced by means of a pyrolysis experiment, i.e.  $\lambda = 0$ , experiment 2 in Table 1.

The impact of NO presence on NH<sub>3</sub> conversion and the reduction of NO that can be attained have been considered by adding approximately 1000 ppm NO (experiments 7 to 10). Also, the influence of NO concentration has been studied for a given  $\lambda$  value,  $\lambda = 1.67$  (experiments 9, 11 and 12 for the residence time of 195/T(K) s in installation A, and 14, 16, and 17 for the residence time of 3100/T(K) s in installation B), resulting in different NH<sub>3</sub>/NO ratios, between approximately 0.5 and 2.

#### 3. Kinetic modelling

Model calculations have been performed using the PFR model of the Chemkin Pro suite [24]. The mechanism used for simulations is largely based on the model of Glarborg et al. [25], with minor updates and extended to include CH<sub>3</sub>CN conversion [26]. This reaction mechanism includes submechanisms for C1-C2 hydrocarbons [27–29], amines [30] cyanides [31] and hydrocarbon/nitrogen interactions [14,15,32], and the interaction of NH<sub>3</sub> and NH<sub>2</sub> to produce N<sub>2</sub>H<sub>3</sub> and H<sub>2</sub>, which was found to be important to describe NH<sub>3</sub> oxidation in an air atmosphere [8], as described in the work of Dove and Nip [33].

Under oxy-fuel combustion conditions, the concentration levels of  $CO_2$  are comparatively high. Thermal decomposition of  $CO_2$  is highly endothermic and is relevant only at high temperatures,

$$CO_2 + M \rightleftharpoons CO + O + M$$
 (r2)

Reactions of  $CO_2$  with the radical pool can occur at comparatively lower temperatures. The main reaction of  $CO_2$  with radicals is its interaction with H radicals, which is important even at moderate

Table 1Experimental conditions.

Set	NH <sub>3</sub> (ppm)	NO (ppm)	O <sub>2</sub> (ppm)	H <sub>2</sub> O (%)	Ar (%)	CO <sub>2</sub> (%)	N <sub>2</sub> (%)	λ	t <sub>r</sub> (s)
1	1008	0	882	0.6	0	0	99.21	1.17	195/T(K)
2	1040	0	0	0	24.9	75	0	0	195/T(K)
3	923	0	231	0.6	25	74.28	0	0.33	195/T(K)
4	906	0	793	0.6	25	74.22	0	1.17	195/T(K)
5	1050	0	1312	0.6	25	74.17	0	1.67	195/T(K)
6	1159	0	2898	0.6	25	74.03	0	3.33	195/T(K)
7	1125	1005	281	0.6	25	74.16	0	0.33	195/T(K)
8	1177	1010	1030	0.6	25	74.10	0	1.17	195/T(K)
9	1143	992	1429	0.6	25	74.06	0	1.67	195/T(K)
10	1122	995	2805	0.6	25	73.94	0	3.33	195/T(K)
11	1131	505	1414	0.6	25	74.11	0	1.67	195/T(K))
12	1090	1975	1363	0.6	25	73.97	0	1.67	195/T(K)
13	1008	1029	875	0.6	25	74.11	0	1.17	3100/T(K)
14	1022	1029	1277	0.6	25	74.07	0	1.67	3100/T(K)
15	1033	1015	2582	0.6	25	73.95	0	3.33	3100/T(K)
16	1048	502	1312	0.6	25	74.08	0	1.67	3100/T(K)
17	1010	2005	1262	0.6	25	74.03	0	1.67	3100/T(K)

temperatures, and is responsible for the partial equilibrium between CO and  $CO_2$  under combustion conditions,

$$CO_2 + H \rightleftharpoons CO + OH$$
 (r3)

This reaction has been extensively studied and it is well-known in a large temperature and pressure interval (see for instance the revision of Rasmussen et al. [34]). For this reaction, the value recommended by Baulch and Drysdale [35] has been used, which roughly agrees with the later determinations for the kinetic constant of this reaction.

Reactions of  $CO_2$  with other radicals, such as O and OH, reactions 4 and 5, are too slow to be relevant under the studied conditions:

$$CO_2 + O \rightleftharpoons CO + O_2$$
 (r4)

$$CO_2 + OH \rightleftharpoons CO + HO_2$$
 (r5)

 $\mathrm{CO}_2$  is able to interact with some reactive radicals of the N/O/H system. The CO<sub>2</sub> + N reaction has been reported to contribute to NO formation under fuel-rich conditions, in CH<sub>4</sub> flames with various O<sub>2</sub> concentrations [36]. The first studies of the kinetic parameters for the  $CO_2 + N \Rightarrow CO + NO$  (r6) reaction were those of Avramenko and Krasnen'kov [37] who obtained an experimental determination in a vacuum system, assuming that the reaction was relatively fast. Later, Herron and Huie [38] performed a mass spectrometric study of that reaction in the 298-523 K temperature range, and later Fernández et al. [39] studied that reaction in a high temperature photochemical reactor. The two latter works indicated that the reaction was slow and little important under combustion conditions. This was confirmed by the theoretical study of Manaa and Chabalowsky [40] who obtained a high energy barrier for this reaction. The present work uses the rate constant proposed by Glarborg et al. [6] which agrees well with the upper limit proposed by Fernández et al. [39] and with the shock wave measurements of Lindackers et al. [41].

Röhrig and Wagner [42] determined the rate constants of the  $CO_2$  + NH interaction (reaction r7) from incident shock wave experiments, finding that the reaction was significantly fast and that the only products thermodynamically stable for that reaction (r7) were HNO + CO in the studied temperature range of 1200–1900 K.

$$CO_2 + NH \Rightarrow HNO + CO$$
 (r7)

Later, Fontijn et al. [43] disregarded those species as reaction products, based on their theoretical studies which indicate much higher activation energies compared to the determinations of Röhrig and Wagner [42]. Fontijn et al. [43] proposed the formation of reaction adducts able to react, but they were not finally able to specify the products of that reaction. Similar conclusions were reached by Mackie and Backsay [44], who concluded that the formation of HNO + CO as products was only significant above 1500 K. Since uncertainty in relation to this reaction is high, we have chosen HNO + CO as the only product channel for the  $CO_2$  + NH reaction, using the kinetic parameters of Mackie and Backsay [44] for the global reaction, which are significantly lower than those proposed by Röhrig and Wagner [42]. While this reaction has not been to be particularly relevant under the conditions of the present work, it exhibits a high uncertainty and would deserve a proper determination.

An experimental shock wave study by Wooldridge et al. [45] allowed to detect the formation of  $CO_2 + NH_2$  as a minority product channel of the HNCO + OH interaction. The authors established a maximum contribution of 10 % for that reaction channel in the 1250–1860 K temperature range. This low value for the reaction constant is consistent with other studies [46,47], which suggest NCO and H<sub>2</sub>O as majority products, reactions:

$HNCO + OH \Rightarrow NCO + H_2O$	(r8)
	(

$$HNCO + OH \rightleftharpoons CO_2 + NH_2 \tag{r9}$$

In the present work, the channel producing  $CO_2$  is kept, despite its expected low contribution in the high  $CO_2$  atmosphere.

The reaction of  $CO_2$  with NO may affect the NO/NO<sub>2</sub> equilibrium. The study for the direct interaction by Clark et al. [48] establishes an upper limit for the constant of this reaction of  $3.01 \times 10^9$  cm<sup>3</sup>/mol s. On the contrary, the reverse reaction has been more studied. In the present work, we use the theoretical determination by Kroupnov and Pogosbekian [49], for this reverse step, reaction r10, which agrees roughly with other literature determinations [50–52].

$$\rm CO + \rm NO_2 \rightleftharpoons \rm CO_2 + \rm NO$$
 (r10)

In the high  $CO_2$  concentration environment, the interaction between  $NH_3$  and  $CO_2$  may potentially be important, and is included in the mechanism, reaction:

$$HNCO + H_2O \rightleftharpoons NH_3 + CO_2 \tag{r11}$$

Cheng et al. [53] used density functional theory (DFT) calculations to investigate the synthesis mechanism of cyanuric acid from  $NH_3 + CO_2$ and provided the mechanism for reaction and energy barriers involved, indicating that those are relatively high. Later, Glarborg et al. [25] included the reaction of HNCO + H<sub>2</sub>O as a possible reaction for HNCO conversion during the oxidation of this compound, indicating that the interaction of HNCO with water would result in  $NH_3 + CO_2$ . Under the conditions of the present work, and because of the high concentrations of  $CO_2$ , this reaction is seen play a role, as is later discussed.

The full mechanism is included in Chemkin format in the supplementary material.

# 4. Results and discussion

The influence of  $N_2$  and  $CO_2$  atmospheres can be seen in Fig. 1, where comparison of experimental results carried out in similar conditions with and without  $CO_2$  and roughly stoichiometric conditions is seen. The conversion of ammonia in the air atmosphere seems to start at slightly lower temperatures compared to the  $CO_2$  environment, and the presence of high amounts of  $CO_2$  seems to slow down ammonia conversion even more at the highest temperatures studied in the present work. The model is able to catch this feature, even though the specific values of concentrations are not well matched. Calculations indicate that the reason for the different behavior obtained in the presence of  $CO_2$  is attributed to the main reactions consuming  $NH_3$  in each case.

In the presence of  $CO_2$ ,  $NH_3$  reacts with OH and  $NH_2$  radicals and with  $CO_2$ , through reactions r11 to r13:

$$NH_3 + OH \rightleftharpoons NH_2 + H_2O \tag{r12}$$

$$NH_3 + NH_2 \rightleftharpoons N_2H_3 + H_2 \tag{r13}$$

$$NH_3 + CO_2 \Rightarrow HNCO + H_2O$$
 (-r11)

In the CO<sub>2</sub> environment, and at the temperature of 1350 K, reaction r12 implies a 48 % of NH<sub>3</sub> conversion, followed by r13 (25 %) and r11 proceeding in the reverse sense, i.e. -r11, (23 %). Similar reactions dominate the high temperature conversion of NH<sub>3</sub> at high temperatures in the same atmosphere, even though the contribution of the reactions is different. It is thus seen that a direct interaction of ammonia and carbon dioxide occurs under the studied experimental conditions.

In the air  $(O_2/N_2)$  environment, NH<sub>3</sub> reacts with different components of the radical pool (OH, NH<sub>2</sub>, H and O radicals) and logically the interaction with CO<sub>2</sub> can not occur. In this case, NH<sub>3</sub> is converted through reactions r12 and r13, and r14 and r15 at the onset of conversion, i.e. ca 1300 K:

$$NH_3 + H \rightleftharpoons NH_2 + H_2 \tag{r14}$$

$$NH_3 + O \Rightarrow NH_2 + OH$$
 (r15)

but, as the temperature increases, only interaction with  $NH_3$  with the O/H radical pool happens, with negligible reaction with  $NH_2$  radicals (r13).

The influence of stoichiometry on ammonia oxidation in a CO<sub>2</sub> atmosphere, representative of oxy-fuel combustion conditions, is shown in Fig. 2, and the formation of NO and N<sub>2</sub> are shown in Figs. 3 and 4 respectively. Conditions represented range from pyrolysis ( $\lambda = 0$ ) to significantly fuel lean ones ( $\lambda = 3.33$ ), sets 2–6 in Table 1. In agreement



Fig. 1. NH<sub>3</sub> and NO concentration profiles as a function of temperature, for a stoichiometry of  $\lambda = 1.17$  in either a N<sub>2</sub>/air (Set 1 in Table 1) or an oxy/CO<sub>2</sub> atmosphere (Set 4 in Table 1).



Fig. 2.  $NH_3$  concentration profiles as a function of temperature, for different stoichiometries in an oxy/CO<sub>2</sub> atmosphere (Sets 2–6 in Table 1).



Fig. 3. NO concentration profiles as a function of temperature, for different stoichiometries in an  $oxy/CO_2$  atmosphere (Sets 2–6 in Table 1).



Fig. 4.  $N_2$  concentration profiles as a function of temperature, for different stoichiometries in an oxy/CO<sub>2</sub> atmosphere (Sets 2–6 in Table 1).

with the known effect of a high availability of oxygen promoting conversion of ammonia [e.g. 8], the higher radical pool reserve at lean conditions is responsible for a higher conversion of NH<sub>3</sub> at a given temperature. However, differences in the conversion profiles of NH<sub>3</sub>, as a function of temperature for the different stoichiometries studied, are not very important, with maybe the exception of  $\lambda = 3.33$ , where the availability of O/H free radicals is considerably higher. The higher conversion of NH<sub>3</sub> occurring with the increase in O<sub>2</sub> concentration is accompanied by a higher formation of NO, despite the low values of this compound attained for all the temperature range considered, except for the most oxidizing conditions studied, where the higher NH<sub>3</sub> consumption facilitates a higher formation of NO.

The experimental nitrogen atoms balance, calculated as the sum of nitrogen in  $NH_3$ , NO and  $N_2$  closes very well until a temperature of approximately 1300 K. Above this temperature, deviations in the balance occur, which seems to indicate that other species than those mentioned appear in the reaction system. Model calculations above this temperature indicate the formation of a high number of different nitrogen species in small concentrations, which may be the reason of the unbalanced nitrogen.

While the model used in simulations reproduces fairly well the  $NH_3$  concentration profiles, the formation of NO is underestimated. However, as it has been mentioned above, the amount of NO formed is low, below 40 ppm under the conditions studied, except for the leanest case in which NO reaches approximately 80 ppm. On the contrary, the formation of  $N_2$  is overestimated by the model, even though again the measured experimental concentrations are low, with maximum values measured of around 100 ppm. Anyway, while the main trends are fairly well reproduced by the model, it seems that it is not fully able to predict properly the conversion of  $NH_3$  into  $N_2$  or NO. At present, we do not know the specific reason for such issue, but it may be related to the conversion of the intermediate species in the presence of the high concentrations of  $CO_2$ .

Figs. 5 to 7 show the impact of the presence of approximately 1000 ppm NO on the conversion of around 1000 ppm NH<sub>3</sub> as a function of temperature for different stoichiometries. Results indicate that the presence of NO has different effects depending on the stoichiometry. Under fuel rich conditions, NO acts to promote NH<sub>3</sub> oxidation. However, for stoichiometric and fuel-lean conditions, the presence of NO inhibits the conversion of NH<sub>3</sub> at the temperatures of its initiation. This happens because of a complex mechanism involving the radical pool concentration and speciation and the competition of NO for NH<sub>2</sub> radicals and other species, in the experiments in the presence of NO.

As mentioned before, in the absence of NO, conversion of NH<sub>3</sub>



Fig. 5. Normalized NH<sub>3</sub> concentration (NH<sub>3</sub>/NH<sub>3</sub> inlet) as a function of temperature, in the absence and presence of NO for a stoichiometry of  $\lambda = 0.33$  (Sets 3 and 7 in Table 1).



Fig. 6. Normalized NH<sub>3</sub> concentration (NH<sub>3</sub>/NH<sub>3</sub> inlet) as a function of temperature, in the absence and presence of NO for a stoichiometry of  $\lambda = 1.17$  (Sets 4 and 8 in Table 1).



Fig. 7. Normalized NH<sub>3</sub> concentration (NH<sub>3</sub>/NH<sub>3</sub> inlet) as a function of temperature, in the absence and presence of NO for a stoichiometry of  $\lambda = 3.33$  (Sets 6 and 10 in Table 1).

mainly proceeds through reactions r11 to r15. While in the absence of NO, conversion of  $NH_3$  is shifted to lower temperatures as the stoichiometry becomes more oxidant, which is attributed to the higher availability of radicals because of the radical building up with temperature, in the presence of NO a different behavior is observed.

When NO is present, reaction r13 (NH<sub>3</sub> + NH<sub>2</sub>  $\Rightarrow$  N<sub>2</sub>H<sub>3</sub> + H<sub>2</sub>) is not active since NO competes with NH<sub>3</sub> for NH<sub>2</sub> radicals and in fact NO dominates the consumption of NH<sub>2</sub> radicals which are present in high levels through reactions r16 and r17, mainly for high oxygen concentrations. Under reducing conditions, the amount of NH<sub>2</sub> radicals is comparatively limited and NO reacts through reactions r16 to r19:

$$NH_2 + NO \rightleftharpoons N_2 + H_2O \tag{r16}$$

- $NH_2 + NO \Rightarrow NNH + OH$  (r17)
- $NO + HO_2 \rightleftharpoons NO_2 + OH$  (r18)

$$N_2H_2 + NO \rightleftharpoons N_2O + NH_2 \tag{r19}$$

leading to an increase of radicals, mainly OH that, in turn, favors conversion of  $NH_3$  into  $NH_2$  radicals. Figure SM1 in the supplementary material summarizes the reactions happening during the conversion of

ammonia under the conditions of the present work in the form of a reaction pathway diagram.

In order to analyze the full temperature range where ammonia conversion occurs, experiments in a different setup with a ceramic reactor were performed, sets 13 to 17 of Table 1. Figs. 8 and 9 show respectively the results of NH3 and NO for different stoichiometries and as a function of temperature obtained with a significantly higher gas residence time. As seen, model calculations (lines) reproduce the main experimental trends (as symbols) even though are shifted about 40 K to higher temperatures. It is seen both experimentally and by model calculations that the conversion at  $\lambda = 1.17$  and 1.67 occurs at a approximately the same temperatures, while increasing  $O_2$  availability up to  $\lambda$ = 3.33 shifts conversion to comparatively lower temperatures. The increase in residence time to 3100 /T(K) of Figs. 8 and 9 compared to the value of 195/T(K) of earlier representations (e.g. Figs. 5-7, solid symbols) acts to shift the conversion regime of ammonia more than 100 K, which indicates the importance of this parameter and its influence on the building up of the radical pool under the studied conditions.

Figs. 10–13 show the impact of varying the amount of NO keeping constant the inlet  $NH_3$  concentration, i.e. varying the the  $NH_3/NO$  ratio, on the concentrations of  $NH_3$  and NO for two different residence times, one approximately 16 times higher compared to the other. The results obtained with the low residence time (195/T(K)) are shown in Figs. 10 and 11 and those corresponding to the high residence time (3100/T(K)) in Figs. 12 and 13.

For the low residence times considered, results indicate that the effect of varying the NH<sub>3</sub>/NO ratio is negligible on both the NH<sub>3</sub> and NO concentrations, while for the high residence time studied, the different NO concentrations do have a slight effect on the conversion regime of NH<sub>3</sub> and NO. Compared to the NH<sub>3</sub>/NO ratio of approximately 1, increasing the NH<sub>3</sub>/NO ratio (i.e. decreasing NO concentration) shifts NO conversion to lower temperatures, while NH<sub>3</sub> conversion is slightly shifted to higher temperatures. As NO concentration increases, reaction NH<sub>3</sub> + OH  $\rightleftharpoons$  NH<sub>2</sub> + H<sub>2</sub>O is comparatively more important compared to the NH<sub>3</sub> + CO<sub>2</sub>  $\rightleftharpoons$  HNCO + H<sub>2</sub>O reaction, which implies a comparatively high concentration of NH<sub>2</sub> radicals. Thus, under these conditions the extent of the NH<sub>2</sub> + NO reaction is increased and the chain branching step where NH<sub>2</sub> is implicated is activated. This effect is more important for high residence times.

## 5. Conclusions

An experimental and simulation study of the oxidation of ammonia in a  $CO_2$  atmosphere, representative of biogas or oxy-fuel combustion has been performed in a variety of experimental conditions. The study







Fig. 9. NO concentration as a function of temperature for different stoichiometries and a high residence time (Sets 13 to 15 in Table 1).



**Fig. 10.**  $NH_3$  concentration as a function of temperature for different  $NH_3/NO$  ratios at the low gas residence time, 195/T(K) s. Sets 11, 9 and 12 in Table 1.



Fig. 11. Normalized NO concentration (NO/NO inlet) as a function of temperature for different  $NH_3/NO$  ratios at the low gas residence time, 195/T(K) s. Sets 11, 9 and 12 in Table 1.



Fig. 12.  $NH_3$  concentration as a function of temperature for different  $NH_3/NO$  ratios at the high gas residence time, 3100/T(K) s. Sets 16, 14 and 17 in Table 1.



Fig. 13. Normalized NO concentration (NO/NO inlet) as a function of temperature for different  $NH_3/NO$  ratios at the high gas residence time, 3100/T(K) s. Sets 16, 14 and 17 in Table 1.

has been carried out at atmospheric pressure using two different experimental set-ups, including respectively quartz and ceramic tubular flow reactors that allow to work in different ranges of temperature (up to 1500 K) and gas residence time. Different oxygen stoichiometries, from pyrolysis to significantly fuel-lean, were considered ( $\lambda = 0$  to 3.33). The experimental results have been interpreted with an updated literature detailed chemical-kinetic mechanism.

The results obtained have allowed to determine that the presence of a  $CO_2/Ar$  atmosphere has an effect on ammonia oxidation compared to what happens in a  $N_2$  atmosphere. Presence of  $CO_2$  shifts the onset of ammonia conversion to lower temperatures and slows down its conversion at the higher temperatures studied. Presence of  $CO_2$  is seen to induce a complex interaction of  $NH_3$  with the O/H radical pool or with  $CO_2$ , and this affects the main conversion channels of ammonia evolution, and thus affects the conversion of ammonia.

In the presence of  $CO_2$ , conversion of ammonia is favored as the availability of oxygen increases. Both NO and N<sub>2</sub> are formed from NH<sub>3</sub>, while N<sub>2</sub>O and NO<sub>2</sub> are found to be negligible under the studied conditions. However, a number of nitrogen species, in very low concentrations, seems to be formed, according to non-closure of the nitrogen balance at high temperatures and to model calculations.

Presence of NO in the NH<sub>3</sub>/CO<sub>2</sub> reaction system results in the

promotion of  $NH_3$  conversion under fuel-rich conditions, while it inhibits the onset of the conversion initiation. Under the studied conditions, NO is reduced by action of  $NH_2$  radicals through the well-known SNCR reactions.

Direct interaction of  $CO_2$  and  $NH_3$  plays a role under given conditions, in particular when those reactants are the dominant ones and the gas residence time is high. However, in the presence of increasing NO concentrations, the impact of the  $CO_2$  atmosphere decreases. The complex observations and behavior attributed to the  $CO_2$  presence is an indication that further studies dealing with the impact of a  $CO_2$  atmosphere by comparison to a  $N_2/air$  atmosphere may be of interest.

Even though further work is still necessary in order to predict with precision the different species, the detailed kinetic mechanism used in calculations is able to reproduce the main trends observed experimentally, as well as the ammonia concentration profiles.

## CRediT authorship contribution statement

María U. Alzueta: Conceptualization, Formal analysis, Investigation, Writing – review & editing. Jorge Giménez-López: Investigation. Víctor D. Mercader: . Rafael Bilbao: Writing – review & editing.

#### **Declaration of Competing Interest**

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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