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GREEN SYNTHESIS OF SILVER NANOPARTICLES USING MICROALGAL EXTRACT AND ITS APPLICATION IN METAL ION REMOVAL FROM AQUEOUS SOLUTION

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ABSTRACT

Heavy metal pollution is of great concern and cannot be ignored as heavy metals are highly mobile in soil and are recalcitrant. Nanotechnology provides a novel sustainable approach for synthesizing materials of desired properties, composition, and structure, it is however expected to adsorb heavy metals and play a significant role in water treatment. Green chemistry is the cost-effective, non-toxic, and environment friendly approach that involves the use of biological components as reducing and stabilizing agents for the synthesis of nanoparticles. In the present study, heavy metals such as cadmium (Cd II) and lead (Pb II) were successfully removed from its aqueous solution by an adsorption process using the silver nanoparticles of size ~15nm biosynthesized using freshwater algal extract. The adsorption capacity (23.98 mg/g) was observed for Pb (II). The higher value of R² showed that the experimental data were fitted best with Langmuir isotherm. The rate kinetics study showed that Pb (II) was attached to the nanoparticles surface through electrostatic force of attraction, also referred to as chemisorption whereas Cd (II) adsorption on Ag nanoparticles followed pseudo-first order kinetics (R²>0.8) indicating physical adsorption between adsorbate and adsorbent.

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1 Introduction

Recent Advancement in nanotechnology has enabled researchers to further explore the unique physical and chemical properties of the nanomaterial. This has expanded its application in various domains such as biomedicine, catalysis, energy storage, and sensors (Xu et al., 2020). Nanoparticles can be engineered by methods such physical as irradiation, diffusion, thermal decomposition, and chemical methods such as condensation, reduction, sol-gel method precipitation, laser pyrolosis, chemical reduction method etc. (Sandhu et al., 2019). However, the use of these methods for the synthesis of nanoparticles are known to impact human health and the environment adversely as it involves the use of hazardous chemicals and releases harmful radiations. In this view, researchers have devised green chemistry as a new approach for the synthesis of nanomaterials that involves the use of biological entities such as plants and microorganism (Mandal et al., 2021). For the synthesis of nanoparticles, bio-components such as cellulose, enzymes, exopolysaccharide, flavonoids, alkaloids, terpenoids, alcohol etc. are responsible for the bio-reduction of metal salt to its elemental form (Xu et al., 2020). The microorganism mediated synthesis of nanoparticles offers particles of smaller size, high surface area to volume ratio, thermal, magnetic, catalytic, and biological properties which pave the way to synthesize different metal and metal oxide nanoparticles successfully (Sandhu et al., 2019). However, the potential contribution of different macromolecules in the formation of nanoparticles is still not known (Karthik et al., 2020).

Furthermore, various nanoparticles namely metal nanoparticles, carbon-based nanoparticles, ceramic nanoparticles, semi-conductor nanoparticles, polymeric, and lipid based nanoparticles have been engineered and are available in different compositions, shapes, sizes, morphologies, chemical, and physical properties. Among different types of noted nanoparticles, metal nanoparticles have drawn a considerable amount of attention due to their exceptional properties including, optical properties, magnetic properties, mechanical strengths, high surface area, high ordered structure, and thermal strength (Abegunde et al., 2020). The kinetics of metal nanoparticles such as silver nanoparticles (AgNPs) is slower, and better manipulation is achieved by crystal growth and stabilization, hence nanoparticles of desired shape and size can be obtained for their wide application for different purposes (Dağlıoğlu & Öztürk, 2019).

Recently, various treatment technologies such as photo–Fenton processes, chemical coagulation/flocculation, have been adopted for the removal of pollutants from water and wastewater (Khalid et al., 2017), chemical precipitation (Zhang & Duan, 2020), nano-filtration (Zhao et al., 2021), ultra-filtration (Yousefi et al., 2021), and ion exchange (Goyal et al., 2021). However, factors such as high operational cost, rigorous use of chemicals, and poor sludge

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Microalgae are well known for tremendous rapid growth and double their biomass around 10 times quicker than higher plants (El-Sheekh & El-Kassas, 2016). In general, algae are divided taxonomically into diverse groups such as, Phaeophyta, Chrysophyta, Bacillariophyta, , Rhodophyta, Chlorophyta, Cyanophyta and Dinophyta; and containing essential biomolecules such as antioxidants, polysaccharides, carotenoids, pigments, amino acids, proteins, minerals, lipids, phytohormones, immune stimulators as well as vitamins (El-Sheekh et al., 2020) that plays a major role in reduction and accumulation of molecules. Therefore, in the present study, we aimed to examine the potential of freshwater microalgae *Tetradesmus* sp. to produce silver nanoparticles and to determine the efficiency of synthesized silver nanoparticles in adsorption of Pb (II) and Cd (II) metal ions from its aqueous solution.

2 Materials and methods

2.1 Chemicals and reagents

All the chemicals and reagents used in the present study were of analytical grade.

2.2 Phycosynthesis of AgNPs

Patel et al. (2015) methodology was adopted with slight modifications for the biosynthesis of Ag nanoparticles using microalgae. The freshwater microalgae sample was collected from freshwater lake during the summer season. The collected strains were isolated from the freshwater sample by serial dilution method and were cultivated in BG11 medium. 100µl of the freshwater sample was spreaded on agar plates and was incubated for 5-10 days at 28±1°C at 3000 lux light intensity and 16:8 hr light and dark cycle to isolate pure colonies. The growth of microalgae was observed within 2 weeks subsequently three different colonies were picked up. The streak plate method was used to obtain pure colonies of microalgae and then single colony was further transferred in afresh medium (BG-11 broth) at temperature 28±1°C to maintain the purity of the culture. The flask containing microalgal inoculum was illuminated with 3000-4000 lux light intensity and L: D cycles of 16:8 hrs. Proper aeration was provided throughout the growth period. The morphological identification of isolated microalgal strain was carried out microscopically and identify by the taxonomic key given by Prescott (1978). The selected isolate was outsourced to determine the microalgal strain at the molecular level. Microalgae were harvested on the 14th day of their growth period and centrifuged at 5000 rpm for 10 min. Pellets were removed from the supernatant and the former was repeatedly washed with distilled water to remove the traces of culture media. Crude extract of algae was prepared using water as an aqueous medium as it is one of the highly polar solvent. The pure microalgal pellets were dried at room temperature ($28 \pm 1^{\circ}$ C) and were crushed to obtain a fine powder. For the extraction process, 1.0 g of dried biomass powder was mixed with 100 ml distilled water and heated at 60°C for 25 min using a hot plate with magnetic stirrer. The mixture was allowed to stand at room temperature for cooling; subsequently, filtration was done using 0.2-micron pore size filter and the resultant extract was stored in the refrigerator at 4°C for further use. For the biosynthesis of AgNPs 1mM of silver nitrate (AgNO3) was used as a precursor agent. The aqueous extract was added drop-wise to 80 ml of 1mM AgNO3 solution till the colorless solution turned into brown color, also the solution was shaken well to ensure the complete mixing of the components. The total volume of extract utilized was found to be 16.5 ml; double distilled water was added to make up the volume to 100 ml. The mixture was kept in dark to prevent the photoactivation of AgNO3. 1mM AgNO3 ⁺. Aqueous extract of microalgae were mixed homogeneously and reduction of Ag+ to Capped zerovalent Ag (Ag⁰) was carried out followed by Spray drying of colloidal solution of nanoparticles.

2.3 Characterization of biosynthesized nanoparticles

High resolution SEM analysis (JEOL, JSM-6490LA) of synthesized nanoparticles was carried out at an accelerating voltage of 20 KV. The samples were dried for 24 hrs to remove moisture content. The powdered samples were placed in the grid and the high quality images were obtained at different resolutions. A powerful analytical system transmission electron microscopy of JEOL, 2100F model was used to identify the shapes and sizes of the biosynthesized nanoparticles. TEM images were used to determine the size of the nanoparticles with the help of image J software. Nanoparticles were selected randomly to create a Gaussian size distribution histogram. The XRD patterns for synthesized nanoparticles were obtained using D8-Advance, Bruker X-ray diffractometer included Cu K alpha as a source of radiation (equals to 1.5406 Å) and LINXEYE XE as a detector. The instrument was operated at a voltage 40 kV with 40 mA functional current. The powdered NPs samples were placed on the sample holder carefully and were analyzed using a 2θ range from 20° to 80°. The obtained patterns were compared with ICDD-JCPDS (International Centre for Diffraction Data-Joint Committee of Powder Diffraction Standards) to determine the crystallinity of nanoparticles. The other fractions may be originated due to the presence of impurities of organic components in the samples. In FTIR spectroscopy, a spectra of % transmission of radiation is obtained when a beam of infrared radiation passes through the sample. In the present study, the Nicolet-6600 FT-IR spectrometer model was used to determine the IR spectra of the samples using wavelength ranged from 400 to 4000 cm⁻¹ with a spectral resolution of 4cm^{-1} . The powdered samples of nanoparticles and algal pellets were oven dried overnight to eliminate the moisture content and were ground with KBr before the analysis.

2.4 Adsorption studies

The experiment was conducted in batch mode to study the adsorption of Pb (II) and Cd (II) on the surface of synthesized Ag nanoparticles (Ifijen et al., 2020). The influence of contact time, pH, temperature, adsorbate, and adsorbent dose on the adsorption capacity of the nanoparticles was evaluated in the study. The suspensions containing a concentration of individual metal ions were agitated at 150 rpm in a shaker for predetermined period. The samples were centrifuged at 8000 rpm for 30 min and resulted supernatant was filtered through Whatman 42 filter paper for the estimation of heavy metal concentration. All the experiments were conducted in triplicate and average values were taken into consideration. The % removal of the heavy metal ion from the solution was calculated using the simple expression as given below (Al-Qahtani, 2017):

$$\% \text{ Removal } = \frac{C_o - C_f}{C_o} \times 100 \tag{1}$$

where C_o (mg/l) and C_f (mg/l) are the initial and final concentrations of the metal ion in the solution respectively.

Salts of $CdCl_2$ and $PbCl_2$ were dissolved in distilled water to prepare 1000 mg L⁻¹ stock solutions. The different concentrations required to study the adsorption process were prepared using stock solutions (Ge et al., 2012).

2.5 Batch mode adsorption experiment

The experiments were conducted in batch mode to determine the effects of various factors on the heavy metal removal efficiency of nanoadsorbents (Chandra & Khan, 2020). For the experiment, 0.5 g/L of nanoadsorbent was mixed with heavy metal solution (100 ml of 100 mg/l) and was shaken at 150 rpm in an orbital shaker for different time of contact: 5, 10, 30, 50, 70, 90, 120, 140 and 180 minutes. On no account adjustment in the pH and temperature of the solution was done for the adsorption study. The concentration of hydrogen ions in the solution affects the adsorption process. Therefore to study the effect of pH on the removal efficiency of an adsorbent varying range of pH was selected: 2, 4, 6, 8, and 10. The

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desired pH of the reaction mixture (adsorbate: 100mg/L; adsorbent: 0.5g/L) was maintained by 0.1N NaOH and 0.1N HCl. The previously determined time of contact was employed in the experiment at room temperature. The concentration of adsorbent plays a significant role in the effective and rapid removal of the pollutants hence optimization of the adsorbent dose is highly preferable. In the experiment, nanoadsorbent of different concentrations i.e. 0.05, 0.10, 0.20, 0.50, 0.90 g/L was mixed with heavy metal solution (100 ml of 100 mg/l) and was shaken at 150 rpm in an orbital shaker for the period as identified in the previous experiment. Optimized contact time and pH were employed in the experiment at room temperature.

To study the effect of initial concentration of heavy metal ions on the adsorption capacity of the nanoparticles varied concentration of the metal solution was chosen for the study i.e. 50, 100, 150 and 200, 250 and 300 mg/L. The previously determined contact time and pH were employed in the experiment at room temperature. To determine the effect of temperature on adsorption, the batch experiment was conducted at different temperature (5, 20, 35, 50, 65, and 80°C), initial metal ion concentration (100 mg/ L), adsorbent dose (0.5 g/L) contact time (120 min) and pH (as determined in the above experiments).

After performing each above mentioned experiment the samples were centrifuged (at 8000 rpm for 30 minutes) and the supernatant was used to determine the heavy metal concentration by AAS analytical technique (Sens-AA, GBC- Scientific equipment). The sorption efficiency of nanoparticles can be calculated using the following equation (Ali et al., 2018):

$$q_e = \frac{C_o - C_e}{M} \times V \tag{2}$$

where, qe is the concentration of adsorbed metal ion onto the nanoparticles at equilibrium in mg/g, C_o and C_e are initial and equilibrium concentration of metal ions in mg/l, V is the volume of the reaction mixture in Liters and M is the concentration of nanoparticles used in the reaction mixture in grams.

2.6 Adsorption isotherm

The results obtained from the metal ion removal studies using varied concentrations of heavy metals (adsorbate) were fitted to different models to identify the adsorption capacity of nanoparticles. Langmuir isotherm (Langmuir, 1918) and Freundlich isotherm (Freundlich, 1907) models were fitted to the experimental data obtained from the metal removal study. The empirical equation of the Freundlich isotherm model can be expressed as Freundlich (1907):

$$\log q_e = \log K_F + \frac{1}{n} \log C_e \tag{3}$$

where, K_F and n are the two Freundlich constant that denotes the adsorption capacity of adsorbent (mg/g) and adsorption intensity (affinity with which adsorbate are attached to adsorbent) respectively; where, 1 < n < 10 shows that adsorption if favourable and n < 1 shows that the adsorption is unfavourable (Venkateswarlu et al., 2015). The numerical values of both the constants can be computed by plotting ln q_e against ln C_e .

The Langmuir equation (Langmuir, 1918) is based on the assumption that a finite number of binding sites is distributed homogeneously over the surface of the adsorbent and also that the adsorbate gets attached to the outer surface of the adsorbent forming a single molecular layer. Once the monolayer is formed no further sorption takes place with zero interaction between the adsorbed materials. It can be further described using the equation as given below:

$$\frac{C_e}{q_e} = \frac{C_e}{q_0} + \frac{1}{q_0 K_L} \tag{4}$$

Where, q is the concentration of adsorbed metal ion onto the nanoparticles at equilibrium in mg/g; Ce is the concentration of adsorbed metal ion in the reaction mixture at equilibrium in mg/L; q0 is the maximum adsorption capacity of adsorbent (mg/g); KL is the Langmuir constant (L/mg) that reflects the affinity of active binding sites quantitatively.

The parameters q0 and KL are determined by plotting $C_{e'}$ q_e against C_{e} . The equilibrium parameter (RL) is dimensionless constant and classified as one of the key features of the Langmuir isotherm model and can be denoted as:

$$R_L = \frac{1}{1 + (1 + K_L C_i)} \tag{5}$$

where, C_i is the initial concentration of adsorbate and K_L is the Langmuir constant related to the energy of adsorption.

The resultant value of the equilibrium parameter (RL) determines whether the nature of adsorption is favourable (when 0<RL<1), unfavourable (RL>1), linear (RL=1), or irreversible (RL=0).

2.7 Rate kinetics

In the present study, the data obtained from the batch experiment was modelled using simplified kinetic models such as pseudo firstorder and pseudo second-order models to determine the rate of adsorption of heavy metal ions onto the surface of biosynthesized nanoparticles. The batch kinetic experiment was carried out to determine the time at which equilibrium condition prevails in the reaction mixture also termed as equilibrium time. For the adsorption kinetic study, 0.5 g/L of nanoparticles was dissolved in 100 mg/L of metal solution at native pH and kept on the orbital shaker (150 rpm) for 180 min. The pseudo first-order order

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equation which is used to describe the sorption rate is given as (Lagergren, 1898):

$$\log (q_e - q_t) = \log (q_e) - \frac{K_1 t}{2.303}$$
(6)

where, qe and qt are the amount of metal adsorbed per unit weight of nanoparticles (mg/g) at equilibrium and at the time "t" respectively and K_1 is the pseudo first-order rate constant in l/min which can be determined by plotting ln (qe- q_t) against time t (min).

The adsorption rate can be determined by using pseudo-secondorder equation as given below (Ho & McKay, 2000):

$$\frac{t}{q_t} = \frac{1}{K_2 q_e^2} + \frac{t}{q_e} \tag{7}$$

where, qe and qt are the amount of metal adsorbed per unit weight of nanoparticles (mg/g) at equilibrium and at time t respectively and K_2 is the pseudo second-order rate constant which can be determined by plotting t/qt against time t.

3 Results and Discussion

3.1 Identification of isolated freshwater microalgal strain

The isolate was selected on the basis of higher optical density (2.46), biomass concentration (168.4 \pm 0.22 mg L-1day-1), and specific growth rate (1.63 \pm 0.31µg/day). The obtained microscopic view of the selected isolate of microalgal strain (Figure 1) was found similar to type *Scenedesmus* when compared with algaeBase database (Guiry & Guiry, 2015). The *Scenedesmus* cells comprise of a single nucleus and one pyrenoid containing chloroplast (Chaidir et al., 2016). The identification of selected microalgal strain was based on an Internal transcribed spacer (ITS) sequence. The ~600bp, ITS fragment was amplified using high–fidelity PCR polymerase. The PCR product was sequenced bi-directionally and the sequence data was aligned and analysed to identify the sample

and its closest neighbours. The molecular identification was done to determine the similarity of the species. The microscopic morphological analysis of isolated microalgal strain confirmed that the isolated microalgal strain belongs to the eukaryotic group. Figure 2 shows the developed phylogenetic tree representing the evolutionary relationship among different organisms. The microalgal sample was found to be *Tetradesmus dimorphus* isolate 959_218S ribosomal RNA gene, partial sequence; internal transcribed spacer 1, 5.8S ribosomal RNA gene, and internal transcribed spacer 2, complete sequence; and 28S ribosomal RNA gene, partial sequence. It was observed that the isolated microalgal strain is clustered with four different strains of *Tetradesmus dimorphus* with sequence similarities of 99 %.

3.2 Synthesis and characterization of AgNPs

The colour of the solution changed from bright yellow to dark brown within an hour of the reaction indicating the presence of Ag nanoparticles (Figure 3). The appearance of silver nanoparticle adsorption peak was periodically monitored at 200-700 nm wavelength range for 1, 5, 15, 20, and 30 days to assess the stability of phycosynthesized AgNPs. Figure 4 shows the maximum peak of UV-Vis spectra for silver nanoparticles at 411 nm with increased absorbance due to the increase in the number of nanoparticles with time, indicating the smaller size. The SPR band position is highly influenced by the shape and size of the nanoparticles. Several studies have reported the presence of silver nanoparticles at an absorbance peak ranging from 400-450 nm (Yousaf et al., 2020; Jabir et al., 2021; Aravind et al., 2021). Therefore, the aforementioned observation showed the successful bioreduction of Ag nanoparticles. The appearance of UV-Vis adsorption spectra of nanoparticles is controlled by SPR and represents the shifts in the adsorption peak from red to blue end due to the slight deviations in the shape and size of the nanoparticles (Raghunandan et al., 2011).



Figure 1 Identified shape morphology of isolated microalgal strain (a) and Scenedesmus (b) (source: algaebase.org)

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Figure 2 Phylogenetic tree showing evolutionary relatedness of isolated microalgal strain with their closely related microalgae species



Figure3 Synthesis of nanoparticles AgNO₃ and algal extract solution after 5 min and 1 hour.

The obtained XRD pattern for AgNPs is shown in Figure 5. The obtained XRD pattern of Ag nanoparticles showed four distinctive diffraction peaks at 2 theta values of 38.12° , 44.06° , 64.31° and 77.58° which is assigned to the planes (111), (200), (220), and (311) of face centered cubic (fcc) structure of Ag nanoparticles respectively. The XRD pattern of the current study was matched with the standard database of the Joint Committee on Powder Diffraction Standards (JCPDS card no. 04-0783 and 84- 0713). The observed peak pattern indicated the crystalline nature of biosynthesized and formation of zerovalent Ag (Ag⁰) nanoparticles.

The SEM images of phycosynthesized AgNPs are presented in Figure 6. It can be seen in the images that the synthesized AgNPs are spherical. The image obtained from SEM imaging displayed agglomeration of larger particles of powdered Ag nanoparticles. To get a better understanding of the morphology of biologically synthesized nanoparticles and to avoid the imaging of unwanted residue, the colloidal solution of the nanoparticle was washed 3-4 times with deionized water.

The biosynthesized AgNPs using aqueous extract of microalgal species were further analyzed by using TEM which is considered the most powerful and advanced analytical technique in determining the surface morphology of nano-sized particles (Figure 7). The average particle size of Ag nanoparticles were calculated by imageJ software using TEM images from which Gaussian distribution of nanoparticle size was obtained (Figure 8). The average particle sizes of AgNPs were found to be 15 ± 4.05 nm and polymorphic in shape.

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Figure 4 UV-Vis spectra of AgNO3 and algal extract solution



WFile: A raw - Type: 2Th/Th locked - Start: 20.000 ° - End: 79.997 ° - Step: 0.019 ° - Step time: 192.1 s - Temp:: 25 °C (Room) - Time Started: 4 s - Theta: 10.000 ° - Chi: 0.00 ° - Phi: 0.00 ° - X: 0.0 mm - Y: 0.0 mm



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Figure 6 SEM image of AgNPs



100 nm HV=200.0kV Direct Mag: 10000x AIRP-JNU

Figure 7 TEM image of AgNPs



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The FT-IR analysis was conducted to determine the presence of functional groups responsible for the bioreduction and stabilization of synthesized nanoparticles. The IR spectrum of dried algal biomass indicated the presence of peaks at 1019.39, 1150.94, 1240.50, 1410.50, 1457.61, 1545.68, 1641.45, 1743.06, 2362.31, 2854.02, 2923.13, and 3282.61 cm⁻¹ (Figure 9). The band positioned at 1743.06 and 3282.61 corresponds to stretching of (C=O) group and (O-H) group of polysaccharides respectively. The spectrum band positioned at 2923.13, 2854.02, and 2584.02

cm⁻¹ is due to the stretching of alkanes (C-H) of Thiol group. The vibrational peak at 1545.68 cm⁻¹, 1019.39 cm⁻¹, and 1641.45 cm⁻¹ are due to the stretching of aliphatic amines (C-N) and vibration of alkene (C=C) (Yian Zheng, 2012; Aboelfetoh et al., 2017). The peak at 1743.06 cm⁻¹ indicating the vibrational (-C=O) stretching of fatty acid ester and lipids. The vibrational peak at 1150.94 cm⁻¹ and 1240.50 cm⁻¹ corresponds to (C-O-C) stretching of carbohydrates and polysaccharides and (P=O) stretching of DNA, RNA, and phospholipids respectively (Jenaa et al., 2014). The

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band is positioned at 1410.50 cm⁻¹ due to the stretching of (C-C) group of aromatic rings. The analogous band spectra were observed in Ag nanoparticles with a slight shift in the band position and at lower intensities which confirms the involvement of functional groups in the bioreduction and stabilization of nanoparticles. The IR spectrum of biosynthesized Ag nanoparticles indicating the presence of peaks at 848.98, 1022.67, 1237.06, 1394.66, 1453.71, 1514.86, 1641.23, 1741.29, 2362.09, 2920.69, 3273.68, 3677.68, 3743.71, 3860.65 and 3894.18 cm⁻¹ (Figure 10). In the case of Ag nanoparticles peak shift of alkanes was observed from 2923.13 to 2920.69. The IR spectra at 1641.23, 1514.86, and 1022.67 cm⁻¹ exhibited the introduction of primary amines, nitro compounds, and aliphatic amines respectively indicating the bioreduction of AgNPs (Khalid et al., 2017). A slight shift of alkenes and O-H stretching was also observed from 1641.45 to 1641.23 cm⁻¹ and 2362.31 to 2362.09 cm⁻¹. The introduction of vibration peak at 1394.66 cm⁻¹ was found related to the amide (II and III) group of aromatic rings possibly due to the association of polyphenols with the synthesized NPs which were extracted from the algal biomass (Rajeshkumar et al., 2012). The slight shift in the peak from 1743.06 to 1741.29 cm⁻¹ and 1240.50 to 1237.06 cm⁻¹ ¹confirms the involvement of lipid biomolecules in the synthesis of AgNPs. The shift in the vibrational peak from 1457.61 to 1453.71 indicates the association of carboxylate ions with the reduced Ag metal ions (Govindaraju et al., 2015). The shift of peak from 3282.61 to 3273.68 cm⁻¹ confirmed the involvement of polysaccharide biomolecules in the reduction of bulk silver to zerovalent AgNPs. The observed peak at 848.98 indicated the presence of the alkyl halides group. The appearance of vibration peak in the range 3400-3000 cm⁻¹ showed stretching of the hydroxyl group.

3.3 Batch experiment adsorption study of Pb (II) using phycosynthesized AgNPs

It was observed that Pb (II) and Cd (II) metal ions were removed from the solution by nanoparticles within 120 min (Figure 11). The maximum experimental Pb (II) and Cd (II) ions removal from the solution using AgNPs were found to be 93.61±0.3 and 76.66±0.41mg/L respectively. The sorption trend for both the metal ions showed adsorption at 120 min which could be due to the unavailability of binding sites on the surface of nanoparticles; also no further increase in the adsorption of metal ions were recorded.

To determine the effect of hydrogen ion concentration on the adsorption process, the batch experiment was carried out at different pH (2.0, 4.0, 6.0, 8.0, and 10.0) with 100 mg/ L of initial metal ion concentration, 0.5 g of adsorbent dose, contact time of 120 min, agitation speed of 150 rpm at room temperature. The percent removal of metal ions by nano adsorbent with varying range of pH is shown in Figure 12. It has been observed that with the increase in pH from 2 to10 the percent removal of Pb (II) and

Cd (II) using Ag nanoparticles was increased. The minimum removal of Pb (II) and Cd (II) ion using Ag nanoparticles as an adsorbent was found to be 10.57 ± 0.37 and 5.68 ± 0.21 mg/L respectively at pH 2.0 while at pH 8.0 the maximum removal of Pb (II) and Cd (II) was found to be 82.82 ± 0.22 and 78.76 ± 0.88 mg/L respectively.

To determine the effect of adsorbent dose on the adsorption of selected metal ions, the batch experiment was conducted at different concentration of Ag nanoparticles (0.05, 0.1, 0.2, 0.5, 0.9 g/L), 100 mg/L of initial metal ion concentration, contact time of 120 min, pH as determined in the above experiment. The maximum removal of metal ions was observed at a concentration of 0.5 g/L. It can be seen in Figure 13 that with an increase in the concentration of adsorbent the removal efficiency increased which could be possibly due to the increase in the number of binding sites which in return facilitates the adsorption process. However, after a certain amount of adsorbent dose, the removal efficiency remains constant and no further removal of metal ion was observed. AgNPs were found responsible for the maximum removal of Pb (II) and Cd (II), i.e. 97.55±0.14 and 93.83±0.29 mg/L respectively and minimum removal of concentration was 8.68±0.32 and 8.72± 0.21 mg/L at 0.05 g/L of adsorbent dose. It has been suggested that at a higher adsorbent dose heavy metal adsorption reduces as metal ions are not sufficient to cover all the adsorption sites available on the adsorbent surface (Sivaprakash et al., 2009).

To determine the influence of concentration of initial metal ion on their adsorption on the surface of adsorbent an experiment was conducted in batch mode using different metal ion concentration (50, 100, 150, 200, 250, and 300 mg/L), adsorbent dose (0.5 g/L), agitation speed (150 rpm), pH as determined in the above experiment. In the case of Ag nanoparticles as an adsorbent, the maximum removal of Pb (II) and Cd (II) ions was 45.67±0.25 and 40.44±0.97 mg/L respectively at 50 mg/L of initial metal concentration (Figure 14). It is also evident from the observations that with the increase in the metal ion concentration the percent removal decreases which could be possibly due to the lesser number of available active binding sites on the adsorbent surface for the higher concentration of adsorbate. It was supported by the fact that at a lower concentration of adsorbate more vacant adsorption sites are available for sorption, hence more uptake of metal ions takes place (Al-Qahtani, 2017).

To determine the effect of temperature on adsorption, the batch experiment was conducted at different temperatures (5, 20, 35, 50, 65, and 80°C), initial metal ion concentration (100 mg/ L), adsorbent dose (0.5 g/L), contact time (120 min) and pH (as determined in the above experiments). The maximum removal of Pb (II) and Cd (II) metal ion by Ag nanoparticles was 96.07 ± 0.31 mg/L and 80.89 ± 0.08 mg/L at 5°C and minimum removal was 2.85 ± 0.45 and 3.63 ± 0.21 mg/L at 80°C respectively (Figure 15).

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Figure 11 Effect of time of contact on adsorption of Pb (II) and Cd (II) by Ag nanoparticles







Figure12 Effect of pH on adsorption of Pb (II) and Cd (II) by Ag nanoparticles



Figure14 Effect of adsorbate dose on adsorption of Pb (II) and Cd (II) by Ag nanoparticles



Figure15 Effect of temperature on adsorption of Pb (II) and Cd (II) by Ag nanoparticles

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The equilibrium study on Pb (II) and Cd (II) metal ion removal was carried out using varying amount of initial metal ion concentration from 50 to 300 mg/ L, adsorbent dose (0.5 g/L), room temperature, the volume of the metal ion solution (100 ml), equilibrium time (120 min), agitation speed (150 rpm) and at their native pH. The parameters/ constants of Freundlich and Langmuir isotherm for Pb (II) and Cd (II) Ag nanoparticles as adsorbents are listed in Table 1. Linear plots of Langmuir isotherm for Pb (II) and Cd (II) using Ag nanoparticles as adsorbents are shown in Figure 16 (a & b). Linear plots of Freundlich isotherm for Pb (II) and Cd (II) using Ag nanoparticles adsorbents are shown in Figure 17 (a & b).

affinity of adsorption sites and maximum adsorbent capacity respectively. The maximum value of metal ion adsorption capacity (23.98 mg/g) was observed for Pb (II) (Table 1). AgNPs showed a stronger binding affinity towards Cd (II). The value of the equilibrium parameter (RL) showed that the nature of adsorption of the metal ion by Ag nanoparticles is favourable as 0 < R < 1. The two parameters of Freundlich isotherm; n and KF, measure the affinity of adsorption sites and maximum adsorbent capacity respectively. The higher value of R^2 showed that the experimental data were fitted best with Langmuir isotherm than with Freundlich isotherm using Ag nanoparticles. Several studies have explained the equilibrium data for Pb (II) and Cd (II) with the help of Langmuir and Freundlich adsorption isotherm using AgNPs (Poguberović et al., 2016; Al-Qahtani, 2017; El-Sabban et al., 2020).

The two parameters of Langmuir isotherm; KL and q0 measure the et

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Metal ion	Langmuir isotherm			Freundlich isotherm		
	$K_L(L/mg)$	$q_0 (mg/g)$	\mathbf{R}^2	n	$K_{\mathrm{F}}\left(\mathrm{mg/g} ight)$	\mathbf{R}^2
Pb (II)	0.0304	23.98	0.990	2.7005	1.622	0.961
Cd(II)	0.041	22.47	0.994	3.0211	1.831	0.959

Table 1Parameters of adsorption isotherms for P	^b (II) and Cd (II) adsorption on Ag nanoparticle
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Figure16 (a) Langmuir adsorption isotherm plot of Pb (II) ion using Ag nanoparticles

Figure16 (b) Langmuir adsorption isotherm plot of Cd (II) ion using Ag nanoparticles

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1.6 1.4 1.2 1 Log qe y = 0.331x + 0.605 $R^2 = 0.959$ 0.8 0.6 0.4 0.2 0 0 0.5 1.5 2 2.5 1 Log Ce

Figure17 (a) Freundlich adsorption isotherm plot of Pb (II) ion using AgNPs

Figure 17 (b) Freundlich adsorption isotherm plot of Cd (II) ion using AgNPs

3.5 Adsorption kinetics

To determine the adsorption rate of metal ions on nanoparticles pseudo-first order and pseudo- second order kinetic models were used in the present study. The equilibrium study on Pb (II) and Cd (II) metal ion removal was carried out at the different time of contact from 5 to 120 min, adsorbent dose (0.5 g/L), the volume of the metal ion solution (100 ml), adsorbate dose (100 mg/L), agitation speed (150 rpm) and at their native pH. The best fitted model was selected based on a higher value of linear correlation coefficient (\mathbf{R}^2) and the similar value obtained from theoretical and experimental data. The experimental data obtained from adsorption process was fitted in the above-mentioned kinetic models to calculate the metal uptake by adsorbents at a different time of contact.

The calculated parameters of pseudo-first order and pseudo-second order rate kinetic models for Pb (II) and Cd (II) using Ag nanoparticles as adsorbent are shown in Table 2 by plotting log (qe-qt) against time t the parameters of pseudo 1st order were determined from slope and intercept of the plot (Figure 18 & 19) while by plotting t/qt against time t, the parameters of pseudo 2nd

order were determined from slope and intercept of the plot (Figure 20 & 21). Both the rate kinetic models were used to simulate the experimentally obtained data for Pb (II) and Cd (II) Ag nanoparticles as adsorbents.

Using Ag nanoparticles the uptake of Pb (II) metal ion showed a higher value of R² i.e.0.93 and maximum adsorption capacity of 23.14 mg/g respectively, hence it can be concluded that Pb (II) followed pseudo 2nd order kinetic model and chemisorption was responsible for the interaction of metal ion with the adsorbent (Alia et al., 2018). On the contrary, it was observed that the linear correlation coefficient of Cd (II) is very low in the case of pseudo 2nd order model and also the experimental qe value and calculated qe value do not match appropriately but fitted well with pseudo 1st order kinetic model hence it can be concluded that physical adsorption is responsible for Cd (II) metal ion uptake by Ag nanoparticles. Few studies have reported the application of kinetic models with pseudo- first order and pseudo- second order equations to examine successful adsorption of Pb (II) and Cd (II) ion using nanoparticles as an adsorbent (Abdollahi et al., 2018; Salam et al., 2020; Lin et al., 2020).

Metal ion	Pseudo 1 st order			Pseudo 2 nd order		
	K	q _e (mg/g)	R^2	K ₂	q _e (mg/g)	R^2
Pb (II)	0.0479	5.697	0.874	0.00186	23.14	0.930
Cd (II)	0.0474	5.476	0.883	-0.0053	-13.62	0.1442

Table 2 Values of parameters of pseudo 1st and 2nd order kinetic models using Ag nanoparticles as an adsorbent

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Figure 18 Pseudo-first order kinetic model for Pb (II) using Ag nanoparticles

Figure 19 Pseudo-first order kinetic model for Cd (II) using Ag nanoparticles





Figure 20 Pseudo-second order kinetic model for Pb (II) using Ag nanoparticles

Figure 21 Pseudo-second order kinetic model for Cd (II) using Ag nanoparticles

Conclusion

An eco-friendly and cost-effective approach has been employed for the synthesis of AgNPs and was examined for their efficiency in the removal of metal ions from an aqueous solution. In the present study, Ag nanoparticles were synthesized using an aqueous extract of microalgae characterized as *Tetradesmus* sp. based on morphological and DNA gene sequencing at the ITS region. The initial change in colour of the reaction mixture confirmed the synthesis of nanoparticles. UV-Vis spectroscopy showed absorption peak at 411 nm indicating a reduction in the size of bulk AgNPs. XRD, SEM, and TEM analysis confirmed the spherical shape and crystalline nature of biosynthesized Ag nanoparticles with an average particle size 15 nm. The FT-IR spectra clearly showed the vibrational stretching of the functional group present on protein, carbohydrate, and lipids confirming the involvement of these biomolecules in size reduction and stabilization of nanoparticles. The equilibrium time of 120 min was recorded for the maximum removal of Pb (II) and Cd (II) ions using Ag nanoparticles. The maximum removal of Pb (II) (II) was observed at pH 8.0 due to higher electronegativity as compared to Cd (II) ion. The study on the effect of initial metal concentration on adsorption of metal ions showed maximum removal at lower metal concentration, i.e. 50 mg/L. The maximum removal of both the metal ions was observed at adsorbent dose of 0.5 g/L. The removal of Pb (II) and Cd (II) by Ag nanoparticles was observed sharply when nanoparticle concentration ranged between 0.05-0.5 g/L. The maximum removal of metal ions at lower temperatures signifies an exothermic adsorption process. The experimental data obtained from the batch experiment were best fitted in the Langmuir isotherm adsorption isotherm model (R²>0.99). The Ag nanoparticles showed maximum adsorption capacity for Pb (II) (23.98 mg/g) and higher affinity towards Cd (II) indicating the selective removal of the metal ion by nanoadsorbent. It was

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observed that the adsorption of Pb (II) and Cd (II) ions on AgNPs followed pseudo-second-order kinetics (R2>0.9) and pseudo-firstorder kinetics (R2>0.8) respectively. Therefore, it can be concluded that Ag nanoparticles can be used efficiently in water/wastewater treatment as an adsorbent. Further research aims at evaluating the usability and stability of catalyst AgNps. Future research needs to be focused on determining the fate of waste generated and evaluating the cost-effectiveness of biologically synthesized nanoparticles for economic designing of water treatment plants.

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Conflict of interest

The authors declare that there is no conflict of interest.

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