



Detection of the phycotoxin pectenotoxin-2 in waters around King George Island, Antarctica

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Abstract

In order to set a base line for the observation of planktonic community changes due to global change, protistan plankton sampling in combination with phycotoxin measurements and solid phase adsorption toxin tracking (SPATT) was performed in two bays of King George Island (KGI) in January 2013 and 2014. In addition, SPATT sampling was performed in Potter Cove during a one-year period from January 2014 until January 2015. Known toxigenic taxa were not firmly identified in plankton samples but there was microscopical evidence for background level presence of *Dinophysis* spp. in the area. This was consistent with environmental conditions during the sampling periods, especially strong mixing of the water column and low water temperatures that do not favor dinoflagellate proliferations. Due to the lack of significant abundance of thecate toxigenic dinoflagellate species in microplankton samples, no phycotoxins were found in net tow samples. In contrast, SPATT sampling revealed the presence of dissolved pectenotoxin-2 (PTX-2) and its hydrolyzed form PTX-2 seco acid in both bays and during the entire one-year sampling period. The presence of dissolved PTX in coastal waters of KGI is strong new evidence for the presence of PTX-producing species, i.e., dinoflagellates of the genus *Dinophysis* in the area. The presence of phycotoxins and their respective producers, even at the low background concentrations found in this study, may be the seed of possible proliferations of these species under changing environmental conditions. Furthermore, phycotoxins can be used as chemotaxonomic markers for a very specific group of plankton thus allowing to track the presence of this group over time.

Keywords Protistan plankton community · Dinoflagellates · Phycotoxin · Liquid chromatography-tandem mass spectrometry (LC–MS/MS) · Solid phase adsorption toxins tracking (SPATT) · Harmful algae bloom (HAB)

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Introduction

During the second half of the twentieth century, the Western Antarctic Peninsula was one of the most affected areas by Global Change, although some cooling trend was evident after 1998 (Turner et al. 2016). These observations can be better described as a delayed warming shaped by the Southern Ocean's meridional overturning circulation and reflect the natural variability superimposed to the long-term warming trend (Armour et al. 2016). However, the overall warming of the sea at the West Antarctic Peninsula remains statistically significant (Henley et al. 2019). Increasing temperatures significantly affect sea and glacier ice dynamics, resulting in increased water column stratification when melting occurs, in turn affecting mixing regimes, nutrient supply, and light availability for Antarctic phytoplankton (Vernet et al. 2008; Schloss et al. 2012; Ducklow et al. 2013).

Surface salinity decrease has a significant impact on coastal Antarctic phytoplankton community composition (Hernando et al. 2015). The aforementioned study evidenced a decrease in the relative abundance of large centric diatoms and an increase in small pennate diatoms such as *Fragilariopsis cylindrus* and *F. nana* after 7 days of exposure to low salinity conditions. In addition, increased temperature and meltwater extend thermal stratification and strengthen the salinity gradient, which may explain some of the changes in phytoplankton dominance (Moline et al. 2004; Mendes et al. 2013). For temperate regions, increased stratification of the water column has been proposed to enhance dinoflagellate growth in coastal assemblages (Margalef 1978; Glibert 2016) and temperature, along with light, influences the germination of dinoflagellate cysts (Anderson et al. 2005). As a consequence, warming trends might result in Harmful Algae Bloom (HAB) seed populations appearing in Antarctic surface coastal waters, reflecting the onset of permissive temperatures for germination (Kremp and Anderson 2000) and increased germination rates favored at higher temperatures (Anderson et al. 2005).

Global Change may lead to a range extension of temperate and subpolar phytoplankton species to Polar Regions (Hallegraeff 2010). Coastal Antarctic waters are relatively isolated from the rest of the world oceans by the Antarctic Circumpolar Current (ACC), and thus an endemic phytoplankton community can be expected in Antarctic waters (Medlin et al. 2000). However, non-endogenous species, including potentially toxic species, may enter Antarctica through ship traffic (Bailey 2015). Successful “invasions” of new HAB species will depend fundamentally on their ability to compete with the local species assemblages, especially under Global Change induced changes

in environmental conditions. Antarctic spring blooms are typically dominated by diatoms (Kim et al. 2018), but almost nothing is known about the occurrence of harmful dinoflagellate species and their associated phycotoxins in Antarctic waters.

The aim of this work was to investigate the presence of potentially toxigenic microalgae and their phycotoxins in coastal waters of King George Island or Isla 25 de Mayo, hereafter KGI, South Shetland Islands, Antarctica, in an effort to detect changes in coastal Antarctic environments due to Global Change. As HAB species levels in Antarctic waters were expected to be very low, integrative sampling of lipophilic phycotoxins was applied in order to be able to detect minute amounts of these molecules. In addition, it was aimed to assess the seasonal distribution of phycotoxins by implementing a phycotoxin survey during an entire year.

Material and methods

Study areas

Different datasets were obtained from Fildes Bay (62.2031° S, 58.9208° W) and Potter Cove (62.2373° S, 58.7009° W), KGI, South Shetland Islands, Western Antarctic Peninsula, during January 2013 (Fildes Bay) and January 2014 (Potter Cove) (Fig. 1). In addition, a year-round measurement of dissolved phycotoxins and CTD parameters was obtained for Potter Cove (2014–2015) that was complemented with a long-term plankton analysis.

Fildes bay

Field sampling

Salinity, fluorescence, oxygen and temperature data were obtained at Fildes Bay (FB) station (62.2031°S, 58.9708°W; Fig. 1) using a SBE 911 plus (Sea-Bird Scientific, Oregon, USA) CTD profiler (11, 12, 14, 17, 18, 21, 22, 23 and 25 January 2013). The CTD casts were processed and validated with the SBE data processing software version 7.23.2. Additionally, on each sampling day, seawater samples were collected at FB station using 5-L Niskin bottles. Samples were taken near the surface (5 m) and at a depth corresponding to 10% of the surface photosynthetically active radiation (PAR) (app. 25 m). Samples were pre-filtered on board through a 150- μ m mesh to remove zooplankton and large particles. Samples were stored in acid-washed plastic carboys and kept in darkness until further processing. Once at the laboratory (< 2 h later), subsamples for microscopy and molecular analyses were taken.

In addition, the upper 20 m of the water column were sampled by six vertical net tows (40 cm diameter, 20- μ m

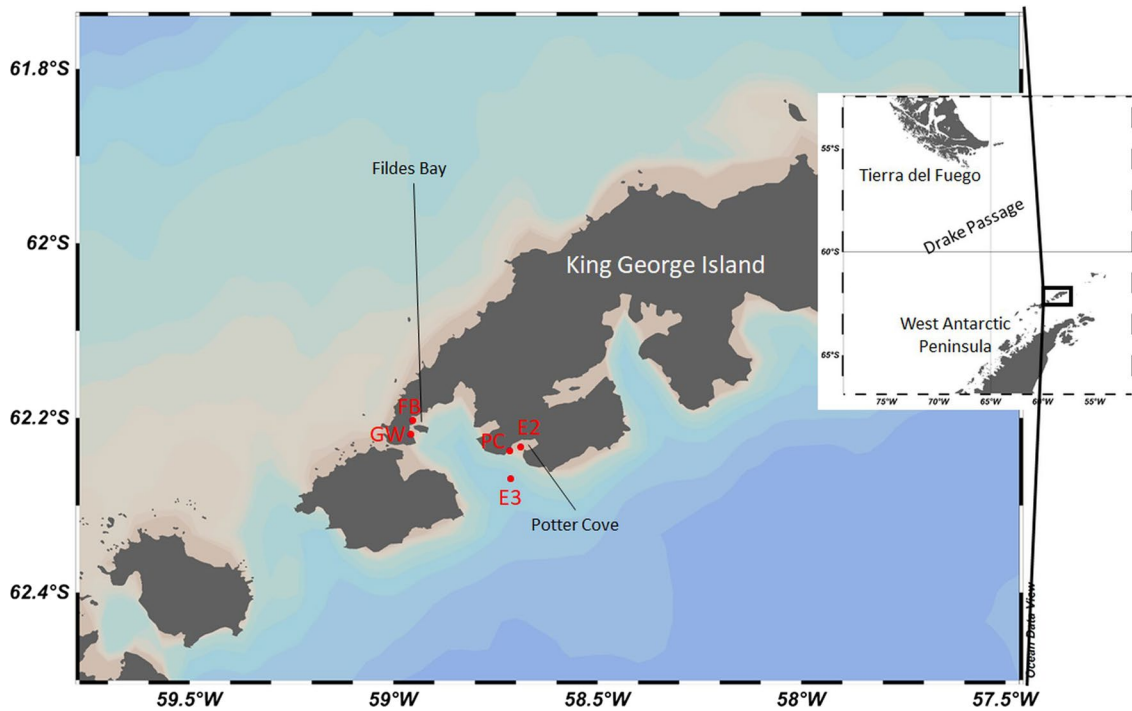


Fig. 1 Geographic locations of SPATT deployments in Fildes Bay (FB), Great Wall station (GW), Potter Cove (PC) and E3 during the two field seasons 2013 and 2014–2015. Protistan plankton for the long-term series was partially sampled at E2

mesh size) and subsequently the combined microplankton concentrates were adjusted with filtered sea water to a final volume of 500 mL and the rest was size fractionated over meshes of 150 µm, 60 µm und 20 µm with a 7553–75 peristaltic pump (Cole-Palmer, Vernon Hills, IL, USA) in order be able to attribute phycotoxins to planktonic groups. Plankton of each mesh were rinsed with 0.2-µm filtered seawater into 50-mL centrifugation tubes (Sarstedt, Nümbrecht, Germany) and adjusted to 45 mL. Each fraction was split into three 15-mL aliquots and each aliquot was centrifuged at 8600×g for 20 min (Sigma 2–16 PK, Sartorius, Göttingen, Germany). Supernatants were removed and pellets re-suspended in ca. 1 mL 0.2-µm filtered seawater, transferred to 2 mL centrifugation tubes and centrifuged again for 20 min

at 21,460×g and 4 °C (Sigma 2–16 PK). Supernatants were removed and two pellets of each station and size fraction (hydrophilic and lipophilic phycotoxins) were frozen at –20 °C for phycotoxin analysis.

Solid phase adsorption toxin tracking sampling

SPATT (MacKenzie et al. 2004) bags (5 × 5 cm) were hand sewed of 50-µm mesh gauze and filled with 1 g Diaion HP20 (Sigma, Deisenhofen, Germany). Filled SPATT bags were conditioned by immersion into methanol and stirring overnight. Conditioned SPATTs were washed with deionized water, individually placed moist into a zip-lock bag and kept

Table 1 SPATT deployment times and locations in January 2013

SPATT #	Deployment		Exposure time (d)	Latitude °S	Longitude °W	Station
	From	To				
1	14. 1. 2013	18. 1. 2013	4	62.2024	58.9538	FB
2	16. 1. 2013	21. 1. 2013	5	62.2186	58.9588	GW
3	22. 1. 2013	29. 1. 2013	7	62.2370	58.7153	PC
4	18. 1. 2013	23. 1. 2013	5	62.2024	58.9538	FB
5	21. 1. 2013	26. 1. 2013	5*	62.2186	58.9588	GW

FB fildes Bay, GW Great Wall station, PC Potter Cove

Asterisk (*) denotes SPATT that was retrieved dry and exact exposure time cannot be determined

at 4 °C until use. SPATTs were placed at different localities in KGI for approximately five days (Fig. 1, Table 1).

Protistan plankton analysis by microscopy

Aliquots of 1 mL from the Niskin bottles without fixation were screened under a CX21FS light microscope (Olympus, Tokyo, Japan) with 40 × magnification. In addition, from each sampling day at Fildes Bay (2013), 50-mL seawater samples from the Niskin bottles were fixed with formalin (1%) and stored at room temperature in the dark, until further analyses. For a non-quantitative analysis (search for *Dinophysis* cells) the whole sample volume was partitioned in several small plastic Petri dishes (5.5 cm in diameter) and after cell settlement completely scanned with a Leica DMIL inverted microscope (Leica Microsystems, Wetzlar, Germany) at 200 × and 400 × magnification.

Molecular analysis

One seawater sample (4.5 L) taken on 18 Jan 2013 was size fractionated using a 7553–75 peristaltic pump (Cole-Palmer) by sequential filtration using 47-mm diameter Swinnex filter holder (Millipore), and 60- μ m, 20- μ m (Nylon, Millipore, Burlington, MA, USA), 12- μ m, 3- μ m and 0.2- μ m (Polycarbonate, Millipore) pore size filters. Filters were stored in 2-mL cryovials at –80 °C until analysis. DNA extraction, quantitation, amplicon sequencing by Next Generation Sequencing (NGS) and bioinformatic analyses were performed according to Rodríguez-Marconi et al. (2015). Briefly, DNA was extracted following a general hexadecyltrimethyl-ammonium bromide (CTAB) and phenol–chloroform method. Tag sequencing was performed on a Illumina MiSeq platform following Caporaso et al. (2011) protocol and using 1391f and EukBr primers (Amaral-Zettler et al. 2009) for amplification of the hypervariable region V9 of the 18S rRNA gene. Sequences were analyzed using Mothur software (Schloss et al. 2009). Raw sequence data were deposited in SRA under BioProject number PRJNA287634, corresponding to SW samples, as indicated in Rodríguez-Marconi et al. (2015).

Potter Cove

Plankton sampling

A SBE CTD (Sea-Bird Scientific, Oregon, USA) was used to record seawater temperature and conductivity (transformed in salinity) profiles from the surface down to around 30 m. Density was estimated, and the Brunt-Väisälä frequency (N^2) was calculated as in Schloss et al. (2014). Additionally, the upper 20 m of the water column were sampled by six vertical net tows (40 cm diameter, 20- μ m mesh size)

and subsequently the combined plankton concentrates were taken to 1000 mL with 0.2- μ m filtered seawater. 50 mL were taken for microscopy (see below) and the rest was size fractionated over meshes of 200 μ m, 50 μ m and 20 μ m by gravity filtration. Plankton of each mesh were rinsed with 0.2- μ m filtered seawater into 50-mL centrifugation tubes (Sarstedt, Nümbrecht, Germany) and adjusted to 45 mL. Each fraction was split into three 15-mL aliquots and each aliquot was centrifuged at 4000 $\times g$ for 15 min (model 5403, rotor 16 A4-44, Eppendorf, Hamburg, Germany). Supernatants were removed and pellets re-suspended in ca. 1 mL 0.2- μ m filtered seawater, transferred to 2-mL centrifugation tubes and centrifuged again at 4000 $\times g$ for 15 min. Supernatants were removed and two pellets of each station for phycotoxin analysis (hydrophilic and lipophilic phycotoxins) were frozen at –20 °C until analysis.

Additionally, discrete water samples were collected at five depths (i.e., 0, 5, 10, 20 and 30 m) with 4.7 L Niskin bottles.

Solid phase adsorption toxin tracking sampling

SPATT bags consisted of 50 × 150-mm filter bags (Carbis Filtration, Stockton-On-Tees, Great Britain), were filled with 10 g Diaion HP20 (Sigma, Deisenhofen, Germany) and sealed with Holdon Midi Clips (Thomann, Burgebrach, Germany). Filled SPATT bags were conditioned by immersion into methanol and stirring overnight. Conditioned SPATTs were washed with deionized water, individually placed moist into a zip-lock bag and kept at 4 °C until use. SPATTs were deployed at the mooring at the permanent sampling station E2 in Potter Cove, KGI (62.2328° S, 58.6886° W, Fig. 1) during the one-year period from February 2014 to January 2015. SPATTs were deployed for approximately one month of exposure time (between 29 and 53 days) depending on accessibility of the sampling site due to weather conditions and ice coverage (Table 2, Fig. S2).

Protistan plankton analysis by microscopy

Microplankton net samples > 20 μ m and bottle samples were used for protistan plankton microscopical observation and documentation. For bottle samples, 1-L sample was gently concentrated by gravity filtration using a 3- μ m polycarbonate filter (47 mm diameter, Whatman, Chicago, IL, USA), and examined using an Diavert inverted microscope (Leica Microsystems, Wetzlar, Germany). Micrographs were taken with a MicrOculareyepiece-camera (Bresser, Rhede, Germany).

Long-term summer plankton composition

Data presented here are part of the sampling conducted close to the Argentinean Carlini Station (formerly called

Table 2 SPATT deployment times and locations from January 2014 to February 2015 at sampling station E2 (62.2328°S, 58.6886°W, 62.2693°S, 58.7136°W)

SPATT #	Deployment		Exposure time [d]
	From	To	
2	18.02.2014	18.03.2014	28
3	18.03.2014	21.04.2014	34
4	21.04.2014	21.05.2014	30
5	25.05.2014	s.l.	–
6	07.06.2014	18.07.2014	41
7	18.07.2014	10.09.2014	53
8	10.09.2014	16.10.2014	35
9	16.10.2014	21.11.2014	36
10	21.11.2014	20.12.2014	29
11	20.12.2014	18.02.2015	40

s.l. sample lost

Jubany) as part of a long-term monitoring program. Water was collected at two sites, E2 (62.2328° S, 58.6886° W) and E3 (62.2693° S, 58.7136° W), with 4.7-L Niskin bottles during the Austral summers 2010, 2011–2012, 2014–2016 (Table S1). Aliquots of 150 mL were fixed with 4% acetic Lugol's solution and kept in dark conditions at room temperature until analysis. For quantitative estimation of plankton assemblage composition, cells were enumerated with a phase contrast Leica DMIL LED inverted microscope (Leica Microsystems, Wetzlar, Germany) according to the procedures described by Utermöhl (1958). Subsamples of 50 mL were settled for 24 h in a composite sedimentation chamber. At least 100 cells of the dominant taxa were counted in one or more strips of the chamber or random fields at 250 x or 400 x, depending on their concentration and size. To count large and sparse species, the whole surface of the chamber was scanned at a magnification of $\times 100$, with a detection limit of 20 cells L^{-1} . Small unidentified specimens other than diatoms, dinoflagellates (phototrophic- and heterotrophic taxa) cryptophytes, prasinophytes and silicoflagellates were included in a single group as 'flagellates' and classified according to their size.

Ice cover data

The annual number of days with fast ice was obtained by daily photographic observations of Potter Cove from January 2014 to March 2015 (Gómez Izquierdo et al. 2014, 2015). With these photographic images it was registered whether the cove was frozen or not.

Phycotoxin extraction from passive samplers

SPATT bags were rinsed with deionized water, placed on tissue paper for removal of excess water and dried in an oven for one hour at 50 °C. Dry SPATT bags were opened and the resin was transferred into 50-mL centrifugation tubes. 25 mL methanol was added and left over night. The next day each resin-methanol mixture was poured into a glass chromatography column (270 mm length, 13 mm internal diameter, filled with 20 mm glass wadding and 10 mm quartz sand). Additional 15 mL methanol were used to rinse the centrifugation tube and added to the column. Methanol was eluted dropwise from the column until the supernatant reached the top of the filling and 25 mL methanol were added to the column and subsequently eluted. The eluate was reduced in a rotary evaporator to approximately 0.5 mL and then taken to 1 mL with methanol.

Phycotoxin measurements

Hydrophilic phycotoxins were measured by ion pair chromatography with fluorescence detection and post-column derivatization as described in Krock et al. (2007). Lipophilic phycotoxins and domoic acid (DA) were measured by liquid chromatography-tandem mass spectrometry in the selected reaction monitoring (SRM) mode as described by Krock et al. (2008).

Results

Environmental conditions in the study areas

In Fildes Bay, sea surface temperature increased during the studied period (Fig. 2), starting around 16 January, with a peak registered on 23 January (1.25 °C). The first increase was accompanied by a slight decrease in surface salinity (from 34.1 to 34.05), and contributed to the observed weak density stratification. The presence of a colder, saltier and denser water mass was evident at depth > 30 m probably entering from Maxwell Bay (see Höfer et al. 2019). No evident water mixing event was observed during the study period in this enclosed bay. In Potter Cove sea surface water temperature followed the seasonal cycle, although a cold-water mass was evident in February (Fig. 3a). Cold water characterized the fall and winter waters, and warming of the surface layers started in late October. Water column salinity (Fig. 3b) showed a homogeneous distribution in summer. A winter warming effect was evident in salinity, when a cold, fresh surface water mass entered the cove, probably as a result of air warming and glacial melting, which adds freshwater to the Cove. The effect of melting was also evident again in the following summer season, in January, when

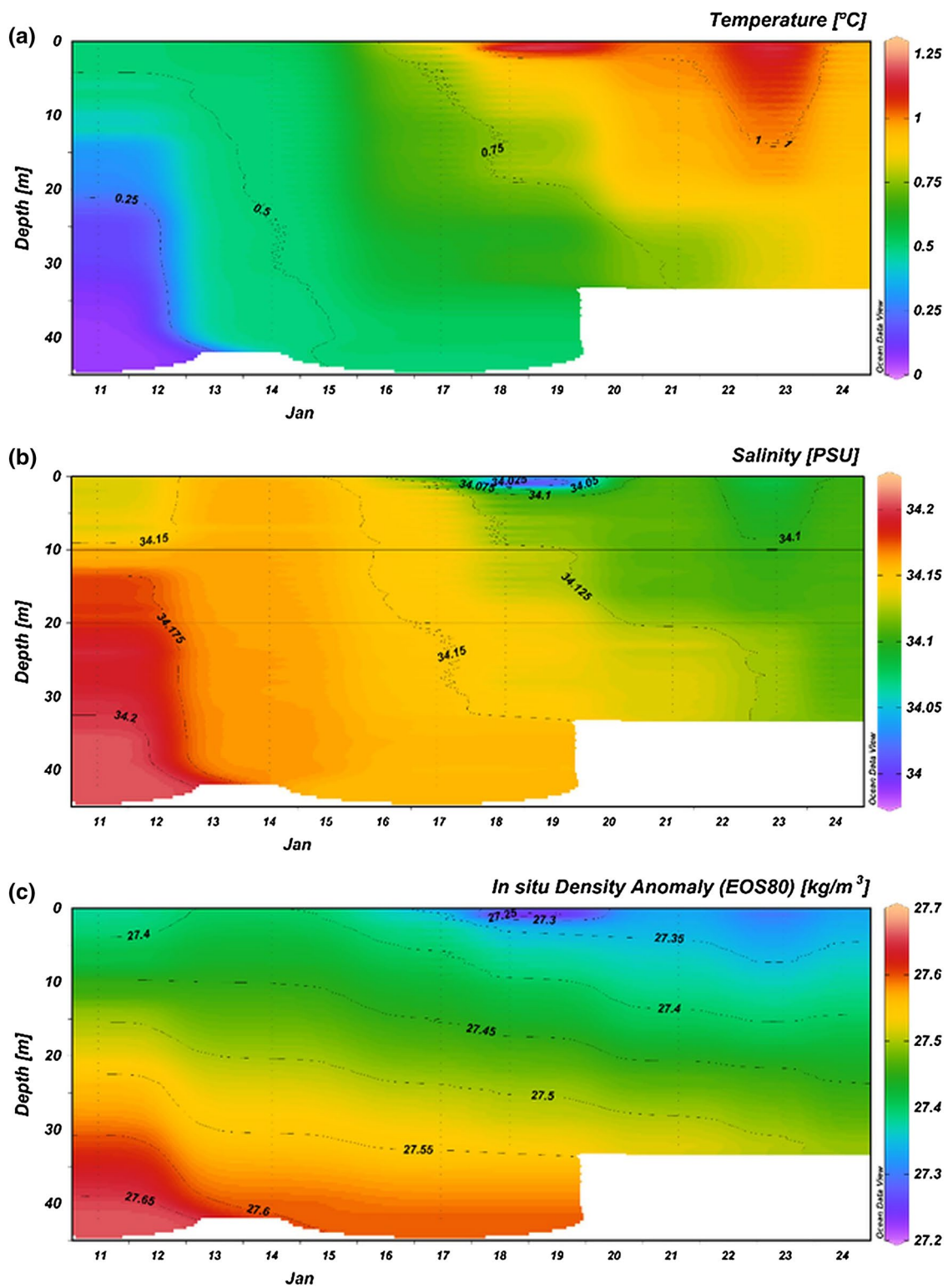
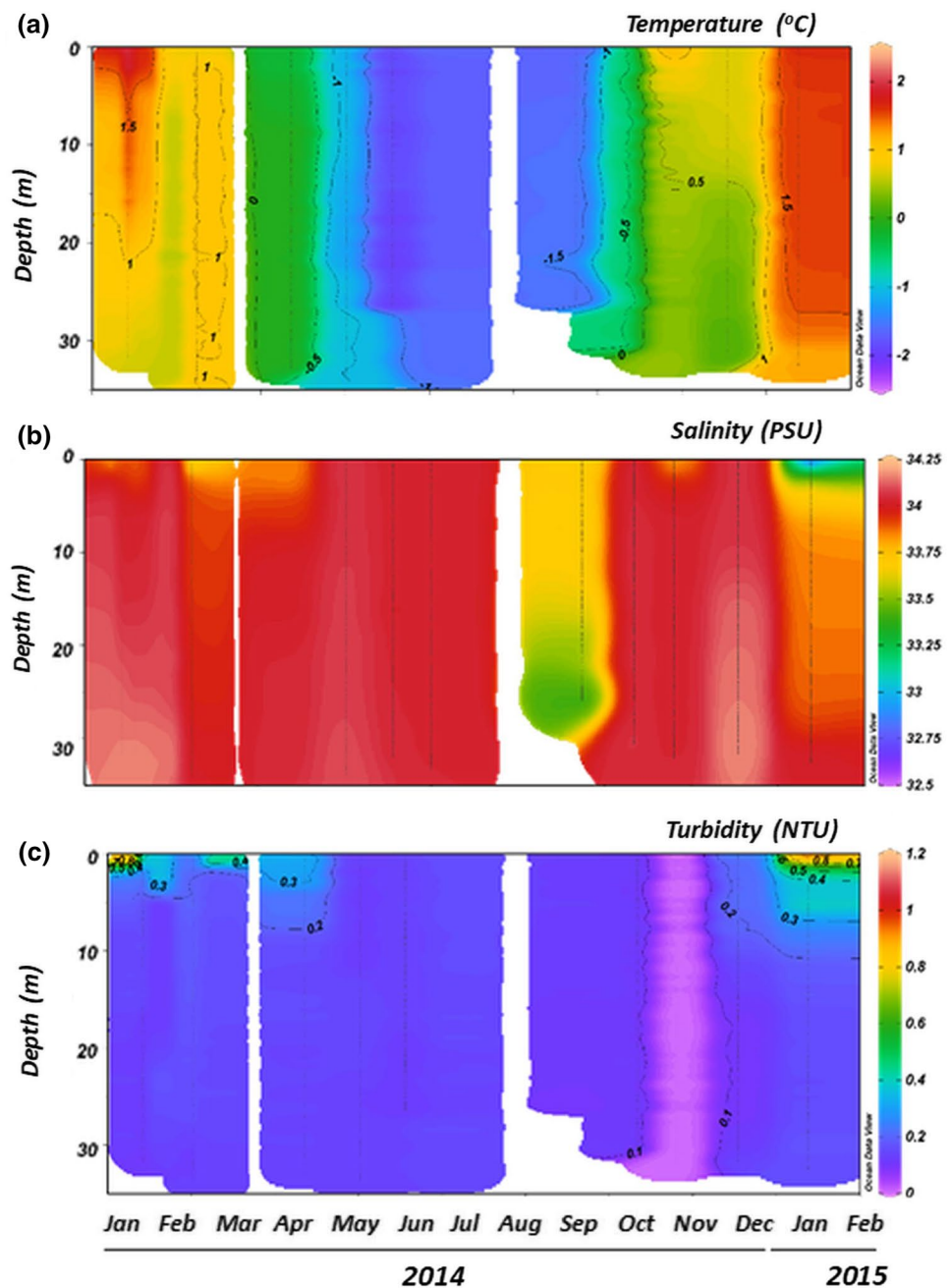


Fig. 2 a Sea water temperature, b salinity, and c density at Fildes Bay, KGI in January 2013

Fig. 3 **a** Sea water temperature, **b** salinity, **c** turbidity at Potter Cove, KGI during the studied period. White areas correspond to periods for which no vertical CTD profiles are available



low salinity and density surface waters were evident in the profiles. This was also apparent in the turbidity profiles at that time (Fig. 3c). The Brunt-Väisälä frequency, which estimates the static stability of the water column, was further considered as a measure of stratification strength (as in Geyer et al 2008). Although values were somewhat lower in January and February 2014, there were in the same order of the climatic (1991–2016) average. In general, except for the winter months, the studied period presented a lower N^2 value than the average (Fig. S1). Potter Cove was only frozen in total for less than two months (54 days) from June to October (Fig. S2). Moreover, between these months, there were

several periods in which the cove was free of ice, and it was mainly frozen especially from August to mid-September.

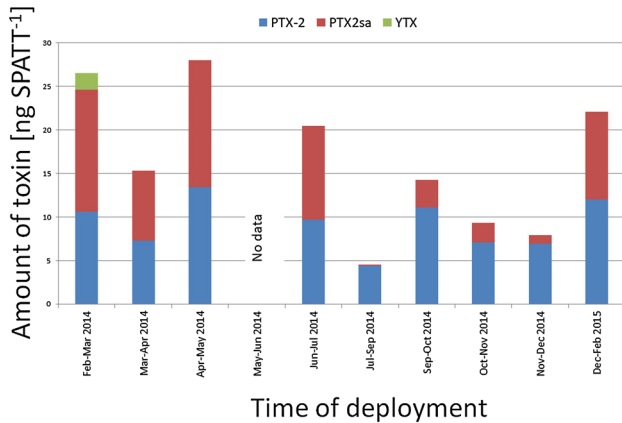
Phycotoxins

Neither hydrophilic PSP toxins and domoic acid (DA) nor lipophilic phycotoxins (gymnodimines, spirolides, okadaic acid, dinophysistoxins, yessotoxin (YTX), pectenotoxins (PTX), azaspiracids) were detected in any of the nano- and microplankton concentrates taken during the summer expeditions in KGI in 2013 and 2014. The detection limits of these measurements ranged between 0.01 and 1 ng

Table 3 Pectenotoxin-2 (PTX-2) amounts [ng] per SPATT bag in Fildes Bay in January 2013

SPATT #	PTX-2 (ng SPATT ⁻¹)	Station
1	7.2	FB
2	3.3	GW
3	14.7	PC
4	11.9	FB
5	nd	GW

For geographic locations of SPATT deployments see Fig. 1
nd not detected

**Fig. 4** Amounts [ng] of phycotoxins per SPATT bag detected in the time series from January 2014 to February 2015 at sampling station E2 in Potter Cove: pectenotoxin-2 (PTX-2, blue bars), pectenotoxins-2 seco acid (PTX-2sa, red bars) and yessotoxin (YTX, green bar)

per net tow depending on the sensitivity of each individual phycotoxin (Table S2). On the other hand, pectenotoxin-2 (PTX-2) was detected on four out of five SPATT samplers during the austral summer expeditions in Fildes Bay and Potter Cove, KGI in January 2013. Amounts ranged from 3 to 15 ng per SPATT bag and were present in almost all samplers (Table 3). This unexpected finding motivated the start of a continuous, long-time series of SPATT sampling for lipophilic phycotoxins in the area, which was performed in Potter Cove, KGI. This one-year long SPATT sampling from January 2014 to February 2015 revealed the presence of PTX-2 during the entire sampling period (Fig. 4), same as the sporadic SPATT samplings in Fildes Bay in the previous year (Table 3). All SPATT samplers contained PTX-2 in abundances from 4 to 14 ng per SPATT bag.

In addition to PTX-2, also pectenotoxin-2 seco acid (PTX-2sa) was detected on all SPATT bags except for one corresponding to the winter period, i.e., July–September 2014. PTX-2sa levels were in the same range (0.12–14.6 ng SPATT⁻¹) as PTX-2, but tended to be lower in the winter-spring period. In contrast to PTX-2sa, abundances of PTX-2 did not show any clear seasonal distribution patterns. The only

other phycotoxin not related to PTX was YTX. YTX was only detected once at an abundance of 1.9 ng SPATT⁻¹ on a SPATT sampler deployed in the summer months, i.e., February–March 2014 (Fig. 4).

Protistan plankton composition in Fildes Bay in 2013

The detection of PTX-2 prompted the search for the toxin-producing organisms in these coastal waters. For this purpose, live plankton samples and fixed water samples obtained from Fildes Bay in January 2013 were inspected. Fixed water samples generally showed relatively low cell abundances. Microplankton communities in these samples were dominated by diatoms of the genera *Thalassiosira* and *Corethron* and contained to a lesser extent small *Chaetoceros* species, whereas dinoflagellates were not detected except for one occasion, where a *Dinophysis* cell was found by in situ microscopy (Fig. 5a). This cell contained plastids, was relatively large, laterally flattened with an epitheca hidden behind the upwards directed precingular list and the hypotheca being deepest about the middle level of its length, pointing posteriorly. The hypotheca was dorsally slightly convex and more straight ventrally below the sulcal list. The left sulcal list was longer than half the hypotheca length and the third rib was slightly pointing posteriorly. The theca had strong reticulate-foveate ornamentation (Fig. 5a). Based on these characters the cell was classified as *Dinophysis norvegica*.

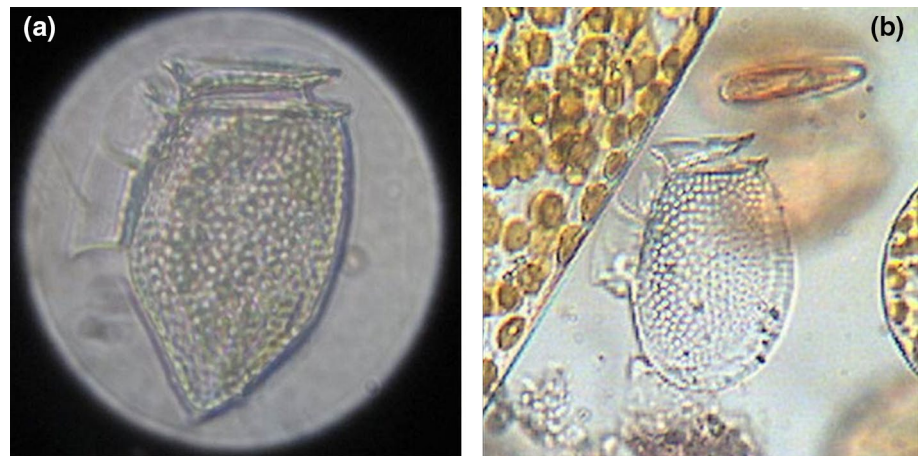
Molecular data from seawater samples of Fildes Bay in 2013

After quality filtering and removal of undesired sequences (metazoan, chloroplast, mitochondria), a total of 1,599,680 sequences and 1600 OTUs were obtained. Stramenopiles captured most of the sequences (1,321,926), followed by Alveolata (126,768) and Hacrobia (24,344) (Fig. S3). Unclassified sequences were 7.2% of the total sequences (115,484). At higher taxonomical levels, Bacillariophyta was the most abundant group, with 88.5% of relative abundance. Dinophyceae corresponded to 8.3% of relative abundance. However, most of the sequences in this class were assigned to unclassified Dinophyceae, and sequences belonging to the genus *Dinophysis* nor even to the order Dinophysales could not be identified.

Protistan plankton observations in Potter Cove in January 2014

Microplankton communities during the sampling period were generally dominated by diatoms with high densities of various species of *Thalassiosira* and *Corethron*. Among

Fig. 5 **a** Live *Dinophysis norvegica* cell photographed through the ocular of a microscope from a Niskin bottle sample taken in Fildes Bay in January 2013. **b** Light microscopy picture of a plankton concentrate from Potter Cove in January 2014 including an empty theca of *Dinophysis*



the diatoms, cells of *Pseudo-nitzschia* were also regularly detected. Although relatively low in abundance, a high diversity of dinoflagellates was detected, but these almost exclusively belonged to the large group of athecate dinoflagellates, i.e., species without rigid plates, which could not be identified to the species level. Among thecate phototrophic dinoflagellates cells of a small *Prorocentrum* species were encountered quite regularly. One single empty theca of a species of *Dinophysis* resembling *D. acuminata* was encountered (Fig. 5b), but the extremely low abundance prevented a further species determination.

Long-term summer protistan plankton composition in Potter Cove (2010–2016)

Nano- and microplankton composition was characterized by the dominance of unidentified tiny flagellates (< 5 μm) and diatoms in all samples collected from 2010 to 2016, usually followed in relative abundance by cryptophytes and prasinophytes. Dinoflagellates were observed all along the sampling period, except in December 2013, with densities ranging from 80 to 40,000 cells L^{-1} , but representing a small

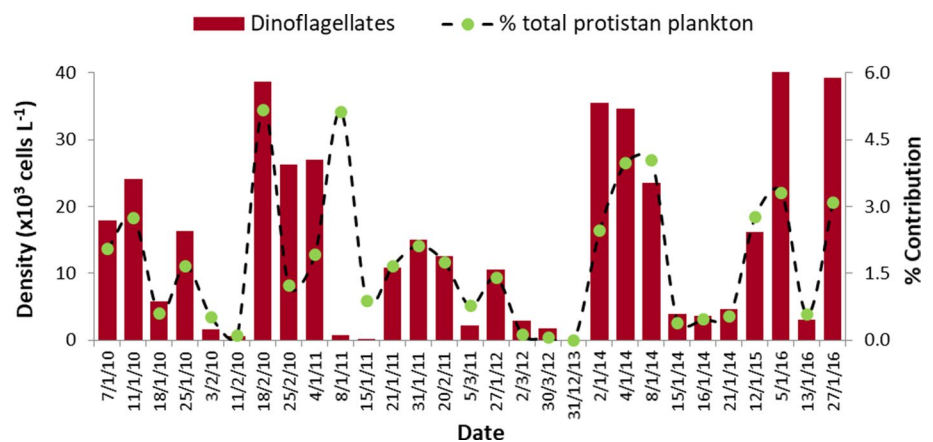
percentage (0.1–5.2%) of total cells (Fig. 6). Dinoflagellate assemblages were mostly dominated by small (< 20 μm) naked gymnodinioid forms, which represented an average of 61% of total dinoflagellate abundance. *Prorocentrum* aff. *balticum* was another conspicuous and frequent taxon, representing up to 90% of total dinoflagellate abundance on 21 January 2014. *Gyrodinium*, *Amphidinium* and *Protoperidinium* species were also usually observed in lower densities. By contrast, cells of potentially toxic species of *Dinophysis* were not detected in any of the analyzed samples.

Discussion

Environmental parameters

Environmental conditions in coastal waters of KGI generally do not favor dinoflagellate blooms due to high water column mixing as inferred from the Brunt-Väisälä frequency (Fig. S1) and observed by in situ measurements (Fig. 2). The water column in Potter Cove usually is well mixed throughout the year, except for short periods during

Fig. 6 Density of dinoflagellates (cells L^{-1}) and their relative contribution to total cell densities of protistan plankton based upon cell counts from Niskin bottle samples collected at Potter Cove during summer



the summer months, when glacial ice melt generates a surface fresh water layer. Water temperature was quite uniform throughout the year ranging from maxima of + 2 °C in summer to minima of – 2 °C in winter (Fig. 2a). Previous inorganic nutrient data show that neither nitrate nor phosphate or silicate limit phytoplankton growth in this area (Schloss et al. 2014) further characterized by high concentrations of bioavailable iron (Henkel et al. 2018). The prevailing mixing and irradiance conditions do not normally favor phytoplankton bloom developments (e.g. Schloss et al. 2002). In particular, cold and well-mixed water bodies generally do not favor high abundances of dinoflagellates (Margalef 1978; Bresnan et al. 2009). Nevertheless, the occurrence of PTX was observed during the entire course of the year 2014 (Fig. 4), which is very strong evidence that there is a constant, albeit low production of these phycotoxins in the area. Interestingly, no seasonality of PTX amounts on SPATTs was observed, which is probably due to the integrative sampling mechanism of the SPATT technique that only adsorbs extracellular phycotoxins. Considering that polyketides are chemically stable compounds that partially survive harsh physical and biological conditions such as boiling and shellfish metabolism, the presence of PTX on SPATTs throughout a 12 month period can be explained by a long persistence of this phycotoxin in the Antarctic cold-water system and is only an indirect proof of the presence of PTX-producing species and not directly linked to their spatial and temporal abundances. Other seasonal factors, such as the availability of the ciliate *Mesodinium rubrum*, the obligate prey of most, if not all *Dinophysis* species (Reguera et al. 2012), or light availability, which in high latitudes is reduced especially in winter and under ice coverage (Campana et al. 2018), likely play a role for the seasonal dynamic of *Dinophysis*, although mixotrophy and photophysiology of *Dinophysis* indicate that these species seem to be rather well adapted to low light intensities (Hansen et al. 2016).

Global change in Antarctica

The West Antarctic Peninsula is one of the fastest warming areas of the world and for this reason, an expansion of non-indigenous temperate taxa, including toxigenic species, may occur with raising temperatures under a global change scenario (Wells et al. 2015). Nonetheless, to our knowledge, only two studies have detected potentially toxigenic dinoflagellate species in Antarctic waters reporting *Alexandrium pacificum* (reported as *A. tamarense*) (Ho et al. 2003), later revised as Group IV of the *A. tamarense* species complex (Ho et al. 2012), and finally renamed as *A. pacificum* (John et al. 2014) and several species of the genus *Dinophysis* (Kuzmenko 2004). In contrast, in the Arctic, a region also highly impacted by global change and with evidence of increasing temperatures, most toxigenic protists known

from temperate regions have been found, including records of the associated phycotoxins (Okolodkov 2005; Vershinin et al. 2006; Gu et al. 2013; Harðardóttir et al. 2015; Lefebvre et al. 2016; Tillmann et al. 2016). This discrepancy could be due to the fact that Arctic waters are much more accessible and surrounded by inhabited land. For this reason, the high number of reports of toxigenic species and related phycotoxins in the Arctic may be a result of more rigorous monitoring programs and/or the introduction of toxigenic plankton by anthropogenic activities, an ongoing discussion also for other world regions (Mardones et al. 2017; Hallegraeff 2015). This lack of knowledge prompted the research of toxigenic species and phycotoxins in Antarctic waters.

The results presented here indicate very low abundance of *Dinophysis* (one cell of a PTX producer and one empty theca of an unidentified *Dinophysis*) and the unambiguous detection of dissolved PTX-2 and PTX-2sa in two different bays of KGI at a distance of app. 15 km. This finding opens the question if the presence of toxigenic species in Antarctic coastal waters is a result of the range expansion of species from neighboring areas of the Southern Ocean, as the barrier of the Circumpolar Current could be more permeable than previously assumed (Fraser et al. 2017, 2018). Alternatively, recent introduction by human activities might be another explanation (Hughes and Ashton 2016, Hughes et al. 2019). A third possibility is that phycotoxin-producing species always have been present in Antarctic waters at background levels. However, independent of the origin of this potentially toxic plankton species, the risk of phycotoxin proliferation with raising temperatures and increasing water stratification due to higher meltwater input to coastal Antarctic areas exists. Accordingly, this possible change in the specific composition of plankton assemblages may be a potential problem for local fauna such as birds, whales or seals, especially in an area, where the entire food chain depends on marine primary production.

Phycotoxin monitoring

No phycotoxins were detected in plankton concentrates of as much as six pooled plankton net hauls at any station in the two subsequent austral summer expeditions of 2013 and 2014. Nevertheless, the presence of phycotoxins on SPATTs is an indication that there is an at least very low presence of toxigenic plankton species in coastal waters of KGI. On the other hand, plankton species are known to show patchy temporal and spatial distributions, which also could explain the absence of phycotoxins in the summer net tow samples and is consistent with only one coincidental observation of *Dinophysis* in a water sample from Fildes Bay. But this record prompted the employment of a more integrative phycotoxin passive sampling technique instead of punctual microplankton sampling at discrete locations

and time points. SPATT originally has been developed in New Zealand for HAB monitoring to simulate phycotoxin accumulation in bivalves of shellfish harvesting areas by adsorption of lipophilic organic molecules dissolved in seawater (MacKenzie et al. 2004). SPATT samplers specifically adsorb large, lipophilic molecules dissolved in the water and thereby are ideal for the sampling of lipophilic phycotoxins with these properties. Due to this physical property, the low cost and easy applicability, its utilization rapidly has been expanded and successfully applied in several other countries (Shea et al. 2006; Lindahl et al. 2007; Fux et al. 2009; Li et al. 2010; García-Altare et al. 2014). Most phycotoxins are intracellular metabolites that are not actively excreted by healthy and well growing cells, but usually become part of the dissolved organic material (DOM) pool after cell death and lysis. Consequently, phycotoxin detection by SPATT samplers is an indirect proof of the presence of the producing species. Nevertheless, chemical stability of most marine biotoxins and their large molecular weight favor their selective adsorption to the adsorbent allowing for highly selective and sensitive detection of this class of biomolecules in seawater. Despite the above advantages, it has to be considered that SPATT sampling is not quantitative in a strict sense, even though phycotoxins can be precisely quantified on SPATT samplers. Adsorption is a reversible kinetic equilibrium process and adsorption kinetics for SPATT has not been investigated. Adsorption of phycotoxins also depends on several parameters, which cannot be controlled in the field, such as temperature, phycotoxin concentration, ratio of phycotoxins and other lipophilic components in the water, currents and other causes of water movement, bio-fouling, pH, salinity, among others, make an extrapolation of phycotoxin contents of SPATT samplers to analyze concentration in the seawater impossible. However, it can be assumed that when most of the above mentioned parameters are relatively constant during the sampling period, at least relative comparisons among values derived from SPATT under similar conditions can be made in a semi-quantitative way and in any case give a qualitative picture of phycotoxins present in a given area during the sampling period. Deployment times, mostly in austral winter during the one-year sampling in 2014/2015, were relatively variable due to bad weather conditions and poor accessibility of deployment sites to allow for quantitative comparisons among phycotoxin values of individual SPATT samplers. Nonetheless, the passive sampling clearly demonstrated that PTX was present in coastal waters of KGI all year round.

The presence of PTX-2 in turn is a strong indication for the presence of at least one of several PTX-producing species. To date, some *Dinophysis* species are the only known producers of PTX. There are several PTX variants reported in the literature (Sasaki et al. 1998; Miles et al. 2004), but PTX-2 dominates the PTX profiles of most species (Reguera

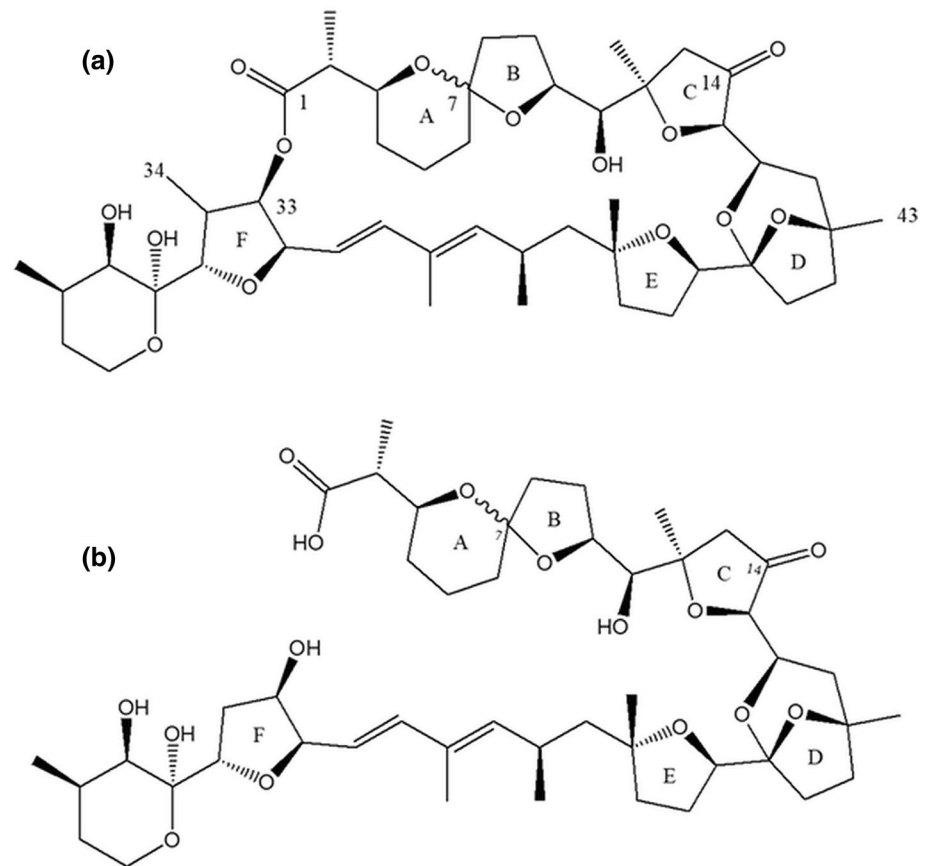
et al. 2014, and references herein), which seems also to be true for PTX-producing *Dinophysis* in the KGI area. The same observation was made during the one-year SPATT sampling in 2014/2015, where PTX-2 was detected on all SPATT samplers. In contrast to the short SPATT samplings during the summer 2013 expedition where only PTX-2 was detected, on the SPATT samplers exposed for longer times (usually more than 4 weeks, each) during the 2014/2015 sampling period also pectenotoxin-2 seco acid (PTX-2sa) in addition to PTX-2 was detected. PTX-2sa is the hydrolysed form of PTX-2, in which the lactone moiety forming the macrocycle of the molecule is opened by hydrolysis and results in a linear structure (Fig. 7). The levels of PTX-2/PTX-2sa in the Antarctic samples did not exceed 30 ng g^{-1} resin (Fig. 4, Table 3), which is one to two orders of magnitude below maximum values of PTX-2 found in areas with recurrent *Dinophysis* blooms such as in Southwest Ireland (220 ng g^{-1} resin) (Fux et al. 2010) and in Northwest Spain (3000 ng g^{-1} resin) (Pizarro et al. 2013). Even though SPATTs can only be regarded as semi-quantitative, these results clearly indicate a low abundance of PTX-producing species in coastal water of KGI.

PTX-2sa is hardly found in phytoplankton samples, but is a well-documented PTX-2 metabolite in mussels (Suzuki et al. 2001). The conversion of PTX-2 to PTX-2sa via hydrolysis is not only a result of metabolic shellfish activity, but also occurring spontaneously under high and low pH. The fact that PTX-2sa was detected at higher proportions on SPATTs with longer exposure times may be an indication that hydrolysis of PTX-2 may also be catalyzed by the surface of the HP20 resin in addition to pH driven hydrolysis in the marine environment. In addition to PTX, one SPATT contained yessotoxin (YTX) known to be produced by gonyaulacoid dinoflagellate species *Protoceratium reticulatum*, *Lingulodinium polyedra*, *Gonyaulax spinifera* (Paz et al. 2008) and *G. taylorii* (Álvarez et al. 2016). Although none of these species were detected in any of the phytoplankton samples analyzed by microscopy, *P. reticulatum* is a species known to occur and to be adapted to polar waters and is frequently found in the Arctic (Sala-Pérez et al. 2016). The absence of a producing species and the fact that there was only one positive SPATT sample for YTX is a weak indication for the presence of this phycotoxin. On the other hand, YTX has a relatively high detection limit and would hardly be detected at the concentrations that were found for PTX, which are detected more sensitively.

Protistan plankton

Using a molecular approach based on amplicon sequencing by NGS, a dominance of diatoms followed by dinoflagellates was detected in this study in Fildes Bay, KGI. There have been some reports of phytoplankton community

Fig. 7 Chemical structures of **a** pectenotoxin-2 (PTX-2) and **b** pectenotoxin-2 seco acid (PTX-2sa)



composition obtained using NGS in Antarctica (Luria et al. 2014; Wolf et al. 2015) and specifically in KGI (Luo et al. 2016; Moreno-Pino et al. 2016; Egas et al. 2017). Although it is known that NGS approaches for phytoplankton characterization have higher resolution than microscopy or pigments-based methods, especially in the case of small and inconspicuous protists, this approach depends on updated and curated databases that ensure that the annotation process of the sequences is accurate and precise. Yet, there are two important pitfalls for the application of this powerful tool in plankton and protist surveys. The first point is that the high variation of the 18S rRNA copy number in marine protistan species, the taxonomic marker traditionally used, can vary from one to hundreds of thousands copies per cell (Zhu et al. 2005). The second point is that the selection of a genetic marker that is sufficiently resolute in terms of taxonomy, but at the same time general enough to allow the identification of the entire protistan plankton community is difficult. In this study, it was found that most of the sequences assigned to Dinophyceae were undetermined dinoflagellates, which suggest that the second element is particularly important here, combined with a general lack of polar planktonic protists in molecular databases.

Dinoflagellates found by microscopy in Potter Cove during summers 2010–2016 were mainly dominated by small

naked species, here collectively grouped as “gymnodinoids”, as it usually occurs in other Antarctic areas (Nuccio et al. 2000; Korb et al. 2010). By contrast, no cells of *Dinophysis* were recorded during the examination of the 29 bottle samples, with a detection limit of 20 cells L^{-1} . This is in accordance with the molecular data presented here, which showed a great abundance of dinoflagellates in the plankton community but that were not able to identify *Dinophysis* sp. within its members. *Dinophysis* species are usually observed at very low cell densities and therefore are seldom found by routine microscopic observations (Reguera et al. 2012). Moreover, *Dinophysis* cells are present at higher densities in thin layers, which usually require the use of depth-integrated hose samples to be detected (Escalera et al. 2012). Additional sampling efforts considering a higher temporal resolution and a more detailed depth-integrated sampling of the water column could contribute to the detection of *Dinophysis* species in the study area.

According to the compendium of Antarctic dinoflagellates by McMinn and Scott (2005), around 10 species of *Dinophysis* including species currently regarded as *Phalacroroma* have been recorded in Antarctica. They mostly include insufficiently known species, such as *Dinophysis antarctica*, *D. contracta*, *D. cornuta*, *D. meteori*, *D. operculoides*, *D. punctata*, *D. tenuivelata* and *D. tuberculata*, which have

been never associated with phycotoxin production. In contrast to the Southern Ocean, where nothing is known about phycotoxin-producing species, toxigenic *Dinophysis* species are known to occur from tropical to temperate and boreal waters (Reguera et al. 2012). Thus, the single finding of *Dinophysis norvegica* in a Niskin bottle sample from Fildes Bay suggests that toxigenic species of the genus *Dinophysis* could also be present in Antarctic waters. This confirms a previous record from the Bransfield Strait also based on microscopy data by Kuzmenko (2004). In addition, the distribution of *D. norvegica* has been recently extended to the southern southwestern Atlantic (Argentine Sea), where it was recorded at temperatures from 6 to 8.9 °C (Fabro et al. 2016).

Conclusion

In this study, the first description of dissolved phycotoxins in Antarctic coastal waters gave indirect evidence of the presence of PTX-producing species and likely also YTX-producing species. This very low but recurrent detection of at least one phycotoxin in coastal waters of KGI, West Antarctic Peninsula indicates the potential risk of proliferation of HAB species in the Southern Ocean. Despite the finding of the presence of HAB species and their respective phycotoxins around KGI, there is a huge knowledge gap of the presence and geographic distribution of HAB species in other locations of the Southern Ocean and, certainly, a broader data base is needed for the development of a risk assessment of HAB events in Antarctica. Furthermore, phycotoxins can serve as sensitive chemotaxonomic markers for areas of very low HAB species abundances as they can be monitored by passive samplers, which is especially useful in remote areas with limited scientific infrastructure.

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Compliance with ethical standards

Conflict of interest All authors declare that they have no conflicts of interest.

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