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Adaptive strategies of sponges to deoxygenated oceans

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Keywords:	climate change, Porifera, marine benthic hypoxia, hypoxic events, oxygen depletion, eutrophication, phenotypic plasticity, evolution
Abstract:	Ocean deoxygenation is one of the major consequences of climate change. In coastal waters, this process can be exacerbated by eutrophication, which is contributing to an alarming increase in the so-called "dead zones" globally. Despite its severity, the effect of reduced dissolved oxygen has only been studied for a very limited number of organisms, compared to other climate change impacts such as ocean acidification and warming. Here we experimentally assessed the response of sponges to moderate and severe simulated hypoxic events. We ran three laboratory experiments on four species from two different temperate oceans (NE Atlantic and SW Pacific). Sponges were exposed to a total of five hypoxic treatments, with increasing severity (3.3, 1.6, 0.5, 0.4 and 0.13 mg O ₂ L ⁻¹ , over 7–12 days). We found that sponges are generally very tolerant of hypoxia. All the sponges survived in the experimental conditions, except <i>Polymastia croceus</i> , which showed significant mortality at the lowest oxygen concentration (0.13 mg O ₂ L ⁻¹ , lethal median time: 286 h). In all species except <i>Suberites carnosus</i> , hypoxic conditions do not significantly affect respiration rate down to 0.4 mg O ₂ L ⁻¹ , showing that sponges can uptake oxygen at very low concentrations in the surrounding environment. Importantly, sponges displayed species-specific phenotypic modifications in response to the hypoxic treatments, including physiological, morphological, and behavioural changes. This phenotypic plasticity likely represents an adaptive strategy to live in reduced or low oxygen water. Our results

	also show that a single sponge species (i.e., <i>Suberites australiensis</i>) can display different strategies at different oxygen concentrations. Compared to other sessile organisms, sponges generally showed higher tolerance to hypoxia, suggesting that sponges could be favoured and survive in future deoxygenated oceans.

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Adaptive strategies of sponges to deoxygenated oceans

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1 Adaptive strategies of sponges to deoxygenated oceans

2 Abstract

3 Ocean deoxygenation is one of the major consequences of climate change. In coastal waters, this process
4 can be exacerbated by eutrophication, which is contributing to an alarming increase in the so-called “dead
5 zones” globally. Despite its severity, the effect of reduced dissolved oxygen has only been studied for a very
6 limited number of organisms, compared to other climate change impacts such as ocean acidification and
7 warming. Here we experimentally assessed the response of sponges to moderate and severe simulated
8 hypoxic events. We ran three laboratory experiments on four species from two different temperate oceans
9 (NE Atlantic and SW Pacific). Sponges were exposed to a total of five hypoxic treatments, with increasing
10 severity (3.3, 1.6, 0.5, 0.4 and 0.13 mg O₂ L⁻¹, over a 7–12-day period). We found that sponges are
11 generally very tolerant of hypoxia. All the sponges survived in the experimental conditions, except
12 *Polymastia croceus*, which showed significant mortality at the lowest oxygen concentration (0.13 mg O₂ L⁻¹,
13 lethal median time: 286 h). In all species except *Suberites carnosus*, hypoxic conditions do not significantly
14 affect respiration rate down to 0.4 mg O₂ L⁻¹, showing that sponges can uptake oxygen at very low
15 concentrations in the surrounding environment. Importantly, sponges displayed species-specific phenotypic
16 modifications in response to the hypoxic treatments, including physiological, morphological, and
17 behavioural changes. This phenotypic plasticity likely represents an adaptive strategy to live in reduced or
18 low oxygen water. Our results also show that a single sponge species (i.e., *Suberites australiensis*) can
19 display different strategies at different oxygen concentrations. Compared to other sessile organisms,
20 sponges generally showed higher tolerance to hypoxia, suggesting that sponges could be favoured and
21 survive in future deoxygenated oceans.

22 KEYWORDS

23 climate change, Porifera, adaptation evolution, marine benthic hypoxia, hypoxic events, oxygen depletion,
24 eutrophication, sessile organisms, dead zones, phenotypic plasticity

25 1 INTRODUCTION

26 Anthropogenic emissions of carbon dioxide and other greenhouse gasses have increased exponentially
27 since the industrial revolution, causing significant changes in the Earth's climate (Raupach & Canadell, 2010;
28 IPCC, 2021). Climate change has three main effects on the marine environment: warming, acidification, and
29 oxygen decline (Bijma et al., 2013). While most ecological and physiological research has targeted the first
30 two stressors, deoxygenation remains comparatively neglected (Limburg et al., 2017). Despite the
31 scant little attention, recent research shows that oxygen loss is a major anthropogenic stressor for marine
32 biota that may exceed the severity of the combined effects of ocean warming and acidification (Sampaio et
33 al., 2021).

34 Oxygen is essential to all aerobic life, and ocean deoxygenation has the potential to affect all
35 biogeochemical and biological processes within the oceans (Semenza, 2007; Levin & Breitburg, 2015). In the
36 open sea, warming is considered the main cause of O₂ reduction: an increase in sea temperature leads to
37 decreased O₂ solubility, increased water stratification, and alterations to oceanic circulation, which reduces
38 O₂ supply to the ocean interior (Doney, 2010; Keeling et al., 2010). Higher temperatures also enhance
39 microbial respiration, which can further deplete oxygen in marine ecosystems (Altieri & Diaz, 2019;
40 Robinson, 2019). Oxygen levels in the global oceans have already declined by 2% during the last 50 years,
41 with more significant O₂ declines in the North Pacific and tropical oxygen minimum zones (OMZ) (Levin &
42 Breitburg, 2015). This is likely to get worse in the future, with models predicting a global ocean reduction in
43 O₂ of up to 7% by the end of the century (Keeling et al., 2010).

44 In coastal waters, climate-driven deoxygenation can be intensified by eutrophication (Nixon, 1995; Altieri &
45 Gedan, 2015). The input of anthropogenic nutrients, such as fertilizers and human/livestock wastes, can
46 increase algal growth resulting in an accumulation of organic material on the seafloor. This excess of
47 organic matter is then degraded by bacteria, causing O₂ depletion that can lead to hypoxic conditions
48 (Smith et al., 2006). In shallow and well-mixed waters, eutrophication-driven hypoxia is generally caused by
49 nocturnal heterotrophic respiration, resulting in daily oscillations in oxygen concentration. In contrast, long-
50 term hypoxic events are more likely to occur in enclosed seas or basins (Levin et al., 2009). Hypoxia has

51 widespread and severe impacts across taxonomic and functional groups. The intensity and duration of
52 oxygen depletion are the main factors influencing the severity of hypoxic events on benthic organisms
53 (Levin et al., 2009; Altieri & Diaz, 2019). Mild hypoxia can alter behavioural patterns, decrease feeding rates
54 and cause changes in physiological processes (Vaquer-Sunyer & Duarte, 2008). Severe hypoxic events can
55 cause mass mortalities, leading to the formation of the so-called “dead zones”, areas largely devoid of
56 macrofauna (Diaz & Rosenberg, 2008). Dead zones have been reported in small water bodies such as
57 harbours, fjord and inlets, and large basins, such as the Baltic Sea, spreading over 60,000 km² (Altieri & Diaz,
58 2019). As climate and land use continue to change, coastal hypoxia is expected to worsen, with the
59 increased occurrence, frequency, intensity, and duration of hypoxic events (Diaz & Rosenberg, 2011).

60 Despite the extent of the problem and the dramatic effects caused by ocean deoxygenation, the response
61 of many groups of organisms to hypoxia is still poorly studied. This lack of knowledge limits our ability to
62 model the effects of declining oxygen availability on marine ecosystems (Seibel, 2011). To date, research on
63 tolerance to reduced levels of dissolved oxygen has primarily focused on fish, crustaceans and molluscs
64 (Vaquer-Sunyer & Duarte, 2008), while very little is known about other groups, especially sessile organisms.
65 Sessile organisms are particularly vulnerable to hypoxic events because they cannot move or migrate to
66 well-oxygenated water. Furthermore, sessile organisms include many important habitat-forming species, so
67 any change in their abundance could have major consequences for the ecosystems they support (Vergés et
68 al., 2019; Woodhead et al., 2019; Piazzini et al., 2021). Therefore, it is critical to understand how these
69 organisms respond to hypoxia to predict possible future changes and effectively manage marine
70 ecosystems.

71 Sponges are the dominant sessile organisms in many marine ecosystems and are found in high abundance
72 in tropical, temperate, and polar ecosystems (Ayling, 1983, Bell et al., 2020). They perform many important
73 ecological functions, including contributing to nutrient cycling, bioerosion, enhancing ecosystem complexity
74 and providing habitats for a wide range of associated organisms (Wulff, 2006; Bell, 2008; Maldonado et al.,
75 2012). Despite being important components of marine ecosystems, sponge tolerance to hypoxia has been
76 poorly investigated to date. Mills et al. (2014) showed that *Halichondria panicea* can feed and respire with

77 oxygen levels down to 4% of air saturation. However, the authors did not provide information on the
78 duration of the treatments and replication; furthermore, in the same study, information on the
79 temperature and salinity of the water was unavailable, so it is not possible to derive the actual oxygen
80 concentrations to which sponges were exposed. Two other relevant experiments have investigated the
81 short-term response of sponges to hypoxia. Mills et al. (2018) exposed *Tethya wilhelma* to a step
82 decreasing oxygen concentration (30–40 h with O₂ lower than 10% a.s., 0.7 mg L⁻¹). They found that
83 sponges continued to perform periodic full-body contractions down to 0.27 mg O₂ L⁻¹, but ceased below
84 that concentration. Leys & Kahn (2018) exposed *Geodia barretti* to 6.5 h of hypoxia (7% air saturation, 0.6
85 mg O₂ L⁻¹), and found that sponge respiration rate remained largely unchanged. However, filtration rates
86 dropped almost immediately after the oxygen level was reduced. Despite these earlier studies, we still have
87 very little insight into how sponges may cope with hypoxic events caused by ocean and coastal
88 deoxygenation.

89 Here we provide the first comprehensive assessment of sponge response to hypoxia. Specifically, we
90 experimentally investigated the physiological, behavioural, and morphological responses of four temperate
91 sponge species to moderate and severe hypoxic conditions. We ran the first experiment to expose sponges
92 to moderate hypoxic conditions for seven days, including a wide range of dissolved oxygen concentrations
93 (0.5, 1.6 and 3.3 mg O₂ L⁻¹). Subsequently, we investigated sponge response to severe hypoxia (0.13 and 0.4
94 mg O₂ L⁻¹) for 12 days with two additional experiments. Finally, we discuss sponge tolerance to low
95 dissolved oxygen compared to other sessile organisms in the context of future climatic conditions.

96 2 MATERIALS AND METHODS

97 2.1 Study area and species

98 Experiment 1 (moderate hypoxia) was performed in Ireland (Renouf Laboratory, Lough Hyne) on two
99 abundant North-East Atlantic sponge species: *Cliona celata* Grant, 1826 and *Suberites carnosus* (Johnston,
100 1842). Experiments 2 and 3 (severe hypoxia) were performed in New Zealand (Wellington University
101 Coastal Ecology Laboratory, Wellington) on two abundant temperate Australasian species: *Polymastia*
102 *croceus* Kelly-Borges and Bergquist, 1997 and *Suberites australiensis* Bergquist, 1968.

103 2.1 Experiment 1: moderate hypoxia

104 In the first experiment, we investigated the response of *Cliona- celata* and *Suberites- carnosus* to a wide
105 range of oxygen concentrations, using an air-tight system with a continuous flow of seawater. Sponges
106 were exposed to ~95% ($7.71 \pm 0.19 \text{ mg O}_2 \text{ L}^{-1}$), ~40% ($3.34 \pm 0.17 \text{ mg O}_2 \text{ L}^{-1}$), ~20% ($1.56 \pm 0.19 \text{ mg O}_2 \text{ L}^{-1}$)
107 and ~6% ($0.48 \pm 0.09 \text{ mg O}_2 \text{ L}^{-1}$) air saturation (a.s.) for seven days (a summary of the seawater parameters
108 is provided in Table S1).

109 The experimental set-up (see scheme in Figure S1) consisted of two independent replicate modules for
110 each treatment, randomly distributed in the experimental set-up. To condition water, we used two header
111 tanks for each experimental module: one providing water and one reservoir. Header tanks were filled with
112 10- μm -filtered seawater. The oxygen level was then lowered and maintained to the desired dissolved
113 oxygen concentration by bubbling specific mixtures of N_2 (BOC, food-grade) and air, through glass-ceramic
114 diffusers. Hypoxic gas blends were prepared by decanting food-grade N_2 and air in 15 L scuba cylinders
115 using an oxygen decanting assembly (Undersea Ltd, 5215) with a DPM-300 digital gauge (0.25% accuracy).
116 Oxygen concentration was then checked with a Nuair Pro O_2 Analyser and adjusted, if necessary.

117 Conditioned water was delivered to two replicate experimental chambers (2.3 L) for each system at a rate
118 of 25 L per day, ensuring 100% water replacement every 2 h and 15 min. Water circulation within each
119 experimental chamber was provided by the gravity-driven water flow (~3 cm/s). Temperature was kept
120 constant using a water bath controlled by an aquarium chiller.

121 *Cliona celata* was collected from the Kedges (51°27'41.4 "N 9°20'44.2 "W), whereas *Suberites S.-carnosus*
122 was collected from the rocky cliffs of Lough Hyne (51°30'00.4"N 9°18'03.9"W). For both species, sampling
123 was carried out at 10–18 m in June 2019 and sponges collected were at least 2 m apart. Sponges were then
124 left to recover for two months from harvesting stress in a 1 m³ underwater cage placed at 8 m of depth.
125 Sponges were then transferred to the experimental system and randomly distributed across the
126 experimental chambers. The experimental design consisted of 32 experimental chambers (two replicate
127 chambers for each species for each replicate module, and two replicate modules for each treatment). A
128 diagram of the experimental design is reported in Figure S2. Three sponges were placed in each chamber (6
129 sponges in total for each replicate module and 12 for each treatment). Sponges belonging to different
130 species were not mixed but kept in separate chambers. Sponges were left to acclimate with oxygen
131 saturated air for five days before oxygen was lowered by introducing hypoxic water into the chambers.
132 Oxygen concentration was lowered in 24h and was then maintained for seven days until the end of the
133 experiment (a graph showing the oxygen concentration in the different treatments over time is provided in
134 Figure S3). In natural ecosystems, hypoxic conditions can develop in short times ranging from hours to a
135 few days (Breitburg, 1990; Nezlin et al., 2009), so we consider these acclimation times appropriate and
136 ecologically relevant. Temperature and oxygen concentration inside the experimental chamber were
137 measured twice a day using a Fibox 4 oxygen meter with a dipping probe (Presence GmbH, Germany). A
138 two-spot calibration was performed on the oxygen probe every three days, using sodium dithionite for 0 %
139 oxygen and air-saturated water for 100% oxygen.

140 2.2 Experiments 2 and 3: severe hypoxia

141 We also investigated the response of sponges (*Polymastia P.-croceus* and *Suberites S.-australiensis*) to
142 severe hypoxia through two separate experiments using an air-tight system. In experiment 2, sponges were
143 exposed to ~5% (0.4 ± 0.04 mg O₂ L⁻¹), and ~100% a.s. (8.34 ± 0.13 mg O₂ L⁻¹), while in eExperiment 3,
144 sponges were exposed ~1.5% (0.13 ± 0.02 mg O₂ L⁻¹), and ~100% a.s. (8.15 ± 0.16 mg O₂ L⁻¹). The different
145 oxygen concentration in the controls was due to the small difference in temperature between the two

146 experiments (13.3 ± 0.5 °C in experiment 2 compared to 14.3 ± 0.6 °C in experiment 3). A summary of the
147 seawater parameters is provided in Table S1.

148 Sponges were kept in independent cylindrical air-sealed polypropylene chambers (10 L), randomly
149 distributed inside a water bath. Every two days, ~70% of the water was replaced using 10- μ m-filtered
150 seawater, preconditioned to the desired oxygen concentration in independent conditioning tanks. Oxygen
151 concentration was then maintained by bubbling air or air-N₂ blends through glass-ceramic diffusers (see
152 section 2.1.2 for more details on gas blends). Custom made de-bubbler devices were used to eliminate
153 bubbles coming from the ceramic diffusers that could affect sponges (Figure S4). The sponges were fed
154 twice a day with *Nannochloropsis* microalgae (1–2 μ m cell diameter; Nanno 3600™ Reed Mariculture, US.).
155 Water circulation within each experimental chamber was provided by the de-bubbler device and an
156 additional water pump located on the side of the chamber, which provided a constant circular water flow.
157 The chambers were placed in a water bath to control water temperature.

158 *Polymastia croceus* was collected from Barrett Reef (Wellington South Coast, 41°20'31.1"S 174°50'09.7"E)
159 by cutting fragments (~8 cm³) from separate sponges (at least 5 m apart). Whole specimens of *Suberites S-*
160 *australiensis* were collected from Mahanga Bay (Wellington Harbour, 41°17'32.2"S 174°50'06.5"E),
161 attached to a fragment of their respective substrate. Sponges were then left to recover for three weeks
162 after sampling and cutting stress in water tables with 10- μ m-filtered flow-through seawater.

163 Sponges were then transferred to the experimental system, consisting of 12 experimental chambers (3
164 independent replicate chambers for each species and treatment combination). Five sponges were placed in
165 each chamber (15 sponges in total for each treatment). Sponges were left to acclimate with oxygen
166 saturated air for five days. Oxygen was then lowered by bubbling a specific Air-N₂ mixture. In experiment 2,
167 oxygen was lowered in ~24 hours and then maintained at 5% a.s. (0.4 mg O₂ L⁻¹) for 12 days until the end of
168 the experiment. While in eExperiment 3, oxygen was firstly gradually lowered to an intermediate
169 concentration (10% a.s., 1 mg O₂ L⁻¹) for two days, then decreased lowered to 1.5% a.s. (0.13 mg O₂ L⁻¹) in
170 ~72 hours (which included a preacclimation at 10% a.s., 1 mg O₂ L⁻¹) and maintained for 12 days until the
171 end of the experiment (Fig. S3). This further acclimation in hypoxic conditions was made because of the

172 very low O₂ concentration of the treatment. In experiment 3, due to the very low concentration of O₂ (0.13
173 ± 0.02 mg L⁻¹), O₂ increased to ~0.3 mg L⁻¹ for about 20 minutes during daily examinations. Temperature
174 and oxygen concentration inside the experimental chamber were measured twice a day using Fibox 4
175 oxygen meter with a dipping probe (Presence GmbH, Germany). A two-spot calibration was performed on
176 the oxygen probe every three days, using sodium dithionite for 0% oxygen and air-saturated water for
177 100% oxygen.

178 2.3 Response variables

179 2.3.1 Survival and health monitoring

180 Sponge health was monitored daily during the experiment. Sponges showing ≥ 25% of external necrosis
181 were considered dead and ~~immediately~~ removed from their treatment tanks during the daily checks, so as
182 not to impact other sponges in the treatments. At the end of the experiments, all sponges were sectioned
183 to assess the presence of any internal necrosis.

184 2.3.2 Respiration Rate

185 For all the experiments, respiration rate was measured on the same specimens at *T*₀ (before the beginning
186 of the experiment), *T*_{1/2} (after two days from the beginning of the final treatment in experiment 1, and
187 after five days in experiments 2 and 3) and *T*-end (end of the experiment). In experiment 1 (moderate
188 hypoxia), we measured respiration rates of three sponges in each replicate module (*n* = 6 for each
189 treatment). In experiments 2 and 3 (severe hypoxia), respiration rates were measured on three sponges in
190 each experimental chamber (*n* = 9 for each treatment). To measure respiration rate, sponges were placed
191 in sealed cylindrical glass respiration chambers (150 ml for *Cliona C.-celata*; 80 ml for *Suberites S.-carnosus*;
192 250 ml for *Polymastia P.-croceus* and *Suberites S.-australiensis*) with PreSens oxygen sensor spots (SP-PSt3-
193 NAU) attached to their inner surface. Experimental chambers contained either oxygen saturated water
194 (pre-experimental measurements and controls) or water at a slightly higher oxygen concentration than the
195 experimental treatment (+20–50%, depending on the treatment) collected from the respective header
196 tanks. Respiration rates were not performed on sponges from ~~e~~Experiment 3 (1.5% a.s.). The incubations
197 were performed in controlled temperature (water bath) and dark conditions. The water inside the

198 respirometry chambers was gently stirred using a magnetic stir bar. After 20 min of acclimation, oxygen
199 concentration inside the chambers was measured every 10 min for 1 hour, using a Fibox 4 oxygen meter
200 with a polymer optical fibre (POF). Respiration measurements were ended prematurely if the oxygen level
201 fell below 70% of the treatment concentration to avoid any detrimental effect on the sponges. Blank
202 incubations, containing only seawater were performed every respiration run and used to correct for any
203 microbial community respiration in the seawater. A two-point calibration was performed on the oxygen
204 sensor spots before each measurement session.

205 Respiration measurements were standardized to sponge ash-free dry weight (*AFDW*) from buoyant weight
206 (*BW*) measurements (Fig. S5). For *Suberites S. australiensis*, it was not possible to estimate *AFDW* from *BW*
207 due to the abundant external material accumulated by the sponge inside the tissue that influences the *BW*.
208 For this species, we measured the *AFDW* of all the specimens used in the respirations at the *T-end*, and we
209 assumed that sponges had the same weight at *T0* and *T1/2*.

210 2.3.3 Changes in weight, size, and morphology

211 Changes in weight and size over time relative to the initial values were estimated by calculating the
212 buoyant weight variation (*BWV*) and contracted area variation (*CAV*). For all of the experiments, buoyant
213 weights (*BW*) of all experimental sponges (except *Suberites S. australiensis*) were taken at *T0* and *T-end* and
214 used to calculate relative buoyant weight variation as $BWV = [(BW_{T-end} - BW_{T0}) / BW_{T0}] \cdot 100$. Buoyant weight
215 was measured with a digital scale (A&D FX-200i) following the methods of Osinga et al. (1999). For
216 experiments 2 and 3 (severe hypoxia), photographs of contracted sponges were taken at *T0* and *T-end* to
217 measure sponge contracted area (*CA*) and calculate contracted area variation as $CAV = [(CA_{T-end} - CA_{T0}) /$
218 $CA_{T0}] \cdot 100$ (following Osinga et al., 1999). Contraction was achieved by disturbing sponges with a blunt
219 plastic rod (being careful not to damage the sponge) and waiting for one hour for the sponge to react to the
220 stimulus. All the photographs were analysed using ImageJ (US National Institutes of Health, Bethesda, Md,
221 USA).

222 During experiments 2 and 3, treatment conditions induced the development of peculiar morphological
223 structures in some specimens of both *Polymastia P.-croceus* and *Suberites S.-australiensis*. Sponges were
224 photographed and monitored daily to calculate the percentage of specimens developing these structures
225 and the median time of occurrence.

226 2.3.4 Sponge contractile behaviour

227 During experiments 2 (5% a.s.) and 3 (1.5% a.s.), sponge contractile behaviour was monitored daily from T_0
228 to $T\text{-end}$, on all experimental sponges through photographic analysis. For *Suberites S.-australiensis*, the
229 contractile behaviour was estimated using an “expansion ratio” ($EXPR$) calculated as $EXPR = A_{T_i} / CA_{T_0}$,
230 where A_{T_i} is the area occupied at T_i and CA_{T_0} is the contracted area at T_0 . Area was preferred over volume
231 because of the low invasiveness of the measurements. In *Polymastia P.-croceus*, contraction/expansion
232 mainly occur at the papillae level, so the contractile behaviour was estimated from the ratio of expanded
233 papillae (REP) calculated as $REP = P_E / P_{tot}$, where P_E is the number of visible expanded papillae and P_{tot} is
234 total number of visible papillae. Expanded papillae were defined as papillae whose length was at least two
235 and a half times the width.

236 2.3.5 Pumping rate

237 Pumping rate was only calculated for *Suberites S.-australiensis* from experiments 2 and 3 (severe hypoxia).
238 Having only one osculum of relatively large size, this species was particularly suitable for investigating
239 changes in pumping rate. To minimize sponge disturbance during the experiment, pumping rate (PR) was
240 derived from the measurement of the sponge osculum cross-sectional area (OSA). In sponges, pumping
241 rate (PR) is correlated with OSA (e.g. Goldstein et al., 2019; Morganti et al., 2021). In the case of *S.*
242 *australiensis*, this relationship was calculated on 20 sponges (following Yahel et al., 2005) and was found as
243 $PR = 6.55 \cdot OSA^{1.43}$ (Fig. S6). Photographs of the oscula with scale were taken daily from T_0 to $T\text{-end}$, on
244 three sponges in each experimental chamber (the same specimens each time point, $n = 9$ per treatment).
245 Since *S. australiensis* has only one osculum, pumping rate was then standardized per sponge volume.

246 2.3.6 Histology

247 Histological sections of *Suberites S. australiensis* from the severe hypoxia experiments were analyzed to
248 calculate the percentage of the sponge body occupied by the aquiferous system (system of connected
249 water channels inside the sponge). At *T-end*, two contracted sponges for each experimental chamber (n = 6
250 per treatment) were fixed and processed following the methods of Strano et al. (2021). Three replicate
251 sections for each sponge were then photographed under a dissecting microscope (Olympus SZ61) and
252 photographed using a Canon EOS 70D digital camera. To calculate the area occupied by the aquiferous
253 system, pictures were analyzed using ImageJ.

254 2.4 Data analysis

255 All the statistical data analyses were performed in R version 3.1.3 (R Core Team, 2013), except
256 PERMANOVA models, which were performed using PRIMER v7 with PERMANOVA+ add-on (Anderson et al.,
257 2008; Clarke & Gorley, 2015). Experiments 2 and 3 were analyzed separately. To investigate respiration
258 rate, pumping rate and expansion ratio in *Suberites S. australiensis* from experiment 2, we used linear
259 mixed-effects models with normally distributed errors and random intercepts (lmer, *lme4* package; Bates et
260 al., 2015). For pumping rate, we added a constant variance function structure (*varIdent*) to the linear
261 mixed-effects models to allow different variances for each treatment at each time point (*lme*, *R* package
262 *nlme*; Pinheiro et al., 2021). The constant variance function structure was necessary because the variance
263 of the response variable differed across treatments and experimental days. To investigate the effect of time
264 and treatment on the expansion ratio of *Polymastia P. croceus* in eExperiment 3, we used a generalized
265 linear mixed model with beta regression and logit link (*glmmTMB*, Brooks et al. 2017). In all the mixed
266 models, treatment and time were considered fixed effects, while experimental chamber and sponge
267 specimen were considered random effects. The experimental chamber effect was included to address
268 pseudo-replication. For these models, fixed- and random-effect terms were tested using the function *anova*
269 and *ranova* (*R* package *lmerTest*, Kuznetsova et al., 2017), respectively; while post hoc pairwise
270 comparisons were computed on estimated marginal means using *emmeans* (*R* package *emmeans*; Lenth,
271 2021). The ratio of expanded papillae in *P. croceus* from experiment 2 and expansion ratio in *S. australiensis*

272 from Experiment 3 were investigated using repeated measure univariate PERMANOVA (Anderson, 2001;
273 2014), because did not meet the normality assumption for mixed-effects models. Pairwise tests were then
274 calculated using permutation *t*-tests (*R* package *RVAideMemoire*; Hervé, 2021). PERMANOVA and
275 permutation *t*-tests were also used to supplement mixed-effects models when there were concerns about
276 the normality of the residuals (pumping rate in *S. australiensis*). Change over time relative to the initial
277 value of buoyant weight and contracted area, and differences in percentage occupied by the aquiferous
278 system were investigated using Kruskal–Wallis H tests, Welch’s *t*-tests or Wilcoxon Signed-Rank Tests,
279 depending on the variable. Respiration rates from experiment 1 were $\log(x + 1)$ transformed, and pumping
280 rates were ~~square-square-root~~ transformed to meet normality assumptions. The goodness of fit, normality
281 and homoscedasticity of the errors were checked for all models by inspecting plots of the normalized
282 residuals and the quantile-quantile plots. All the multiple comparisons were corrected using Benjamini-
283 Hochberg Procedure, but uncorrected *p*-values are reported in the text. All the statistical analyses made for
284 each variable are reported and summarized in Table S2.

285 Time to event analysis for sponge survival and development of peculiar morphological structures (modified
286 papillae and protruding oscular membranes) was performed using Kaplan-Meier Method, and *p*-values
287 were calculated using the Log Rank Test implemented in the survival *R* package (Therneau, 2021). Median
288 lethal time (LT₅₀) and median time to the development of modified morphological structures were
289 calculated using a logistic model.

290 3 RESULTS

291 3.1 Sponge responses to moderate hypoxia

292 All the sponges of experiment 1 survived the seven days of treatment, except one specimen of *Suberites S.*
293 *carneus* in the lower DO treatment (6% a.s.), which presented internal necrosis on the final day of the
294 experiment.

295 Mean buoyant weight variation between *T0* and *T-end* ranged between -1% and -1.6% for *Cliona C. celata*
296 and +2.1% and -0.5% in *Suberites S. carneus*. There were no differences in buoyant weight variation
297 among treatments for both species, but for *C. celata* there was a significant slight decrease in weight in the
298 40% a.s. (-1.6%, $p = 0.008$) and 20% a.s. (-1.4%, $p = 0.008$) treatments (Tab. S3; Fig. S7).

299 For *Cliona C. celata*, there was no significant effect of time or treatment on the respiration rate (Tab. S4).
300 However, pairwise comparisons ~~found~~ revealed a significant decrease ($p = 0.028$) in the 20% a.s. treatment
301 between day 0 and 7, and a significant increase ($p = 0.029$) in the 6% a.s. treatment between day 2 and day
302 7, but both became non-significant after the correction for multiple comparisons (Tab. S4). However, the
303 data suggest a coherent temporal pattern in the respiration rate in both 20% a.s. and 6% a.s. treatments. *C.*
304 *celata* respiration rate decreased after two days from the start of the experiment and then increased until
305 the end of the experiment. In contrast, in both the 100% a.s. and 40% a.s. treatments, respiration rate
306 remained stable for the whole duration of the experiment (Fig. 1a).

307 For *Suberites S. carneus*, there was a significant ~~effect of time ($p = 0.002$), and the~~ interaction of time and
308 treatment ($p = 0.007$) on the respiration rate (Tab. S5). Pairwise comparisons ~~revealed~~ found a significant
309 decrease in respiration rate between day 0 and 7 ($p < 0.0001$), and day 2 and 7 ($p < 0.0001$) (Tab. S5). The
310 respiration rate also slightly decreased towards the end of the experiment in the 20% a.s. treatment (but
311 not significantly), while in both the 100% a.s. and 40% a.s. treatments, respiration rate remained stable for
312 the duration of the experiment (Fig. 1b).

313 3.2 Sponge responses to severe hypoxia

314 3.2.1 Survival

315 Sponge survival differed among species, with *Suberites S. australiensis* more tolerant than *Polymastia P.*
316 *croceus*. No mortality was observed for *S. australiensis* in both experiments 2 (5% a.s.) and 3 (1.5% a.s.). In
317 contrast, for *P. croceus*, significant mortality ($p = 0.001$) was observed in sponges exposed to the 1.5% a.s.
318 treatment, starting from day 10 (day 12 when including the hypoxic acclimation), and with a median lethal
319 time of 11.9 ± 0.3 days (Fig S8–9). Eight out of 15 sponges had died by the end of the experiment. No
320 mortality was observed for *P. croceus* in the 5% a.s. treatment.

321 3.2.2 Change in weight and size

322 For *Polymastia P. croceus*, buoyant weight variation between T_0 and T_{end} differed among treatments in
323 Experiment 3 (1.5% a.s.) ($t = 2.82$, $p = 0.012$), but not in experiment 2 (5% a.s.). Sponges from the 1.5% a.s.
324 treatment experienced a significant decrease in buoyant weight (-7.1% , $t = -5.17$, $p = 0.002$), while the
325 controls did not experience any significant change (Tab. S6; Fig. S10).

326 The relative variation in area of contracted sponges (after stimulating contraction) between T_0 and T_{end}
327 differed significantly between treatments and controls for both *Polymastia P. croceus* ($W = 12$, $p < 0.0001$)
328 and *Suberites S. australiensis* ($W = 13$, $p < 0.0001$), but only in experiment 2 (5% a.s.). Both *P. croceus* and *S.*
329 *australiensis*, from the 5% a.s. treatment, experienced an increase in contracted area ($+18.9\%$, $W = 120$, $p =$
330 0.0001 and $+18.4\%$, $W = 105$, $p = 0.008$, respectively). While *S. australiensis* from the control treatment
331 (experiment 2) experienced a decrease in contracted area (-15.3% , $W = 3$, $p = 0.0003$) (Tab. S7; Fig. S11).

332 3.2.3 Sponge contractile behaviour

333 Low DO treatments generally induced sponge expansion, but the response differed between species, and it
334 was generally more marked in the 5% a.s. treatment. In *Polymastia P. croceus*, the ratio of expanded
335 papillae was significantly affected by ~~time and~~ the interaction between time and treatment in both
336 experiments 2 ($p = 0.0001$) and 3 ($p < 0.0001$ and $p = 0.03$) ~~experiments~~ (Tab. S8–9). During experiment 2
337 (5% a.s.), the treatment induced a progressive expansion of papillae from day 2. The ratio of expanded

338 papillae in sponges from the hypoxic treatment ~~become~~became significantly higher than control sponges
339 from day 6 to the end of the experiment ($p = 0.0002$ – 0.005) (Tab. S8; Fig. 2a). A similar trend was found in
340 ~~e~~Experiment 3 (1.5% a.s.), but the ratio of expanded papillae of the treatment sponges was more variable
341 and became significantly different only at day 9 ($p = 0.003$) (Tab. S9; Fig. 2b). In this experiment, we also
342 found a correlation between the ratio of expanded papillae and mortality. Sponges that survived the
343 treatment had a significantly higher maximum ratio of expanded papillae compared to sponges that died,
344 both when the maximum ratio was calculated at the end of the experiment (Welch t -test: $t = 5.3$, $p =$
345 0.0005) and at day ten, before sponges started to die (Welch t -test: $t = 4.6$, $p = 0.0007$).

346 In *Suberites S. australiensis*, there was a significant ~~interactive effect~~effect of treatment ~~and~~ ($p = 0.001$),
347 time ($p < 0.0001$) ~~and their interaction~~ ($p < 0.0001$) on the expansion ratio in experiment 2 (5% a.s.), but
348 only an effect of time ($p = 0.01$) in ~~e~~Experiment 3 (1.5% a.s.) (Tab. S10–11). For experiment 2 (5% a.s.),
349 pairwise comparisons found significant expansion in sponges (+60%, $p < 0.0001$) between day 0 and 1.
350 Sponges then remained expanded for the whole duration of the experiment, and the expansion ratio was
351 significantly higher in the treatments compared to the controls from the first to the last day of the
352 experiment ($p < 0.0003$) (Tab. S10; Fig. 2c, d).

353 3.2.4 Morphological modifications

354 During experiments 2 (5% a.s.) and 3 (1.5% a.s.), some *Polymastia P. croceus* and *Suberites S. australiensis*
355 sponges exposed to hypoxic treatments underwent morphological modifications (Fig. 3; S12). In some *P.*
356 *croceus*, the conical papillae showed a progressive elongation, flattening, and, in some cases, spiralization
357 (Fig. S12a–f). This process occurred in both the 5% a.s. and 1.5% a.s. treatments, but morphological
358 changes were more pronounced in lower DO treatment (Fig. S12d–e). Exposed to the 1.5% a.s. treatment,
359 some sponges developed papillae so slender that they could not sustain their weight (Fig. S12d–e). The
360 development of these modified papillae was also associated with an apparent increase in the porosity of
361 the sponge external surface (Fig. S12e). In the 5% a.s. treatment, 73% of the sponges developed modified
362 papillae, starting from day 6. In the 1.5% a.s. treatment, 60% of sponges developed modified papillae,
363 starting from day 2, from the beginning of the final treatment (day 4 considering hypoxic acclimation

364 period) (Fig. S13). The median time of development of these morphological structures (considering only the
365 sponges that developed them) was 7.2 ± 0.2 days in the 5% a.s. treatment and 4.6 ± 0.4 days in the 1.5%
366 a.s. treatment (Fig. S14). Although not significant ($\chi^2 = 3.62$, $p = 0.057$), a relationship between modified
367 papillae and survival was found. Among the *P. croceus* that survived the 1.5% a.s. treatment, six had
368 developed modified papillae, while one had not. While among the sponges that died following the 1.5% a.s.
369 treatment, three had developed modified papillae, while five had not.

370 In the the 5% a.s. treatment, 53% of *Suberites S-australiensis* developed a semi-transparent protruding
371 membrane surrounding the oscula. This membrane progressively reduced the oscular-cross sectional area
372 (Fig. 3; S12g–i). The median time it took for these protruding oscular membranes to ~~became~~ become
373 noticeable (considering only the sponges that developed them) was 5.1 ± 0.2 days (Fig. S14). By the end of
374 the experiment, 53% of sponges had developed these structures (Fig. S13).

375 3.2.5 Histology

376 Histological analyses indicated that hypoxia influences the percentage of the sponge body occupied by the
377 aquiferous system in *Suberites S-australiensis*. At the end of experiment 2 (5% a.s.), treatment sponges had
378 a significantly higher percentage of aquiferous system ($t = -9.82$, $p < 0.0001$), compared to the controls
379 ($35.9 \pm 7.1\%$ vs $6.4 \pm 1.8\%$). No significant differences were found for ~~e~~Experiment 3 (1.5% a.s.) (Fig. 3;
380 S12j–m; S15).

381 3.2.6 Pumping rate

382 Oxygen concentration significantly affected the pumping rate of *Suberites S-australiensis* in both
383 experiments 2 (5% a.s.) and 3 (1.5% a.s.) (Tab. S12–15). In experiment 2 (5% a.s.), there was significant
384 ~~effect~~ interaction of treatment ~~and~~ ($p = 0.02$), time ($p = 0.0001$), ~~and the interaction between time and~~
385 ~~treatment~~ ($p = 0.0001$) (linear mixed-effects model; Tab. S12). Pumping rate significantly increased from
386 day 0 to 1 ($p < 0.0001$), remained stable from day 1 to 2, and then decreased from day 2 to 3 ($p < 0.0001$)
387 and from day 3 to 4 ($p = 0.005$) (Tab. S12). Sponges from the 5% a.s. treatment had a significantly higher
388 pumping rate than the control at day 1 ($p = 0.007$) and 2 ($p = 0.002$) (Tab. S12; Fig. 2e). Similar results were

389 given by PERMANOVA (Tab. S13). For ~~e~~Experiment 3 (1.5% a.s.), both the linear mixed-effects model and
390 PERMANOVA ~~revealed~~found significant ~~effect of time ($p < 0.0001$ and $p = 0.0006$, respectively) and the~~
391 interaction between time and treatment ($p = 0.049$ and $p = 0.002$, respectively) on the pumping rate of *S.*
392 *australiensis* (Tab. S14–15). However, differences were less marked compared to experiment 2 and
393 pairwise comparisons only ~~found~~revealed a slight decrease of pumping rate of treatment sponges between
394 day 0 and 14 ($p = 0.0002$) (Tab. S14; Fig. 2f).

395 3.2.7 Respiration rate

396 In experiment 2 (5% a.s.), linear mixed-effects models only ~~revealed~~found a significant effect of time on the
397 respiration rate ~~but not treatment or interaction between treatment and time~~, for both *Polymastia P.*
398 *croceus* and *Suberites S. australiensis*~~arnosus~~ (Tab. S16–17; Fig. 4). In *P. croceus*, pairwise comparisons
399 ~~revealed~~found a slightly higher respiration rate of the controls at day 12 compared to day 0 ($p = 0.008$) and
400 day 5 ($p = 0.004$), but no differences between controls and treatments at any time. In *S. australiensis*,
401 pairwise comparisons ~~revealed~~found a slightly lower respiration rate at day 12 compared to day 0 ($p =$
402 0.008) and day 5 ($p = 0.005$) in control sponges; while in treatment sponges, respiration rate was slightly
403 lower at day 5 ($p = 0.032$) and 12 ($p = 0.016$) compared to day 0, but also in this case, there was no
404 significant difference between treatments and controls at any time point.

405 4 DISCUSSION

406 Hypoxia has become an increasingly common problem in the marine environment and will likely become
407 worse in the future (Diaz & Rosenberg, 2011). Nevertheless, the direct effects of hypoxia on marine
408 organisms are still very poorly studied (Vaquer-Sunyer & Duarte, 2008). We describe the first multi-species
409 experiment from two oceans to test sponge tolerance, behaviour, and physiological responses to oxygen
410 concentrations as low as 1.5% a.s. ($0.13 \text{ mg O}_2 \text{ L}^{-1}$) for up to 12 days. We found that ~~irrespective of species~~
411 ~~or location, sponges are generally very tolerant to low DO~~ sponges are generally very tolerant to low DO
412 irrespective of species or location. Only, and only *Polymastia P.-croceus* showed mortality in the lower DO
413 treatment ($0.13 \text{ mg O}_2 \text{ L}^{-1}$, $\text{LT}_{50} = 286 \text{ h}$). Furthermore, our results suggest that sponges can display species-
414 specific acclimation, including physiological, morphological and behavioural changes, in response to severe
415 hypoxia that might help them survive periods of very low oxygen. Our study also suggests that the same
416 species can show different adaptive strategies for different degrees of hypoxia.

417 4.1 Sponge response to hypoxia

418 Our results suggest that sub-lethal oxygen thresholds for most sponges are in the range of 6–20% a.s.
419 ($0.48\text{--}1.56 \text{ mg O}_2 \text{ L}^{-1}$), while lethal thresholds are lower than 5% a.s. ($0.4 \text{ mg O}_2 \text{ L}^{-1}$). These pieces of
420 evidence are consistent with Mills et al. (2014) for *Halichondria panicea*, which showed a sub-lethal
421 response starting from 17% air saturation. However, our results contrast with Mills et al. (2018) studying
422 *Tethya wilhelma*, which did not show any response down to 4% a.s. ($0.27 \text{ mg O}_2 \text{ L}^{-1}$). The very high
423 tolerance of *T. wilhelma* could be explained by the extremely low metabolism of *Tethya* species generally
424 (Leys & Kahn, 2018), and by their very small size (0.5–1 cm) (Sarà et al., 2001). Of the two species we
425 exposed to the lowest DO concentration (1.5% a.s., $0.13 \text{ mg O}_2 \text{ L}^{-1}$), only *Polymastia P.-croceus* showed
426 mortality, while all the *Suberites S.-australiensis* survived the 12 days of treatment conditions. This
427 differential response could be due to the different habitats where these species are usually found.
428 *Polymastia croceus* lives on rocky reefs, while *S. australiensis* lives on sediments in bays and semi-enclosed
429 basins, where hypoxic events are more likely to occur (Diaz & Rosenberg, 2008; de Cook, 2010).

430 ~~Therefore, it is possible that *S. australiensis* have evolved adaptations to survive with very little oxygen.~~

431 Some sponges can live in anoxic conditions for several months, such as the sponges of the family
432 Raspailidae found in the deeper cliffs of Lough Hyne (Bell & Barnes, 2000; McAllen et al., 2009). Schuster et
433 al. (2021) suggested that this tolerance could be conferred by specific bacterial symbionts, which are able
434 to carry out anaerobic metabolism. In addition, these sponges living in anoxia are all thin crusts, with a very
435 high surface-to-volume ratio, which could favour the exchange of gases and the release of metabolic waste
436 (Levin et al., 1991). Other examples of sponges living in very low oxygen conditions are the ones found at
437 the edges of Oxygen Minimum Zones (OMZ) (Mosch et al., 2012). These sponges can live with a consistent
438 oxygen concentration as low as $0.13\text{mg O}_2\text{ L}^{-1}$ (Wishner et al., 1995, Murty et al., 2009). Sponges are not the
439 only organisms able to live in OMZs. Many representatives of other phyla live in these extremely hypoxic
440 conditions, where they benefit from the rich supply of organic matter. However, since OMZs have existed
441 over geological timescales, organisms have had the time to evolve specific adaptations to cope with
442 permanent hypoxia (Levin, 2003). Therefore, these organisms cannot be used to generalize tolerance to
443 periodic hypoxic events experienced by organisms usually living in fully oxygenated waters.

444 The degree of hypoxia tolerance in sponges could also be influenced by the abundance and diversity of
445 sponge-associated microbial symbionts-. Based on bacterial biomass, sponges are generally divided into
446 “low microbial abundance” (LMA) or “high microbial abundance” (HMA) species (Hentschel et al., 2003).
447 Bacterial densities in HMA sponges are generally two to four orders of magnitude higher than in LMA
448 sponges and can constitute up to 35% of the total sponge biomass (Vacelet, 1975; Hentschel et al., 2006).
449 Sponges with HMA tend to have a lower choanocyte chamber density, and a slower pumping rate
450 compared LMA sponges (Lavy et al., 2016), which means HMA species might have a lower ability to
451 ventilate in low oxygen conditions. Furthermore, HMA species generally have a higher metabolic cost than
452 LMA species, and therefore a higher oxygen requirement (Leys & Kahn, 2018). Although these differences
453 suggest that LMA sponges might be better adapted to hypoxic conditions, HMA species have a higher
454 diversity of microbial symbionts that could help them cope with low oxygen conditions (Hoffmann et al.,
455 2005, Lavy et al., 2016). All the sponges for which responses to hypoxia has been investigated so far are
456 LMA species (or are likely to be, based on known congeners, see Kamke et al., 2010; Mills et al., 2014;

457 Moitinho-Silva et al., 2017). Therefore, future research is needed to investigate the response of HMA
458 sponges to hypoxia and shed light on possible differences between LMA and HMA sponges and the
459 mechanisms involved.

460 Some organism's abilities to tolerate hypoxia result from their physiological ability to lower metabolism and
461 oxygen demand (McAllen et al., 1999; Altieri, 2019). Instead, other species switch from aerobic to
462 anaerobic metabolism or a combination of the two (Altieri & Diaz, 2019). Our results suggest that all our
463 species (except *Suberites carnosus*) have respiration rates at 5-6% a.s. that are comparable to sponges in
464 normoxic conditions. This is consistent with what was found in *Geodia barretti* and *H. panicea*, suggesting
465 that sponges have a common ability to uptake oxygen at very low concentrations in the surrounding
466 environment (Leys & Kahn, 2018). In *Cliona celata*, hypoxic water initially resulted in a decrease in the
467 respiration rate, which then increased back to pre-treatment levels after seven days of exposure. This
468 suggests that the sponges gradually adjusted to hypoxic conditions. In *S. carnosus*, instead, the respiration
469 rate remained stable after two days of exposure to low dissolved oxygen, but it more than halved after
470 seven days. This response may allow *S. carnosus* to cope with long periods of hypoxia, in which sponges
471 decrease their metabolism, as has been reported for other organisms (Hagerman, 1998; Mentel et al.,
472 2014). Although our study shows that sponges can perform aerobic metabolism when exposed to
473 extremely low oxygen concentrations, the presence of anaerobic metabolism cannot be excluded and
474 needs further investigation.

475 Sponge species exposed to the lowest DO concentrations (0.4 and 0.13 mg O₂ L⁻¹) also showed other
476 phenotypic modifications that could represent adaptive strategies to cope with hypoxia. In *Suberites S.*
477 *australiensis*, hypoxic water (0.4 mg O₂ L⁻¹) induced expansion of the sponge body and the aquiferous
478 system that lasted for the duration of the experiment. This expansion was likely semi-permanent as it
479 persisted after inducing the contraction and corresponded to a reorganization of the sponge aquiferous
480 system at the histological level. These behavioural and morphological changes are likely beneficial for the
481 sponge, as higher internal water flow corresponds to an increase in oxygen that can be taken up. The body
482 expansion was accompanied by a marked increase in the pumping rate that then dropped after two days.

483 The pumping rate increase could be a strategy to increase ventilation and oxygen availability, similarly to
484 other animals when exposed to hypoxic waters (Hagerman, 1998). However, the successive decrease in
485 pumping rate (after two days) and the gradual production of a membrane to close the oscula remains
486 unclear but could represent a trade-off between increasing ventilation and keeping the energetic cost of
487 pumping reasonable. In *S. australiensis*, body expansion is correlated with an increase in osculum area, and
488 osculum area is the main determinant of pumping rate in this and many other species (Morganti et al.,
489 2021; Goldstein et al., 2019). Perhaps the increase in pumping rate only represents a physiological
490 consequence of the body expansion and is then quickly brought back to normal, decreasing the osculum
491 size by producing an oscular membrane. These physiological and morphological changes of *S. australiensis*
492 described above was not present on sponges exposed to more severe hypoxia ($0.13 \text{ mg O}_2 \text{ L}^{-1}$). This could
493 mean that the same sponge species may display different adaptive strategies to cope with decreased
494 oxygen depending on the oxygen concentration. At 0.4 mg L^{-1} , oxygen might still be sufficient to support
495 regular metabolism, but sponges may need to increase the amount of water flowing through their bodies
496 to absorb the oxygen needed. However, 0.13 mg L^{-1} might be too low a DO concentration, and sponges
497 might decrease their metabolism to cope with lack of oxygen, similarly to other metazoans (Hagerman
498 1998, Mentel et al., 2014).

499 *Polymastia croceus* also showed a behavioural change in response to hypoxic conditions: hypoxic water at
500 0.4 mg L^{-1} induced the progressive expansion of sponge papillae (where inhalant and exhalant channels are
501 found), that was significantly greater than in the control sponges. It is unlikely that the papillae expansion
502 represents an increase in sponge filtering activity because the respiration rate was very similar in the
503 treatments and the controls. Therefore, sponges might expand their papillae to increase the volume
504 occupied by the aquiferous systems, as in the case of *Suberites S.-australiensis*, but also to access more
505 oxygenated water further from the bottom. A similar response occurred in sponges exposed to 0.13 mg L^{-1}
506 but with much more variability across specimens, and the statistical test did not detect any change.
507 Interestingly, sponges that survived after the 12-day treatment had a significantly higher ratio of expanded
508 papillae than sponges that died, suggesting that expansion might help cope with severe hypoxic conditions.

509 Along with behavioural changes, *Polymastia* underwent morphological modifications that could help to
510 tolerate low DO. Papillae become thinner and flattened, and some even spiralized. These modifications of
511 the papillae could increase the surface-to-volume ratio and help oxygen diffusion (Levin et al., 1991). The
512 elongation of papillae, which accompanies the thinning, could be an evolutionary relic of a process that
513 moved the inhalant pores of the papillae as far as possible from the surface. However, in the lowest DO
514 treatment, papillae often lost their vertical orientation and laid horizontally on the sponge surface. We
515 hypothesized that the new orientation of papillae was a consequence of their thinning process: probably
516 papillae ~~become~~ became so thin that they could not support their weight anymore. Interestingly, sponges
517 that developed modified papillae showed less mortality than sponges that did not, although the evidence is
518 not strong enough to claim this with confidence ($p = 0.057$). Therefore, these structures may not only
519 represent a stress response, but could provide an advantage to the sponge. Further research is needed on
520 this topic needed to elucidate the function of these structures.

521 Despite the remarkable tolerance of sponges to hypoxia observed in laboratory conditions, field
522 observations suggest that severe hypoxic/anoxic events can catastrophically affect sponge populations.
523 Mass mortalities of sponges following hypoxic/anoxic events have been reported both in temperate and
524 tropical ecosystems (Stachowitsch, 1984, Altieri et al., 2017; Chu et al., 2018; Johnson et al., 2018; Kealoha
525 et al., 2020). For example, in a hypoxic/anoxic event in the Gulf of Trieste, all the sponges living in several
526 hundred km² died within 2-3 days (Stachowitsch, 1984). Some anemones survived up to a week, but
527 virtually all organisms were dead within two weeks from the onset. -Altieri et al. (2017) also reported
528 widespread mortality of sponges and corals following a hypoxic event (~ 0.5 mg O₂ L⁻¹) that occurred in
529 Bocas del Toro, Panama. Since this study focused on corals, it is unclear what proportion of the sponges
530 were affected and if some species were more tolerant than others. These reports highlight that hypoxic
531 events, in their most severe form, leave no survivors.
532 Furthermore, it is possible that in natural conditions, other factors combine with low dissolved oxygen. For
533 example, a recent meta-analysis showed that in marine organisms, increased temperature reduces survival
534 times under hypoxia by 74% on average and increased median lethal concentration by 16% on average

535 (Vaquer-Sunyer & Duarte, 2011). Another meta-analysis showed that hydrogen sulphide (H₂S) also reduces
536 survival time of marine organisms under hypoxia by an average of 30% (Vaquer-Sunyer & Duarte, 2010).
537 Acidification was shown to have additive or synergistic negative effects combined with hypoxia (Gobler &
538 Baumann, 2016; Steckbauer et al., 2020). Since all these factors usually co-occur during hypoxic events, *in*
539 *situ* sponge thresholds to hypoxia could be lower than determined through single stressor laboratory
540 experiments (Diaz & Rosenberg, 1995; Steckbauer et al., 2020). Future experiments that evaluate the
541 combined effect of these factors will be crucial to understand the full response of sponges to hypoxia in
542 natural ecosystems.

543 Diel oxygen variation is another factor that could influence an organism's tolerance to hypoxia in natural
544 conditions. In the photic zone of marine ecosystems, dissolved oxygen generally increases during the day
545 because of photosynthesis and decreases at night because of aerobic respiration (Kroeker et al. 2019). The
546 amplitude of these diel fluctuations can sometimes lead to hypoxia or complete anoxia at night and,
547 supersaturation in peak sunny hours, or both (Diaz & Breitburg, 2009). These extreme oxygen dynamics
548 have been reported from a wide variety of macro- and micro-habitats from both tropical and temperate
549 ecosystems, such as intertidal reef platforms, tide pools, semi-enclosed basins, tropical lagoons and the
550 boundary layer around macroalgal canopies (Morris and Taylor, 1983, Frieder et al. 2012, Cornwall et al.,
551 2013, Gruber et al., 2017, Trowbridge et al. 2017, Hughes et al., 2020). Diurnal fluctuations in oxygen can
552 produce different responses from static exposure in a laboratory experiments, which may either
553 overestimate or underestimate the emergent effects of hypoxia in natural environments (Bumett & Stickle,
554 2001). Therefore, future experiments will need to account for current and future temporal variability in
555 oxygen concentration to accurately forecast the emergent ecological effects of deoxygenation (Kroeker et
556 al., 2019).

557 **4.2 Hypoxia Tolerance of sponges compared to other sessile organisms**

558 Marine organisms have very variable tolerance to low dissolved oxygen, with lethal thresholds ranging from
559 8.6 mg O₂ L⁻¹ for the first larval zoea stage of the crustacean *Cancer irroratus*, to resistance to complete
560 anoxia as in the case of the sea anemone *Metridium senile* and the oyster *Crassostrea virginica* (Wahl 1984;

561 Vaquer-Sunyer and Duarte, 2008). Sessile organisms are generally more tolerant than mobile ones, which is
562 likely due to them not being able to escape hypoxic conditions (Altieri & Diaz, 2019). Therefore, sessile
563 organisms that experience these conditions must have evolved other adaptive strategies to cope with
564 reduced oxygen (Diaz & Rosenberg, 1995).

565 Here we provide new evidence to support the hypothesis that sponges are one of the groups of sessile
566 organisms that are more tolerant to hypoxia and could be favoured in future deoxygenated oceans. Other
567 phyla, such as cnidarians and bivalves, include very tolerant species that can cope with prolonged periods
568 of anoxia (Fig. 13). This is not surprising since tolerance to severe hypoxia/anoxia is a widespread feature in
569 the animal world, and many organisms independently evolved this feature to cope with local conditions
570 (Hochachka & Lutz, 2001; Nilsson & Renshaw, 2004; Vaquer-Sunyer & Duarte, 2008). This ability is not
571 restricted to invertebrates and includes higher animals such as fish and reptiles (Milton & Prentice, 2007;
572 Vornanen et al., 2009).

573 What makes sponges unique as a phylum is their widespread tolerance to hypoxia, ~~with all the species~~
574 ~~investigated so far being.~~ All the species investigated so far have been -shown to cope with very low levels
575 of dissolved oxygen. In contrast, other phyla have a much wider range of tolerances, with some species
576 resistant to anoxia and others very sensitive to decreased oxygen (Fig. 5). For example, in sessile cnidarians,
577 lethal hypoxia thresholds range between 0 and 4 mg O₂ L⁻¹, while sublethal ones are between 0.71 and 4.56
578 mg O₂ L⁻¹ (Mangum, 1980; Dodds et al., 2007). In sessile bivalves, lethal thresholds range between 0 and 2
579 mg O₂ L⁻¹, with the sub-lethal threshold being 3.1 mg O₂ L⁻¹ for *Mytilus galloprovincialis* (de Zwaan et al.,
580 1991; Woo et al., 2013). Sponges, instead, show much less variation with known lethal thresholds that are
581 lower than 0.5 mg O₂ L⁻¹, and sublethal thresholds that range between 0.27 and 1.56 mg O₂ L⁻¹ (Mills et al.,
582 2014; 2018) (Fig. 5). It is worth noting that lethal thresholds are highly dependent on the time of exposure.
583 In the studies we considered, these ranged from a few days to weeks. However, there were no noticeable
584 differences evidence in the experimental duration and the median lethal time for the different organisms.
585 Therefore, we believe that differences in time of exposure do not represent a bias in our comparison. In

586 contrast, sublethal responses (e.g. changes in respiration rate, behaviour, and feeding activity) usually have
587 rapid time-to-onset, so they will likely be independent of exposure time.

588 The high tolerance of sponges to hypoxia compared to other organisms can be explained by the
589 evolutionary history of this group. Sponges are one of the most ancient groups of metazoans. ~~There is some~~
590 ~~evidence that they~~ They likely evolved before the Marinoan glaciation (657-645 million years
591 ago) ~~Neoproterozoic oxygenation event (635–630 Mya)~~, when oxygen was perhaps less than 10% of present
592 atmospheric concentration. ~~(Kump, 2008; Love et al., 2009; Maloof et al., 2010; Brocks et al., 2017; Whelan~~
593 ~~et al., 2017; Cole et al., 2020; Turner, 2021).~~ ~~Therefore, it is possible that M~~modern sponges might have
594 retained an ancestral condition concerning oxygen requirements (Mills et al., 2014; 2018). Therefore, it is
595 more likely that sSponges unable to survive severe hypoxia today (e.g., *P. croceus*) have lost certain key
596 ancestral adaptations to hypoxia, rather than hypoxia-tolerant lineages (e.g., *S. australiensis*) having
597 evolved relatively new capacities for hypoxia tolerance (Müller et al., 2012). Likewise, other animals which
598 might have evolved in similar conditions, such as ctenophores, also show great resistance to hypoxia
599 (Thuesen et al., 2005). Therefore, we speculate that sponges' long evolutionary history could give these
600 organisms an adaptive advantage in future deoxygenated oceans, since they may have experienced similar
601 conditions in past geological eras.

602

603

604 CONCLUSIONS

605 Overall, sponges show high tolerance to low dissolved oxygen compared to all the other phyla of sessile
606 marine organisms that have been studied. Species-specific phenotypic plasticity appears to help these
607 organisms to overcome hypoxic events, and future research will need to elucidate the mechanisms behind
608 these changes. This exceptional adaptive capacity of sponges could derive from their ancient evolutionary
609 origin and could confer sponges a competitive advantage in future deoxygenated oceans over other
610 organisms (Mills et al., 2014; Schuster et al., 2021).

For Review Only

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619 AUTHOR CONTRIBUTIONS

620 V.M., J.J.B and R.M. designed the study. V.M., F.S. and L.H. realized the experimental set-up. V.M. and F.S.
621 conducted the experiments. V.M. and L.W. analyzed the data. V.M. and J.J.B wrote the original draft. All the
622 authors participated in interpreting the results and contributed to the revision of the manuscript.

623 DATA AVAILABILITY STATEMENT

624 All data and the R code used in this paper are available on Figshare repository:
625 <https://doi.org/10.6084/m9.figshare.15169662>.

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633 **REFERENCES**

- 634 Altieri, A. H., & Diaz, R. J. (2019). Dead zones: oxygen depletion in coastal ecosystems. In: C. Sheppard
635 (Ed.), *World Seas: An Environmental Evaluation* (pp. 453–473). Academic Press.
636 <https://doi.org/10.1016/b978-0-12-805052-1.00021-8>
- 637 Altieri, A. H., & Gedan, K. B. (2015). Climate change and dead zones. *Global Change Biology*, 21(4),
638 1395–1406. <https://doi.org/10.1111/gcb.12754>
- 639 Altieri, A. H., Harrison, S. B., Seemann, J., Collin, R., Diaz, R. J., & Knowlton, N. (2017). Tropical dead
640 zones and mass mortalities on coral reefs. *Proceedings of the National Academy of Sciences*, 114(14),
641 3660–3665. <https://doi.org/10.1073/pnas.1621517114>
- 642 Anderson, M. J. (2001). A new method for non-parametric multivariate analysis of variance. *Austral*
643 *Ecology*, 26(1), 32–46. <https://doi.org/10.1111/j.1442-9993.2001.01070.pp.x>
- 644 Anderson, M. J. (2014). Permutational multivariate analysis of variance (PERMANOVA). *Wiley statsref:*
645 *statistics reference online*, 1–15. <https://doi.org/10.1002/9781118445112.stat07841>
- 646 Anderson, M. J., Gorley, R. N. & Clarke, K. R. (2008). *PERMANOVA+ for PRIMER: Guide to software and*
647 *statistical methods*. PRIMER-E, Plymouth, UK.
- 648 Ayling, A. L. (1983). Factors affecting the spatial distributions of thinly encrusting sponges from
649 temperate waters. *Oecologia*, 60(3), 412–418. <https://doi.org/10.1007/bf00376861>
- 650 Bates, D., Maechler, M., Bolker, B., & Walker, S. (2015). Fitting Linear Mixed-Effects Models Using lme4.
651 *Journal of Statistical Software*, 67(1), 1–48. <https://doi:10.18637/jss.v067.i01>
- 652 Bell, J. J. (2008). The functional roles of marine sponges. *Estuarine, Coastal and Shelf Science*, 79(3),
653 341–353. <https://doi.org/10.1016/j.ecss.2008.05.002>

- 654 Bell, J. J., & Barnes, D. K. (2000). The distribution and prevalence of sponges in relation to
655 environmental gradients within a temperate sea lough: vertical cliff surfaces. *Diversity and*
656 *Distributions*, 6(6), 283–303. <https://doi.org/10.1046/j.1472-4642.2000.00091.x>
- 657 Bell, J. J., McGrath, E., Kandler, N. M., Marlow, J., Beepat, S. S., Bachtiar, R., Shaffer, M. R., Mortimer, C.,
658 Micaroni, V., Mobilia, V., Rovellini, A., Harris, B., Farnham, E., Strano, F., & Carballo, J. L. (2020).
659 Interocean patterns in shallow water sponge assemblage structure and function. *Biological*
660 *Reviews*, 95(6), 1720–1758. <https://doi.org/10.1111/brv.12637>
- 661 Bijma, J., Pörtner, H. O., Yesson, C., & Rogers, A. D. (2013). Climate change and the oceans—What does
662 the future hold?. *Marine Pollution Bulletin*, 74(2), 495–505.
663 <https://doi.org/10.1016/j.marpolbul.2013.07.022>
- 664 [Breitburg, D. L. \(1990\). Near-shore hypoxia in the Chesapeake Bay: patterns and relationships among](https://doi.org/10.1016/0272-7714(90)90095-9)
665 [physical factors. *Estuarine, Coastal and Shelf Science*, 30\(6\), 593–609. \[https://doi.org/10.1016/0272-\]\(https://doi.org/10.1016/0272-7714\(90\)90095-9\)](https://doi.org/10.1016/0272-7714(90)90095-9)
666 [7714\(90\)90095-9](https://doi.org/10.1016/0272-7714(90)90095-9)
- 667 [Brooks, J. J., Jarrett, A. J., Sirantoine, E., Hallmann, C., Hoshino, Y., & Liyanage, T. \(2017\). The rise of](https://doi.org/10.1038/nature23457)
668 [algae in Cryogenian oceans and the emergence of animals. *Nature*, 548\(7669\), 578-581.](https://doi.org/10.1038/nature23457)
669 [https://doi.org/doi:10.1038/nature23457](https://doi.org/10.1038/nature23457)
- 670 Brooks, M. E., Kristensen, K., Van Benthem, K. J., Magnusson, A., Berg, C. W., Nielsen, A., Skaug, H. J.,
671 Machler, M. & Bolker, B.M. (2017). glmmTMB balances speed and flexibility among packages for zero-
672 inflated generalized linear mixed modeling. *The R Journal*, 9(2), 378–400. [https://doi.org/10.32614/rj-](https://doi.org/10.32614/rj-2017-066)
673 [2017-066](https://doi.org/10.32614/rj-2017-066)
- 674 [Bumett, L. E., & Stickle, W. B. \(2001\). Physiological responses to hypoxia. In: N. N. Rabalais, R. E. Turner](https://doi.org/10.1029/ce058p0101)
675 [\(Eds\). *Coastal and Estuarine Studies, Volume 58. Coastal hypoxia: consequences for living resources and*](https://doi.org/10.1029/ce058p0101)
676 [*ecosystems*, \(pp. 101–114\). AGU Publications. <https://doi.org/10.1029/ce058p0101>](https://doi.org/10.1029/ce058p0101)

- 677 Chu, J. W., Curkan, C., & Tunnicliffe, V. (2018). Drivers of temporal beta diversity of a benthic
678 community in a seasonally hypoxic fjord. *Royal Society Open Science*, 5(4), 172284.
679 <https://doi.org/10.1098/rsos.172284>
- 680 Clarke, K. R., & Gorley, R. N. (2015). *Getting started with PRIMER v7*. PRIMER-E, Plymouth, UK.
- 681 [Cole, D. B., Mills, D. B., Erwin, D. H., Sperling, E. A., Porter, S. M., Reinhard, C. T., & Planavsky, N. J.](#)
682 [\(2020\). On the co-evolution of surface oxygen levels and animals. *Geobiology*, 18\(3\), 260–281.](#)
683 <https://doi.org/10.1111/gbi.12382>
- 684 Cook, S. D. C. (2010). *New Zealand coastal marine invertebrates*. Canterbury University Press.
- 685 [Cornwall, C. E., Hepburn, C. D., Pilditch, C. A., & Hurd, C. L. \(2013\). Concentration boundary layers](#)
686 [around complex assemblages of macroalgae: Implications for the effects of ocean acidification on](#)
687 [understory coralline algae. *Limnology and Oceanography*, 58\(1\), 121–130.](#)
688 <https://doi.org/10.4319/lo.2013.58.1.0121>
- 689 De Zwaan, A., Cortesi, P., Van den Thillart, G., Roos, J., & Storey, K. B. (1991). Differential sensitivities to
690 hypoxia by two anoxia-tolerant marine molluscs: a biochemical analysis. *Marine Biology*, 111(3), 343–
691 351. <https://doi.org/10.1007/bf01319405>
- 692 Desai, D. V., & Prakash, S. (2009). Physiological responses to hypoxia and anoxia in *Balanus amphitrite*
693 (Cirripedia: Thoracica). *Marine Ecology Progress Series*, 390, 157–166.
694 <https://doi.org/10.3354/meps08155>
- 695 [Diaz, R. J., & Breitburg, D. L. \(2009\). The hypoxic environment. In: J. G. Richards, A. P. Farrell, C. J.](#)
696 [Brauner \(Eds\) *Fish physiology, Volume 27* \(pp. 1–23\). Academic Press. \[12-809633-8.03213-1\]\(https://doi.org/10.1016/b978-0-
697 <a href=\)](#)
- 698 Diaz, R. J., & Rosenberg, R. (1995). Marine benthic hypoxia: a review of its ecological effects and the
699 behavioural responses of benthic macrofauna. *Oceanography and marine biology. An annual review*,
700 33, 245–03.

- 701 Diaz, R. J., & Rosenberg, R. (2008). Spreading dead zones and consequences for marine
702 ecosystems. *Science*, 321(5891), 926–929. <https://doi.org/10.1126/science.1156401>
- 703 Diaz, R. J., & Rosenberg, R. (2011). Introduction to environmental and economic consequences of
704 hypoxia. *International Journal of Water Resources Development*, 27(1), 71–82.
705 <https://doi.org/10.1080/07900627.2010.531379>
- 706 Dodds, L. A., Roberts, J. M., Taylor, A. C., & Marubini, F. (2007). Metabolic tolerance of the cold-water
707 coral *Lophelia pertusa* (Scleractinia) to temperature and dissolved oxygen change. *Journal of*
708 *Experimental Marine Biology and Ecology*, 349(2), 205–214.
709 <https://doi.org/10.1016/j.jembe.2007.05.013>
- 710 Doney, S. C. (2010). The growing human footprint on coastal and open-ocean biogeochemistry. *Science*,
711 328(5985), 1512–1516. <https://doi.org/10.1126/science.1185198>
- 712 Ellington, W. R. (1982). Metabolic responses of the sea anemone *Bunodosoma cavernata* (Bosc) to
713 declining oxygen tensions and anoxia. *Physiological Zoology*, 55(3), 240–249.
714 <https://doi.org/10.1086/physzool.55.3.30157888>
- 715 [Frieder, C. A., Nam, S. H., Martz, T. R., & Levin, L. A. \(2012\). High temporal and spatial variability of](#)
716 [dissolved oxygen and pH in a nearshore California kelp forest. *Biogeosciences*, 9\(10\), 3917–3930.](#)
717 <https://doi.org/10.5194/bg-9-3917-2012>
- 718 Gobler, C. J., & Baumann, H. (2016). Hypoxia and acidification in ocean ecosystems: coupled dynamics
719 and effects on marine life. *Biology Letters*, 12(5), 20150976. <https://doi.org/10.1098/rsbl.2015.0976>
- 720 Goldstein, J., Riisgård, H. U., & Larsen, P. S. (2019). Exhalant jet speed of single-ostium explants of the
721 demosponge *Halichondria panicea* and basic properties of the sponge-pump. *Journal of Experimental*
722 *Marine Biology and Ecology*, 511, 82–90. <https://doi.org/10.1016/j.jembe.2018.11.009>

- 723 [Gruber, R. K., Lowe, R. J., & Falter, J. L. \(2017\). Metabolism of a tide-dominated reef platform subject to](#)
724 [extreme diel temperature and oxygen variations. *Limnology and Oceanography*, 62\(4\), 1701–1717.](#)
725 <https://doi.org/10.1002/lno.10527>
- 726 Haas, A. F., Smith, J. E., Thompson, M., & Deheyn, D. D. (2014). Effects of reduced dissolved oxygen
727 concentrations on physiology and fluorescence of hermatypic corals and benthic algae. *PeerJ*, 2, e235.
728 <https://doi.org/10.7717/peerj.235>
- 729 Hagerman, L. (1998). Physiological flexibility; a necessity for life in anoxic and sulphidic habitats.
730 *Hydrobiologia*, 375, 241–254. https://doi.org/10.1007/978-94-017-2864-5_20
- 731 Hervé, M. (2021). *RVAideMemoire: testing and plotting procedures for biostatistics*. R package version
732 0.9-79. <https://cran.r-project.org/web/packages/RVAideMemoire/index.html>
- 733 [Hentschel, U., Fieseler, L., Wehrl, M., Gernert, C., Steinert, M., Hacker, J., & Horn, M. \(2003\). Microbial](#)
734 [diversity of marine sponges. In: W. E. G. Müller \(Ed\) *Progress in Molecular and Subcellular Biology.*](#)
735 [*Sponges \(Porifera\)* \(pp 59–88\). Springer. \[https://doi.org/10.1007/978-3-642-55519-0_3\]\(https://doi.org/10.1007/978-3-642-55519-0_3\)](#)
- 736 [Hentschel, U., Usher, K. M., & Taylor, M. W. \(2006\). Marine sponges as microbial fermenters. *FEMS*](#)
737 [*microbiology ecology*, 55\(2\), 167–177. <https://doi.org/10.1111/j.1574-6941.2005.00046.x>](#)
- 738 Hochachka, P. W., & Lutz, P. L. (2001). Mechanism, origin, and evolution of anoxia tolerance in animals.
739 *Comparative Biochemistry and Physiology Part B: Biochemistry and Molecular Biology*, 130(4), 435–459.
740 [https://doi.org/10.1016/s1096-4959\(01\)00408-0](https://doi.org/10.1016/s1096-4959(01)00408-0)
- 741 [Hoffmann, F., Larsen, O., Thiel, V., Rapp, H. T., Pape, T., Michaelis, W., & Reitner, J. \(2005\). An](#)
742 [anaerobic world in sponges. *Geomicrobiology Journal*, 22\(1–2\), 1–10.](#)
743 <https://doi.org/10.1080/01490450590922505>
- 744 Hughes, D. J., Alderdice, R., Cooney, C., Kühl, M., Pernice, M., Voolstra, C. R., & Suggett, D. J. (2020).
745 Coral reef survival under accelerating ocean deoxygenation. *Nature Climate Change*, 10(4), 296–307.
746 <https://doi.org/10.1038/s41558-020-0737-9>

- 747 IPCC (2021). Summary for Policymakers. In: V. Masson-Delmotte, P. Zhai, A. Pirani, S. L. Connors, C.
748 Péan, S. Berger, N. Caud, Y. Chen, L. Goldfarb, M. I. Gomis, M. Huang, K. Leitzell, E. Lonnoy, J.B.R.
749 Matthews, T. K. Maycock, T. Waterfield, O. Yelekçi, R. Yu & B. Zhou (Eds.). *Climate Change 2021: The*
750 *Physical Science Basis. Contribution of Working Group I to the Sixth Assessment Report of the*
751 *Intergovernmental Panel on Climate Change*. Cambridge University Press. In Press.
- 752 Johnson, M. D., Rodriguez, L. M., & Altieri, A. H. (2018). Shallow-water hypoxia and mass mortality on a
753 Caribbean coral reef. *Bulletin of Marine Science*, 94(1), 143–144.
754 <https://doi.org/10.5343/bms.2017.1163>
- 755 Johnson, P. D., & McMahon, R. F. (1998). Effects of temperature and chronic hypoxia on survivorship of
756 the zebra mussel (*Dreissena polymorpha*) and Asian clam (*Corbicula fluminea*). *Canadian Journal of*
757 *Fisheries and Aquatic Sciences*, 55(7), 1564–1572. <https://doi.org/10.1139/f98-030>
- 758 [Kamke, J., Taylor, M. W., & Schmitt, S. \(2010\). Activity profiles for marine sponge-associated bacteria](#)
759 [obtained by 16S rRNA vs 16S rRNA gene comparisons. *The ISME Journal*, 4\(4\), 498–508.](#)
760 <https://doi.org/10.1038/ismej.2009.143>
- 761 Kealoha, A. K., Doyle, S. M., Shamberger, K. E., Sylvan, J. B., Hetland, R. D., & Di Marco, S. F. (2020).
762 Localized hypoxia may have caused coral reef mortality at the Flower Garden Banks. *Coral Reefs*, 39(1),
763 119–132. <https://doi.org/10.1007/s00338-019-01883-9>
- 764 Keeling, R. F., Körtzinger, A., & Gruber, N. (2010). Ocean deoxygenation in a warming world. *Annual*
765 *Review of Marine Science*, 2, 199–229. <https://doi.org/10.1146/annurev.marine.010908.163855>
- 766 [Kroeker, K. J., Bell, L. E., Donham, E. M., Hoshijima, U., Lummis, S., Toy, J. A., & Willis-Norton, E. \(2020\).](#)
767 [Ecological change in dynamic environments: Accounting for temporal environmental variability in](#)
768 [studies of ocean change biology. *Global change biology*, 26\(1\), 54–67.](#)
769 <https://doi.org/10.1111/gcb.14868>
- 770 [Kump, L. R. \(2008\). The rise of atmospheric oxygen. *Nature*, 451\(7176\), 277–278.](#)
771 <https://doi.org/10.1038/nature06587>

- 772 Kuznetsova, A., Brockhoff, P. B., & Christensen, R. H. (2017). lmerTest package: tests in linear mixed
773 effects models. *Journal of Statistical Software*, 82(13), 1–26. <https://doi.org/10.18637/jss.v082.i13>
- 774 [Lavy, A., Keren, R., Yahel, G., & Ilan, M. \(2016\). Intermittent hypoxia and prolonged suboxia measured](#)
775 [in situ in a marine sponge. *Frontiers in Marine Science*, 3, 263.](#)
776 <https://doi.org/10.3389/fmars.2016.00263>
- 777 Lenth, R. (2021). *emmeans: estimated marginal means, aka least-squares means*. R package version
778 1.6.0. <https://CRAN.R-project.org/package=emmeans>
- 779 Levin, L. A. (2003). Oxygen minimum zone benthos: adaptation and community response to hypoxia.
780 *Oceanography and Marine Biology: An Annual Review*. 41, 1–45.
781 <https://doi.org/10.1201/9780203180570-3>
- 782 Levin, L. A., & Breitburg, D. L. (2015). Linking coasts and seas to address ocean deoxygenation. *Nature*
783 *Climate Change*, 5(5), 401–403. <https://doi.org/10.1038/nclimate2595>
- 784 Levin, L. A., Ekau, W., Gooday, A. J., Jorissen, F., Middelburg, J. J., Naqvi, S. W. A., Neira, C., Rabalais, N.
785 N. & Zhang, J. (2009). Effects of natural and human-induced hypoxia on coastal benthos.
786 *Biogeosciences*, 6(10), 2063–2098. <https://doi.org/10.5194/bg-6-2063-2009>
- 787 Levin, L. A., Huggett, C. L., & Wishner, K. F. (1991). Control of deep-sea benthic community structure by
788 oxygen and organic-matter gradients in the eastern Pacific Ocean. *Journal of Marine Research*, 49(4),
789 763–800. <https://doi.org/10.1357/002224091784995756>
- 790 Leys, S. P., & Kahn, A. S. (2018). Oxygen and the energetic requirements of the first multicellular
791 animals. *Integrative and Comparative Biology*, 58(4), 666–676. <https://doi.org/10.1093/icb/icy051>
- 792 Limburg, K. E., Breitburg, D., & Levin, L. A. (2017). Ocean deoxygenation-a climate-related problem.
793 *Frontiers in Ecology and the Environment*, 15(9), 479–479. <https://doi.org/10.1002/fee.1728>
- 794 Love, G.D., Grosjean, E., Stalvies, C., Fike, D.A., Grotzinger, J.P., Bradley, A.S., Kelly, A.E., Bhatia, M.,
795 Meredith, W., Snape, C.E. and Bowring, S.A., Condon, D. J., & Summons, R. E. (2009). Fossil steroids

- 796 record the appearance of Demospongiae during the Cryogenian period. *Nature*, 457(7230), 718–721.
797 <https://doi.org/10.1038/nature07673>
- 798 Lunden, J. J., McNicholl, C. G., Sears, C. R., Morrison, C. L., & Cordes, E. E. (2014). Acute survivorship of
799 the deep-sea coral *Lophelia pertusa* from the Gulf of Mexico under acidification, warming, and
800 deoxygenation. *Frontiers in Marine Science*, 1, 78. <https://doi.org/10.3389/fmars.2014.00078>
- 801 Maldonado, M., Ribes, M., & van Duyl, F. C. (2012). Nutrient fluxes through sponges: biology, budgets,
802 and ecological implications. *Advances in Marine Biology*, 62, 113–182. [https://doi.org/10.1016/b978-0-](https://doi.org/10.1016/b978-0-12-394283-8.00003-5)
803 [12-394283-8.00003-5](https://doi.org/10.1016/b978-0-12-394283-8.00003-5)
- 804 Maloof, A. C., Rose, C. V., Beach, R., Samuels, B. M., Calmet, C. C., Erwin, D. H., Poirier, G. R., Yao, N., &
805 Simons, F.J. (2010). Possible animal-body fossils in pre-Marinoan limestones from South Australia.
806 *Nature Geoscience*, 3(9), 653–659. <https://doi.org/10.1038/ngeo934>
- 807 Mangum, D. C. (1980). Sea anemone neuromuscular responses in anaerobic conditions. *Science*,
808 208(4448), 1177–1178. <https://doi.org/10.1126/science.6103580>
- 809 McAllen, R., Davenport, J., Bredendieck, K., & Dunne, D. (2009). Seasonal structuring of a benthic
810 community exposed to regular hypoxic events. *Journal of Experimental Marine Biology and Ecology*,
811 368(1), 67–74. <https://doi.org/10.1016/j.jembe.2008.10.019>
- 812 McAllen, R., Taylor, A. C., & Davenport, J. (1999). The effects of temperature and oxygen partial
813 pressure on the rate of oxygen consumption of the high-shore rock pool copepod *Tigriopus brevicornis*.
814 *Comparative Biochemistry and Physiology Part A: Molecular & Integrative Physiology*, 123(2), 195–202.
815 [https://doi.org/10.1016/s1095-6433\(99\)00050-1](https://doi.org/10.1016/s1095-6433(99)00050-1)
- 816 Mentel, M., Röttger, M., Leys, S., Tielens, A. G., & Martin, W. F. (2014). Of early animals, anaerobic
817 mitochondria, and a modern sponge. *Bioessays*, 36(10), 924–932.
818 <https://doi.org/10.1002/bies.201400060>

- 819 Mills, D. B., Francis, W. R., Vargas, S., Larsen, M., Elemans, C. P., Canfield, D. E., & Wörheide, G. (2018).
820 The last common ancestor of animals lacked the HIF pathway and respired in low-oxygen
821 environments. *Elife*, 7, e31176. <https://doi.org/10.7554/eLife.31176.001>
- 822 Mills, D. B., Ward, L. M., Jones, C., Sweeten, B., Forth, M., Treusch, A. H., & Canfield, D. E. (2014).
823 Oxygen requirements of the earliest animals. *Proceedings of the National Academy of Sciences*,
824 111(11), 4168–4172. <https://doi.org/10.1073/pnas.1400547111>
- 825 Milton, S. L., & Prentice, H. M. (2007). Beyond anoxia: the physiology of metabolic downregulation and
826 recovery in the anoxia-tolerant turtle. *Comparative Biochemistry and Physiology Part A: Molecular and*
827 *Integrative Physiology*, 147(2), 277–290. <https://doi.org/10.1016/j.cbpa.2006.08.041>
- 828 [Moitinho-Silva, L., Steinert, G., Nielsen, S., Hardoim, C. C., Wu, Y. C., McCormack, G. P., López-Legendil,](#)
829 [S., Marchant, R., Webster, N., Thomas, T., & Hentschel, U. \(2017\). Predicting the HMA-LMA status in](#)
830 [marine sponges by machine learning. *Frontiers in microbiology*, 8, 752.](#)
831 <https://doi.org/10.3389/fmicb.2017.00752>
- 832 Morganti, T. M., Ribes, M., Moskovich, R., Weisz, J. B., Yahel, G., & Coma, R. (2021). In situ Pumping
833 Rate of 20 Marine Demosponges Is a Function of Osculum Area. *Frontiers in Marine Science*, 8, 583188.
834 <https://doi.org/10.3389/fmars.2021.583188>
- 835 [Morris, S., & Taylor, A. C. \(1983\). Diurnal and seasonal variation in physico-chemical conditions within](#)
836 [intertidal rock pools. *Estuarine, Coastal and Shelf Science*, 17, 339–355. \[https://doi.org/10.1016/0272-\]\(https://doi.org/10.1016/0272-7714\(83\)90026-4\)](#)
837 [7714\(83\)90026-4](#)
- 838 Mosch, T., Sommer, S., Dengler, M., Noffke, A., Bohlen, L., Pfannkuche, O., Liebetrau, V. & Wallmann, K.
839 (2012). Factors influencing the distribution of epibenthic megafauna across the Peruvian oxygen
840 minimum zone. *Deep Sea Research Part I: Oceanographic Research Papers*, 68, 123–135.
841 <https://doi.org/10.1016/j.dsr.2012.04.014>
- 842 [Müller, M., Mentel, M., van Hellemond, J. J., Henze, K., Woehle, C., Gould, S. B., Yu, R.Y., van der](#)
843 [Giezen, M., Tielens, A.G., Martin, W. F. \(2012\). Biochemistry and evolution of anaerobic energy](#)

- 844 [metabolism in eukaryotes. *Microbiology and Molecular Biology Reviews*, 76\(2\), 444–495.](#)
- 845 <https://doi.org/10.1128/MMBR.05024-11>
- 846 Murty, S. J., Bett, B. J., & Gooday, A. J. (2009). Megafaunal responses to strong oxygen gradients on the
- 847 Pakistan margin of the Arabian Sea. *Deep Sea Research Part II: Topical Studies in Oceanography*, 56(6-
- 848 7), 472–487. <https://doi.org/10.1016/j.dsr2.2008.05.029>
- 849 Nagasoe, S., Tokunaga, T., Yurimoto, T., & Matsuyama, Y. (2020). Survival and behavior patterns
- 850 associated with hypoxia at different life stages of the pen shell *Atrina cf. japonica*. *Aquatic Toxicology*,
- 851 227, 105610. <https://doi.org/10.1016/j.aquatox.2020.105610>
- 852 [Nezlin, N. P., Kamer, K., Hyde, J., & Stein, E. D. \(2009\). Dissolved oxygen dynamics in a eutrophic](#)
- 853 [estuary, Upper Newport Bay, California. *Estuarine, Coastal and Shelf Science*, 82\(1\), 139–151.](#)
- 854 <https://doi.org/10.1016/j.ecss.2009.01.004>
- 855 Nilsson, G. E., & Renshaw, G. M. (2004). Hypoxic survival strategies in two fishes: extreme anoxia
- 856 tolerance in the North European crucian carp and natural hypoxic preconditioning in a coral-reef shark.
- 857 *Journal of Experimental Biology*, 207(18), 3131–3139. <https://doi.org/10.1242/jeb.00979>
- 858 Nixon, S. W. (1995). Coastal marine eutrophication: a definition, social causes, and future concerns.
- 859 *Ophelia*, 41(1), 199–219. <https://doi.org/10.1080/00785236.1995.10422044>
- 860 Osinga, R., Tramper, J., & Wijffels, R. H. (1999). Cultivation of marine sponges. *Marine Biotechnology*,
- 861 1(6), 509–532. <https://doi.org/10.1007/PL00011807>
- 862 Piazzzi, L., Atzori, F., Cadoni, N., Cinti, M. F., Frau, F., Pansini, A., Pinna, F., Stipcich, P. & Ceccherelli, G.
- 863 (2021). Animal Forest Mortality: Following the Consequences of a Gorgonian Coral Loss on a
- 864 Mediterranean Coralligenous Assemblage. *Diversity*, 13(3), 133. <https://doi.org/10.3390/d13030133>
- 865 Pinheiro, J., Bates, D., DebRoy, S., Sarkar, D., & R Core Team (2021). *nlme: Linear and Nonlinear Mixed*
- 866 *Effects Models*. R package version 3.1-152. <https://CRAN.R-project.org/package=nlme>.
- 867 R Core Team (2013). *R: A language and environment for statistical computing*.

- 868 Rao, D. P., & Ganapati, P. N. (1968). Respiration as a function of oxygen concentration in intertidal
869 barnacles. *Marine Biology*, 1(4), 309–310. <https://doi.org/10.1007/BF00360781>
- 870 Raupach, M. R., & Canadell, J. G. (2010). Carbon and the Anthropocene. *Current Opinion in*
871 *Environmental Sustainability*, 2(4), 210–218. <https://doi.org/10.1016/j.cosust.2010.04.003>
- 872 Robinson, C. (2019). Microbial respiration, the engine of ocean deoxygenation. *Frontiers in Marine*
873 *Science*, 5, 533. <https://doi.org/10.3389/fmars.2018.00533>
- 874 Sampaio, E., Santos, C., Rosa, I. C., Ferreira, V., Pörtner, H. O., Duarte, C. M., Levin, L. A. & Rosa, R.
875 (2021). Impacts of hypoxic events surpass those of future ocean warming and acidification. *Nature*
876 *Ecology and Evolution*, 5(3), 311–321. <https://doi.org/10.1038/s41559-020-01370-3>
- 877 Sarà, M., Sarà, A., Nickel, M., & Brümmer, F. (2001). Three New Species of *Tethya* (Porifera:
878 Demospongiae) From German Aquaria. *Stuttgarter Beiträge zur Naturkunde* 631, 1–15.
- 879 Sassaman, C., & Mangum, C. P. (1972). Adaptations to environmental oxygen levels in infaunal and
880 epifaunal sea anemones. *The Biological Bulletin*, 143(3), 657–678. <https://doi.org/10.2307/1540189>
- 881 Schuster, A., Strehlow, B. W., Eckford-Soper, L., McAllen, R., & Canfield, D. E. (2021). Effects of Seasonal
882 Anoxia on the Microbial Community Structure in Demosponges in a Marine Lake in Lough Hyne, Ireland.
883 *Mosphere*, 6(1), e00991-20. <https://doi.org/10.1101/2020.09.09.290791>
- 884 Seibel, B. A. (2011). Critical oxygen levels and metabolic suppression in oceanic oxygen minimum
885 zones. *Journal of Experimental Biology*, 214(2), 326–336. <https://doi.org/10.1242/jeb.049171>
- 886 Semenza, G. L. (2007). Life with oxygen. *Science*, 318(5847), 62–64.
887 <https://doi.org/10.1126/science.1147949>
- 888 Smith, V. H., Joye, S. B., & Howarth, R. W. (2006). Eutrophication of freshwater and marine ecosystems.
889 *Limnology and Oceanography*, 51, 351–355. https://doi.org/10.1007/978-90-481-3385-7_1
- 890 Stachowitsch, M. (1984). Mass mortality in the Gulf of Trieste: the course of community destruction.
891 *Marine Ecology*, 5(3), 243–264. <https://doi.org/10.1111/j.1439-0485.1984.tb00124.x>

- 892 Steckbauer, A., Klein, S. G., & Duarte, C. M. (2020). Additive impacts of deoxygenation and acidification
893 threaten marine biota. *Global Change Biology*, 26(10), 5602–5612. <https://doi.org/10.1111/gcb.15252>
- 894 Stickle, W. B., Kapper, M. A., Liu, L. L., Gnaiger, E., & Wang, S. Y. (1989). Metabolic adaptations of
895 several species of crustaceans and molluscs to hypoxia: tolerance and microcalorimetric studies. *The*
896 *Biological Bulletin*, 177(2), 303–312. <https://doi.org/10.2307/1541945>
- 897 Strano, F., Micaroni, V., Davy, S. K., Maldonado, M., & Bell, J. J. (2021). Reproduction and early life
898 stages of the poecilosclerid sponge *Crella incrustans*. *Invertebrate Biology*, e12335.
899 <https://doi.org/10.1111/ivb.12335>
- 900 Therneau, T. (2021). *A Package for Survival Analysis in R*. R package version 3.2-10. [https://CRAN.R-](https://CRAN.R-project.org/package=survival)
901 [project.org/package=survival](https://CRAN.R-project.org/package=survival).
- 902 Thuesen, E. V., Rutherford Jr, L. D., & Brommer, P. L. (2005). The role of aerobic metabolism and
903 intragel oxygen in hypoxia tolerance of three ctenophores: *Pleurobrachia bachei*, *Bolinopsis*
904 *infundibulum* and *Mnemiopsis leidyi*. *Journal of the Marine Biological Association of the United*
905 *Kingdom*, 85(3), 627. <https://doi.org/10.1017/s0025315405011550>
- 906 [Trowbridge, C. D., Davenport, J., Cottrell, D. M., Harman, L., Plowman, C. Q., Little, C., & McAllen, R.](#)
907 [\(2017\). Extreme oxygen dynamics in shallow water of a fully marine Irish sea lough. *Regional Studies in*](#)
908 [Marine Science](#), 11, 9–16. <https://doi.org/10.1016/j.rsma.2017.01.008>
- 909 Turner, E. C. (2021). Possible poriferan body fossils in early Neoproterozoic microbial reefs. *Nature*,
910 596, 87–91. <https://doi.org/10.1038/s41586-021-03773-z>
- 911 [Vacelet, J. \(1975\). Étude en microscopie électronique de l'association entre bactéries et spongiaires du](#)
912 [genre *Verongia* \(Dictyoceratida\).](#)
- 913 Vaquer-Sunyer, R., & Duarte, C. M. (2008). Thresholds of hypoxia for marine biodiversity. *Proceedings*
914 *of the National Academy of Sciences*, 105(40), 15452–15457.
915 <https://doi.org/10.1073/pnas.0803833105>

- 916 Vaquer-Sunyer, R., & Duarte, C. M. (2010). Sulfide exposure accelerates hypoxia-driven mortality.
917 *Limnology and Oceanography*, 55(3), 1075–1082. <https://doi.org/10.4319/lo.2010.55.3.1075>
- 918 Vaquer-Sunyer, R., & Duarte, C. M. (2011). Temperature effects on oxygen thresholds for hypoxia in
919 marine benthic organisms. *Global Change Biology*, 17(5), 1788–1797. <https://doi.org/10.1111/j.1365-2486.2010.02343.x>
- 921 Vergés, A., McCosker, E., Mayer-Pinto, M., Coleman, M. A., Wernberg, T., Ainsworth, T., & Steinberg, P.
922 D. (2019). Tropicalisation of temperate reefs: implications for ecosystem functions and management
923 actions. *Functional Ecology*, 33(6), 1000–1013. <https://doi.org/10.1111/1365-2435.13310>
- 924 Vornanen, M., Stecyk, J. A., & Nilsson, G. E. (2009). The anoxia-tolerant crucian carp (*Carassius*
925 *carassius* L.). In: J. Richards, A. Farrell, & C. Brauner (Eds.), *Fish Physiology: Hypoxia, Volume 27* (pp.
926 397–441). Academic Press. [https://doi.org/10.1016/s1546-5098\(08\)00009-5](https://doi.org/10.1016/s1546-5098(08)00009-5)
- 927 Whelan, N. V., Kocot, K. M., Moroz, T. P., Mukherjee, K., Williams, P., Paulay, G., Moroz, L. L. &
928 Halanych, K. M. (2017). Ctenophore relationships and their placement as the sister group to all other
929 animals. *Nature Ecology & Evolution*, 1(11), 1737–1746. <https://doi.org/10.1038/s41559-017-0331-3>
- 930 Wishner, K. F., Ashjian, C. J., Gelfman, C., Gowing, M. M., Kann, L., Levin, L. A., Mullineaux, L. S. &
931 Saltzman, J. (1995). Pelagic and benthic ecology of the lower interface of the Eastern Tropical Pacific
932 oxygen minimum zone. *Deep Sea Research Part I: Oceanographic Research Papers*, 42(1), 93–115.
933 [https://doi.org/10.1016/0967-0637\(94\)00021-j](https://doi.org/10.1016/0967-0637(94)00021-j)
- 934 Woo, S., Denis, V., Won, H., Shin, K., Lee, G., Lee, T. K., & Yum, S. (2013). Expressions of oxidative stress-
935 related genes and antioxidant enzyme activities in *Mytilus galloprovincialis* (Bivalvia, Mollusca) exposed
936 to hypoxia. *Zoological Studies*, 52(1), 1–8. <https://doi.org/10.1186/1810-522x-52-15>
- 937 Woodhead, A. J., Hicks, C. C., Norström, A. V., Williams, G. J., & Graham, N. A. (2019). Coral reef
938 ecosystem services in the Anthropocene. *Functional Ecology*, 33(6), 1023–1034.
939 <https://doi.org/10.1111/1365-2435.13331>

940 Wulff, J. L. (2006). Ecological interactions of marine sponges. *Canadian Journal of Zoology*, 84(2), 146–
941 166. <https://doi.org/10.1139/z06-019>

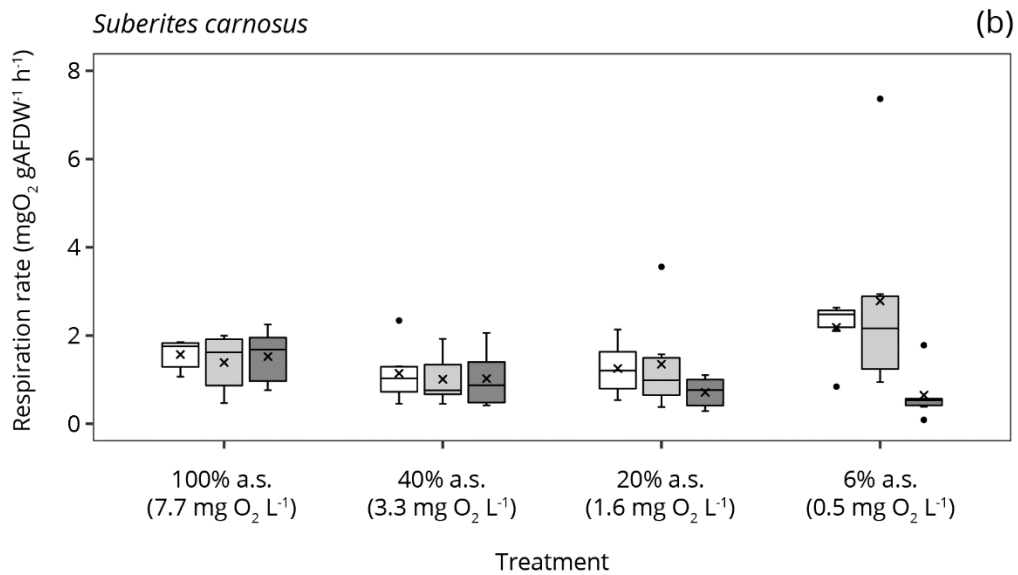
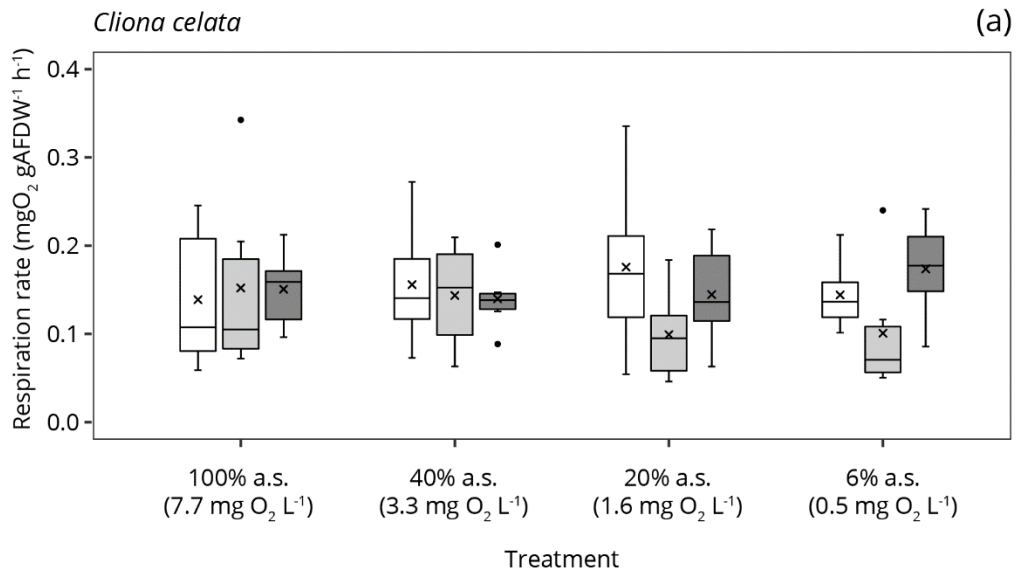
942 Yahel, G., Marie, D., & Genin, A. (2005). InEx—a direct in situ method to measure filtration rates,
943 nutrition, and metabolism of active suspension feeders. *Limnology and Oceanography: Methods*, 3(2),
944 46–58. <https://doi.org/10.4319/lom.2005.3.46>


945

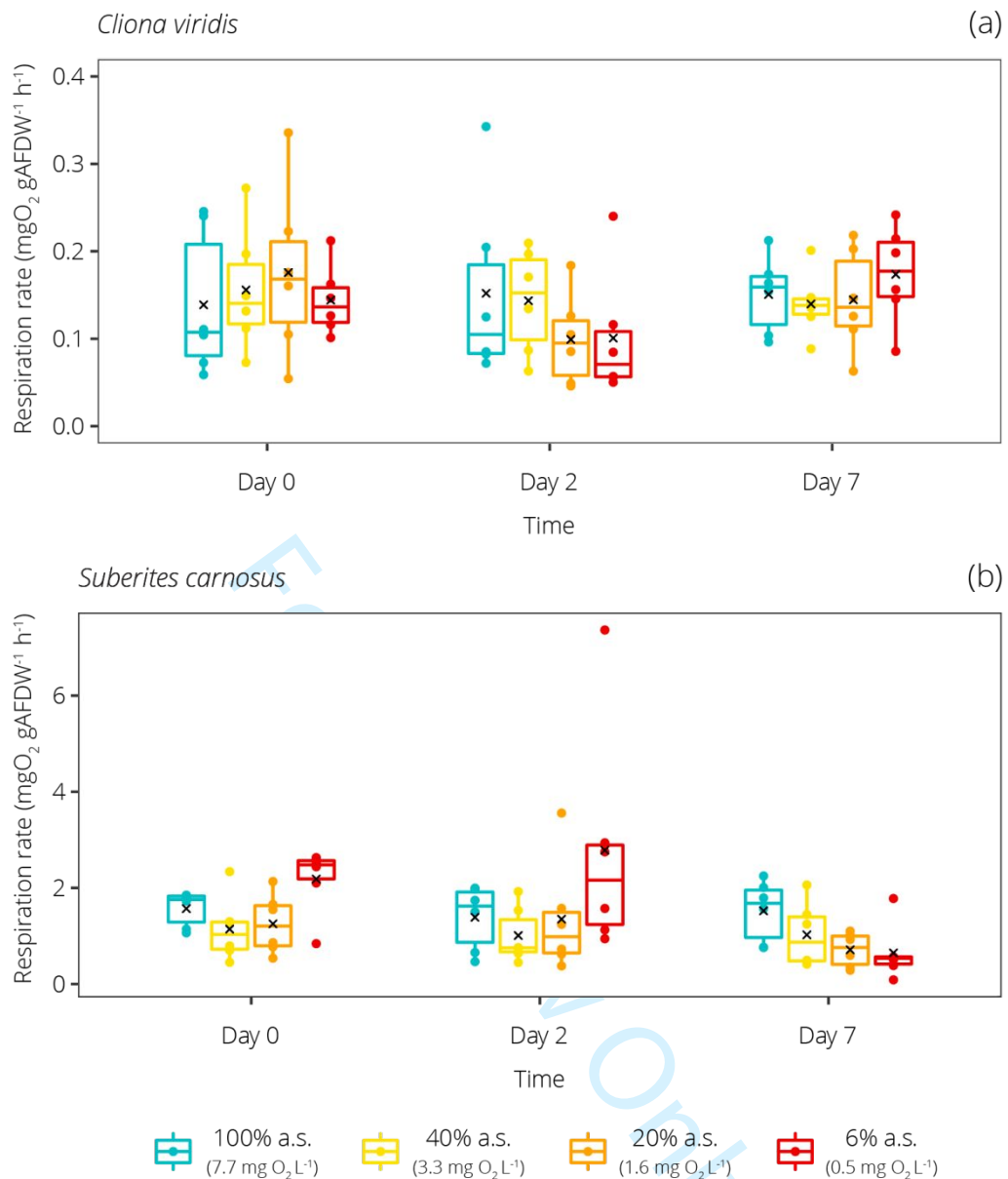
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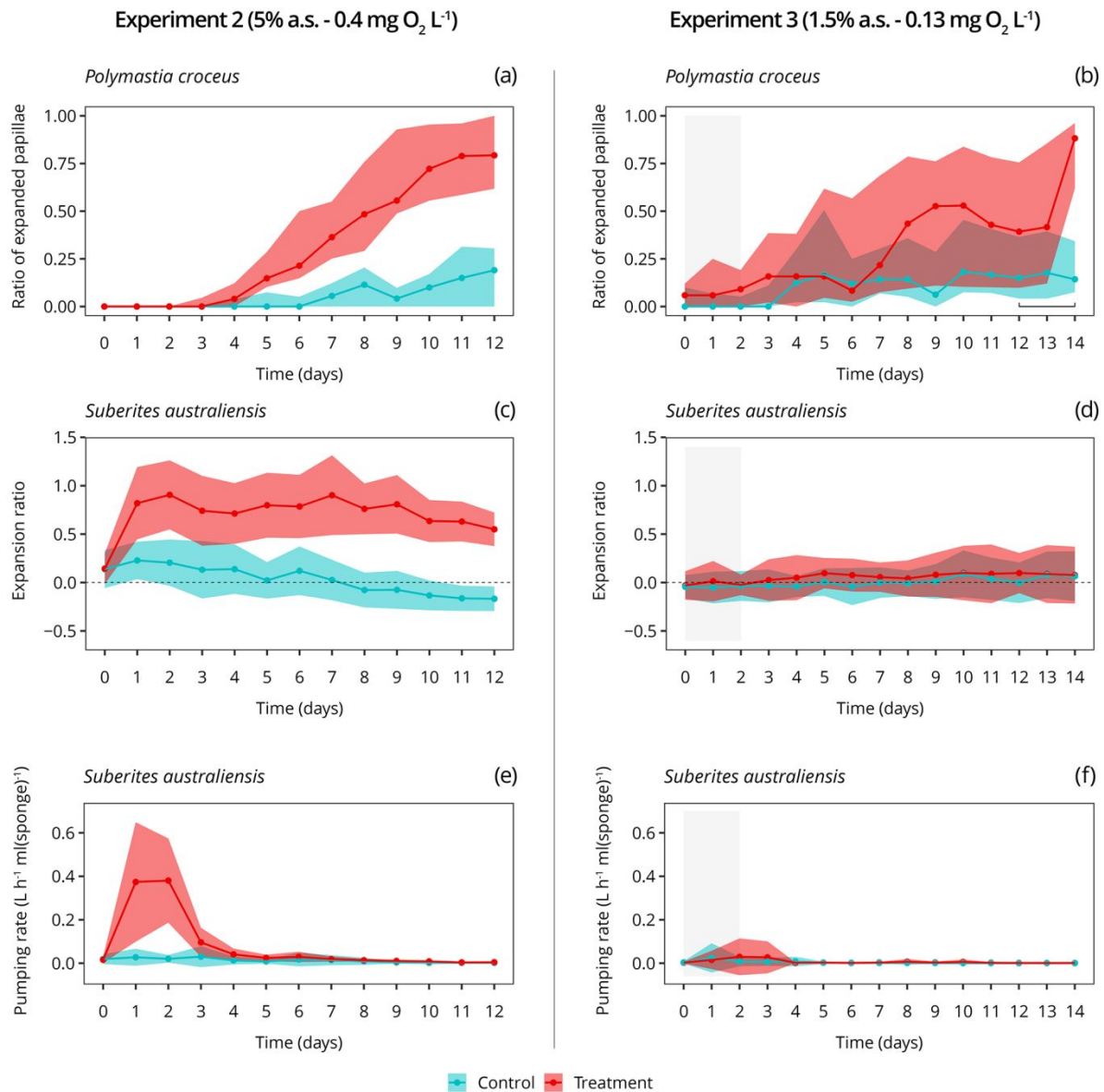
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Figure 1. Respiration rates in (a) *Cliona- celata* and (b) *Suberites- carnosus* from experiment 1 (moderate hypoxia) measured at T_0 , $T_{1/2}$ and $T\text{-end}$. Note: x-axis and y-axis scales differ between species. The oxygen concentration of the different treatments is expressed as % air saturation (% a.s.). Horizontal bars inside the boxplots represent medians; the symbol x represents means. Lower and upper hinges of the boxplots correspond to the first and third quartiles, respectively. Lower and upper whiskers represent the smallest and largest values, respectively. Single dots represent outliers data points.



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959 Figure 2. Contractile behaviour and pumping rate during experiments 2 and 3 (severe hypoxic conditions).

960 Changes in the ratio of expanded papillae over time in *Polymastia P.-croceus* in each treatment in961 experiments 2 (a) and 3 (b). Changes in the expansion ratio over time in *Suberites S.-australiensis* in each

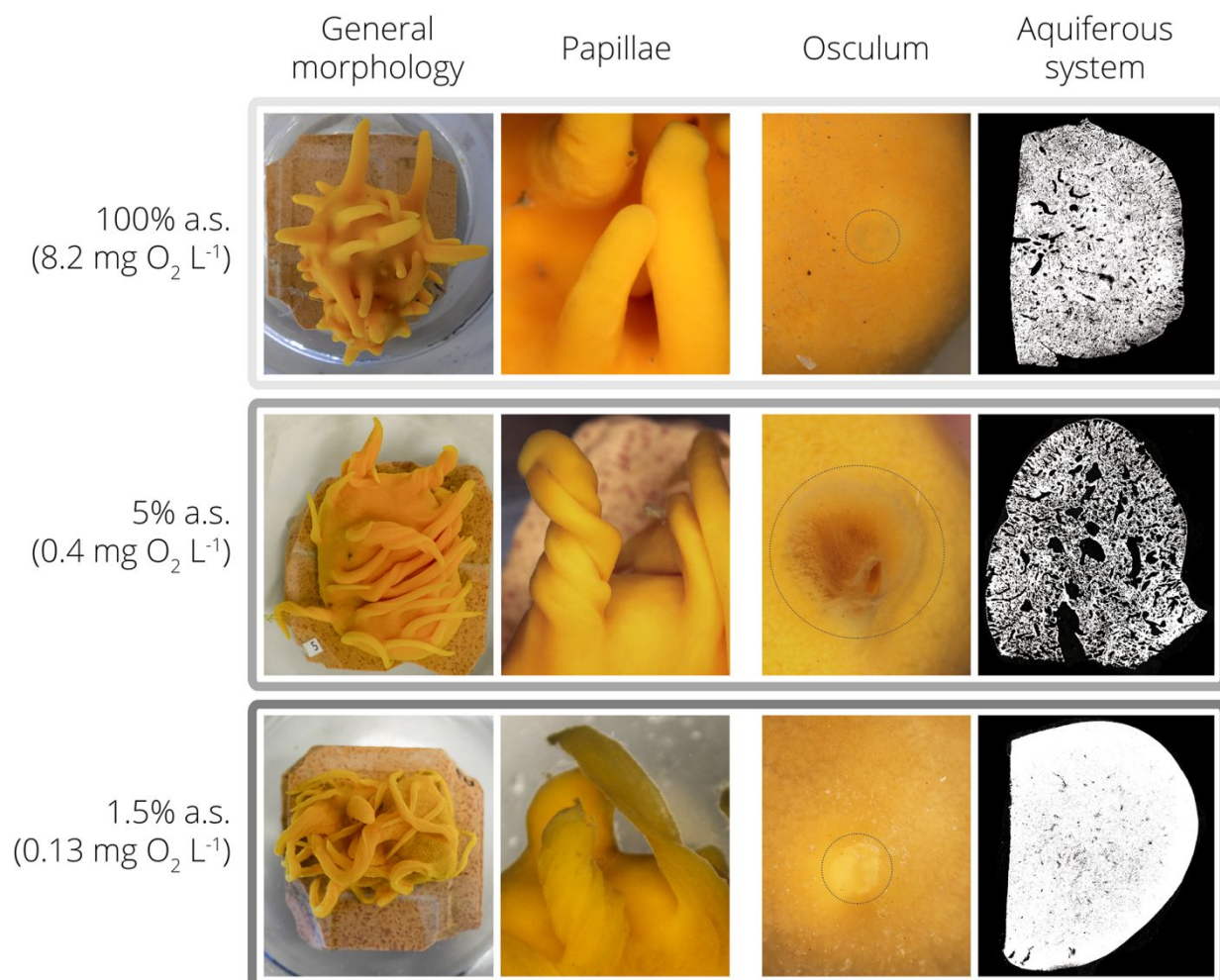
962 treatment in experiments 2 (c) and 3 (d). Changes in the pumping rate over time (estimated from the

963 osculum cross-sectional area) in *S. australiensis* in each treatment in experiments 2 (e) and 3 (f). In (a) and964 (b), points represent the median, while lower and upper edges of the ribbons represent the 75th and 25th

965 percentile, respectively. In (c), (d), (e) and (f), points represent the means while lower and upper edges of

966 the ribbon represent the standard deviation. Days of hypoxic acclimation (10% a.s.) are highlighted in grey.

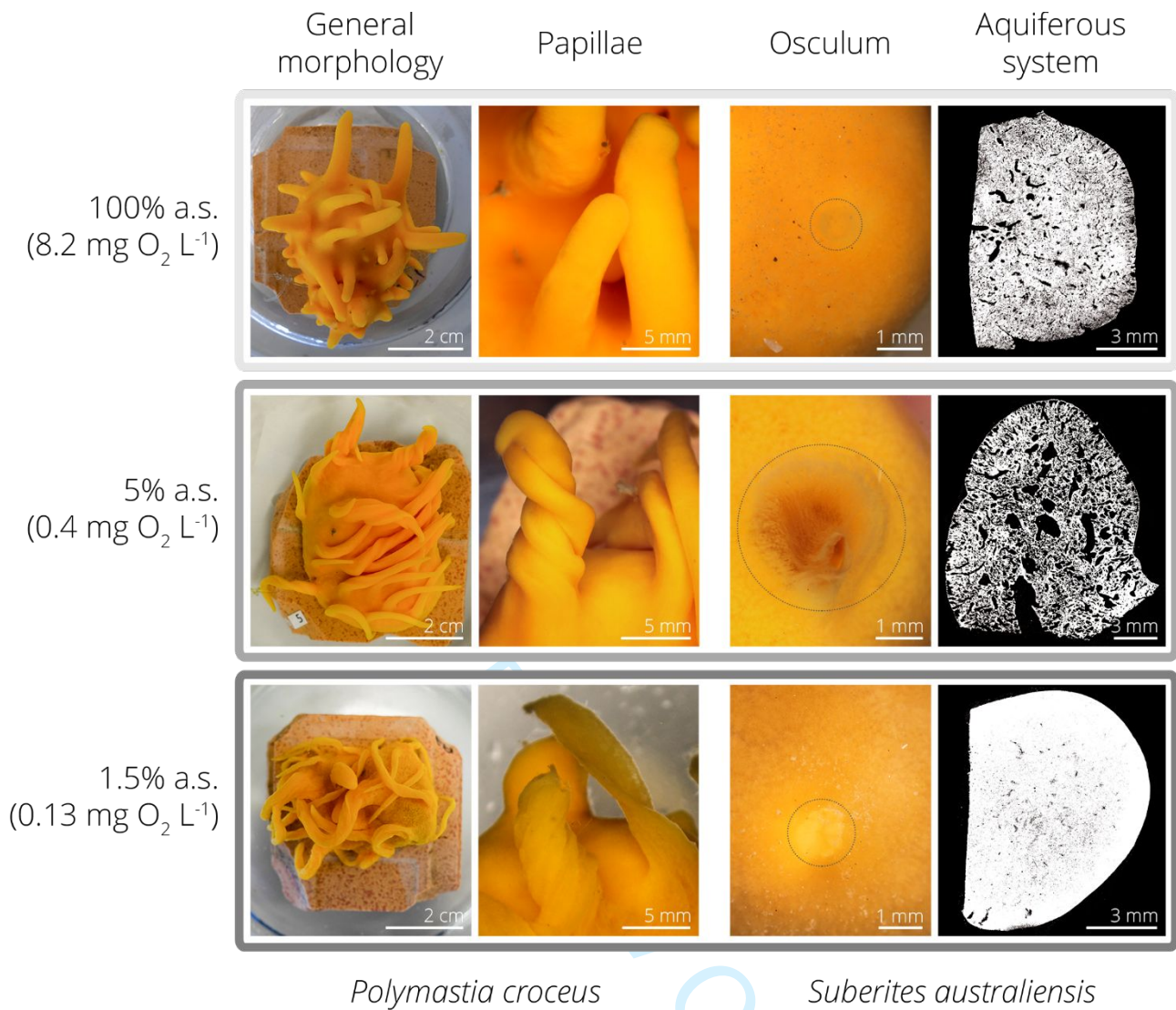
967 In (b), a black line is used to highlight days when sponges experienced mortality.



Polymastia croceus

Suberites australiensis

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Figure 3. Examples of the morphological modifications reported in sponges exposed to low dissolved oxygen in the severe hypoxia treatments compared to the controls. From left to right: general external morphology, and details of papillae in *Polymastia P-croceus*; details of the osculum (evidenced with a dotted line), and transverse histological section (sponge tissue is in white and empty spaces representing the aquiferous system are in black) in *Suberites S-australiensis*. An extended version of this figure is found in the supplemental material (Fig. S12).

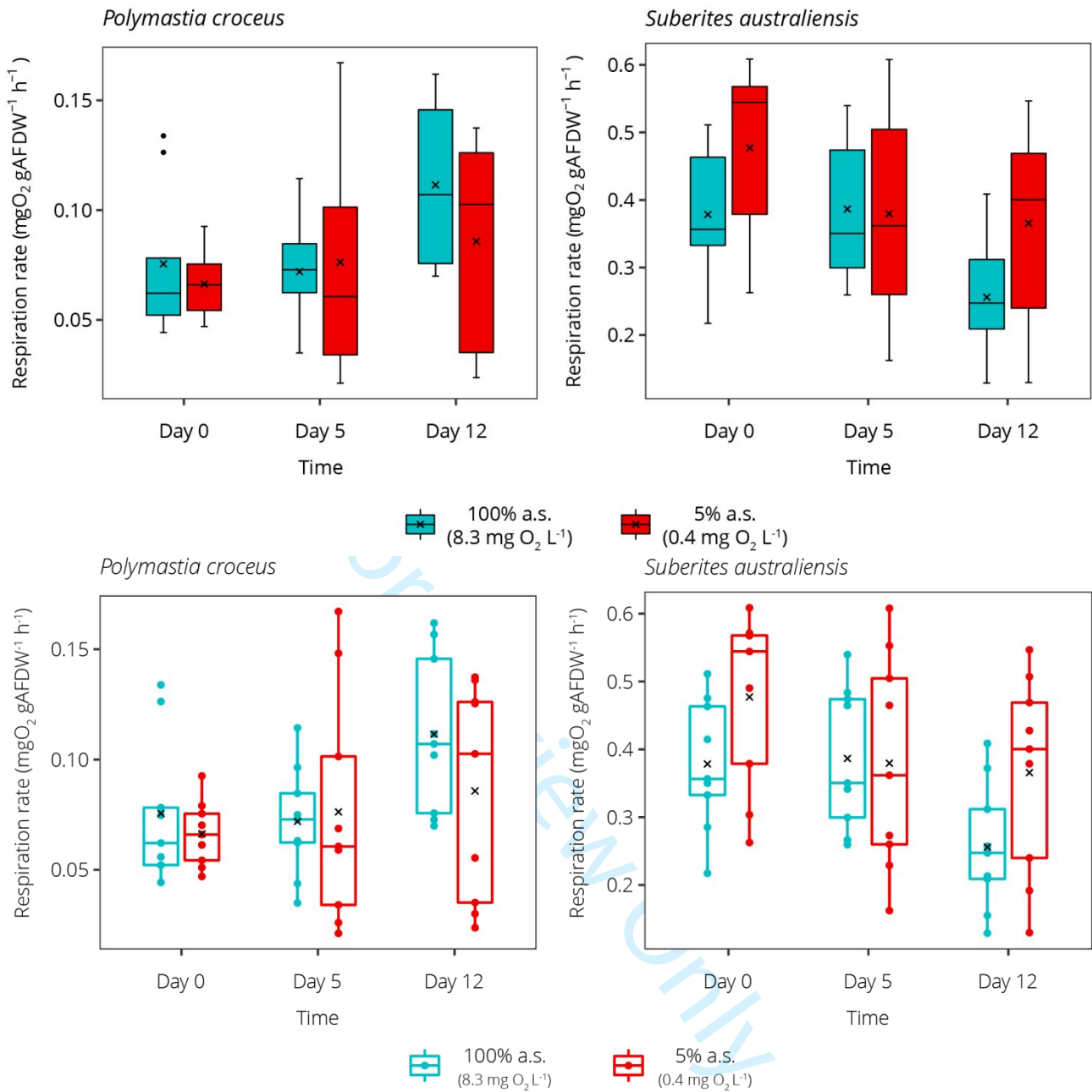
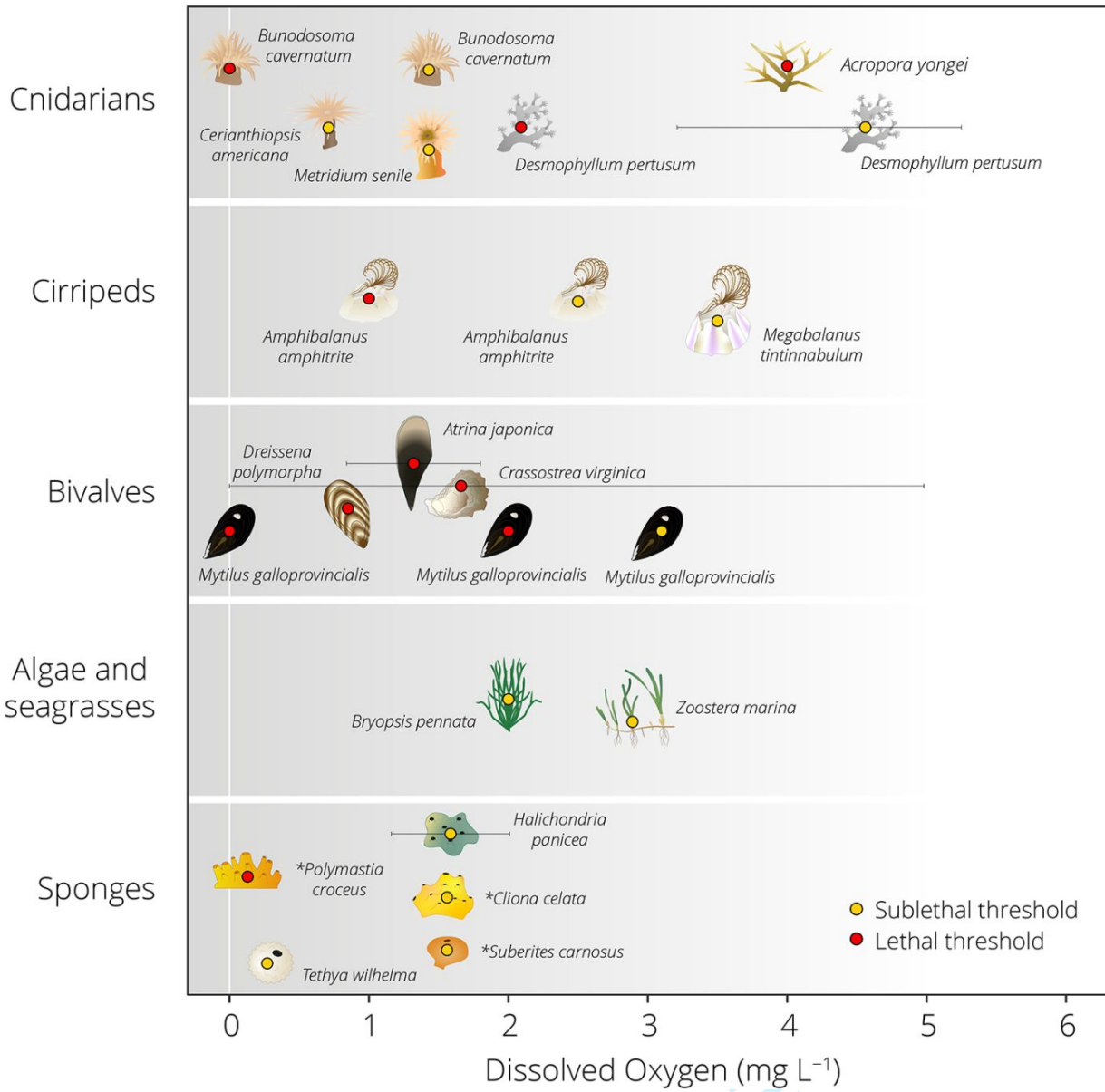
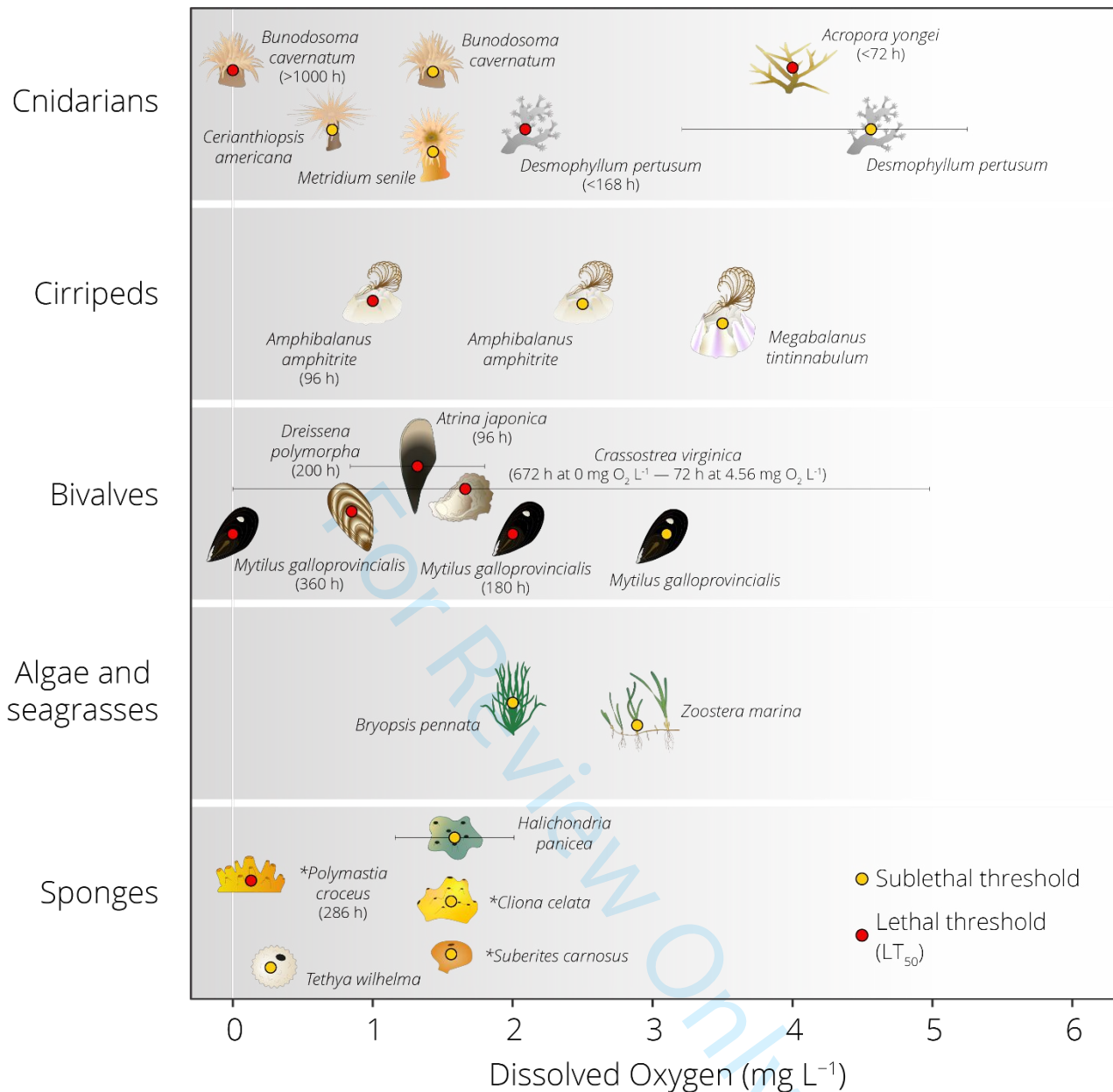


Figure 4. Respiration rates in *Polymastia P. croceus* and *Suberites S. australiensis* from experiment 2 (5% a.s.) measured at T_0 , $T_{1/2}$ and $T\text{-end}$. Note: x-axis and y-axis scales differ between species. Horizontal bars inside the boxplots represent medians; the symbol \times represents means. Lower and upper hinges of the boxplots correspond to the first and third quartiles, respectively. Lower and upper whiskers represent the smallest and largest values, respectively. Points represent data points. Single dots represent outliers.





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987 Figure 5. The tolerance of marine sessile organisms to hypoxia. Red dots indicate lethal thresholds, while
 988 yellow dots indicate sub-lethal thresholds. Organisms whose values were found in the present study are
 989 labelled with an asterisk. For studies that report multiple values for the same species according to other
 990 abiotic conditions (i.e., temperature and salinity), we report a range where the dots represent the mean
 991 value, and the edges of the whiskers represent minimum and maximum values. For lethal thresholds, we
 992 report in bracket the median Lethal time (LT₅₀, hours) at that specific oxygen concentration (or at the
 993 extremes of the range). The symbol < was used when LT₅₀ was not reported, but more than 50% of the
 994 organisms died after a certain amount of time; while > was used when LT₅₀ was not reached by the end of

995 [the experiment](#). For *H. panicea*, Mills et al. (2014) only report oxygen measurement as per cent air
996 saturation without reporting temperature and salinity, so the actual oxygen concentration is unknown. We,
997 therefore, estimated the oxygen content using the range of temperatures and salinity found where these
998 sponges were sampled (Salinity 8.9–29.5; Temperature: 5–25 °C; Thomassen & Riisgård, 1995) and we
999 provide the mean and the range of possible values. List of references associated with each species: *A.*
1000 *yongei* (Haas et al., 2014), *A. amphitrite* (Rao & Ganapati, 1968; Desai & Prakash, 2009), *A. japonica*
1001 (Nagasoe et al., 2020), *B. pennata* (Haas et al., 2014), *B. cavernatum* (Mangum, 1980; Ellington, 1982), *C.*
1002 *americana* (Vaquer-Sunyer & Duarte, 2008), *C. virginica* (Stickle et al., 1989), *D. pertusum* (Dodds et al.,
1003 2007; Lunden et al., 2014), *D. polymorpha* (Johnson & McMahon, 1998), *H. panicea* (Mills et al., 2014), *M.*
1004 *tinnabulum* (Rao and Ganapati 1968), *M. senile* (Sassaman & Mangum 1972), *M. galloprovincialis* (De
1005 Zwaan et al., 1991; Woo et al., 2013), *T. wilhelma* (Mills et al., 2018), *Z. marina* (Hughes et al., 2020). [Figure](#)
1006 [inspired by Hughes et al. \(2020\)](#).

1 **Adaptive strategies of sponges to deoxygenated oceans**

2 **Abstract**

3 Ocean deoxygenation is one of the major consequences of climate change. In coastal waters, this process
4 can be exacerbated by eutrophication, which is contributing to an alarming increase in the so-called “dead
5 zones” globally. Despite its severity, the effect of reduced dissolved oxygen has only been studied for a very
6 limited number of organisms, compared to other climate change impacts such as ocean acidification and
7 warming. Here we experimentally assessed the response of sponges to moderate and severe simulated
8 hypoxic events. We ran three laboratory experiments on four species from two different temperate oceans
9 (NE Atlantic and SW Pacific). Sponges were exposed to a total of five hypoxic treatments, with increasing
10 severity (3.3, 1.6, 0.5, 0.4 and 0.13 mg O₂ L⁻¹, over 7–12-days). We found that sponges are generally very
11 tolerant of hypoxia. All the sponges survived in the experimental conditions, except *Polymastia croceus*,
12 which showed significant mortality at the lowest oxygen concentration (0.13 mg O₂ L⁻¹, lethal median time:
13 286 h). In all species except *Suberites carnosus*, hypoxic conditions do not significantly affect respiration
14 rate down to 0.4 mg O₂ L⁻¹, showing that sponges can uptake oxygen at very low concentrations in the
15 surrounding environment. Importantly, sponges displayed species-specific phenotypic modifications in
16 response to the hypoxic treatments, including physiological, morphological, and behavioural changes. This
17 phenotypic plasticity likely represents an adaptive strategy to live in reduced or low oxygen water. Our
18 results also show that a single sponge species (i.e., *Suberites australiensis*) can display different strategies at
19 different oxygen concentrations. Compared to other sessile organisms, sponges generally showed higher
20 tolerance to hypoxia, suggesting that sponges could be favoured and survive in future deoxygenated
21 oceans.

22 **KEYWORDS**

23 climate change, Porifera, evolution, marine benthic hypoxia, hypoxic events, oxygen depletion,
24 eutrophication, sessile organisms, dead zones, phenotypic plasticity

25 **1 INTRODUCTION**

26 Anthropogenic emissions of carbon dioxide and other greenhouse gasses have increased exponentially
27 since the industrial revolution, causing significant changes in the Earth's climate (Raupach & Canadell, 2010;
28 IPCC, 2021). Climate change has three main effects on the marine environment: warming, acidification, and
29 oxygen decline (Bijma et al., 2013). While most ecological and physiological research has targeted the first
30 two stressors, deoxygenation remains comparatively neglected (Limburg et al., 2017). Despite the scant
31 attention, recent research shows that oxygen loss is a major anthropogenic stressor for marine biota that
32 may exceed the severity of the combined effects of ocean warming and acidification (Sampaio et al., 2021).

33 Oxygen is essential to all aerobic life, and ocean deoxygenation has the potential to affect all
34 biogeochemical and biological processes within the oceans (Semenza, 2007; Levin & Breitburg, 2015). In the
35 open sea, warming is considered the main cause of O₂ reduction: an increase in sea temperature leads to
36 decreased O₂ solubility, increased water stratification, and alterations to oceanic circulation, which reduces
37 O₂ supply to the ocean interior (Doney, 2010; Keeling et al., 2010). Higher temperatures also enhance
38 microbial respiration, which can further deplete oxygen in marine ecosystems (Altieri & Diaz, 2019;
39 Robinson, 2019). Oxygen levels in the global oceans have already declined by 2% during the last 50 years,
40 with more significant O₂ declines in the North Pacific and tropical oxygen minimum zones (OMZ) (Levin &
41 Breitburg, 2015). This is likely to get worse in the future, with models predicting a global ocean reduction in
42 O₂ of up to 7% by the end of the century (Keeling et al., 2010).

43 In coastal waters, climate-driven deoxygenation can be intensified by eutrophication (Nixon, 1995; Altieri &
44 Gedan, 2015). The input of anthropogenic nutrients, such as fertilizers and human/livestock wastes, can
45 increase algal growth resulting in an accumulation of organic material on the seafloor. This excess of
46 organic matter is then degraded by bacteria, causing O₂ depletion that can lead to hypoxic conditions
47 (Smith et al., 2006). In shallow and well-mixed waters, eutrophication-driven hypoxia is generally caused by
48 nocturnal heterotrophic respiration, resulting in daily oscillations in oxygen concentration. In contrast, long-
49 term hypoxic events are more likely to occur in enclosed seas or basins (Levin et al., 2009). Hypoxia has
50 widespread and severe impacts across taxonomic and functional groups. The intensity and duration of

51 oxygen depletion are the main factors influencing the severity of hypoxic events on benthic organisms
52 (Levin et al., 2009; Altieri & Diaz, 2019). Mild hypoxia can alter behavioural patterns, decrease feeding rates
53 and cause changes in physiological processes (Vaquer-Sunyer & Duarte, 2008). Severe hypoxic events can
54 cause mass mortalities, leading to the formation of the so-called “dead zones”, areas largely devoid of
55 macrofauna (Diaz & Rosenberg, 2008). Dead zones have been reported in small water bodies such as
56 harbours, fjord and inlets, and large basins, such as the Baltic Sea, spreading over 60,000 km² (Altieri & Diaz,
57 2019). As climate and land use continue to change, coastal hypoxia is expected to worsen, with the
58 increased occurrence, frequency, intensity, and duration of hypoxic events (Diaz & Rosenberg, 2011).

59 Despite the extent of the problem and the dramatic effects caused by ocean deoxygenation, the response
60 of many groups of organisms to hypoxia is still poorly studied. This lack of knowledge limits our ability to
61 model the effects of declining oxygen availability on marine ecosystems (Seibel, 2011). To date, research on
62 tolerance to reduced levels of dissolved oxygen has primarily focused on fish, crustaceans and molluscs
63 (Vaquer-Sunyer & Duarte, 2008), while very little is known about other groups, especially sessile organisms.
64 Sessile organisms are particularly vulnerable to hypoxic events because they cannot move or migrate to
65 well-oxygenated water. Furthermore, sessile organisms include many important habitat-forming species, so
66 any change in their abundance could have major consequences for the ecosystems they support (Vergés et
67 al., 2019; Woodhead et al., 2019; Piazzini et al., 2021). Therefore, it is critical to understand how these
68 organisms respond to hypoxia to predict possible future changes and effectively manage marine
69 ecosystems.

70 Sponges are the dominant sessile organisms in many marine ecosystems and are found in high abundance
71 in tropical, temperate, and polar ecosystems (Ayling, 1983, Bell et al., 2020). They perform many important
72 ecological functions, including contributing to nutrient cycling, bioerosion, enhancing ecosystem complexity
73 and providing habitats for a wide range of associated organisms (Wulff, 2006; Bell, 2008; Maldonado et al.,
74 2012). Despite being important components of marine ecosystems, sponge tolerance to hypoxia has been
75 poorly investigated to date. Mills et al. (2014) showed that *Halichondria panicea* can feed and respire with
76 oxygen levels down to 4% of air saturation. However, the authors did not provide information on the

77 duration of the treatments and replication; furthermore, in the same study, information on the
78 temperature and salinity of the water was unavailable, so it is not possible to derive the actual oxygen
79 concentrations to which sponges were exposed. Two other relevant experiments have investigated the
80 short-term response of sponges to hypoxia. Mills et al. (2018) exposed *Tethya wilhelma* to a step
81 decreasing oxygen concentration (30–40 h with O_2 lower than 10% a.s., 0.7 mg L^{-1}). They found that
82 sponges continued to perform periodic full-body contractions down to $0.27 \text{ mg O}_2 \text{ L}^{-1}$, but ceased below
83 that concentration. Leys & Kahn (2018) exposed *Geodia barretti* to 6.5 h of hypoxia (7% air saturation, 0.6
84 $\text{mg O}_2 \text{ L}^{-1}$), and found that sponge respiration rate remained largely unchanged. However, filtration rates
85 dropped almost immediately after the oxygen level was reduced. Despite these earlier studies, we still have
86 very little insight into how sponges may cope with hypoxic events caused by ocean and coastal
87 deoxygenation.

88 Here we provide the first comprehensive assessment of sponge response to hypoxia. Specifically, we
89 experimentally investigated the physiological, behavioural, and morphological responses of four temperate
90 sponge species to moderate and severe hypoxic conditions. We ran the first experiment to expose sponges
91 to moderate hypoxic conditions for seven days, including a wide range of dissolved oxygen concentrations
92 (0.5 , 1.6 and $3.3 \text{ mg O}_2 \text{ L}^{-1}$). Subsequently, we investigated sponge response to severe hypoxia (0.13 and 0.4
93 $\text{mg O}_2 \text{ L}^{-1}$) for 12 days with two additional experiments. Finally, we discuss sponge tolerance to low
94 dissolved oxygen compared to other sessile organisms in the context of future climatic conditions.

95 2 MATERIALS AND METHODS

96 2.1 Study area and species

97 Experiment 1 (moderate hypoxia) was performed in Ireland (Renouf Laboratory, Lough Hyne) on two
98 abundant North-East Atlantic sponge species: *Cliona celata* Grant, 1826 and *Suberites carnosus* (Johnston,
99 1842). Experiments 2 and 3 (severe hypoxia) were performed in New Zealand (Wellington University
100 Coastal Ecology Laboratory, Wellington) on two abundant temperate Australasian species: *Polymastia*
101 *croceus* Kelly-Borges and Bergquist, 1997 and *Suberites australiensis* Bergquist, 1968.

102 2.1 Experiment 1: moderate hypoxia

103 In the first experiment, we investigated the response of *Cliona celata* and *Suberites carnosus* to a wide
104 range of oxygen concentrations, using an air-tight system with a continuous flow of seawater. Sponges
105 were exposed to ~95% ($7.71 \pm 0.19 \text{ mg O}_2 \text{ L}^{-1}$), ~40% ($3.34 \pm 0.17 \text{ mg O}_2 \text{ L}^{-1}$), ~20% ($1.56 \pm 0.19 \text{ mg O}_2 \text{ L}^{-1}$)
106 and ~6% ($0.48 \pm 0.09 \text{ mg O}_2 \text{ L}^{-1}$) air saturation (a.s.) for seven days (a summary of the seawater parameters
107 is provided in Table S1).

108 The experimental set-up (see scheme in Figure S1) consisted of two independent replicate modules for
109 each treatment, randomly distributed in the experimental set-up. To condition water, we used two header
110 tanks for each experimental module: one providing water and one reservoir. Header tanks were filled with
111 10- μm -filtered seawater. The oxygen level was then lowered and maintained to the desired dissolved
112 oxygen concentration by bubbling specific mixtures of N_2 (BOC, food-grade) and air, through glass-ceramic
113 diffusers. Hypoxic gas blends were prepared by decanting food-grade N_2 and air in 15 L scuba cylinders
114 using an oxygen decanting assembly (Undersea Ltd, 5215) with a DPM-300 digital gauge (0.25% accuracy).
115 Oxygen concentration was then checked with a Nuair Pro O_2 Analyser and adjusted, if necessary.

116 Conditioned water was delivered to two replicate experimental chambers (2.3 L) for each system at a rate
117 of 25 L per day, ensuring 100% water replacement every 2 h and 15 min. Water circulation within each
118 experimental chamber was provided by the gravity-driven water flow (~3 cm/s). Temperature was kept
119 constant using a water bath controlled by an aquarium chiller.

120 *Cliona celata* was collected from the Kedges (51°27'41.4 "N 9°20'44.2 "W), whereas *Suberites carnosus* was
121 collected from the rocky cliffs of Lough Hyne (51°30'00.4"N 9°18'03.9"W). For both species, sampling was
122 carried out at 10–18 m in June 2019 and sponges collected were at least 2 m apart. Sponges were then left
123 to recover for two months from harvesting stress in a 1 m³ underwater cage placed at 8 m of depth.
124 Sponges were then transferred to the experimental system and randomly distributed across the
125 experimental chambers. The experimental design consisted of 32 experimental chambers (two replicate
126 chambers for each species for each replicate module, and two replicate modules for each treatment). A
127 diagram of the experimental design is reported in Figure S2. Three sponges were placed in each chamber (6
128 sponges in total for each replicate module and 12 for each treatment). Sponges belonging to different
129 species were not mixed but kept in separate chambers. Sponges were left to acclimate with oxygen
130 saturated air for five days before oxygen was lowered by introducing hypoxic water into the chambers.
131 Oxygen concentration was lowered in 24h and was then maintained for seven days until the end of the
132 experiment (a graph showing the oxygen concentration in the different treatments over time is provided in
133 Figure S3). In natural ecosystems, hypoxic conditions can develop in short times ranging from hours to a
134 few days (Breitburg, 1990; Nezlin et al., 2009), so we consider these acclimation times appropriate and
135 ecologically relevant. Temperature and oxygen concentration inside the experimental chamber were
136 measured twice a day using a Fibox 4 oxygen meter with a dipping probe (Presence GmbH, Germany). A
137 two-spot calibration was performed on the oxygen probe every three days, using sodium dithionite for 0 %
138 oxygen and air-saturated water for 100% oxygen.

139 **2.2 Experiments 2 and 3: severe hypoxia**

140 We also investigated the response of sponges (*Polymastia croceus* and *Suberites australiensis*) to severe
141 hypoxia through two separate experiments using an air-tight system. In experiment 2, sponges were
142 exposed to ~5% (0.4 ± 0.04 mg O₂ L⁻¹), and ~100% a.s. (8.34 ± 0.13 mg O₂ L⁻¹), while in experiment 3,
143 sponges were exposed ~1.5% (0.13 ± 0.02 mg O₂ L⁻¹), and ~100% a.s. (8.15 ± 0.16 mg O₂ L⁻¹). The different
144 oxygen concentration in the controls was due to the small difference in temperature between the two

145 experiments (13.3 ± 0.5 °C in experiment 2 compared to 14.3 ± 0.6 °C in experiment 3). A summary of the
146 seawater parameters is provided in Table S1.

147 Sponges were kept in independent cylindrical air-sealed polypropylene chambers (10 L), randomly
148 distributed inside a water bath. Every two days, ~70% of the water was replaced using 10- μ m-filtered
149 seawater, preconditioned to the desired oxygen concentration in independent conditioning tanks. Oxygen
150 concentration was then maintained by bubbling air or air-N₂ blends through glass-ceramic diffusers (see
151 section 2.1.2 for more details on gas blends). Custom made de-bubbler devices were used to eliminate
152 bubbles coming from the ceramic diffusers that could affect sponges (Figure S4). The sponges were fed
153 twice a day with *Nannochloropsis* microalgae (1–2 μ m cell diameter; Nanno 3600™ Reed Mariculture, US.).
154 Water circulation within each experimental chamber was provided by the de-bubbler device and an
155 additional water pump located on the side of the chamber, which provided a constant circular water flow.
156 The chambers were placed in a water bath to control water temperature.

157 *Polymastia croceus* was collected from Barrett Reef (Wellington South Coast, 41°20'31.1"S 174°50'09.7"E)
158 by cutting fragments (~8 cm³) from separate sponges (at least 5 m apart). Whole specimens of *Suberites*
159 *australiensis* were collected from Mahanga Bay (Wellington Harbour, 41°17'32.2"S 174°50'06.5"E),
160 attached to a fragment of their respective substrate. Sponges were then left to recover for three weeks
161 after sampling and cutting stress in water tables with 10- μ m-filtered flow-through seawater.

162 Sponges were then transferred to the experimental system, consisting of 12 experimental chambers (3
163 independent replicate chambers for each species and treatment combination). Five sponges were placed in
164 each chamber (15 sponges in total for each treatment). Sponges were left to acclimate with oxygen
165 saturated air for five days. Oxygen was then lowered by bubbling a specific Air-N₂ mixture. In experiment 2,
166 oxygen was lowered in ~24 hours and then maintained at 5% a.s. (0.4 mg O₂ L⁻¹) for 12 days until the end of
167 the experiment. While in experiment 3, oxygen was lowered to 1.5% a.s. (0.13 mg O₂ L⁻¹) in ~72 hours
168 (which included a preacclimation at 10% a.s., 1 mg O₂ L⁻¹) and maintained for 12 days until the end of the
169 experiment (Fig. S3). This further acclimation in hypoxic conditions was made because of the very low O₂
170 concentration of the treatment. In experiment 3, due to the very low concentration of O₂ (0.13 ± 0.02 mg L⁻¹)

171 ¹), O₂ increased to ~0.3 mg L⁻¹ for about 20 minutes during daily examinations. Temperature and oxygen
172 concentration inside the experimental chamber were measured twice a day using Fibox 4 oxygen meter
173 with a dipping probe (Presence GmbH, Germany). A two-spot calibration was performed on the oxygen
174 probe every three days, using sodium dithionite for 0% oxygen and air-saturated water for 100% oxygen.

175 **2.3 Response variables**

176 **2.3.1 Survival and health monitoring**

177 Sponge health was monitored daily during the experiment. Sponges showing ≥ 25% of external necrosis
178 were considered dead and removed from their treatment tanks during the daily checks, so as not to impact
179 other sponges in the treatments. At the end of the experiments, all sponges were sectioned to assess the
180 presence of any internal necrosis.

181 **2.3.2 Respiration Rate**

182 For all the experiments, respiration rate was measured on the same specimens at *T*₀ (before the beginning
183 of the experiment), *T*_{1/2} (after two days from the beginning of the final treatment in experiment 1, and
184 after five days in experiments 2 and 3) and *T*-end (end of the experiment). In experiment 1 (moderate
185 hypoxia), we measured respiration rates of three sponges in each replicate module (*n* = 6 for each
186 treatment). In experiments 2 and 3 (severe hypoxia), respiration rates were measured on three sponges in
187 each experimental chamber (*n* = 9 for each treatment). To measure respiration rate, sponges were placed
188 in sealed cylindrical glass respiration chambers (150 ml for *Cliona celata*; 80 ml for *Suberites carnosus*; 250
189 ml for *Polymastia croceus* and *Suberites australiensis*) with PreSens oxygen sensor spots (SP-PSt3-NAU)
190 attached to their inner surface. Experimental chambers contained either oxygen saturated water (pre-
191 experimental measurements and controls) or water at a slightly higher oxygen concentration than the
192 experimental treatment (+20–50%, depending on the treatment) collected from the respective header
193 tanks. Respiration rates were not performed on sponges from experiment 3 (1.5% a.s.). The incubations
194 were performed in controlled temperature (water bath) and dark conditions. The water inside the
195 respirometry chambers was gently stirred using a magnetic stir bar. After 20 min of acclimation, oxygen
196 concentration inside the chambers was measured every 10 min for 1 hour, using a Fibox 4 oxygen meter

197 with a polymer optical fibre (POF). Respiration measurements were ended prematurely if the oxygen level
198 fell below 70% of the treatment concentration to avoid any detrimental effect on the sponges. Blank
199 incubations, containing only seawater were performed every respiration run and used to correct for any
200 microbial community respiration in the seawater. A two-point calibration was performed on the oxygen
201 sensor spots before each measurement session.

202 Respiration measurements were standardized to sponge ash-free dry weight (*AFDW*) from buoyant weight
203 (*BW*) measurements (Fig. S5). For *Suberites australiensis*, it was not possible to estimate *AFDW* from *BW*
204 due to the abundant external material accumulated by the sponge inside the tissue that influences the *BW*.
205 For this species, we measured the *AFDW* of all the specimens used in the respirations at the *T-end*, and we
206 assumed that sponges had the same weight at *T0* and *T1/2*.

207 **2.3.3 Changes in weight, size, and morphology**

208 Changes in weight and size over time relative to the initial values were estimated by calculating the
209 buoyant weight variation (*BWV*) and contracted area variation (*CAV*). For all of the experiments, buoyant
210 weights (*BW*) of all experimental sponges (except *Suberites australiensis*) were taken at *T0* and *T-end* and
211 used to calculate relative buoyant weight variation as $BWV = [(BW_{T-end} - BW_{T0}) / BW_{T0}] \cdot 100$. Buoyant weight
212 was measured with a digital scale (A&D FX-200i) following the methods of Osinga et al. (1999). For
213 experiments 2 and 3 (severe hypoxia), photographs of contracted sponges were taken at *T0* and *T-end* to
214 measure sponge contracted area (*CA*) and calculate contracted area variation as $CAV = [(CA_{T-end} - CA_{T0}) /$
215 $CA_{T0}] \cdot 100$ (following Osinga et al., 1999). Contraction was achieved by disturbing sponges with a blunt
216 plastic rod (being careful not to damage the sponge) and waiting for one hour for the sponge to react to the
217 stimulus. All the photographs were analysed using ImageJ (US National Institutes of Health, Bethesda, Md,
218 USA).

219 During experiments 2 and 3, treatment conditions induced the development of peculiar morphological
220 structures in some specimens of both *Polymastia croceus* and *Suberites australiensis*. Sponges were

221 photographed and monitored daily to calculate the percentage of specimens developing these structures
222 and the median time of occurrence.

223 **2.3.4 Sponge contractile behaviour**

224 During experiments 2 (5% a.s.) and 3 (1.5% a.s.), sponge contractile behaviour was monitored daily from T_0
225 to $T\text{-end}$, on all experimental sponges through photographic analysis. For *Suberites australiensis*, the
226 contractile behaviour was estimated using an “expansion ratio” ($EXPR$) calculated as $EXPR = A_{T_i} / CA_{T_0}$,
227 where A_{T_i} is the area occupied at T_i and CA_{T_0} is the contracted area at T_0 . Area was preferred over volume
228 because of the low invasiveness of the measurements. In *Polymastia croceus*, contraction/expansion mainly
229 occur at the papillae level, so the contractile behaviour was estimated from the ratio of expanded papillae
230 (REP) calculated as $REP = P_E / P_{tot}$, where P_E is the number of visible expanded papillae and P_{tot} is total
231 number of visible papillae. Expanded papillae were defined as papillae whose length was at least two and a
232 half times the width.

233 **2.3.5 Pumping rate**

234 Pumping rate was only calculated for *Suberites australiensis* from experiments 2 and 3 (severe hypoxia).
235 Having only one osculum of relatively large size, this species was particularly suitable for investigating
236 changes in pumping rate. To minimize sponge disturbance during the experiment, pumping rate (PR) was
237 derived from the measurement of the sponge osculum cross-sectional area (OSA). In sponges, pumping
238 rate (PR) is correlated with OSA (e.g. Goldstein et al., 2019; Morganti et al., 2021). In the case of *S.*
239 *australiensis*, this relationship was calculated on 20 sponges (following Yahel et al., 2005) and was found as
240 $PR = 6.55 \cdot OSA^{1.43}$ (Fig. S6). Photographs of the oscula with scale were taken daily from T_0 to $T\text{-end}$, on
241 three sponges in each experimental chamber (the same specimens each time point, $n = 9$ per treatment).
242 Since *S. australiensis* has only one osculum, pumping rate was then standardized per sponge volume.

243 **2.3.6 Histology**

244 Histological sections of *Suberites australiensis* from the severe hypoxia experiments were analyzed to
245 calculate the percentage of the sponge body occupied by the aquiferous system (system of connected

246 water channels inside the sponge). At *T-end*, two contracted sponges for each experimental chamber (n = 6
247 per treatment) were fixed and processed following the methods of Strano et al. (2021). Three replicate
248 sections for each sponge were then photographed under a dissecting microscope (Olympus SZ61) and
249 photographed using a Canon EOS 70D digital camera. To calculate the area occupied by the aquiferous
250 system, pictures were analyzed using ImageJ.

251 **2.4 Data analysis**

252 All the statistical data analyses were performed in R version 3.1.3 (R Core Team, 2013), except
253 PERMANOVA models, which were performed using PRIMER v7 with PERMANOVA+ add-on (Anderson et al.,
254 2008; Clarke & Gorley, 2015). Experiments 2 and 3 were analyzed separately. To investigate respiration
255 rate, pumping rate and expansion ratio in *Suberites australiensis* from experiment 2, we used linear mixed-
256 effects models with normally distributed errors and random intercepts (lmer, *lme4* package; Bates et al.,
257 2015). For pumping rate, we added a constant variance function structure (varIdent) to the linear mixed-
258 effects models to allow different variances for each treatment at each time point (lme, *R* package *nlme*;
259 Pinheiro et al., 2021). The constant variance function structure was necessary because the variance of the
260 response variable differed across treatments and experimental days. To investigate the effect of time and
261 treatment on the expansion ratio of *Polymastia croceus* in experiment 3, we used a generalized linear
262 mixed model with beta regression and logit link (*glmmTMB*, Brooks et al. 2017). In all the mixed models,
263 treatment and time were considered fixed effects, while experimental chamber and sponge specimen were
264 considered random effects. The experimental chamber effect was included to address pseudo-replication.
265 For these models, fixed- and random-effect terms were tested using the function *anova* and *ranova* (*R*
266 package *lmerTest*, Kuznetsova et al., 2017), respectively; while post hoc pairwise comparisons were
267 computed on estimated marginal means using *emmeans* (*R* package *emmeans*; Lenth, 2021). The ratio of
268 expanded papillae in *P. croceus* from experiment 2 and expansion ratio in *S. australiensis* from experiment
269 3 were investigated using repeated measure univariate PERMANOVA (Anderson, 2001; 2014), because did
270 not meet the normality assumption for mixed-effects models. Pairwise tests were then calculated using
271 permutation *t*-tests (*R* package *RVAideMemoire*; Hervé, 2021). PERMANOVA and permutation *t*-tests were

272 also used to supplement mixed-effects models when there were concerns about the normality of the
273 residuals (pumping rate in *S. australiensis*). Change over time relative to the initial value of buoyant weight
274 and contracted area, and differences in percentage occupied by the aquiferous system were investigated
275 using Kruskal–Wallis H tests, Welch’s t-tests or Wilcoxon Signed-Rank Tests, depending on the variable.
276 Respiration rates from experiment 1 were $\log(x + 1)$ transformed, and pumping rates were square-root
277 transformed to meet normality assumptions. The goodness of fit, normality and homoscedasticity of the
278 errors were checked for all models by inspecting plots of the normalized residuals and the quantile-quantile
279 plots. All the multiple comparisons were corrected using Benjamini-Hochberg Procedure, but uncorrected
280 p-values are reported in the text. All the statistical analyses made for each variable are reported and
281 summarized in Table S2.

282 Time to event analysis for sponge survival and development of peculiar morphological structures (modified
283 papillae and protruding oscular membranes) was performed using Kaplan-Meier Method, and *p*-values
284 were calculated using the Log Rank Test implemented in the survival *R* package (Therneau, 2021). Median
285 lethal time (LT_{50}) and median time to the development of modified morphological structures were
286 calculated using a logistic model.

287 3 RESULTS

288 3.1 Sponge responses to moderate hypoxia

289 All the sponges of experiment 1 survived the seven days of treatment, except one specimen of *Suberites*
290 *carneus* in the lower DO treatment (6% a.s.), which presented internal necrosis on the final day of the
291 experiment.

292 Mean buoyant weight variation between *T0* and *T-end* ranged between -1% and -1.6% for *Cliona celata* and
293 +2.1% and -0.5% in *Suberites carneus*. There were no differences in buoyant weight variation among
294 treatments for both species, but for *C. celata* there was a significant slight decrease in weight in the 40%
295 a.s. (-1.6%, $p = 0.008$) and 20% a.s. (-1.4%, $p = 0.008$) treatments (Tab. S3; Fig. S7).

296 For *Cliona celata*, there was no significant effect of time or treatment on the respiration rate (Tab. S4).
297 However, pairwise comparisons revealed a significant decrease ($p = 0.028$) in the 20% a.s. treatment
298 between day 0 and 7, and a significant increase ($p = 0.029$) in the 6% a.s. treatment between day 2 and day
299 7, but both became non-significant after the correction for multiple comparisons (Tab. S4). However, the
300 data suggest a coherent temporal pattern in the respiration rate in both 20% a.s. and 6% a.s. treatments. *C.*
301 *celata* respiration rate decreased after two days from the start of the experiment and then increased until
302 the end of the experiment. In contrast, in both the 100% a.s. and 40% a.s. treatments, respiration rate
303 remained stable for the whole duration of the experiment (Fig. 1a).

304 For *Suberites carneus*, there was a significant interaction of time and treatment ($p = 0.007$) on the
305 respiration rate (Tab. S5). Pairwise comparisons revealed a significant decrease in respiration rate between
306 day 0 and 7 ($p < 0.0001$), and day 2 and 7 ($p < 0.0001$) (Tab. S5). The respiration rate also slightly decreased
307 towards the end of the experiment in the 20% a.s. treatment (but not significantly), while in both the 100%
308 a.s. and 40% a.s. treatments, respiration rate remained stable for the duration of the experiment (Fig. 1b).

309 **3.2 Sponge responses to severe hypoxia**

310 **3.2.1 Survival**

311 Sponge survival differed among species, with *Suberites australiensis* more tolerant than *Polymastia croceus*.
312 No mortality was observed for *S. australiensis* in both experiments 2 (5% a.s.) and 3 (1.5% a.s.). In contrast,
313 for *P. croceus*, significant mortality ($p = 0.001$) was observed in sponges exposed to the 1.5% a.s. treatment,
314 starting from day 10 (day 12 when including the hypoxic acclimation), and with a median lethal time of 11.9
315 ± 0.3 days (Fig S8–9). Eight out of 15 sponges had died by the end of the experiment. No mortality was
316 observed for *P. croceus* in the 5% a.s. treatment.

317 **3.2.2 Change in weight and size**

318 For *Polymastia croceus*, buoyant weight variation between T_0 and T_{end} differed among treatments in
319 experiment 3 (1.5% a.s.) ($t = 2.82$, $p = 0.012$), but not in experiment 2 (5% a.s.). Sponges from the 1.5% a.s.
320 treatment experienced a significant decrease in buoyant weight (-7.1%, $t = -5.17$, $p = 0.002$), while the
321 controls did not experience any significant change (Tab. S6; Fig. S10).

322 The relative variation in area of contracted sponges (after stimulating contraction) between T_0 and T_{end}
323 differed significantly between treatments and controls for both *Polymastia croceus* ($W = 12$, $p < 0.0001$)
324 and *Suberites australiensis* ($W = 13$, $p < 0.0001$), but only in experiment 2 (5% a.s.). Both *P. croceus* and *S.*
325 *australiensis*, from the 5% a.s. treatment, experienced an increase in contracted area (+18.9%, $W = 120$, $p =$
326 0.0001 and +18.4%, $W = 105$, $p = 0.008$, respectively). While *S. australiensis* from the control treatment
327 (experiment 2) experienced a decrease in contracted area (-15.3%, $W = 3$, $p = 0.0003$) (Tab. S7; Fig. S11).

328 **3.2.3 Sponge contractile behaviour**

329 Low DO treatments generally induced sponge expansion, but the response differed between species, and it
330 was generally more marked in the 5% a.s. treatment. In *Polymastia croceus*, the ratio of expanded papillae
331 was significantly affected by the interaction between time and treatment in both experiments 2 ($p =$
332 0.0001) and 3 ($p < 0.0001$ and $p = 0.03$) (Tab. S8–9). During experiment 2 (5% a.s.), the treatment induced a
333 progressive expansion of papillae from day 2. The ratio of expanded papillae in sponges from the hypoxic

334 treatment became significantly higher than control sponges from day 6 to the end of the experiment ($p =$
335 0.0002–0.005) (Tab. S8; Fig. 2a). A similar trend was found in experiment 3 (1.5% a.s.), but the ratio of
336 expanded papillae of the treatment sponges was more variable and became significantly different only at
337 day 9 ($p = 0.003$) (Tab. S9; Fig. 2b). In this experiment, we also found a correlation between the ratio of
338 expanded papillae and mortality. Sponges that survived the treatment had a significantly higher maximum
339 ratio of expanded papillae compared to sponges that died, both when the maximum ratio was calculated at
340 the end of the experiment (Welch t -test: $t = 5.3$, $p = 0.0005$) and at day ten, before sponges started to die
341 (Welch t -test: $t = 4.6$, $p = 0.0007$).

342 In *Suberites australiensis*, there was a significant interactive effect of treatment and time ($p < 0.0001$) on
343 the expansion ratio in experiment 2 (5% a.s.), but only an effect of time ($p = 0.01$) in experiment 3 (1.5%
344 a.s.) (Tab. S10–11). For experiment 2 (5% a.s.), pairwise comparisons found significant expansion in sponges
345 (+60%, $p < 0.0001$) between day 0 and 1. Sponges then remained expanded for the whole duration of the
346 experiment, and the expansion ratio was significantly higher in the treatments compared to the controls
347 from the first to the last day of the experiment ($p < 0.0003$) (Tab. S10; Fig. 2c, d).

348 **3.2.4 Morphological modifications**

349 During experiments 2 (5% a.s.) and 3 (1.5% a.s.), some *Polymastia croceus* and *Suberites australiensis*
350 sponges exposed to hypoxic treatments underwent morphological modifications (Fig. 3; S12). In some *P.*
351 *croceus*, the conical papillae showed a progressive elongation, flattening, and, in some cases, spiralization
352 (Fig. S12a–f). This process occurred in both the 5% a.s. and 1.5% a.s. treatments, but morphological
353 changes were more pronounced in lower DO treatment (Fig. S12d–e). Exposed to the 1.5% a.s. treatment,
354 some sponges developed papillae so slender that they could not sustain their weight (Fig. S12d–e). The
355 development of these modified papillae was also associated with an apparent increase in the porosity of
356 the sponge external surface (Fig. S12e). In the 5% a.s. treatment, 73% of the sponges developed modified
357 papillae, starting from day 6. In the 1.5% a.s. treatment, 60% of sponges developed modified papillae,
358 starting from day 2, from the beginning of the final treatment (day 4 considering hypoxic acclimation
359 period) (Fig. S13). The median time of development of these morphological structures (considering only the

360 sponges that developed them) was 7.2 ± 0.2 days in the 5% a.s. treatment and 4.6 ± 0.4 days in the 1.5%
361 a.s. treatment (Fig. S14). Although not significant ($\chi^2 = 3.62$, $p = 0.057$), a relationship between modified
362 papillae and survival was found. Among the *P. croceus* that survived the 1.5% a.s. treatment, six had
363 developed modified papillae, while one had not. While among the sponges that died following the 1.5% a.s.
364 treatment, three had developed modified papillae, while five had not.

365 In the the 5% a.s. treatment, 53% of *Suberites australiensis* developed a semi-transparent protruding
366 membrane surrounding the oscula. This membrane progressively reduced the oscular-cross sectional area
367 (Fig. 3; S12g–i). The median time it took for these protruding oscular membranes to become noticeable
368 (considering only the sponges that developed them) was 5.1 ± 0.2 days (Fig. S14). By the end of the
369 experiment, 53% of sponges had developed these structures (Fig. S13).

370 3.2.5 Histology

371 Histological analyses indicated that hypoxia influences the percentage of the sponge body occupied by the
372 aquiferous system in *Suberites australiensis*. At the end of experiment 2 (5% a.s.), treatment sponges had a
373 significantly higher percentage of aquiferous system ($t = -9.82$, $p < 0.0001$), compared to the controls (35.9
374 $\pm 7.1\%$ vs $6.4 \pm 1.8\%$). No significant differences were found for experiment 3 (1.5% a.s.) (Fig. 3; S12j–m;
375 S15).

376 3.2.6 Pumping rate

377 Oxygen concentration significantly affected the pumping rate of *Suberites australiensis* in both experiments
378 2 (5% a.s.) and 3 (1.5% a.s.) (Tab. S12–15). In experiment 2 (5% a.s.), there was significant interaction of
379 treatment and time ($p = 0.0001$, linear mixed-effects model; Tab. S12). Pumping rate significantly increased
380 from day 0 to 1 ($p < 0.0001$), remained stable from day 1 to 2, and then decreased from day 2 to 3 ($p <$
381 0.0001) and from day 3 to 4 ($p = 0.005$) (Tab. S12). Sponges from the 5% a.s. treatment had a significantly
382 higher pumping rate than the control at day 1 ($p = 0.007$) and 2 ($p = 0.002$) (Tab. S12; Fig. 2e). Similar
383 results were given by PERMANOVA (Tab. S13). For experiment 3 (1.5% a.s.), both the linear mixed-effects
384 model and PERMANOVA revealed significant interaction between time and treatment ($p = 0.049$ and $p =$

385 0.002, respectively) on the pumping rate of *S. australiensis* (Tab. S14–15). However, differences were less
386 marked compared to experiment 2 and pairwise comparisons only revealed a slight decrease of pumping
387 rate of treatment sponges between day 0 and 14 ($p = 0.0002$) (Tab. S14; Fig. 2f).

388 3.2.7 Respiration rate

389 In experiment 2 (5% a.s.), linear mixed-effects models only revealed a significant effect of time on the
390 respiration rate, for both *Polymastia croceus* and *Suberites australiensis* (Tab. S16–17; Fig. 4). In *P. croceus*,
391 pairwise comparisons revealed a slightly higher respiration rate of the controls at day 12 compared to day 0
392 ($p = 0.008$) and day 5 ($p = 0.004$), but no differences between controls and treatments at any time. In *S.*
393 *australiensis*, pairwise comparisons revealed a slightly lower respiration rate at day 12 compared to day 0
394 ($p = 0.008$) and day 5 ($p = 0.005$) in control sponges; while in treatment sponges, respiration rate was
395 slightly lower at day 5 ($p = 0.032$) and 12 ($p = 0.016$) compared to day 0, but also in this case, there was no
396 significant difference between treatments and controls at any time point.

397 4 DISCUSSION

398 Hypoxia has become an increasingly common problem in the marine environment and will likely become
399 worse in the future (Diaz & Rosenberg, 2011). Nevertheless, the direct effects of hypoxia on marine
400 organisms are still very poorly studied (Vaquer-Sunyer & Duarte, 2008). We describe the first multi-species
401 experiment from two oceans to test sponge tolerance, behaviour, and physiological responses to oxygen
402 concentrations as low as 1.5% a.s. ($0.13 \text{ mg O}_2 \text{ L}^{-1}$) for up to 12 days. We found that sponges are generally
403 very tolerant to low DO irrespective of species or location. Only *Polymastia croceus* showed mortality in the
404 lower DO treatment ($0.13 \text{ mg O}_2 \text{ L}^{-1}$, $\text{LT}_{50} = 286 \text{ h}$). Furthermore, our results suggest that sponges can
405 display species-specific acclimation, including physiological, morphological and behavioural changes, in
406 response to severe hypoxia that might help them survive periods of very low oxygen. Our study also
407 suggests that the same species can show different adaptive strategies for different degrees of hypoxia.

408 4.1 Sponge response to hypoxia

409 Our results suggest that sub-lethal oxygen thresholds for most sponges are in the range of 6–20% a.s.
410 ($0.48\text{--}1.56 \text{ mg O}_2 \text{ L}^{-1}$), while lethal thresholds are lower than 5% a.s. ($0.4 \text{ mg O}_2 \text{ L}^{-1}$). These pieces of
411 evidence are consistent with Mills et al. (2014) for *Halichondria panicea*, which showed a sub-lethal
412 response starting from 17% air saturation. However, our results contrast with Mills et al. (2018) studying
413 *Tethya wilhelma*, which did not show any response down to 4% a.s. ($0.27 \text{ mg O}_2 \text{ L}^{-1}$). The very high
414 tolerance of *T. wilhelma* could be explained by the extremely low metabolism of *Tethya* species generally
415 (Leys & Kahn, 2018), and by their very small size (0.5–1 cm) (Sarà et al., 2001). Of the two species we
416 exposed to the lowest DO concentration (1.5% a.s., $0.13 \text{ mg O}_2 \text{ L}^{-1}$), only *Polymastia croceus* showed
417 mortality, while all the *Suberites australiensis* survived the 12 days of treatment conditions. This differential
418 response could be due to the different habitats where these species are usually found. *Polymastia croceus*
419 lives on rocky reefs, while *S. australiensis* lives on sediments in bays and semi-enclosed basins, where
420 hypoxic events are more likely to occur (Diaz & Rosenberg, 2008; de Cook, 2010).

421 Some sponges can live in anoxic conditions for several months, such as the sponges of the family
422 Raspailidae found in the deeper cliffs of Lough Hyne (Bell & Barnes, 2000; McAllen et al., 2009). Schuster et

423 al. (2021) suggested that this tolerance could be conferred by specific bacterial symbionts, which are able
424 to carry out anaerobic metabolism. In addition, these sponges living in anoxia are all thin crusts, with a very
425 high surface-to-volume ratio, which could favour the exchange of gases and the release of metabolic waste
426 (Levin et al., 1991). Other examples of sponges living in very low oxygen conditions are the ones found at
427 the edges of Oxygen Minimum Zones (OMZ) (Mosch et al., 2012). These sponges can live with a consistent
428 oxygen concentration as low as $0.13\text{mg O}_2\text{ L}^{-1}$ (Wishner et al., 1995, Murty et al., 2009). Sponges are not the
429 only organisms able to live in OMZs. Many representatives of other phyla live in these extremely hypoxic
430 conditions, where they benefit from the rich supply of organic matter. However, since OMZs have existed
431 over geological timescales, organisms have had the time to evolve specific adaptations to cope with
432 permanent hypoxia (Levin, 2003). Therefore, these organisms cannot be used to generalize tolerance to
433 periodic hypoxic events experienced by organisms usually living in fully oxygenated waters.

434 The degree of hypoxia tolerance in sponges could also be influenced by the abundance and diversity of
435 sponge-associated microbial symbionts. Based on bacterial biomass, sponges are generally divided into
436 “low microbial abundance” (LMA) or “high microbial abundance” (HMA) species (Hentschel et al., 2003).
437 Bacterial densities in HMA sponges are generally two to four orders of magnitude higher than in LMA
438 sponges and can constitute up to 35% of the total sponge biomass (Vacelet, 1975; Hentschel et al., 2006).
439 Sponges with HMA tend to have a lower choanocyte chamber density, and a slower pumping rate
440 compared LMA sponges (Lavy et al., 2016), which means HMA species might have a lower ability to
441 ventilate in low oxygen conditions. Furthermore, HMA species generally have a higher metabolic cost than
442 LMA species, and therefore a higher oxygen requirement (Leys & Kahn, 2018). Although these differences
443 suggest that LMA sponges might be better adapted to hypoxic conditions, HMA species have a higher
444 diversity of microbial symbionts that could help them cope with low oxygen conditions (Hoffmann et al.,
445 2005, Lavy et al., 2016). All the sponges for which responses to hypoxia has been investigated so far are
446 LMA species (or are likely to be, based on known congeners, see Kamke et al., 2010; Mills et al., 2014;
447 Moitinho-Silva et al., 2017). Therefore, future research is needed to investigate the response of HMA

448 sponges to hypoxia and shed light on possible differences between LMA and HMA sponges and the
449 mechanisms involved.

450 Some organism's abilities to tolerate hypoxia result from their physiological ability to lower metabolism and
451 oxygen demand (McAllen et al., 1999; Altieri, 2019). Instead, other species switch from aerobic to
452 anaerobic metabolism or a combination of the two (Altieri & Diaz, 2019). Our results suggest that all our
453 species (except *Suberites carnosus*) have respiration rates at 5-6% a.s. that are comparable to sponges in
454 normoxic conditions. This is consistent with what was found in *Geodia barretti* and *H. panicea*, suggesting
455 that sponges have a common ability to uptake oxygen at very low concentrations in the surrounding
456 environment (Leys & Kahn, 2018). In *Cliona celata*, hypoxic water initially resulted in a decrease in the
457 respiration rate, which then increased back to pre-treatment levels after seven days of exposure. This
458 suggests that the sponges gradually adjusted to hypoxic conditions. In *S. carnosus*, instead, the respiration
459 rate remained stable after two days of exposure to low dissolved oxygen, but it more than halved after
460 seven days. This response may allow *S. carnosus* to cope with long periods of hypoxia, in which sponges
461 decrease their metabolism, as has been reported for other organisms (Hagerman, 1998; Mentel et al.,
462 2014). Although our study shows that sponges can perform aerobic metabolism when exposed to
463 extremely low oxygen concentrations, the presence of anaerobic metabolism cannot be excluded and
464 needs further investigation.

465 Sponge species exposed to the lowest DO concentrations (0.4 and 0.13 mg O₂ L⁻¹) also showed other
466 phenotypic modifications that could represent adaptive strategies to cope with hypoxia. In *Suberites*
467 *australiensis*, hypoxic water (0.4 mg O₂ L⁻¹) induced expansion of the sponge body and the aquiferous
468 system that lasted for the duration of the experiment. This expansion was likely semi-permanent as it
469 persisted after inducing the contraction and corresponded to a reorganization of the sponge aquiferous
470 system at the histological level. These behavioural and morphological changes are likely beneficial for the
471 sponge, as higher internal water flow corresponds to an increase in oxygen that can be taken up. The body
472 expansion was accompanied by a marked increase in the pumping rate that then dropped after two days.
473 The pumping rate increase could be a strategy to increase ventilation and oxygen availability, similarly to

474 other animals when exposed to hypoxic waters (Hagerman, 1998). However, the successive decrease in
475 pumping rate (after two days) and the gradual production of a membrane to close the oscula remains
476 unclear but could represent a trade-off between increasing ventilation and keeping the energetic cost of
477 pumping reasonable. In *S. australiensis*, body expansion is correlated with an increase in osculum area, and
478 osculum area is the main determinant of pumping rate in this and many other species (Morganti et al.,
479 2021; Goldstein et al., 2019). Perhaps the increase in pumping rate only represents a physiological
480 consequence of the body expansion and is then quickly brought back to normal, decreasing the osculum
481 size by producing an oscular membrane. These physiological and morphological changes of *S. australiensis*
482 described above was not present on sponges exposed to more severe hypoxia ($0.13 \text{ mg O}_2 \text{ L}^{-1}$). This could
483 mean that the same sponge species may display different adaptive strategies to cope with decreased
484 oxygen depending on the oxygen concentration. At 0.4 mg L^{-1} , oxygen might still be sufficient to support
485 regular metabolism, but sponges may need to increase the amount of water flowing through their bodies
486 to absorb the oxygen needed. However, 0.13 mg L^{-1} might be too low a DO concentration, and sponges
487 might decrease their metabolism to cope with lack of oxygen, similarly to other metazoans (Hagerman
488 1998, Mentel et al., 2014).

489 *Polymastia croceus* also showed a behavioural change in response to hypoxic conditions: hypoxic water at
490 0.4 mg L^{-1} induced the progressive expansion of sponge papillae (where inhalant and exhalant channels are
491 found), that was significantly greater than in the control sponges. It is unlikely that the papillae expansion
492 represents an increase in sponge filtering activity because the respiration rate was very similar in the
493 treatments and the controls. Therefore, sponges might expand their papillae to increase the volume
494 occupied by the aquiferous systems, as in the case of *Suberites australiensis*, but also to access more
495 oxygenated water further from the bottom. A similar response occurred in sponges exposed to 0.13 mg L^{-1}
496 but with much more variability across specimens, and the statistical test did not detect any change.
497 Interestingly, sponges that survived after the 12-day treatment had a significantly higher ratio of expanded
498 papillae than sponges that died, suggesting that expansion might help cope with severe hypoxic conditions.

499 Along with behavioural changes, *Polymastia* underwent morphological modifications that could help to
500 tolerate low DO. Papillae become thinner and flattened, and some even spiralized. These modifications of
501 the papillae could increase the surface-to-volume ratio and help oxygen diffusion (Levin et al., 1991). The
502 elongation of papillae, which accompanies the thinning, could be an evolutionary relic of a process that
503 moved the inhalant pores of the papillae as far as possible from the surface. However, in the lowest DO
504 treatment, papillae often lost their vertical orientation and laid horizontally on the sponge surface. We
505 hypothesized that the new orientation of papillae was a consequence of their thinning process: probably
506 papillae became so thin that they could not support their weight anymore. Interestingly, sponges that
507 developed modified papillae showed less mortality than sponges that did not, although the evidence is not
508 strong enough to claim this with confidence ($p = 0.057$). Therefore, these structures may not only represent
509 a stress response, but could provide an advantage to the sponge. Further research is needed on this topic
510 needed to elucidate the function of these structures.

511 Despite the remarkable tolerance of sponges to hypoxia observed in laboratory conditions, field
512 observations suggest that severe hypoxic/anoxic events can catastrophically affect sponge populations.
513 Mass mortalities of sponges following hypoxic/anoxic events have been reported both in temperate and
514 tropical ecosystems (Stachowitsch, 1984; Altieri et al., 2017; Chu et al., 2018; Johnson et al., 2018; Kealoha
515 et al., 2020). For example, in a hypoxic/anoxic event in the Gulf of Trieste, all the sponges living in several
516 hundred km² died within 2-3 days (Stachowitsch, 1984). Some anemones survived up to a week, but
517 virtually all organisms were dead within two weeks from the onset. Altieri et al. (2017) also reported
518 widespread mortality of sponges and corals following a hypoxic event ($\sim 0.5 \text{ mg O}_2 \text{ L}^{-1}$) that occurred in
519 Bocas del Toro, Panama. Since this study focused on corals, it is unclear what proportion of the sponges
520 were affected and if some species were more tolerant than others. These reports highlight that hypoxic
521 events, in their most severe form, leave no survivors.

522 Furthermore, it is possible that in natural conditions, other factors combine with low dissolved oxygen. For
523 example, a recent meta-analysis showed that in marine organisms, increased temperature reduces survival
524 times under hypoxia by 74% on average and increased median lethal concentration by 16% on average

525 (Vaquer-Sunyer & Duarte, 2011). Another meta-analysis showed that hydrogen sulphide (H₂S) also reduces
526 survival time of marine organisms under hypoxia by an average of 30% (Vaquer-Sunyer & Duarte, 2010).
527 Acidification was shown to have additive or synergistic negative effects combined with hypoxia (Gobler &
528 Baumann, 2016; Steckbauer et al., 2020). Since all these factors usually co-occur during hypoxic events, *in*
529 *situ* sponge thresholds to hypoxia could be lower than determined through single stressor laboratory
530 experiments (Diaz & Rosenberg, 1995; Steckbauer et al., 2020). Future experiments that evaluate the
531 combined effect of these factors will be crucial to understand the full response of sponges to hypoxia in
532 natural ecosystems.

533 Diel oxygen variation is another factor that could influence organism tolerance to hypoxia in natural
534 conditions. In the photic zone of marine ecosystems, dissolved oxygen generally increases during the day
535 because of photosynthesis and decreases at night because of aerobic respiration (Kroeker et al. 2019). The
536 amplitude of these diel fluctuations can sometimes lead to hypoxia or complete anoxia at night and
537 supersaturation in peak sunny hours, or both (Diaz & Breitburg, 2009). These extreme oxygen dynamics
538 have been reported from a wide variety of macro- and micro-habitats from both tropical and temperate
539 ecosystems, such as intertidal reef platforms, tide pools, semi-enclosed basins, tropical lagoons and the
540 boundary layer around macroalgal canopies (Morris and Taylor, 1983, Frieder et al. 2012, Cornwall et al.,
541 2013, Gruber et al., 2017, Trowbridge et al. 2017, Hughes et al., 2020). Diurnal fluctuations in oxygen can
542 produce different responses from static exposure in laboratory experiments, which may either
543 overestimate or underestimate the emergent effects of hypoxia in natural environments (Bumett & Stickle,
544 2001). Therefore, future experiments will need to account for current and future temporal variability in
545 oxygen concentration to accurately forecast the emergent ecological effects of deoxygenation (Kroeker et
546 al., 2019).

547 **4.2 Hypoxia Tolerance of sponges compared to other sessile organisms**

548 Marine organisms have very variable tolerance to low dissolved oxygen, with lethal thresholds ranging from
549 8.6 mg O₂ L⁻¹ for the first larval zoea stage of the crustacean *Cancer irroratus*, to resistance to complete
550 anoxia as in the case of the sea anemone *Metridium senile* and the oyster *Crassostrea virginica* (Wahl 1984;

551 Vaquer-Sunyer and Duarte, 2008). Sessile organisms are generally more tolerant than mobile ones, which is
552 likely due to them not being able to escape hypoxic conditions (Altieri & Diaz, 2019). Therefore, sessile
553 organisms that experience these conditions must have evolved other adaptive strategies to cope with
554 reduced oxygen (Diaz & Rosenberg, 1995).

555 Here we provide new evidence to support the hypothesis that sponges are one of the groups of sessile
556 organisms that are more tolerant to hypoxia and could be favoured in future deoxygenated oceans. Other
557 phyla, such as cnidarians and bivalves, include very tolerant species that can cope with prolonged periods
558 of anoxia (Fig. 13). This is not surprising since tolerance to severe hypoxia/anoxia is a widespread feature in
559 the animal world, and many organisms independently evolved this feature to cope with local conditions
560 (Hochachka & Lutz, 2001; Nilsson & Renshaw, 2004; Vaquer-Sunyer & Duarte, 2008). This ability is not
561 restricted to invertebrates and includes higher animals such as fish and reptiles (Milton & Prentice, 2007;
562 Vornanen et al., 2009).

563 What makes sponges unique as a phylum is their widespread tolerance to hypoxia. All the species
564 investigated so far have been shown to cope with very low levels of dissolved oxygen. In contrast, other
565 phyla have a much wider range of tolerances, with some species resistant to anoxia and others very
566 sensitive to decreased oxygen (Fig. 5). For example, in sessile cnidarians, lethal hypoxia thresholds range
567 between 0 and 4 mg O₂ L⁻¹, while sublethal ones are between 0.71 and 4.56 mg O₂ L⁻¹ (Mangum, 1980;
568 Dodds et al., 2007). In sessile bivalves, lethal thresholds range between 0 and 2 mg O₂ L⁻¹, with the sub-
569 lethal threshold being 3.1 mg O₂ L⁻¹ for *Mytilus galloprovincialis* (de Zwaan et al., 1991; Woo et al., 2013).
570 Sponges, instead, show much less variation with known lethal thresholds that are lower than 0.5 mg O₂ L⁻¹,
571 and sublethal thresholds that range between 0.27 and 1.56 mg O₂ L⁻¹ (Mills et al., 2014; 2018) (Fig. 5). It is
572 worth noting that lethal thresholds are highly dependent on the time of exposure. In the studies we
573 considered, these ranged from a few days to weeks. However, there were no noticeable differences in the
574 experimental duration and the median lethal time for the different organisms. Therefore, we believe that
575 differences in time of exposure do not represent a bias in our comparison. In contrast, sublethal responses
576 (e.g. changes in respiration rate, behaviour, and feeding activity) usually have rapid time-to-onset, so they

577 will likely be independent of exposure time. The high tolerance of sponges to hypoxia compared to other
578 organisms can be explained by the evolutionary history of this group. Sponges are one of the most ancient
579 groups of metazoans. They likely evolved before the Marinoan glaciation (657-645 million years ago), when
580 oxygen was perhaps less than 10% of present atmospheric concentration (Love et al., 2009; Maloof et al.,
581 2010; Brocks et al., 2017; Whelan et al., 2017; Cole et al., 2020; Turner, 2021). Modern sponges might have
582 retained an ancestral condition concerning oxygen requirements (Mills et al., 2014; 2018). Therefore, it is
583 more likely that sponges unable to survive severe hypoxia today (e.g., *P. croceus*) have lost certain key
584 ancestral adaptations to hypoxia, rather than hypoxia-tolerant lineages (e.g., *S. australiensis*) having
585 evolved relatively new capacities for hypoxia tolerance (Müller et al., 2012). Likewise, other animals which
586 might have evolved in similar conditions, such as ctenophores, also show great resistance to hypoxia
587 (Thuesen et al., 2005). Therefore, we speculate that sponges' long evolutionary history could give these
588 organisms an adaptive advantage in future deoxygenated oceans, since they may have experienced similar
589 conditions in past geological eras.

590

591 **CONCLUSIONS**

592 Overall, sponges show high tolerance to low dissolved oxygen compared to all the other phyla of sessile
593 marine organisms that have been studied. Species-specific phenotypic plasticity appears to help these
594 organisms to overcome hypoxic events, and future research will need to elucidate the mechanisms behind
595 these changes. This exceptional adaptive capacity of sponges could derive from their ancient evolutionary
596 origin and could confer sponges a competitive advantage in future deoxygenated oceans over other
597 organisms (Mills et al., 2014; Schuster et al., 2021).

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606 AUTHOR CONTRIBUTIONS

607 V.M., J.J.B and R.M. designed the study. V.M., F.S. and L.H. realized the experimental set-up. V.M. and F.S.
608 conducted the experiments. V.M. and L.W. analyzed the data. V.M. and J.J.B wrote the original draft. All the
609 authors participated in interpreting the results and contributed to the revision of the manuscript.

610 DATA AVAILABILITY STATEMENT

611 All data and the R code used in this paper are available on Figshare repository:
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620 **REFERENCES**

- 621 Altieri, A. H., & Diaz, R. J. (2019). Dead zones: oxygen depletion in coastal ecosystems. In: C. Sheppard
622 (Ed.), *World Seas: An Environmental Evaluation* (pp. 453–473). Academic Press.
623 <https://doi.org/10.1016/b978-0-12-805052-1.00021-8>
- 624 Altieri, A. H., & Gedan, K. B. (2015). Climate change and dead zones. *Global Change Biology*, 21(4),
625 1395–1406. <https://doi.org/10.1111/gcb.12754>
- 626 Altieri, A. H., Harrison, S. B., Seemann, J., Collin, R., Diaz, R. J., & Knowlton, N. (2017). Tropical dead
627 zones and mass mortalities on coral reefs. *Proceedings of the National Academy of Sciences*, 114(14),
628 3660–3665. <https://doi.org/10.1073/pnas.1621517114>
- 629 Anderson, M. J. (2001). A new method for non-parametric multivariate analysis of variance. *Austral*
630 *Ecology*, 26(1), 32–46. <https://doi.org/10.1111/j.1442-9993.2001.01070.pp.x>
- 631 Anderson, M. J. (2014). Permutational multivariate analysis of variance (PERMANOVA). *Wiley statsref:*
632 *statistics reference online*, 1–15. <https://doi.org/10.1002/9781118445112.stat07841>
- 633 Anderson, M. J., Gorley, R. N. & Clarke, K. R. (2008). *PERMANOVA+ for PRIMER: Guide to software and*
634 *statistical methods*. PRIMER-E, Plymouth, UK.
- 635 Ayling, A. L. (1983). Factors affecting the spatial distributions of thinly encrusting sponges from
636 temperate waters. *Oecologia*, 60(3), 412–418. <https://doi.org/10.1007/bf00376861>
- 637 Bates, D., Maechler, M., Bolker, B., & Walker, S. (2015). Fitting Linear Mixed-Effects Models Using lme4.
638 *Journal of Statistical Software*, 67(1), 1–48. <https://doi:10.18637/jss.v067.i01>
- 639 Bell, J. J. (2008). The functional roles of marine sponges. *Estuarine, Coastal and Shelf Science*, 79(3),
640 341–353. <https://doi.org/10.1016/j.ecss.2008.05.002>

- 641 Bell, J. J., & Barnes, D. K. (2000). The distribution and prevalence of sponges in relation to
642 environmental gradients within a temperate sea lough: vertical cliff surfaces. *Diversity and*
643 *Distributions*, 6(6), 283–303. <https://doi.org/10.1046/j.1472-4642.2000.00091.x>
- 644 Bell, J. J., McGrath, E., Kandler, N. M., Marlow, J., Beepat, S. S., Bachtiar, R., Shaffer, M. R., Mortimer, C.,
645 Micaroni, V., Mobilia, V., Rovellini, A., Harris, B., Farnham, E., Strano, F., & Carballo, J. L. (2020).
646 Interocean patterns in shallow water sponge assemblage structure and function. *Biological*
647 *Reviews*, 95(6), 1720–1758. <https://doi.org/10.1111/brv.12637>
- 648 Bijma, J., Pörtner, H. O., Yesson, C., & Rogers, A. D. (2013). Climate change and the oceans—What does
649 the future hold?. *Marine Pollution Bulletin*, 74(2), 495–505.
650 <https://doi.org/10.1016/j.marpolbul.2013.07.022>
- 651 Breitburg, D. L. (1990). Near-shore hypoxia in the Chesapeake Bay: patterns and relationships among
652 physical factors. *Estuarine, Coastal and Shelf Science*, 30(6), 593–609. [https://doi.org/10.1016/0272-](https://doi.org/10.1016/0272-7714(90)90095-9)
653 [7714\(90\)90095-9](https://doi.org/10.1016/0272-7714(90)90095-9)
- 654 Brocks, J. J., Jarrett, A. J., Sirantoine, E., Hallmann, C., Hoshino, Y., & Liyanage, T. (2017). The rise of
655 algae in Cryogenian oceans and the emergence of animals. *Nature*, 548(7669), 578-581.
656 <https://doi.org/doi:10.1038/nature23457>
- 657 Brooks, M. E., Kristensen, K., Van Benthem, K. J., Magnusson, A., Berg, C. W., Nielsen, A., Skaug, H. J.,
658 Machler, M. & Bolker, B.M. (2017). glmmTMB balances speed and flexibility among packages for zero-
659 inflated generalized linear mixed modeling. *The R Journal*, 9(2), 378–400. [https://doi.org/10.32614/rj-](https://doi.org/10.32614/rj-2017-066)
660 [2017-066](https://doi.org/10.32614/rj-2017-066)
- 661 Bumett, L. E., & Stickle, W. B. (2001). Physiological responses to hypoxia. In: N. N. Rabalais, R. E. Turner
662 (Eds). *Coastal and Estuarine Studies, Volume 58. Coastal hypoxia: consequences for living resources and*
663 *ecosystems*, (pp. 101–114). AGU Publications. <https://doi.org/10.1029/ce058p0101>

- 664 Chu, J. W., Curkan, C., & Tunnicliffe, V. (2018). Drivers of temporal beta diversity of a benthic
665 community in a seasonally hypoxic fjord. *Royal Society Open Science*, 5(4), 172284.
666 <https://doi.org/10.1098/rsos.172284>
- 667 Clarke, K. R., & Gorley, R. N. (2015). *Getting started with PRIMER v7*. PRIMER-E, Plymouth, UK.
- 668 Cole, D. B., Mills, D. B., Erwin, D. H., Sperling, E. A., Porter, S. M., Reinhard, C. T., & Planavsky, N. J.
669 (2020). On the co-evolution of surface oxygen levels and animals. *Geobiology*, 18(3), 260–281.
670 <https://doi.org/10.1111/gbi.12382>
- 671 Cook, S. D. C. (2010). *New Zealand coastal marine invertebrates*. Canterbury University Press.
- 672 Cornwall, C. E., Hepburn, C. D., Pilditch, C. A., & Hurd, C. L. (2013). Concentration boundary layers
673 around complex assemblages of macroalgae: Implications for the effects of ocean acidification on
674 understory coralline algae. *Limnology and Oceanography*, 58(1), 121–130.
675 <https://doi.org/10.4319/lo.2013.58.1.0121>
- 676 De Zwaan, A., Cortesi, P., Van den Thillart, G., Roos, J., & Storey, K. B. (1991). Differential sensitivities to
677 hypoxia by two anoxia-tolerant marine molluscs: a biochemical analysis. *Marine Biology*, 111(3), 343–
678 351. <https://doi.org/10.1007/bf01319405>
- 679 Desai, D. V., & Prakash, S. (2009). Physiological responses to hypoxia and anoxia in *Balanus amphitrite*
680 (Cirripedia: Thoracica). *Marine Ecology Progress Series*, 390, 157–166.
681 <https://doi.org/10.3354/meps08155>
- 682 Diaz, R. J., & Breitburg, D. L. (2009). The hypoxic environment. In: J. G. Richards, A. P. Farrell, C. J.
683 Brauner (Eds) *Fish physiology, Volume 27* (pp. 1–23). Academic Press. <https://doi.org/10.1016/b978-0-12-809633-8.03213-1>
684
- 685 Diaz, R. J., & Rosenberg, R. (1995). Marine benthic hypoxia: a review of its ecological effects and the
686 behavioural responses of benthic macrofauna. *Oceanography and marine biology. An annual review*,
687 33, 245–03.

- 688 Diaz, R. J., & Rosenberg, R. (2008). Spreading dead zones and consequences for marine
689 ecosystems. *Science*, 321(5891), 926–929. <https://doi.org/10.1126/science.1156401>
- 690 Diaz, R. J., & Rosenberg, R. (2011). Introduction to environmental and economic consequences of
691 hypoxia. *International Journal of Water Resources Development*, 27(1), 71–82.
692 <https://doi.org/10.1080/07900627.2010.531379>
- 693 Dodds, L. A., Roberts, J. M., Taylor, A. C., & Marubini, F. (2007). Metabolic tolerance of the cold-water
694 coral *Lophelia pertusa* (Scleractinia) to temperature and dissolved oxygen change. *Journal of*
695 *Experimental Marine Biology and Ecology*, 349(2), 205–214.
696 <https://doi.org/10.1016/j.jembe.2007.05.013>
- 697 Doney, S. C. (2010). The growing human footprint on coastal and open-ocean biogeochemistry. *Science*,
698 328(5985), 1512–1516. <https://doi.org/10.1126/science.1185198>
- 699 Ellington, W. R. (1982). Metabolic responses of the sea anemone *Bunodosoma cavernata* (Bosc) to
700 declining oxygen tensions and anoxia. *Physiological Zoology*, 55(3), 240–249.
701 <https://doi.org/10.1086/physzool.55.3.30157888>
- 702 Frieder, C. A., Nam, S. H., Martz, T. R., & Levin, L. A. (2012). High temporal and spatial variability of
703 dissolved oxygen and pH in a nearshore California kelp forest. *Biogeosciences*, 9(10), 3917–3930.
704 <https://doi.org/10.5194/bg-9-3917-2012>
- 705 Gobler, C. J., & Baumann, H. (2016). Hypoxia and acidification in ocean ecosystems: coupled dynamics
706 and effects on marine life. *Biology Letters*, 12(5), 20150976. <https://doi.org/10.1098/rsbl.2015.0976>
- 707 Goldstein, J., Riisgård, H. U., & Larsen, P. S. (2019). Exhalant jet speed of single-ostium explants of the
708 demosponge *Halichondria panicea* and basic properties of the sponge-pump. *Journal of Experimental*
709 *Marine Biology and Ecology*, 511, 82–90. <https://doi.org/10.1016/j.jembe.2018.11.009>

- 710 Gruber, R. K., Lowe, R. J., & Falter, J. L. (2017). Metabolism of a tide-dominated reef platform subject to
711 extreme diel temperature and oxygen variations. *Limnology and Oceanography*, 62(4), 1701–1717.
712 <https://doi.org/10.1002/lno.10527>
- 713 Haas, A. F., Smith, J. E., Thompson, M., & Deheyn, D. D. (2014). Effects of reduced dissolved oxygen
714 concentrations on physiology and fluorescence of hermatypic corals and benthic algae. *PeerJ*, 2, e235.
715 <https://doi.org/10.7717/peerj.235>
- 716 Hagerman, L. (1998). Physiological flexibility; a necessity for life in anoxic and sulphidic habitats.
717 *Hydrobiologia*, 375, 241–254. https://doi.org/10.1007/978-94-017-2864-5_20
- 718 Hervé, M. (2021). *RVAideMemoire: testing and plotting procedures for biostatistics*. R package version
719 0.9-79. <https://cran.r-project.org/web/packages/RVAideMemoire/index.html>
- 720 Hentschel, U., Fieseler, L., Wehrl, M., Gernert, C., Steinert, M., Hacker, J., & Horn, M. (2003). Microbial
721 diversity of marine sponges. In: W. E. G. Müller (Ed) *Progress in Molecular and Subcellular Biology*.
722 *Sponges (Porifera)* (pp 59–88). Springer. https://doi.org/10.1007/978-3-642-55519-0_3
- 723 Hentschel, U., Usher, K. M., & Taylor, M. W. (2006). Marine sponges as microbial fermenters. *FEMS*
724 *microbiology ecology*, 55(2), 167–177. <https://doi.org/10.1111/j.1574-6941.2005.00046.x>
- 725 Hochachka, P. W., & Lutz, P. L. (2001). Mechanism, origin, and evolution of anoxia tolerance in animals.
726 *Comparative Biochemistry and Physiology Part B: Biochemistry and Molecular Biology*, 130(4), 435–459.
727 [https://doi.org/10.1016/s1096-4959\(01\)00408-0](https://doi.org/10.1016/s1096-4959(01)00408-0)
- 728 Hoffmann, F., Larsen, O., Thiel, V., Rapp, H. T., Pape, T., Michaelis, W., & Reitner, J. (2005). An
729 anaerobic world in sponges. *Geomicrobiology Journal*, 22(1–2), 1–10.
730 <https://doi.org/10.1080/01490450590922505>
- 731 Hughes, D. J., Alderdice, R., Cooney, C., Kühl, M., Pernice, M., Voolstra, C. R., & Suggett, D. J. (2020).
732 Coral reef survival under accelerating ocean deoxygenation. *Nature Climate Change*, 10(4), 296–307.
733 <https://doi.org/10.1038/s41558-020-0737-9>

- 734 IPCC (2021). Summary for Policymakers. In: V. Masson-Delmotte, P. Zhai, A. Pirani, S. L. Connors, C.
735 Péan, S. Berger, N. Caud, Y. Chen, L. Goldfarb, M. I. Gomis, M. Huang, K. Leitzell, E. Lonnoy, J.B.R.
736 Matthews, T. K. Maycock, T. Waterfield, O. Yelekçi, R. Yu & B. Zhou (Eds.). *Climate Change 2021: The*
737 *Physical Science Basis. Contribution of Working Group I to the Sixth Assessment Report of the*
738 *Intergovernmental Panel on Climate Change*. Cambridge University Press. In Press.
- 739 Johnson, M. D., Rodriguez, L. M., & Altieri, A. H. (2018). Shallow-water hypoxia and mass mortality on a
740 Caribbean coral reef. *Bulletin of Marine Science*, 94(1), 143–144.
741 <https://doi.org/10.5343/bms.2017.1163>
- 742 Johnson, P. D., & McMahon, R. F. (1998). Effects of temperature and chronic hypoxia on survivorship of
743 the zebra mussel (*Dreissena polymorpha*) and Asian clam (*Corbicula fluminea*). *Canadian Journal of*
744 *Fisheries and Aquatic Sciences*, 55(7), 1564–1572. <https://doi.org/10.1139/f98-030>
- 745 Kamke, J., Taylor, M. W., & Schmitt, S. (2010). Activity profiles for marine sponge-associated bacteria
746 obtained by 16S rRNA vs 16S rRNA gene comparisons. *The ISME Journal*, 4(4), 498–508.
747 <https://doi.org/10.1038/ismej.2009.143>
- 748 Kealoha, A. K., Doyle, S. M., Shamberger, K. E., Sylvan, J. B., Hetland, R. D., & Di Marco, S. F. (2020).
749 Localized hypoxia may have caused coral reef mortality at the Flower Garden Banks. *Coral Reefs*, 39(1),
750 119–132. <https://doi.org/10.1007/s00338-019-01883-9>
- 751 Keeling, R. F., Körtzinger, A., & Gruber, N. (2010). Ocean deoxygenation in a warming world. *Annual*
752 *Review of Marine Science*, 2, 199–229. <https://doi.org/10.1146/annurev.marine.010908.163855>
- 753 Kroeker, K. J., Bell, L. E., Donham, E. M., Hoshijima, U., Lummis, S., Toy, J. A., & Willis-Norton, E. (2020).
754 Ecological change in dynamic environments: Accounting for temporal environmental variability in
755 studies of ocean change biology. *Global change biology*, 26(1), 54–67.
756 <https://doi.org/10.1111/gcb.14868>
- 757 Kuznetsova, A., Brockhoff, P. B., & Christensen, R. H. (2017). lmerTest package: tests in linear mixed
758 effects models. *Journal of Statistical Software*, 82(13), 1–26. <https://doi.org/10.18637/jss.v082.i13>

- 759 Lavy, A., Keren, R., Yahel, G., & Ilan, M. (2016). Intermittent hypoxia and prolonged suboxia measured
760 in situ in a marine sponge. *Frontiers in Marine Science*, 3, 263.
761 <https://doi.org/10.3389/fmars.2016.00263>
- 762 Lenth, R. (2021). *emmeans: estimated marginal means, aka least-squares means*. R package version
763 1.6.0. <https://CRAN.R-project.org/package=emmeans>
- 764 Levin, L. A. (2003). Oxygen minimum zone benthos: adaptation and community response to hypoxia.
765 *Oceanography and Marine Biology: An Annual Review*. 41, 1–45.
766 <https://doi.org/10.1201/9780203180570-3>
- 767 Levin, L. A., & Breitburg, D. L. (2015). Linking coasts and seas to address ocean deoxygenation. *Nature*
768 *Climate Change*, 5(5), 401–403. <https://doi.org/10.1038/nclimate2595>
- 769 Levin, L. A., Ekau, W., Gooday, A. J., Jorissen, F., Middelburg, J. J., Naqvi, S. W. A., Neira, C., Rabalais, N.
770 N. & Zhang, J. (2009). Effects of natural and human-induced hypoxia on coastal benthos.
771 *Biogeosciences*, 6(10), 2063–2098. <https://doi.org/10.5194/bg-6-2063-2009>
- 772 Levin, L. A., Huggett, C. L., & Wishner, K. F. (1991). Control of deep-sea benthic community structure by
773 oxygen and organic-matter gradients in the eastern Pacific Ocean. *Journal of Marine Research*, 49(4),
774 763–800. <https://doi.org/10.1357/002224091784995756>
- 775 Leys, S. P., & Kahn, A. S. (2018). Oxygen and the energetic requirements of the first multicellular
776 animals. *Integrative and Comparative Biology*, 58(4), 666–676. <https://doi.org/10.1093/icb/icy051>
- 777 Limburg, K. E., Breitburg, D., & Levin, L. A. (2017). Ocean deoxygenation-a climate-related problem.
778 *Frontiers in Ecology and the Environment*, 15(9), 479–479. <https://doi.org/10.1002/fee.1728>
- 779 Love, G.D., Grosjean, E., Stalvies, C., Fike, D.A., Grotzinger, J.P., Bradley, A.S., Kelly, A.E., Bhatia, M.,
780 Meredith, W., Snape, C.E. and Bowring, S.A., Condon, D. J., & Summons, R. E. (2009). Fossil steroids
781 record the appearance of Demospongiae during the Cryogenian period. *Nature*, 457(7230), 718–721.
782 <https://doi.org/10.1038/nature07673>

- 783 Lunden, J. J., McNicholl, C. G., Sears, C. R., Morrison, C. L., & Cordes, E. E. (2014). Acute survivorship of
784 the deep-sea coral *Lophelia pertusa* from the Gulf of Mexico under acidification, warming, and
785 deoxygenation. *Frontiers in Marine Science*, 1, 78. <https://doi.org/10.3389/fmars.2014.00078>
- 786 Maldonado, M., Ribes, M., & van Duyl, F. C. (2012). Nutrient fluxes through sponges: biology, budgets,
787 and ecological implications. *Advances in Marine Biology*, 62, 113–182. <https://doi.org/10.1016/b978-0-12-394283-8.00003-5>
- 789 Maloof, A. C., Rose, C. V., Beach, R., Samuels, B. M., Calmet, C. C., Erwin, D. H., Poirier, G. R., Yao, N., &
790 Simons, F.J. (2010). Possible animal-body fossils in pre-Marinoan limestones from South Australia.
791 *Nature Geoscience*, 3(9), 653–659. <https://doi.org/10.1038/ngeo934>
- 792 Mangum, D. C. (1980). Sea anemone neuromuscular responses in anaerobic conditions. *Science*,
793 208(4448), 1177–1178. <https://doi.org/10.1126/science.6103580>
- 794 McAllen, R., Davenport, J., Bredendieck, K., & Dunne, D. (2009). Seasonal structuring of a benthic
795 community exposed to regular hypoxic events. *Journal of Experimental Marine Biology and Ecology*,
796 368(1), 67–74. <https://doi.org/10.1016/j.jembe.2008.10.019>
- 797 McAllen, R., Taylor, A. C., & Davenport, J. (1999). The effects of temperature and oxygen partial
798 pressure on the rate of oxygen consumption of the high-shore rock pool copepod *Tigriopus brevicornis*.
799 *Comparative Biochemistry and Physiology Part A: Molecular & Integrative Physiology*, 123(2), 195–202.
800 [https://doi.org/10.1016/s1095-6433\(99\)00050-1](https://doi.org/10.1016/s1095-6433(99)00050-1)
- 801 Mentel, M., Röttger, M., Leys, S., Tielens, A. G., & Martin, W. F. (2014). Of early animals, anaerobic
802 mitochondria, and a modern sponge. *Bioessays*, 36(10), 924–932.
803 <https://doi.org/10.1002/bies.201400060>
- 804 Mills, D. B., Francis, W. R., Vargas, S., Larsen, M., Elemans, C. P., Canfield, D. E., & Wörheide, G. (2018).
805 The last common ancestor of animals lacked the HIF pathway and respired in low-oxygen
806 environments. *Elife*, 7, e31176. <https://doi.org/10.7554/eLife.31176.001>

- 807 Mills, D. B., Ward, L. M., Jones, C., Sweeten, B., Forth, M., Treusch, A. H., & Canfield, D. E. (2014).
808 Oxygen requirements of the earliest animals. *Proceedings of the National Academy of Sciences*,
809 *111*(11), 4168–4172. <https://doi.org/10.1073/pnas.1400547111>
- 810 Milton, S. L., & Prentice, H. M. (2007). Beyond anoxia: the physiology of metabolic downregulation and
811 recovery in the anoxia-tolerant turtle. *Comparative Biochemistry and Physiology Part A: Molecular and*
812 *Integrative Physiology*, *147*(2), 277–290. <https://doi.org/10.1016/j.cbpa.2006.08.041>
- 813 Moitinho-Silva, L., Steinert, G., Nielsen, S., Hardoim, C. C., Wu, Y. C., McCormack, G. P., López-Legentil,
814 S., Marchant, R., Webster, N., Thomas, T., & Hentschel, U. (2017). Predicting the HMA-LMA status in
815 marine sponges by machine learning. *Frontiers in microbiology*, *8*, 752.
816 <https://doi.org/10.3389/fmicb.2017.00752>
- 817 Morganti, T. M., Ribes, M., Moskovich, R., Weisz, J. B., Yahel, G., & Coma, R. (2021). In situ Pumping
818 Rate of 20 Marine Demosponges Is a Function of Osculum Area. *Frontiers in Marine Science*, *8*, 583188.
819 <https://doi.org/10.3389/fmars.2021.583188>
- 820 Morris, S., & Taylor, A. C. (1983). Diurnal and seasonal variation in physico-chemical conditions within
821 intertidal rock pools. *Estuarine, Coastal and Shelf Science*, *17*, 339–355. [https://doi.org/10.1016/0272-](https://doi.org/10.1016/0272-7714(83)90026-4)
822 [7714\(83\)90026-4](https://doi.org/10.1016/0272-7714(83)90026-4)
- 823 Mosch, T., Sommer, S., Dengler, M., Noffke, A., Bohlen, L., Pfannkuche, O., Liebetrau, V. & Wallmann, K.
824 (2012). Factors influencing the distribution of epibenthic megafauna across the Peruvian oxygen
825 minimum zone. *Deep Sea Research Part I: Oceanographic Research Papers*, *68*, 123–135.
826 <https://doi.org/10.1016/j.dsr.2012.04.014>
- 827 Müller, M., Mentel, M., van Hellemond, J. J., Henze, K., Woehle, C., Gould, S. B., Yu, R.Y., van der
828 Giezen, M., Tielens, A.G., Martin, W. F. (2012). Biochemistry and evolution of anaerobic energy
829 metabolism in eukaryotes. *Microbiology and Molecular Biology Reviews*, *76*(2), 444–495.
830 <https://doi.org/10.1128/MMBR.05024-11>

- 831 Murty, S. J., Bett, B. J., & Gooday, A. J. (2009). Megafaunal responses to strong oxygen gradients on the
832 Pakistan margin of the Arabian Sea. *Deep Sea Research Part II: Topical Studies in Oceanography*, 56(6-
833 7), 472–487. <https://doi.org/10.1016/j.dsr2.2008.05.029>
- 834 Nagasoe, S., Tokunaga, T., Yurimoto, T., & Matsuyama, Y. (2020). Survival and behavior patterns
835 associated with hypoxia at different life stages of the pen shell *Atrina cf. japonica*. *Aquatic Toxicology*,
836 227, 105610. <https://doi.org/10.1016/j.aquatox.2020.105610>
- 837 Nezlin, N. P., Kamer, K., Hyde, J., & Stein, E. D. (2009). Dissolved oxygen dynamics in a eutrophic
838 estuary, Upper Newport Bay, California. *Estuarine, Coastal and Shelf Science*, 82(1), 139–151.
839 <https://doi.org/10.1016/j.ecss.2009.01.004>
- 840 Nilsson, G. E., & Renshaw, G. M. (2004). Hypoxic survival strategies in two fishes: extreme anoxia
841 tolerance in the North European crucian carp and natural hypoxic preconditioning in a coral-reef shark.
842 *Journal of Experimental Biology*, 207(18), 3131–3139. <https://doi.org/10.1242/jeb.00979>
- 843 Nixon, S. W. (1995). Coastal marine eutrophication: a definition, social causes, and future concerns.
844 *Ophelia*, 41(1), 199–219. <https://doi.org/10.1080/00785236.1995.10422044>
- 845 Osinga, R., Tramper, J., & Wijffels, R. H. (1999). Cultivation of marine sponges. *Marine Biotechnology*,
846 1(6), 509–532. <https://doi.org/10.1007/PL00011807>
- 847 Piazzzi, L., Atzori, F., Cadoni, N., Cinti, M. F., Frau, F., Pansini, A., Pinna, F., Stipcich, P. & Ceccherelli, G.
848 (2021). Animal Forest Mortality: Following the Consequences of a Gorgonian Coral Loss on a
849 Mediterranean Coralligenous Assemblage. *Diversity*, 13(3), 133. <https://doi.org/10.3390/d13030133>
- 850 Pinheiro, J., Bates, D., DebRoy, S., Sarkar, D., & R Core Team (2021). *nlme: Linear and Nonlinear Mixed*
851 *Effects Models*. R package version 3.1-152. <https://CRAN.R-project.org/package=nlme>.
- 852 R Core Team (2013). *R: A language and environment for statistical computing*.
- 853 Rao, D. P., & Ganapati, P. N. (1968). Respiration as a function of oxygen concentration in intertidal
854 barnacles. *Marine Biology*, 1(4), 309–310. <https://doi.org/10.1007/BF00360781>

- 855 Raupach, M. R., & Canadell, J. G. (2010). Carbon and the Anthropocene. *Current Opinion in*
856 *Environmental Sustainability*, 2(4), 210–218. <https://doi.org/10.1016/j.cosust.2010.04.003>
- 857 Robinson, C. (2019). Microbial respiration, the engine of ocean deoxygenation. *Frontiers in Marine*
858 *Science*, 5, 533. <https://doi.org/10.3389/fmars.2018.00533>
- 859 Sampaio, E., Santos, C., Rosa, I. C., Ferreira, V., Pörtner, H. O., Duarte, C. M., Levin, L. A. & Rosa, R.
860 (2021). Impacts of hypoxic events surpass those of future ocean warming and acidification. *Nature*
861 *Ecology and Evolution*, 5(3), 311–321. <https://doi.org/10.1038/s41559-020-01370-3>
- 862 Sarà, M., Sarà, A., Nickel, M., & Brümmer, F. (2001). Three New Species of *Tethya* (Porifera:
863 Demospongiae) From German Aquaria. *Stuttgarter Beiträge zur Naturkunde* 631, 1–15.
- 864 Sassaman, C., & Mangum, C. P. (1972). Adaptations to environmental oxygen levels in infaunal and
865 epifaunal sea anemones. *The Biological Bulletin*, 143(3), 657–678. <https://doi.org/10.2307/1540189>
- 866 Schuster, A., Strehlow, B. W., Eckford-Soper, L., McAllen, R., & Canfield, D. E. (2021). Effects of Seasonal
867 Anoxia on the Microbial Community Structure in Demosponges in a Marine Lake in Lough Hyne, Ireland.
868 *Mosphere*, 6(1), e00991-20. <https://doi.org/10.1101/2020.09.09.290791>
- 869 Seibel, B. A. (2011). Critical oxygen levels and metabolic suppression in oceanic oxygen minimum
870 zones. *Journal of Experimental Biology*, 214(2), 326–336. <https://doi.org/10.1242/jeb.049171>
- 871 Semenza, G. L. (2007). Life with oxygen. *Science*, 318(5847), 62–64.
872 <https://doi.org/10.1126/science.1147949>
- 873 Smith, V. H., Joye, S. B., & Howarth, R. W. (2006). Eutrophication of freshwater and marine ecosystems.
874 *Limnology and Oceanography*, 51, 351–355. https://doi.org/10.1007/978-90-481-3385-7_1
- 875 Stachowitsch, M. (1984). Mass mortality in the Gulf of Trieste: the course of community destruction.
876 *Marine Ecology*, 5(3), 243–264. <https://doi.org/10.1111/j.1439-0485.1984.tb00124.x>
- 877 Steckbauer, A., Klein, S. G., & Duarte, C. M. (2020). Additive impacts of deoxygenation and acidification
878 threaten marine biota. *Global Change Biology*, 26(10), 5602–5612. <https://doi.org/10.1111/gcb.15252>

- 879 Stickle, W. B., Kapper, M. A., Liu, L. L., Gnaiger, E., & Wang, S. Y. (1989). Metabolic adaptations of
880 several species of crustaceans and molluscs to hypoxia: tolerance and microcalorimetric studies. *The*
881 *Biological Bulletin*, *177*(2), 303–312. <https://doi.org/10.2307/1541945>
- 882 Strano, F., Micaroni, V., Davy, S. K., Maldonado, M., & Bell, J. J. (2021). Reproduction and early life
883 stages of the poecilosclerid sponge *Crella incrustans*. *Invertebrate Biology*, e12335.
884 <https://doi.org/10.1111/ivb.12335>
- 885 Therneau, T. (2021). *A Package for Survival Analysis in R*. R package version 3.2-10. [https://CRAN.R-](https://CRAN.R-project.org/package=survival)
886 [project.org/package=survival](https://CRAN.R-project.org/package=survival).
- 887 Thuesen, E. V., Rutherford Jr, L. D., & Brommer, P. L. (2005). The role of aerobic metabolism and
888 intragel oxygen in hypoxia tolerance of three ctenophores: *Pleurobrachia bachei*, *Bolinopsis*
889 *infundibulum* and *Mnemiopsis leidyi*. *Journal of the Marine Biological Association of the United*
890 *Kingdom*, *85*(3), 627. <https://doi.org/10.1017/s0025315405011550>
- 891 Trowbridge, C. D., Davenport, J., Cottrell, D. M., Harman, L., Plowman, C. Q., Little, C., & McAllen, R.
892 (2017). Extreme oxygen dynamics in shallow water of a fully marine Irish sea lough. *Regional Studies in*
893 *Marine Science*, *11*, 9–16. <https://doi.org/10.1016/j.rsma.2017.01.008>
- 894 Turner, E. C. (2021). Possible poriferan body fossils in early Neoproterozoic microbial reefs. *Nature*,
895 *596*, 87–91. <https://doi.org/10.1038/s41586-021-03773-z>
- 896 Vacelet, J. (1975). Étude en microscopie électronique de l'association entre bactéries et spongiaires du
897 genre *Verongia* (Dictyoceratida).
- 898 Vaquer-Sunyer, R., & Duarte, C. M. (2008). Thresholds of hypoxia for marine biodiversity. *Proceedings*
899 *of the National Academy of Sciences*, *105*(40), 15452–15457.
900 <https://doi.org/10.1073/pnas.0803833105>
- 901 Vaquer-Sunyer, R., & Duarte, C. M. (2010). Sulfide exposure accelerates hypoxia-driven mortality.
902 *Limnology and Oceanography*, *55*(3), 1075–1082. <https://doi.org/10.4319/lo.2010.55.3.1075>

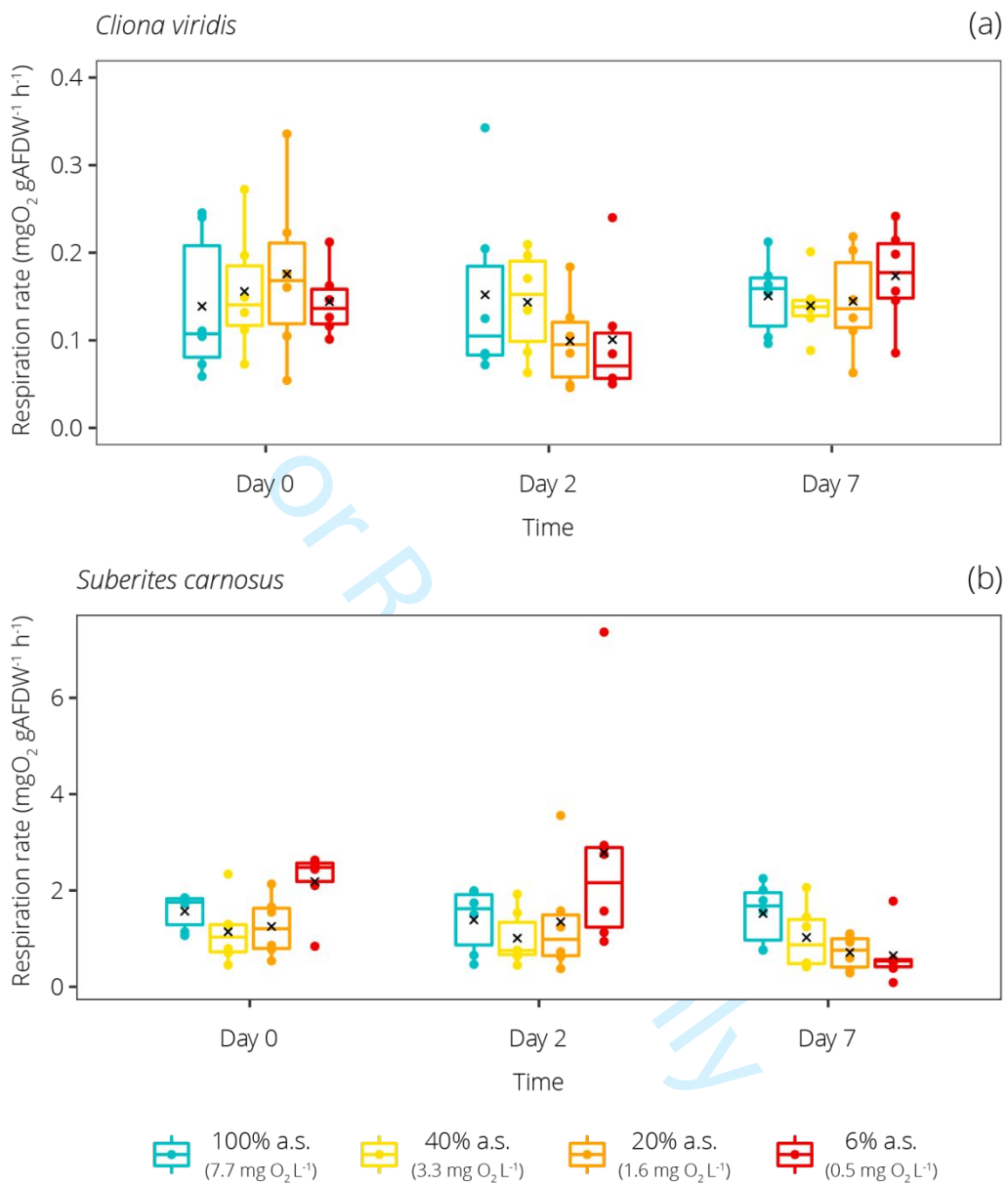
- 903 Vaquer-Sunyer, R., & Duarte, C. M. (2011). Temperature effects on oxygen thresholds for hypoxia in
904 marine benthic organisms. *Global Change Biology*, 17(5), 1788–1797. <https://doi.org/10.1111/j.1365-2486.2010.02343.x>
- 906 Vergés, A., McCosker, E., Mayer-Pinto, M., Coleman, M. A., Wernberg, T., Ainsworth, T., & Steinberg, P.
907 D. (2019). Tropicalisation of temperate reefs: implications for ecosystem functions and management
908 actions. *Functional Ecology*, 33(6), 1000–1013. <https://doi.org/10.1111/1365-2435.13310>
- 909 Vornanen, M., Stecyk, J. A., & Nilsson, G. E. (2009). The anoxia-tolerant crucian carp (*Carassius*
910 *carassius* L.). In: J. Richards, A. Farrell, & C. Brauner (Eds.), *Fish Physiology: Hypoxia, Volume 27* (pp.
911 397–441). Academic Press. [https://doi.org/10.1016/s1546-5098\(08\)00009-5](https://doi.org/10.1016/s1546-5098(08)00009-5)
- 912 Whelan, N. V., Kocot, K. M., Moroz, T. P., Mukherjee, K., Williams, P., Paulay, G., Moroz, L. L. &
913 Halanych, K. M. (2017). Ctenophore relationships and their placement as the sister group to all other
914 animals. *Nature Ecology & Evolution*, 1(11), 1737–1746. <https://doi.org/10.1038/s41559-017-0331-3>
- 915 Wishner, K. F., Ashjian, C. J., Gelfman, C., Gowing, M. M., Kann, L., Levin, L. A., Mullineaux, L. S. &
916 Saltzman, J. (1995). Pelagic and benthic ecology of the lower interface of the Eastern Tropical Pacific
917 oxygen minimum zone. *Deep Sea Research Part I: Oceanographic Research Papers*, 42(1), 93–115.
918 [https://doi.org/10.1016/0967-0637\(94\)00021-j](https://doi.org/10.1016/0967-0637(94)00021-j)
- 919 Woo, S., Denis, V., Won, H., Shin, K., Lee, G., Lee, T. K., & Yum, S. (2013). Expressions of oxidative stress-
920 related genes and antioxidant enzyme activities in *Mytilus galloprovincialis* (Bivalvia, Mollusca) exposed
921 to hypoxia. *Zoological Studies*, 52(1), 1–8. <https://doi.org/10.1186/1810-522x-52-15>
- 922 Woodhead, A. J., Hicks, C. C., Norström, A. V., Williams, G. J., & Graham, N. A. (2019). Coral reef
923 ecosystem services in the Anthropocene. *Functional Ecology*, 33(6), 1023–1034.
924 <https://doi.org/10.1111/1365-2435.13331>
- 925 Wulff, J. L. (2006). Ecological interactions of marine sponges. *Canadian Journal of Zoology*, 84(2), 146–
926 166. <https://doi.org/10.1139/z06-019>

927 Yahel, G., Marie, D., & Genin, A. (2005). InEx—a direct in situ method to measure filtration rates,
928 nutrition, and metabolism of active suspension feeders. *Limnology and Oceanography: Methods*, 3(2),
929 46–58. <https://doi.org/10.4319/lom.2005.3.46>

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For Review Only

931 FIGURES

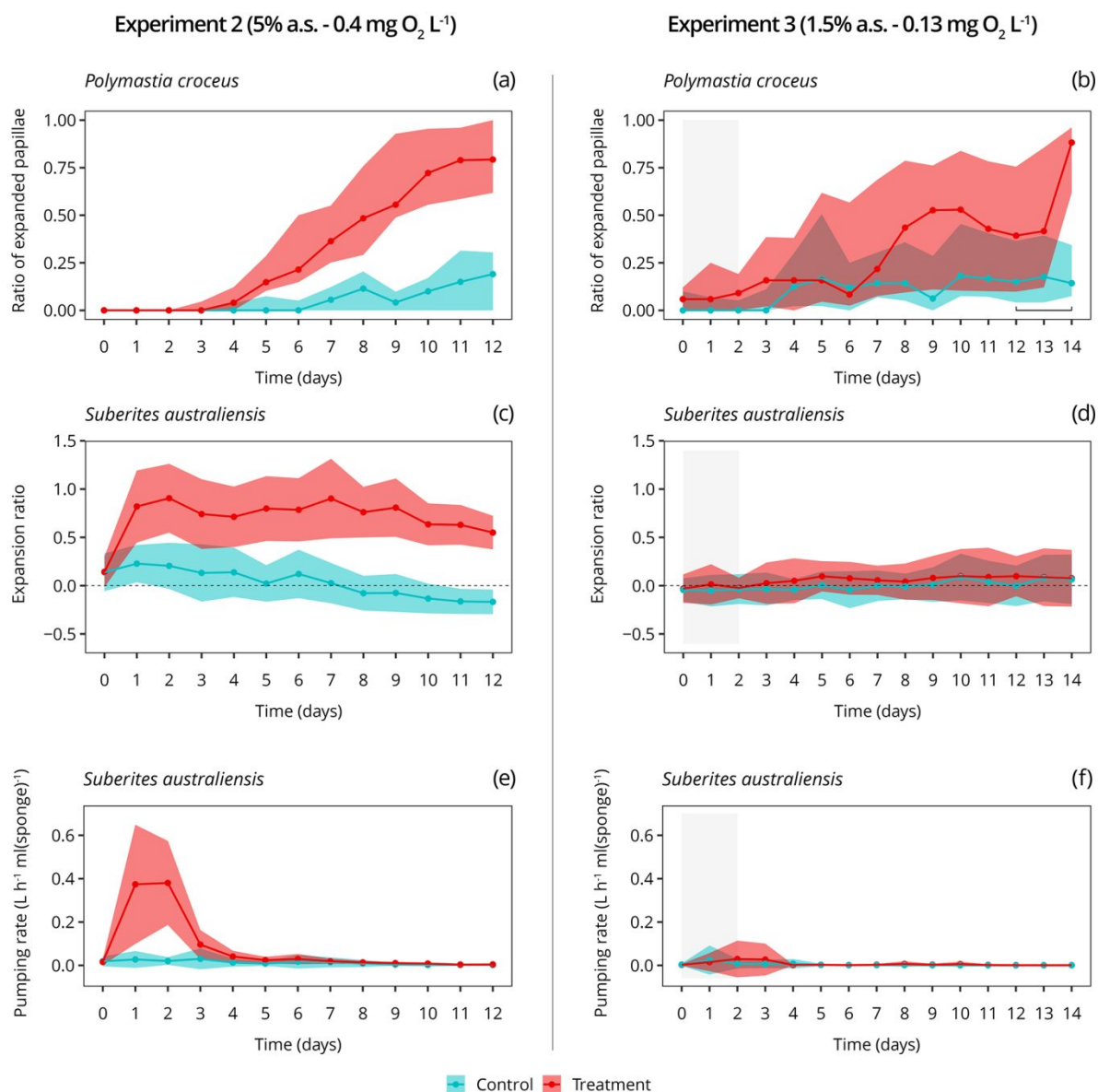


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934 Figure 1. Respiration rates in (a) *Cliona celata* and (b) *Suberites carnosus* from experiment 1 (moderate935 hypoxia) measured at T_0 , $T_{1/2}$ and $T\text{-end}$. Note: x-axis and y-axis scales differ between species. Horizontal936 bars inside the boxplots represent medians; the symbol \times represents means. Lower and upper hinges of the

937 boxplots correspond to the first and third quartiles, respectively. Points represent data points.



938

939

940 Figure 2. Contractile behaviour and pumping rate during experiments 2 and 3 (severe hypoxic conditions).

941 Changes in the ratio of expanded papillae over time in *Polymastia croceus* in each treatment in experiments942 2 (a) and 3 (b). Changes in the expansion ratio over time in *Suberites australiensis* in each treatment in

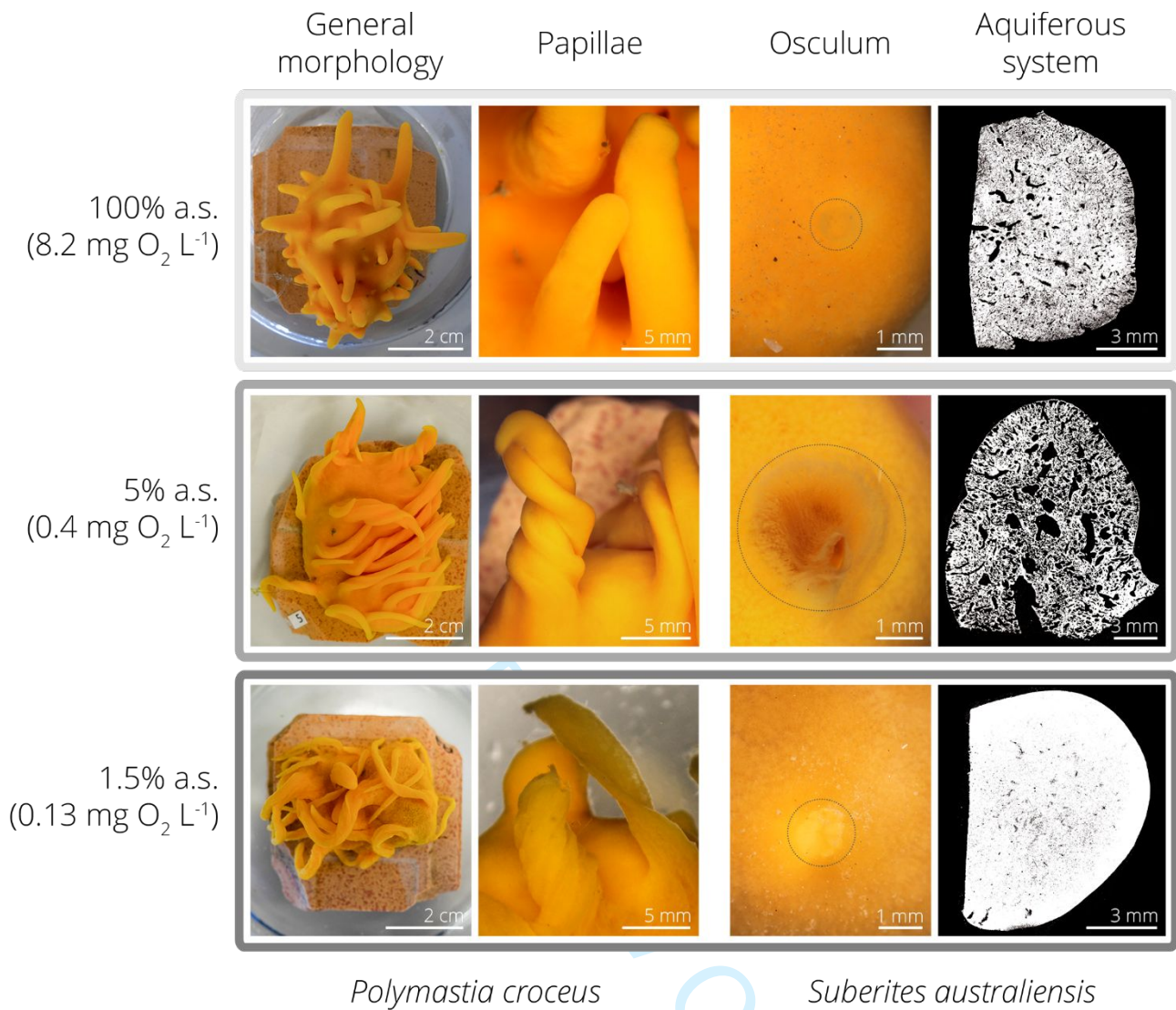
943 experiments 2 (c) and 3 (d). Changes in the pumping rate over time (estimated from the osculum cross-

944 sectional area) in *S. australiensis* in each treatment in experiments 2 (e) and 3 (f). In (a) and (b), points945 represent the median, while lower and upper edges of the ribbons represent the 75th and 25th percentile,

946 respectively. In (c), (d), (e) and (f), points represent the means while lower and upper edges of the ribbon

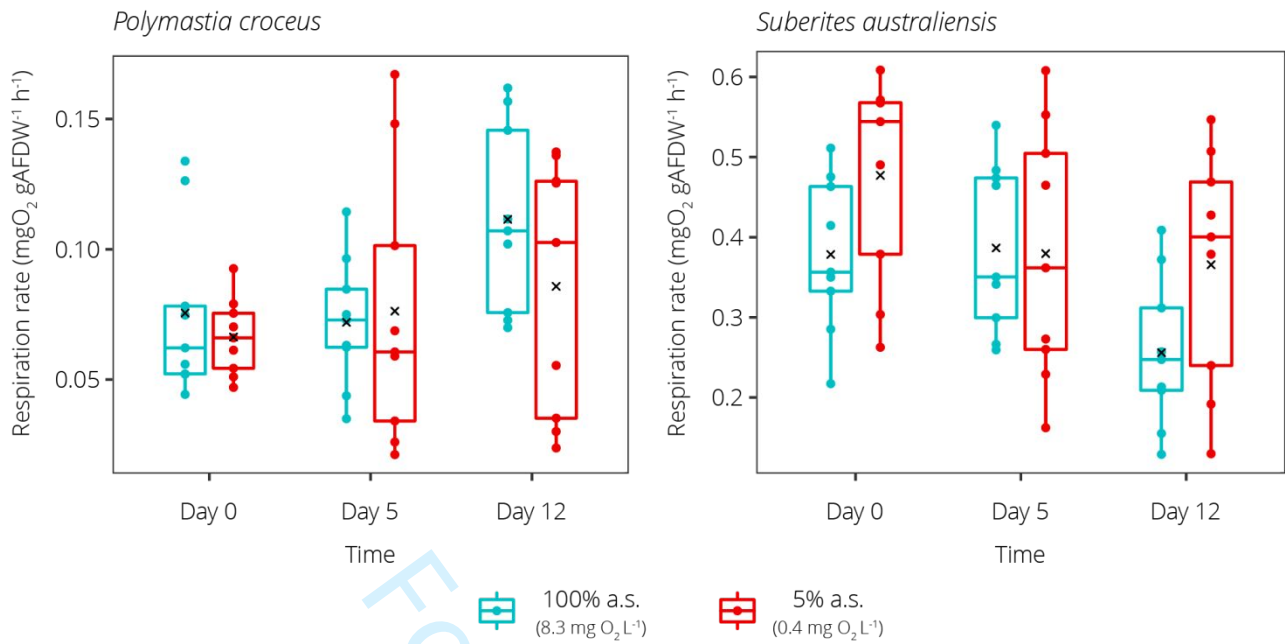
947 represent the standard deviation. Days of hypoxic acclimation (10% a.s.) are highlighted in grey. In (b), a

948 black line is used to highlight days when sponges experienced mortality.



949

950 Figure 3. Examples of the morphological modifications reported in sponges exposed to low dissolved
 951 oxygen in the severe hypoxia treatments compared to the controls. From left to right: general external
 952 morphology, and details of papillae in *Polymastia croceus*; details of the osculum (evidenced with a dotted
 953 line), and transverse histological section (sponge tissue is in white and empty spaces representing the
 954 aquiferous system are in black) in *Suberites australiensis*. An extended version of this figure is found in the
 955 supplemental material (Fig. S12).

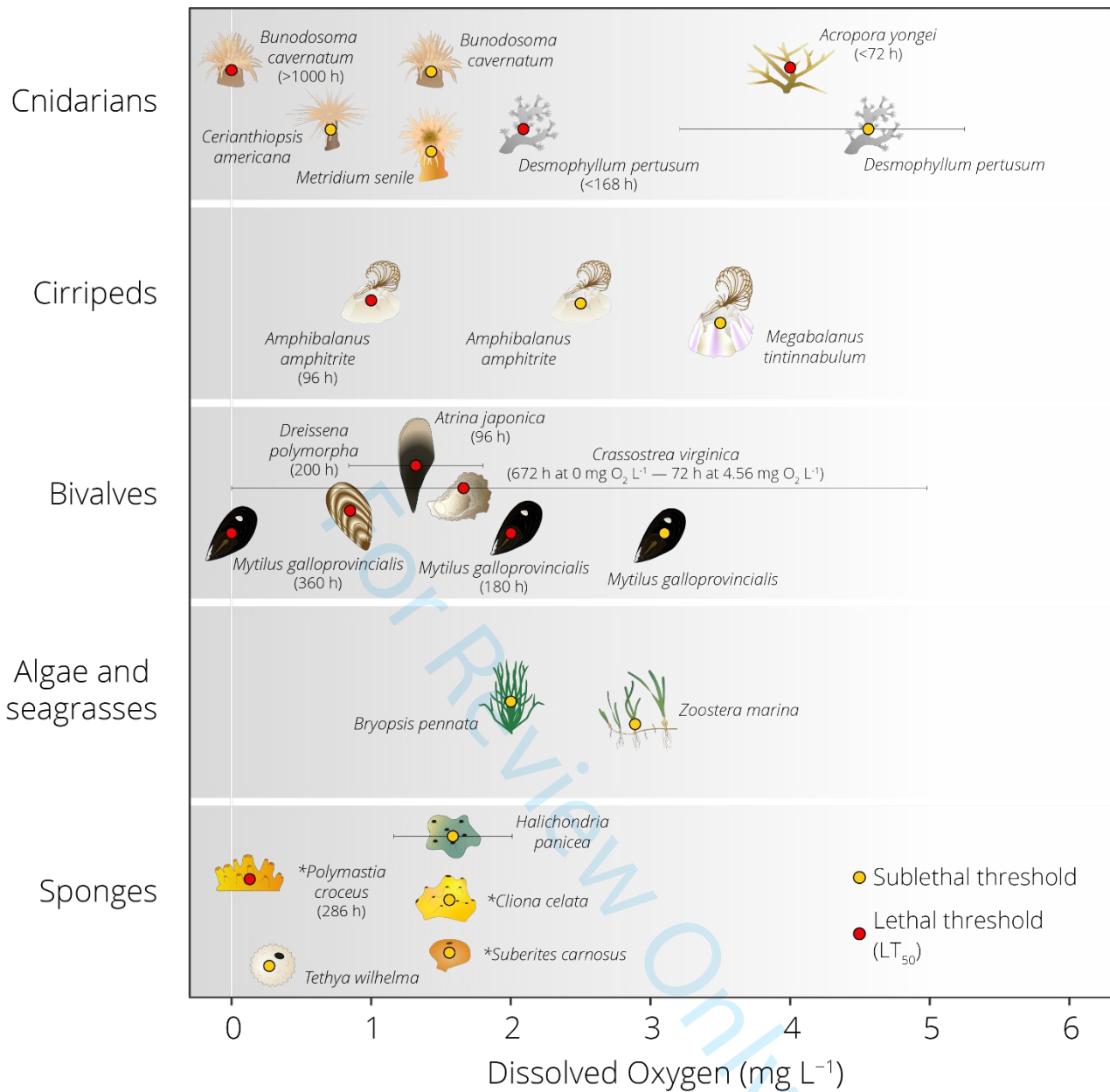


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958 Figure 4. Respiration rates in *Polymastia croceus* and *Suberites australiensis* from experiment 2 (5% a.s.)959 measured at T_0 , $T_{1/2}$ and $T\text{-end}$. Note: x-axis and y-axis scales differ between species. Horizontal bars960 inside the boxplots represent medians; the symbol \times represents means. Lower and upper hinges of the

961 boxplots correspond to the first and third quartiles, respectively. Points represent data points.

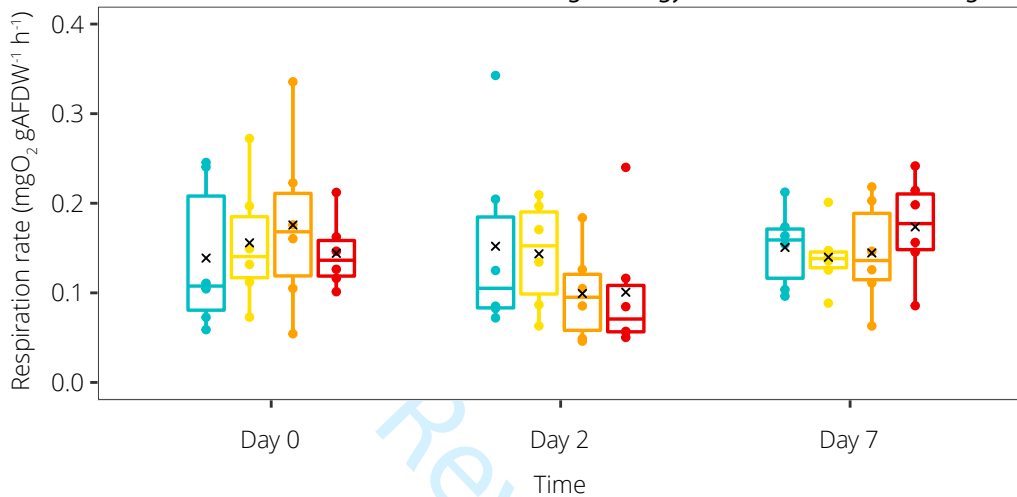


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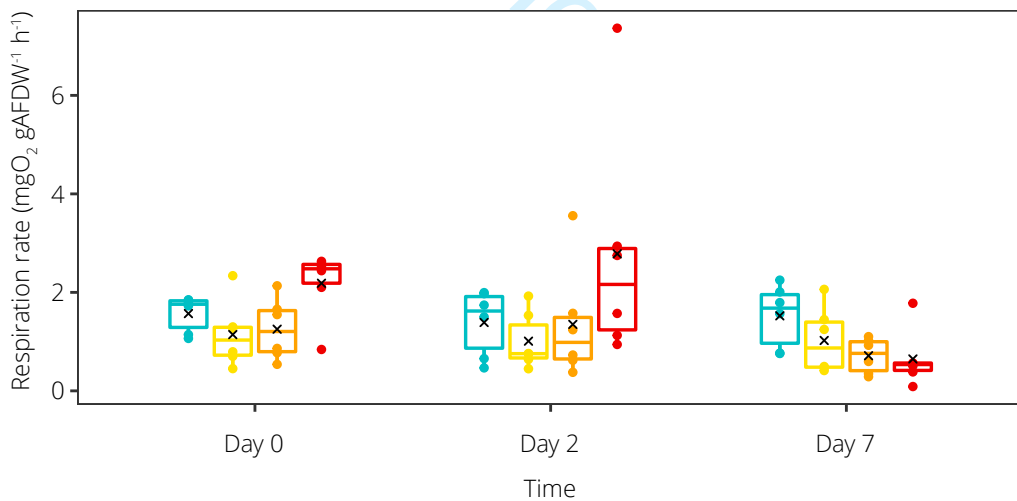
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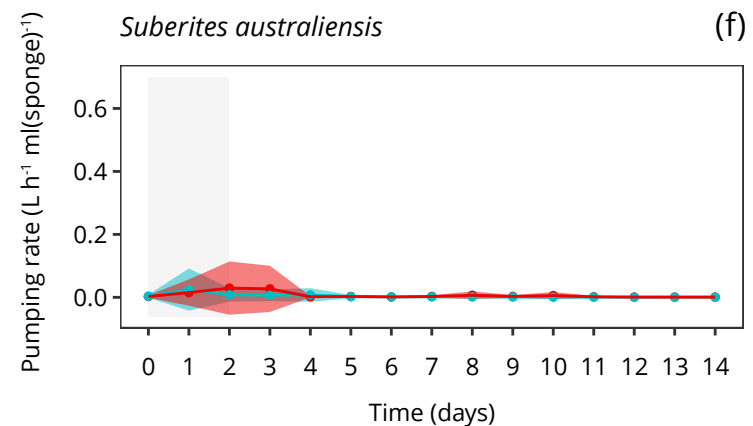
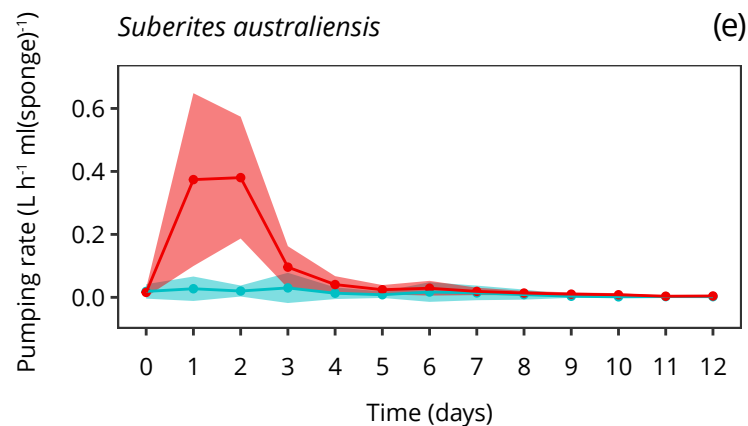
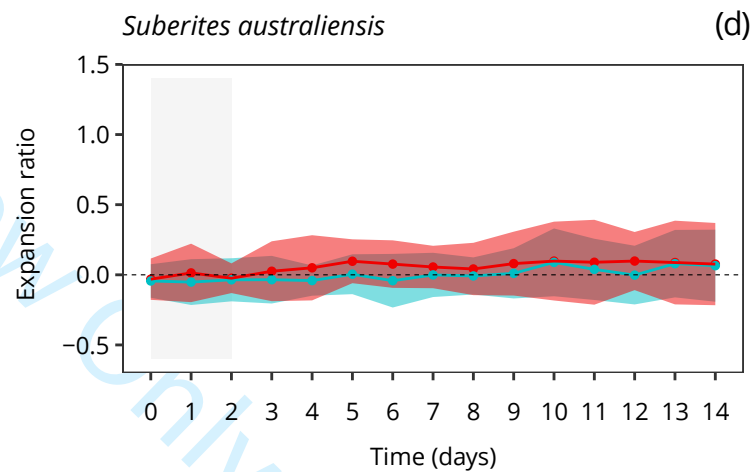
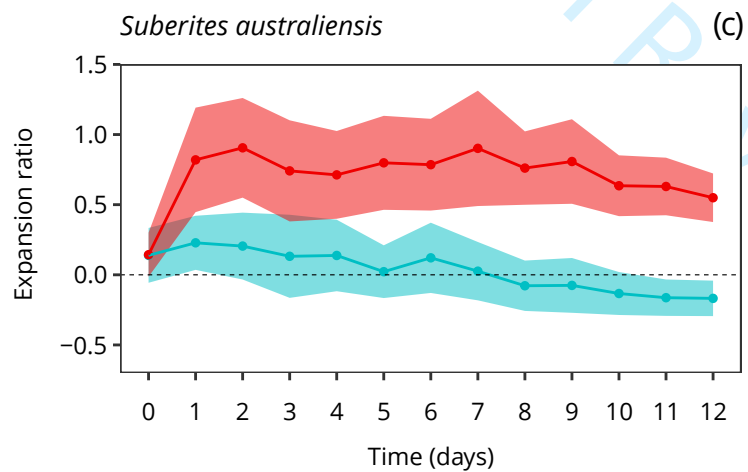
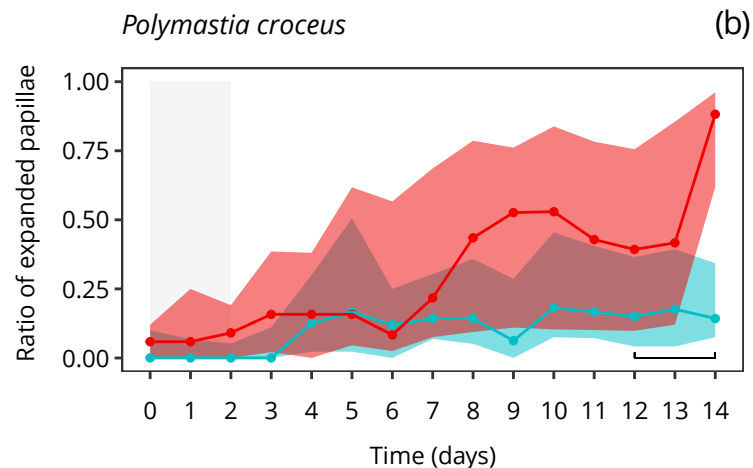
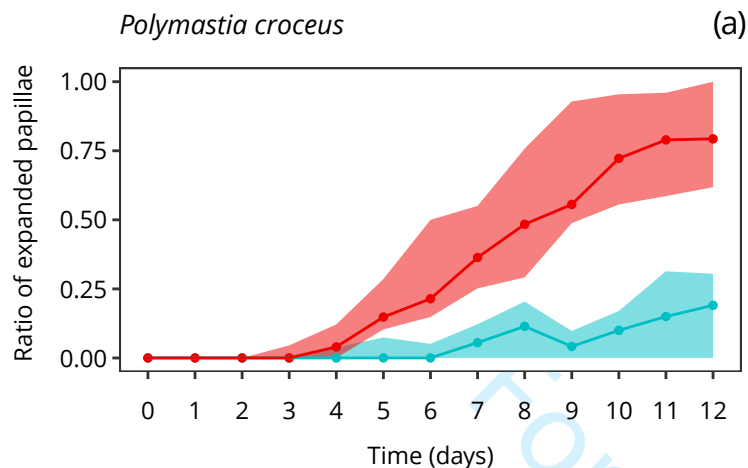
964 Figure 5. The tolerance of marine sessile organisms to hypoxia. Red dots indicate lethal thresholds, while
 965 yellow dots indicate sub-lethal thresholds. Organisms whose values were found in the present study are
 966 labelled with an asterisk. For studies that report multiple values for the same species according to other
 967 abiotic conditions (i.e., temperature and salinity), we report a range where the dots represent the mean
 968 value, and the edges of the whiskers represent minimum and maximum values. For lethal thresholds, we
 969 report in bracket the median Lethal time (LT₅₀, hours) at that specific oxygen concentration (or at the
 970 extremes of the range). The symbol < was used when LT₅₀ was not reported, but more than 50% of the
 971 organisms died after a certain amount of time; while > was used when LT₅₀ was not reached by the end of

972 the experiment. For *H. panicea*, Mills et al. (2014) only report oxygen measurement as per cent air
973 saturation without reporting temperature and salinity, so the actual oxygen concentration is unknown. We,
974 therefore, estimated the oxygen content using the range of temperatures and salinity found where these
975 sponges were sampled (Salinity 8.9–29.5; Temperature: 5–25 °C; Thomassen & Riisgård, 1995) and we
976 provide the mean and the range of possible values. List of references associated with each species: *A.*
977 *yongei* (Haas et al., 2014), *A. amphitrite* (Rao & Ganapati, 1968; Desai & Prakash, 2009), *A. japonica*
978 (Nagasoe et al., 2020), *B. pennata* (Haas et al., 2014), *B. cavernatum* (Mangum, 1980; Ellington, 1982), *C.*
979 *americana* (Vaquer-Sunyer & Duarte, 2008), *C. virginica* (Stickle et al., 1989), *D. pertusum* (Dodds et al.,
980 2007; Lunden et al., 2014), *D. polymorpha* (Johnson & McMahon, 1998), *H. panicea* (Mills et al., 2014), *M.*
981 *tintinnabulum* (Rao and Ganapati 1968), *M. senile* (Sassaman & Mangum 1972), *M. galloprovincialis* (De
982 Zwaan et al., 1991; Woo et al., 2013), *T. wilhelma* (Mills et al., 2018), *Z. marina* (Hughes et al., 2020). Figure
983 inspired by Hughes et al. (2020).

*Suberites carnosus*

(b)



Experiment 2 (5% a.s. - 0.4 mg O₂ L⁻¹)Experiment 3 (1.5% a.s. - 0.13 mg O₂ L⁻¹)

Control Treatment

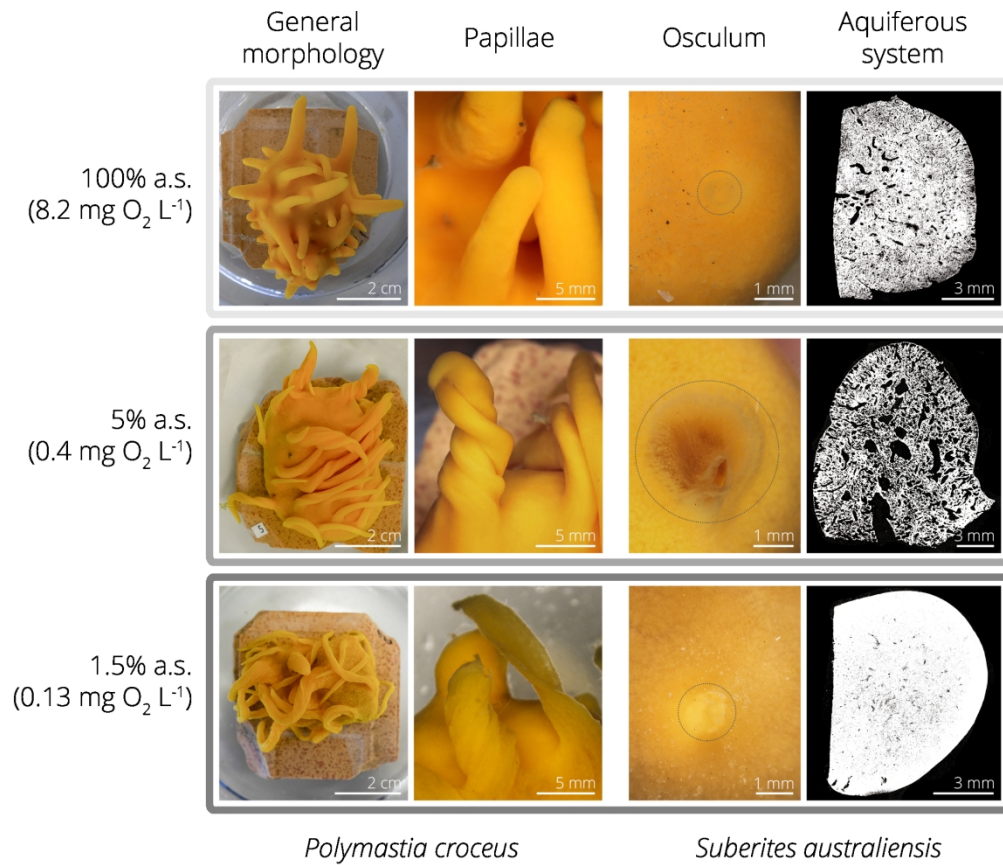
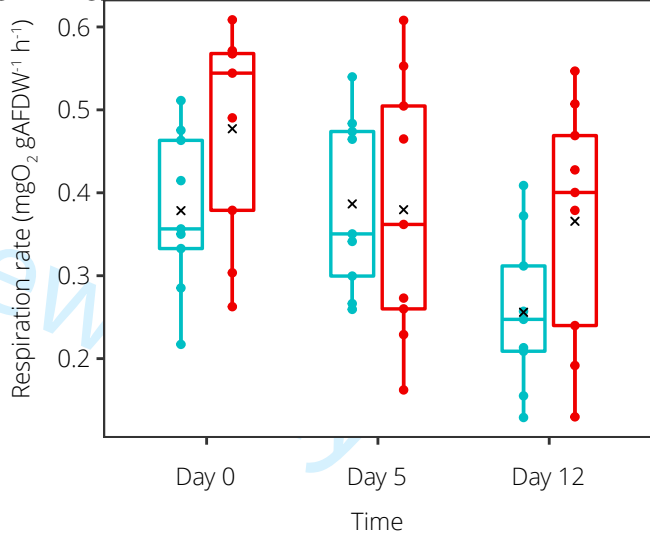
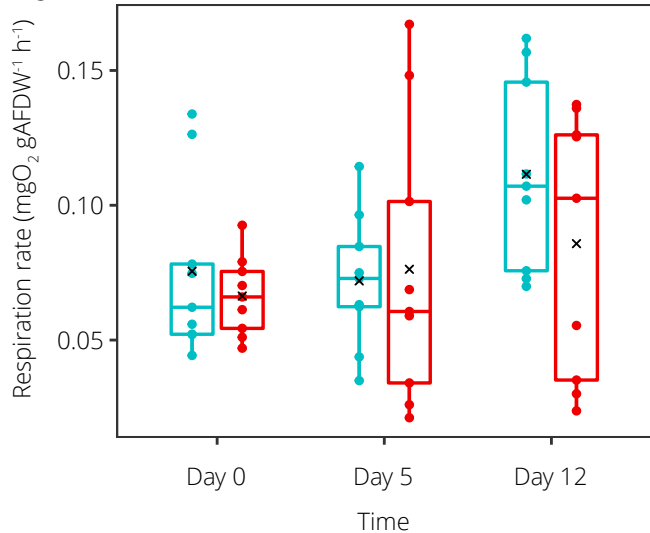
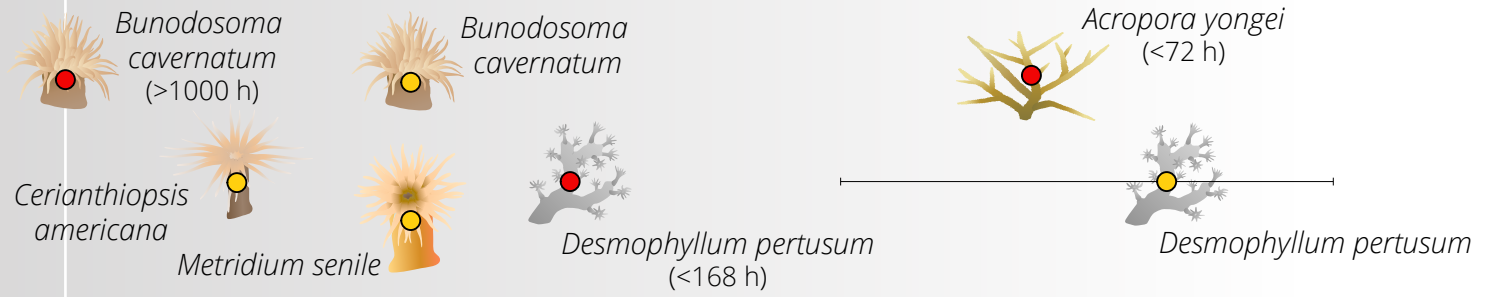


Figure 3. Examples of the morphological modifications reported in sponges exposed to low dissolved oxygen in the severe hypoxia treatments compared to the controls. From left to right: general external morphology, and details of papillae in *Polymastia croceus*; details of the osculum (evidenced with a dotted line), and transverse histological section (sponge tissue is in white and empty spaces representing the aquiferous system are in black) in *Suberites australiensis*. An extended version of this figure is found in the supplemental material (Fig. S12).



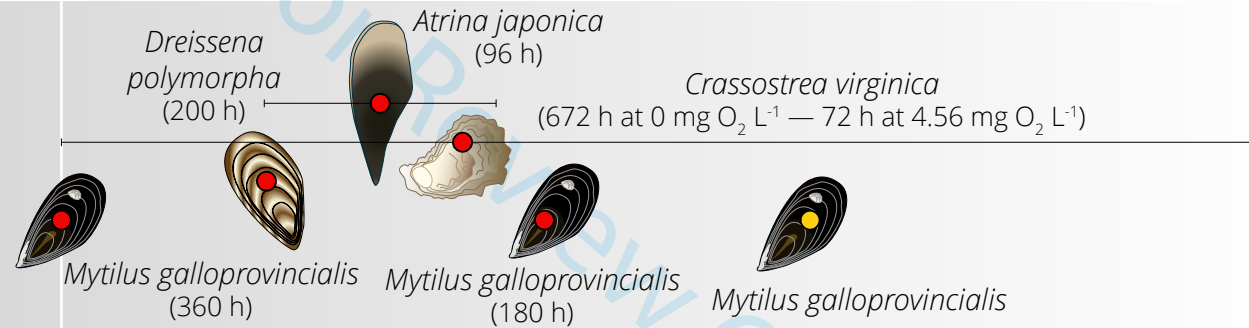
Cnidarians



Cirripeds



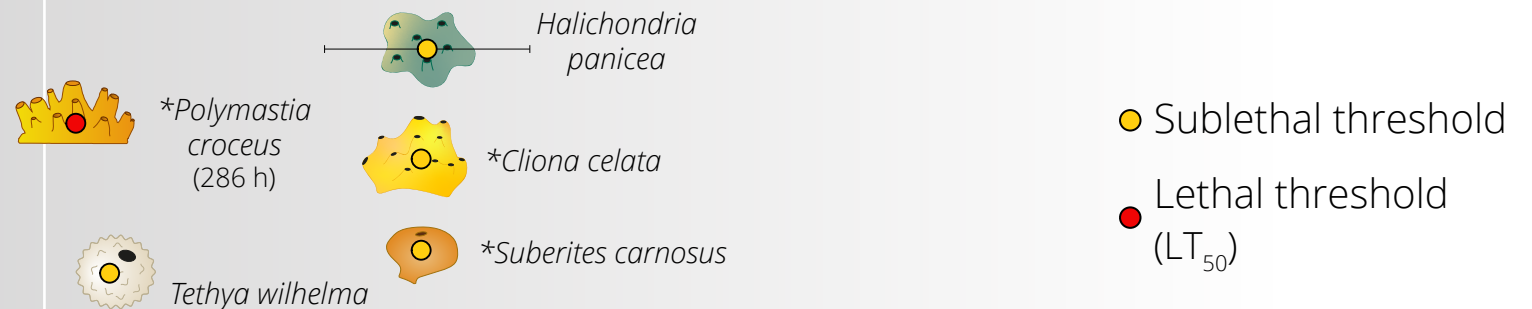
Bivalves



Algae and seagrasses



Sponges



0 1 2 3 4 5 6

Dissolved Oxygen (mg L⁻¹)