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Title: Impact of bisphenol A (BPA) on early embryo development in the marine bivalve *Mytilus galloprovincialis*: effects on gene transcription

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Abstract: Bisphenol A (BPA), a monomer used in plastic manufacturing, is weakly estrogenic and a potential endocrine disruptor in mammals. Although it degrades quickly, it is pseudo-persistent in the environment because of continual inputs, with reported concentrations in aquatic environments between 0.0005 µg/L and 12 µg/L. BPA represents a potential concern for aquatic ecosystems, as shown by its reproductive and developmental effects in aquatic vertebrates.

In invertebrates, endocrine-related effects of BPA were observed in different species and experimental conditions, with often conflicting results, indicating that the sensitivity to this compound can vary considerably among related taxa. In the marine bivalve *Mytilus galloprovincialis* BPA was recently shown to affect early development at environmental concentrations. In this work, the possible effects of BPA on mussel embryos were investigated at the molecular level by evaluating transcription of 13 genes, selected on the basis of their biological functions in adult mussels. Gene expression was first evaluated in trocophorae and D-veligers (24 and 48 h post fertilization) grown in physiological conditions, in comparison with unfertilized eggs. Basal expressions showed a general up-regulation during development, with distinct transcript levels in trocophorae and D-veligers. Exposure of fertilized eggs to BPA (10 µg/L) induced a general upregulation at 24 h pf, followed by down regulation at 48 h pf. *Mytilus* Estrogen Receptors serotonin receptor and genes involved in biomineralization (Carbonic Anhydrase and Extrapallial Protein) were the most affected by BPA exposure. At 48 h pf, changes in gene expression were associated with irregularities in shell formation, as shown by scanning electron microscopy (SEM), indicating that the formation of the first shelled embryo, a key step in bivalve development, represents a sensitive target for BPA. Similar results were obtained with the natural estrogen 17β-estradiol. The results demonstrate that estrogenic chemicals can affect *Mytilus* early development through dysregulation of gene transcription.

DEPARTMENT OF BIOLOGICAL, GEOLOGICAL,
AND ENVIRONMENTAL SCIENCESJuly 4th, 2016

Dear Editor,

I send you the manuscript “*Impact of bisphenol A (BPA) on early embryo development in the marine bivalve *Mytilus galloprovincialis*: effects on gene transcription*” by Teresa Balbi, Silvia Franzellitti, Rita Fabbri, Michele Montagna, Elena Fabbri, Laura Canesi, to be considered for publication in *Environmental Pollution*.

This work deals with the potential concern for endocrine disrupting chemicals (EDCs), and in particular for the estrogenic chemical Bisphenol A (BPA), in aquatic ecosystems. Although BPA is considered as one of the most common EDC that can affect reproduction and development in vertebrates, little information is available on the developmental effects of this compound in aquatic invertebrate species. The results here reported demonstrate that in the model marine invertebrate, the bivalve *Mytilus*, environmental concentrations of BPA significantly affect early embryo development through dysregulation of gene transcription. Moreover, the formation of the first calcified embryo, a key step in bivalve development, is showed as a sensitive target for BPA as well as for the natural estrogen 17 β -estradiol (E₂).

To the best of our knowledge, this study represents the first investigation on transcriptional changes occurring during early development of *M. galloprovincialis*, as well as the first indication on the molecular mechanisms of estrogenic compounds impacts in embryos of marine bivalves.

The results underline how exposure during early life stages can help identifying potential effects and modes of action of potential EDCs in aquatic invertebrates. Moreover, since calcifying larvae of marine species are particularly vulnerable to abiotic stressors, studies on the effects of potential EDCs on early life stages of sensitive species may contribute to the development of future multiple stressor studies in a global change scenario.

I do hope the manuscript meets the requirements for publication on *Environmental Pollution*.

Looking forward to hearing from you,

Sincerely yours

Prof. Elena Fabbri

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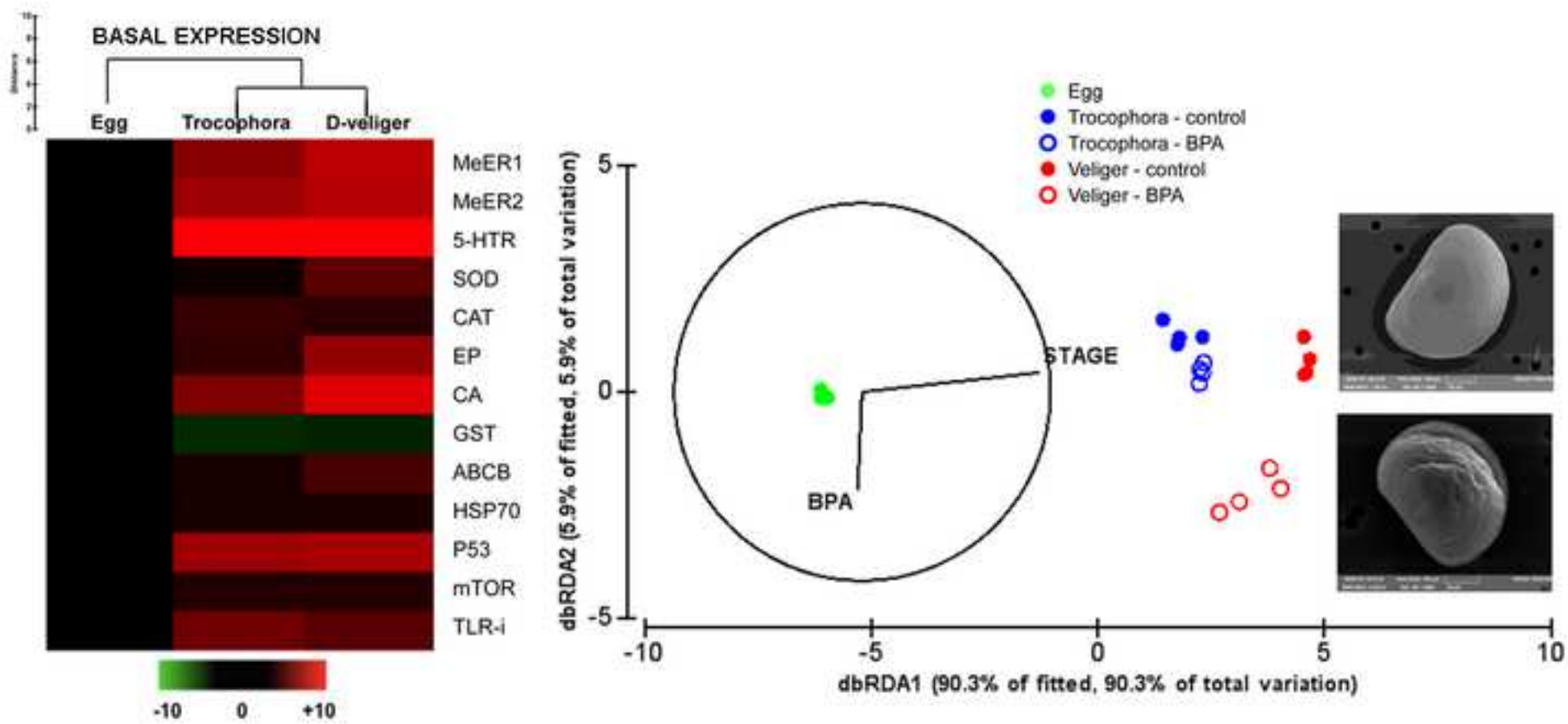
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Highlights

- The effects of BPA on *Mytilus* early embryo development were investigated
- Changes in transcription of 13 genes were observed at 24 and 48 h pf
- At 48 h pf, gene downregulation was associated with alterations in shell formation
- Similar results were obtained with the natural estrogen 17 β -estradiol
- Estrogenic chemicals affect bivalve embryos by dysregulation of gene transcription

1 **Impact of bisphenol A (BPA) on early embryo development in the marine bivalve *Mytilus***
2 ***galloprovincialis*: effects on gene transcription**

3

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24 **Abstract.**

25 Bisphenol A (BPA), a monomer used in plastic manufacturing, is weakly estrogenic and a potential
26 endocrine disruptor in mammals. Although it degrades quickly, it is pseudo-persistent in the
27 environment because of continual inputs, with reported concentrations in aquatic environments
28 between 0.0005 µg/L and 12 µg/L. BPA represents a potential concern for aquatic ecosystems, as
29 shown by its reproductive and developmental effects in aquatic vertebrates.

30 In invertebrates, endocrine-related effects of BPA were observed in different species and
31 experimental conditions, with often conflicting results, indicating that the sensitivity to this
32 compound can vary considerably among related taxa. In the marine bivalve *Mytilus*
33 *galloprovincialis* BPA was recently shown to affect early development at environmental
34 concentrations. In this work, the possible effects of BPA on mussel embryos were investigated at
35 the molecular level by evaluating transcription of 13 genes, selected on the basis of their biological
36 functions in adult mussels. Gene expression was first evaluated in trocophorae and D-veligers (24
37 and 48 h post fertilization) grown in physiological conditions, in comparison with unfertilized eggs.
38 Basal expressions showed a general up-regulation during development, with distinct transcript
39 levels in trocophorae and D-veligers. Exposure of fertilized eggs to BPA (10 µg/L) induced a
40 general upregulation at 24 h pf, followed by down regulation at 48 h pf. *Mytilus* Estrogen Receptors
41 serotonin receptor and genes involved in biomineralization (Carbonic Anydrase and Extrapallial
42 Protein) were the most affected by BPA exposure. At 48 h pf, changes in gene expression were
43 associated with irregularities in shell formation, as shown by scanning electron microscopy (SEM),
44 indicating that the formation of the first shelled embryo, a key step in bivalve development,
45 represents a sensitive target for BPA. Similar results were obtained with the natural estrogen 17β-
46 estradiol. The results demonstrate that estrogenic chemicals can affect *Mytilus* early development
47 through dysregulation of gene transcription.

48

49 **Keywords:** endocrine disrupters; bisphenol; marine mussel; larvae; gene transcription; shell

50 formation

51

52 **Capsule:** *Bisphenol A effects on larval stages of Mytilus galloprovincialis*

53

54 **1. Introduction**

55 Aquatic environments are the ultimate reservoirs for many environmental anthropogenic
56 chemicals, including those that can interfere with the functions of natural hormones. These
57 compounds, defined as endocrine disrupting chemicals (EDCs) can adversely impact reproduction
58 and development in humans and wildlife (Zoeller et al., 2012). Fish and other aquatic organisms
59 often have the greatest exposures to such chemicals during critical periods of development or even
60 their entire life cycles.

61 Bisphenol A (BPA), generally considered one of the most common EDC (Rubin, 2011; Vandenberg
62 et al., 2013), is a monomer used in the manufacture of polycarbonate plastics and in a variety of
63 plastic products, including food and beverage packaging, flame retardants, adhesives, building
64 materials, electronic components, and paper coatings (EFSA FIP, 2015). Due to its widespread
65 applications, BPA has become one of the highest production-volume chemicals (5.5 million tons
66 per year) in the last decades (Alonso-Magdalena et al., 2015; Flint et al., 2012). Although it
67 degrades quickly, it is pseudo-persistent in the environment because of continual inputs, leading to
68 reported concentrations in aquatic environments between 0.0005 µg/L and 12 µg/L (Flint et al.,
69 2012). Available data underline that environmental concentrations of BPA may represent a potential
70 concern for aquatic ecosystems (Tišler et al., 2016). In aquatic vertebrates, BPA causes
71 developmental and reproductive effects, as well disturbances in immune function and metabolism
72 (Canesi and Fabbri, 2015). In invertebrates, endocrine-related effects of BPA have been observed in
73 different species and experimental conditions, with often conflicting results. Available data
74 underline how some invertebrates appear to be quite sensitive to BPA (freshwater and marine
75 mollusks, insect larvae, marine copepods in particular), with effects documented at environmentally
76 relevant concentrations (reviewed in Canesi and Fabbri, 2015; Flint et al., 2012).

77 Research on the impact of BPA in model aquatic organisms has major regulatory
78 implications towards the ecosystem health. In this light, studies involving exposure during early life

79 stages in invertebrates, which can be highly sensitive to environmental perturbations, would greatly
80 help identifying potential adverse effects of BPA as an EDC, in analogy with what reported in
81 vertebrates (Hamlin and Guillette, 2011; Santangeli et al., 2016; Tse et al., 2013).

82 We have recently shown that in the marine bivalve *Mytilus galloprovincialis* exposure of
83 fertilized eggs to BPA at environmentally relevant concentrations affects the formation of fully
84 developed D-larvae at 48 h post fertilization (pf) (Fabbri et al., 2014). Larval development in
85 mussels chronologically unfolds following a mechanical remodeling of the form and behavior,
86 which requires controlled mechanisms engaged in tissue distribution and organ system
87 differentiation (Ackerman et al., 1994). In *Mytilus*, after external oocyte fertilization embryos
88 undergo fast cell divisions that lead to their segmentation, and differentiate into motile forms
89 (trochophorae, 24 h pf), with successive development of various shelled forms (D-veligers) from 48
90 h pf (Marin et al., 2008). Although a number of genetic regulators during late-stages of larval
91 development have been recently identified (Bassim et al., 2014), no information on the molecular
92 mechanisms underlying the physiological processes involved in the early critical stages, such as the
93 formation of the first shelled embryo, is available.

94 In this work, the effects of BPA on gene transcription in early embryo development of *M.*
95 *galloprovincialis* were investigated. Fertilized eggs were exposed to BPA 1 µg/L (< EC₅₀, LOEC
96 10x) and 10 µg/L (> EC₅₀, LOEC 100x) as previously described (Fabbri et al., 2014), and the
97 differential expression of 13 gene products related to known biological functions in adult mussels
98 was assessed. These include neuroendocrine signaling, antioxidant defense, biomineralization,
99 stress response, apoptosis, autophagy, and immune response. Transcriptional profiles were
100 evaluated in trochophorae and D-veligers grown under physiological conditions or in the presence of
101 BPA at environmentally realistic concentrations (Corrales et al., 2015). Morphology of control and
102 BPA-exposed embryos was evaluated by scanning electron microscopy (SEM). The effects of BPA
103 on embryo development, gene expression and morphology were compared with those of the natural
104 estrogen 17β-estradiol (E₂).

105

106 **2. Methods**

107 *2.1. Animal holding and larval rearing*

108 Sexually mature mussels (*M. galloprovincialis* Lam.), purchased from an aquaculture farm
109 in the Ligurian Sea (La Spezia, Italy) between November 2014 and March 2014, were transferred to
110 the laboratory and acclimatized in static tanks containing aerated artificial sea water (ASTM, 2004),
111 pH 7.9 - 8.1, 36 ppt salinity (1 L/animal), at $16 \pm 1^\circ\text{C}$.

112 Gamete collection and fertilization were performed following the procedure described by
113 Fabbri et al. (2014). Details are reported in Fig S1 (Supporting Information Appendix 1) with the
114 description of the 96-microwell embryotoxicity assay performed to assess the effects of E₂.
115 Unfertilized eggs (about 24,000 eggs/mL) obtained from at least 6 female individuals were
116 collected by centrifugation at 400 xg for 10 min at 4°C, and the resulting pellet was frozen in liquid
117 nitrogen. Eggs were fertilized with an egg:sperm ratio 1:10 in polystyrene 6-well plates and a final
118 6 mL volume. After 30 min, fertilization success (n. fertilized eggs / n. total eggs × 100) was
119 verified by microscopical observation (>85%). At 30 min pf, BPA or E₂ were added to fertilized
120 eggs in each well from 1 mg/L concentrated stock solutions prepared in ethanol, and suitably
121 diluted to reach the final nominal concentrations of 1 and 10 µg/L for BPA, and 10 µg/L for E₂.
122 Control wells (negative controls) contained only ASW. Four replicates for each experimental
123 condition were made. In all cases, ethanol final concentration did not significantly affect the
124 biological endpoints analysed (data not shown). Concentrations of BPA and E₂ in test solutions
125 were checked by LC/MS (Table S1, Supporting Information Appendix 2).

126

127 *2.2. RNA extraction and cDNA preparation*

128 At 24 and 48 h pf larvae were collected by a nylon mesh (40 µm pore-filter) and washed with ASW.
129 Three wells for each condition were pooled in order to obtain approximately 7000

130 embryos/replicate. The larval suspension was centrifuged at 800 xg for 10 min at 4°C. Larval
131 pellets and unfertilized eggs were lysed in 1 mL of the TRI Reagent (Sigma Aldrich, Milan, Italy)
132 and total RNA was further extracted using the DirectZol kit (Zymo Research, Freiburg, Germany)
133 following manufacturer's instructions. RNA concentration and quality were verified using the Qubit
134 RNA assay (Thermo Fisher, Milan, Italy) and electrophoresis using a 1.2% agarose gel under
135 denaturing conditions. First strand cDNA for each sample was synthesized from 1 µg total RNA
136 according to Balbi et al. (Balbi et al., 2014).

137

138 2.3. qPCR

139 Transcript expressions were assessed by real-time PCR (qPCR). Primers pairs were as reported in
140 previous studies or were designed with Primer Express (Thermo Fisher, Milan, Italy) using
141 nucleotide sequences retrieved from the GeneBank database
142 (<https://www.ncbi.nlm.nih.gov/genbank/>) for *Mytilus galloprovincialis* (Table S2, Supporting
143 Information Appendix 2). qPCR reactions were performed in duplicate in a final volume of 10 µl
144 containing 5 µl iTaq universal master mix with ROX (BioRad Laboratories, Milan, Italy), 2 µl
145 diluted cDNA, and 0.2 µM specific primers (Table S2, Supporting Information Appendix 2). A
146 control lacking cDNA template (no-template) and a minus-reverse transcriptase (no-RT) control
147 were included in the qPCR analysis to determine the specificity of target cDNA amplification.
148 Amplifications were performed in a StepOne real time PCR system apparatus (Thermo Fisher,
149 Milan, Italy) using a standard “fast mode” thermal protocol. For each target mRNA, melting
150 curves and agarose gel electrophoresis were utilized to verify the specificity of the amplified
151 products and the absence of artifacts. The amplification efficiency of each primer pair was
152 calculated using a dilution series of cDNA (Table S2, Supporting Information Appendix 2). A
153 preliminary stability analysis of 6 established candidate reference transcripts (Cubero-Leon et al.,
154 2012; Franzellitti and Fabbri, 2013) was performed to achieve a robust normalization of qPCR data
155 in mussel larval stages. Detailed methods and results are reported in Supporting Information

156 Appendix 3. *HEL* and *EF- α 1* were selected as the best performing combination of reference gene
157 products for data normalization. Calculations of relative expression of target mRNAs was
158 performed by a comparative C_T method (Schmittgen and Livak, 2008) using the StepOne software
159 tool (Thermo Fisher, Milan, Italy). Data were reported as relative expression (fold change or \log_2 -
160 transformed fold changes according to the data ranges) with respect to unfertilized eggs (basal gene
161 expression across larval development) or control samples within each life stage (BPA or E_2
162 treatments).

163

164 2.4. *Scanning Electron Microscopy (SEM)*

165 At different times post-fertilization, depending on the experiment, control and treated embryos
166 grown in 96 microwells were fixed in 3% glutaraldehyde in ASW. After fixation, samples from 6
167 microwells were pooled, placed onto Whatman 22 μ m filters, dehydrated in an ascending series of
168 ethanol washes (50% - 80% - 90% - 100%) and air-dried. Then samples were sputter-coated with
169 gold, and observed at 20 kV with a Vega3 - Tescan scanning electron microscope.

170

171 2.5. *Statistics*

172 Statistical comparisons between unfertilized eggs vs trocophora/D-veligers, or controls vs
173 BPA/ E_2 treated larvae (within each developmental stage) were performed using the non-parametric
174 one-way ANOVA (Kruskal-Wallis test) followed by the Mann-Whitney U-test ($P < 0.05$), after
175 deviations from parametric ANOVA assumptions being verified (Normality: Shapiro-Wilk's test;
176 equal variance: F-test). Data analyses were performed using GraphPad Prism 5 software (GraphPad
177 Inc.).

178 Additional statistical approaches were utilized for analysis of qPCR data to evaluate:

179 i) Basal gene expression across larval development. Ordination analysis (carried out by Principal
180 Coordinates, PCO) and clustering of basal transcript expressions across the larval stages were

181 performed using the PRIMER v6 software (PRIMER-E Ltd, UK), which employed \log_2 -
182 transformed fold change variations of the target transcripts to calculate similarity matrices based on
183 the Euclidean distance (999 permutations). The similarity matrices were also used to perform a
184 canonical analysis of principal coordinates ordination (CAP) to attribute the specific transcripts
185 responsible for condition differences (i.e. differences between larval stages). A heatmap describing
186 the overall transcriptional profile was generated using the Gene Cluster software v2.0 (Eisen et al.,
187 1998) and the TreeView software for heatmap visualization. Similarity was measured by Euclidean
188 distance.

189 ii) Effects of BPA or E₂ treatments on target transcript expressions in different larval stages.
190 Relative expression data from BPA or E₂ treatments were submitted to permutation multivariate
191 analysis of variance (PERMANOVA) using the PERMANOVA+ add-on in PRIMER v6 (Anderson
192 et al., 2008). Log-transformed copy number variations of the target transcripts were used to
193 calculate similarity matrices based on the Euclidean distance (999 permutations). Factors
194 considered were “developmental stage” and “treatment” (BPA or E₂). Pseudo-F values in the
195 PERMANOVA main tests were evaluated in terms of significance (Anderson et al., 2008). When
196 the main test revealed statistical differences ($P < 0.05$), PERMANOVA pairwise comparisons were
197 carried out (using the Euclidean distance matrix and 999 permutations). Distance-based redundancy
198 linear modeling (DISTLM) followed by a redundancy analysis (dbRDA) in PRIMER with
199 PERMANOVA+ extension was also performed on BPA or E₂ treatments to examine the
200 relationship between the multivariate dataset (i.e. the suite of transcripts assayed and their
201 expression levels) and the predictor variables (life stage and BPA/E₂ treatment). Numerical metric
202 for life stage progression was indicated by the post fertilization time (0: unfertilized egg; 24 h:
203 trocophora; 48 h: D-veliger). Treatment was indicated by the nominal concentrations of exposure
204 (BPA/E₂: 0 and 10 $\mu\text{g/L}$). DISTLM used the BEST selection procedure and adjusted R^2 selection
205 criteria.

206

207 3. Results

208 3.1. Transcriptional profiles during early larval development

209 Transcription of selected genes related to neuroendocrine signaling (estrogen receptors *MeER1*,
210 *MeER2*, serotonin receptor, *5-HTR*), antioxidant defense (superoxide dismutase, *SOD*, catalase,
211 *CAT*), biomineralization (extrapallial protein, *EP*, carbonic anhydrase, *CA*), detoxification response
212 (glutathione-s-transferase, *GST-π*, ABC transporter p-glycoprotein, *ABCB*), stress-response (70kDa
213 heat shock proteins, *HSP70*), apoptosis (*p53*), autophagy (serine/threonine-protein kinase mTOR,
214 *mTor*), immune response (Toll-like receptor, *TLR-i*), was evaluated at both 24 and 48 h pf in
215 embryos grown under normal physiological conditions. Data on relative expression of each
216 transcript across larval development (fold changes with respect to unfertilized eggs) are reported in
217 Fig S2 (Supporting Information Appendix 1). To help comparing overall transcript levels among
218 datasets, fold change variations (\log_2 -transformed) of target transcripts were subjected to a cluster
219 analysis and to an ordination analysis using a Principal Coordinate Ordination (PCO) extraction
220 procedure (Fig 1). Basal transcript expressions showed substantial time-course variations across
221 developmental stages with general and significant up-regulation (Fig 1A; Fig S2, Supporting
222 Information Appendix 1). The highest mRNA levels were reported for *MeERs*, *5-HTR*, *CA* and *p53*
223 (Fig S2, Supporting Information Appendix 1). On the contrary, down-regulation was observed for
224 *GST-π* in both larval stages (Fig1A; Fig S2, Supporting Information Appendix 1). The PCO
225 analysis (Fig 1B) detected two principal coordinates (PCO1 and PCO2) explaining about 97% of
226 total variation (89.8% and 5.8%, respectively). Samples from unfertilized eggs formed one group
227 with evident separation from those belonging to trocophora and D-veliger stages, which also
228 formed distinct clusters. The canonical analysis of principal coordinates ordination (CAP analysis)
229 attributed much of the differences between these latter two stages to *5-HTR*, *EP*, *CA*, *MeER1*, *p53*,
230 *CAT*, *SOD*, *ABCB* profiles (Fig. 1C).

231

232 3.2. *Effects of BPA on gene transcription*

233 Fertilized eggs were exposed to BPA 10 $\mu\text{g/L}$ ($> \text{EC}_{50}$, LOEC 100x), as previously described
234 (Fabbri et al., 2014) and gene transcription was evaluated at different times pf (24 h and 48 h) in
235 comparison with untreated samples (Fig 2). Distinct changes in gene transcription at different stages
236 were observed. In trocophorae, all transcripts were significantly up-regulated, with the exception of
237 *HSP70*, which showed down-regulation (Fig. 2A). A maximum 3.5-fold increase was observed for
238 the *EP* transcript. In contrast, in D-veligers 10 out of 13 transcripts were significantly down-
239 regulated, with mRNA levels ranging from -30% to -54% with respect to controls (Fig.2A). Only
240 *SOD* and *CA* were significantly up-regulated, with *CA* showing up to 4.2-fold increase with respect
241 to controls (Fig. 2A). To obtain further details on concentration-dependent effects of BPA on both
242 larvae developmental stages, parallel experiments were performed exposing fertilized eggs to BPA
243 1 $\mu\text{g/L}$ ($< \text{EC}_{50}$, LOEC 10x) (Fabbri et al., 2014). The results, reported in Fig S3 (Supporting
244 Information Appendix 1), showed a small though significant up-regulation of most transcripts (from
245 + 35 % to + 55 %, depending on the transcript and the developmental stage). Significant down-
246 regulation was observed for *SOD* and *EP* in trocophorae, and for *SOD* and *CA* in D-veligers (from -
247 30 % to - 60 %).

248 Distance-based linear model (DISTLM) analysis revealed that expression profiles were
249 strongly dependent on embryo development (time from pf) that explained about 90 % total
250 variation. Nevertheless, BPA treatment accounted for 6% of total variation, which mostly explained
251 the observed changes in transcript expressions at 48 h pf (Fig. 2B). Results from PERMANOVA
252 and permutation t-test analyses demonstrated that the effects of BPA were statistically significant in
253 both larval stages ($P < 0.05$; Table S3, Supporting Information Appendix 2). PERMANOVA
254 analysis also showed a significant interaction between the two factors (developmental stage and
255 BPA treatment; $P < 0.05$; Table S3, Supporting Information Appendix 2).

256

257 3.3. SEM analysis

258 Larval development was observed by Scanning electron microscopy (SEM) first in larvae grown
259 under physiological conditions, and representative images are reported in **Fig 3a-f**. Fig. 3a and Fig.
260 3b show a trocophora at 24 h pf, and the detail of the ciliated epithelium, respectively. During the
261 trocophora stage, the onset of shell mineralization occurs (generally at 20 h pf). The process is
262 initiated by a specialized group of ectodermal cells, the shell field, that are responsible for
263 production of the first outer organic shell layer, the periostracum (Weiss et al., 2002). The saddle-
264 shaped shell field is well visible in Fig. 3c. From about 32 h pf, first shelled embryos could be
265 observed (Fig. 3d); at this stage, most shells were extremely fragile, and damaged by the sample
266 preparation procedure for SEM observations. Micrometric spherules of amorphous calcium
267 carbonate (ACC) could be observed both at the surface and margins of the valvae (Fig. 3e) typical
268 of initial shell formation (Cartwright et al., 2012; Weiss et al., 2002). Finally, at 48 h pf the D-
269 Veliger stage is reached (Fig. 3f), with shells showing straight hinge, symmetric valvae, and
270 uniform surface, indicating that formation of the calcified shell occurred simultaneously in the
271 whole shell field. In samples exposed to BPA (10 µg/L), no appreciable morphological changes
272 were observed at 24 h pf (Trocophora) (not shown). On the other hand, clear changes were
273 observed in the morphology of D-larvae at 48 h pf (Fig. 3g-i). Common shell alterations were rough
274 surface, thinner valvae, convex hinge, asymmetry and/or gap between valvae; moreover, many
275 samples showed the externalized velum (ciliated swimming and feeding organ), a feature typical of
276 the pre-veliger stage (Fig. 3i and insert to Fig. 3i).

277

278 3.4. Effects of 17β-estradiol on mussel embryo development and gene transcription

279 The effects of BPA were compared with those of the natural estrogen 17β-estradiol (E₂). Fig S1a
280 (Supporting Information Appendix 1) shows results from the 48-h embryotoxicity assay performed
281 utilizing E₂ in the same concentration range as that previously utilized for BPA (Fabbri et al., 2014).

282 E₂ induced a dose-dependent decrease in normal larval development that was significant from 0.1
283 µg/L, with a calculated EC₅₀ value of 6.904 µg/L (95% confidence interval of 3.284 - 14.511 µg/L).
284 Optical microscopy observations indicated that, as previously observed with BPA (Fabbri et al.,
285 2014), E₂ mainly induced a delay in development, with a prevalent increase in the percentage of
286 trocophorae/not fully developed D-veligers, and a variable percentage of malformed embryos at
287 different concentrations. At the highest concentration tested (1,000 µg/L) more than 90% embryos
288 were at the pre-veliger stage or showed gross shell malformations (not shown). At 10 µg/L E₂, SEM
289 observations showed that most common shell alterations induced by E₂ were hinge malformations,
290 asymmetric and thinned valvae and rough surface (Fig. S1b-d, Supporting Information Appendix
291 1).

292 The effects of E₂ (10 µg/L) on gene transcription are illustrated in Fig. 4. The results show
293 that most transcripts were significantly down-regulated both at 24 h and 48 h pf, to a different
294 extent depending on the type of gene and developmental stage. In particular, at 24 h pf E₂ induced a
295 general decrease in gene transcription, except for *5-HTR*, *ABCB*, *EP* and *CA*. Dramatic down-
296 regulation ($\leq 70\%$ with respect to controls) was observed for *MeERI*, *CAT*, *p53*, and *mTor*, and
297 almost complete inhibition resulted for *TLR-i* transcription. At 48 h pf, a significant down-
298 regulation was observed for all transcripts except for *HSP70* and *mTor*, whose mRNA levels were
299 unchanged, and for *ABCB*, that was up-regulated.

300 DISTLM analysis revealed that E₂ treatment accounted for 7.5 % of total variation, and
301 affected both trocophora and D-veliger stages (Fig. 4B). Results from PERMANOVA and
302 permutation t-test analyses demonstrated that the effects of E₂ were significant in both larvae stages
303 ($P < 0.05$; Table S4, Supporting Information Appendix 2). PERMANOVA analysis also showed a
304 significant interaction between the two factors (developmental stage and E₂ treatment; $P < 0.05$;
305 Table S4, Supporting Information Appendix 2).

306

307 4. Discussion

308 We have previously shown that BPA significantly affected 48 h embryo development in *M.*
309 *galloprovincialis*, with a LOEC and EC₅₀ of 0.1 µg/L and 3.68 µg/L, respectively (Fabbri et al.,
310 2014). Since these values fall within the concentration range of BPA found in coastal environments
311 (Flint et al., 2012), these data underlined the sensitivity of mussel embryos towards exposure to
312 environmental concentrations of BPA, and supported the hypothesis that this compound could act
313 as an EDC in marine bivalves. The results here obtained demonstrated that BPA affects gene
314 transcriptions in early mussel embryos, with effects depending on the BPA concentration, the target
315 transcript and the developmental stage. These data represent the first indication of molecular
316 mechanisms for BPA impacts on developmental stages in a marine bivalve.

317 Basal transcription of 13 genes involved in different biological processes was first evaluated
318 in larvae grown under physiological conditions at both 24 and 48 h pf, and relative expression was
319 quantified with respect to unfertilized eggs. Genes were selected on the basis of their recognized
320 physiological functions in *Mytilus spp.* and/or of their sensitivity to different environmental
321 chemicals, including BPA (i.e. neuroendocrine signaling, antioxidant defenses, biomineralization,
322 detoxification processes, stress response, apoptosis, autophagy, immunity). These pathways are
323 reported to provide potent defenses in embryos against common stressors, including changes in
324 temperature, hypoxia, and pollutants, to permit optimal development despite relevant environmental
325 challenges (Hamdoun and Epel, 2007). To the best of our knowledge, the results represent the first
326 data on basal mRNA expressions in early development of *M. galloprovincialis*. All transcripts were
327 significantly up-regulated in both trocophora and D-veliger stages with respect to unfertilized eggs,
328 except for *GST-π*, which resulted significantly down-regulated in both trocophore and D-veligers
329 compared to eggs, thus suggesting that early embryos rely on maternal GST, if any, for Phase II
330 biotransformation pathways. PCO and cluster analyses disclosed a further distinct clustering
331 between trocophore and D-veliger samples, consistent with the progressive differentiation of

332 different processes. Main contribution to the observed data ordination was provided by the
333 increased expressions of *5-HTR*, *EP*, *CA*, *MeER1*, *p53*, *CAT*, *SOD*, *ABCB* in D-veligers compared
334 to trocophorae, as suggested by CAP analysis. If on one hand the results obtained for *5-HTR*
335 confirm the early and progressive development of serotonergic transmission in *Mytilus spp.*
336 reported by previous immuno-histochemical investigations (Voronezhskaya et al., 2008), on the
337 other hand increased expressions of *SOD*, and to a lesser extent, of other transcripts with
338 cytoprotective functions (i.e. *ABCB*, *CAT*, *p53*), indicate a reinforcement of defenses in the first D-
339 shelled embryo stage.

340 Although a number of genes involved in biomineralization are being described in different
341 bivalve species (Bassim et al., 2015; Vendrami et al., 2016), those involved in the initial formation
342 of a first calcified shell (prodissoconch I), a key step in early development, have not been identified
343 yet. Carbonic anhydrase (*CA*) is known to regulate matrix mineralization by generating an acidic
344 environment (Clark et al., 2010). *Mytilus EP* (Extrapallial Protein), an acidic calcium binding
345 protein, regulates the production of different polymorphs of calcium carbonate (Yin et al., 2009).
346 Both *EP* and *CA* showed the highest expression in D-veligers, supporting the results from SEM
347 analyses, showing that the formation of a complete normal shell occurred only at 48 h pf, and
348 confirming the key physiological role of both transcripts in initial biomineralization (Medakovic,
349 2000; Yin et al., 2009, 2005).

350 Exposure of fertilized eggs to 1 µg/L BPA resulted in small but significant up-regulation of
351 most transcripts. However, down-regulation was observed for *SOD*, and for *EP* and *CA* in
352 trocophorae and D-veligers, respectively. More evident changes were observed at 10 µg/L. In these
353 conditions, distinct trends in gene transcription were observed at different post-fertilization periods,
354 with a general up-regulation in the trocophorae and down-regulation in D-veligers for all genes
355 tested, except for *CA* and *SOD*, whose mRNA levels were further increased at 48 h pf. Interestingly,
356 transcription of both *EP* and *CA* was strongly affected by BPA exposure at both developmental
357 stages.

358 The effects of BPA on mussel embryos were compared with those of the natural estrogen
359 17β -estradiol (E_2). E_2 significantly affected larval development in the 48 h embryotoxicity tests
360 ($EC_{50} = 6.904 \mu\text{g/L}$), with effects similar to those of BPA (Fabbri et al., 2014). Therefore, gene
361 transcription was evaluated after exposure to $10 \mu\text{g/L } E_2$. At 24 h pf, E_2 induced a general decrease
362 in gene transcription, in particular of genes involved in apoptosis, autophagy and immunity,
363 showing a distinct effect with respect to BPA. At 48 h pf, E_2 , like BPA, induced down-regulation of
364 most genes, including those involved in biomineralization. An exception to this trend was
365 represented by *ABCB*, whose transcription was unaffected in trocophorae, and significantly
366 increased in D-veligers.

367 At the concentration utilized for gene expression studies ($10 \mu\text{g/L}$) the main morphological
368 effects induced by BPA (Fabbri et al., 2014) and E_2 (this work) were represented by the presence of
369 immature, not fully developed D-shell larvae at 48 h pf. Similar results were obtained when BPA or
370 E_2 were added at 24 h pf, instead of at 30 min post oocyte fertilization (not shown), indicating that
371 the transition from the trocophora stage to the first D-shelled larva represents a critical
372 developmental step affected by both compounds. Accordingly, a common effect of BPA and E_2
373 observed at 48 h pf was the general down-regulation of gene expression and alterations in shell
374 formation, as indicated by SEM analysis.

375 Overall, the results demonstrate that both the estrogenic chemical BPA and the natural
376 estrogen E_2 affect *Mytilus* early development through dysregulation of gene transcription. In the
377 marine gastropod *Haliotis diversicolor*, BPA affected different stages of larval development, with
378 an EC_{50} of $1.02 \mu\text{g/L}$ at 96 h (completion of metamorphosis) (Liu et al., 2011). Furthermore,
379 proteomic analysis indicated that both chemicals interfered with different physiological pathways,
380 including energy and substance metabolism, cell signaling, formation of cytoskeleton and cilium,
381 immune and stress responses at the same time, leading to the failure of metamorphosis (Liu et al.,
382 2011).

383 The results of the present work underline that in *Mytilus MeERs* and *5-HTR* represent
384 significant targets for both BPA and E₂ during early embryo development. Similar results on *MeER*
385 expression were observed in the hepatopancreas of adult mussels exposed to BPA (Canesi et al.,
386 2011, 2007, 2006). Taken together the results support a common neuroendocrine-related effect of
387 these compounds in mussels. However, from an accurate analysis on the literature on the effects of
388 steroids and EDCs Scott (2013) concluded that there is no indisputable bioassay evidence that
389 vertebrate sex steroids (or their agonist/antagonists) have endocrine or reproductive roles in
390 mollusks. Estrogen receptors (ERs) generally show constitutive transcriptional expressions in
391 mollusks, and are not activated by either natural estrogens or estrogenic compounds, this rising a
392 debate on their physiological function (Bridgham et al., 2014). In this respect, molluscan ERs seem
393 to resemble mammalian orphan estrogen-related receptors (ERRs), that play multiple and
394 overlapping roles in cell growth, differentiation and metabolism (Huss et al., 2015). However,
395 cloning and phylogenetic analysis of estrogen-like receptors in *Mytilus edulis* and *M.*
396 *galloprovincialis* revealed the Me/MgER2 is an ER, whereas the Me/MgER1 is an ERR (Nagasawa
397 et al., 2015). Our results show that in physiological conditions both transcripts encoding the MeER1
398 and MeER2 isoforms were among the most up-regulated across different developmental stages,
399 while these receptors were down-regulated by exposure to both BPA and E₂ in D-veligers. These
400 data indicate that *Mytilus* ERs may play a role in early development and represent significant targets
401 for estrogenic chemicals. Similar results were obtained for the *5-HTR* transcript, encoding a
402 putative type 1 serotonin receptor (Cubero-Leon et al., 2010; Franzellitti and Fabbri, 2013),
403 indicating that BPA and E₂ have a significant impact on early development of the serotonergic
404 signaling in mussels (Kreiling et al., 2001; Voronezhskaya et al., 2008).

405 In addition, the results underline how estrogenic compounds significantly affected
406 transcription of *EP* and *CA* involved in mussel biomineralization, with down-regulation of both
407 transcripts observed at 48 h pf. These data were supported by SEM observations, which clearly
408 showed similar irregularities in shell formation in both BPA and E₂ -treated embryos at 48 h pf.

409 These effects are in line with the effects of estrogenic chemicals on bone metabolism observed in
410 vertebrate systems (Agas et al., 2013). Overall, these represent the first data on the effects of EDCs
411 on early biomineralization processes in marine invertebrates. The effects of BPA on the shell
412 structure are under investigation. Calcifying larvae of marine species are particularly vulnerable to
413 abiotic stressors (Przeslawski et al., 2015). Interestingly, the shell malformations induced by BPA
414 and E₂ were similar to those observed in *M. galloprovincialis* embryos by inhibiting of the synthesis
415 of chitin, a key biological component that forms the framework for other macromolecules that guide
416 initial shell deposition (Schönitzer and Weiss, 2007; Weiss and Schönitzer, 2006), as well as by
417 exposure to high pCO₂ or low pH (Kadar et al., 2010; Kurihara et al., 2008). In this light, chitin
418 synthesis may represent an important target for the effects of both chemical pollutants and altered
419 environmental parameters in the developing embryo. Studies on the effects of individual chemicals
420 that represent potential EDCs on early life stages of sensitive marine species may contribute to the
421 development of future multiple stressor studies in a global change scenario.

422

423 **5. Conclusions**

424 The results of the present study show that environmental levels of estrogenic compounds may
425 adversely affect physiological processes involved in early embryo development of marine mussels
426 through dysregulation of transcription for genes involved in different key regulatory pathways

427 In particular, the results address the transition from the trocophora stage to the first D-shelled
428 larva as a very sensitive target for estrogenic compounds. The results indicate that, although
429 physiological processes in early embryos are well buffered for environmental challenges (Hamdoun
430 and Epel, 2007), anthropogenic stressors can overwhelm this intrinsic robustness and compromise
431 developmental processes.

432

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575

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579

580

581 **Figure legends**

582 **Fig. 1 – Transcriptional profiles of selected gene products during early larval development in**
583 **mussels. (A)** Treatment ranking according to the qPCR data describing transcript levels at different
584 developmental stages. Hierarchical clustering was performed using fold change variations (\log_2 -
585 transformed) of target transcripts at each developmental stage (In-detail data are reported in Fig. S2,
586 Supporting Information Appendix 1). Colors represent relative expression levels with respect to
587 unfertilized eggs. **(B)** Principal coordinates ordination (PCO) bi-plot of mRNA expression levels
588 with super-imposed cluster analysis by condition (Euclidean distance matrix; 999 permutations).
589 **(C)** Canonical analysis of principal coordinates ordination (CAP analysis) illustrating transcripts
590 whose expression patterns are likely to discriminate the differences between life stages. Transcripts
591 to be displayed in the CAP bi-plot were filtered according to their Spearman correlation's
592 coefficients ($r > 0.6$ with at least CAP1 or CAP2).

593

594 **Fig. 2 - Effects of BPA (10 $\mu\text{g/L}$) on gene transcription in *Mytilus* embryos at 24 h and 48 h**
595 **post fertilization times. (A)** Expression levels of target transcripts evaluated by qPCR. Data are
596 reported as mean \pm SD of the relative expression with respect to untreated samples within each life
597 stage (N = 4). *P < 0.05, **P < 0.01 according to the Mann-Whitney U test. **(B)** DISTLM modeling
598 with distance-based redundancy analysis (dbRDA) to explore the amount of the variation in gene
599 transcription to be attributed to BPA treatment (Euclidean Distance resemblance matrix, 999
600 permutations).

601

602 **Fig. 3 - Scanning electron microscopy (SEM) of *Mytilus* embryos.** a-f: representative images of
603 different stages of larvae grown under physiological conditions: a) trocophora at 24 h pf, and b)
604 detail of the ciliated epithelium; c) trocophora with super-imposed margins of the saddle-shaped
605 shell field; d) early D-veliger at 36 h pf, showing the presence of amorphous calcium carbonate
606 (ACC) spherules (e); f) normal D-veliger at 48 h pf, with straight hinge, symmetric valvae and

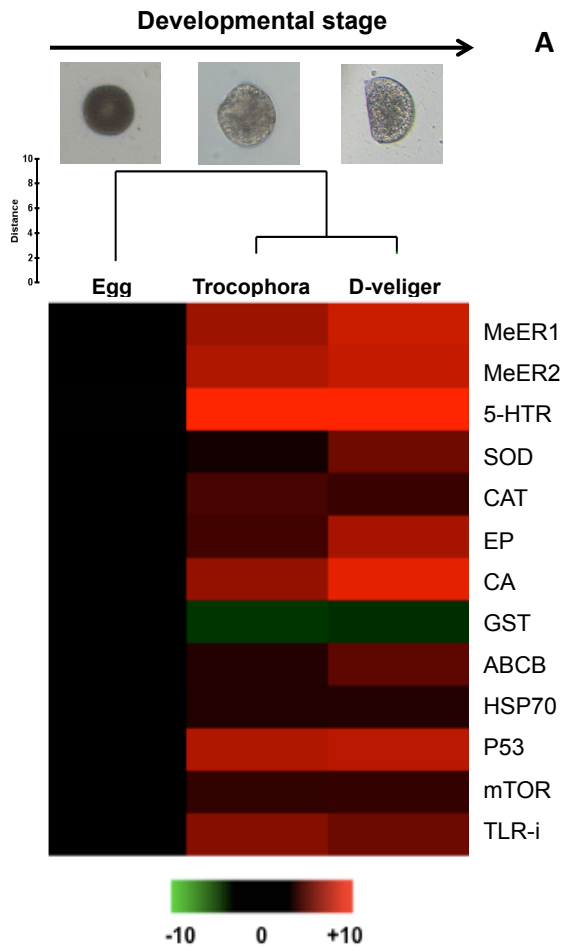
607 uniform surface. g-i: representative images of the effects of BPA (10 µg/L) of shell morphology in
608 D-veligers at 48 h pf, showing asymmetric valvae and rough surface (g), convex hinge and thinner
609 valvae (h), externalized velum (i).

610

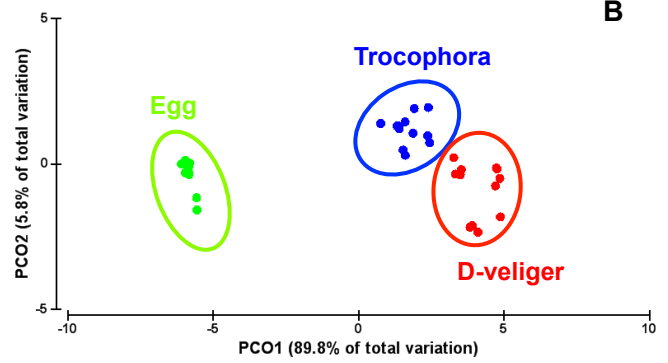
611 **Fig. 4 – Effects of E₂ (10 µg/L) on gene transcription in *Mytilus* embryos at 24 h and 48 h post**
612 **fertilization times. (A)** Expression levels of target transcripts evaluated by qPCR. Data are reported
613 as the mean ± SD of the relative expression with respect to untreated samples within each life stage
614 (N = 4). *P < 0.05, **P < 0.01 according to the Mann-Whitney U test. **(B)** DISTLM modeling with
615 distance-based redundancy analysis (dbRDA) to explore the amount of the variation in gene
616 transcription is attributable to E₂ treatment (Euclidean Distance resemblance matrix, 999
617 permutations).

618

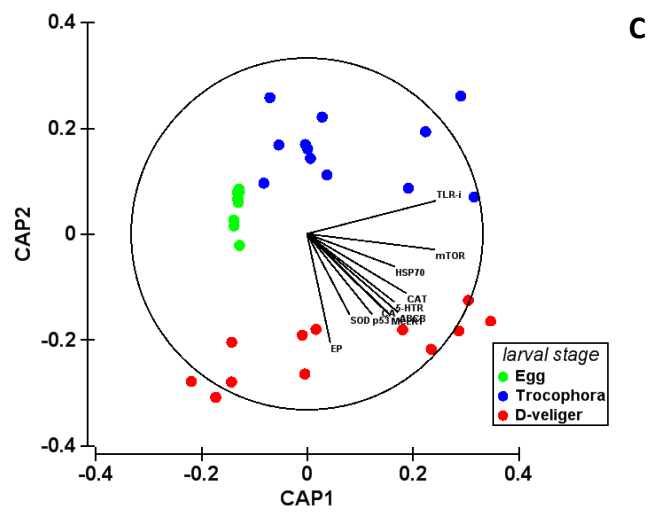
Figure 1
[Click here to download Figure: Fig1.pdf](#)



A



B



C

Figure 2
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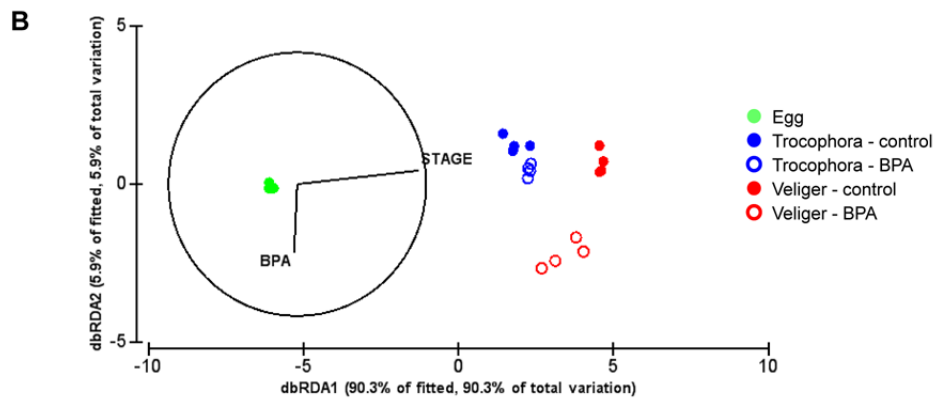
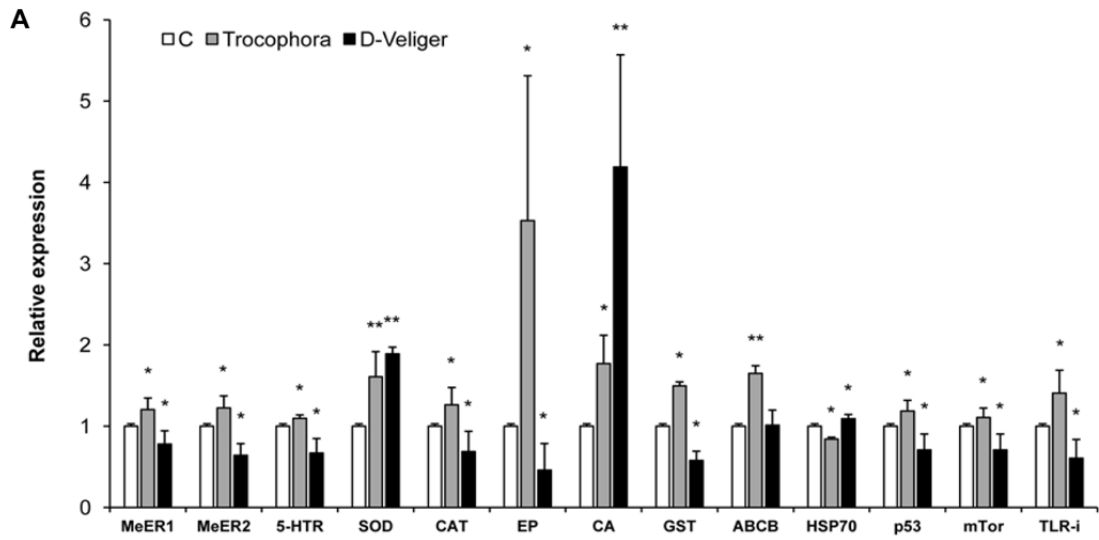


Figure 3
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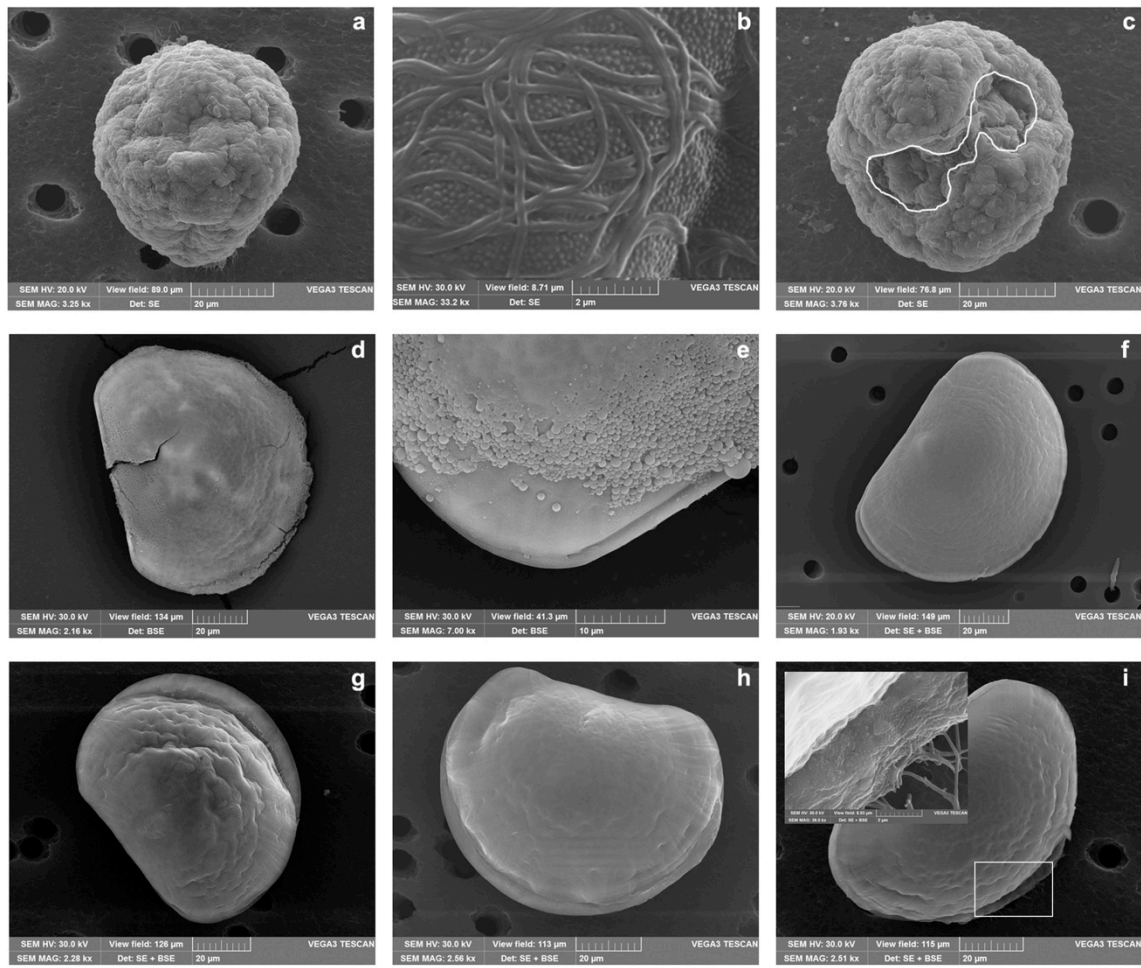
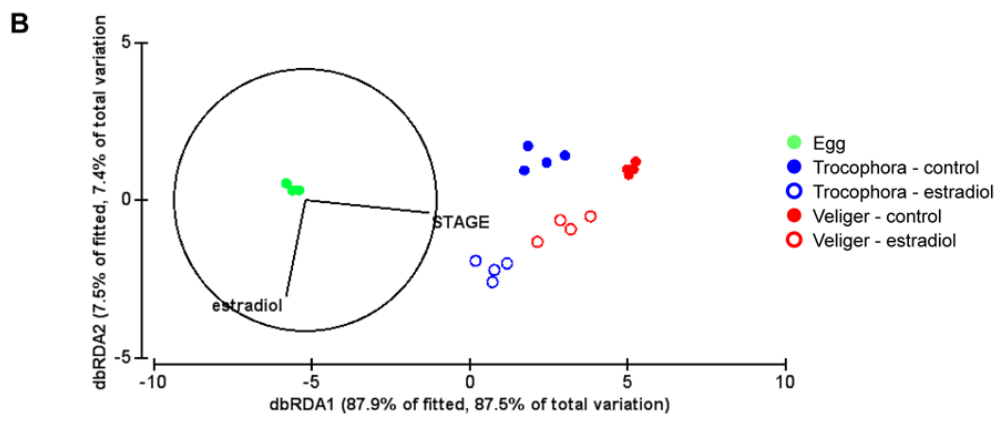
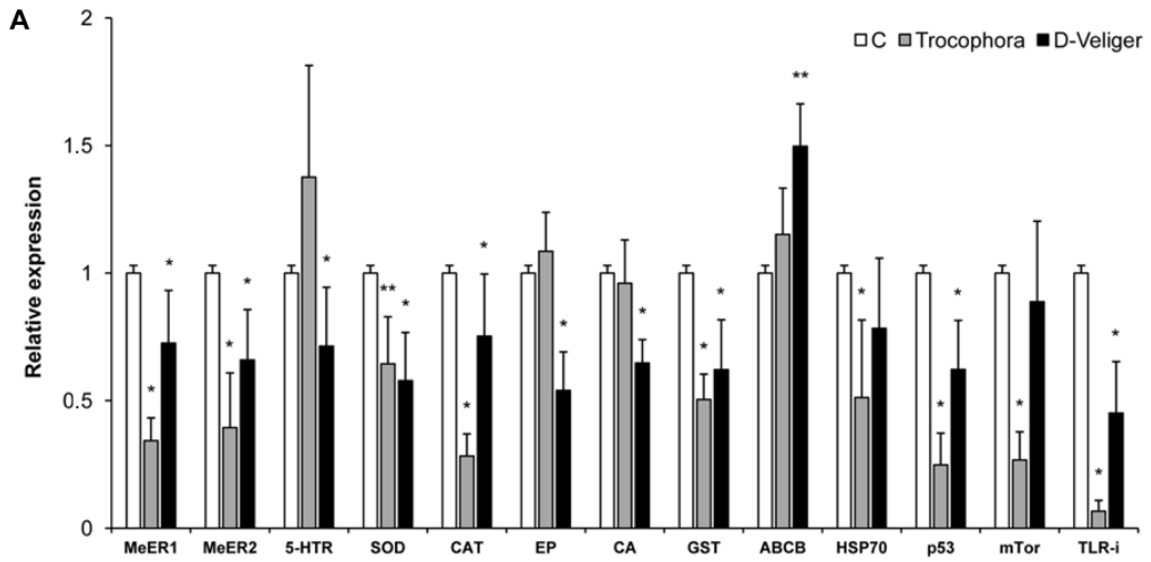


Figure 4
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