1	Biodegradability improvement of clopyralid wastes through electrolysis
2	using different diamond anodes
3	
4	
5	
6	Géssica O. S. Santos ^{a,b,c} , Katlin I. B. Eguiluz ^{a,b} , Giancarlo R. Salazar-Banda ^{a,b} , Cristina Saez ^c ,
7	Manuel A. Rodrigo ^{c,*}
8	
9	
10	
11	^a Electrochemistry and Nanotechnology Laboratory, Research and Technology Institute - ITP,
12	Aracaju, SE, Brazil
13	^b Processes Engineering Post-graduation - PEP, Universidade Tiradentes, 49037-580
14	Aracaju, SE, Brazil
15	^c Department of Chemical Engineering, Universidad de Castilla-La Mancha, Campus
16	Universitario s/n, 13071 Ciudad Real, Spain
17	
18	
19	
20	
21	
22	
23	
24	*Author to whom all correspondence should be addressed: manuel.rodrigo@uclm.es
25	
26	

27 Abstract

28 The use of boron-doped (BDDs) anodes for efficient removal of complex organic molecules, 29 such as organochlorine compounds, is well stated in the literature. However, the role of the 30 different characteristics of this anode on the transformation of these type of contaminants into 31 more biodegradable molecules is a topic of interest that need to be clarified when aimed an 32 efficient combination of an electrochemical system as a previous step to biological treatment. 33 In this work, improvement in the biodegradability of synthetic wastes polluted with 34 clopyralid, as an organochlorine model compound, is studied after electrolysis with different 35 BDDs in the presence of the two most common supporting electrolytes (containing sulfate or 36 chloride ions). For that, clopyralid removal, mineralization, aromatics intermediates, short-37 chain carboxylic acids, and inorganic ions were monitored. Improved results were found in 38 sulfate media for BDD with 200 ppm, capable of removing 88.7 % of contaminants and 85 % 39 of TOC, resulting in an improvement in biodegradability of almost 7-fold compared to the 40 initial sample. These findings point out that lower doping levels are preferable when coupling 41 studied technologies. 42 43 44 45

46 Keywords

47 Biodegradability; electrolysis; diamond anodes; combined process; clopyralid

48

49

50

52 **1. Introduction**

53 Mainly because of the massive amount of pesticides used in agriculture, the contamination of 54 water by organochlorine compounds generated from industrial and agricultural activities is 55 currently a matter of great concern [1]. These contaminants are typically highly toxic and 56 non-biodegradable [2-4]. Therefore, it is essential to find effective ways to remove them from 57 water [5-8] to protect the environment. The herbicide clopyralid (3,6-dichloro-2-pyridine-58 carboxylic acid) is a widely used organochlorinated hydrocarbon applied for the removal of 59 broadleaf weed in crops [9-12]. Due to its high solubility in water, clopyralid has been 60 frequently detected in the environment. As many other persistent organic pollutants, its 61 resistance to biodegradation makes this contaminant hard to be removed by conventional technologies [12-14]. Because of that, electrochemical oxidation with boron-doped diamond 62 63 (BDD) anodes appears as one of the most promising treatment alternatives [14-16], in 64 particular when taking into account the unique properties of BDD, which have allowed, 65 during the last two decades, its efficient application in the electrochemical oxidation of wastes polluted with many different types of complex organic molecules [17-21]. This 66 67 technology is known to generate high amounts of highly reactive and non-selective hydroxyl 68 radicals (Eq. 1) capable of attaining the complete mineralization of contaminants, that is, to 69 transform them into carbon dioxide, water and inorganics molecules [7, 22, 23].

70

71 BDD (H₂O) \rightarrow BDD($^{\bullet}$ OH) + H⁺ + e⁻ (1)

72

Moreover, if sulfates and chlorides salts are contained in the wastewater (which is a common situation in real wastes), these species can be oxidized either by direct electrontransfer on the surface of the electrode or by the electrogenerated •OH, forming more stable oxidants, which can act in the bulk solution during the treatment [22, 24]. This production of 77 oxidants helps to minimize the problems associated with mass transport limitations, typically 78 found during the electrochemical treatment of lowly concentrated wastewater. As it is known, the principal mediator species for chloride and sulfate-containing solutions are hypochlorite 79 80 and peroxosulfate, respectively [25-28]. These stable oxidants and other less stable species 81 derived from them (such as the chlorine and the sulfate radicals) are expected to be formed 82 when these salts are present in the waste, and they may affect the electrochemical oxidation 83 of clopyralid differently. Besides, the efficiency of this process may also depend on the 84 electrode properties, such as the boron doping level, substrate resistivity, and thickness of the 85 diamond layer [29, 30].

86 Regardless of the variability associated with the different characteristics of diamond 87 or the composition of the salts in the waste, the efficiency of the electrolysis with diamond 88 electrodes is typically very high. However, the harsh oxidation conditions produced in bulk 89 during the treatment are associated with large consumption of electricity, which, in turn, leads 90 to high-energy costs. The combination of electrochemical technology with cheaper biological 91 treatments is a challenge, which can help to reduce the operation cost importantly [31]. Thus, 92 the pre-treatment of wastes containing a hardly-to-degrade anthropogenic pollutant to break 93 its molecules into easier to degrade shorter molecules (mainly carboxylic acids) is one of the 94 most critical challenges faced nowadays by the environmental electrochemical technology.

Considering this background, the goal of this work is to determine the role of borondoped diamond anodes with distinct characteristics on the capability of enhancement of the biodegradability of synthetic wastes polluted with clopyralid, used as a model organochlorine pollutant. For this, electrochemical oxidation of clopyralid in the two most common supporting electrolytes (containing sulfate or chloride ions) was carried out. No real waters were used because the aim of the study is prospective, trying to determine conditions for a further successful application of the combined technology, and this goal requires a good 102 understanding of the composition of the waste. The performance of the anodes in the 103 degradation of organochlorines was analyzed through High-Performance Liquid 104 Chromatography (HPLC), chemical oxygen demand (COD), total organic carbon (TOC), ion 105 chromatography and novel rapid biodegradability assays, trying to shed light on the 106 mechanisms and electrode characteristics that promote a successful technology combination.

107

108 2. Experimental

109 2.1 Chemicals

110 Clopyralid was of analytical grade (99%), purchased from Sigma Aldrich. Sodium chloride 111 and anhydrous sodium sulfate used as supporting electrolytes were of analytical grade and 112 purchased from Panreac. Methanol and formic acid used for the mobile phase of HPLC were 113 purchased from Sigma-Aldrich. All aqueous solutions were prepared using high-purity water 114 (Millipore Milli-Q system, resistivity > 18 MΩ). All reactants were used as received.

115

116 2.2 Electrodes

117 The experiments were performed in an electrochemical cell using different compositions of 118 BDD supplied by Adamant Technologies (Neuchatel, Switzerland) and synthesized by the 119 hot filament chemical vapor deposition (CVD-HF) on monocrystalline substrate p-type Si (1 120 0 0) wafers (Siltronix).

121 The characterization of conductive-diamond lots was carried out by Adamant 122 Technologies and information about boron contents, sp^3/sp^2 ratio, the thickness of the 123 diamond layer, the roughness of the p-Si substrate and resistivity of the p-Si substrate were 124 provided in a report. From this report, the main characteristics of these anodes are 125 summarized in Table 1.

127 **Table 1.** Characteristics of different BDD anodes used in this work (provided from Adamant

Anode	Boron	Ratio	Thickness of the	Si-resistivity	Roughness, Si-	
	content	sp ³ /sp ²	BDD layer / µm	/ Ohm	Surfinra / µm	
	/ppm					
BDD8000	8000	80	1.05	10	< 0.1	
BDD2500	2500	43	1.13	10	< 0.1	
BDD1300	1300	77	2.27	10	0.3 – 0.5	
BDD200	200	75	1.14	100	0.3 – 0.5	
BDD100	100	89	1.03	10	0.3 – 0.5	

128 Technologies).

129

130 **2.2 Electrochemical measurements**

131 The electrochemical measurements were performed in a conventional three-electrode cell 132 connected to a computer connect with an Autolab potentiostat/galvanostat model PGCTAT 133 302N. The cell used an Ag/AgCl (3 M KCl) as the reference electrode and a platinum wire 134 (Pt) as the counter electrode. The working electrodes used were BDD anodes with a geometric surface area of 4 cm². All electrolysis of clopyralid wastes were performed in the 135 136 three-electrode cell using 150 mL of the electrolyte solution (synthetic wastewater) that was 137 kept in a double-jacket cell at a constant temperature of 25 °C. Electrolyzes were performed 138 in galvanostatic mode at a fixed current density of 30 mA cm². During the process, no pH 139 control was performed.

140

141 **2.3 Analytical techniques**

142 The evolution of the concentration of clopyralid and its aromatic intermediates was143 monitored by high-performance liquid chromatography (HPLC). An Agilent 1100, equipped

144 with an Eclipse Plus C-18 column (4.6 mm \times 100 mm; 3.5 µm) was used. The mobile phase 145 was a mixture of 30% methanol and 70% formic acid (0.1%) at a flow rate of 1 mL min⁻¹, 146 with the detection at 280 nm. The injection volume was 20.0 µL, the column temperature was 147 set to 20°C, and the retention time was around 2.7 min for clopyralid. Acids were identified 148 by an HPLC equipped with a Zorbax SB-Aq, column (4.6 mm \times 150 mm), the mobile phase 149 consists of 4 mM H₂SO₄ with detection at 210 nm. The retention time comparison identified 150 the intermediates with standard solutions. Inorganic ions (nitrite, nitrate, ammonium, 151 chlorate, and perchlorate) concentrations were measured by ion chromatography using a 152 Metrohm 930 Compact IC Flex coupled to a conductivity detector. The total organic carbon 153 (TOC) concentration was monitored using a Multi N/C 3100 Analytik Jena analyzer. 154 Chemical oxygen demand (COD) analyses were conducted by using a Spectroquant[®] test 155 CSB from Merck.

156

157 **2.4 Biodegradability tests**

158 Unacclimated activated sludge, was taken from a local municipal wastewater treatment plant 159 (WTTP), located in Ciudad Real, Spain, and was used for short-term biodegradability 160 assessment. Prior to the tests, the collected sludge was aerated for more than one night in the 161 absence of external substrate addition. Changes in the oxygen concentration were measured 162 with an oximetry WWR-OX4100H. For the tests, a volume of waste (5 mL) was added to 100 163 ml bioreactor filled (95 mL) of activated sludge under continuous stirrer by magnetic bars. 164 The short-term biodegradability is then estimated taking into account the primary decay related to the concentration of dissolved oxygen (DO) associated with endogenous 165 166 consumption and variation from linear to exponential decay when these external compounds are added provoking an increased consumption rate associated with the consumption of the 167 168 organics contained. Then, the primary decay is reestablished (at the new sludge 169 concentration), and a linear decay rate is kept related to the initial one. Thus, the oxygen 170 consumption associated with the substrate contained in the samples is calculated by mass 171 balance added. The obtained value correspondent to the short-term biological oxygen demand 172 (BOD_{st}), and when related to the COD, the ratio indicates the biodegradability of the sample. 173 Before the biodegradability analysis, pH values were adjusted to 7.0.

174

175 **3. Results and discussion**

Figure 1 shows the linear sweep voltammetry curves of the BDD anodes taken in a 0.5 mol L^{-1} H₂SO₄ solution. These measurements are essential to confirm that the electrochemical anodic potential interval of the diamond electrodes decreases with the increase in the boron doping level. This trend is pointed out in several works employing BDD anodes with distinct doping levels [29, 32, 33].

181 As can be seen, the oxygen evolution reaction (OER) occurs at the same potential at 182 highly-doped anodes (BDD8000 and BDD2500), which means that this property does not 183 influence the potential window at this range of boron doping. On the contrary, OER 184 overpotential shifts more significantly to values that are more positive when boron doping 185 decreases from 1300 ppm to 100 ppm. The BDD200 anode presented a larger potential 186 window than BDD100. This behavior can be explained in terms of the influence of the silicon 187 substrate that is 10-folds more resistive than for the BDD100 anode. This observation is in 188 agreement with findings reported by Moraleda et al. [23] that recently showed that the use of 189 different substrates during the diamond deposition influences the formation of hydroxyl 190 radicals. The significant difference observed points out that the resistance of the silicon 191 substrate is a more meaningful input that the boron content, at least within the range of low 192 boron doping levels studied. It is essential to consider that the broadest potential window is 193 desirable when working with BDD anodes, because it is related to the more critical

194 production of oxidants and, hence, it is expected to influence the oxidation behavior at these

anodes.

196

197



Figure 1. Linear sweep voltammetric curves performed at 10 mV s⁻¹ in 0.5 mol L⁻¹ H₂SO₄.

Figure 2 shows the clopyralid removal with different BDD anodes as a function of the 200 current charge passed for electrolysis carried out in 3.0 g L^{-1} Na₂SO₄ at a fixed current 201 density of 30 mA cm⁻². Note that clopyralid is incompletely removed during 8 h for all 202 203 anodes, which confirms the stability of this molecule. The highest removal percentage of 204 clopyralid is 88.6%, obtained using the BDD200 anode, suggesting the generation of higher 205 concentration of oxidants onto the surface of this anode that, in turn, can react with sulfate species to form more stable oxidants in the bulk solution. On the other hand, for the rest of 206 207 the anodes, the percentage of removal remained between 72.2% and 83.0% after 8 h of 208 treatment. Inset in Figure 2 shows the exponential decay of the clopyralid concentration, 209 suggesting that the clopyralid removal obeys a pseudo-first-order kinetic. This result indicates 210 that the mass transport of reactants towards the anodes limited the process, which is an 211 expected response for electrochemical diffusion-controlled systems.



Figure 2. Clopyralid removal on different BDD anodes in Na₂SO₄ (3 g L^{-1}) at a fixed current density: 30 mA cm⁻². Conditions: pH: 3.6; Clopyralid₀: 100 mg L^{-1} . Inset: Kinectic for clopyralid removal.

217

Figure 3 shows the total dissolved organic carbon (DOC) removal on the different BDD anodes. The use of different types of diamonds results in different DOC removals after the same charge passed (6.4 A h dm⁻³). In Na₂SO₄ media, the BDD200 leads to higher DOC removal of 88.7%. The inset shows the logarithmic plot, which also fits well with a pseudofirst-order kinetic model. The values of k_{DOC} and R^2 are in **Table 2**. The fastest DOC kinetics decay is observed for BDD200.



225

Figure 3. DOC removal on different BDD anodes in Na₂SO₄ (3.0 g L^{-1}) at a fixed current density of 30 mA cm⁻². Conditions: pH: 3.6; Clopyralid₀: 100 mg L^{-1} . Inset: Kinetic for DOC removal.

Table 2. Pseudo-first-order kinetics constants calculated for clopyralid and DOC removal and

the percentage of clopyralid and DOC removal.

Anode	Na ₂ SO ₄	R ²	%	Na ₂ SO ₄	\mathbb{R}^2	%
	$k_{ m Clop}$ /min ⁻¹		clopyralid	K_{DOC} / min ⁻¹		DOC
BDD8000	0.202	90.7	81.8	0.174	99.7	76.0
BDD2500	0.207	99.1	83.0	0.161	94.2	78.2
BDD1300	0.138	99.8	72.2	0.128	99.2	67.7
BDD200	0.255	99.3	88.6	0.227	99.3	85.0
BDD100	0.150	99.2	72.2	0.154	99.5	69.0

232

To evaluate the influence of the salts contained in the waste, **Figure 4** shows the removal of clopyralid with different BDD anodes, as a function of current charge passed for electrolysis carried out in an electrolyte containing 3.7 g L^{-1} NaCl at a fixed current density of 30 mA cm⁻². The inset of Figure 4 displays the semilogarithmic plot, made for easy kinetic analysis. Note that BDD200 removed 71.1% of clopyralid, while other anodes removed the herbicide in percentages that range from 55.6% to 62.0%. This outcome agrees with the observed in sulfate media, which indicates that mediated oxidation is playing an essential role in the clopyralid removal.

241



242

Figure 4. Clopyralid degradation as a function of time and current charge passed for electrolysis carried out in NaCl 3.7 g L^{-1} at a fixed current of 30 mA cm⁻².

245

The removal of the total dissolved organic carbon (DOC) (**Figure 5**) seen for sulfate media, shows that the mineralization process obeys pseudo-first-order kinetics (as indicated by straight lines seen in the inset of Figure 5). **Table 3** summarizes the values of k_{DOC} and R². The highest constant rate is observed for the BDD200 anode, also pointing out that this anode is more efficient in producing hydroxyl radicals as well as other reactive oxidative species.



Figure 5. DOC removal on different BDD anodes in NaCl (3.7 g L^{-1}) at a fixed current density of 30 mA cm⁻². Conditions: pH: 3.6; Clopyralid₀: 100 mg L^{-1} . Inset: Kinetic for DOC removal.

257

Table 3. Pseudo-first-order kinetics constants calculated for clopyralid and DOC removal andthe percentage of clopyralid and DOC removal.

Anode NaCl		R ²	% clopyralid	NaCl	R ²	% DOC
	k _{clop} / min ⁻¹			k_{DOC} / min^{-1}		
BDD8000	0.1027	99.9	55.6	0.041	99.6	25.6
BDD2500	0.1214	98.1	62.0	0.094	94.3	42.8
BDD1300	0.1112	99.2	56.7	0.061	99.3	38.1
BDD200	0.1504	98.4	71.1	0.126	98.8	67.6
BDD100	0.1032	99.1	56.4	0.074	98.8	52.4

260

Hence, in chloride and sulfate media, the resistivity of the substrate influences the results, which agrees with the voltammetric observations (**Figure 1**). Regarding the boron doping, the influence is not as evident, indicating that this parameter is not as important in the oxidation of the model herbicide tested, although in comparing the extreme values, more critical oxidation in the case of the low-doped anodes can be suggested.

266 HPLC analysis of the treated solution of clopyralid showed the formation of two primary by-products: picolinic acid and 3-picolinic acid, corresponding to the dehalogenation 267 268 of clopyralid, which may be developed on the cathode of the electrochemical cell and not on 269 the anode (cathodic hydrodehalogenation process). Figure 6 presents the evolution of these 270 reaction intermediates during the different electrolytic tests. The absence of other peaks 271 indicates that oxidation does not form stable aromatic intermediates in significant 272 concentrations, that is, that oxidation intermediates formed are rapidly oxidized. 273 Consequently, only carboxylic acids are going to be found as intermediates coming from the 274 oxidation process in the cell.

Again, as a general trend, and regardless of the supporting electrolyte, the anodes with high boron doping level favor the concentration of these dehalogenated intermediates, suggesting the more difficult oxidation of these cathodically formed intermediates with those anodes. It helps to explain the more limited mineralization rate observed for these anodes. In turn, anodes with low doping levels, do not favor the accumulation of intermediates, indicating that they are oxidized to carboxylic acids.

281



Figure 6. Evolution of byproducts concentration during the electrochemical oxidation of clopyralid on different BDD anodes in Na₂SO₄ (a and b) and NaCl (c and d) at a fixed current density: 30 mA cm⁻². Conditions: pH: 3.6; Clopyralid₀: 100 mg L⁻¹.

288 Figure 7 and Figure 8 show, respectively, the evolution of the formation of short-289 chain carboxylic acids and inorganic ions. Oxalic acid and oxamic acid were detected in 290 sulfate media (Figure 7a and 7b), while only oxalic acid was detected in chloride media 291 (Figure 7c). Typically, the concentration of these intermediates (except for the oxalic acid 292 with the BDD8000 anode) increases as the boron doping decreases, suggesting more effective 293 oxidation of the aromatic intermediates, but the more considerable difficulties in the 294 oxidation of these acid intermediates formed. Accumulation of carboxylic acids is positive 295 from the viewpoint of a further combination of technologies because, typically, these species 296 are readily biodegradable.



Figure 7. Evolution of main carboxylic acid concentration during the electrochemical oxidation of clopyralid on different BDD anodes in Na₂SO₄ (a and b) and NaCl (c) at a fixed current density: 30 mA cm⁻². Conditions: pH: 3.6; Clopyralid₀: 100 mg L⁻¹.

304 Concerning inorganic species, ion chromatography indicates the formation of 305 ammonium, chlorides, and nitrate ions (**Figure 8**). In sulfate media, it can be easily seen the 306 formation rate of chloride ions, reaching a maximum for the BDD200 anode and a minimum 307 for the BDD8000 anode (Figure 8a). The same trend is observed regarding the ammonium and nitrate ions released in sulfate media, which is consistent with previously discussed 308 309 regarding the reduction of total dissolved organic carbon (DOC). Nevertheless, for chloride 310 media, it is difficult to find differences in chloride ions, since a large amount of this salt in the 311 electrolyte makes it difficult to get any conclusion in this media. On the other hand, for 312 ammonium and nitrate, it is possible to note interesting differences among anodes. The main difference observed, again, for the BDD200 anode, which was responsible for the formation 313 314 of higher quantities of nitrate and almost no detected ammonium.





Figure 8. Evolution of inorganic ions formed during the electrochemical oxidation of clopyralid on different BDD anodes in Na₂SO₄ (a, c and e) and NaCl (b, d, and f) at a fixed current density: 30 mA cm⁻². Conditions: pH: 3.6; Clopyralid₀: 100 mg L⁻¹.

323 Samples obtained before and after the different electrolysis underwent respirometric 324 tests, consisting of monitoring the oxygen consumption evolution in an activated sludge 325 sample before and after the addition of a volume of waste. Figure 9 shows that before the 326 addition of the waste, the sludge, obtained from the biological reactor of a municipal WWTP 327 and previously aerated for more than one night in the absence of external substrate addition, 328 presents a linear decay in the concentration of oxygen. It is associated with endogenous 329 consumption, which is the consuming of dead microorganisms by the living microorganisms, 330 which value depends on the concentration of the sludge. It means that when diluted, the 331 oxygen consumption rate decreases linearly with the dilution rate. Then, by adding an 332 increased volume of waste, the decay rate reduces (because of the dilution of the 333 concentration of the microorganisms), but increases the rate associated with the consumption 334 of the organics contained. Changes in the oxygen concentration vary from linear to 335 exponential decay. After consumption of these external compounds, the endogenous rate (at the new sludge concentration) is again the primary decay, and a linear decay rate is kept 336 (whose slope is related to the initial one taking into account the dilution undergone). By mass 337

338 balance, the oxygen consumption associated with the substrate contained in the samples 339 added to the sludge can be easily calculated. This value corresponds to the short-term 340 biological oxygen demand, and when related to the COD, the ratio indicates the 341 biodegradability of the sample.

342





Figure 9. Dissolved oxygen *versus* time graphs used to calculate biodegradability of the
samples obtained before (a) and after electrolysis (b) for a representative sample. Conditions:

347 BDD200 anode; pH: 3.6; Clopyralid₀: 100 mg L⁻¹; current density: 30 mA cm⁻²; NaCl: 3.7 g 348 L⁻¹.

349

Table 4 shows the calculated short-term oxygen demand, the COD of the samples,
and the biodegradability. It also shows the total aromatic and aliphatic chromatographic area
of intermediates.

Table 4. Obtained values for COD of the samples, the total aromatic and aliphaticchromatographic area of intermediates, short-term oxygen demand, and the biodegradability.

			Total	Total		
Anode	Supporting	COD	aromatics	carboxylic	DBOst	Biodegradability
	electrolyte		HPLC	acids		(DBOst/COD)
			area	HPLC area		
BDD8000	Na ₂ SO ₄	60.0	592.0	2015.0	0.4	13.3
BDD2500	$(3.0 \text{ g } \text{L}^{-1})$	87.9	602.1	529.0	0.9	20.4
BDD1300		36.0	516.9	986.7	0.5	27.7
BDD200		28.2	393.3	2254.9	0.8	57.0
BDD100		19.7	107.4	1774.1	0.6	60.8
BDD8000	NaCl	90.0	252.1	843.2	0.5	11.1
BDD2500	(3.7 g L^{-1})	70.1	305.0	4485.9	0.8	22.8
BDD1300		80.8	205.7	3195.9	0.8	19.8
BDD200		62.6	31.50	3966.7	0.8	25.5
BDD100		51.4	133.5	3641.6	0.9	35.0

357 Figure 10 shows significant differences among anodes studied that strongly depend 358 on the doping level. For both electrolytes studied, the BDD100 and BDD200 anodes provided 359 better responses in terms of improved biodegradability, which is more than 4.5 times better 360 than that obtained by the diamond doped with 8000 ppm of boron. Comparing the 361 biodegradability of the treated wastes and the initial wastes, we see that it improved 362 significantly, as biodegradability of the initial waste was found to be as low as 8.3 %. In 363 general, in sulfate media, the biodegradability is higher than in chloride media. It can be 364 explained in terms of the negative effect of the electrochemically generated chlorine on the 365 microorganisms. Although peroxosulfates can also affect negatively to microorganisms, 366 chlorine, and derivatives lead to a more negative effect.

367





Figure 10. Biodegradability tests after treatment in Na₂SO₄ (3.0 g L^{-1}) (a) and NaCl (3.7 g L^{-1}) (b). Conditions: pH: 3.6; Clopyralid₀: 100 mg L⁻¹; current density: 30 mA cm⁻². White discontinuous line: ratio before degradation tests

369

The results obtained confirm that electrolysis with diamond anodes can be a good pretreatment for a later biological stage, although, in order to improve this combination, electrodes with low boron doping have to be used, because these electrodes, besides, to attain higher mineralization, lead to effluents with a higher biodegradability.

378

379 Conclusions

The electrochemical oxidation using boron-doped diamonds with different characteristics contributes differently to the improvement of biodegradability of the poorly biodegradable herbicide, clopyralid. Distinct boron-doped diamonds provide different efficiencies in terms of clopyralid removal and mineralization. These efficiencies, in turn, depend on the composition of the supporting electrolyte. For both supporting electrolytes (chloride or sulfate) evaluated, the anode of 200 ppm content in boron (BDD200) was the most efficient in removing and to transform clopyralid into more biodegradable compounds. The resistivity 387 of the diamond substrate seems to be a significant parameter in order to explain the results of 388 electrolysis. On the other hand, BDD with a high doping level favors the formation of 389 intermediates, while low doping levels only favor a small accumulation of intermediates, 390 which favors mineralization rate. Regardless of the electrolyte media, lower doping levels enhanced 1.8–4.5-times, for BDD100 and BDD200, respectively, 391 the solution biodegradability for the model contaminant here studied. Finally, these findings point out that 392 393 lower doping levels are preferable for improving biodegradability in order to make the sequenced process using electrolysis before biodegradation with activated sludge, 394 395 economically attractive.

396

397 Conflicts of interest

398 There are no conflicts to declare.

399

400 Acknowledgments

401 Financial support from the Agencia Estatal de Investigación through project CTM2016-

402 76197-R (AEI/FEDER, UE) is gratefully acknowledged. The authors also acknowledge the

403 financial support from the Brazilian agencies CNPq (305438/2018-2 and 310282/2013-6),

404 CAPES (88882.365552/2018-01 and 88881.187890/2018-01) and FAPITEC/SE.

405

406 **References**

- 407 [1] M.A. Rodrigo, N. Oturan, M.A. Oturan, Electrochemically Assisted Remediation of 408 Pesticides in Soils and Water: A Review, Chemical Reviews, 114 (2014) 8720-8745.
- [2] K.Y. Foo, B.H. Hameed, Detoxification of pesticide waste via activated carbon adsorption
 process, Journal of Hazardous Materials, 175 (2010) 1-11.
- 411 [3] Y.S. Ma, C.F. Sung, J.G. Lin, Degradation of carbofuran in aqueous solution by
- 412 ultrasound and Fenton processes: Effect of system parameters and kinetic study, Journal of413 Hazardous Materials, 178 (2010) 320-325.
- 414 [4] S. Randazzo, O. Scialdone, E. Brillas, I. Sires, Comparative electrochemical treatments of
- 415 two chlorinated aliphatic hydrocarbons. Time course of the main reaction by-products,
- 416 Journal of Hazardous Materials, 192 (2011) 1555-1564.

- 417 [5] R. Dewil, D. Mantzavinos, I. Poulios, M.A. Rodrigo, New perspectives for advanced 418 oxidation processes, Journal of environmental management, 195 (2017) 93-99.
- 419 [6] I. Sirés, E. Brillas, M.A. Oturan, M.A. Rodrigo, M. Panizza, Electrochemical advanced
- 420 oxidation processes: today and tomorrow. A review, Environmental Science and Pollution
- 421 Research, 21 (2014) 8336-8367.
- 422 [7] F.C. Moreira, R.A. Boaventura, E. Brillas, V.J. Vilar, Electrochemical advanced oxidation
- 423 processes: a review on their application to synthetic and real wastewaters, Applied Catalysis
- 424 B: Environmental, 202 (2017) 217-261.
- 425 [8] C.A. Martinez-Huitle, M.A. Rodrigo, I. Sires, O. Scialdone, Single and coupled 426 electrochemical processes and reactors for the abatement of organic water pollutants: a 427 critical review, Chemical reviews, 115 (2015) 13362-13407.
- 428 [9] M.B. Carboneras, P. Cañizares, M.A. Rodrigo, J. Villaseñor, F.J. Fernandez-Morales,
 429 Improving biodegradability of soil washing effluents using anodic oxidation, Bioresource
 430 technology, 252 (2018) 1-6.
- 431 [10] C. Berberidou, V. Kitsiou, D.A. Lambropoulou, A. Antoniadis, E. Ntonou, G.C. Zalidis,
- 432 I. Poulios, Evaluation of an alternative method for wastewater treatment containing pesticides
- 433 using solar photocatalytic oxidation and constructed wetlands, Journal of environmental
- 434 management, 195 (2017) 133-139.
- [11] A. Özcan, N. Oturan, Y. Şahin, M.A. Oturan, Electro-Fenton treatment of aqueous
 clopyralid solutions, International Journal of Environmental and Analytical Chemistry, 90
- 437 (2010) 478-486.
- [12] D.V. Šojić, V.B. Anderluh, D.Z. Orčić, B.F. Abramović, Photodegradation of clopyralid
 in TiO2 suspensions: Identification of intermediates and reaction pathways, Journal of
 hazardous materials, 168 (2009) 94-101.
- 441 [13] B. Carboneras, J. Villaseñor, F.J. Fernandez-Morales, Modelling aerobic biodegradation
- of atrazine and 2, 4-dichlorophenoxy acetic acid by mixed-cultures, Bioresource technology,
 243 (2017) 1044-1050.
- 444 [14] M. Muñoz-Morales, C. Sáez, P. Cañizares, M. Rodrigo, Enhanced electrolytic treatment 445 for the removal of clopyralid and lindane, Chemosphere, (2019).
- 446 [15] E. Dos Santos, C. Sáez, P. Cañizares, C. Martínez-Huitle, M. Rodrigo, UV assisted
 447 electrochemical technologies for the removal of oxyfluorfen from soil washing wastes,
 448 Chemical Engineering Journal, 318 (2017) 2-9.
- 449 [16] E.V. Dos Santos, S.F.M. Sena, D.R. da Silva, S. Ferro, A. De Battisti, C.A. Martínez-
- 450 Huitle, Scale-up of electrochemical oxidation system for treatment of produced water 451 generated by Brazilian petrochemical industry, Environmental Science and Pollution
- 452 Research, 21 (2014) 8466-8475.
- 453 [17] S. Cotillas, M.J.M. de Vidales, J. Llanos, C. Sáez, P. Cañizares, M.A. Rodrigo,
- 454 Electrolytic and electro-irradiated processes with diamond anodes for the oxidation of 455 persistent pollutants and disinfection of urban treated wastewater, Journal of hazardous 456 materials, 319 (2016) 93-101.
- [18] D.A. Coledam, M.M. Pupo, B.F. Silva, A.J. Silva, K.I. Eguiluz, G.R. Salazar-Banda,
 J.M. Aquino, Electrochemical mineralization of cephalexin using a conductive diamond
 anode: A mechanistic and toxicity investigation, Chemosphere, 168 (2017) 638-647.
- 460 [19] M. Panizza, G. Cerisola, Direct And Mediated Anodic Oxidation of Organic Pollutants, 461 Chemical Reviews, 109 (2009) 6541-6569.
- 462 [20] A.M. Polcaro, M. Mascia, S. Palmas, A. Vacca, Electrochemical degradation of diuron
- 463 and dichloroaniline at BDD electrode, Electrochimica Acta, 49 (2004) 649-656.
- 464 [21] F. de Souza, C. Sáez, P. Cañizares, A. de Motheo, M. Rodrigo, Electrochemical removal
- 465 of dimethyl phthalate with diamond anodes, Journal of Chemical Technology &
- 466 Biotechnology, 89 (2014) 282-289.

- 467 [22] B. Marselli, J. Garcia-Gomez, P.-A. Michaud, M. Rodrigo, C. Comninellis,
 468 Electrogeneration of hydroxyl radicals on boron-doped diamond electrodes, Journal of the
 469 Electrochemical Society, 150 (2003) D79-D83.
- 470 [23] E. Brillas, C.A. Martínez-Huitle, Decontamination of wastewaters containing synthetic
- 471 organic dyes by electrochemical methods. An updated review, Applied Catalysis B:
 472 Environmental, 166 (2015) 603-643.
- 473 [24] M. Panizza, G. Cerisola, Application of diamond electrodes to electrochemical
 474 processes, Electrochimica Acta, 51 (2005) 191-199.
- 475 [25] S.O. Ganiyu, C.A. Martínez-Huitle, Nature, Mechanisms and Reactivity of 476 Electrogenerated Reactive Species at Thin-Film Boron-Doped Diamond (BDD) Electrodes
- 477 During Electrochemical Wastewater Treatment, ChemElectroChem, 6 (2019) 2379-2392.
- 478 [26] P. Cañizares, C. Sáez, A. Sánchez-Carretero, M. Rodrigo, Influence of the characteristics
- 479 of p-Si BDD anodes on the efficiency of peroxodiphosphate electrosynthesis process,
 480 Electrochemistry Communications, 10 (2008) 602-606.
- 481 [27] K. Serrano, P.A. Michaud, C. Comninellis, A. Savall, Electrochemical preparation of
- 482 peroxodisulfuric acid using boron doped diamond thin film electrodes, Electrochimica Acta,
 483 48 (2002) 431-436.
- 484 [28] K. Groenen-Serrano, E. Weiss-Hortala, A. Savall, P. Spiteri, Role of Hydroxyl Radicals
- 485 During the Competitive Electrooxidation of Organic Compounds on a Boron-Doped 486 Diamond Anode, Electrocatalysis, 4 (2013) 346-352.
- [29] E. Guinea, F. Centellas, E. Brillas, P. Cañizares, C. Sáez, M.A. Rodrigo, Electrocatalytic
 properties of diamond in the oxidation of a persistant pollutant, Applied Catalysis B:
 Environmental, 89 (2009) 645-650.
- 490 [30] C. Carvalho de Almeida, M. Munoz-Morales, C. Saez, P. Canizares, C.A. Martinez-
- 491 Huitle, M.A. Rodrigo, Electrolysis with diamond anodes of the effluents of a combined soil
- 492 washing ZVI dechlorination process, Journal of Hazardous Materials, 369 (2019) 577-583.
- 493 [31] J. Radjenovic, D.L. Sedlak, Challenges and opportunities for electrochemical processes
- 494 as next-generation technologies for the treatment of contaminated water, Environmental
 495 Science & Technology, 49 (2015) 11292-11302.
- 496 [32] K. Schwarzová-Pecková, J. Vosáhlová, J. Barek, I. Šloufová, E. Pavlova, V. Petrák, J.
- 497 Zavázalová, Influence of boron content on the morphological, spectral, and electroanalytical
- characteristics of anodically oxidized boron-doped diamond electrodes, Electrochimica Acta,
 243 (2017) 170-182.
- 500 [33] R. Šelešovská, B. Kränková, M. Štěpánková, P. Martinková, L. Janíková, J. Chýlková,
- 501 M. Vojs, Influence of boron content on electrochemical properties of boron-doped diamond
- 502 electrodes and their utilization for leucovorin determination, Journal of Electroanalytical
- 503 Chemistry, 821 (2018) 2-9.
- 504
- 505