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Response of cattle manure anaerobic digestion to zinc oxide  
nanoparticles: methane production, microbial community, and functions

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## Abstract

The increasing use of zinc oxide nanoparticles (ZnO NPs) as feed additives has raised huge environmental concerns in the anaerobic digestion (AD) of livestock wastes. In this study the 30-day effect of ZnO NPs on AD performance of cattle manure was investigated under optimal conditions (temperature=55°C, initial pH=10, total solids contents=10%), which were obtained from response surface methodology. Results showed that ZnO NPs (5-100 mg/g TS) promoted the accumulation of SCOD, SP and SC. Removal of SP and SC or production of VFAs were not significantly decreased in the presence of 5 mg ZnO NPs/g TS, but they were negatively affected as ZnO NPs increased to 30 and 100 mg/g TS. Besides, ZnO NPs had negative effects on VFAs consumption, which led to decreased methane production with 84.55% , 92.39% , and 93.72% in the presence of 5, 30 and 100 mg ZnO NPs /g TS, respectively. The shift of microbial community demonstrated that the decrease in the abundance of functional bacteria from 72.11% to 11.24% (in families *Ruminococcaceae* and *Lachnospiraceae*), () and of methanogens from 96.82% to <1% (in genus *Methanothermobacter*) () led to poor fermentation and methanogenesis, respectively. Functional analysis indicated that ZnO NPs can enhance abundances of genes related to AD like transport and metabolism of amino acid and carbohydrate, and energy conversion. The actual genetic expression, DNA integrity, cellular membrane, and intercellular communication may contribute to the low methane yield. This study provides new insights into livestock manure reduction and reutilization in the presence of exogenous pollutants.

## Keywords

Anaerobic digestion, functional analysis, livestock wastes, microbial community, zinc oxide

nanoparticles

## 1 **1 Introduction**

2 The rapidly developing intensive livestock industry in China has led to a series of  
3 environmental issues due to the occurrence of persistent pollutants (initially used as animal  
4 food additives) in livestock waste. The Chinese Government reported that the amount of  
5 chemical oxygen demand (COD) discharged from the livestock industry was increased from  
6 41.87% in 2007 to 46.67% in 2017 (Ministry of Ecological Environment of the PRC et al.,  
7 2010; Ministry of Ecological Environment of the PRC et al., 2020). Chinese production of  
8 livestock and poultry manure reached  $1.64 \times 10^{12}$  kg (fresh weight) in 2017 (Liu et al., 2020). ,  
9 which can cause substantial environmental and public health problems due to its content of  
10 heavy metals (Liu et al., 2020), antibiotics (Gaballah et al., 2021), excess nitrogen and  
11 phosphorus (Li et al., 2020; Sakadevan & Nguyen, 2017), if the manure is not treated properly.

12 Biogas can be generated from livestock manure with high organic matter mainly through  
13 anaerobic digestion (AD) process. The potential energy generated from manure-produced  
14 biogas in China was about  $5.74\text{-}6.73 \times 10^{12}$  MJ in 2017, which is equivalent to 4-5% of the  
15 country's energy demands (Wang et al., 2021b). Also, a high methane potential of livestock  
16 manure was estimated for Europe at 26 million  $\text{m}^3$  biomethane/year (Scarlat et al., 2018). AD  
17 has been considered as one of the most efficient technologies for biogas generation AD  
18 performance can be affected by many factors, including the co-existence of exogenous  
19 substances that are widely used as animal food additives (Luo et al., 2020).

20 Zinc oxide nanoparticles (ZnO NPs) have been widely used in many fields, such as  
21 cosmetic, medication, textile and automotive, and finally led to their ubiquitous occurrence in  
22 the environment (Jin & Jin, 2019). ZnO NPs with key advantages, such as the better

23 antibacterial and bioavailability properties, compared with conventional ZnO, were recently  
24 used as feed supplement in livestock industry (Fawzy et al., 2021). This has resulted in the  
25 presence of ZnO NPs in livestock manure with Zn concentrations ranging from 39.5 to 11379.0  
26 mg/kg in livestock manure (Hui Wang, 2013; Liu et al., 2020). Improper application or  
27 management of ZnO NPs can even deduce an extreme case which we must take into  
28 consideration.

29 On the one hand, addition of zinc ions ( $Zn^{2+}$ ) at 50-100 mg/L in an anaerobic co-  
30 digestion system was found to enhance COD removal rate and methanation.(Chan et al., 2019).  
31 On the other hand, it has been pointed out that ZnO NPs could significantly affect AD process  
32 by inhibiting the activity of methanogenesis due to the enhanced volatile fatty acids (VFAs)  
33 accumulation (Zhang et al., 2017b). Also, Zheng et al. investigated the short- and long- term  
34 effects of ZnO NPs concentrations (6, 30, and 150 mg/g TSS.) on AD performance, showing  
35 that inhibition of methanogenesis process could be observed even within short-term exposure,  
36 and would not mitigate over time (Zheng et al., 2015). Furthermore, the abundance of  
37 methanogenic archaea and enzyme activity could also be negatively influenced by the presence  
38 of ZnO NPs (Mu & Chen, 2011). The activities of protease, cellulase, acetated kinase, and  
39 coenzyme F-420 were adversely affected by ZnO NPs (30-150 mg/g VSS) in waste activated  
40 sludge AD system (Wang et al., 2021a). The toxic effects induced by ZnO NPs can be attributed  
41 not only to the released  $Zn^{2+}$ , which negatively impacts the substances transfer and enzyme  
42 system (Mu & Chen, 2011), but also to the intracellular reactive oxygen species (ROS) that are  
43 toxic to cytoplasmic lipids, proteins, and other intracellular intermediates (Sharma et al., 2009).  
44 Last but not least, other pollutants (e.g. norfloxacin, sulfamethazine, etc.) co-existing with ZnO

45 NPs in livestock manure could yield a greater impact on methane production than ZnO NPs  
46 alone (Zhao et al., 2019).

47 Existing literature is focusing on the effects of ZnO NPs on municipal sludge AD  
48 performance and to a lesser extent on livestock wastes. The influence of Zn on swine manure  
49 AD has been previously studied but with a focus on multi-pollutant effects and antibiotics  
50 resistance genes (Yang et al., 2020; Zhang et al., 2018; Zhang et al., 2017a). Studies illustrating  
51 the mechanisms of the effect of ZnO NPs on livestock manure are still missing elements in  
52 literature. Taking into account that municipal sludge differs a lot in terms of physical, chemical  
53 and biological characteristics from livestock wastes, as well as the extensive use of ZnO NPs  
54 in livestock industry, it is of particular importance to understand how ZnO NPs can influence  
55 livestock manure AD process.

56 AD is driven by microorganisms and highly depends on microbial structure and activity  
57 (Sasaki et al., 2011). Previous studies revealed that the shift of microbial community would be  
58 enhanced by nano-additions, and inhibition on methanogenesis contributed a lot to decreased  
59 methane production (Luo et al., 2020). However, another study found that occurrence of nano-  
60 materials could influence methane production but had no significant impact on microbial  
61 community structure during AD processes (Zhang et al., 2019). Owing to the development of  
62 proteomics and metabolomics technologies, the Clusters of Orthologous Groups of proteins  
63 (COGs) database provides a great tool of exploring underlying differences among functional  
64 proteins (Wang et al., 2019). Therefore, a comprehensive investigation of the effects of ZnO  
65 NPs on manure AD taking into consideration microbial communities as well as their functions  
66 can be carried out.

67 The influence of ZnO NPs on manure AD process was comprehensively investigated in  
68 this study. First, the optimal operational conditions were determined by response surface  
69 methodology (RSM) based on Box-Behnken design (BBD). Secondly, the effect of various  
70 concentrations of ZnO NPs on manure AD performance was evaluated. Also, the shift of  
71 microbial community was detected by means of high-throughput sequencing. To deeply reveal  
72 the biochemical mechanism of ZnO NPs on manure AD process, analysis of functional proteins  
73 was carried out.

## 74 **2 Materials and Methods**

### 75 **2.1 Nanoparticles and livestock wastes**

76 Cattle manure was collected from a manure tank in a farm in Sichuan Province, China,  
77 and then stored at -20 °C. The main characteristics of cattle manure were analyzed as follows:  
78 total solid (TS) of  $27.38 \pm 5.23\%$  (mass ratio), volatile solids (VS) of  $23.97 \pm 4.56\%$  (mass  
79 ratio), total organic carbon (TOC) of  $352.72 \pm 25.7$  mg/g TS.

80 Dry ZnO NPs (99.8%, metal basis) was purchased from Aladdin Reagent Co. Ltd., China,  
81 and the particle size ranged from 80 nm to 100 nm as shown by scanning electron microscopy  
82 (Figure S2). ZnO NPs suspension (10 g/L) was prepared by magnetic stirring and  
83 ultrasonication (1 h, 25 °C, 250 W, 40 kHz) according to the methods described elsewhere  
84 (Zhang et al., 2013).

### 85 **2.2 Response surface methodology design**

86 RSM was used to evaluate the effects of temperature (*A*), initial pH (*B*), and TS (*C*) on  
87 cattle manure AD performance. The range of those factors were set as follows: temperature of  
88 30-55 °C, initial pH of 6-10, and TS of 6-15%. BBD was used to optimize these three



89 independent variables (each at three levels) and 15 experiments with the actual form of three  
90 independent variables were conducted (Table S1). Methane yield ( $Y$ ) was set as the response  
91 and was measured by gas bags. The statistical design of experiments and analyses were  
92 performed using Design Expert (Version 8.0.6, Stat-Waes, Inc USA).

### 93 **2.3 Anaerobic digestion batch experiment**

94 AD was carried out using Automatic Methane Potential Test System II (Bioprocess  
95 Control, Sweden) equipped with serum bottles, under the optimal conditions obtained from  
96 RSM. Four reactors, marked as G0, G1, G2, and G3, were set up to deeply explore the effects  
97 of ZnO NPs concentrations (0, 5, 30 and 100 mg/g TS, respectively) on cattle manure AD  
98 performance. The working volume of each reactor was  $400 \pm 10$  mL with TS of 6%. 3 mmol  
99 NaOH was used to adjust the initial pH. Anaerobic conditions in the reactors were achieved by  
100 infusing nitrogen gas at the beginning of the batch experiments, and stratification was avoided  
101 by stirring reactors twice a day. Each AD experiment was carried out for 30 days and samples  
102 were collected for physico-chemical and microbial analyses on the 1<sup>st</sup>, 6<sup>th</sup>, 14<sup>th</sup>, 22<sup>nd</sup>, and 30<sup>th</sup>  
103 day of treatment. All AD batch experiments were run in parallel in triplicates ( $n=3$ ). The  
104 schematic diagram of the experimental setup is illustrated in Figure S3.

### 105 **2.4 High-throughput sequencing analysis**

106 The effect of ZnO NPs on the shift of microbial community structure in cattle manure AD  
107 process was also investigated using Illumina Miseq sequencing analysis (labeled G0, G1, G2,  
108 and G3, respectively). All samples were centrifuged for 10 min at 8000 rpm and then the Fast  
109 DNA Spin Kit for Soil (MP bio, USA) was used to extract DNA. The extracted genomic DNA  
110 was detected by 1% agarose gel electrophoresis. The total DNA was subjected to PCR

111 amplification using the primers 338F/806R and 524F10extF/Arch958RmodR. The primer sets  
112 of 338F (5'-ACTCCTACGGGAGGCAGCAG-3') and 806R (5'-  
113 GGACTACHVGGGTWTCTAAT-3') were used for analyzing bacterial strains, while the  
114 primers set of 524F10extF (5'-TGYCAGCCGCCGCGGTAA-3') and Arch958RmodR (5'-  
115 YCCGGCGTTGAVTCCAATT-3') were used for archaeal analysis (Liu et al., 2016; Xu et al.,  
116 2016). The processes of constructing 16S rRNA library and sequencing were conducted on the  
117 Illumina MiSeq PE 300 by Majorbio Co. (Shanghai, China) with three biological replicates.

118 The operational taxonomic unit (OTU) representative sequence was selected and the  
119 species information ID was obtained by comparing the Greengene database with BLASTn tool.  
120 After that, the software PICRUSt (Version 1.0.0) was used for the functional prediction of 16S  
121 rRNA gene data (Langille et al., 2013). EggNOG (<http://eggnog.embl.de/>) databases were  
122 employed to construct the abundance of COGs of microorganisms in manure AD in the  
123 presence of ZnO NPs concentrations.

## 124 **2.5 Chemical analyses**

125 The collected samples were centrifuged for 10 min at 8000 rpm and one part of the  
126 supernatant liquid was immediately filtered through 0.45  $\mu\text{m}$  membrane to measure its soluble  
127 chemical oxygen demand (SCOD), soluble protein (SP), soluble polysaccharides (SC) and  
128 ammonia nitrogen ( $\text{NH}_4^+\text{-N}$ ) content. The other part of the supernatant was filtered through a  
129 0.22  $\mu\text{m}$  membrane for VFAs determination after being acidized (to  $\text{pH} < 3$ ) with 10%  
130 phosphoric acid. TS, SCOD, and  $\text{NH}_4^+\text{-N}$  were measured according to the standard methods  
131 (APHA, 1995). SP and SC were analyzed by Lowry-Folin method and phenol-sulfuric method  
132 (Lowry et al., 1951; Dubois et al., 1951), respectively. VFAs were measured by high-

133 performance liquid chromatography (HPLC) equipped with UV detector (at 450 nm), and  
134 chromatographic column (SCR-101H, Shimadzu) based on a method developed previously  
135 (Sun et al., 2011). The total VFAs (TVFAs) concentration was obtained by summing up the  
136 concentrations of three VFAs, namely acetic acid, propionic acid, and butyric acid.

### 137 **3 Results and discussion**

#### 138 **3.1 Optimal condition for livestock manure AD**

139 RSM was applied to evaluate the individual and interactive effects of temperature (*A*),  
140 initial pH (*B*), and total solids (*C*) on methane yield in manure AD. The BBD matrix with  
141 replication of central runs (Run 07, 10, 13) was used to estimate the error in the model and in  
142 experimental runs and the detailed information of each tested factor is listed in Table S1. Based  
143 on results from 15 designed experiments, a second order regression equation was derived using  
144 coded factors for methane yield (*Y*) (Equation 1):

$$Y=287.93-19.312A+10.58B-3.05C+0.38AB+0.32AC-0.05BC+0.15A^2+0.12B^2+0.09C^2$$

145 (Equation 1)

146 Where *A*, *B*, *C* are the coded values of initial pH, temperature, and TS content, respectively.  
147 *Y* is predicted response of methane yield. The response surface contours and 3D figures are  
148 illustrated in Figure S1.

149 The second order regression model fitted experimental data significantly with *P* value of  
150 0.0101 and *F* of 10.10. The lack of fit was not significant with *F* of 17 and *P* value of 0.0561.  
151 It was observed that the linear and quadratic effects of temperature were significant, with its  
152 linear effect being more pronounced (*P* = 0.0005) than the quadratic effect (*P* = 0.0086) (Table  
153 S2), whereas other factors were not significant. Therefore, the effect of temperature on

154 methane yield was found to be the most significant parameter in manure AD process. This is  
155 consistent with a previous study dealing with the co-AD of cattle manure, canola residues, and  
156 inoculum (Safari et al., 2018).

157 Interactions among the three factors of initial pH, temperature, and TS are shown in  
158 response surface contours and 3D figures (Figure S1). The closer the contour line is, the greater  
159 the interaction between the two factors is. Response surfaces of initial pH-temperature (TS  
160 content=10.5%) and TS-temperature (initial pH=8) suggest that the effect of temperature on  
161 methane production was enhanced when TS content or initial pH increased. Also, surfaces of  
162 TS-temperature (initial pH=8) and TS-initial pH (temperature=42.5 °C) indicate that the effect  
163 of TS on methane yield was more significant when the initial pH declined or when the  
164 temperature increased. As for the initial pH, its impact on methane generation was more intense  
165 at high temperature or TS content, according to the response surfaces of initial pH-TS  
166 (temperature=42.5 °C) and temperature-initial pH (TS=10.5%).

167 The optimal conditions obtained from RSM modeling were as follows: temperature of  
168 55 °C, TS content of 6%, and initial pH of 10, which can result in 73.05 mL/g TS methane  
169 yield. This optimal initial pH (pH 10) differs from other studies that estimated optimal values  
170 at pH 7-8 (Zhai et al., 2015). This can be attributed to the fact that cattle manure, has higher  
171 buffer capacity compared to other substrates like sludge and food wastes, and was therefore  
172 easier to alleviate the negative effects created by VFAs and  $\text{NH}_4^+$  under the initial pH of 10  
173 (Xing et al., 2020). Besides, pH 10 could perform better in terms of decomposition of hard-  
174 degradable organic matter like lignin substances compared with other initial pH values (Ma et  
175 al., 2019). Validation of the findings was conducted by carrying out three parallel replicate

176 experiments under the temperature of 55 °C, TS content of 6%, and initial pH of 10, that  
177 resulted in methane yield of  $74.21 \pm 3.34$  mL/g TS, which indicates that the second order  
178 regression model (73.05 mL/g TS) fits well with the experimental data.

### 179 **3.2 Effects of ZnO NPs on AD performance**

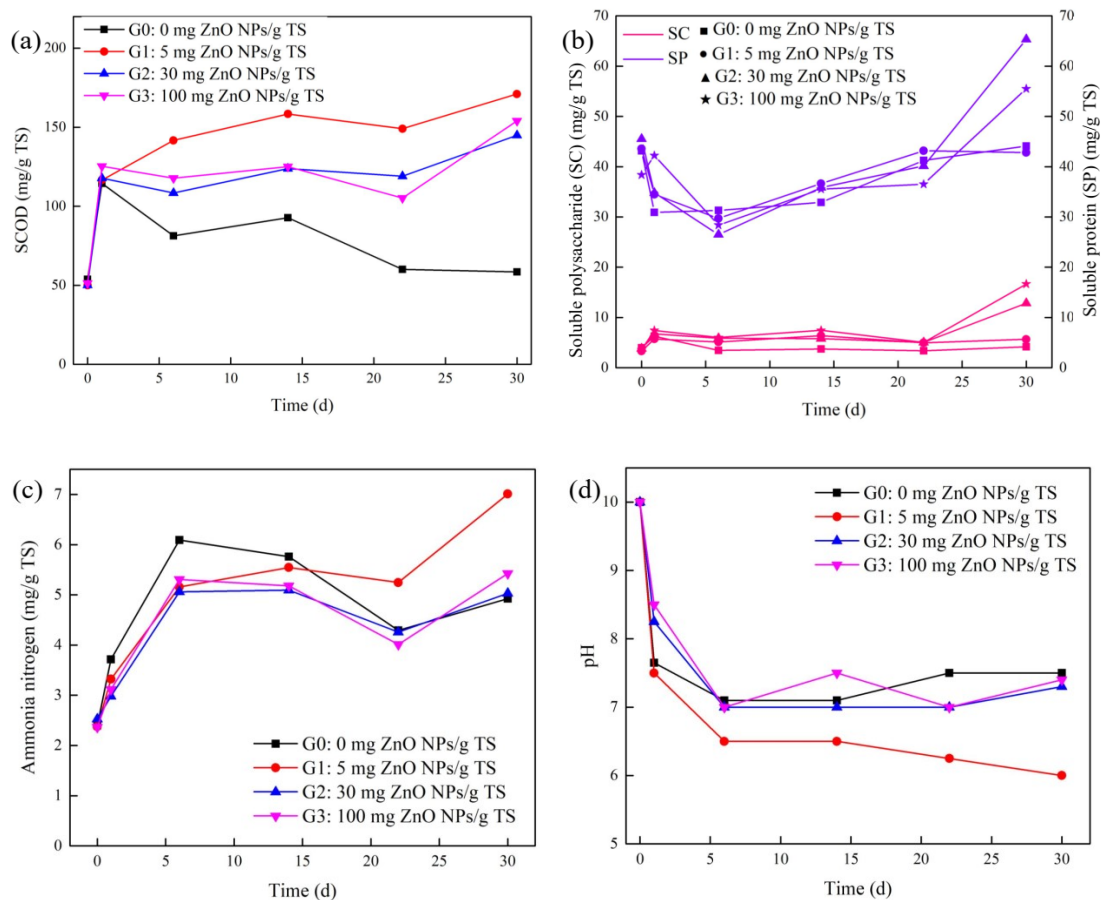
#### 180 **3.2.1 Effects on hydrolysis products**

181 Profiles of SCOD concentrations in the presence of different ZnO NPs amounts are shown  
182 in Fig. 1a. No obvious difference was observed among the four groups on the first day of  
183 digestion indicating that ZnO NPs had no acute inhibition on hydrolysis. As AD process  
184 progressed, the SCOD concentration in control group (G0, without ZnO NPs) was invariably  
185 lower than in the experimental groups (G1-G3, with ZnO NPs), while SCOD concentration in  
186 G1 group with a relatively low amount of ZnO NPs (5 mg/g TS ) was the highest among the  
187 four groups after the 1<sup>st</sup> day of treatment. The SCOD accumulation was much faster than its  
188 utilization in G1-G3 indicating that subsequent biochemical reactions were suppressed by ZnO  
189 NPs, while SCOD derived from organics solubilization and hydrolysis was enhanced (Fig. 1a).

190 Variation of SP and SC concentrations, two main hydrolysis products, were also monitored  
191 and are depicted in Fig. 1b. In the first 22 days, similar variation trend of SP was observed in  
192 all four groups, while SC concentrations in the experimental groups with 5-100 mg ZnO NPs/g  
193 TS were slightly higher than that in the control group. Also, significant accumulations of SP  
194 and SC in G2 and G3 were gained in the last day of digestion. Higher concentrations of SP and  
195 SC in G1-G3 groups suggest that ZnO NPs may promote the hydrolysis of organic matter on  
196 the one hand, and may inhibit SP and SC degradations on the other hand in such dynamic  
197 processes (Chen et al., 2020). Inhibition of ZnO NPs on SP degradation was also proved by

198 profiles of  $\text{NH}_4^+\text{-N}$  concentration illustrated in Fig. 1c which shows that concentrations of  
 199  $\text{NH}_4^+\text{-N}$  were lower in G2 and G3 than in G1 group. Besides, the lowest pH associated with SP  
 200 and SC decomposition was gained in G1 (Fig. 1d). This revealed that the inhibited effects on  
 201 SP and SC degradation were more severe in the presence of high ZnO NPs levels (30 and 100  
 202 mg/g TS) compared with G1 group.

203



204

205

206 **Fig. 1** Changes of SCOD (a), hydrolysis products (b), ammonia nitrogen (c), and pH (d) during  
 207 AD of cattle manure with different concentrations of ZnO NPs.

### 208 3.2.2 Effects on VFAs fermentation

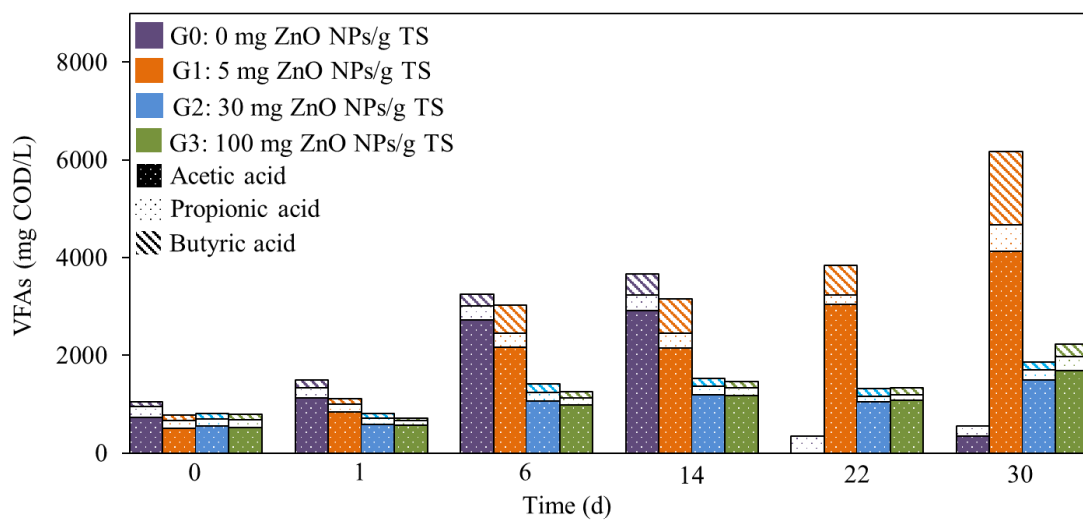
209 Since VFAs are vital substrates for methanogenesis in AD process, the variations of VFAs  
 210 concentration in the four groups were also investigated, in terms of TVFAs and VFAs

211 composition (Fig. 2). TVFAs concentration in the control group decreased after the 14<sup>th</sup> day,  
212 while in experimental groups showed an upward trend as digestion went on. The highest TVFAs  
213 accumulation occurred in G1 with the growth multiple of 6.91, followed by 1.81 and 1.31 in  
214 G3 and G2, respectively, on the 30<sup>th</sup> day of treatment (Fig. 2).

215 During the first two weeks, the TVFAs concentrations in the three experimental groups  
216 were lower than in the control group, which was probably attributed to the adverse effect on  
217 acidification led by ZnO NPs (Zheng et al., 2015). This is also in agreement with the lower SC  
218 and SP concentrations in G0. After the 14<sup>th</sup> day, concentrations of TVFAs kept increasing in  
219 G1-G3 experimental groups, especially in G1, whereas a significant decrease was observed in  
220 the control group. This indicated that VFAs consumption might be inhibited by the presence of  
221 ZnO NPs. Besides, TVFAs accumulation was most significant in G1 among the three  
222 experimental groups along with the lowest pH (Fig. 1d). Since VFAs are mainly produced from  
223 SP and SC, the lower accumulation of VFAs in G2 and G3 also indicates the inhibitory effects  
224 of ZnO NPs on SP and SC degradation (Chen et al., 2020), which also corresponded to the  
225 significant accumulation of SP and SC in groups with 30 (G2) and 100 (G3) mg/g TS of ZnO  
226 NPs (Fig. 1b).

227 Acetic acid is essential for methanogenesis and accounted for the largest proportion in  
228 TVFAs in all groups, followed by butyric and propionic acid (Fig. 2). Comparing the  
229 accumulation of individual acids up to the 14<sup>th</sup> day of treatment, it can be seen that acetic acid  
230 was higher in G0 than in the other groups containing ZnO NPs. This indicated that there might  
231 be an inhibition resulting from ZnO NPs on accumulation of acetic acid. Besides, no significant  
232 difference in propionic acid was showed among the control and the other experimental groups

233 during the first two weeks of treatment, which was consistent with the similar trends of SP,  
 234 whose most important product is propionic acid (Feng et al., 2009). In contrast, butyric acid  
 235 accumulation could be improved in G1 by the presence of ZnO NPs of 5 mg/g TS, probably  
 236 due to the suppressed conversion of butyric acid, an important intermediate in producing acetic  
 237 acid, at a certain concentration of ZnO NPs, and similar results were also obtained for sludge  
 238 AD process (Chen et al., 2020; Zheng et al., 2019).



239  
 240 **Fig. 2** Changes of TVFAs during AD of cattle manure in the presence of various concentrations  
 241 of ZnO NPs  
 242

### 243 3.2.3 Effects on methane production

244 Cumulative methane production (CMP) and daily methane production (DMP) were  
 245 monitored to evaluate the methanogenic activity and potential (Fig. 3). An inhibiting effect on  
 246 CMP was highly related to ZnO NPs concentration. The final CMP was decreased by 84.55%  
 247 (G1), 92.39%(G2), and 93.72% (G3), respectively, compared to the control group G0. Also,  
 248 CMP remained almost constant during the first 3 days of digestion and after that inhibitory  
 249 effects occurred on day 4 in G2 and G3 groups, which is earlier than day 6 when inhibition

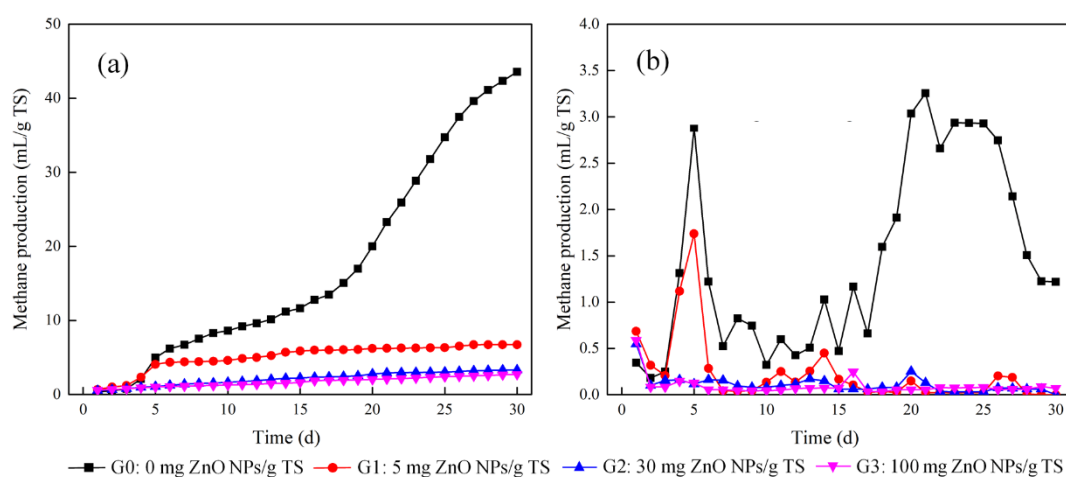


250 occurred for G1 group with the lowest ZnO NPs content. It can be suggested that a delayed  
251 inhibitory impact on CMP was induced by ZnO NPs, and such an inhibition showed positive  
252 correlation with ZnO NPs concentration within a certain range (0-30 mg ZnO NPs/g TS in this  
253 study).

254 To further investigate the effect of ZnO NPs on cattle manure AD performance, DMP  
255 variations in all experimental groups were also compared (Fig. 3b). It was found that DMP in  
256 the groups with ZnO NPs was higher than in the control group (0.35 mL/g TS) within the 1<sup>st</sup>  
257 day, with the highest methane yields being 0.69 mL/g TS, 0.55 mL/g TS, and 0.59 mL/g TS in  
258 G1, G2 and G3, respectively. This based on the profile of fermentation products in Fig. 2 might  
259 be attributed to two reasons. First, the appropriate and low amount of ZnO NPs (5 mg/g TS)  
260 could be beneficial to acetic acid formation and stimulate methanation in the short-term (within  
261 1 day). This could be proved by the higher increasing rate of acetic acid in G1 (64.04%) than  
262 that in G0 (52.72%). Secondly, the utilization of organic substrates might be stimulated by ZnO  
263 NPs within the 1<sup>st</sup> day. Furthermore, comparison of DMP, in this study, showed that the faster  
264 inhibition of methane production could be attributed to the presence ZnO NPs at concentrations  
265 higher than 5 mg/g TS.

266 In addition, two DMP peaks were observed on the 5<sup>th</sup> and 21<sup>st</sup> day in G0, while there was  
267 only one DMP peak on the 5<sup>th</sup> day in G1 with 5 mg ZnO NPs/g TS. Higher VFAs generation  
268 and conversion rates contributed to these peaks when taking into account changes of SCOD,  
269 SC, SP, and TVFAs concentrations, given in Fig. 1a, 1b, and 2. The lower peak value of methane  
270 yield in G1 on the 5<sup>th</sup> day was related to the lower VFAs production resulting from the inhibitory  
271 impact of ZnO NPs on the degradation of SC and SP (Fig. 1b). Large VFAs consumption in G0

272 led to the formation of another peak on the 21<sup>st</sup> day. Moreover, the accumulation of VFAs was  
 273 the reason for the inhibited methane yield in G1 group with 5 mg ZnO NPs/g TS, and can also  
 274 be proved by Fig. 1d that shows that the pH in G1 on the 21<sup>st</sup> day was 6.25, which does not  
 275 favour methane generation. In short, no significant effect on methanogenesis was induced by  
 276 ZnO NPs at the short-term (1 day of digestion), but severe inhibitions on methane production  
 277 did occur in the presence of high concentrations of ZnO NPs after 30 days of treatment



278  
 279 **Fig. 3** Cumulative methane production (a) and daily methane production (b) during AD of cattle  
 280 manure at various concentrations of ZnO NPs

### 282 3.3 Effects on microbial community

283 AD process is driven by microbes, thus high throughput sequencing was employed to  
 284 evaluate the influence of ZnO NPs on the microbial (bacterial and archaeal) community  
 285 structure in cattle manure AD process.

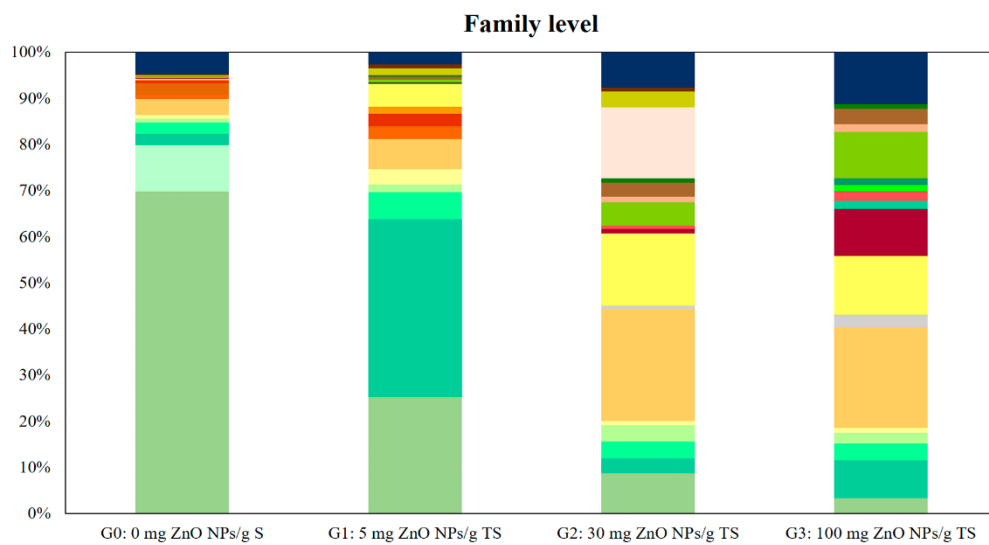
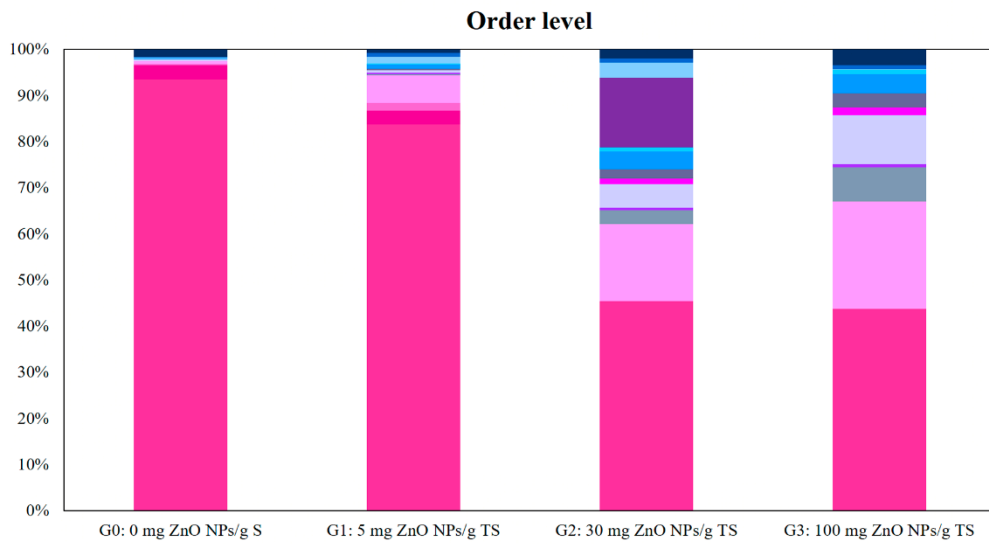
#### 286 3.3.1 The effects on bacterial community

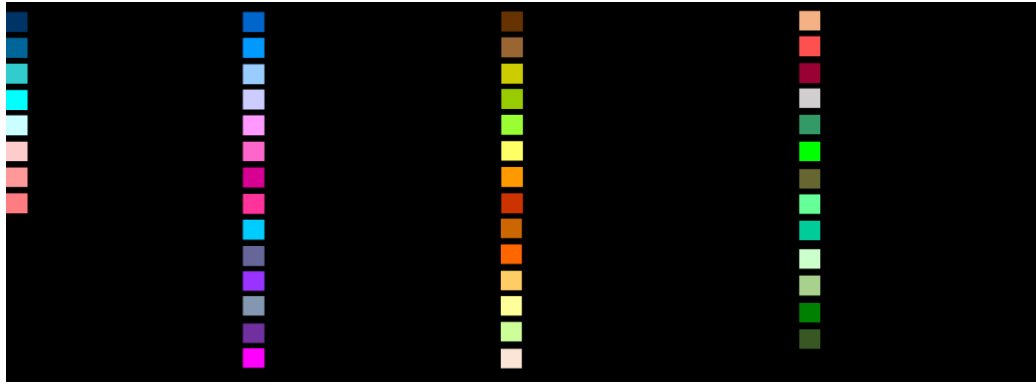
287 The comparison of the main bacterial community responsible for hydrolysis, acidogenesis,  
 288 acetogenesis, and hydrogenesis in cattle manure AD process was illustrated in Fig. 4. It was

289 observed that bacterial diversity was enhanced in the presence of ZnO NPs. Bacteria in classes  
290 *Clostridia*, *Bacilli*, *Alphaproteobacteria*, and *Actinobacteria* were dominant in all experimental  
291 groups. The relative abundance of bacteria in Classes *Bacilli*, *Actinobacteria* and  
292 *Alphaproteobacteria* was promoted by ZnO NPs. In contrast, percentages of bacteria belonging  
293 to class *Clostridia* in the four groups were 97.37% (G0), 88.23% (G1), 45.38% (G2), and 43.64%  
294 (G3) of total bacterial OTUs, respectively. Strains in *Clostridia* are capable of degradation of  
295 organic compounds and acid formation during AD (Yang et al., 2015). Its negative correlation  
296 with ZnO NPs concentrations might be the reason for the inferior SP and SC degradation and  
297 VFAs formation in G1-G3 groups with ZnO NPs (Fig. 1b and 2). At the order level, bacteria in  
298 order *Clostridiales* belonging to class *Clostridia* dominated in all groups with relative  
299 abundances of 93.32% (G0), 83.52% (G1), 45.38% (G2) and 43.64% (G3) of total bacterial  
300 OTUs, respectively, suggesting a negative relationship between bacterial abundance of  
301 *Clostridiales* and increasing ZnO NPs concentrations. The obvious decrease of bacterial  
302 abundance in order *Clostridiales* was also probably the reason for lower VFAs accumulation  
303 and CMP in G1-G3 groups with ZnO NPs. A previous study has also pointed out that strains in  
304 *Clostridiales* were widely related to VFAs production, and biogas production was directly  
305 connected with VFAs generation (Straeuber et al., 2016).

306 Furthermore, ZnO NPs also led to the change of the composition of bacteria belonging to  
307 order *Clostridiales* at the family level. *Lachnospiraceae* in order *Clostridia* was the highest in  
308 G1 (38.27%), while proportion of family *Ruminococcaceae* was negatively related to  
309 increasing ZnO NPs concentrations. Relative abundances of family *Ruminococcaceae* in all  
310 groups were 69.68% (G0), 25.05% (G1), 8.54% (G2), and 3.17% (G3), respectively. The

311 relative abundance of family *Family\_XI\_o\_\_Clostridiales* took a large part in order *Clostridiales*  
312 in G2 and G3. Strains in families *Ruminococcaceae* and *Lachnospiraceae* can both promote  
313 degradation of cellulose and hydrogen production, which play key roles in hydrolysis and  
314 hydrogenesis in AD process (Biddle et al., 2013). The large component of *Ruminococcaceae*  
315 (25.05%) and *Lachnospiraceae* (38.27%) in G1 (5 mg ZnO NPs/g TS) supported the substantial  
316 amounts of VFAs during anaerobic fermentation which is consistent with results shown in  
317 Figure 2. In contrast, the large proportion of 23.69% and 21.44%, of family  
318 *Family\_XI\_o\_\_Clostridiales*, in G3 and G4 groups, respectively, may spoil anaerobic  
319 fermentation performance with lower VFAs generation leading to SP and SC accumulation.





321

322 **Fig. 4** Relative abundance of bacteria at the class, order and family levels in AD of cattle manure  
 323 with different concentrations of ZnO NPs.

324

### 325 3.3.2 Effects on archaeal community

326 The comparison of the archaeal flora community among the four groups is shown in Fig.

327 5. Genera *Methanothermobacter*, *Methanobrevibacter*, *unclassified\_k\_\_noranked\_d\_\_archaea*,

328 and *Methanosphaera* dominated in all groups. Therefore, hydrogenotrophic methanogenesis

329 was the main pathway in thermophilic AD of cattle manure in this study. Although the diversity

330 did not significantly change in the presence of ZnO NPs, the archaeal community were shifted

331 in G0-G3. Strains in *Methanothermobacter* made up the main parts of archaea with 96.82% in

332 the control group. Archaea in the presence of 5 mg ZnO NPs/g TS was mainly consisted of

333 genera *Methanothermobacter* (22.45%), *Methanobrevibacter* (15.72%), and

334 *unclassified\_k\_\_noranked\_d\_\_archaea* (60.11%). Strains in genus *Methanobrevibacter* were

335 dominant in G2 (89.94%) and G3 (89.45%) with higher concentrations of ZnO NPs. Archaea

336 in *Methanothermobacter* mainly transform  $H_2/CO_2$  into methane, and archaea in

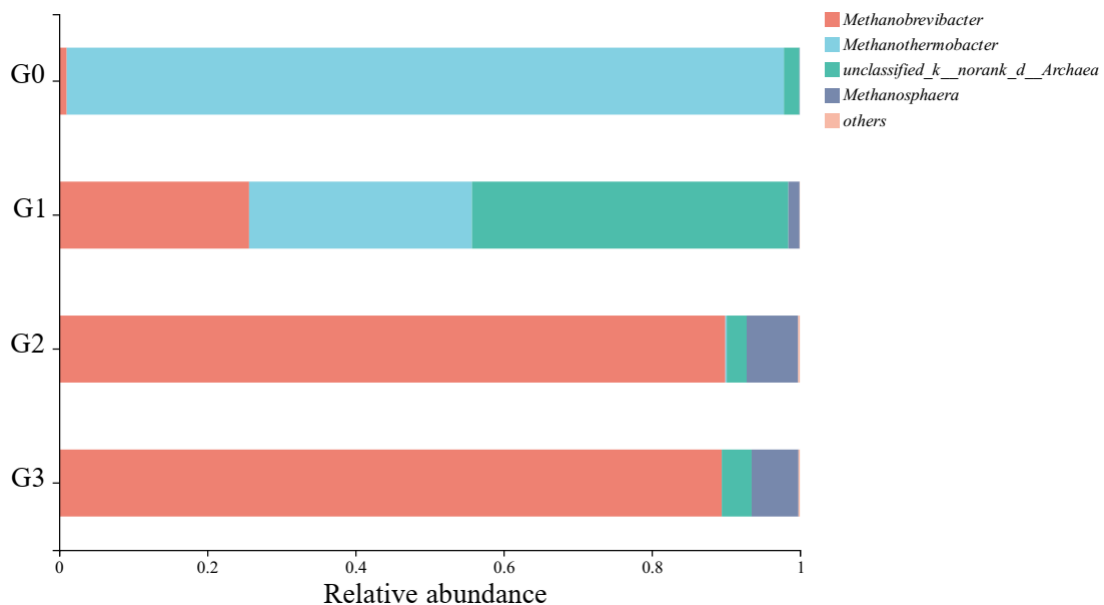
337 *Methanobrevibacter* uses  $H_2$  and/or formate as substrates (Danielsson et al., 2017; Liu et al.,

338 2019). Therefore, the variance of archaea indicated that the conversion of  $H_2/CO_2$  to methane

339 might be inhibited due to the presence of ZnO NPs in this study, since  $H_2/CO_2$  methanation was

340 found to be the fastest methanogenic step in digested manure system (Pan et al., 2016). A vast  
 341 array of methanogenic precursors were also responsible for the highest DMP in the control  
 342 group (Fig. 3b). The presence of F<sub>420</sub>H<sub>2</sub> oxidase in *Methanobrevibacter* species may be the  
 343 reason for its prosperity in G2 and G3 (Seedorf et al., 2004). The advantages under the oxidation  
 344 stress led by ZnO NPs will support a potential application in severe environment with higher  
 345 amount of ZnO NPs. However, it is known that strains of *Methanobrevibacter* can be key  
 346 methanogens in the AD process (Sun et al., 2021), which was inconsistent with methanogenesis  
 347 performance as shown in Fig. 3. Therefore, functional analysis was conducted for further  
 348 investigation.

349



350

351 **Fig. 5** Abundance of archaea at genus level in AD of cattle manure with different concentrations  
 352 of ZnO NPs

353

### 354 3.4 Functional analysis based on 16s RNA data

355 The variation among the abundances of COGs with different ZnO NPs concentrations has

356 been illustrated in Fig. 6. Functional proteins could be classified into three categories including  
357 metabolic pathways, information storage and processing, and cellular processes and signaling.

358 As for the bacterial COGs, amino acid transport and metabolism (AATM) and  
359 carbohydrate transport and metabolism (CTM) functions were related to SP and SC conversions.  
360 However, the variation of its abundances in the four groups was not consistent with SP and SC  
361 concentrations given in Fig. 1b. This might be attributed to the fact that although the abundances  
362 of functional proteins related to AATM and CTM could be boosted by ZnO NPs, the high ZnO  
363 NPs concentration probably inhibited its performance under exposure for 30 days. In this study  
364 the three concentration levels of ZnO NPs had similar effects on genetic abundances of AATM,  
365 while the abundance of genes coded CTM was highest in G1 with ZnO NPs of 5 mg/g TS.  
366 These results may explain the higher capacity of SC degradation compared with G2 and G3  
367 (Fig. 1b). Also, this can indicate that ZnO NPs at 30 and 100 mg/g TS may have greater negative  
368 effects on gene expression than at 5 mg/g TS. Further investigation about response of gene  
369 expression to ZnO NPs was required.

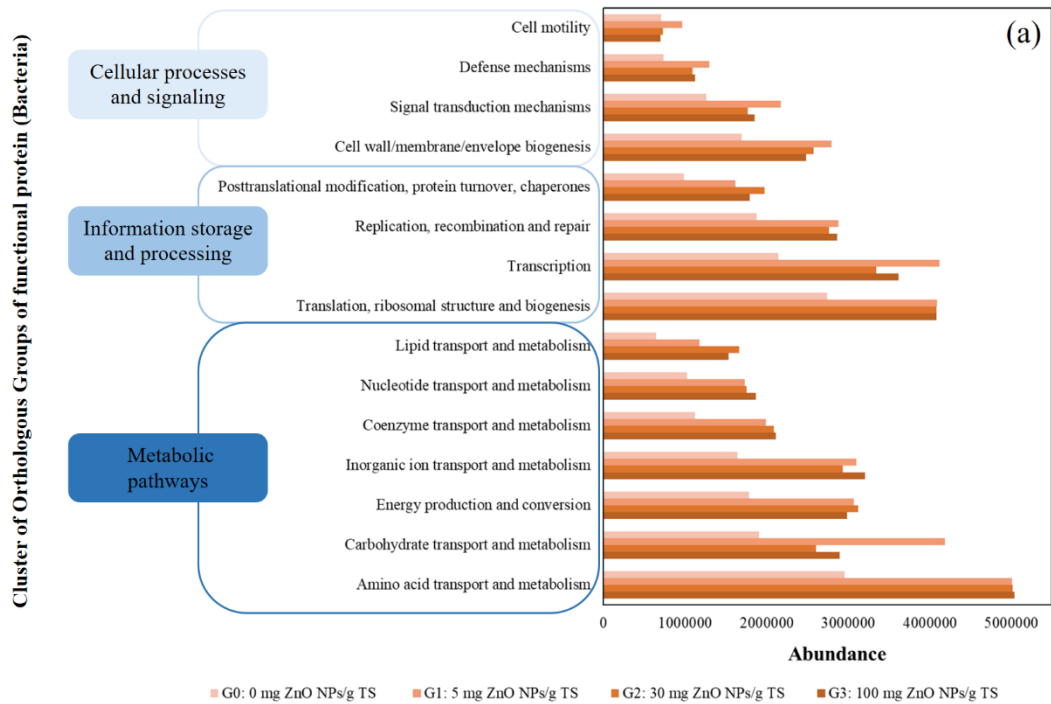
370 For archaea, abundances of COGs related to energy production and conversion, AATM,  
371 and coenzyme transport and metabolism were dominant in metabolism pathways. This was  
372 attributed to the fact that methanogens need to maintain special mechanism for efficient energy  
373 conservation with unique enzymes, electron carriers, and cofactors, in reaction to the inefficient  
374 synthesis of energy in methanogenesis using H<sub>2</sub>/CO<sub>2</sub>, formate, methanol, and acetate as  
375 substrates (Welte & Deppenmeier, 2014). The high abundances of COGs associated with energy  
376 conservation revealed the amount of energy demand when ZnO NPs are present in cattle  
377 manure. However, similarly to bacterial COGs, the higher genetic abundances related to



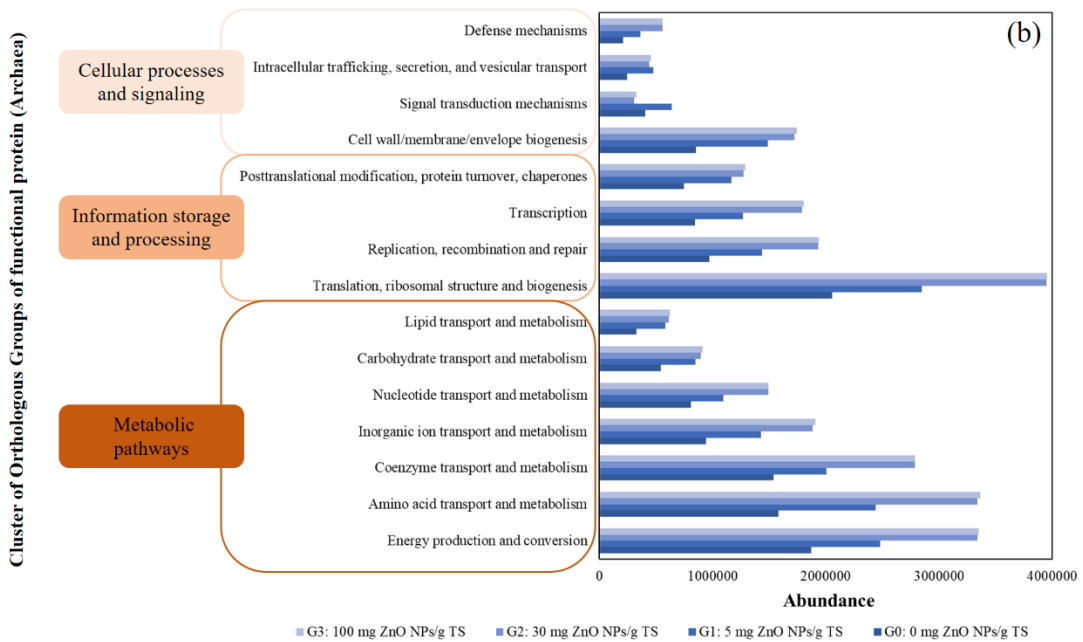
378 methanogenesis in G1-G3 were not corresponded with the lower methane yield shown in Fig.  
379 3. Therefore, although genes conferring functional proteins related to AD were boosted, the  
380 final genetic expression may be inhibited by ZnO NPs, and gene expression in methanogens  
381 were more sensitive to ZnO NPs than bacteria. This might also be the reason that considerable  
382 relative abundances of potential functional microbes in *Methanobrevibacter* and low methane  
383 yield occurred at the same time in this study.

384 Functional proteins involved in the category of information storage and processing are  
385 essential to maintain microbial vital activity (Lin et al., 2016). Total abundances of those  
386 proteins associated with transcription, translation, ribosomal structure and biogenesis, and  
387 replication, recombination and repair were relatively higher in groups with ZnO NPs, especially  
388 in G2 and G3 as shown in Fig.6. This may be attributed to DNA destruction by nanoparticles,  
389 which decreased normal performance of anaerobic fermentation and subsequent methanation  
390 (Lacerda et al., 2007).

391 Genes related to cellular processes and signaling mainly included cell motility, cell  
392 wall/membrane/envelope biogenesis, defense mechanism, and signal transduction metabolism.  
393 Current research has pronounced that the cell membrane transport system and the associated  
394 membrane metabolism are connected with cell recovery under ZnO NPs stress (Wu et al., 2017).  
395 Defense mechanism supports the tolerance for intercellular oxidative stress resulted from the  
396 presence of ZnO NPs (Yang et al., 2009). Signal transduction can be modified by Zn<sup>2+</sup> binding  
397 sites on proteins (Maret, 2006). Up-warded abundances for these functional genes in microbes  
398 indicated cellular membrane and intercellular communication may be damaged by ZnO NPs  
399 which are also responsible for the abnormal performance of AD.



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**Fig. 6** Abundance of the Cluster of Orthologous Groups of functional categories in the AD of cattle manure with different concentrations of ZnO NPs. COGs identified for bacteria (a) and archaea (b) are represented separately. The abundances of unknown functions and those less than 1% were excluded.

407 **Conclusion**

408 The role of ZnO NPs in cattle manure AD process has been comprehensively investigated.  
409 On the one hand, ZnO NPs (5-100 mg/g TS) promoted hydrolysis and the accumulations of  
410 organic matters. On the other hand, the production of VFAs, especially acetic acids, was  
411 inhibited in the presence of 30 and 100 mg ZnO NPs/g TS, while no inhibition was observed at  
412 5 mg ZnO NPs/g TS. As for methanogenesis, ZnO NPs (5-100 mg/g TS) exhibited severe  
413 impacts on methane production with decreasing rates of 84.55% (G1), 92.39% (G2), and 93.72%  
414 (G3). The poor performance of AD in the presence of NPs was a result of various microbial  
415 community and cellular damage. Further investigation of its actual expression should be carried  
416 out. Finally, other future work should involve the study of interactions of ZnO NPs with other  
417 pollutants, usually found in livestock wastes, and their effects on AD systems in order to  
418 resemble complexity of real-world applications.

419

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425 **References**

- 426 APHA (1995) Standard methods for the examination of water and wastewater. 19th ed.  
427 American Public Health Association, Washington DC.
- 428 Biddle, A., Stewart, L., Blanchard, J., Leschine, S.J.D. 2013. Untangling the Genetic Basis of  
429 Fibrolytic Specialization by Lachnospiraceae and Ruminococcaceae in Diverse Gut  
430 Communities. *5*(3), 627-640.
- 431 Chan, P.C., de Toledo, R.A., Iu, H.I., Shim, H. 2019. Effect of Zinc Supplementation on Biogas  
432 Production and Short/Long Chain Fatty Acids Accumulation During Anaerobic Co-  
433 digestion of Food Waste and Domestic Wastewater. *Waste and Biomass Valorization*,  
434 *10*(12), 3885-3895.
- 435 Chen, Y., Yang, Z., Zhang, Y., Xiang, Y., Xu, R., Jia, M., Cao, J., Xiong, W. 2020. Effects of  
436 different conductive nanomaterials on anaerobic digestion process and microbial  
437 community of sludge. *Bioresource Technology*, 304.
- 438 Danielsson, R., Dicksved, J., Sun, L., Gonda, H., Muller, B., Schnurer, A., Bertilsson, J. 2017.  
439 Methane Production in Dairy Cows Correlates with Rumen Methanogenic and Bacterial  
440 Community Structure. *Frontiers in Microbiology*, 8.
- 441 Fawzy, M., Khairy, G.M., Hesham, A., Rabaan, A.A., El-Shamy, A.G., Nagy, A. 2021.  
442 Nanoparticles as a novel and promising antiviral platform in veterinary medicine. *Arch*  
443 *Virol*, *166*(10), 2673-2682.
- 444 Feng, L., Chen, Y., Zheng, X. 2009. Enhancement of Waste Activated Sludge Protein  
445 Conversion and Volatile Fatty Acids Accumulation during Waste Activated Sludge  
446 Anaerobic Fermentation by Carbohydrate Substrate Addition: The Effect of pH.  
447 *Environmental Science & Technology*, *43*(12), 4373-4380.
- 448 Gaballah, M.S., Guo, J., Sun, H., Aboagye, D., Sobhi, M., Muhmood, A., Dong, R. 2021. A  
449 review targeting veterinary antibiotics removal from livestock manure management  
450 systems and future outlook. *Bioresour Technol*, *333*, 125069.
- 451 Hui Wang, Y.D., Yunya Yang, Gurpal S. Toor, Xumei Zhang. 2013. Changes in heavy metal  
452 contents in animal feeds and manures in an intensive animal production region of China.  
453 *Journal of Environmental Sciences*, *25*(12), 2435-2442.
- 454 Jin, S.-E., Jin, H.-E. 2019. Synthesis, Characterization, and Three-Dimensional Structure

455        Generation of Zinc Oxide-Based Nanomedicine for Biomedical Applications.  
456        *Pharmaceutics*, 11(11).

457        Lacerda, C.M.R., Choe, L.H., Reardon, K.F. 2007. Metaproteomic analysis of a bacterial  
458        community response to cadmium exposure. *Journal of Proteome Research*, 6(3), 1145-  
459        1152.

460        Langille, M.G.I., Zaneveld, J., Caporaso, J.G., McDonald, D., Knights, D., Reyes, J.A.,  
461        Clemente, J.C., Burkepille, D.E., Thurber, R.L.V., Knight, R., Beiko, R.G., Huttenhower,  
462        C. 2013. Predictive functional profiling of microbial communities using 16S rRNA marker  
463        gene sequences. *Nature Biotechnology*, 31(9), 814-+.

464        Li, S., Zou, D., Li, L., Wu, L., Liu, F., Zeng, X., Wang, H., Zhu, Y., Xiao, Z. 2020. Evolution  
465        of heavy metals during thermal treatment of manure: A critical review and outlooks.  
466        *Chemosphere*, 247.

467        Lin, Y.-W., Nguyen Ngoc, T., Huang, S.-L. 2016. Metaproteomic analysis of the microbial  
468        community present in a thermophilic swine manure digester to allow functional  
469        characterization: A case study. *International Biodeterioration & Biodegradation*, 115, 64-  
470        73.

471        Liu, C., Li, H., Zhang, Y., Si, D., Chen, Q. 2016. Evolution of microbial community along with  
472        increasing solid concentration during high-solids anaerobic digestion of sewage sludge.  
473        *Bioresource Technology*, 216, 87-94.

474        Liu, C., Mao, L., Zheng, X., Yuan, J., Hu, B., Cai, Y., Xie, H., Peng, X., Ding, X. 2019.  
475        Comparative proteomic analysis of *Methanothermobacter thermautotrophicus* reveals  
476        methane formation from H<sub>2</sub> and CO<sub>2</sub> under different temperature conditions.  
477        *Microbiologyopen*, 8(5).

478        Liu, W.R., Zeng, D., She, L., Su, W.X., He, D.C., Wu, G.Y., Ma, X.R., Jiang, S., Jiang, C.H.,  
479        Ying, G.G. 2020. Comparisons of pollution characteristics, emission situations, and mass  
480        loads for heavy metals in the manures of different livestock and poultry in China. *Sci Total*  
481        *Environ*, 734, 139023.

482        Luo, J., Zhang, Q., Zha, J., Wu, Y., Wu, L., Li, H., Tang, M., Sun, Y., Guo, W., Feng, Q., Cao,  
483        J., Wang, D. 2020. Potential influences of exogenous pollutants occurred in waste  
484        activated sludge on anaerobic digestion: A review. *Journal of Hazardous Materials*, 383.

485 Ma, S., Hu, H., Wang, J., Liao, K., Ma, H., Ren, H. 2019. The characterization of dissolved  
486 organic matter in alkaline fermentation of sewage sludge with different pH for volatile  
487 fatty acids production. *Water Research*, 164.

488 Maret, W. 2006. Zinc coordination environments in proteins as redox sensors and signal  
489 transducers. *Antioxidants & Redox Signaling*, 8(9-10), 1419-1441.

490 Mu, H., Chen, Y. 2011. Long-term effect of ZnO nanoparticles on waste activated sludge  
491 anaerobic digestion. *Water Research*, 45(17), 5612-5620.

492 Pan, X., Angelidaki, I., Alvarado-Morales, M., Liu, H., Liu, Y., Huang, X., Zhu, G. 2016.  
493 Methane production from formate, acetate and H<sub>2</sub>/CO<sub>2</sub>; focusing on kinetics and  
494 microbial characterization. *Bioresource Technology*, 218, 796-806.

495 Safari, M., Abdi, R., Adl, M., Kafashan, J. 2018. Optimization of biogas productivity in lab-  
496 scale by response surface methodology. *Renewable Energy*, 118, 368-375.

497 Sakadevan, K., Nguyen, M.L. 2017. Livestock Production and Its Impact on Nutrient Pollution  
498 and Greenhouse Gas Emissions. in: *Advances in Agronomy*, Vol 141, (Ed.) D.L. Sparks,  
499 Vol. 141, pp. 147-184.

500 Sasaki, K., Morita, M., Sasaki, D., Nagaoka, J., Matsumoto, N., Ohmura, N., Shinozaki, H.  
501 2011. Syntrophic degradation of proteinaceous materials by the thermophilic strains  
502 *Coprothermobacter proteolyticus* and *Methanothermobacter thermautotrophicus*. *Journal*  
503 *of Bioscience and Bioengineering*, 112(5), 469-472.

504 Scarlat, N., Fahl, F., Dallemand, J.-F., Monforti, F., Motola, V. 2018. A spatial analysis of biogas  
505 potential from manure in Europe. *Renewable & Sustainable Energy Reviews*, 94, 915-930.

506 Seedorf, H., Dreisbach, A., Hedderich, R., Shima, S., Thauer, R.K. 2004. F<sub>420</sub>H<sub>2</sub> oxidase  
507 (FprA) from *Methanobrevibacter arboriphilus*, a coenzyme F<sub>420</sub>-dependent enzyme  
508 involved in O<sub>2</sub> detoxification. *Archives of Microbiology*, 182(2-3), 126-137.

509 Sharma, V., Shukla, R.K., Saxena, N., Parmar, D., Das, M., Dhawan, A. 2009. DNA damaging  
510 potential of zinc oxide nanoparticles in human epidermal cells. *Toxicology Letters*, 185(3),  
511 211-218.

512 Straeuber, H., Lucas, R., Kleinstueber, S. 2016. Metabolic and microbial community dynamics  
513 during the anaerobic digestion of maize silage in a two-phase process. *Applied*  
514 *Microbiology and Biotechnology*, 100(1), 479-491.

515 Sun, H., Yang, Z., Shi, G., Arhin, S.G., Papadakis, V.G., Goula, M.A., Zhou, L., Zhang, Y., Liu,  
516 G., Wang, W. 2021. Methane production from acetate, formate and H<sub>2</sub>/CO<sub>2</sub> under high  
517 ammonia level: Modified ADM1 simulation and microbial characterization. *Science of the*  
518 *Total Environment*, 783.

519 Sun, Z.-Y., Tang, Y.-Q., Iwanaga, T., Sho, T., Kida, K. 2011. Production of fuel ethanol from  
520 bamboo by concentrated sulfuric acid hydrolysis followed by continuous ethanol  
521 fermentation. *Bioresource Technology*, 102(23), 10929-10935.

522 Wang, H., Li, H.-X., Fang, F., Guo, J.-s., Chen, Y.-P., Yan, P., Yang, J.-X. 2019. Underlying  
523 mechanisms of ANAMMOX bacteria adaptation to salinity stress. *Journal of Industrial*  
524 *Microbiology & Biotechnology*, 46(5), 573-585.

525 Wang, S., Chen, L., Yang, H., Liu, Z. 2021a. Influence of zinc oxide nanoparticles on anaerobic  
526 digestion of waste activated sludge and microbial communities. *Rsc Advances*, 11(10),  
527 5580-5589.

528 Wang, Y., Zhang, Y., Li, J., Lin, J.G., Zhang, N., Cao, W. 2021b. Biogas energy generated from  
529 livestock manure in China: Current situation and future trends. *J Environ Manage*, 297,  
530 113324.

531 Welte, C., Deppenmeier, U. 2014. Bioenergetics and anaerobic respiratory chains of aceticlastic  
532 methanogens. *Biochimica Et Biophysica Acta-Bioenergetics*, 1837(7), 1130-1147.

533 Wu, J., Lu, H., Zhu, G., Chen, L., Chang, Y., Yu, R. 2017. Regulation of membrane fixation and  
534 energy production/conversion for adaptation and recovery of ZnO nanoparticle impacted  
535 *Nitrosomonas europaea*. *Applied Microbiology and Biotechnology*, 101(7), 2953-2965.

536 Xing, B.-S., Han, Y., Wang, X.C., Ma, J., Cao, S., Li, Q., Wen, J., Yuan, H. 2020. Cow manure  
537 as additive to a DMBR for stable and high-rate digestion of food waste: Performance and  
538 microbial community. *Water Research*, 168.

539 Xu, N., Tan, G., Wang, H., Gai, X. 2016. Effect of biochar additions to soil on nitrogen leaching,  
540 microbial biomass and bacterial community structure. *European Journal of Soil Biology*,  
541 74, 1-8.

542 Yang, C., Zhou, A., He, Z., Jiang, L., Guo, Z., Wang, A., Liu, W. 2015. Effects of ultrasonic-  
543 assisted thermophilic bacteria pretreatment on hydrolysis, acidification, and microbial  
544 communities in waste-activated sludge fermentation process. *Environmental Science and*

545 Pollution Research, 22(12), 9100-9109.

546 Yang, H., Liu, C., Yang, D., Zhang, H., Xi, Z. 2009. Comparative study of cytotoxicity,  
547 oxidative stress and genotoxicity induced by four typical nanomaterials: the role of particle  
548 size, shape and composition. *Journal of Applied Toxicology*, 29(1), 69-78.

549 Yang, S., Wen, Q., Chen, Z. 2020. Impacts of Cu and Zn on the performance, microbial  
550 community dynamics and resistance genes variations during mesophilic and thermophilic  
551 anaerobic digestion of swine manure. *Bioresource Technology*, 312.

552 Zhai, N., Zhang, T., Yin, D., Yang, G., Wang, X., Ren, G., Feng, Y. 2015. Effect of initial pH  
553 on anaerobic co-digestion of kitchen waste and cow manure. *Waste Management*, 38, 126-  
554 131.

555 Zhang, J., Wang, Z., Lu, T., Liu, J., Wang, Y., Shen, P., Wei, Y. 2019. Response and mechanisms  
556 of the performance and fate of antibiotic resistance genes to nano-magnetite during  
557 anaerobic digestion of swine manure. *Journal of Hazardous Materials*, 366, 192-201.

558 Zhang, L., Li, Y., Liu, X., Zhao, L., Ding, Y., Povey, M., Cang, D. 2013. The properties of ZnO  
559 nanofluids and the role of H<sub>2</sub>O<sub>2</sub> in the disinfection activity against *Escherichia coli*. *Water*  
560 *Research*, 47(12), 4013-4021.

561 Zhang, R., Gu, J., Wang, X., Zhang, L., Tuo, X., Guo, A. 2018. Influence of combined  
562 sulfachloropyridazine sodium and zinc on enzyme activities and biogas production during  
563 anaerobic digestion of swine manure. *Water Science and Technology*, 77(11), 2733-2741.

564 Zhang, R., Wang, X., Gu, J., Zhang, Y. 2017a. Influence of zinc on biogas production and  
565 antibiotic resistance gene profiles during anaerobic digestion of swine manure.  
566 *Bioresource Technology*, 244, 63-70.

567 Zhang, Z.-Z., Xu, J.-J., Shi, Z.-J., Cheng, Y.-F., Ji, Z.-Q., Deng, R., Jin, R.-C. 2017b. Short-  
568 term impacts of Cu, CuO, ZnO and Ag nanoparticles (NPs) on anammox sludge: CuNPs  
569 make a difference. *Bioresource Technology*, 235, 281-291.

570 Zhao, L., Ji, Y., Sun, P., Deng, J., Wang, H., Yang, Y. 2019. Effects of individual and combined  
571 zinc oxide nanoparticle, norfloxacin, and sulfamethazine contamination on sludge  
572 anaerobic digestion. *Bioresource Technology*, 273, 454-461.

573 Zheng, L., Zhang, Z., Tian, L., Zhang, L., Cheng, S., Li, Z., Cang, D. 2019. Mechanistic  
574 investigation of toxicological change in ZnO and TiO<sub>2</sub> multi-nanomaterial systems during



575 anaerobic digestion and the microorganism response. *Biochemical Engineering Journal*,  
576 147, 62-71.

577 Zheng, X., Wu, L., Chen, Y., Su, Y., Wan, R., Liu, K., Huang, H. 2015. Effects of titanium  
578 dioxide and zinc oxide nanoparticles on methane production from anaerobic co-digestion  
579 of primary and excess sludge. *Journal of Environmental Science and Health Part a-  
580 Toxic/Hazardous Substances & Environmental Engineering*, 50(9), 913-921.

581

582