

<https://helda.helsinki.fi>

---

## Food purchase behaviour in a Finnish population : patterns, carbon footprints and expenditures

Meinilä, Jelena

2022-11

---

Meinilä , J , Hartikainen , H , Tuomisto , H L , Uusitalo , L , Vepsäläinen , H , Saarinen , M , Kinnunen , S , Lehto , E , Saarijärvi , H , Katajajuuri , J-M , Erkkola , M , Nevalainen , J & Fogelholm , M 2022 , ' Food purchase behaviour in a Finnish population : patterns, carbon footprints and expenditures ' , Public Health Nutrition , vol. 25 , no. 11 , 1368980022001707 , pp. 3265-3277 . <https://doi.org/10.1017/S1368980022001707>

---

<http://hdl.handle.net/10138/350677>

<https://doi.org/10.1017/S1368980022001707>

---

cc\_by

publishedVersion

---

*Downloaded from Helda, University of Helsinki institutional repository.*

*This is an electronic reprint of the original article.*

*This reprint may differ from the original in pagination and typographic detail.*

*Please cite the original version.*



# Food purchase behaviour in a Finnish population: patterns, carbon footprints and expenditures

Jelena Meinilä<sup>1,2,\*</sup>, Hanna Hartikainen<sup>3</sup>, Hanna L Tuomisto<sup>3,4,5</sup>, Liisa Uusitalo<sup>1</sup>, Henna Vepsäläinen<sup>1</sup>, Merja Saarinen<sup>3</sup>, Satu Kinnunen<sup>1</sup>, Elviira Lehto<sup>6</sup>, Hannu Saarijärvi<sup>2</sup>, Juha-Matti Katajajuuri<sup>3</sup>, Maijaliisa Erkkola<sup>1</sup>, Jaakko Nevalainen<sup>2</sup> and Mikael Fogelholm<sup>1</sup>

<sup>1</sup>Department of Food and Nutrition, University of Helsinki, PO Box 66, Helsinki 00014, Finland: <sup>2</sup>Tampere University, Tampere, Finland: <sup>3</sup>Natural Resources Institute Finland, Helsinki, Finland: <sup>4</sup>Department of Agricultural Sciences, University of Helsinki, Helsinki, Finland: <sup>5</sup>Helsinki Institute of Sustainability Science, University of Helsinki, Helsinki, Finland: <sup>6</sup>Department of Teacher Education, University of Helsinki, Helsinki, Finland

Submitted 22 October 2021: Final revision received 19 May 2022: Accepted 8 August 2022: First published online 18 August 2022

## Abstract

**Objective:** To identify food purchase patterns and to assess their carbon footprint and expenditure.

**Design:** Cross-sectional.

**Setting:** Purchase patterns were identified by factor analysis from the annual purchases of 3435 product groups. The associations between purchase patterns and the total purchases' carbon footprints (based on life-cycle assessment) and expenditure were analysed using linear regression and adjusted for nutritional energy content of the purchases.

**Participants:** Loyalty card holders ( $n$  22 860) of the largest food retailer in Finland.

**Results:** Eight patterns explained 55% of the variation in food purchases. The *Animal-based* pattern made the greatest contribution to the annual carbon footprint, followed by the *Easy-cooking*, and *Ready-to-eat* patterns. *High-energy*, *Traditional* and *Plant-based* patterns made the smallest contribution to the carbon footprint of the purchases. *Animal-based*, *Ready-to-eat*, *Plant-based* and *High-energy* patterns made the greatest contribution, whereas the *Traditional* and *Easy-cooking* patterns made the smallest contribution to food expenditure. Carbon footprint per euros spent increased with stronger adherence to the *Traditional*, *Animal-based* and *Easy-cooking* patterns.

**Conclusions:** The *Animal-based*, *Ready-to-eat* and *High-energy* patterns were associated with relatively high expenditure on food, suggesting no economic barrier to a potential shift towards a plant-based diet for consumers adherent to those patterns. Strong adherence to the *Traditional* pattern resulted in a low energy-adjusted carbon footprint but high carbon footprint per euro. This suggests a preference for cheap nutritional energy rather than environment-conscious purchase behaviour. Whether a shift towards a plant-based pattern would be affordable for those with more traditional and cheaper purchase patterns requires more research.

**Keywords**  
Nutrition  
Greenhouse gas emission  
Diet  
Food consumption  
Environmental impact

Reducing health risks caused by an unhealthy diet (CHD, type 2 diabetes and cancer (WHO 2013)) and reducing the carbon footprint of food consumption require changes in food consumption patterns<sup>(1)</sup> which in turn might require changes in food prices<sup>(2,3)</sup>.

Several studies based on theoretical models suggest that changing dietary habits could reduce the carbon footprint

of a diet by up to 50–80%<sup>(4,5)</sup>. Comparisons between diets such as omnivorous, vegetarian and vegan diets only partially reflect the current reality in Western societies, where the proportions of vegetarians and vegans are still low<sup>(6–8)</sup>. Furthermore, we do not know exactly what alternative diets are taking shape and what they contain. To support the climate change mitigation goals and to monitor the effects of

\*Corresponding author: Email jelena.meinila@helsinki.fi

© The Author(s), 2022. Published by Cambridge University Press on behalf of The Nutrition Society. This is an Open Access article, distributed under the terms of the Creative Commons Attribution licence (<https://creativecommons.org/licenses/by/4.0/>), which permits unrestricted re-use, distribution, and reproduction in any medium, provided the original work is properly cited.

any dietary change that is already under way, it is essential to know the carbon footprint of the current food consumption patterns beyond the rarely followed dietary patterns such as vegetarian or vegan, or national averages. In a few studies assessing real-life food consumption patterns, the differences between the carbon footprints of common self-selected dietary patterns have varied from negligible to major<sup>(9,10)</sup>.

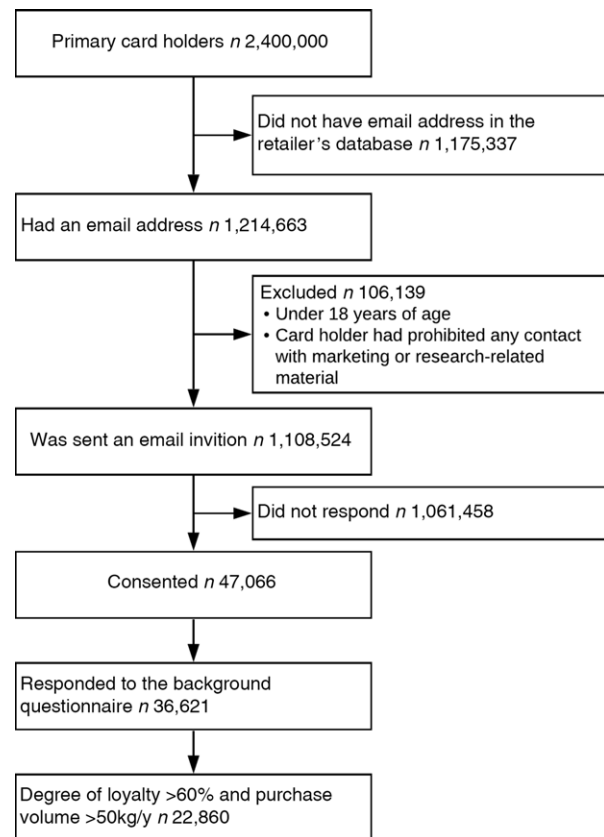
To make healthy and environmentally sustainable food available to all, reasonable pricing is important: it can make sustainable food consumption possible for households with low incomes. Several studies suggest that healthy food is more expensive than unhealthy food<sup>(11–14)</sup>. On the other hand, the prices of legumes and grains, considered both healthy and climate-friendly, can be substantially less expensive per kJ than meat<sup>(15)</sup>. Our previous study showed that plant-based protein sources were bought, on average, for a cheaper price than meat<sup>(16)</sup>. In addition, in a US study, vegetarians spent less money on food purchases than meat eaters<sup>(17)</sup>. To make healthy and sustainable foods more attractive to consumers, raising the prices of unhealthy and environmentally unsustainable foods such as red and processed meat could also be a solution<sup>(2,18)</sup>.

Food retailers' customer loyalty card data provide a unique tool for gaining insights into dietary patterns. We have previously shown that food purchase data are a valid instrument for ranking consumers according to their self-reported food<sup>(19)</sup> and beer<sup>(20)</sup> consumption. In this study, the detailed data on purchased product groups over a 1-year period enabled an objective assessment of food expenditure and the allocation of carbon footprints for large sets of product groups on a household level. This automatically accumulating data enabled relatively easy access to unusually large food purchase datasets. Thus, the aim of this study was to identify food purchase patterns by using customer loyalty card data and to study their contribution to the carbon footprints of and expenditure on total food purchases.

## Methods

### Recruitment

This study utilises large-scale loyalty card data from the largest grocery chain in Finland (S Group)<sup>(21)</sup>. The S Group sells groceries through five retail chains, which are convenience stores, supermarkets, hypermarkets, and one upper-market concept with an extended focus on high-quality and special products. The selection of food items varies between chains, from only a few thousand to over 20 000 items. The retail chains follow an 'Everyday, low-pricing' strategy (as opposed to a 'high-low pricing' strategy). At the time of the data collection (2018), 2.4 million households in Finland held the S Group's customer loyalty card, which accounted for 88 % of all Finnish households. Loyalty card holders across



**Fig. 1** Participant flow

Finland received an invitation to the study by email if they had given permission to be approached for research purposes, and if they were aged 18 years or above (Fig. 1). Those who gave their consent for the use of their purchase data for research purposes received an invitation to respond to an additional electronic questionnaire with complementary data on, for example, household structure and income<sup>(22)</sup>.

### Study sample and participants

Initial food purchase data were obtained from *n* 47 066 participants<sup>(22)</sup>. This study comprises data from the year 2018. In the questionnaire, the participants were asked to assess their degree of loyalty to the retailer (i.e. the proportion of purchases from the retailer's stores of the total food purchases of the household). Only those who had a self-reported degree of loyalty of at least 61 % (i.e. participants who made a large proportion of their food purchases from the food retailer) and who made at least 50 kg of purchases during 2018 were included in the analysis. Our previous analyses showed that purchases associated more strongly with dietary intake among the most loyal (degree of loyalty >60 %) customers<sup>(19,20)</sup> and therefore build a more complete picture of relative food purchases<sup>(22)</sup>. In this study, we used purchase data aggregated to annual consumption in both volume (kg) and expenditure (€).

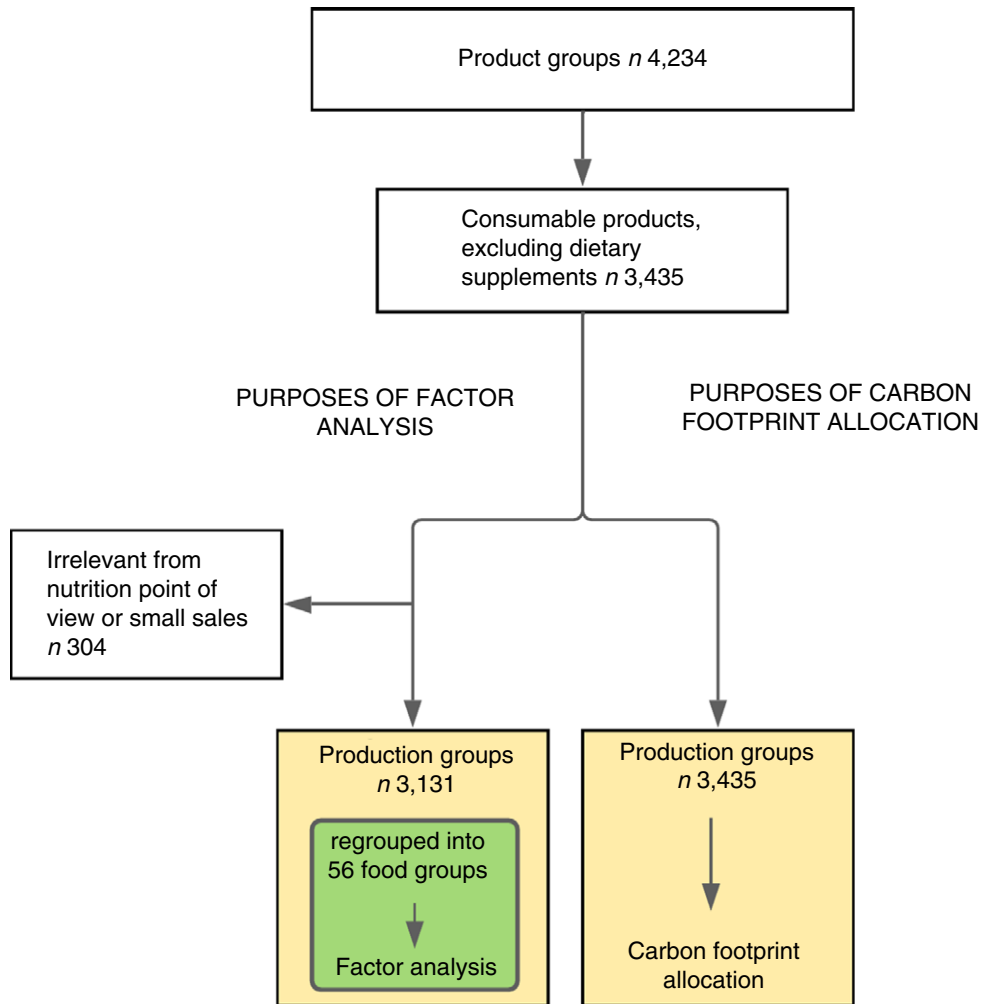


Fig. 2 Food item flow

**Background data**

The retailer’s database provided data on the sex and age of the participants. In the additional questionnaire, the participants reported their number of household members and how many of these were aged 0–6, 7–17, 18–24, 25–64 and 65 years or older. We combined the data on these two questions into a family structure variable that consisted of five categories: single-adult households, one adult and a child/children, two adults, two adults and a child/children, or other (households with three or more adults and households with an unknown family structure). The participants reported their loyalty level to the retailer by choosing from the options of <20 %, 21–40 %, 41–60 %, 61–80 % and >80 %, but as explained in the previous section, only participants in the upper two categories (61–80 %, >80 %) were included in this study.

The participants selected the monthly income of their household from five predefined categories ranging from household income less than 1500 €/month to 9000 €/month or more. Dividing the income (here, the mean of each income category) by the square root of the household size produced the monthly household income (OECD

square root scale). This income is thus presented in five categories (less than 1000 €/month, 1000–1999 €/month, 2000–2999 €/month, 3000–3999 €/month and 4000 €/month or more).

**Carbon footprint assessment**

The food retailer originally had 4234 different product groups for their products (Fig. 2). A total of 3435 product groups were assigned a carbon footprint (kg CO<sub>2</sub>-equivalent), using 1 kg of food purchased in retail as the functional unit. As carbon footprint values are not available for all the product groups, indicator products were chosen to represent the 3435 product groups, meaning that one indicator product represented several product groups. Based on the available and suitable LCA studies, we used about 100 different indicator products. As an example of the indicator product approach, all fruits were assigned the same carbon footprint, which was estimated on the basis of the weighted average of the carbon footprints of the five most sold fruits – more than 80 % of the fruits sold. Thus, the carbon footprint of an indicator product stems from

carbon footprints of several products. The method for calculating the carbon footprints of the indicator products is described in detail elsewhere (Hartikainen, Heusala, Harrison, Katajajuuri and Silvenius, unpublished results).

The main life cycle phases of the indicator products until retail were included in the system boundaries of the carbon footprint assessment, which comprised the production of inputs to agriculture, agricultural primary production, food processing, packaging, storage (before retail) and transportation. Food waste, land use changes and changes in soil carbon stocks were excluded due to a lack of data.

Data for producing the carbon footprints of the indicator products consisted of a database for food products sold in Finland, compiled by the Natural Resources Institute Finland. The database contains 170 scientific studies of the carbon footprint assessment of food products, and Natural Resources Institute Finland's LCA database is supplemented with expert estimates. For most of the indicator products, the carbon footprints were medians of the carbon footprints available in data sources. For ready-to-eat meals and beef, additional modifications were made. We determined the meal's category on the basis of the recipes of the retailer's most sold meals, using the available carbon footprints for the ingredients. The 'beef' category was determined by calculating the carbon footprints of, on the one hand, combined milk and beef production, and on the other hand, suckler beef production from the literature (medians of results from chosen studies), based on how much of the beef was sourced from the combined and suckler beef production systems<sup>(23)</sup>. The data for storing, according to the three options of dry, cold and frozen, were from the EcoInvent database<sup>(24)</sup>, and the data for packaging were from Plastics Europe<sup>(25)</sup>, the European Aluminium Association<sup>(26)</sup>, the World Steel Association<sup>(27)</sup> and the FEFCO<sup>(28)</sup>. The distance between the producer country and the logistics centres needed for the assessment of emissions from transportation were calculated based on six main production areas: (1) Finland; (2) other Nordic countries and Estonia; (3) the rest of Europe; (4) America; (5) Africa and the Middle East; and (6) Asia. The emission factors for transportation were taken from the Lipasto database<sup>(29)</sup>, except for the trans-oceanic container ship, which was from the EcoInvent database<sup>(24)</sup>.

After the carbon footprints of the indicator products had been determined for the retailer's product groups, the purchase volume of each loyalty card holder for each product group (kg) was multiplied by the corresponding carbon footprint (kg CO<sub>2</sub>-eq) to obtain customers' product group-specific and total-purchase carbon footprints. The sum of the carbon footprints of all the product groups represents the carbon footprint of the total food purchases per person.

### **Nutritional energy content of the purchases**

Throughout the text, energy refers to the nutritional energy content of the food purchases (not, e.g., energy utilised in

food production). The energy content of 1 kg of each product group (e.g. cucumber, skimmed milk and vegetarian lasagna) was derived from the nutrition calculation software on [www.fineli.fi](http://www.fineli.fi). This webpage utilises the food composition database Fineli which is maintained by the Finnish Institute for Health and Welfare. The purchase volume (kg) of each product group was multiplied by the energy content per 1 kg of the group to obtain the absolute energy content of the purchase. The energy contents of all the purchased product groups were summed to obtain the annual energy content of the total purchases.

### **Grouping of food purchase data for factor analysis**

For factor analysis, a major regrouping was conducted, based on the purpose of use (combined fresh vegetables such as cucumber, tomato etc., as fresh vegetables; soya milk, soya yoghurt, oat milk, oat yoghurt etc., as different plant-based dairy alternatives, etc.). The aggregation of food groups was restricted to a level that enabled differentiation on the basis of nutritional content and carbon footprint. This was driven by the differences in nutrient content that are relevant for public health in Finland and that are reflected in the food-based dietary guidelines of the Nordic Nutrition Recommendations (e.g. separating high-fibre from low-fibre breads and high-fat from low-fat dairy), the degree of processing (e.g. separating fresh potato from frozen potato) and the carbon footprints of the product groups (e.g. separating meat types such as beef, pork and poultry) (see online Supplemental Table 1). Examples of the aggregated food groups and product groups included were as follows: 'skimmed milk and sour milk': regular, low-lactose, and lactose-free skimmed milk and skimmed sour milk, and 'sugar-sweetened beverages': soft drinks, energy drinks, juices, ice teas and seasonal drinks. Product groups that were not relevant to overall diet quality (e.g. tea, bottled water, chewing gum and spices) or product groups with very low purchases (e.g. game, reindeer and horse meat) were excluded. The final number of food groups to be used in factor analysis was 56.

### **Statistical methods**

Participants' characteristics are presented as means and standard deviations, or frequencies and percentages.

To estimate the total household food purchases, we multiplied the volume of total purchases (kg) by the inverse of the self-reported degree of loyalty, for which we used the midpoints of the category intervals. Deviation from normal distribution, detected by visual inspection of their empirical distributions, led to the logarithmic transformation of the food group variables, the energy content of the total food purchases, the carbon footprint (CO<sub>2</sub>-eq. values) of and expenditure (€) on total food purchases and the carbon footprint to expenditure ratio. Before the subsequent factor analysis on food purchases measured in kilograms, we



performed a 98 % winsorisation of the food group variables to diminish the effect of outliers, that is, outliers below the 1st percentile and above the 99th percentile were truncated into the 1st and 99th percentile, respectively.

Bartlett's test of sphericity ( $\chi^2(1540) = 567\,540$ ,  $P < 0.001$ ) suggested the appropriateness of the factor analysis and the Kaiser–Meyer–Olkin test ( $KMO = 0.96$ ) indicated good sampling adequacy. Food purchase patterns were derived from principal component analysis, based on the correlation matrix of food groups<sup>(30)</sup>. The number of principal components was decided by simultaneously examining the scree plot (see online Supplemental Fig. 1), the Kaiser criterion (eigenvalue  $> 1$ ), the percentage of explained variation (our aim was  $> 50\%$ ) and the interpretability of the factors. We chose eight components and used an orthogonal varimax rotation to produce the final factors, from which we then identified and named the food purchase patterns. All the participants were assigned standardised factor scores to represent food purchase patterns, that is, weighted combinations of the purchased food groups. The pattern scores showed how strongly empirically derived purchase patterns (and the food groups defining it) were reflected in a participant's shopping basket; the higher the score, the stronger the adherence to the purchase pattern. As customary in nutrition research, the patterns were named based on a feature that was common for the food groups that had high loadings for the factor and that separated the factor from the other factors.

The associations between the food purchase patterns and carbon footprint or expenditure were analysed using linear regression analysis with purchase pattern scores as explanatory variables and log-transformed carbon footprints or log-transformed expenditure as the response variable. The model had one purchase pattern at a time and the log-transformed energy content of the total purchases as explanatory variables, meaning that each pattern was analysed separately. Thus, as the analyses were adjusted for the (log-) energy content, the regression coefficients can be interpreted as the difference between the log-carbon footprint or log-expenditure of two individuals with the same energy content of the total purchases but a unit's (SD) difference in their purchase pattern. To illustrate the magnitude of the effects of the patterns in a more perceivable manner, we calculated the estimated carbon footprint and expenditure in the lowest and highest thirds and in the lowest and highest 10 % of the pattern scores of each pattern using the regression equation:

$$Y = \exp[\alpha + \beta_1 \times Q + \beta_2 \times T],$$

where Y is either carbon footprint or expenditure,  $\alpha$  is the intercept term,  $\beta_1$  equals the regression coefficient of the pattern score, Q equals the mean of the pattern score in a given quantile (lowest third, highest third, lowest 10 % or highest 10 %) and  $\beta_2$  equals the regression coefficient of the log-transformed energy content at its mean value (T).

The associations between the patterns and the ratio of carbon footprint (kg CO<sub>2</sub>-eq.) to expenditure (€) were analysed using simple regression analysis, of which the log-transformed carbon footprint:expenditure ratio was the outcome and one pattern at a time an explanatory variable. The patterns that could not be considered overall dietary patterns, namely *Skimmed milk and margarine* and *Alcohol*, were excluded from further analyses.

To gain further insights into the association between purchase patterns and carbon footprint and expenditure, we calculated (1) the individual food groups' sum of the annual carbon footprint (kg CO<sub>2</sub>-eq.), (2) the percentage of each food group's carbon footprint of the annual total, (3) the sum of annual expenditure (€) on the individual food groups, (4) the percentage of expenditure on each food group of the total annual expenditure, and (5) the ratio of the annual carbon footprint and the annual expenditure (see online Supplemental Table 2).

### Sensitivity analyses

To investigate the sensitivity of different decisions to the result of the factor analysis, we conducted factor analysis with several different choices: (1) included only participants with an  $\geq 80\%$  degree of loyalty; (2) no winsorisation of the food variables before factor analysis and (3) 5 % winsorisation of the food variables before factor analysis. The results of these factor analyses were similar to the one presented; the same patterns with similar explained variations were identified. Therefore, these results are not shown.

## Results

The majority of the participants were women (66 %) (Table 1). A two-adult household was the most common family structure (34 %), followed by single-adult households (25 %), and two adults with a child/children (23 %). The majority of the households fell into the scaled monthly income range of 2000–2999 € or 3000–3999 € (29 % and 23 %, respectively), and the majority (61 %) bought 81–100 % of all of their food purchases from the retailer.

### Purchase patterns

Eight factors were derived, which explained altogether 55 % of the variation of the fifty-six food groups (Fig. 3). In descending order of explained variation, the patterns were named *Traditional* (11.0 % of variation), *High-energy* (9.7 %), *Plant-based* (8.0 %), *Animal-based* (8.0 %), *Ready-to-eat* (5.7 %), *Easy-cooking* (3.9 %), *Skimmed milk and margarine* (3.7 %) and *Alcohol* (3.3 %). Figure 3 shows the food group loadings in each of the patterns.

**Table 1** Background characteristics of participants (n 22 860)

|  | n      | %          |
|--|--------|------------|
| Sex  |        |            |
| Men  | 7745   | 33.9       |
| Women  | 15 115 | 66.1       |
| Age  | Mean   | SD         |
|  | 47.9   | 15.2       |
|  | n      | %          |
| Family structure   |        |            |
| Single-adult households  | 5717   | 25.0       |
| One adult and a child/children   | 1040   | 4.5        |
| Two adults   | 7875   | 34.4       |
| Two adults and a child/children  | 5272   | 23.1       |
| Other  | 1707   | 7.5        |
| Missing  | 1249   | 5.5        |
| Scaled household income (€/month)*   |        |            |
| Less than 1000   | 1967   | 8.6        |
| 1000–1999  | 3398   | 14.9       |
| 2000–2999  | 6633   | 29.0       |
| 3000–3999  | 5147   | 22.5       |
| 4000 or more   | 4237   | 18.5       |
| Missing  | 1478   | 6.5        |
| Percentage of purchases from retailer  |        |            |
| 61–80 %  | 8978   | 39.3       |
| 81–100 %   | 13 882 | 60.7       |
|  | Median | IQR        |
| CO <sub>2</sub> of total food purchases from S Group (kg CO <sub>2</sub> -eq/year) | 2750   | 1676, 4339 |
| Total purchase volume (kg/year)  | 733    | 463, 1129  |
| Energy content of the total purchases (MJ/year)                                    | 4244   | 2721, 6506 |
| Expenditure on food (€/year)   | 3141   | 2015, 4691 |

\*Income (here the mean of each income category) divided by the square root of the household size produced (OECD square root scale).

### Carbon footprint of the purchases and association with purchase patterns

An investigation of the food groups behind the patterns showed that 12 % of the total carbon footprint of the purchases originated from beef and processed beef and 9 % from cheese (see online Supplemental Table 2). Both food groups were strongly loaded in the *Animal-based* pattern (factor loading for beef and processed beef 0.65). The third largest food group that contributed to the total carbon footprint was fresh vegetables (6 %), which loaded strongly in the *Plant-based* (factor loading 0.61) and *Animal-based* (0.51) patterns. In contrast, peas, beans and lentils, which loaded strongly (0.73) in the *Plant-based* pattern, made only a small contribution to the total carbon footprint (0.4 %). All the meat food groups together contributed to 29 % of the total carbon footprint, all the dairy food groups to 28 %, and all the vegetables, fruits and berries to 12 % of the total carbon footprint.

When adjusted for the energy content of the annual purchases, the difference in the carbon footprint (log-kg CO<sub>2</sub>-eq.) of 1 SD difference in pattern scores was the largest for the *Animal-based* pattern ( $\beta$  0.134, 95 % CI (0.132, 0.137)), followed by *Easy-cooking* ( $\beta$  0.039, 95 % CI (0.036, 0.042))

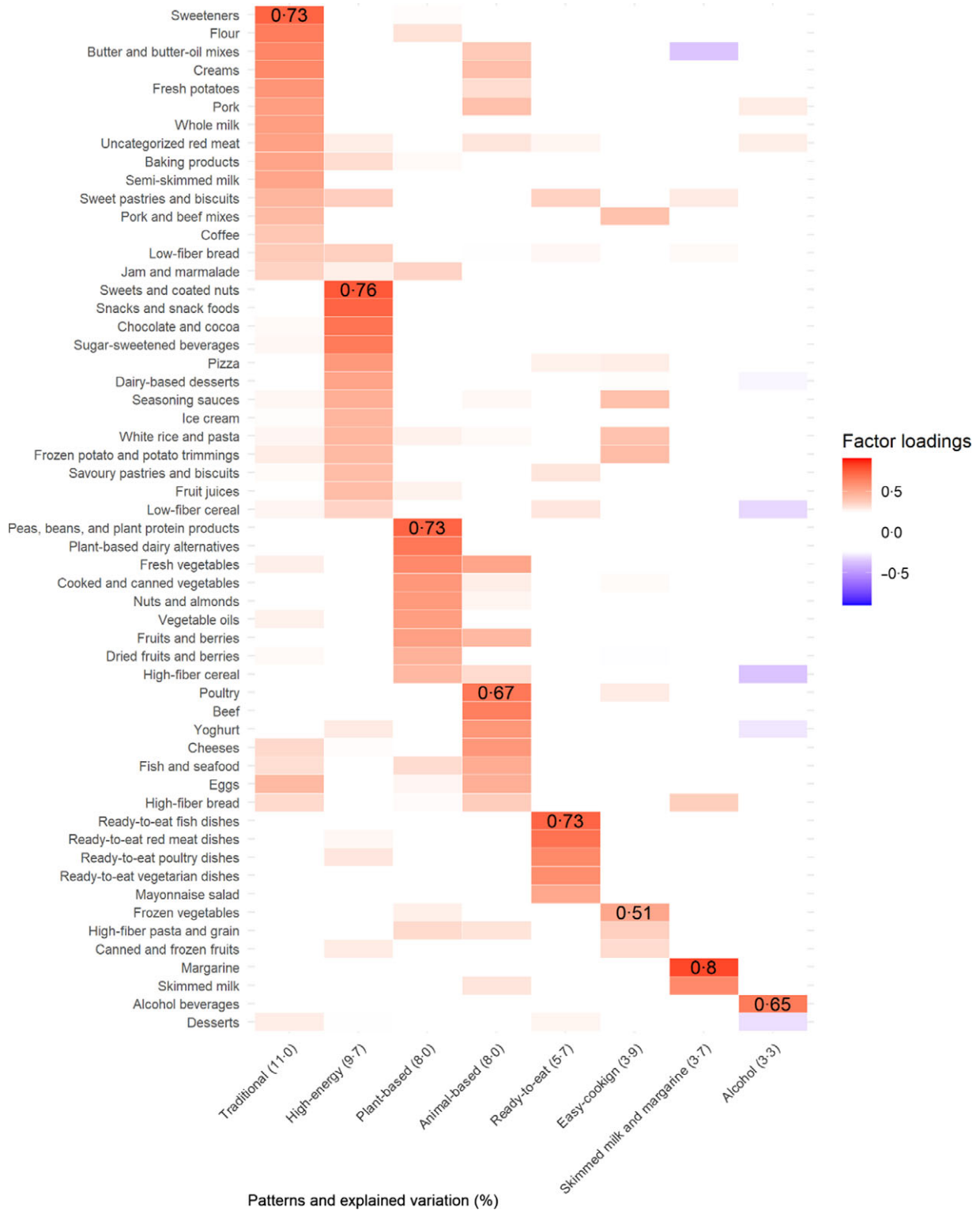
and *Ready-to-eat* ( $\beta$  0.016, 95 % CI (0.013, 0.019)) (Table 2). For the *High-energy* ( $\beta$  -0.032, 95 % CI (-0.035, -0.029)), *Traditional* ( $\beta$  -0.036, 95 % CI (-0.039, -0.032)) and *Plant-based* patterns ( $\beta$  -0.047, 95 % CI (-0.050, -0.044)), the relationship was inverse; a 1 SD higher *Plant-based* score was associated with a significant decrease in the carbon footprint of the total purchases. In other words, the *Animal-based*, *Easy-cooking* and *Ready-to-eat* patterns were positively associated, and the *Plant-based*, *Traditional* and *High-energy* patterns were inversely associated with the carbon footprint of the total purchases.

The comparison of the highest 10 % of *Animal-based* v. *Plant-based* revealed a + 869 kg CO<sub>2</sub>-eq annual difference, which means a 27 % lower annual food purchase carbon footprint among those strongly adhering to the *Plant-based* pattern than among those strongly adhering to the *Animal-based* pattern. Those with the highest 10 % of *Easy-cooking* and *Ready-to-eat* scores had a 14 % and 17 % smaller carbon footprint, respectively, than those in the highest 10 % of the *Animal-based* scores, but *Easy-cooking* and *Ready-to-eat* had a 17 % and 13 % larger carbon footprint, respectively, than those with the highest 10 % of the *Plant-based* scores. The carbon footprint of those in the highest 10 % of *High energy* and in the highest 10 % of *Traditional* was close to that of the *Plant-based* pattern, only 4 % and 3 % higher, respectively.

### Food expenditure and association with purchase patterns

An investigation of the food groups behind the patterns showed that of the food groups, expenditure was highest on cheese (7 %), fresh vegetables (7 %), fruits and berries (6 %), alcohol beverages (6 %), and yoghurt (5 %) (see online Supplemental Table 2). Cheese and yoghurt were strongly loaded in the *Animal-based* pattern, whereas fresh vegetables, and fruits and berries were strongly loaded in the *Plant-based* and *Animal-based* patterns. Peas, beans and lentils made up only 0.65 % of the total food expenditure.

When adjusted for the annual energy content of the purchases, all the purchase patterns were associated with the total expenditure on food (log-€) (Table 3). The regression indicated a positive correlation for all patterns except those of *Traditional* ( $\beta$  -0.115, 95 % CI (-0.120, -0.111)) or *Easy-cooking* ( $\beta$  -0.029, 95 % CI (-0.033, -0.025)), for which the correlations were inverse. The change in the expenditure on food purchases by a 1 SD increase in the food purchase pattern score (adjusted for total energy content of the purchases) was the largest and inverse in the *Traditional* pattern and the second, third, and fourth largest in the *Animal-based* ( $\beta$  0.064, 95 % CI (0.059, 0.068)), *Ready-to-eat* ( $\beta$  0.063, 95 % CI (0.059, 0.066)) and *Plant-based* ( $\beta$  0.055, 95 % CI (0.051, 0.059)) patterns, but in the opposite direction to that of *Traditional*. Comparison of the highest 10 % of the *Traditional* v. *Animal-based*



**Fig. 3** Illustration of rotated principal components' loading matrix of food purchase patterns. The values in the tiles represent the largest factor loading within each pattern. The percentages of explained variances for the factors are in parenthesis after the pattern names under the x-axis

patterns revealed a + 886 € annual difference between the patterns, which means a 27 % lower annual expenditure among those in the highest 10 % of the *Traditional* pattern

*v.* those in the highest 10 % of the *Animal-based* pattern. The difference between those in the highest 10 % of *Traditional* and *Plant-based* was similar.



**Table 2** Regression coefficients ( $\beta$ ) and 95 % CI for association between food purchase patterns and log-transformed annual carbon footprint with energy from the purchases (MJ) at its annual mean level, and predicted carbon footprint (kg CO<sub>2</sub>-eq/year) in the lowest (T1) and highest thirds (T3), and lowest (D1) and highest deciles (D10) of each purchase pattern

| Pattern      | $\beta$ | 95 % CI |        | D1   | T1   | T3   | D10  |
|--------------|---------|---------|--------|------|------|------|------|
|              |         | Lower   | Upper  |      |      |      |      |
| Animal-based | 0.134   | 0.132   | 0.137  | 2119 | 2225 | 2970 | 3218 |
| Easy-cooking | 0.039   | 0.036   | 0.042  | 2437 | 2471 | 2691 | 2761 |
| Ready-to-eat | 0.016   | 0.013   | 0.019  | 2529 | 2539 | 2625 | 2663 |
| High energy  | -0.032  | -0.035  | -0.029 | 2701 | 2671 | 2488 | 2437 |
| Traditional  | -0.036  | -0.039  | -0.032 | 2710 | 2680 | 2477 | 2415 |
| Plant-based  | -0.047  | -0.05   | -0.044 | 2743 | 2706 | 2446 | 2349 |

### Relationship between carbon footprints and expenditures

When the food groups behind the patterns were examined separately, the largest carbon footprint per euro was for beef and processed beef (2.6), followed by pork and beef mixes (2.1), skimmed milk and sour milk (1.3), butter and butter-oil mixes (1.2), and semi-skimmed milk and sour milk (1.1) (see online Supplemental Table 2). In contrast, peas, beans and lentils had small carbon footprints per euro (0.34). Pork and beef mixes, both milk food groups, and butter-oil mixes loaded strongly in the *Traditional* pattern.

When adjusted for the energy content of the annual purchases, total expenditure (log-€) was positively associated with total carbon footprint (log-kg CO<sub>2</sub>-eq.,  $\beta$ : 0.36, 95 % CI (0.35, 0.37)). Figure 4 shows the relationships between the patterns and the ratio of carbon footprint and expenditure. Stronger adherence to the *Traditional* ( $\beta$ : 0.048, 95 % CI (0.047, 0.050)), *Animal-based* ( $\beta$ : 0.042, 95 % CI (0.040, 0.044)) and *Easy-cooking* ( $\beta$ : 0.037, 95 % CI (0.036, 0.039)) patterns were associated with a higher carbon footprint per spent euro, whereas stronger adherence to the *High-energy* ( $\beta$ : -0.006, 95 % CI (-0.007, -0.004)), *Ready-to-eat* ( $\beta$ : -0.011, 95 % CI (-0.013, -0.010)) and *Plant-based* ( $\beta$ : -0.029, 95 % CI (-0.030, -0.027)) patterns were associated with a lower carbon footprint per spent euro.

### Discussion

We identified eight food purchase patterns, from which we further analysed the six that explained most of the variation. Of all the patterns, the *Animal-based* explained the carbon footprint of total food purchases the most, that is, the purchases of those strongly adhering to the *Animal-based* had the largest carbon footprint, followed by *Easy-cooking* and *Ready-to-eat* patterns. As expected, the purchases of those who adhered strongly to the *Plant-based* pattern had the smallest carbon footprint. Those who adhered strongly to the *Animal-based*, *Ready-to-eat* and *Plant-based* patterns spent the most money on food, whereas those who adhered strongly to the *Traditional* pattern spent the least money on food. Stronger adherence to the *Traditional*,

*Animal-based* and *Easy-cooking* patterns was associated with a larger carbon footprint per euro spent. This was because these patterns had high loadings of animal-based food groups, which had high carbon footprint to expenditure ratios.

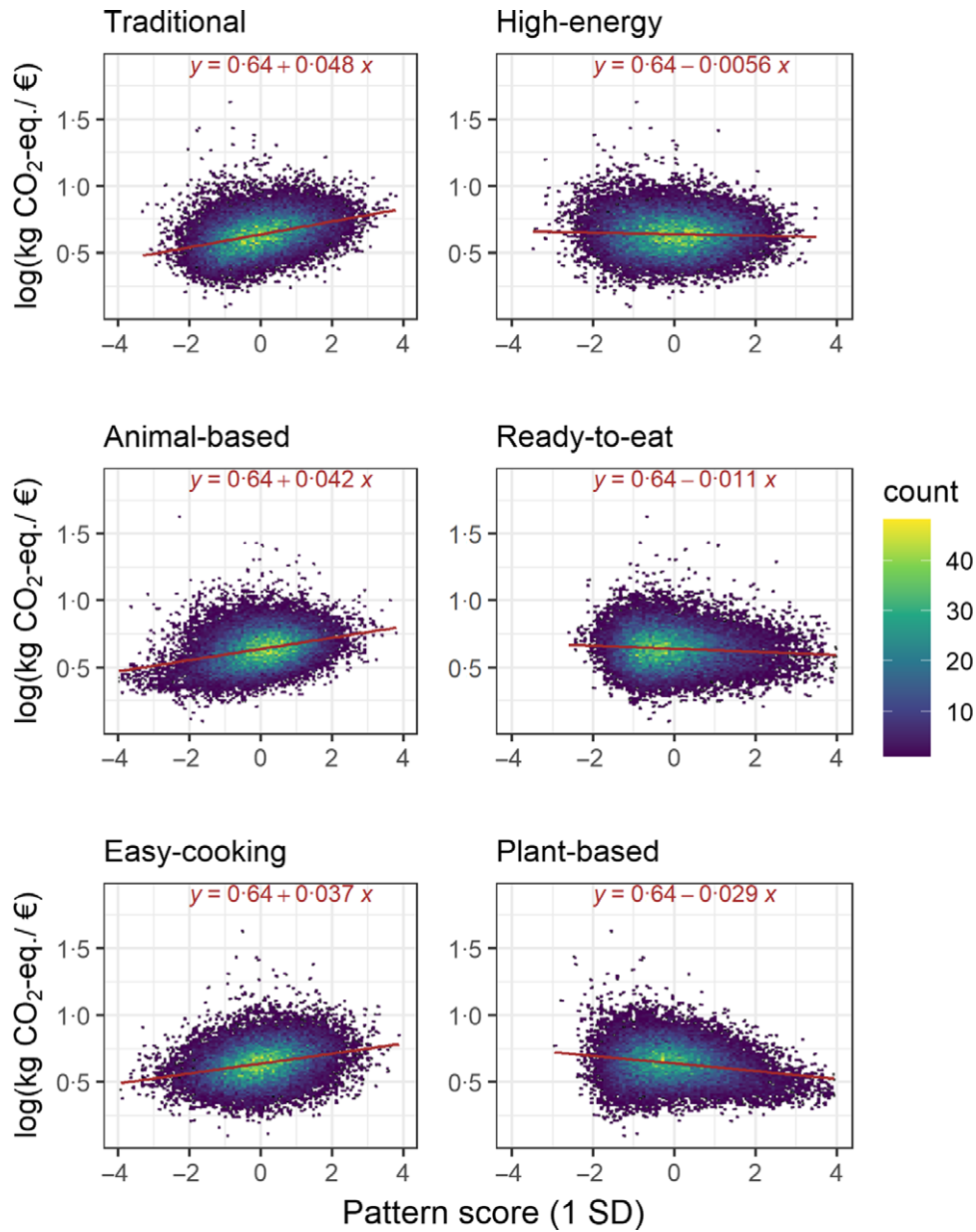
Prior research has used alternative methods such as bar-code scanning to study purchase patterns<sup>(31–33)</sup>, and only one has used customer loyalty card data<sup>(34)</sup>. Because loyalty card data accumulate without any effort required from the participant and is provided by the retailer instead of the customer/participant, they are possibly more objective than data collected by participants using bar-code scanning. Automated accumulation of data without much effort from researchers or participants enables data from a larger number of participants than data collected by participants using bar-code scanning. A detailed comparison of our study and previous studies of purchase patterns *per se* is not feasible because the countries of these studies have different food cultures (Finland, UK, Germany and USA), the analysed food groups consist of different foods and the analytical methods are different. However, one German<sup>(33)</sup> and one US study<sup>(31)</sup> found patterns similar to our *Traditional* pattern, characterised by high loadings for vegetables, fruits, potatoes, high-fat milk, and high-fat meat, and the US study<sup>(31)</sup> found a pattern characterised by ready-to-eat meals, and a pattern characterised by sweets, snacks and deserts. These patterns resembled our *Ready-to-eat* and *High-energy* patterns, respectively.

A few studies have analysed carbon footprints by using data-driven food consumption patterns, but not actual purchase data<sup>(10,35,36)</sup>. Comparing our study to somewhat similar studies is challenging because of differing food consumption assessment methods (dietary intake *v.* food purchases), differing covariates in the models (e.g. energy adjustment), varying LCA methodologies, the individual method choices of assessing CO<sub>2</sub>-eq values (allocations, system boundaries, etc.), and different food production conditions and practices. All these lead to the studies having different carbon footprints. It is also important to note that unlike dietary intake data, food purchase data also include foods that end up in household food waste (an advantage when studying the environmental impacts of food consumption). The direction of the results of the most



**Table 3** Regression coefficients and 95 % CI for association between food purchase patterns and log-transformed annual expenditure on food (€) with energy from the purchases (MJ) at its annual mean level, and predicted expenditure (€) in the lowest (T1) and highest thirds (T3), and lowest (D1) and highest deciles (D10) of each purchase pattern

| Pattern      | $\beta$ | 95 % CI |        | D1   | T1   | T3   | D10  |
|--------------|---------|---------|--------|------|------|------|------|
|              |         | Lower   | Upper  |      |      |      |      |
| Animal-based | 0.064   | 0.059   | 0.068  | 2685 | 2748 | 3151 | 3273 |
| Ready-to-eat | 0.063   | 0.059   | 0.066  | 2730 | 2770 | 3164 | 3353 |
| Plant-based  | 0.055   | 0.051   | 0.059  | 2742 | 2786 | 3134 | 3285 |
| High energy  | 0.029   | 0.024   | 0.033  | 2828 | 2855 | 3042 | 3098 |
| Easy-cooking | -0.029  | -0.033  | -0.025 | 3072 | 3041 | 2855 | 2801 |
| Traditional  | -0.115  | -0.12   | -0.111 | 3458 | 3336 | 2588 | 2387 |



**Fig. 4** Relationship between the purchase patterns and the log-transformed ratio of carbon footprint (kg CO<sub>2</sub>-eq.) and expenditure (€)

similar study to ours<sup>(35)</sup>, however, resembled the direction of ours: a food consumption pattern characterised by a high consumption of meat was associated with a larger carbon footprint than the patterns characterised by less consumption of meat. In the same study, the carbon footprints of the patterns characterised by less meat consumption, such as a plant-based healthy pattern (Lebanese-Mediterranean pattern) and a pattern with high loadings for high-energy/low-nutrient foods, were not fundamentally different to each other. This was similar to our result showing that the carbon footprints of *High-energy* and *Plant-based* patterns differed only a little.

The small size of the carbon footprint related to strong adherence to the *High-energy* pattern, which explained the variation the second most, can be explained by its high loadings for only plant-based foods. These foods, however, were typical ultra-processed foods<sup>(37)</sup>, which are high in sugar, saturated fat and energy, and low in fibre and micronutrients<sup>(38,39)</sup>. The consumption of ultra-processed foods is associated with obesity and other non-communicable diseases<sup>(40–42)</sup>, although it has not been conclusively shown that these adverse health effects are due to ultra-processing *per se*. Thus, despite a small carbon footprint, a *High-energy* pattern is not recommendable as a sustainable alternative to patterns with a large carbon footprint.

The relative increase in the *Ready-to-eat* score was associated with only a moderate increase in the carbon footprint of total food purchases. This was probably because the *Ready-to-eat* pattern is a mixture of animal- and plant-based foods (vegetarian, red meat, poultry and fish), and in Finland, the red meat alternatives of ready-to-eat meals usually contain relatively small quantities of meat. Data on the carbon footprints of ready-to-eat meals are scarce, however, and the results of earlier studies vary. In a Finnish study, ready-to-eat meals had a smaller carbon footprint than home-cooked equivalents, owing to raw material selection in ready-to-eat meals<sup>(43)</sup>. In contrast, in a UK study, ready-to-eat meals had a greater carbon footprint than equivalent home-cooked meals, mainly due to higher waste production during the processing phase<sup>(44)</sup>. Our data on ready-to-eat meals are based on the scarce available LCA data, which is why any interpretation of the carbon footprint of *Ready-to-eat* pattern requires caution. More data on the carbon footprints of ready-to-eat meals are clearly required.

According to nutrition recommendations, the food groups to be consumed the most are fruits, vegetables and high-fibre grains<sup>(45)</sup>. Furthermore, given the climate mitigation goals, consumption and purchases of plant-based products should be much more common, particularly if it leads to a reduction in meat consumption. Previous studies suggest that a plant-based sustainable diet is not affordable for everyone, especially in low- and middle-income countries<sup>(46,47)</sup>. Compared to other food groups, fruits and vegetables were expensive, whereas starchy staple foods (e.g. wheat flour, potatoes and rice)

were the least expensive in all regions of the world<sup>(46)</sup>. Previous studies, however, have not extensively investigated the expenditure on food of those adhering to plant-based consumption patterns in developed countries.

In terms of reducing the carbon footprints of the households in our study sample, a shift from *Animal-based and Easy-cooking* patterns towards *Plant-based* pattern would be beneficial. Those adhering strongly to the *Animal-based* pattern (highest 10% of the total pattern score) spent similar amount of money on food as those adhering strongly to the *Plant-based* pattern, which suggests a lack of economic barrier for a necessary shift towards a plant-based pattern. Those adhering strongly to *Easy-cooking* spent 484 €/year less on food than those adhering strongly to the *Plant-based* pattern. The carbon footprint of those adhering strongly to the *Traditional* pattern was not much larger than that of those strongly adhering to the *Plant-based* pattern, and more plant-based food choices would improve the nutritive value of their purchases. The difference between the expenditures of those adhering strongly to the *Traditional* and those adhering strongly to the *Plant-based* pattern was great – 898 €/year. However, it is worth noting that purchase data are based on actual expenditures. They do not represent the cheapest or most expensive selections. Therefore, cheaper, healthy plant-based food baskets are probably available. What they would contain and on what terms they would appeal to consumers requires more research. Thus, our results suggest that a shift from the *Animal-based* to the *Plant-based* pattern should not be considered as an economic issue. An economic barrier to shifting from the *Traditional* and *Easy-cooking* to a healthy plant-based pattern would be an important research topic.

Those strongly adhering to *High-energy* foods had a relatively high expenditure on food. Because of the unhealthy characteristics of ultra-processed foods, the health authorities usually consider ultra-processed foods, which are typically considered cheap, a threat to the health of the lowest income households in particular<sup>(30)</sup>. Our results suggest that the expenditure on food of at least those who adhered strongly to the *High-energy* pattern was near the average among the loyalty card holders. High palatability, affordability, convenience (often sold as ready-to-consume) and effective marketing may also increase the purchases of *High-energy* foods among those with higher expenditure. Even more so than *High-energy*, the *Ready-to-eat* pattern was not associated with low expenditure on food, which suggests that aiming for low expenditure is not the main motive of purchasing ready-to-eat meals. Our results thus suggest that a shift from *Ready-to-eat* pattern to a *Plant-based* pattern might not be an economic issue. This is supported by a previous study, which showed that convenience was an important food motive among Finnish consumers, especially among younger individuals and households with adults and children<sup>(48)</sup>. Based on these results, one way to acknowledge the convenience motive



but to improve healthiness and reduce the carbon footprint of food consumption could be to increase the availability of healthy plant-based ready-to-eat meals with small carbon footprints.

In our analyses of the carbon footprint per euro in relation to the patterns, the most important results were those of the *Traditional* and *Ready-to-eat* patterns. Although those adhering strongly to the *Traditional* pattern had a lower carbon footprint than the other patterns, their carbon footprint per euro was high. This may indicate primarily cheap-energy-driven rather than environment-conscious purchase behaviour, resulting in an unintentional outcome of a smaller carbon footprint than among the others for the same total energy content. For those strongly adhering to the *Ready-to-eat*, the carbon footprint per euro was somewhat low, which is logical because they had an average carbon footprint but high expenditure on food. The carbon footprints per euro in relation to the other patterns were mostly in line with the energy-adjusted carbon footprints related to the patterns; those with a high carbon footprint, *Animal-based* and *Easy-cooking*, also had a high carbon footprint per euro and those with a low carbon footprint, *Plant-based* and *High-energy*, had a low carbon footprint per euro.

This study has several strengths. The potential of customer loyalty card data for investigating food purchase behaviour patterns has remained largely unexplored. Unlike self-reported food consumption data, customer loyalty card data do not suffer from recall bias or under- and over-reporting. Extensive purchase data are obtainable without substantially burdening participants or researchers. As the data covered an extensive time period, they were more accurate regarding habitual consumption and the inclusion of the seasonal variation of food consumption. We have also previously shown that food purchase data are a valid instrument for ranking consumers according to their self-reported food<sup>(19)</sup> and beer<sup>(20)</sup> consumption. Our data included detailed information on thousands of product groups, which enabled regrouping based on the principles most appropriate for the purpose. No studies have analysed carbon footprints or the expenditure of data-driven food purchase patterns. Conclusively, our study is among the first to display how customer loyalty card data can be used to identify and assess purchase patterns and the associated carbon footprints and expenditure.

Some uncertainties regarding the data are worth discussing. Even though most of the purchases were bought from the retailer in question, some foods may have been bought from different retailers, especially by those who reported buying only 61–80% of their groceries from the retailer. However, our sensitivity analysis showed that the purchase patterns were similar when only those who bought  $\geq 81\%$  of their food from the retailer were included in the factor analysis. This is in line with our previous finding that the proportions of food groups purchased were very similar among customers with high loyalty<sup>(22)</sup>, and that their purchases reflect the loyalty card holder's dietary intake<sup>(19)</sup>.

To estimate the similarity of the purchase data with the purchases of the general Finnish population, the average annual expenditure on groceries and non-alcoholic beverages in Finland was 2916 €, and on alcohol and cigarettes 578 € in 2016<sup>(49)</sup>. These figures are not completely comparable to our expenditure data (median 3141 €/year) because the food expenditure in our data was only that of the primary card holder, because only alcoholic beverages with  $\leq 5.5\%$  of alcohol are available in grocery stores in Finland, because our study data did not include cigarettes, and because of inflation (0.72% from 2016 to 2018). These figures are, however, of somewhat similar magnitude.

Finally, the study also had some limitations. The sample was selected as those who bought most of their grocery shopping from the retailer of the present study and excluded those who preferred other retailers. The customers of different retailers may have different background or purchase profiles. As we have shown earlier, the purchase sample differed slightly from that of the general Finnish population; there were more women, individuals with higher education, and employed individuals, and less individuals aged under 30 years and over 70 years, as well as retired individuals<sup>(22)</sup>. It is therefore possible that the purchase patterns specific only to men, to those with lower education, unemployed, and to those aged under 30 or over 70, may not have been identified by the factor analysis. In the indicator product approach, some categories contained versatile food items, which increases the uncertainty related to the carbon footprint estimates. For instance, carbon footprints for ready-to-eat meals should be considered as rough estimates. However, the indicator product approach enabled reasonably sophisticated estimates, and the results of the study can be considered robust estimates of the relative differences concerning carbon footprint of the purchase patterns. The carbon footprints of product groups did not include the customer phase, which could have a significant impact on the carbon footprints of the purchases due to, for example, transportation from stores to homes or cooking methods. It should be noted that we only covered carbon footprints, and hence other highly relevant environmental impacts of food purchase patterns (e.g. biodiversity, water footprint, eutrophication and acidification) were not assessed, although all environmental impacts should be considered together when planning actions, due to their potential trade-offs.

## Conclusions

The carbon footprint was the greatest in those with strong adherence to the *Animal-based* and the lowest in those with strong adherence to the *Plant-based* pattern. The finding that strong adherence to the *Traditional* pattern resulted in a low energy-adjusted carbon footprint but high carbon footprint per euro suggests primarily

cheap-energy-driven rather than environment-conscious purchase behaviour. The *High-energy* and *Ready-to-eat* patterns, which were both associated with moderate carbon footprints, associated with high expenditure, suggesting motives other than aiming for minimising expenditure. Because those adhering strongly to the *Animal-based* and the *Plant-based* patterns spent nearly equivalent amounts of money on food, a shift towards a recommended plant-based purchase pattern would probably not be an economic issue for those strongly adhering to the *Animal-based* pattern. The characteristics of affordable, healthy, plant-based purchase patterns that would be appealing to those who spend less money on food require further research.

### Acknowledgements

**Acknowledgements:** The authors thank Hannele Heusala, Eric Harrison and Frans Silvenius from Natural Resources Institute Finland who have contributed to producing the carbon footprint database utilised in this study. The authors thank Alice Lehtinen for language editing of the manuscript. **Financial support:** This work was funded by the Emil Aaltonen Foundation, Juho Vainio Foundation, and The Finnish Food Research Foundation. Open access funded by Helsinki University Library. **Conflicts of interest:** Both the research group and the retailer signed a contract on data transfer, ensuring the independence of the research and scientific publishing from business interests. HV received a fee from the S Group. The collaboration included offering professional advice to influencers and writing a blog post about the interpretation of the nutrition calculator in S Group's mobile app. MF is a member of the S Group's Advisory Board for Societal Responsibility. This membership involves no compensation. The authors declare no other relationships or activities that could appear to have influenced the present work. The authors declare that they have no other competing interests. **Authorship:** The authors' responsibilities were as follows: J.M., H.H., H.T., L.U., H.V., M.S., S.K., E.L., J.-M.K., M.E., J.N. and M.F. designed the study; H.H. and J.-M.K. produced the carbon footprint data; J.N., M.F., H.S. and M.E. obtained the purchase data; J.M., H.V., S.K., L.U., M.E., J.N. and M.F. processed the purchase data, J.M. and J.N. analysed the data and performed the statistical analysis; J.M. and H.H. wrote the paper; J.M., J.N., M.F. and J.-M.K. had primary responsibility for the final content. All the authors reviewed and approved the final version of the manuscript. **Ethics of human subject participation:** The study obtained ethical approval from the ethical board of the University of Helsinki Review Board in humanities and social and behavioural sciences. Each participant provided electronic consent for the collection and use of their purchase data and questionnaire data for research.

### Supplementary material

For supplementary material accompanying this paper visit <https://doi.org/10.1017/S1368980022001707>

### References

1. Willett W, Rockstrom J, Loken B *et al.* (2019) Food in the Anthropocene: the EAT-Lancet Commission on healthy diets from sustainable food systems. *Lancet* **393**, 447–492.
2. Springmann M, Mason-D'Croz D, Robinson S *et al.* (2018) Health-motivated taxes on red and processed meat: a modelling study on optimal tax levels and associated health impacts. *PLoS ONE* **13**, e0204139.
3. Wright A, Smith KE & Hellowell M (2017) Policy lessons from health taxes: a systematic review of empirical studies. *BMC Public Health* **17**, 583.
4. Hallström E, Carlsson-Kanyama A & Börjesson P (2015) Environmental impact of dietary change: a systematic review. *J Clean Prod* **91**, 1–11.
5. Aleksandrowicz L, Green R, Joy EJ *et al.* (2016) The impacts of dietary change on greenhouse gas emissions, land use, water use, and health: a systematic review. *PLoS ONE* **11**, e0165797.
6. Vinnari M, Montonen J, Härkänen T *et al.* (2009) Identifying vegetarians and their food consumption according to self-identification and operationalized definition in Finland. *Public Health Nutr* **12**, 481–488.
7. Wozniak H, Larpin C, de Mestral C *et al.* (2020) Vegetarian, pescatarian and flexitarian diets: sociodemographic determinants and association with cardiovascular risk factors in a Swiss urban population. *Br J Nutr* **124**, 844–852.
8. Lehto E, Kaartinen NE, Sääksjärvi K *et al.* (2021) Vegetarians and different types of meat eaters among the Finnish adult population from 2007 to 2017. *Br J Nutr* **127**, 1–13.
9. Vieux F, Darmon N, Touazi D *et al.* (2012) Greenhouse gas emissions of self-selected individual diets in France: changing the diet structure or consuming less? *Ecol Econ* **75**, 91–101.
10. Mogensen L, Hermansen JE & Trolle E (2020) The climate and nutritional impact of beef in different dietary patterns in Denmark. *Foods* **9**, 1176.
11. Rehm CD, Monsivais P & Drewnowski A (2011) The quality and monetary value of diets consumed by adults in the United States. *Am J Clin Nutr* **94**, 1333–1339.
12. Rydén PJ & Hagfors L (2011) Diet cost, diet quality and socioeconomic position: how are they related and what contributes to differences in diet costs? *Public Health Nutr* **14**, 1680–1692.
13. Darmon N & Drewnowski A (2015) Contribution of food prices and diet cost to socioeconomic disparities in diet quality and health: a systematic review and analysis. *Nutr Rev* **73**, 643–660.
14. Jones NR, Tong TY & Monsivais P (2018) Meeting UK dietary recommendations is associated with higher estimated consumer food costs: an analysis using the national diet and nutrition survey and consumer expenditure data, 2008–2012. *Public Health Nutr* **21**, 948–956.
15. Drewnowski A (2009) Defining nutrient density: development and validation of the nutrient rich foods index. *J Am Coll Nutr* **28**, 421S–426S.
16. Erkkola M, Kinnunen SM, Vepsäläinen HR *et al.* (2022) A slow road from meat dominance to more sustainable diets: an analysis of purchase preferences among Finnish loyalty-card holders. *PLOS Sustain Transform* **1**, e0000015. doi: 10.1371/journal.pstr.0000015.
17. Lusk JL & Norwood FB (2016) Some vegetarians spend less money on food, others don't. *Ecol Econ* **130**, 232–242.



18. Funke F, Mattauch L, van den Bijgaart I *et al.* (2021) Is Meat Too Cheap? Towards Optimal Meat Taxation. INET Oxford Working Paper No. [https://www.inet.ox.ac.uk/files/Funke\\_et\\_al\\_2021\\_Towards\\_optimal\\_meat\\_taxation\\_WP.pdf](https://www.inet.ox.ac.uk/files/Funke_et_al_2021_Towards_optimal_meat_taxation_WP.pdf) (accessed June 2021).
19. Vepsäläinen H, Nevalainen J, Kinnunen S *et al.* (2021) Do we eat what we buy? Relative validity of grocery purchase data as an indicator of food consumption in the LoCard study. *Br J Nutr*, 1–24. doi: 10.1017/S0007114521004177.
20. Lintonen T, Uusitalo L, Erkkola M *et al.* (2020) Grocery purchase data in the study of alcohol use – a validity study. *Drug Alcohol Depend* **214**, 108145.
21. Finnish Grocery Trade Association (2019) Päivittäistavarakaupan Myynti Ja Markkinaosuudet 2018 (Sales and Market Share of the Grocery Trade Groups in 2018). Finnish Grocery Trade Association. <https://www.pty.fi/ajankohtaista/tiedotteet/uutinen/article/paivittaeistavara-kaupan-myynti-ja-markkinaosuudet-2018/> (accessed August 2021).
22. Vuorinen A, Erkkola M, Fogelholm M *et al.* (2020) Characterization and correction of bias due to nonparticipation and the degree of loyalty in large-scale Finnish loyalty card data on grocery purchases: cohort study. *J Med Internet Res* **22**, e18059.
23. Hartikainen H & Pulkkinen H. (2016) *Summary of the Chosen Methodologies and Practices to Produce GHGE-Estimates for an Average European Diet*. Helsinki: Natural Resources Bioeconomy Studies.
24. Wernet G, Bauer C, Steubing B *et al.* (2016) The ecoinvent database version 3 (part I): overview and methodology. *Int J Life Cycle Assess* **21**, 1218–1230.
25. Plastics Europe (2019) Eco-Profiles – Program and Methodology 2. PlasticsEurope, Version 3.0. <https://www.plasticseurope.org/en/resources/eco-profiles> (accessed October 2019).
26. The European Aluminium Association (2018) Environmental Profile Report 2018. <https://www.european-aluminium.eu/resource-hub/environmental-profile-report-2018/> (accessed October 2019).
27. The World Steel Association (2019) Life Cycle Inventory Study 2019 Data Release. [https://www.worldsteel.org/en/dam/jcr:c4159749-afab-4476-a09f-59efca686e9e/LCI%2520study\\_2019%2520data%2520release.pdf](https://www.worldsteel.org/en/dam/jcr:c4159749-afab-4476-a09f-59efca686e9e/LCI%2520study_2019%2520data%2520release.pdf) (accessed October 2019).
28. FEFCO (2018) European Database for Corrugated Board Life Cycle Studies. FEFCO – Corrugated Packaging. <https://www.fefco.org/lca> (accessed October 2019).
29. Mäkelä K & Auvinen H (2009) *LIPASTO – Transport Emission Database*. Finland: VTT Technical Research Centre of Finland, Espoo (Finland).
30. Jolliffe IT & Cadima J (2016) Principal component analysis: a review and recent developments. *Philos Trans Royal Soc A* **374**, 20150202.
31. Piernas C, Mendez MA, Ng SW *et al.* (2014) Low-calorie- and calorie-sweetened beverages: diet quality, food intake, and purchase patterns of US household consumers. *Am J Clin Nutr* **99**, 567–577.
32. Peltner J & Thiele S (2018) Convenience-based food purchase patterns: identification and associations with dietary quality, sociodemographic factors and attitudes. *Public Health Nutr* **21**, 558–570.
33. Thiele S, Peltner J, Richter A *et al.* (2017) Food purchase patterns: empirical identification and analysis of their association with diet quality, socio-economic factors, and attitudes. *Nutr J* **16**, 69.
34. Clark SD, Shute B, Jenneson V *et al.* (2021) Dietary patterns derived from UK supermarket transaction data with nutrient and socioeconomic profiles. *Nutrients* **13**, 1481.
35. Naja F, Jomaa L, Itani L *et al.* (2018) Environmental footprints of food consumption and dietary patterns among Lebanese adults: a cross-sectional study. *Nutr J* **17**, 85.
36. Green RF, Joy EJM, Harris F *et al.* (2018) Greenhouse gas emissions and water footprints of typical dietary patterns in India. *Sci Total Environ* **643**, 1411–1418.
37. Monteiro CA, Levy RB, Claro RM *et al.* (2010) A new classification of foods based on the extent and purpose of their processing. *Cad Saude Publica* **26**, 2039–2049.
38. Steele EM, Baraldi LG, Louzada MLDC *et al.* (2016) Ultra-processed foods and added sugars in the US diet: evidence from a nationally representative cross-sectional study. *BMJ Open* **6**, e009892.
39. Louzada MLDC, Ricardo CZ, Steele EM *et al.* (2018) The share of ultra-processed foods determines the overall nutritional quality of diets in Brazil. *Public Health Nutr* **21**, 94–102.
40. Rauber F, Campagnolo PDB, Hoffman DJ *et al.* (2014) Consumption of ultra-processed food products and its effects on children’s lipid profiles: a longitudinal study. *Nutr Metab Cardiovasc Dis* **25**, 116–122.
41. Tavares LF, Fonseca SC, Garcia Rosa ML *et al.* (2012) Relationship between ultra-processed foods and metabolic syndrome in adolescents from a Brazilian Family Doctor Program. *Public Health Nutr* **15**, 82–87.
42. Machado PP, Steele EM, Levy RB *et al.* (2020) Ultra-processed food consumption and obesity in the Australian adult population. *Nutr Diabetes* **10**, 39.
43. Saarinen M, Kurppa S, Virtanen Y *et al.* (2012) Life cycle assessment approach to the impact of home-made, ready-to-eat and school lunches on climate and eutrophication. *J Clean Prod* **28**, 177–186.
44. Schmidt Rivera XC & Azapagic A (2019) Life cycle environmental impacts of ready-made meals considering different cuisines and recipes. *Sci Total Environ* **660**, 1168–1181.
45. Nordic Council of Ministers (2014) *Nordic Nutrition Recommendations 2012 – Integrating Nutrition and Physical Activity*, 5th ed. Copenhagen: Nordic Council of Ministers.
46. Bai Y, Alemu R, Block SA *et al.* (2021) Cost and affordability of nutritious diets at retail prices: evidence from 177 countries. *Food Policy* **99**, 101983.
47. Hirvonen K, Bai Y, Headey D *et al.* (2020) Affordability of the EAT-Lancet reference diet: a global analysis. *Lancet Glob Health* **8**, e59–e66.
48. Kontinen H, Halmesvaara O, Fogelholm M *et al.* (2021) Sociodemographic differences in motives for food selection: results from the LoCard cross-sectional survey. *Int J Behav Nutr Phys Act* **18**, 71.
49. Statistics Finland (2018) Statistics Finland’s Free-of-Charge Statistical Databases. <http://pxnet2.stat.fi/PXWeb/pxweb/en/StatFin/> (accessed March 2022).