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Madagascar's extraordinary biodiversity: Threats and opportunities

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96

97 **Abstract:**

98 Madagascar's unique biota is heavily impacted by human activity and under intense threat. Here,
99 we review the current state of knowledge on the conservation status of Madagascar's terrestrial
100 and freshwater biodiversity by presenting data and analyses on documented and predicted species-
101 level conservation status, the most prevalent and relevant threats, *ex situ* collections and programs,
102 and the coverage and comprehensiveness of protected areas. The existing terrestrial protected area

103 network in Madagascar covers 10.4% of its land area and includes at least part of the range of the
104 majority of described native species of vertebrates with known distributions (97.1% of freshwater
105 fishes, amphibians, reptiles, birds and mammals combined) and plants (67.7%). The overall figures
106 are higher for threatened species (97.7% of threatened vertebrates and 79.6% of threatened plants
107 occurring within at least one protected area). IUCN Red List assessments and Bayesian neural
108 network analyses for plants identify overexploitation of biological resources and unsustainable
109 agriculture as the most prominent threats to biodiversity. We highlight five opportunities for action
110 at multiple levels to ensure that conservation and ecological restoration objectives, programs and
111 activities take account of complex underlying and interacting factors and produce tangible benefits
112 for the biodiversity and people of Madagascar.

113

114 **One Sentence Summary:** Current knowledge on Madagascar's biodiversity and its decline
115 indicates an urgent need for inclusive actions.

116 **Main text:**

117 Madagascar's biota, the result of millions of years of evolution in relative isolation, is both unique
118 and under threat. At the same time as the scientific description of new species is accelerating (1),
119 so is the overall rate of extinction (2), and many species may be disappearing before they are even
120 documented. In this review, we aim to consolidate information on the conservation status of some
121 of the main elements of Madagascar's biodiversity, evaluate the many and varied threats faced by
122 species assessed under the criteria for the International Union for Conservation of Nature (IUCN)
123 Red List of Threatened Species, and provide some perspectives on future opportunities to ensure
124 the future of this hyperdiverse and unique biota.

125

126 **Threats to Madagascar's biodiversity**

127 Madagascar's biodiversity is in decline, with some groups more threatened than others (Fig. 1). In
128 our review of threatened species, we follow the IUCN Red List data (3) and threat categories (4),
129 unless otherwise specified. Threatened species are those listed as Critically Endangered (CR),
130 Endangered (EN) or Vulnerable (VU). At one extreme, 22% (35 species) of assessed birds are
131 threatened, while, at the other end of the scale, approximately 73% (66 species) of freshwater
132 fishes and 75% (173 species) of magnoliid plants are threatened. Trees are particularly important
133 in terms of their broad ecological functions and human uses, and 63% of the 3,118 assessed tree
134 species in Madagascar are threatened (5). Humans have impacted the environment since arrival on
135 Madagascar, not only in recent years. To avoid a shifting baseline effect, it is necessary to view
136 changes in light of human settlement beginning hundreds or even thousands of years ago (1). For
137 example, despite the relatively low proportion of bird species currently threatened with extinction,

138 Madagascar has already lost at least 14 species (7% of all species) that were present when humans
139 first settled the island (Fig. 1). The rate of anthropogenic extinction is even higher in mammals,
140 with 23 species (10%) extirpated since first human settlement. Vertebrate extinctions include the
141 loss of lineages representing millions of years of evolution – e.g., the sloth-, koala- and monkey-
142 lemurs (families Palaeopropithecidae, Megaladapidae, and Archaeolemuridae) and two species of
143 hippopotamus (family Hippopotamidae). The extinction of four species of elephant birds (order
144 Aepyornithiformes) represents the global loss of a functionally unique clade (6, 7). Extinctions,
145 especially those of megafauna such as these, have broad scale implications for ecosystem
146 functioning (6-8).

147 In total, 13 endemic animal species are listed as Extinct (EX), defined as extinctions after 1500
148 AD, and an additional 33 are listed as Extinct Prehistorically [EP], defined as anthropogenic
149 extinctions prior to 1500 AD (see (9) for a full list of documented anthropogenic extinctions before
150 1500 AD). A further nine have been categorized as Critically Endangered (Possibly Extinct) –
151 CR(PE). For plants, no species has been assessed as Extinct, and only one species (*Aloe silicicola*)
152 is categorized as Extinct in the Wild (EW). A further 118 plant species are listed by IUCN as
153 CR(PE) (111 spp.) or Critically Endangered (Possibly Extinct in the Wild) – CR(PEW) (7 spp.).
154 Of those currently listed as CR(PE), five species are present in *ex situ* living collections, and their
155 status should therefore be updated to CR(PEW) (3, 10).

156 Malagasy species feature prominently among animal groups that have been considered by the
157 EDGE of Existence program (11-13), which ranks species according to their evolutionary
158 distinctiveness and the level of threat they face (EDGE = Evolutionary Distinct and Globally
159 Endangered). Almost one in five species of amphibians (18 spp.), reptiles (17 spp.), and mammals

160 (17 spp.) in the top 100 EDGE species of each group are found in Madagascar (13). Yet only one
161 in 20 (4 spp.) of the top 100 EDGE species of birds are found on the island.

162 Given the narrow geographic range of many Malagasy species (e.g., (14)), numerous undetected
163 anthropogenic extinctions are likely to have taken place (15), such as CR *Aloe* species, which may
164 have become extinct in the wild since they were last recorded. This may be especially pronounced
165 in groups with high levels of micro-endemism, for example freshwater fishes and amphibians (16).
166 Ascertaining extinction events is difficult due to sampling biases, insufficient taxonomic
167 knowledge regarding the morphological features of extant species, and the challenges of
168 comparisons with fossil and subfossil remnants in certain groups, such as frogs (e.g., (17)).

169

170 *Reliability of species conservation assessments*

171 Conservation assessments rely on taxonomic classification, and different opinions on species
172 limits and numbers may influence the proportion of threatened species (e.g., (18)). This proportion
173 may also be biased by an over-assessment of well-known and widespread taxa, or, alternatively,
174 range-restricted species that are more likely to be threatened. To investigate indications of bias,
175 we calculated the fraction of threatened species across different plant groups based on two sets of
176 species: taxa with full threat-status assessments in the Red List compiled by the IUCN and their
177 partners (19); and those estimated with a Bayesian neural network approach (Fig. 1; (9, 20)), which
178 inferred the threat status for all remaining species. Using this method, we predicted the threat status
179 of 8,821 species with an estimated test accuracy of >65%. All taxa with a full threat-status
180 assessment were included, although some assessments may be out of date and could underestimate
181 threat levels.

182 The neural network approach combined with current IUCN assessments revealed a similar fraction
183 of species inferred to be threatened across most taxonomic groups (Fig. 1). Large deviations from
184 the proportion of threatened species in the current IUCN assessments occur in the ferns and
185 lycophytes, and to a lesser extent the magnoliids. The neural network results combined with the
186 known IUCN categories predicted a far higher proportion of threatened ferns and lycophytes (146
187 of 306 spp; 47.7% [95%CI: 38.5-56.7%]) than reflected in published IUCN assessments (1 of 33
188 spp; 3.0%), suggesting a bias towards assessing more common species. In the magnoliids, the
189 combined results predict a lower proportion of threatened species (211 of 294 spp; 71.8% [95%CI:
190 68.0-75.9%]) compared to published IUCN assessments alone (173 of 225 spp; 76.9%),
191 suggesting a bias towards assessing rare species in that group.

192

193 *Genetic erosion*

194 The reduction of genetic diversity within species resulting from the extirpation of
195 subpopulations is a crucial, yet easily overlooked, facet of biodiversity loss that is often a precursor
196 to extinction. Genetic erosion has negative effects on the individual fitness, the health of
197 populations, and a species' ability to adapt to changing environments, reducing their resilience to
198 further change, and potentially incurring extinction debt (21, 22). In practice, genetic factors are
199 not directly incorporated into IUCN assessments, which are based on measures of the probability
200 of extinction due to population declines, restricted geographic ranges, and small population sizes
201 (23).

202 The reduction in population sizes of wild plants and animals, together with their fragmentation and
203 isolation, is generally expected to increase inbreeding and genetic load, reducing genetic diversity

204 and fitness over time (22, 24). The few studies of intraspecific diversity in Malagasy species to
205 date reveal that some species have maintained high genetic diversity in spite of habitat
206 fragmentation (e.g., (25, 26)), whereas others have relatively low diversity, possibly as a result of
207 anthropogenic effects (e.g., (25, 27-29)). Results differ even within species, such as in the palm
208 *Beccariophoenix madagascariensis*, in which only some populations show strong signals of
209 inbreeding, reflected by an excess of homozygotes (30). It is important to note that under some
210 circumstances, population decline may outstrip the speed with which genetic diversity is eroded
211 due to inbreeding. Estimates of heterozygosity may therefore not indicate the true genetic health
212 and long-term prospects of populations when considered in isolation (31, 32).

213 A more powerful, although less explored, approach is to use coalescence-based demographic
214 modeling, which uses genome-wide data to estimate the longer-term trends in population size,
215 providing more information than metrics of contemporary genetic diversity alone (25, 33). In
216 *Cheirogaleus* dwarf lemurs, genomic analysis suggests that four species have experienced
217 population size declines in the last 50,000 years, with one decline (*C. cf. medius*) starting as long
218 as 300,000 years ago – all clearly in pre-human times and resulting in lower genetic diversity (29).
219 In contrast, another genomic study shows that five out of ten analyzed plant species with varying
220 extinction risk have experienced substantial population declines since human colonization of
221 Madagascar (25). In the golden-crowned sifaka (*Propithecus tattersalli*) (26), mouse lemurs
222 (*Microcebus* spp.) (28), *Mantella* frogs (34), and the Milne-Edwards' sportive lemur (*Lepilemur*
223 *edwardsi*) (35) demographic declines also appear to have taken place after the arrival of humans
224 on the island (although the inherent uncertainties of mutation rates in the microsatellite data used
225 makes the timing of these declines less certain).

226 The risks of inbreeding and increased genetic load may represent substantial and likely
227 underestimated longer-term threats to the survival of Malagasy species. This is especially relevant
228 considering the high level of fragmentation of native habitats in some vegetation types, such as the
229 humid forests, and is worthy of further investigation.

230

231 **Predicting future extinction: direct drivers of loss**

232 Identifying direct threats is part of the IUCN Red List Assessment process, and even species that
233 are not explicitly threatened (i.e., those that are Least Concern [LC], Near Threatened [NT], or
234 Data Deficient [DD]) can still have threats listed. Here we discuss these threats and how they apply
235 to all species. Our analysis of IUCN assessments indicates that overexploitation and agriculture
236 are the most frequently listed threats to Malagasy fauna (excluding invertebrates) and flora (Fig.
237 2), mirroring global findings (36). Overexploitation is unsustainable biological resource use as
238 defined by the IUCN (37), including hunting and collecting for subsistence use or
239 national/international trade. Overexploitation is linked in some cases to illegal harvesting – for
240 example, the illegal logging of rosewood for trade (*Dalbergia* spp.) – which is banned under the
241 Convention on International Trade in Endangered Species of Wild Fauna and Flora since 2013 and
242 under Malagasy law since 2010.

243 We estimated that 62.1% of vertebrates and 87.1% of plants are threatened by overexploitation
244 and that 56.8% of vertebrates and 87.8% of plants are threatened by agriculture. These two major
245 threats, almost equal in magnitude (Fig. 2), have different modes of impact – overexploitation is
246 more targeted and tends to occur over relatively restricted areas compared to the broad effects of
247 land clearance for agriculture.

248 Agriculture, and to a lesser extent overexploitation, are also the primary causes of deforestation in
249 Madagascar. Approximately 44% of the land area covered by native forest in 1953 was deforested
250 by 2014 (38). The rate of deforestation has steadily increased, reaching 99.0 kha/yr between 2010
251 and 2014 (38), and according to Global Forest Watch remains very high at 72.9 kha/yr (2014–
252 2020) (39). Deforestation in Madagascar reflects global patterns (40) and is primarily driven by
253 the small-scale but widespread practice of swidden agriculture (also known as shifting cultivation;
254 in Madagascar referred to as *tavy* for rice cultivation in humid and subhumid areas, and *hatsake*
255 for cassava and maize in dry and subarid areas). Additionally, cash crop production, particularly
256 maize and peanut, has become a major driver of deforestation (41), alongside the production of
257 products for international markets, such as forest-derived vanilla (42). The most frequent threats
258 listed for plants and vertebrates suggest that this trend of increasing deforestation rates will
259 continue, with forest loss and degradation a consequence of clearance of land for agriculture,
260 potentially associated with small-scale fire activity (43) and overexploitation through selective
261 logging and highly targeted activities such as the collection of palm hearts. Additionally, natural
262 system modifications (threats from actions that convert or degrade habitat, e.g., anthropogenic fire
263 in forests or changes in water management; Fig. 2), adds to deforestation and threatens 23.2% of
264 vertebrates and is estimated to threaten 68.9% of plants. Some predictions indicate that in the
265 absence of an effective strategy against deforestation, 38–93% of forest present in 2000 will be no
266 longer present in 2050 (41).

267 For vertebrates, the greatest threat after overexploitation and agriculture is ‘invasive and
268 problematic species and emerging infectious diseases’ (referred to as invasives/diseases in Fig. 2),
269 which impacts 27% of all species (360 spp.; Fig. 2). This category includes non-native invasive
270 species, as well as problematic native species and diseases of any origin. Changes in habitat due

271 to the spread of non-native plant species can have a large effect, and one study reports that of a
272 total of 546 naturalized non-native plants in Madagascar, 101 have been found to display invasive
273 characteristics (44). Many non-native plants, such as the Mexican yellow pine (*Pinus patula*) in
274 terrestrial systems (45), and common water hyacinth (*Pontederia crassipes*) in freshwater systems
275 (46), are aggressively invasive and transformative in semi-natural habitats, and are clearly
276 impacting native fauna and flora. Even within reserves and protected areas, the issue can be
277 pronounced. For example, three species of invasive/problematic plants – strawberry guava
278 (*Psidium cattleianum*), Molucca raspberry (*Rubus moluccanus*), and wild cardamom (*Aframomum*
279 *angustifolium*) – together occupy 17.6% of the Betampona Nature Reserve (47) and are also
280 widespread in Ranomafana National Park and other protected areas.

281 Not all impacts are negative, however, and there is some evidence to suggest that, due to their
282 potential for faster growth, some non-native plants are better able to combat the rapid
283 fragmentation of native vegetation, and may be beneficial for endemic vertebrates, providing
284 refuge, food, and vegetation corridors, while also improving human livelihoods (48). The potential
285 for such species to become invasive or readily burn must however be fully considered before
286 embarking on any planting initiatives (49). In addition, effects must be considered at different
287 scales. For examples, the presence of strawberry guava has been reported to locally increase
288 species richness in frugivores, but as they are primary dispersers of the seed this further contributes
289 to the spread and to associated changes in floral and faunal community structure and reduction in
290 taxonomic richness (50).

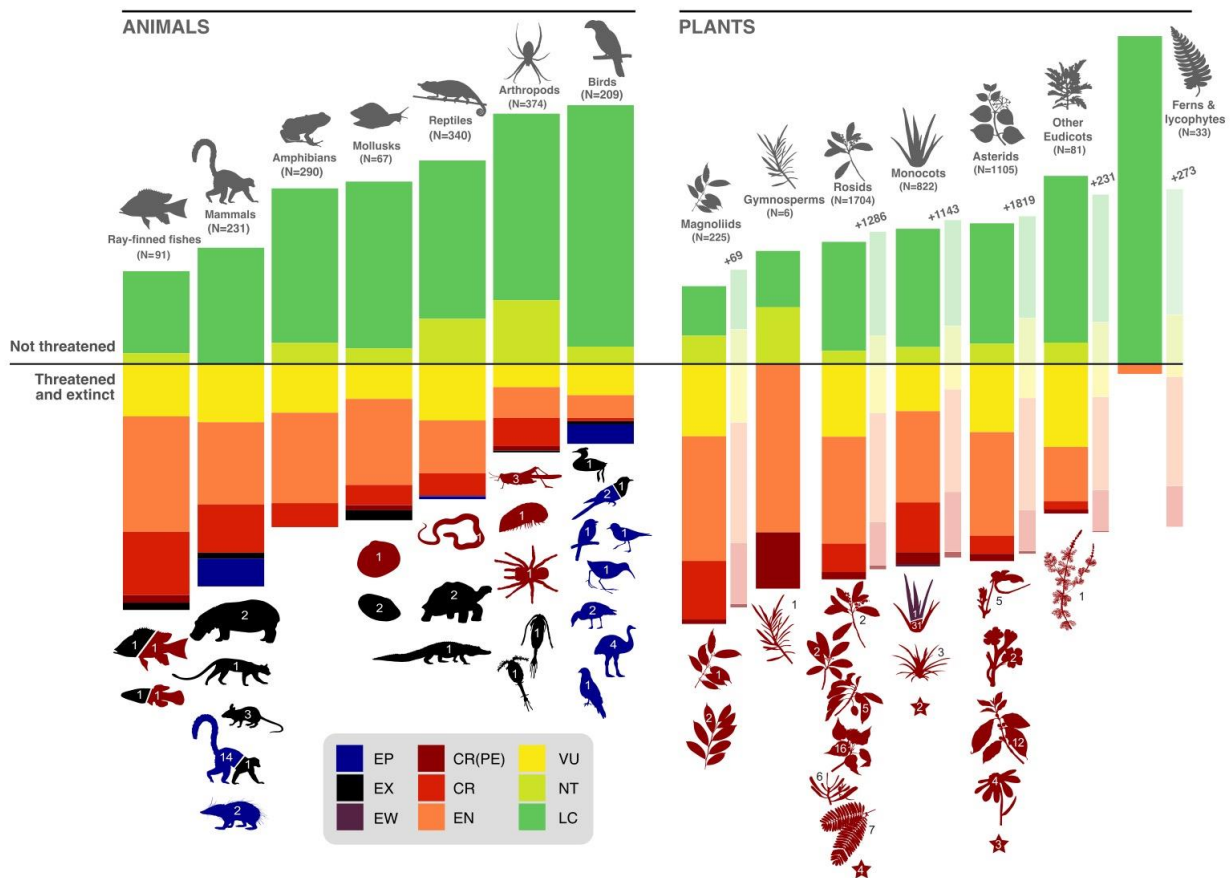
291 Non-native vertebrates have also had marked and diverse impacts, which we also here illustrate
292 with some examples. Introduced rats (*Rattus rattus*; present since at least the 14th century) are
293 now ubiquitous, even in remote areas, and there is evidence that their presence is associated with

294 declines in native small mammals (51). In freshwater habitats, competition and predation by exotic
295 fish species is considered a major factor in the decline of native freshwater fish (52), which have
296 been completely replaced by non-native species across much of the Central Highlands and western
297 areas (53). While not yet listed in current assessments, the recent invasion of the toxic Asian
298 common toad (*Duttaphrynus melanostictus*), along with the predicted vulnerability of most native
299 vertebrates to its toxins (54), is expected to represent a new threat to many nocturnal carnivores.
300 The effects of other introduced and naturalized animals on native biodiversity are not well studied;
301 this includes widely occurring species such as dogs (*Canis familiaris*), cats (*Felis catus*), the
302 common myna (*Acridotheres tristis*), and the marbled crayfish (*Procambarus virginalis*). The
303 threat of emerging infectious diseases is primarily driven by the occurrence of the chytrid fungus
304 *Batrachochytrium dendrobatidis*, widely documented across Madagascar over the last decade and
305 a potential threat to all amphibians, although no mass mortalities associated with chytridiomycosis
306 have been reported in the country (55). Species often face multiple threats at the same time,
307 although the impact of each threat can vary between species (Fig. 2).

308 Among vertebrates, amphibians have the highest number of IUCN-identified threats per species
309 (Fig. 2A), with a mean of 4.8 threats per species, followed by mammals (mean 2.5 threats/species),
310 and reptiles (mean 2.2 threats/species). For plants (Fig. 2B), magnoliids have the most threats per
311 species (mean 2.9 threats/species), followed by rosids (mean 2.8 threats/species), and other
312 eudicots (mean 2.8 threats/species). Although there might be some variation in the perception and
313 documentation of threats between the specialists carrying out assessments, all follow the same
314 protocols (4).

315 The number and relative impact of these threats may change in coming decades. The impact of
316 climate change on Malagasy biodiversity remains understudied and it is not currently indicated in

317 IUCN assessments as a major threat. However, this impact is expected to increase in the future
 318 (56-59), and could potentially result in synergistic negative effects with unsustainable agriculture
 319 associated with land clearance, invasive alien species, and inappropriate management of fire
 320 regimes that can increase future fire risk (43, 56, 57, 60). Extinctions in one group could also have
 321 effects on others that depend on them, such as in cases of strong plant–animal mutualisms (61, 62).
 322 Although coextinction is hard to quantify, with substantial knowledge and data gaps (63), models
 323 suggest that the effects of extinction can be amplified as a result of the interactions between species
 324 within and between trophic levels, with the potential to lead to secondary and even cascading
 325 extinctions (64, 65).



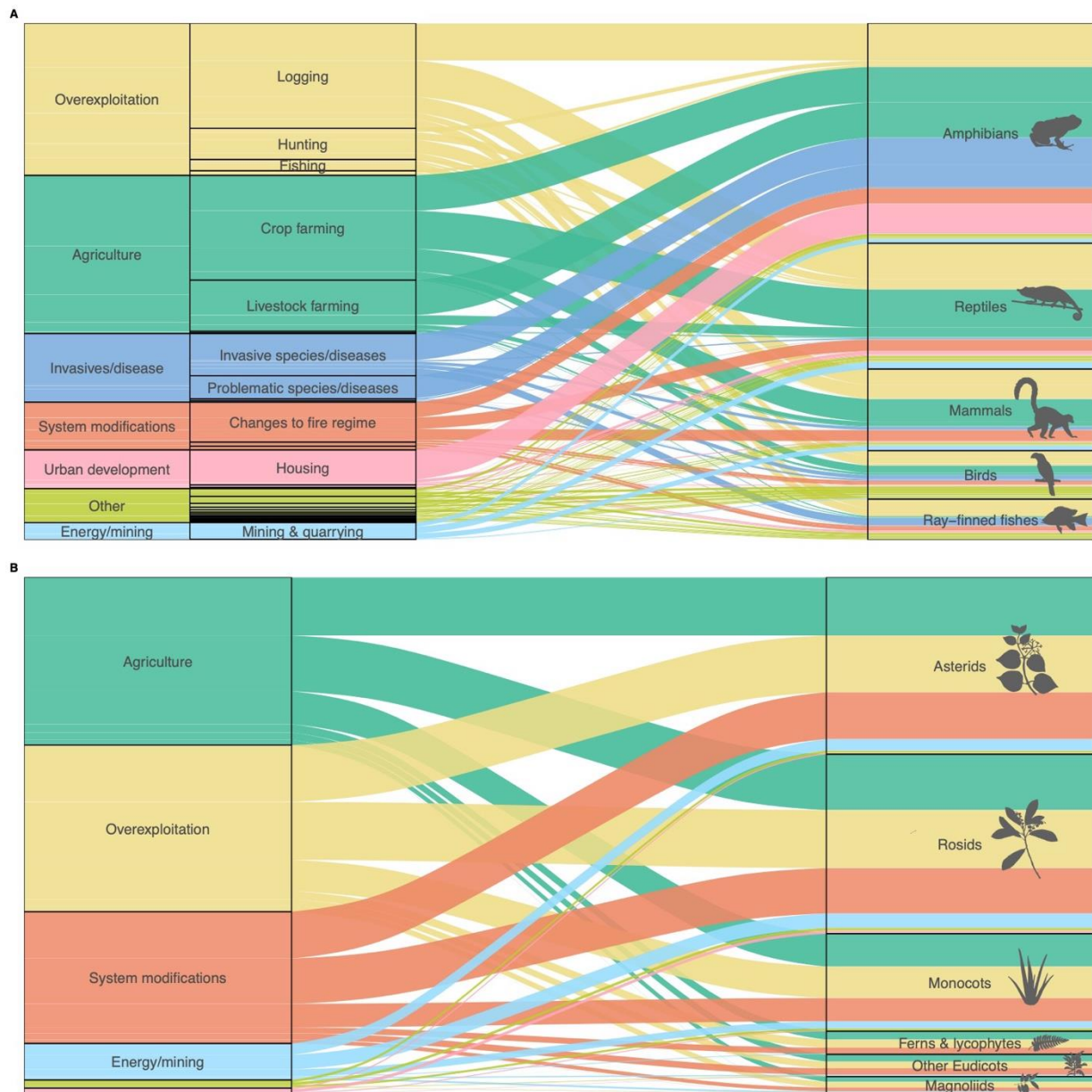
326

327

328 **Fig. 1. Madagascar's threatened and lost biodiversity.** IUCN Red List assessment categories
329 of major groups of plants and animals from Madagascar. Assessment categories and coloration
330 follow the standards used by the IUCN Red List. Category distributions for animal groups
331 include ray-finned fishes, (Actinopterygii, freshwater species only, N=91), mammals
332 (Mammalia, N=231 species), amphibians (Amphibia, N=296), mollusks (Mollusca, N=67),
333 reptiles (Reptilia, N=340), arthropods (Arthropoda, N=374), and birds (Aves, N=209). Category
334 distributions for plants, indicated with saturated, wider bars, include magnoliids (N=225),
335 gymnosperms (N=6), rosids (N=1,704), monocots (N=822), asterids (N=1,105 species), other
336 eudicots (N=81), and ferns & lycophytes (N=33). Thinner, unsaturated bars indicate the relative
337 proportion of plant taxa in each threat category for IUCN Red List assessments combined with
338 the taxa where the threat category was predicted in a Bayesian neural network analysis: asterids
339 (N=2,924 species), rosids (N=2,990), other eudicots (N=312), magnoliids (N=294), monocots
340 (N=1,965), and ferns & lycophytes (N=306). The number indicated above each bar with "+" is
341 the number of taxa for which the threat category was predicted using the neural network analysis.
342 IUCN Red List Assessment categories include Least Concern (LC) and Near Threatened (NT),
343 together making up the "not threatened" category; while Vulnerable (VU); Endangered (EN);
344 Critically Endangered (CR); Critically Endangered, Possibly Extinct (CR(PE)); Extinct in the
345 Wild (EW); Extinct (EX; i.e., extinct after 1500 CE), and Extinct Prehistorically (EP; sensu (66),
346 i.e., extinct before 1500 CE but with dated records within the last 130,000 years) make up the
347 group "threatened and extinct." Silhouettes below the bars depict taxonomic orders with EP, EX,
348 EW, and CR(PE) species, with the number of species in each category per order. For some plant
349 groups, additional orders with single CR(PE) species are indicated with a star. Depicted orders
350 are, from left to right and top to bottom: Perciformes, Cyprinodontiformes, Cetartiodactyla,
351 Carnivora, Rodentia, Primates, Afrosoricida, Venerida, Unionoida, Perciformes,

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352 Cyprinodontiformes, Squamata, Testudines, Crocodylia, Orthoptera, Spirobolida, Araneae,
 353 Calanoida, Cyclopoida, Podicipediformes, Cuculiformes, Coraciiformes, Charadriiformes,
 354 Gruiformes, Anseriformes, Aepyornithiformes, Accipitriformes, Laurales, Magnoliales, Pinales,
 355 Oxalidales, Sapindales, Myrtales, Malvales, Malpighiales, Fabales, Asparagales, Poales,
 356 Ericales, Boraginales, Gentianales, Asterales, Saxifragales.
 357



358

359 **Fig. 2. Threats to Malagasy biodiversity.** Alluvial plots showing threats, as defined by the IUCN,
360 and their associations with major groups of terrestrial and freshwater (A) vertebrates (1,332 species
361 with IUCN assessments, of which 993 species have at least one listed threat) and (B) plants (9,268
362 species with IUCN assessments or predictions, all of which have at least one listed threat; includes
363 gymnosperms [6 species], which could not be visualized). Widths of the boxes/lines reflect the
364 number of species impacted by each threat. Threats for vertebrates are further divided into sub-
365 threats, whereas only the highest threat classification was available for assessed plants. The
366 estimates for plants include predictions for unassessed species based on a Bayesian neural network
367 analysis (9). The color scheme is consistent across panels. The “Other” threat class includes
368 Pollution, Climate change, Transportation, and Human disturbance, plus Invasives/diseases for
369 plants. Some threat classes have been renamed for brevity/clarity, including the IUCN category
370 “biological resource use”, which is labeled “overexploitation” here and in the text, for brevity and
371 in line with IPBES terminology (36).

372

373 **Conservation efforts and effectiveness**

374 *Protected Areas*

375 Protected areas (PAs) are the central political and scientific accomplishment of Madagascar's
376 conservation strategy. The network has been continuously developed since the first PA was
377 established in 1927 (67-71). Our data compilation shows that the network now encompasses 10.4%
378 of the land area of Madagascar, having grown by more than a third over the last two decades (Fig.
379 3). This recent and extensive designation of new PAs was carried out via a multi-stakeholder
380 consultative process, in combination with data and literature analyses, through the Durban Vision

381 initiative conceived in 2003. In addition to preserving diverse ecosystems and landscapes, the
382 focus has been on species groups for which sufficient diversity and distribution data were
383 available, primarily vertebrates (including birds, mammals, amphibians, and reptiles), and some
384 plant groups. Despite the production of considerable new data since the Durban Vision began (e.g.,
385 many newly described species; (1), the network designed during that process remains highly
386 taxonomically comprehensive. From a global perspective, the PA network also excels at capturing
387 the vast majority of Madagascar's many EDGE species: 14 out of 18 amphibians, 15 out of 17
388 reptiles, 16 out of 17 mammals, and all four birds (13).

389 As of November 2020, there were 110 terrestrial PAs with permanent protected status in
390 Madagascar, covering 61,300 km² across the country (Fig. 3) (70, 72, 73). Eleven of these are
391 “orphan PAs” – sites abandoned by their former managers with responsibility reverting to the
392 Ministry of Environment and Sustainable Development (70). An additional 89 sites (15,200 km²),
393 predominantly comprising Key Biodiversity Areas (KBAs), are not under formal protection (70,
394 72, 74, 75).

395 The long-term security and effective management of Madagascar's PAs is therefore crucial to
396 addressing the country's biodiversity challenges. Providing evidence of their effectiveness and co-
397 benefits, such as ecosystem service provision, will be critical to securing ongoing support and
398 management from local communities, as well as from local and national governments. However,
399 measuring PA effectiveness is challenging (e.g., at avoiding deforestation, or providing alternative
400 livelihoods) while accounting for numerous covariates (76), particularly in Madagascar with
401 comparatively little long-term biodiversity monitoring data (77). Recent counterfactual analyses
402 (78) have sought to address this question by identifying protected and non-protected sites that are
403 similar across multiple social and environmental variables, and then comparing indicators of

404 conservation effectiveness, such as deforestation rate. These analyses indicate that PAs have a
405 small, but significant, effect at reducing deforestation (9).

406 We show that since 1990, human impacts have measurably increased across all terrestrial PAs
407 (Table S8 (9)), a trend documented worldwide (76). Human activity by local communities inside
408 PAs is not necessarily detrimental to biodiversity, and land use and conservation are therefore not
409 mutually exclusive. Nevertheless, land conversion and unsustainable exploitation remain major
410 drivers of biodiversity loss. This suggests that protecting and realizing the potential of
411 Madagascar's comprehensive PA network will require the application of rigorous monitoring and
412 evaluation strategies, matched with extensive community collaboration, to understand co-benefits
413 and minimize detrimental human impacts.

414 Scores for deforestation and management effectiveness – for example, from the self-reported
415 Management Effectiveness Tracking Tool (79) – have been the main metrics used to monitor
416 effectiveness to date. However, these are not always reliable indicators of management
417 effectiveness (77). New and expanded capacity of variables such as remote-sensed fire and stable
418 night lights, with increased temporal resolution, offer promising new monitoring opportunities.
419 How fire is associated with land transformation in Madagascar has been discussed in the literature
420 but only recently quantitatively assessed (43), demonstrating that tree loss anomalies are highest
421 in environments where landscapes-scale fire (>21 ha) does not occur, and where the role of small-
422 scale fires (<21 ha) requires close and urgent investigation. We show that trends in anthropogenic
423 fire are variable, increasing in some areas of forest vegetation in the north, east, and west but
424 decreasing in grassland-woodland mosaic vegetation across central Madagascar (Fig. 4A, B).
425 Forest loss also reflects this pattern, primarily occurring in the humid forest biome in the east, but
426 also in dry forest and spiny forest in the west (Fig. 4C, D). Deforestation and land use conversion

427 remain key challenges to conservation in Madagascar, and improved remote-sensing will
428 accelerate monitoring and developing understanding on the effectiveness of PAs and other
429 conservation measures.

430

431 *Ex situ conservation and restoration*

432 Living plant collections in botanic gardens and seed banks represent invaluable sources of
433 taxonomic and genetic diversity for immediate conservation and research, and should continue to
434 support restoration efforts. Globally, 29.6% of all known native Malagasy plant species (23.1% of
435 endemic species and 23.1% of native threatened species) are held in botanic gardens, with 15.5%
436 held in Madagascar (10), where their cultivation is sometimes linked to educational programs and
437 community engagement, essential to raising awareness of biodiversity and conservation issues.
438 The Millennium Seed Bank Partnership in Madagascar, initiated in 1996, hosts collections of an
439 estimated 3,500 native Malagasy species, including members of four of the five endemic plant
440 families and all seven of the iconic baobab species (*Adansonia* spp.). The single Malagasy plant
441 species listed as Extinct in the Wild, *Aloe silicicola*, now only survives in one living collection
442 outside Madagascar.

443 For native terrestrial and freshwater vertebrates, 9% of amphibians, 17% of mammals, 20% of
444 reptiles, 21% of freshwater fishes, and 33% of birds are currently held in zoological collections
445 (18% overall) (9, 80). Many are part of active breeding programs: a subset of these (3% of
446 amphibians, 7% of reptiles, 11% of freshwater fishes, 13% of mammals, and 23% of birds) were
447 successfully bred during 2020 (9). Unsurprisingly, the species held in captive breeding facilities
448 are biased towards the more charismatic, well-known taxa (81). For example, among amphibians,

449 13 of the 34 species in zoos belong to the genus *Mantella*, a group of strikingly colored diurnal
450 frogs, even though *Mantella* contains only 4% of Madagascar's amphibian fauna. Freshwater
451 fishes, amphibians, and reptiles are highly suitable for targeted *ex situ* breeding and reintroduction
452 programs (82-85). For species in these groups and others with high levels of micro-endemism,
453 such conservation programs continue to represent a major safeguard against extinction (86). This
454 complies with the One Plan Approach to species conservation proposed by the IUCN SSC
455 Conservation Planning Specialist Group, which supports the development of conservation and
456 management plans for all populations of a species, even outside of their natural range (87). It
457 should be noted that the success of reintroduction relies also on the maintenance of natural habitat
458 and functional diversity at potential reintroduction sites, along with minimizing risks associated
459 with invasive species and infectious diseases. In addition, particularly for mammals, vulnerability
460 of captive-bred populations to predation can also jeopardize the success of reintroductions (88).

461 *Progress towards international conservation commitments*

462 Madagascar continues to make progress towards Convention on Biological Diversity targets, but
463 like most countries falls short of meeting them in full (89). Of particular relevance here is that
464 Madagascar did not formally meet Aichi Target 11 to protect at least 17% of its total land area
465 (Fig. 3) – as was the case for 48% of the parties reporting their progress (89). If areas designated
466 as important for biodiversity but not currently under formal protection were also given protection,
467 the total percentage of PA coverage would rise from the current 10.4% to 13% (Fig. 3B). However,
468 given that even the existing network is widely considered to be chronically under-resourced, this
469 action is not a priority for the near future (90, 91).

470 Target 4 of the Global Strategy for Plant Conservation seeks to protect 15% of each vegetation
471 type. This has been achieved for mangrove (currently at 29.4%), spiny forest (21.5%), humid forest
472 (18.5%), and tapia (17.9%), but not for dry forest (13.3%), subhumid forest (5.7%), and grassland-
473 woodland mosaic (1.8%) (Table S6 (9)). However, expansion of the areas of those vegetation types
474 under protection may not be feasible due to limited financial resources, the large degree of
475 fragmentation and geographical spread of habitats, and the long administrative process involved
476 in extending PAs or designating additional areas, as well as a lack of political will. It also may not
477 be desirable until it can be demonstrated that the existing PAs are well-resourced, achieving
478 conservation objectives and providing benefits to communities. Restoration within currently
479 protected areas may provide a longer-term pathway to meeting this goal, particularly where there
480 are rapidly realizable socio-economic benefits such as sustainable silk production from wild native
481 silkworms (*Borocera cajani*) associated with tapia (*Uapaca bojeri*) in the Itremo Massif PA and
482 Ambatofinandrahana KBA. Other targets are more difficult to assess due to lack of data. For
483 example, there is very little evidence to assess success in the control of invasive alien species, with
484 some exceptions such as the ongoing but promising house crow (*Corvus splendens*) eradication
485 (92).

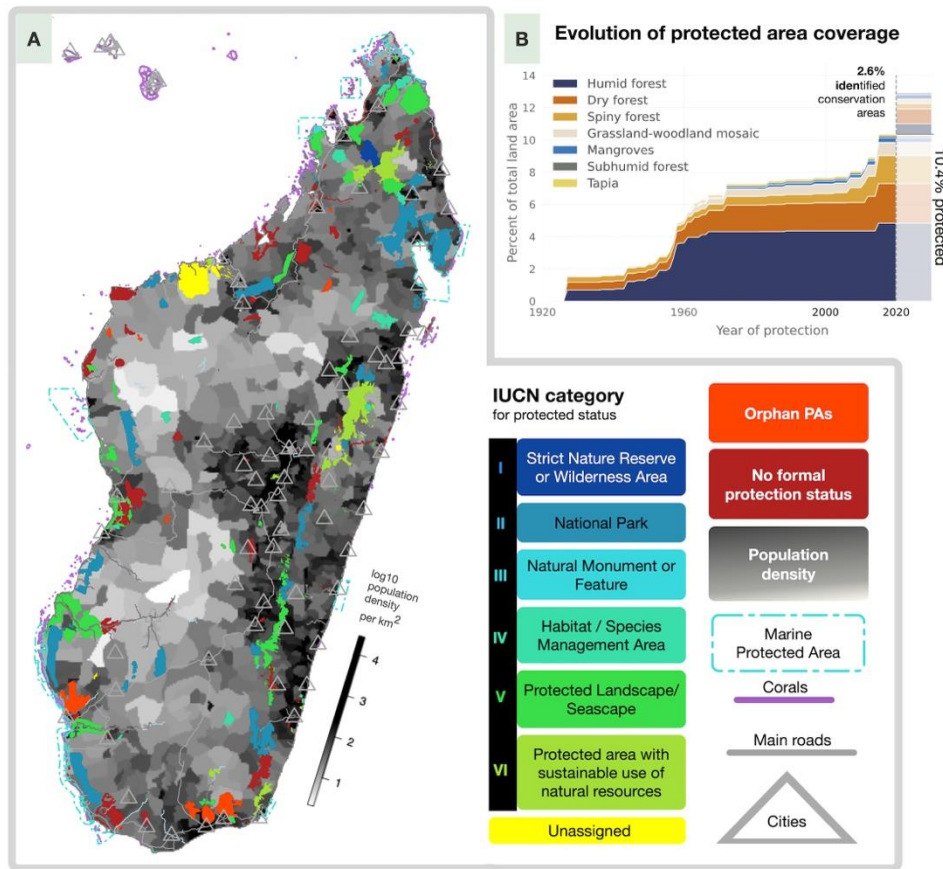
486 Although most of the Aichi and GSPC targets were either not achieved or cannot be assessed, a
487 marked success is that Madagascar has comfortably achieved GSPC Target 7 (at least 75% of
488 known threatened plant species conserved *in situ*), with our analyses indicating this percentage is
489 currently at 80%.

490

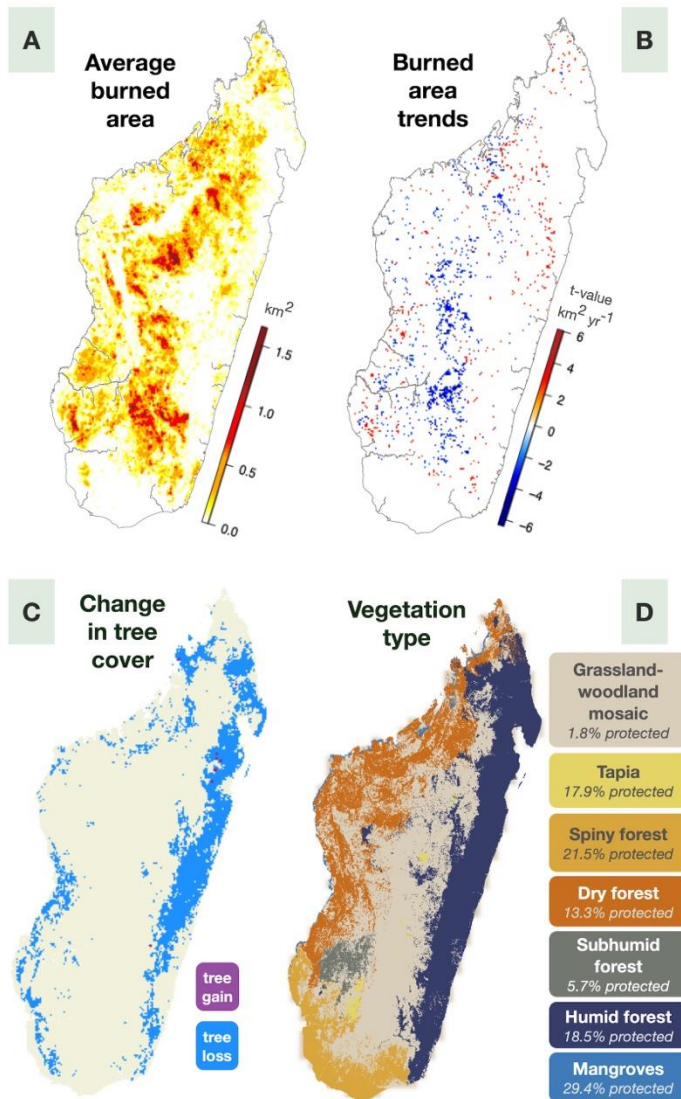
491 *Realizing benefits of biodiversity for people*

492 The majority of Madagascar's over 28 million inhabitants live outside of, but often very close to,
493 PAs (93) (Figs. 3A; S1). These communities face challenges connected to widespread poverty,
494 which itself is related to degradation of natural capital in the landscape, limited access to formal
495 education and health care, crime, corruption, weak governance, and regulatory issues including
496 land tenure (15, 94, 95). For example, southern Madagascar is severely affected by food and water
497 insecurity, which catalyzes political and social instability, exacerbates economic insecurity, and
498 has led to large-scale migration within the country (96). This instability likewise hampers the
499 operations of local, national, and international conservation organizations, which could be
500 compounded further by adverse effects from climate change (59). As the human population in the
501 country is expected to reach 42–105 million by the end of this century, of which half will be under
502 15 years of age, and with the majority under the poverty threshold (97), the conservation success
503 of PAs will be inextricably linked to the effective provision of livelihoods, food security, and
504 natural capital – a situation echoed across all Malagasy ecosystems and the world over (98).

505



506 **Fig. 3. Madagascar's terrestrial protected areas (PAs) in the context of human population**
 507 **density and changes in coverage of vegetation type over time.** (A) PAs with IUCN protected
 508 status (99), "orphan" status, or no formal protection status (e.g., unprotected Key Biodiversity
 509 Areas [KBAs]), shown in the context of nearby marine PAs, surrounding bathymetry (100), coral
 510 reefs (101), cities, roads, and population density (102). (B) The evolution of PA coverage over
 511 time, showing the potential increase in area protected that could be gained if the designated areas
 512 (those identified as important for biodiversity but not currently under formal protection, mostly
 513 KBAs) were protected in the future (74, 75).



514

515 **Fig 4. Recent changes and patterns in burned area and tree cover in Madagascar.** (A)

516 Average burned area in the period 2003–2019. (B) Statistically significant trends in burned area

517 (MODIS) (103) from 2006–2016, not explained by precipitation change (TRMM) (104), dates

518 chosen for comparison with Goodman et al. (72). Red indicates an increasing trend; blue indicates

519 a decreasing trend. (C) Change in tree cover from 2000–2012 (105). (D) Vegetation map, inferred

520 and simplified from Moat & Smith (106). The legend indicates the percentage of each vegetation
521 category currently covered by the protected area network.

522

523 **Looking back, moving forward**

524 Despite decades of research and applied conservation programs supported through substantial
525 financial investments (95, 107), Madagascar's remarkable biodiversity continues to face severe
526 challenges (Figs. 1, 2). It is reasonable to ask whether more of the same – even if better resourced
527 and underpinned with greater scientific understanding and technology – is likely to deliver a
528 tangible reversal in Madagascar's trajectory of biodiversity loss, or whether new approaches are
529 required to bring transformative change (108), including greater emphasis on monitoring
530 interventions and addressing underlying drivers through key leverage points. The responsibility
531 for averting humanitarian and biodiversity crises is a shared global challenge (36, 109), with
532 solutions needed at all societal levels – including via local communities, engagement of the private
533 sector, sound leadership and policy from regional and national government, steady international
534 support for conservation, and increased recognition of how historic and ongoing global and
535 national inequalities have contributed to the current situation. Scientific data and evidence will
536 continue to make a vital contribution, but it is crucial that this is done in an interdisciplinary
537 context, with open communication channels to relevant government departments and third sector
538 organizations.

539

540 *Decades of progress in biodiversity science and conservation*

541 We now have a clearer and more detailed understanding than ever before of the past and present
542 diversity and distribution of Madagascar's biodiversity, and the threats it faces (1) (Fig. 1). The
543 underlying data are the product of decades of research – with an increasing number of Malagasy
544 biologists involved. This body of research and the evidence we have collated and presented here
545 makes a clear case for Madagascar as one of the world's foremost conservation priorities.

546 Despite multiple competing demands on land, the Malagasy government, in collaboration with a
547 broad group of conservation organizations and donors, has succeeded in designating 10.4% of the
548 country as terrestrial PAs in a network that is largely representative of Madagascar's diverse
549 biomes (Fig. 3, 4). Most terrestrial and freshwater vertebrate species with known distributions have
550 ranges that overlap with least one PA (94.7% of reptiles, 97.2% of amphibians, 98.1% of
551 mammals, 98.9% of freshwater fishes, 100% of birds, and 97.1% for all groups combined), as do
552 the majority of plants, but to a lesser extent (67.7%) (9). For threatened species with known
553 distributions, the percentages are similar for vertebrates (94.3% of reptiles, 99.3% of amphibians,
554 97.7% of mammals, 100% of freshwater fishes, 100% of birds, and 97.7% for all groups combined)
555 and markedly higher for plants (79.6%). Nonetheless, there are still many threatened species with
556 ranges that do not overlap with existing PA network, including one amphibian, three mammals,
557 seven reptiles, and 559 plants (9), and many more that have not yet been assessed but may be
558 threatened. The ranges of all birds overlapped with at least one PA; this was also true when we
559 filtered the analysis to only include resident and breeding areas (9).

560 Since the loss of Madagascar's terrestrial megafauna (here defined as vertebrates above 10 kg),
561 there have been few documented modern extinctions, but many species have perilously reduced
562 population sizes. The continued increase in new species descriptions suggests there may be
563 undocumented extinctions, especially in poorly studied taxa (1). Despite this, with limited

564 resources and/or capacity, Madagascar has made important progress towards achieving
565 international climate, biodiversity, and sustainable development goals, providing a foundation on
566 which to build in the coming decades.

567 Success stories for individual species highlight how positive collaborative efforts can avert
568 extinction. Examples include work on the Madagascar pochard (*Aythya innotata*) (110), which
569 shows a 30% probability that extinction was prevented due to conservation action, the success
570 story of the community-based protection of the tahina palm or dimaka (*Tahina spectabilis*) where
571 local communities were involved in propagation and population reinforcement (111), and the work
572 to prevent the extinction of the ploughshare tortoise (*Astrochelys yniphora*) through a captive
573 breeding program (112).

574 Other notable successes have come from Madagascar's "biodiversity conservation boom", which
575 started in the 1980s, including a growth in the number of students pursuing university-level
576 education in environmental sciences, biodiversity conservation and management, and related
577 fields, at both public and private universities. The result is an increasingly robust national capacity
578 for the conservation and management of biodiversity that extends to international conservation
579 organizations, which have been able to actively recruit Malagasy professionals to the highest
580 administrative and executive positions. Going beyond this, the gap in scientific leadership that
581 underpins conservation evidence is being incrementally filled by Malagasy biodiversity scientists.
582 Researchers from outside Madagascar are increasingly collaborating with Malagasy researchers
583 for mutual benefit. The requirement for international collaborators to provide financial and
584 technical support for Malagasy researchers and their research infrastructure via collaboration
585 protocols, set out in the national strategy for scientific research in Madagascar (113), reinforces
586 the importance of this.

587 As in many low-income countries, insufficient public funding means that the number of Malagasy
588 professionals is still insufficient to serve the country's needs, there are relatively few PhD positions
589 available to students, and those that are trained at higher levels often move away from academia
590 and into the private sector. Access to up-to-date biodiversity data has also been a limiting factor
591 (15). A further challenge is how to successfully engage multiple parts of society in conservation.
592 Efforts that are genuinely socially integrated have been shown to produce more effective and
593 resilient practices, policies, and decision-making, especially in the face of unstable environmental,
594 political, and health situations (114). The Madagascar Fauna and Flora Group, the Lemur
595 Conservation Foundation, Durrell Wildlife Conservation Trust, The Peregrine Fund Madagascar,
596 Madagascar Biodiversity Center, and Madagasikara Voakajy, as well as the work of the Royal
597 Botanic Gardens, Kew, and Missouri Botanical Garden, are all examples of successful
598 collaborations involving researchers, conservation partners and local communities to protect
599 biodiversity and empower local people.

600 *The future of biodiversity in Madagascar*

601 Meeting the Convention on Biological Diversity's Post-2020 Global Biodiversity Framework
602 2030 targets and milestones and achieving the 2050 goals (115) will be challenging – in
603 Madagascar and globally. Evaluating successes and failures over previous decades and learning
604 from these to prioritize effective conservation investment will be particularly important. To
605 embrace diverse views and promote inclusivity in the identification of future directions, we
606 discussed our results and current literature among our co-authors and consulted with Malagasy and
607 external researchers, conservation leaders, and politicians, to arrive at five main opportunities for
608 the future, which we now present.

609 1) Investment in conservation and restoration must be based on evidence, effectiveness, and future
610 challenges. Since the 1980s, billions of US dollars from international donors and conservation
611 organizations, in cooperation with the Malagasy government, have been dedicated to protecting
612 the country's biodiversity and creating today's network of PAs (107, 116). However, the
613 effectiveness of many interventions is poorly understood because impact evaluations are absent or
614 lacking rigor. Evaluating the effectiveness of conservation activities is challenging, but the subject
615 of increasingly sophisticated research efforts (76, 78, 117). Nevertheless, it is imperative
616 investments reinforce evidence-based and regularly evaluated interventions, requiring greater
617 collaboration and co-design between local communities, regional and national authorities,
618 researchers, the private sector, and other stakeholders. A particular opportunity is to frame these
619 evaluations around community-based conservation interventions that address challenges faced by
620 people and nature in unison. For example, nature-based solutions (118) for diversified, locally
621 adapted and sustainable agriculture can help address livelihood needs, while more efficient stoves
622 can substantially decrease the demand on charcoal from native forests for cooking and heating,
623 and further may reduce the health hazards of smoke inhalation. Such initiatives increase food and
624 energy security (119) while providing resilience to climate stochasticity (120). Similarly,
625 coordinated, community-based fire management and awareness raising can be used to help
626 mitigate risk to fire-sensitive forests. On-site management is especially important for fire
627 mitigation, as a study during the COVID 19 pandemic has shown (121). Fire management also
628 presents the opportunity to mitigate the impact of exotic species by targeting the removal of
629 flammable invasives (e.g., *Pinus*), and guide appropriate tree-planting initiatives to avoid fire-
630 prone plantations near areas of particular biological importance. Such measures can improve the
631 quality of grazing land for livestock, while reducing carbon emissions from fire and helping to
632 protect biodiverse habitats.

633

634 2) Expanded biodiversity monitoring is key to safeguarding Madagascar's most valuable natural
635 assets. Existing biodiversity data are sufficient to characterize major conservation challenges and
636 robustly support the orientation of conservation efforts in Madagascar. Calling for the collection
637 of additional data risks delivering diminished returns on investment for conservation planning
638 (122). Nevertheless, from collating the information for this review, we acknowledge a clear need
639 to address gaps in understudied ecosystems, taxa, and genetically distinct populations, noting that
640 many newly described species are already threatened (123) and in need of immediate protection.
641 Monitoring is also crucial for the detection of new non-native and potentially invasive species, as
642 well as providing important data for the management of those that have already taken hold.
643 Increasing connections with international trading partners without concurrent improvements in
644 capacity for biosecurity increases Madagascar's vulnerability to such species (124), and strategies
645 to monitor and mitigate these risks while delivering near-term benefits are needed.

646 Although there are initiatives that provide broad overviews of conservation effectiveness (e.g.
647 (117)), many conservation interventions lack impact evaluations, in part due to a lack of robust,
648 long-term monitoring data for biodiversity and social outcomes. The major gap is a lack of capacity
649 for robust biodiversity monitoring. An example of the increasing value of data and coherency in
650 conservation efforts is the development of the Madagascar Protected Areas website (125), which
651 consolidates much of the information about Madagascar's extensive network of PAs. But as with
652 many initiatives, the key is in long-term financing and maintenance of these portals and ensuring
653 that data flows freely and openly to similar, global initiatives like Protected Planet (73).

654 Biological monitoring needs to be based on consistent, repeatable methodologies, with shared data.
655 This information provides the science-based evidence needed to leverage international funding

656 and government policy support. Monitoring is one area where new technologies will play a key
657 role, such as through the increasing availability of near real-time satellite images and small and
658 cost-effective unmanned aerial vehicles, which can increase visual access to remote areas (126).
659 Similarly, DNA-based biodiversity surveys, including environmental sampling, can greatly
660 improve the speed of site-inventories and identification of unknown and understudied taxa.
661 Advances in monitoring must be delivered with improved and centralized management. This
662 should include open-source and transdisciplinary data on biodiversity, social and conservation
663 governance and performance. These data should be in formats that are accessible and useful to
664 practitioners, to identify relevant baselines, and support evidence-based decisions for conservation
665 and restoration.

666 3) Improving the effectiveness of existing PAs is more important than creating new ones.
667 Madagascar has an extensive, evidence-based, and highly representative network of terrestrial PAs
668 (Fig. 3, 4). Madagascar's existing PAs already include at least partial ranges of a substantial
669 proportion of Malagasy taxa, including most Malagasy EDGE species. Focusing on improving
670 their quality and effectiveness will likely lead to positive biodiversity outcomes (127), further
671 increasing the already measurable impact that PAs have had on biodiversity. By strengthening
672 PAs, biodiversity can be conserved across ecosystem, species, and genetic levels, all of which are
673 integral in long-term conservation, as discussed above. Investment in restoration of degraded areas
674 within and beyond the existing network (see Opportunity 4 below) will provide multiple benefits
675 for biodiversity and people. This could help increase the resilience of habitats to future drivers of
676 biodiversity loss including climate change, while increasing potential ranges of many species in
677 parallel. Demonstrating the benefits of strengthened PAs to people is a likely prerequisite for
678 societal support to maintain and improve upon the existing network, while mitigating risk of future

679 downgrading, downsizing, or degazettement (legal removal of conservation status) (128).
680 Financial benefits that come with strengthened PAs must be distributed appropriately and
681 equitably within the country's political and social contexts, with the full inclusion of local
682 communities at all stages (127, 129).

683

684 4) Conservation and restoration should not focus solely on the PA network. Madagascar's PAs are
685 islands of natural capital in a landscape of degraded natural resources (130) and therefore provide
686 vital resources for communities living adjacent to them. Traditional "fortress conservation"
687 – seeking to protect areas by limiting access – is therefore both undesirable and unlikely to be
688 effective. To further reduce the detrimental human impacts that exist in all PAs (107) (Table S8
689 (9)), we argue for strategies to enhance the natural capital of the surrounding landscapes, to reduce
690 pressure on PAs as providers of basic resources, and to increase buffer zones for the species that
691 live in and around them. This could include increasing ecosystem provision, such as productive
692 soils, food, fibers, and other materials and services such as water flow regulation and carbon
693 capture. Such measures would serve to address some of the largest threats to species, including the
694 expansion of agriculture and overexploitation (Fig. 2).

695 In particular, ecological restoration could benefit people and biodiversity, particularly when
696 targeted to the 89.6% of the country that is not protected. It offers potential to provide new
697 livelihood opportunities that are far from, and independent of, the resources within PAs, further
698 reducing pressure on the system (131). Importantly, restoration should not only target those
699 ecosystems that traditionally receive the most conservation attention because they hold the greatest
700 biodiversity, for example forests. Other vegetation types such as grasslands, where most
701 agriculture takes place, are equally vital. Restoration should be carried out following best practice

702 and in places where people will benefit most, not necessarily only adjacent to PAs. Further,
703 restoration should include maximizing biodiversity recovery to meet multiple goals, using resilient
704 species, and working together with local communities (49, 132).

705 For the species and their inherent genetic diversity not covered by the PA network, particularly
706 those that are challenging to conserve, such as freshwater fishes and palms, *ex situ* conservation in
707 zoological and botanical gardens is a vital tool to support conservation and restoration. For plants,
708 efforts should especially focus on the 32.3% of plant species that fall outside of the PA network,
709 and the species that have cultural or economic value for people (e.g., crop wild relatives).
710 Promoting biobanking for animals and intensifying it for seeds, spores, and fungi will not only
711 support conservation but also contribute material and knowledge to restoration and research (88).

712

713 5) Conservation actions must address the root causes of biodiversity loss. Our analysis showed
714 that the most frequently listed threats to Madagascar's biodiversity come from overexploitation
715 and agriculture, predominantly a result of forest loss and potentially tied to increases in small-scale
716 anthropogenic fire in forests (Fig. 4A, B; see also (43)), significantly affecting humid forest areas
717 in the east and dry forest and spiny forest in the west (Fig. 4C, D). This trend is likely to continue
718 unless the root causes of this forest loss are addressed. Conservationists and their funders must
719 recognize that food, social security, health, and well-being are the utmost priorities for rural
720 communities, and that PAs will always be vulnerable when surrounded by impoverished people
721 living in landscapes with eroded natural capital (133). Politicians and economists must recognize
722 that sustainable and equitable development in Madagascar is inextricably linked to, and dependent
723 on, the maintenance of ecosystem function and the goods and services they provide. Initiatives that
724 address these issues by working with local communities to identify tailored solutions in health,

725 education, and green entrepreneurship are increasingly successful and should be expanded, but
726 generally lack data and evidence from monitoring (see Opportunity 2). Promising approaches
727 include voluntary savings and loans; inclusive, sustainable agricultural development schemes that
728 promote stable land ownership and build – rather than destroy – natural capital and the ecosystem
729 services it provides; implementation of conservation interventions, including research and
730 monitoring; and PA management that maximizes local employment (107, 132). Such efforts will
731 facilitate improved livelihoods for many, while reducing pressure on the PAs themselves, bringing
732 tangible benefits to communities, and contributing to sustainable management (107, 134).

733

734 **Conclusions**

735 The alarming status of Madagascar's biodiversity is the result of multifaceted, unsustainable
736 practices including historic and contemporary exploitation. In the eyes of much of the world,
737 Madagascar's biodiversity is a unique global asset that needs “saving”; in the daily lives of many
738 of the Malagasy people, it is a rapidly diminishing source of the most basic needs for subsistence.
739 Achieving a sustainable future that benefits people and biodiversity is possible by building on, and
740 expanding, integrated, inclusive conservation efforts. Biodiversity is the greatest opportunity and
741 most valuable asset for Madagascar's future development.

742

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1448 Materials and Methods

1449 Figure S1

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