

1 **Fast pyrolysis of agroindustrial wastes blends: hydrocarbon production**  
2 **enhancement**

3 A. Alcazar-Ruiz, F. Dorado, L. Sanchez- Silva\*

4 Department of Chemical Engineering, University of Castilla –La Mancha,

5 Avda. Camilo José Cela 12, 13071 Ciudad Real, Spain

6 \*Corresponding author phone: +34 926 29 53 00 ext. 6307; fax: +34 926 29 52 56;

7 e-mail: [marialuz.sanchez@uclm.es](mailto:marialuz.sanchez@uclm.es)

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22 **Abstract**

23 Fast pyrolysis of waste from agroindustry may be an alternative choice for sustainable  
24 use of enhanced biofuels. Plastics are one option for improving the hydrogen to carbon  
25 efficiency ratio ( $H/C_{\text{eff}}$ ) of biomass feedstock. Waste from agroindustry in blends with  
26 biomass could modify the reaction mechanism for removing oxygen by substituting  
27 decarbonylation and decarboxylation with dehydration. Firstly, fast pyrolysis was  
28 performed to find the optimal mass blending ratio for olive pomace (OP) and  
29 agroindustrial polymers (polyethylene (PE), polystyrenes (PS) and polyvinyl chloride  
30 (PVC)) according to hydrocarbon production. Experimental results for the 1.5:1 OP/PE,  
31 1:1.5 OP/PS and 1:1.5 OP/PVC mass blending ratios at 500 °C, showed synergistic  
32 enhancement of hydrocarbon yields. Alkenes yield were enhanced for 1.5:1 OP/PE,  
33 where the light hydrocarbons fraction ( $C_6-C_{10}$ ) first increased and then decreased with  
34 temperature, reaching a maximum at 650 °C. For 1:1.5 OP/PS and 1:1.5 OP/PVC, it was  
35 improved the aromatic compounds formation, being 500 °C and 650 °C the optimal  
36 reaction temperature for the former and the later, respectively. Benzene, toluene and  
37 xylene were in large quantities obtained for PS and PVC blends with OP. Additionally,  
38 the synergistic effect on pyrolysis of the blends did not show any clear trend for pyrolytic  
39 gas emissions. In general, as reaction temperature increased, CO and CO<sub>2</sub> emissions fell  
40 and CH<sub>4</sub> was enhanced. Finally, olefin and aromatic yields were promoted for blends with  
41 a higher  $H/C_{\text{eff}}$ .

42 **Keywords:** Fast pyrolysis; Olive pomace; Agroindustrial polymers; Hydrocarbon  
43 production; Synergistic effect;  $H/C_{\text{eff}}$

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## 46        **1. Introduction**

47    Growing concerns with the imminent depletion of fossil fuels and their adverse  
48    environmental effects have led to a focus on using renewable lignocellulosic biomass as  
49    an alternative energy resource. Olive oil production and related industries are of great  
50    importance in terms of wealth, health and tradition in the Mediterranean area [1].  
51    However, this industry generates olive pomace in high amounts as waste from  
52    agroindustry and there have been moves to valorise this. Dried olive pomace contains a  
53    high amount of organic matter, water-soluble fats, proteins, water-soluble carbohydrates  
54    and water-soluble phenolic substances [2]. Therefore, this olive waste could be used as a  
55    sustainable biomass feedstock for creating valuable products, thereby reducing the impact  
56    the olive oil industry has. In addition, the hydrogen to carbon effective ( $H/C_{\text{eff}}$ ) ratio plays  
57    a significant role in the conversion efficiency of biomass into valuable products. Thus, it  
58    is logical to add high hydrogen yield co-reactants to lignocellulosic biomass, hence  
59    modifying the reaction mechanism of oxygen removal by substituting decarbonylation  
60    and decarboxylation with dehydration, which enhances the hydrogen deficient biomass  
61    [3].

62    Plastic waste is seen as one option for improving the hydrogen to carbon efficiency of  
63    sustainable biomass feedstock due to its hydrogen-rich nature [4]. Furthermore, plastic  
64    materials have similar properties to those in fossil fuels in terms of heating value and lack  
65    of oxygenated compounds. Currently, the most widespread plastics used are polyethylene  
66    (PE), polypropylene (PP), polystyrenes (PS), polyvinyl chloride (PVC) and polyethylene  
67    terephthalate (PET). For over 50 years, plastics have been made for improving our  
68    standard of living [5]. Consequently, annual global consumption of plastics has surpassed  
69    300 million tonnes [6]. Due to its high consumption, managing this waste material is of  
70    great social and environmental concern. At present, recycling these residues is a growth

71 area. In Europe, only 32.5% of plastic waste generated in 2019 was recycled, while the  
72 rest was landfilled or used in energy recovery processes [7]. To be specific, plastic waste  
73 from agroindustry represents 2% of total plastic waste generated [8].

74 To reduce such wastes, different thermochemical methods such as pyrolysis have been  
75 promoted as one solution. Pyrolysis of plastic waste in an inert atmosphere is regarded as  
76 one of the most viable recycling methods for obtaining valuable products [9]. Three  
77 pyrolysis processes can be differentiated depending on operating conditions, each of  
78 which is aimed at producing desirable products. Fast pyrolysis leads to high bio-oil  
79 production (approximately 75 wt.%) and occurs at high heating rates, moderate  
80 temperatures and short residence time [4,10,11]. Operating conditions for fast pyrolysis  
81 can easily be adjusted to maximise bio-oil production and the quality of the products  
82 obtained [12]. Furthermore, in this process plastic could act as a source of hydrogen  
83 atoms, producing hydrocarbons with a similar composition found in conventional  
84 gasoline [13]. Therefore, co-pyrolysis could be a promising solution since it can easily  
85 process waste from biomass and plastics, tackling different challenges simultaneously  
86 and converting them into high value fuels or chemicals [14]. In co-pyrolysis, high oil  
87 yields are achieved at temperatures from 400 to 600 °C, and the optimum temperature for  
88 maximum oil yield depends on the proportions and characteristics of the feedstock [4].  
89 Moreover, co-pyrolysis of woody biomass and plastic waste produces high yields in  
90 terms of forming hydrocarbons while oxygenated compounds such as ketones, aldehydes  
91 and acids are suppressed [15]. However, any improvements in the yield and quality of  
92 bio-oil is attributed to various synergistic interactions between the feedstocks [4,16,17].

93 Research data on biomass and plastic pyrolysis usually focus on bio-oil properties and  
94 yield, rather than the synergistic mechanism. The synergy effect between biomass and  
95 polymers in co-pyrolysis is the main factor behind bio-oil quality and any improved

96 quantity [3]. It usually occurs when the combined effects of the components are greater  
97 than the sum of their individual ones. Ephraim et al. analysed the synergistic effect and  
98 product yield for various plastic materials and saw that product composition depends on  
99 feedstock [18]. Li et al. studied the maximum content of hydrocarbons from catalytic fast  
100 co-pyrolysis of rice husk and plastic films waste from greenhouses, at a maximum  
101 temperature of 600 °C with a mass ratio of 1:1.5 [19]. Mixtures of biomass sawdust and  
102 waste polyolefins from packaging for dairy products revealed maximum hydrocarbon  
103 production were achieved with 75% of PS blends [13]. In general, the literature has  
104 established a positive synergy between biomass and plastics, which results in enhanced  
105 aromatic hydrocarbon yields due to the higher calorific values and  $H/C_{\text{eff}}$  ratio in plastics  
106 compared to biomass [20,21]. Nevertheless, there are no studies that focus on the  
107 synergistic enhancement of olive pomace in blends with plastic wastes in which pyrolysis  
108 product distribution is valorised.

109 In this study, olive pomace is co-pyrolysed with waste polymers (PE, PS and PVC) to  
110 valorise these agroindustrial subproducts and to research the influence of plastic content  
111 on bio-oil properties. Furthermore, olive pomace and the polymers mass ratio (OP/P ratio)  
112 and the effect of reaction temperature were studied to evaluate the synergies, hydrocarbon  
113 composition in bio-oils, and the hydrogen to carbon effective ratios derived from fast  
114 pyrolysis mixtures.

## 115 **2. Materials and methods**

### 116 **2.1. Materials**

117 Olive pomace (OP) was supplied by Aceites García de la Cruz olive oil mill (Madrirdejos,  
118 Castilla-La Mancha, Spain). Prior to the experiments, OP was first oven-dried for 24 h at  
119 100 °C and milled and sieved to obtain an average particle size ranging from 100 to 150

120  $\mu\text{m}$ . The polymers used (PE, PS and PVC) were purchased from Sigma-Aldrich, US. The  
121 plastics received were ground and sieved to a particle size of under 0.1 mm.

## 122 2.2. Sample characterisation

123 Materials were first characterized by elemental and thermogravimetric analysis in a TGA  
124 apparatus (TGA-DSC 1, Mettler Toledo). The ultimate and proximate analysis were  
125 carried out according to standards UNE 15104:2011, UNE-EN ISO18123:2016, UNE 32-  
126 004-84 and UNE 32-002-95, in a Thermo Fischer Scientific Flash 2000 elemental  
127 analyser, equipped with a thermal conductivity detector. The proximate analysis provided  
128 data on moisture, ash, volatile matter and fixed carbon content. Also, the ultimate analysis  
129 was used to find the concentration of carbon, hydrogen, nitrogen, oxygen and sulphur in  
130 the sample. The higher heating value (HHV) can be calculated using the ultimate analysis  
131 data with the following empirical correlation (eq. 1, [22]; Table 1):

$$132 \quad HHV \text{ (MJ/kg)} = 0.3491 \times C + 1.1783 \times H + 0.1005 \times S - 0.1034 \times O - 0.0151 \times N - \\ 133 \quad 0.0211 \times A \text{ (eq. 1)}$$

134 In which C, H, S, O and N are the weight percentages of carbon, hydrogen, sulphur,  
135 oxygen and nitrogen respectively, whereas A is the weight percentage of ash.

136 In addition, metal content was determined by Inductively Coupled Plasma-Optical  
137 Emission Spectrometry (ICP-OES) with Varian 720-ES equipment (which was  
138 previously calibrated using standard stock solutions). The ratios of extractives,  
139 hemicellulose and Klason lignin in OP were tested according to the TAPPI 204 om-97  
140 and TAPPI T222 om-02 method [23,24] whose results are shown in Table 1.

141 The chemical structure of the obtained residual carbon after fast pyrolysis was  
142 characterized by: Resolution Scanning Electron Microscope (HRSEM), carried out using  
143 a Gemini SEM 500 High from ZEISS brand (Oberkochen, Germany); Energy Dispersive

144 X-Ray Analyses (EDX), carried out in an Oxford brand instrument at 15.00 kV  
145 accelerated voltage and 127x magnifications; and finally, FTIR spectroscopy, performed  
146 with a Perkin-Elmer FTIR Spectrum-two spectrophotometer provided with a Universal  
147 Attenuated Total Reflectance accessory (UATR). The spectra accumulated 64 scans with  
148 a range between 500 and 4000  $\text{cm}^{-1}$  and a resolution of 8  $\text{cm}^{-1}$ .

### 149 2.3. Fast co-pyrolysis experiments, conditions and procedure analysis

150 The co-pyrolysis experiments were performed with OP and PE, PS and PVC via Py-  
151 GC/MS-FGA (CDS Pyroprobe 6200 pyrolyser; Agilent Technologies 7890B/5977B  
152 GC/MS, and CDS Analytical Model 5500 Fixed Gas Analyser). A schema of the  
153 experimental setup is presented in Fig. 1.

154 The GC/MS injector temperature was kept at 280 °C. An Elite-35MS capillary column  
155 (30 m x 0.25  $\mu\text{m}$ ) was used for chromatographic separation. Helium (99.999 %) was  
156 selected as the carrier gas at a constant flow rate of 1 mL/min and a 1:80 split ratio. This  
157 was carried out to separate and identify the chemical composition of the bio-oil. Oven  
158 temperature was programmed from 40 °C (3 min) to 280 °C at a heating rate of 5 °C/min.  
159 The chromatograms were integrated, and relative peak areas were calculated and  
160 subsequently identified using the NIST library as a reference and only those with over an  
161 80% matching quality were considered. The FGA used a 1/8" packed column and thermal  
162 conductivity detector to analyse the gases produced during fast pyrolysis that were not  
163 easily assayed using capillary GC/MS. The absorbent trap of the pyrolyser collected the  
164 organic products from pyrolysis and transferred them to the GC/MS as usual.

165 Different experiments were performed by varying the OP/P ratio (1.5:1, 1:1, 1:1.5, and  
166 1:2), by keeping the total feedstock mass at 1 mg  $\pm$  0.05 mg and placing it in the middle  
167 of a quartz tube (2 mm diameter and 20 mm long) with a quartz wool base. The heating

168 rate and the reaction time of fast pyrolysis were optimized, set at 20 °C/ms, and 20 s [2].  
169 Firstly, fast pyrolysis took place at 500 °C to determine the optimum mass ratio.  
170 Subsequently, the influence of temperature on product distribution was performed at 500,  
171 550, 600, 650, and 700 °C. The experiments were carried out in triplicate for each sample  
172 to ensure reproducibility.

173 The peak area based on the Py-GC/MS and Py-FGA analysis could not show the real  
174 content in the target compounds. However, if the mass of the sample was the same in each  
175 pyrolysis experiment, the corresponding chromatographs could be compared to reveal the  
176 proportion of components in the bio-oils produced.

#### 177 2.4. Effective hydrogen to carbon molar ratio ( $H/C_{eff}$ )

178 The effective hydrogen to carbon molar ratio ( $H/C_{eff}$ ) is used to describe whether a  
179 feedstock can be upgraded or processed at existing refineries [25]. This factor shows the  
180 potential the feedstock has for converting it economically to hydrocarbons, and helps  
181 estimating the overall yield of olefin and aromatics [26]. The definition of  $H/C_{eff}$  is  
182 calculated by using the number of moles of the feedstock with the following equation  
183 [16]:

$$184 \quad H/C_{eff} = (H - 2O - 3N - 2S) / C \quad (\text{eq. 2})$$

185 Higher  $H/C_{eff}$  ratios in the feedstock implies more efficient conversion at refineries. This  
186 ratio is also shown by many researchers to estimate how a feedstock can be economically  
187 converted to hydrocarbons using a potential catalyst. The  $H/C_{eff}$  ratio of petroleum-  
188 derived feedstocks is between 1 and 2 and that for lignocellulosic biomass is between 0  
189 and 0.3 [27]. Therefore, pyrolysis oil from the latter feedstock may be considered to have  
190 hydrogen-deficient molecules when compared to those in the feedstock at the current  
191 petroleum refinery [25]. Therefore, to obtain pyrolysis oil with a high  $H/C_{eff}$  ratio,

192 agroindustrial plastics could be a potential hydrogen donor feedstock, to produce  
193 enhanced bio-oil.

#### 194 2.5.Synergistic evaluation

195 To obtain a comprehensive analysis of the results, theoretical values expected from each  
196 experiment were calculated by considering the proportion of mixtures as an arithmetic  
197 sum of the values obtained by single-feedstock fast pyrolysis, assuming there were no  
198 chemical interactions between either feedstock during devolatilisation. This method may  
199 be considered as one of the best approaches for analysing synergistic effects in properties  
200 after any pyrolytic process, since bio-oils from lignocellulosic biomass are not miscible  
201 with polyolefin oils [28].

### 202 3. Results and discussion

#### 203 3.1. Thermal behaviour of olive pomace and agroindustrial plastics

204 Fig. 2 shows the thermogravimetric (TG) and the derivate thermogravimetric (DTG)  
205 curves obtained for OP, PE, PS and PVC. The main pyrolysis stage for OP took place at  
206 temperatures between 120 and 450 °C, as seen from their DTG profile in Figure 2b. For  
207 OP, the shoulders observed at around 280 and 350 °C were attributed to hemicellulose  
208 and cellulose decomposition, respectively. These peaks were followed by a tail, which  
209 was ascribed to lignin decomposition, leading to char formation. However, a small peak  
210 was observed for olive pomace at nearly 400 °C, which could be attributed to lipid  
211 decomposition of the olive oil [29]. One outstanding observation was that the thermal  
212 stability of the biomass was lower than that for the plastics during pyrolysis. Thus, the  
213 free radicals arising when the biomass degraded might have promoted decomposition of  
214 plastic-derived macromolecules [3]. Moreover, thermal degradation of PE took place at  
215 a higher temperature range (410-520 °C) and was completed at 520 °C. As seen with the

216 behaviour of PE, the PS sample was thermally degraded from 350 to 500 °C. In general,  
217 thermal decomposition of most synthetic polymers (such as PE and PS) occurs by random  
218 scission and radical mechanisms: initiation, propagation and termination [30]. The  
219 carbon-to-hydrogen (C–H) bonds of primary, secondary, and tertiary carbon atoms, and  
220 carbon-to-carbon (C–C) sigma bond are the only ones that makes probability distribution  
221 in bond selection for radical formation or breakage [3]. However, in straight chain  
222 polymers, the main products formed are hydrogen gas and hydrocarbons [4]. In contrast  
223 with the thermal behaviour seen with PVC, two peaks are observed that correspond to  
224 two degradation stages. The first one occurred in the 240-410 °C temperature range. The  
225 dominant reaction here was dehydrochloration in which HCl and other chlorinated  
226 hydrocarbons were released. The second peak occurred in the 410-540 °C interval, in  
227 which rearrangement and cyclisation of conjugated polyene were the main reactions and  
228 the benzene derivatives generated. This indicates that carbon-chloride bonds during  
229 pyrolysis has less energy than that in C-C bonds [31]. Furthermore, note that the loss in  
230 mass could have occurred earlier when PVC was in the blend, indicating that promoted  
231 pyrolysis products could have occurred at lower temperatures [26]. In conclusion, those  
232 polymers selected could provide hydrogen, which is lacking in OP, due to the range of  
233 temperatures in fast pyrolysis. Also, there may have been improvements to product  
234 distributions for the OP/P blends.

### 235 3.2. Effect of OP/P mass ratio on product distribution

236 To analyse the influence of adding agroindustrial polymers (PE, PS and PVC) to product  
237 distribution, different OP/P mass blending ratios were evaluated. The optimal OP/P mass  
238 ratio was selected taking into account the hydrogen to carbon effective  $(H/C)_{\text{eff}}$  ratio  
239 which efficiently converted OP into elaborate biofuels. However, this data could  
240 demonstrate synergistic enhancement on hydrocarbon product distribution. Also,

241 polymers with different  $(H/C)_{\text{eff}}$  ratios were selected to be studied in a mixture with OP  
242 through fast pyrolysis to observe whether there was an improvement in hydrocarbon  
243 production efficiency or not. It must be stressed that fast pyrolysis temperature was set at  
244 500 °C. First, raw samples of the materials under study were pyrolysed to establish their  
245 product composition. The results showed significant differences in pyrolysis product  
246 distribution for OP and agroindustrial polymer samples. The relative percentage of  
247 hydrocarbons in OP bio-oil reached a maximum of 25%, while the remainder were the  
248 oxygenates formed based on alcohol, aldehydes, acids, esters, ketones and phenols. The  
249 most representative group identified in OP pyrolysis were phenolics from the thermal  
250 decomposition of lignin [23]. In contrast, from pyrolysis product distribution of  
251 agroindustrial polyolefins at 500 °C, hydrocarbons were the main functional group  
252 detected, and they accounted for nearly 100%. The theoretical (Theo) and experimental  
253 (Exp) data for the carbon yield (%) of each functional group detected after fast pyrolysis  
254 of the different OP/P mass ratios were compared to study the synergistic effects.  
255 Theoretical carbon yield ( $CY_{\text{th}}$ ) can be calculated as follows [29]:

$$256 \quad CY_{\text{th}} = CY_{\text{OP}} \cdot F_{\text{OP}} + CY_{\text{P}} \cdot F_{\text{P}} \quad (\text{eq. 3})$$

257 Where  $F_{\text{OP}}$  and  $F_{\text{P}}$  are the mass fraction of olive pomace and polyolefin used (varied PE,  
258 PS and PVC), and  $CY_{\text{OP}}$  and  $CY_{\text{P}}$  are the carbon yields (%) obtained from fast pyrolysis.  
259 A positive rise in carbon yield production in comparison with the theoretical value of the  
260 blend could be deemed a synergistic effect in which more hydrocarbons are formed. Fig.  
261 3 summarised this effect on hydrocarbon distribution for the different OP/P mass ratios  
262 with the fast pyrolysis experiments performed at 500 °C.

263 Figure 3 compares theoretical and experimental carbon yields of total hydrocarbons  
264 produced from fast pyrolysis for different OP/P mass blending ratios. With PE, its product

265 distribution is mainly based on hydrocarbons, with alkanes being the most representative  
266 group detected, with a yield of 64.17%. The PE composition gave rise to a greater HHV  
267 of 41 MJ/kg (Table 1), similar to those in commercial fuels, such as gasoline or diesel  
268 [32]. However, PE in blends with biomass feedstock was expected to reduce the oxygen  
269 content at the organic phase whilst carbon yield increased, leading to a higher HHV. The  
270 synergistic effect between the OP/PE mass blending ratios during fast pyrolysis has been  
271 discussed. In Fig. 3a, synergy was detected on comparing theoretical and experimental  
272 hydrocarbon formation for OP/PE blends. Aliphatic hydrocarbons derived from pyrolysis  
273 of PE might have interacted with the oxygenate fraction present in OP through a series of  
274 cracking, cyclization and isomerization reactions, thereby increasing hydrocarbon  
275 diversity (alkenes and cyclic hydrocarbons) [16]. According to the results, the 1.5:1  
276 OP/PE mass blending ratio produced the highest percentage of hydrocarbons. The  
277 maximum percentage of hydrocarbons reaches a maximum of 89.73% at 500 °C. The  
278 hydrocarbons obtained are mostly aliphatic compounds, especially long chain olefins, and  
279 they amount to 79.44% of total condensable compounds. Thus, it may be concluded that  
280 mixtures of OP and PE by fast pyrolysis can modify the reaction mechanism, thereby  
281 removing oxygenates by substituting decarbonylation and decarboxylation with  
282 dehydration [3]. Previous studies have demonstrated the modification of reaction pathway  
283 of the co-feeding of lignocellulosic biomass with hydrogen-rich feedstocks in catalytic  
284 pyrolysis [33–35]. In addition, hydrogen abstraction by OP oxygenates and reactive free  
285 radicals could facilitate thermal degradation of PE, which explains the synergy on  
286 increasing aliphatic hydrocarbons [3].

287 Regarding Fig. 3b, from the OP/PS blends, positive synergy was observed with an  
288 increase in hydrocarbon, specifically the carbon yield of aromatic compounds. Bio-oil  
289 composition obtained from PS fast pyrolysis was shown to be oxygen-free. It has

290 significant hydrocarbon content which amount to 64.54% of carbon yield. In comparison  
291 with PE, PS has a low aliphatic content due to its aromatic ring. Also, PS has a similar  
292 HHV to commercial liquid fuels, and has the capacity to reducing oxygen increasing  
293 carbon yield. Different OP/PS mass ratios were studied to observe the synergistic effect  
294 in fast pyrolysis by comparing theoretical and experimental results. The aromatic yields  
295 obtained are discussed between ratio blends. Moreover, the 1:1.5 OP/PS mass blend ratio  
296 provided the highest percentage of hydrocarbons. The relative content of hydrocarbons  
297 in this blend represented 96.23% of carbon yield at 500 °C, in which 68.57% of the  
298 condensable compounds detected were aromatics. These results were coherent with the  
299 literature, in which adding PS in blends with lignocellulosic biomass enhanced the quality  
300 of the bio-oil obtained, as oxygenated compounds transformed into aromatics [36–38].

301 Some special polymers, such as PVC, has attracted attention for use in co-pyrolysis. PVC  
302 was selected despite its low  $H/C_{\text{eff}}$  in comparison with other polyolefins. In Fig. 3c, a  
303 synergistic effect in the OP/PVC blends by fast pyrolysis at 500 °C was assumed. Results  
304 revealed hydrocarbon compounds were the main product obtained from PVC pyrolysis,  
305 with aromatics being the most representative hydrocarbon with 53.43% of total carbon  
306 yield. Zhou et al. researched the pyrolysis mechanism for PVC, and explained that  
307 formations of aromatics can be ascribed to molecular rearrangement and cyclisation of  
308 fragments of polyene [31]. PVC has low HHV, which may not help to reduce oxygen  
309 content in blends with OP. Moreover, OP has a higher  $H/C_{\text{eff}}$  ratio in comparison to PVC.  
310 The hydrogen content in OP could act as a hydrogen donor for PVC in fast pyrolysis. In  
311 this case, water evolved from OP pyrolysis could also act as a reactive compound, thereby  
312 accelerating further cracking of PVC, and hence, promoting yields of bio-oil [3].  
313 Furthermore, PVC chloride atoms are unstable and easily split at higher temperatures.  
314 Dehydrochlorinated and chlorinated hydrocarbons were formed at temperatures below

315 500 °C. As the pyrolysis temperature was set at 500 °C, chloride compounds were not  
316 detected in this study. Additionally, removing HCl typically includes alkali and alkaline  
317 earth metal substances as absorbents, as the inherent metals presents in the organic matrix  
318 of OP [31]. This will be discussed in more detail below. The blend with the 1:1.5 OP/PVC  
319 mass ratio produced the highest carbon yield of hydrocarbon compounds (85.23%) at 500  
320 °C, in which 54.04% of the components detected were aromatics.

321 These results demonstrated the synergy observed on comparing theoretical and  
322 experimental data. This proved to have an apparent effect on bio-oil hydrocarbon  
323 production, thus enhancing use of OP [16]. Additionally, one of the forms to evaluate the  
324 bio-oil quality is through its calorific value. Closely, this depends on carbon, hydrogen  
325 and oxygen contents, as well as the  $H/C_{eff}$  ratio. When plastics are used in the reaction,  
326 this ratio increases due to the contribution of the H atoms that the plastics provide, and  
327 consequently the O atoms decrease, thus increasing the calorific value [39]. For this  
328 reason, synergy was obtained between OP and agroindustrial polyolefins (PE, PS and  
329 PVC), where the oxygenates were reduced and hydrocarbon compounds fomented.  
330 Indeed, this effect was supported by the effective hydrogen to carbon ratio values  
331 obtained for the blends. The carbon yields of hydrocarbon compounds in bio-oil with  
332  $H/C_{eff}$  ratios are given in Fig. 4. The convex parabolic curves in Fig. 4a and 4b  
333 demonstrate there is a synergistic effect between OP and PE or PS blends in fast pyrolysis.  
334 Alkenes production reached a maximum carbon yield for the 1.5:1 OP/PE blend ratio, in  
335 which the  $H/C_{eff}$  mixture marked 0.88. This result suggests that the 1.5:1 OP/PE ratio  
336 should mean there is enough hydrogen added to the blend to maximise hydrocarbon  
337 products. However, this small amount of PE in the blend could stabilize most of the  
338 unstable compounds in the bio-oil by hydrogenation. The unstable molecules that tend to  
339 increase coke formation include oxygenates with unsaturated bonds [27]. Conversely,

340 aromatic compounds reached maximum yields when OP/PS and OP/PVC blend ratios  
341 were 1:1.5, in which  $H/C_{\text{eff}}$  was 0.66 and 0.11, respectively. Remarkably, in OP/PVC  
342 mixtures aromatic production synergy was observed in the bio-oils produced, even though  
343 PVC showed a low hydrogen to carbon effective value. Despite this, PVC was  
344 characterised as hydrogen-rich and intermediate in a blend with OP during fast pyrolysis.  
345 Furthermore, oxygenated chemical bonds from the OP organic matrix could favour chain  
346 scission and breakages in long-chain organic matter in polyolefins [3]. The convex curves  
347 shown demonstrate synergy in the OP/P blends. Thus, this may indicate that  
348 thermochemical processes efficiently convert more undefined oxygenates into  
349 hydrocarbons when producing bio-oil from these blends. The best options for OP/P mass  
350 blending ratios were 1.5:1 OP/PE, 1:1.5 OP/PS and 1:1 OP/PVC. These were the optimal  
351 proportion of polyolefins and OP feedstock for enhancing olefin and aromatic compounds  
352 production yields.

### 353 3.3. Influence of reaction temperature on fast pyrolysis products for the OP/P mass 354 ratios selected

355 To study the influence of fast pyrolysis temperature on hydrocarbon composition, 1.5:1  
356 OP/PE, 1:1.5 OP/PS and 1:1.5 OP/PVC blends were evaluated at a range of 500-700 °C.  
357 The experimental hydrocarbon composition obtained by varying reaction temperature are  
358 shown in Fig. 5.

359 Fig. 5 shows how hydrocarbon production is promoted as fast pyrolysis temperature  
360 increases. This trend varied in carbon yields depending on the feedstock selected. From  
361 the literature, higher temperatures in fast pyrolysis indicate that more energy is available  
362 for rupturing organic bonds. This could promote devolatilisation in feedstocks and  
363 facilitate endothermic co-pyrolysis reactions [16]. Regarding the 1.5:1 OP/PE mass blend

364 ratio, maximum carbon yield was detected at 650 °C, with a maximum production of  
365 aliphatic compounds (93.97% of total hydrocarbons). Remarkably in Fig. 5a for 1.5:1  
366 OP/PE at 650 °C the following carbon yields were observed: 91.02% of alkenes and  
367 2.94% of alkynes. Alkene yields initially accounted for 61% at 500 °C and reached a  
368 maximum value at 650 °C. Strikingly, in the case of 1:1.5 OP/PS blend (Fig. 5b), 500 °C  
369 was selected as the optimal temperature due to the positive results obtained in aromatic  
370 hydrocarbons production yields. When PS was used in the blend, aromatic carbon yields  
371 declined as temperature increased, reaching minimum yields at 600 °C. By contrast, at  
372 600 °C alkene yields reached a maximum (52.35%) for the 1:1.5 OP/PS mass blending  
373 ratio. Further studies on pyrolysis established that 500-550 °C was the optimum  
374 temperature range for maximising hydrocarbon composition from waste plastic based on  
375 polystyrene [5]. In Fig. 5c, carbon yields obtained from a temperature analysis of the 1:1.5  
376 OP/PVC mass blending ration is shown. The results revealed a maximum yield for the  
377 total sum of hydrocarbons at 650 °C which accounted for 95.16%. Aromatics were the  
378 main product detected with 64.62% of carbon yield. At lower temperatures they fell to  
379 the detriment of alkenes. In contrast, when pyrolysis temperature increased, this trend  
380 was reverted. This might have been due to reforming and aromatisation of alkenes, and  
381 the side chains of the aromatic ring structure cracking during depolymerisation under high  
382 temperature conditions [31]. In this study, no organic chloride compounds were found  
383 above 500 °C for OP/PVC blends. According to the research of Chen et al. [40],  
384 interactions might occur between the pyrolysis products of PVC and the volatiles of  
385 biomass pyrolysis, or between the products of PVC and the solid char residue of biomass  
386 pyrolysis. To clarify this finding, ultimate and mineral content analysis were carried out  
387 to the solid char residue generated by 1:1.5 OP/PVC at 650 °C sample (Table 2). The  
388 carbon content greatly increased (76.40 wt.%) while oxygen content was reduced (17.25

389 w.%) in comparison with the data from raw OP (Table 1). In addition, HSEM coupled  
390 with EDX was carried out to study the residual carbon obtained after fast pyrolysis at 650  
391 °C for OP/PVC blends. As shown in Fig. 6a, the carbon residue obtained for 1:1.5  
392 OP/PVC showed a porous structure. It is due to the amount of volatile matter, as well as  
393 the chemical composition of OP, in which hemicellulose content was detected in higher  
394 proportion (Table 1) [2,41,42]. EDX results (Fig. 6b) demonstrate that carbon, in  
395 agreement with ultimate analysis (Table 2), was the major contributor to the sample.  
396 Interestingly, chlorine was detected in the same weight percentage as inherent AAEMs  
397 (K). Fig. 6c shows the FTIR spectrum of the obtained carbon residue. A strong absorption  
398 peak at 3100-2600  $\text{cm}^{-1}$  corresponding to the asymmetrical stretching of H-Cl was  
399 observed. Moreover, an absorption peak at 750-660  $\text{cm}^{-1}$  verified the existence of -C-Cl  
400 bonds in the char residue [31]. Thus, these results suggest that the rests of chlorinated  
401 compounds from PVC were retained by the char residue. It has been proposed that  
402 biomass materials can act as catalyst which inhibits the dehydrochlorination process or  
403 promotes the chain scission of PVC. Thus, dehydrochlorination during PVC fast pyrolysis  
404 is only partly completed [43]. Similar results were obtained from Kuramochi et al. [44],  
405 where chlorine emissions were reduced by the presence of wood during pyrolysis of PVC.  
406 The reason suggested by these authors was that hemicellulose may reduce HCl emission  
407 by fixing the Cl into the pyrolyzed residue. As commented, our results also indicates that  
408 chlorinated compounds were retained in the carbon residue. Thus, it would be an effective  
409 method for degradation and dechlorination of chlorine-containing plastics to produce high  
410 quality liquid products. The generation of chlorinated hydrocarbons would be avoided  
411 increasing the pyrolysis temperature and with the presence of inherent alkali and alkaline  
412 earth metals (AAEMs) from lignocellulosic biomass, acting as adsorbents. Indeed, one

413 the typical methods to remove HCl involves the use of AAEMs as adsorbents [31, 45–  
414 47].

415 Regarding the hydrocarbon distribution in the bio-oil, a significant increase in the yield  
416 of aliphatic compounds for PE and aromatic hydrocarbons for PS and PVC in blend with  
417 OP can be noted as reaction temperature increased. Moreover, aromatics were the main  
418 products for OP with PS and PVC blends, including monocyclic aromatic hydrocarbons  
419 (such as benzene, toluene and xylene) and polycyclic aromatic hydrocarbons (PAHs).  
420 Both aromatics and olefins are essential as feedstock for the manufacture of valuable  
421 products, such as pharmaceutical compounds, paints, solvents, among others [39].  
422 Moreover, light olefins together with BTX (benzene, toluene and xylene) are the most  
423 common preliminary petrochemicals [48]. To further investigate the effect of temperature  
424 on the production of that valuable components, an analysis of the major fraction obtained  
425 in the bio-oils were carried out. Table S1 shows a detail breakdown of the hydrocarbon  
426 distribution in the bio-oils obtained at different reaction temperatures. The effect of  
427 temperature on the carbon number distribution of aliphatic hydrocarbons for 1.5:1 OP/PE  
428 blends is shown in Fig. 7. The aliphatic fraction was in the range of C<sub>6</sub>-C<sub>20</sub>, which were  
429 mainly derived from the random chain scission of the PE [39, 48]. As temperature  
430 increase, the relative content of light hydrocarbons (C<sub>6</sub>-C<sub>10</sub>) first increased and then  
431 decreased, reaching the maximum at 650 °C, accounting C<sub>6</sub> and C<sub>7</sub> maximum carbon  
432 yields. On the other hand, the hydrocarbon fractions between C<sub>11</sub> and C<sub>20</sub> exhibited a  
433 different trend, first decreasing and then increasing with reaction temperature. These  
434 results suggest that the long C-C chains cracking to form light hydrocarbons (C<sub>6</sub>-C<sub>10</sub>) was  
435 favoured as reaction temperature increased, being optimum at 650 °C.

436 The influence of reaction temperature for 1:1.5 OP/PS was done focused on aromatic  
437 distribution. Both PS and OP (specially lignin, one of its major chemical fractions) are

438 aromatic in nature. Additionally, there is a possibility of PAHs formation from their  
439 degradation products. In this work, PAHs are categorized as naphthalene and its  
440 derivatives, indene and its derivatives, and multiring aromatic components having more  
441 than two rings. Fig. 8a shows the aromatic selectivity of 1:1.5 OP/PS blend as reaction  
442 temperature increased. The obtained aromatic compounds were divided into benzenes  
443 (benzene derivatives include ethylbenzene, propylbenzene, 2-propenylbenzene, 1-  
444 propenylbenzene, cyclopropylbenzene, biphenyl, terphenyl, etc.), toluene and PAHs. The  
445 presence of naphthalene and indene derivatives could be attributed to the generation of  
446 methyl groups during lignin depolymerization [49]. In general, benzenes and toluene  
447 selectivity decreased while PAHs were in the same percentage yield when reaction  
448 temperature increased. As PAHs are undesirable in the bio-oil because they are highly  
449 susceptible to coke formation [49-51], a reaction temperature of 500 °C would be the best  
450 for this blend.

451 For the 1:1.5 OP/PVC blend, the effect of temperature was studied dividing the obtained  
452 aromatic fraction into benzenes and its derivatives, toluene, xylene and PAHs. For this  
453 sample, it was obtained a higher yield of aromatics than the pyrolysis of the raw biomass.  
454 Interestingly, the most valuable monocyclic aromatics (toluene and xylene) were found  
455 for the OP/PVC blend, but not for the raw PVC. Thus, the combination of OP and PVC  
456 improves the final product, generating valuable chemicals like BTX. The aromatic  
457 selectivity of 1:1.5 OP/PVC blend at different reaction temperature is shown in Fig. 8b.  
458 BTX were obtained in a major proportion at 650 °C, accounting 37.9 % of aromatic  
459 selectivity. Among the PAHs observed, two-ring PAH, naphthalenes and indenenes,  
460 dominated in all the samples. Naphthalene selectivity was decreased as reaction  
461 temperature increased, from 9.7 to 6.4 %. In general, monocyclic aromatics were favoured  
462 versus PAHs fractions as reaction temperature increased.

463 To sum up, the fraction of light hydrocarbons ( $C_6-C_{10}$ ) was favoured as reaction  
464 temperature increased for 1.5:1 OP/PE, being the optimum temperature 650 °C. On the  
465 other hand, valuable chemicals like toluene or xylene were found abundant in the bio-  
466 oils, and useful aromatics like styrene or ethylbenzene were amply formed for PS and  
467 PVC mixed with OP. They increased with temperature for 1:1.5 OP/PVC, but not for  
468 1:1.5 OP/PS, being therefore the optimum temperature 650 °C for the former and 500 °C  
469 for the later.

470 Pyrolytic gas was collected during fast pyrolysis experiments and analysed with FGA.  
471 The fixed gas compounds that passed through the trap in fast pyrolysis at 500 °C for the  
472 materials under study are shown in Fig. 9. The thermochemical processes involved high  
473 rates of temperature conducive to endothermic reactions, such as Boudouard, water-gas,  
474 and steam-methane reforming [29]. The main gases given off during pyrolysis were CO,  
475  $CH_4$ ,  $CO_2$ ,  $C_2H_2$ ,  $C_2H_4$  and  $C_2H_6$  and the main ones obtained throughout the whole process  
476 were CO,  $CO_2$  and  $CH_4$  for all samples. Olive pomace had the highest  $CO_2$  yield and  $CH_4$   
477 emissions, which correlated with the elemental analysis (Table 1). CO and  $CO_2$  are  
478 formed from thermal decomposition of oxygen functionality in their inner lignocellulosic  
479 organic matrix [52]. The light hydrocarbons detected came from thermal cracking and  
480 methanisation [53]. The highest  $CH_4$  yield may have been promoted by the inherent  
481 amount of potassium content in the OP sample (Table 1). This metal is reported in the  
482 bibliography as an active catalyst for methanation [29]. However, there is a close  
483 correlation between the  $H/C_{eff}$  ratio and pyrolytic gas emissions yields. Zhang et al. stated  
484 that as the  $H/C_{eff}$  ratio feedstock rose, CO and  $CO_2$  emissions also increased [27]. For that  
485 reason, as PE is one of the polyolefins with the highest ratios of  $H/C_{eff}$ , high amounts of  
486 CO and  $CO_2$  emissions were observed. For OP, the sum of the oxygenated gases  
487 represented the most important fraction detected. Furthermore, as polymers are abundant

488 sources of hydrogen, a greater amount of light hydrocarbons gases ( $C_2H_2$ ,  $C_2H_4$  and  $C_2H_6$ )  
489 were detected in comparison with lignocellulosic OP. However, these gases might have  
490 evolved in aromatisation, thereby enhancing carbon efficiency yields of aromatic  
491 compounds in blends with OP [6].

492 Using biomass and polymers to improve fast pyrolysis product yields and quality is an  
493 important issue for research on pyrolysis gas emissions. The pyrolytic gas produced from  
494 the combination of biomass-polymer pyrolysis comes into contact swiftly before cooling,  
495 and is catalysed into  $CO_2$  or CO. Synergy is observed after these processes, when oxygen  
496 is removed, and high-quality bio-oil is obtained [54]. Regarding the pyrolytic gas  
497 emission results for the blends used, it might be concluded that they did not follow any  
498 clear trend due to the synergy between the raw materials. Table 3 shows the effect of  
499 temperature on yields of the gases given off which were detected for the previously  
500 selected mass blending ratio samples. Overall, a reduction in CO and  $CO_2$  product yields  
501 was seen, when OP was mixed with polymers. Blends with a low  $H/C_{eff}$  ratio should  
502 produce more CO by decarbonylation, as they are hydrogen-deficient compounds, which  
503 probably limit the amount of oxygen transferred into water and favour CO production  
504 [27]. Also,  $CO_2$  formation by the water-gas shift reaction as a function of  $H/C_{eff}$  shows  
505 the same trend as that for CO [27,55]. For the 1.5:1 OP/PE, the range of temperatures  
506 between 600 to 650 °C, limited  $CO_2$  emission was observed. These results are coherent  
507 with hydrocarbon compounds yields at those temperatures at which oxygenates were  
508 suppressed. Moreover,  $CH_4$  became the most representative gas in PE blends, due to the  
509 features this plastic has, as it is a carbon chain. The effect of temperature on the 1:1.5  
510 OP/PS displayed ever increasing formation of  $CH_4$  and reached 0.55 wt.%/g-sample at  
511 700 °C. CO and  $CO_2$  formation was largely constant throughout the range of temperature,  
512 although lower gas yields were detected as temperature increased. These were in keeping

513 with the previously described trend in hydrocarbon yields. With the 1:1.5 OP/PVC  
514 sample, a reduction in CO and CO<sub>2</sub> were observed while temperature increased.  
515 Maximums oxygenate gas yields were given off in the 500-600 °C range, where  
516 hydrocarbon yields ranked low. However, PVC samples displayed similar trends for CH<sub>4</sub>  
517 with maximum yields reached as temperature increased. This was related to the decrease  
518 in alkanes yields obtained.

519 In general, results indicated that CO and CO<sub>2</sub> formation were both reduced during co-  
520 pyrolysis. While olefin and aromatic carbon yields increased for higher H/C<sub>eff</sub> ratio  
521 blends, CO and CO<sub>2</sub> carbon yields decreased. However, these outcomes implied that  
522 hydrogen transfers from polymers to biomass-derived oxygenates may have mitigated  
523 polymerisation and cross-linking reactions and suppressed decarbonylation and  
524 decarboxylation reactions to generate CO and CO<sub>2</sub> [3]. According to these results, thermal  
525 decomposition of CO and CO<sub>2</sub> fell as temperature increased. Also, light hydrocarbon  
526 emissions were promoted with temperature, especially for CH<sub>4</sub>. Thus, in conclusion, as  
527 temperatures rose, thermal cracking and methanisation were promoted as the inherent  
528 potassium in OP become more active as a catalyst.

#### 529 **4. Conclusions**

530 The effects of fast pyrolysis temperature and olive pomace (OP) and agroindustrial  
531 polymers (PE, PS and PVC) mass blending ratios on product distribution and  
532 hydrocarbon composition were researched. Experimental results demonstrated an  
533 apparent synergy with hydrocarbon yields in fast pyrolysis bio-oil at 500 °C for the 1.5:1  
534 OP/PE, 1:1.5 OP/PS and 1:1.5 OP/PVC mass ratios. Moreover, alkenes are maximised  
535 for OP/PE and aromatic compounds for OP/PS and OP/PVC blends. Total hydrocarbon  
536 yields increased as temperatures rose from 500 to 700 °C. For the 1.5:1 OP/PE blend,

537 alkenes formation was improved, where the light hydrocarbons fraction (C<sub>6</sub>-C<sub>10</sub>) first  
538 increased and then decreased with temperature, reaching a maximum at 650 °C. On the  
539 other hand, aromatic compounds production was enhanced for 1:1.5 OP/PS and 1:1.5  
540 OP/PVC, with optimal temperatures at 500 and 600 °C, respectively. The more valuable  
541 aromatic chemicals like benzenes, toluene and xylene were found to be abundantly  
542 formed for both blends. As for pyrolytic gas composition, they did not follow any trend,  
543 owing to the synergy in the blends. In general, while temperature increased, CO and CO<sub>2</sub>  
544 decreased and CH<sub>4</sub> was promoted. Finally, H<sub>2</sub>-rich feedstocks (plastics) and a H<sub>2</sub>-  
545 deficient (OP) offers synergy in improving aliphatic and/or aromatic carbon yields of the  
546 bio-oils formed.

#### 547 **CRedit authorship contribution statement**

548 **A. Alcazar-Ruiz:** Conceptualization, Investigation, Writing - original draft, Data  
549 curation, Supervision. **F. Dorado:** Format analysis, Methodology, Funding acquisition,  
550 Writing - review & editing. **L. Sanchez- Silva:** Validation, Resources, Writing - review  
551 & editing.

#### 552 **Acknowledgements**

553 The authors wish to thank the regional government of Castilla -La Mancha for their  
554 financial support. (Project SBPLY/17/180501/000238).

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Table 1. Characteristic data analysis of the materials.

Sample	Proximate analysis (wt.%) <sup>*daf</sup>				Ultimate analysis (wt.%) <sup>*daf</sup>				HHV	H/C <sub>eff</sub>
	Moisture	Ash	Volatile matter	Fixed carbon <sup>*diff</sup>	C	H	N	O <sup>*diff</sup>	(MJ/kg)	
OP	1.36	3.39	79.92	15.31	49.88	6.12	0	43.99	20.01	0.15
PE	0.45	0.22	99.27	0.06	85.70	14.20	0.05	0.05	46.63	1.98
PS	0.26	2.39	97.30	0.04	92.31	7.72	0	0	41.27	1.01
PVC	0.23	1.04	98.73	0.04	40.03	5.09	0	0.65	19.87	0.09
OP Chemical composition (wt.%) <sup>*db</sup>										
			Klason Lignin		21.2					
			Hemicellulose		31.5					
			Extractives		38.1					
OP Mineral Content (wt.%)										
			Ca		0.45					
			K		3.48					
			Mg		0.068					
			Na		0.059					

749 <sup>\*daf</sup>: dry and ash free basis; O<sup>diff</sup>: % of oxygen calculated from difference in C, H, N and S; Fixed carbon<sup>\*diff</sup>: % of fixed

750 carbon calculated from differences in moisture, ash and volatile matter; <sup>\*db</sup>: dry basis.

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755 Table 2. Characteristic data analysis of 1:1.5 OP/PVC char residue at 650 °C.

Ultimate analysis (wt.%) <sup>*daf</sup>			
C	H	N	O <sup>*diff</sup>
76.40	5.67	0.68	17.25
Mineral content (wt.%)			
Ca	K	Mg	Na
0.24	1.58	0.057	0.049

756 <sup>\*daf</sup>: dry and ash free basis

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771 Table 3. Pyrolytic gas composition results at different fast pyrolysis temperatures.

Sample	Reaction temperature (°C)	wt.% / g-sample					
		CO	CH <sub>4</sub>	CO <sub>2</sub>	C <sub>2</sub> H <sub>2</sub>	C <sub>2</sub> H <sub>4</sub>	C <sub>2</sub> H <sub>6</sub>
1,5:1 OP/PE	500	0.17	0.21	0.26	0.16	0.13	0.09
	550	0.11	0.21	0.26	0.18	0.16	0.09
	600	0.18	0.29	0.17	0.07	0.21	0.08
	650	0.16	0.24	0.17	0.15	0.16	0.12
	700	0.15	0.10	0.26	0.23	0.12	0.13
1:1,5 OP/PS	500	0.30	0.35	0.19	0.11	0.03	0.03
	550	0.26	0.35	0.19	0.11	0.03	0.06
	600	0.07	0.31	0.36	0.14	0.04	0.09
	650	0.06	0.34	0.14	0.33	0.06	0.08
	700	0.14	0.55	0.11	0.07	0.09	0.04
1:1,5 OP/PVC	500	0.34	0.06	0.29	0.21	0.06	0.07
	550	0.32	0.09	0.29	0.17	0.06	0.07
	600	0.62	0.15	0.07	0.05	0.05	0.06
	650	0.14	0.36	0.17	0.20	0.05	0.08
	700	0.19	0.52	0.16	0.06	0.02	0.04

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783 **Figure captions**

784 Figure 1. Schematic experimental setup of Py-GC/MS-FGA used for co-pyrolysis  
785 experiments.

786 Figure 2. TG and DTG profiles of OP, PE, PS and PVC; a) TGA curves and b) DTG  
787 curves for co-pyrolysis.

788 Figure 3. Mass ratio effect on fast pyrolysis hydrocarbon production at 500 °C of a)  
789 OP/PE, b) OP/PS and c) OP/PVC

790 Figure 4. Hydrocarbon composition in the bio-oil obtained through fast pyrolysis with  
791 different  $H/C_{eff}$  ratios of a) OP/PE, b) OP/PS and c) OP/PVC feedstock blend ratios.

792 Figure 5. Influence of temperature fast pyrolysis on a) 1.5:1 OP/PE, b) 1:1.5 OP/PS and  
793 c) 1:1.5 OP/PVC feedstock blend ratios.

794 Figure 6. Characterisation analysis of the obtained 1:1.5 OP/PVC char residue after fast  
795 pyrolysis at 650 °C: a) HRSEM image, b) EDX analysis and c) FTIR spectra.

796 Figure 7. Effect of reaction temperature on the carbon number distribution of aliphatic  
797 hydrocarbons in the bio-oil obtained for the 1.5:1 OP/PE sample.

798 Figure 8. Effect of reaction temperature on the aromatic compounds distribution in the  
799 bio-oil obtained for the a) 1:1.5 OP/PS and b) 1:1.5 OP/PVC sample.

800 Figure 9. Pyrolytic gas yields obtained (wt.%/g-sample) from fast pyrolysis of OP, PE,  
801 PS and PVC at 500 °C.

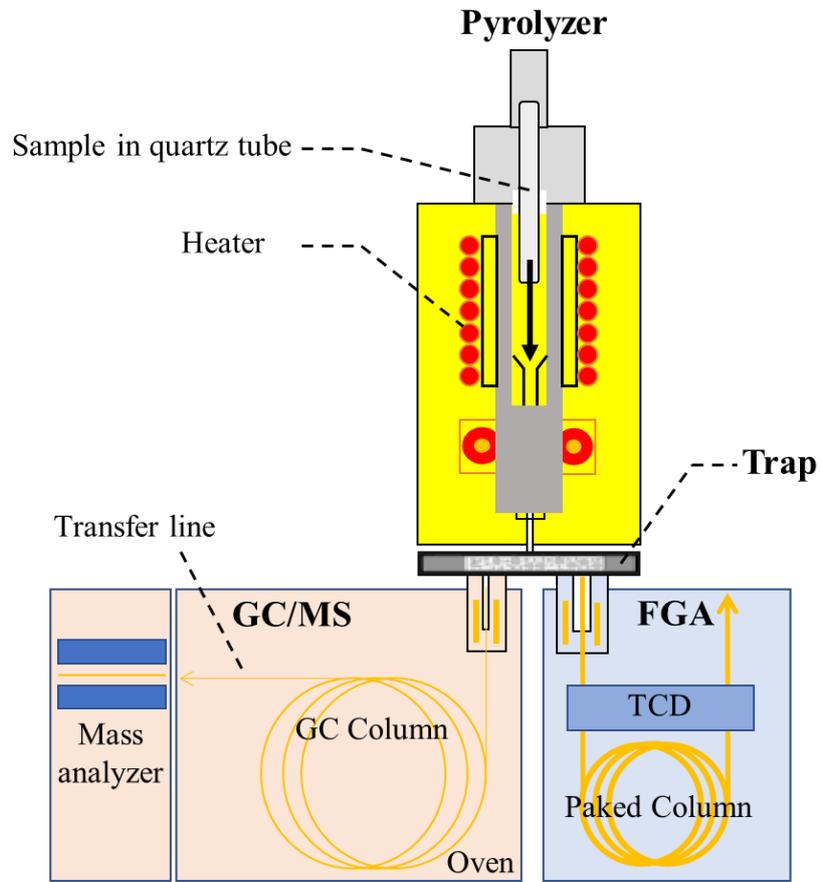
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806 Figure 1.



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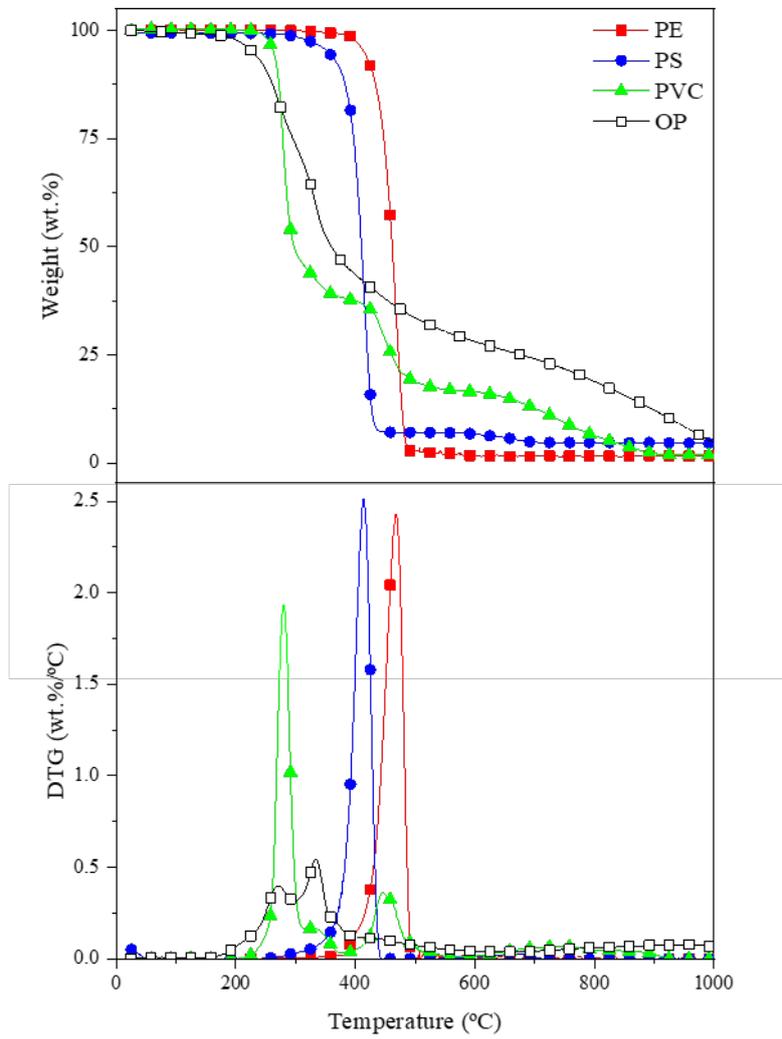
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821 Figure 2.



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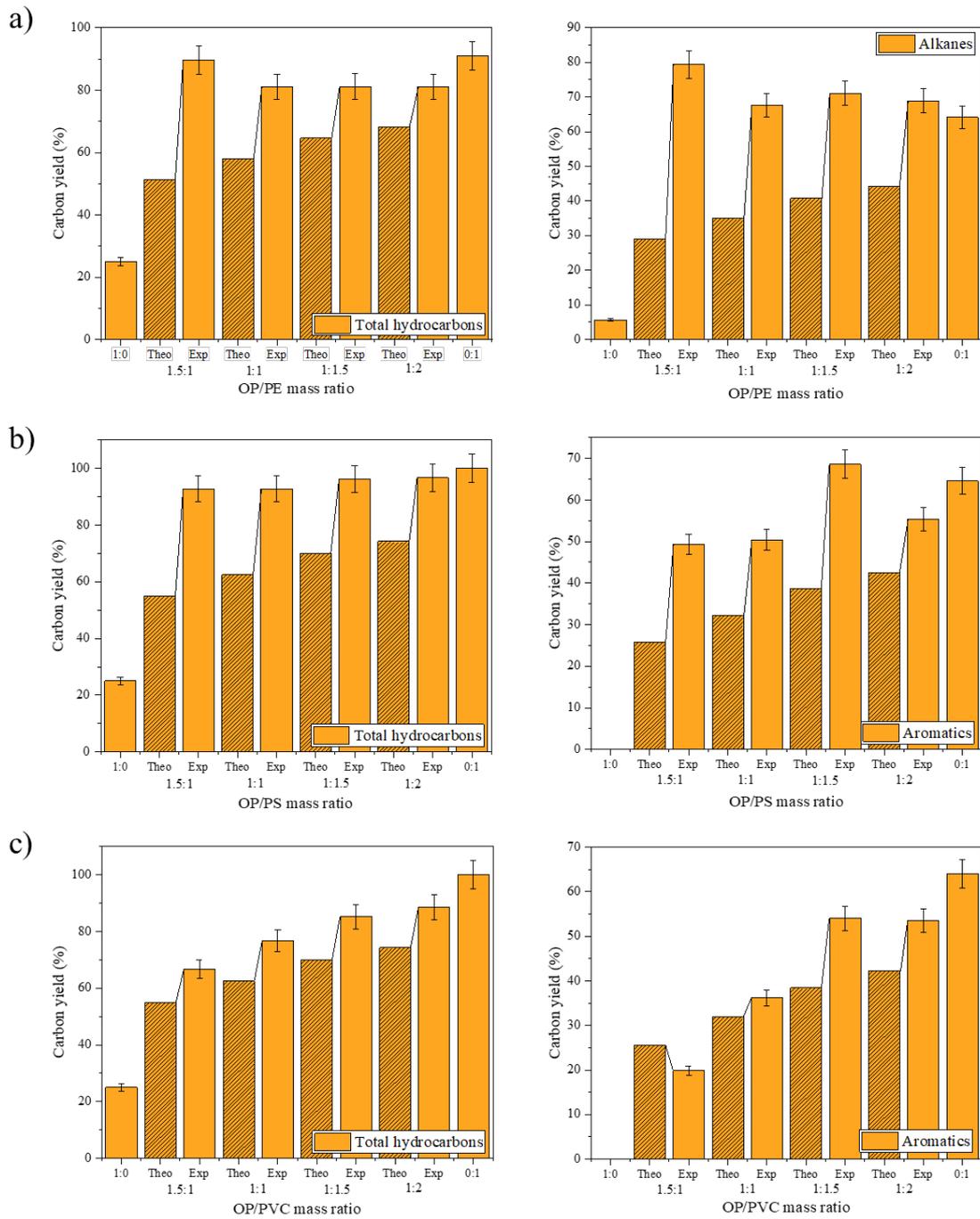
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833 Figure 3.



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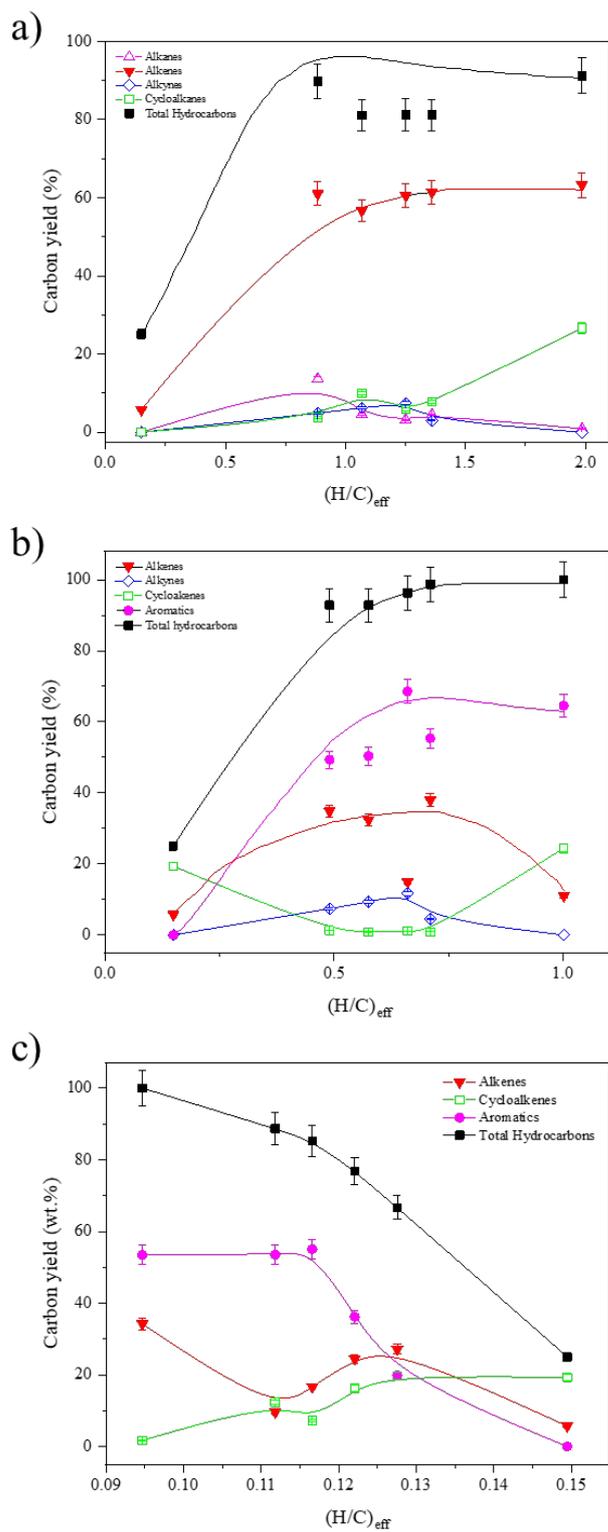
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840 Figure 4.

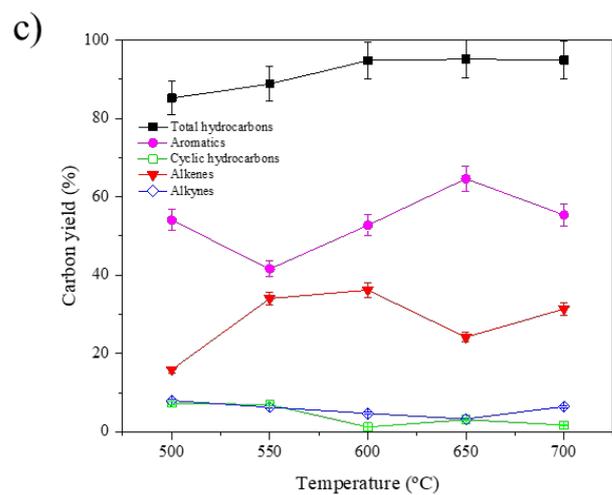
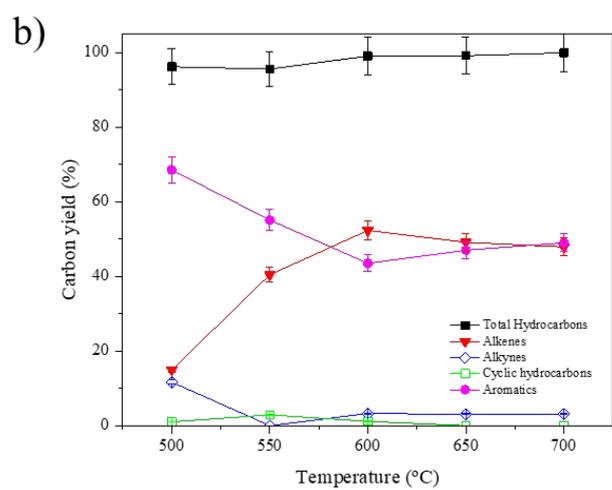
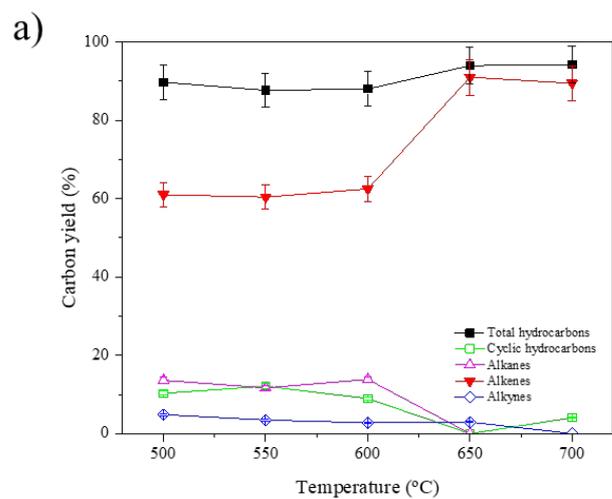


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844 Figure 5.

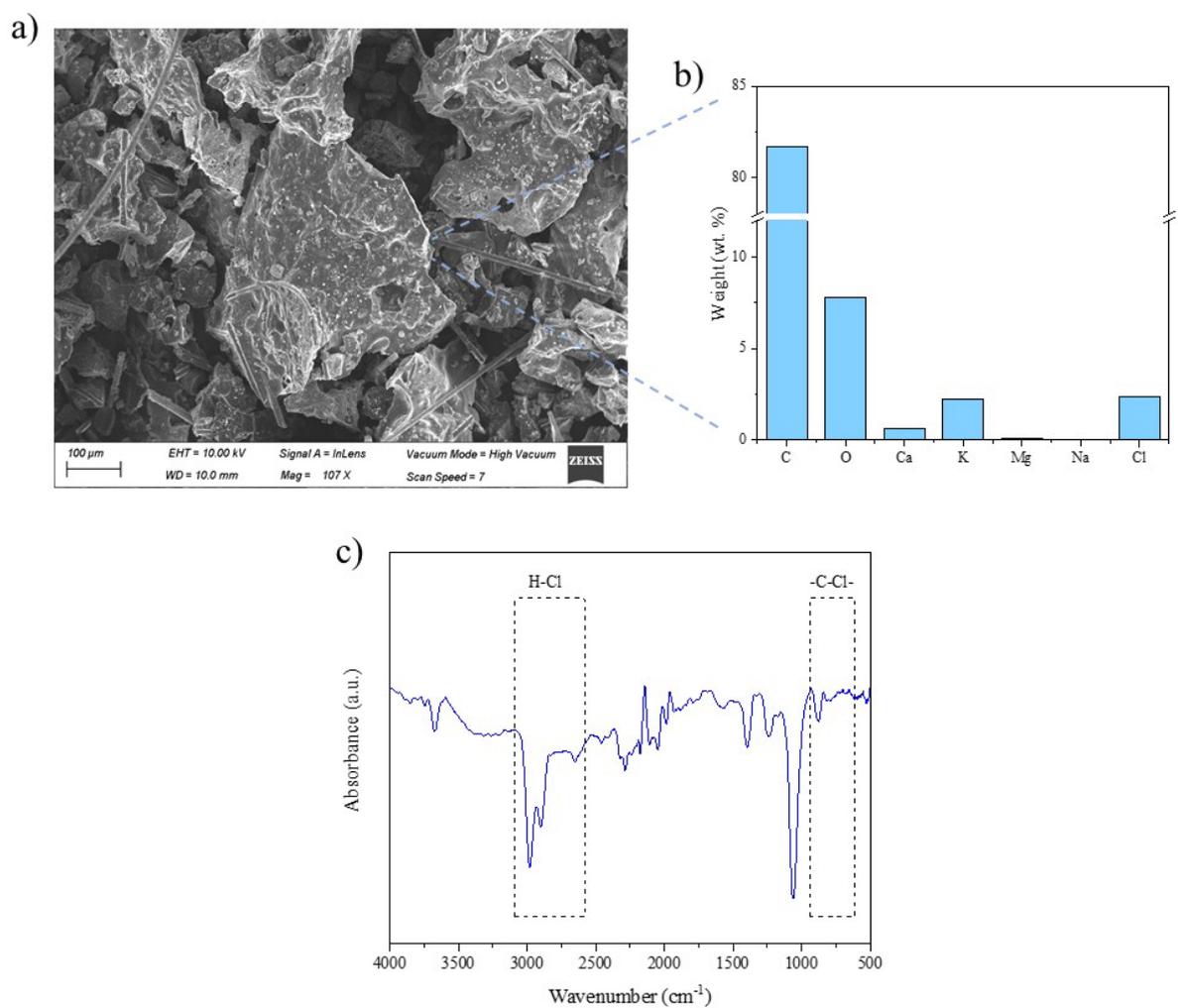


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848 Figure 6.



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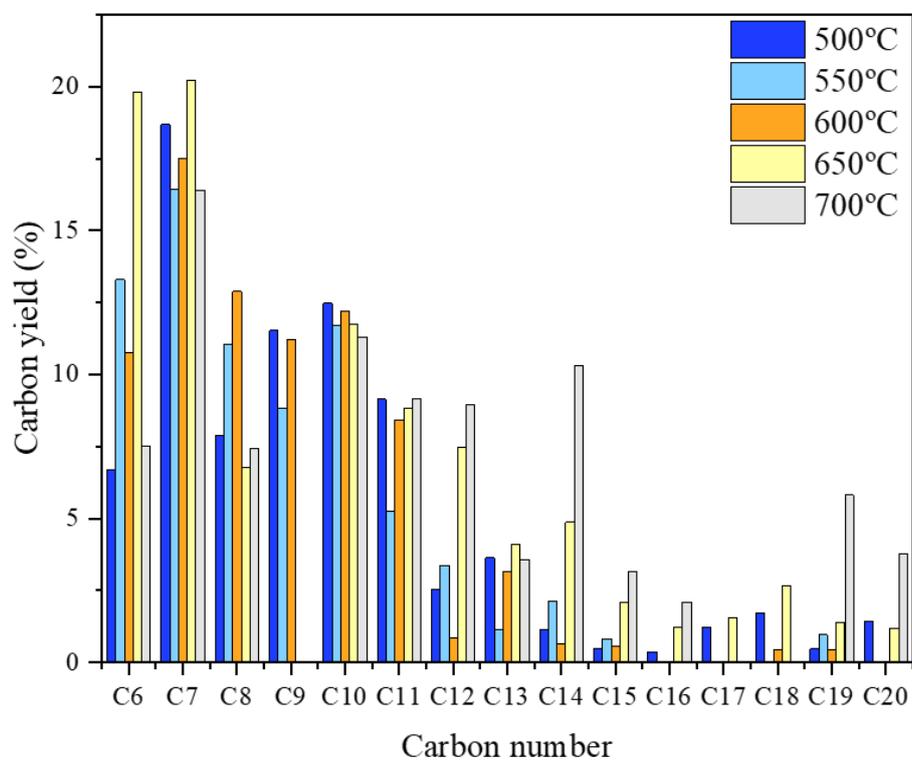
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862 Figure 7.



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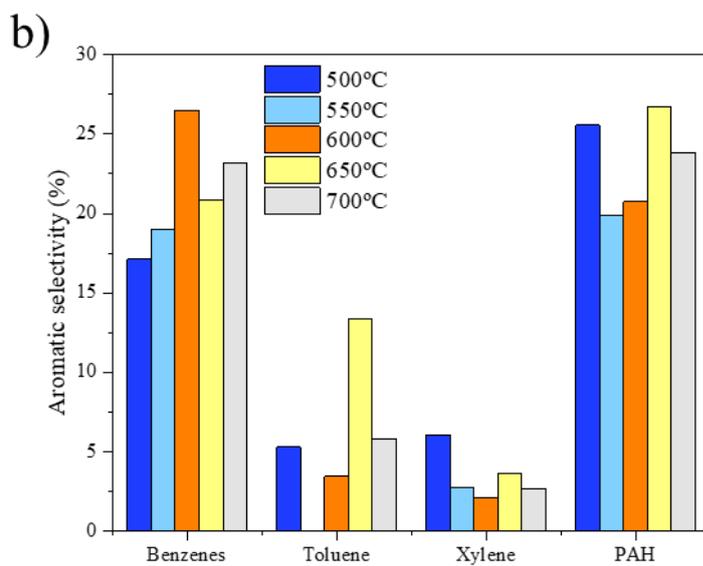
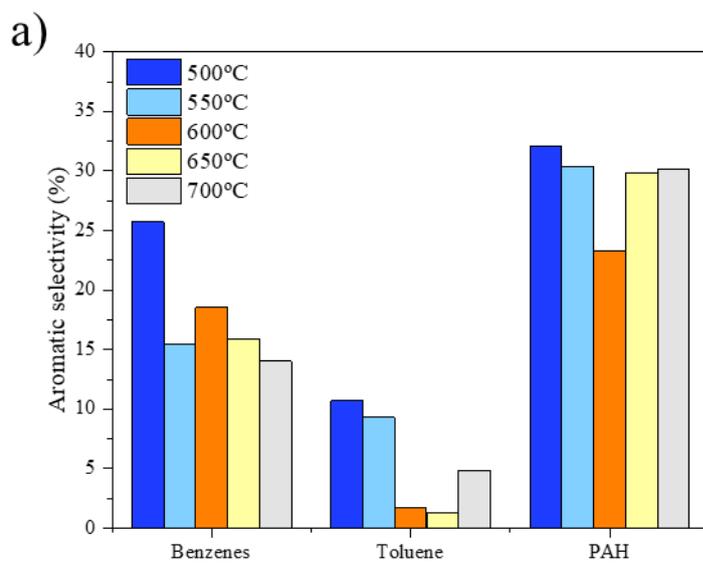
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878 Figure 8.



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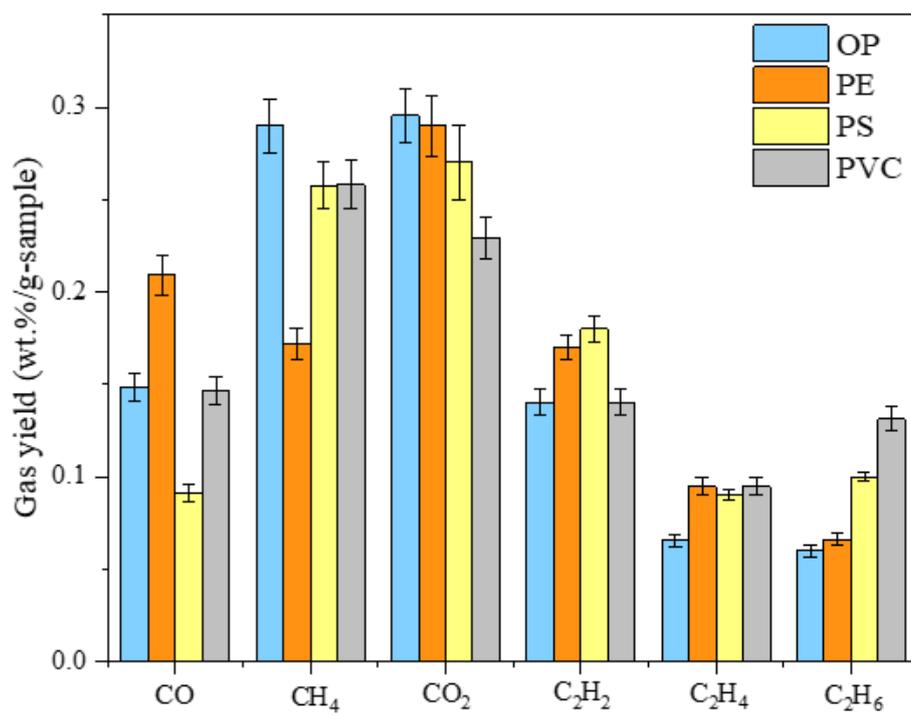
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889 Figure 9.



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