University of Texas Rio Grande Valley

ScholarWorks @ UTRGV

Earth, Environmental, and Marine Sciences Faculty Publications and Presentations

College of Sciences

9-2022

Fine Particulate Matter Concentrations during Independence Day Fireworks Display in the Lower Rio Grande Valley Region, South Texas, USA

Esmeralda Mendez

Owen Temby

Dawid Wladyka

Katarzyna Sepielak

Amit U. Raysoni

Follow this and additional works at: https://scholarworks.utrgv.edu/eems_fac

Part of the Earth Sciences Commons, Environmental Sciences Commons, and the Marine Biology Commons



Research Article

Fine Particulate Matter Concentrations during Independence Day Fireworks Display in the Lower Rio Grande Valley Region, South Texas, USA

Esmeralda Mendez,¹ Owen Temby,¹ Dawid Wladyka,² Katarzyna Sepielak,² and Amit U. Raysoni ¹

¹School of Earth, Environmental, and Marine Sciences, The University of Texas Rio Grande Valley, Brownsville, TX, USA ²Department of Sociology, The University of Texas Rio Grande Valley, Brownsville, TX, USA

Correspondence should be addressed to Amit U. Raysoni; amit.raysoni@utrgv.edu

Received 9 May 2022; Revised 16 August 2022; Accepted 20 August 2022; Published 12 September 2022

Academic Editor: Muhammad Altaf Khan

Copyright © 2022 Esmeralda Mendez et al. This is an open access article distributed under the Creative Commons Attribution License, which permits unrestricted use, distribution, and reproduction in any medium, provided the original work is properly cited.

Fireworks are typically discharged as a mark of celebration and joy in many societies spanning various cultures. In the United States of America, 4th July is celebrated as the Independence Day when the nation overthrew the British colonial yoke in 1776. While this day instills a sense of patriotism in every American's heart, it is also a major $PM_{2.5}$ air pollution concern. This study is first of its type in the Lower Rio Grande Valley (RGV) Region of South Texas, USA, that characterizes fine particulate matter pollution. Using a low-cost sensor (TSI BlueSky Air Quality Monitor), real-time $PM_{2.5}$ measurements were assessed at eleven different locations in four different towns and cities of Lower RGV Region: Brownsville, Edinburg, Weslaco, and Port Isabel. Hourly $PM_{2.5}$ concentrations from July 03–06, 2021 are presented in this research work. Intraurban $PM_{2.5}$ spatial and temporal variations provide an insight on the general population's exposure burden during the festive period. Results indicate an increase in fine particulate matter pollution across the region, but the levels do not exceed the U.S. National Ambient Air Quality Standards (NAAQS). Findings from this study would possibly help in the formulation of effective firework policies to minimize the pollution impact.

1. Introduction

Fireworks display can be a major source of air pollution and has the potential to cause deleterious health effects [1–4]. Of all the major criteria air pollutants, fine particulate matter (aerosol particles with an aerodynamic diameter of less than $2.5 \,\mu$ m), i.e., PM_{2.5} is measured and studied due to its documented health effects [5–7]. PM_{2.5} particles can deposit deep in the lungs thereby resulting in damage to the lower thoracic region [8]. PM_{2.5} exposures have both acute and chronic impacts on human health [2, 3, 9–11]. Continued exposure to fine PM results in the development of relevant cardiovascular and or respiratory diseases [1, 4, 10]. The major sources of fine particulate matter pollution are typically traffic emissions including both tailpipe and break and tire tear and wear, power plant emissions, emissions from smokestacks, and agricultural burning [4, 12, 13].

Firework displays generate harmful concentrations of particulate matter and sulfur dioxide [4, 13–18]. Statistically significant impact between trace metals in $PM_{2.5}$ from firework emissions on human health via reactive oxygen species formations has been well-documented [14, 19]. The assembly of a firecracker consists of multiple inorganic and organic chemicals such as sodium oxalate, strontium nitrate, barium nitrate, potassium perchlorate, potassium nitrates, arsenic, aluminum, charcoal, sulfur, manganese, and iron dust powder [11, 13, 14, 16, 17]. Explosion of fireworks also result in local haze [4].

A study published in 2015 reported that $PM_{2.5}$ concentrations are elevated on 4^{th} July and continue to remain

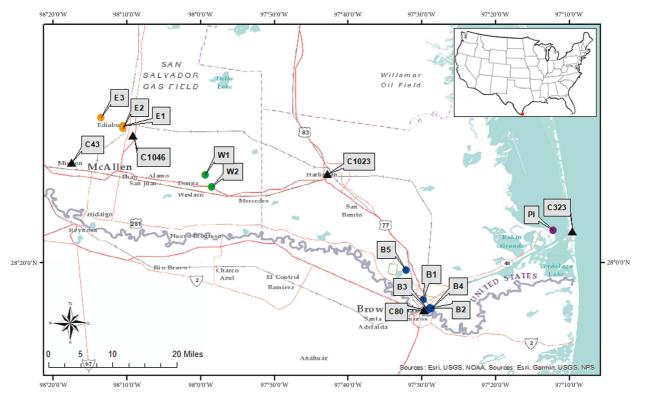


FIGURE 1: Map of Texas commission on environmental quality (TCEQ). Continuous ambient monitoring station (CAMS) sites and BlueSky air quality monitors throughout the study area in the lower RGV region.

high until the morning of 5th July [20]. The authors analyzed PM_{2.5} data from 315 US air quality monitory sites and showed a 42% increase $(5 \mu g/m^3)$ of 24 hr PM_{2.5} concentrations in the United States during Independence Day [4, 20]. During the Chinese Lantern Festival in Beijing, Wang and colleagues showed a six-fold increase in PM_{2.5} concentrations on the Lantern Day as compared to normal days [17]. Another study from Beijing, China quantified PM_{2.5} average concentrations and documented highest levels during firework days at $248.9 \,\mu g \cdot m^{-3}$ in the 2015 Spring Festival. [4, 21, 22]. During the Montreal International Fireworks Competition in Quebec, Canada, researchers showed that $PM_{2.5}$ levels can be as high as 1000 μ g/ m³ during the display period of about 45 minutes [23]. In New Delhi, India, during the day of the Diwali (Festival of Lights) festival, $PM_{2.5}$ levels were 588 μ g/m³ in 2007, and 389 µg/m³ in 2008 [13].

Till date, no study has been conducted till date in the Lower RGV Region of South Texas, USA, to assess an increase in fine particulate matter pollution levels during the US Independence Day celebrations. The present study is first of its kind to use TSI BlueSky low-cost sensor to characterize fine particulate matter pollution in this region. Recently, the usage of low-cost air quality sensors to measure PM pollution has been gaining credence [9, 24]. These sensors are easy to operate, very convenient in terms of price, energy, and mobility, and record data in real time and can store data for future downloading purposes [9, 25].

2. Materials and Methods

2.1. Site Selection and Study Period. The study area is in the Lower RGV Region of South Texas, the USA, with a population of 1,402,512 persons [26]. Eleven BlueSky sensors were deployed in four towns and cities and labeled as follows: five monitors in Brownsville (B1–B5), three monitors in Edinburg (E1–E3), two monitors in Weslaco (W1 and W2), and one monitor in Port Isabel (PI). These intraurban sensor locations are illustrated in Figure 1 with color-coordinated indicators to differentiate the cities. Brownsville locations are marked with blue indicators, Edinburg has orange, Weslaco has green, and lastly, Port Isabel is purple. Each marker signifies a deployed BlueSky monitor with tactical outdoor locations to portray significant areas such as schools, residential area, and other neighborhood locations.

Sensor B1 in Brownsville is right across from Dean Porter Park and adjacent to the Gladys Porter Zoo in a semiresidential area. B2 is deployed in the University of Texas Rio Grande Valley (UTRGV) Brownsville Campus police department, near the campus student dormitories. B3 is located at the Music and Science Learning Center on the University Boulevard at the UTRGV campus, and B4 is situated adjacent to Texas State Highway 69E only 0.04 miles away from the U.S.-Mexico International Port of Entry. The entire university is located on the main highway 69E on the edge of the US-Mexican border, as seen in Figure 1, so exposure to traffic pollutants from cross-border vehicular traffic is a health concern. B5 is in a neighborhood surrounded by Resaca del Rancho Viejo. Resacas are ancient distributary channels of Rio Grande River (the natural international boundary between the United States and Mexico in the State of Texas) and are a unique geological feature of this area.

Figure 1 also shows the locations of the five Continuous Ambient Monitoring Station (CAMS) sites in the region. These sites are maintained by the Texas Commission on Environmental Quality (TCEQ) and they monitor air pollutant such as $PM_{2.5}$ and meteorological parameters such as temperature, resultant wind speed, and solar radiation. The CAMS sites are shown as black triangular symbol in the study map. The five CAMS sites are: C323 (Port Isabel), C80 (Brownsville), C1023 (Harlingen), C1046 (Edinburg), and C43 (Mission). Out of the five CAMS sites, only three record data for $PM_{2.5}$. These are CAMS sites C43, C80, and C323. The other two C1046 and C 1023 do not collect $PM_{2.5}$ data. All five sites collect various meteorological parameters such as temperature, resultant wind speed, and two sites (C43 and C80) provide data on solar radiation.

The UTRGV Edinburg campus has two monitors, E1 and E2 facing commonly used roads (Schunior St. and 107 Texas, respectively). E3 is deployed in a closed gated residential community. In Weslaco, W1 is located further from the city in a residence right off Farm to Market Road 88 (FM88). W2 is deployed at one of the buildings of the Weslaco Police department on the frontage of Texas State Expressway 83. Lastly, the PI sensor is deployed at the UTRGV coastal labs between two neighborhoods near the popular tourist destination of South Padre Island.

Sensor deployment locations were considered such that they are an accurate representation of the daily exposure levels of $PM_{2.5}$ in the neighborhood. The RGV Region has previously hosted multiple fireworks in their cities, as well as supporting many fireworks stands, so residents can easily purchase firecrackers for their celebrations. In 2021, the cities and towns of Brownsville, Edinburg, Harlingen, McAllen, Weslaco, and South Padre Island celebrated the Independence Day with much fanfare by grand firework display shows.

The study period starts the evening of July 3^{rd} , 7:00 pm through July 6^{th} at 6:59 pm. BlueSky monitors were configured to log PM concentrations every 5 minutes, and these data were collected by a built-in microSD card. Those 5-minute concentrations were formulated into hourly data to be further assessed. The day prior and after 4^{th} July are considered primarily as control days to help understand the temporal variation in the PM_{2.5} concentrations.

2.2. Instrumentation. In this study, BlueSky Low-Cost Sensors were used (Model: 8143 by TSI Incorporated, Minnesota, U.S.). This instrumentation is easy to install and weighs about 0.35 lb. The sensor does not require much power, approximately less than 5 W (5 VDC @ 1 Amp). The PM sensor included is precalibrated similarly to other highquality TSI equipment like the DustTrak[™] models (TSI Incorporated, Minnesota, U.S). Self-diagnostic tests are configured to daily cleaning intervals to attain high-quality

data. The PM sensor measures from the range 0 to $1000 \,\mu \text{g}/$ m^3 with the measurement resolution of $1 \mu g/m^3$ and a response time being 1 second [27]. It is prudent to mention here that, albeit the usage of low-cost sensors for air quality monitoring has increased substantially in the last few years, issues such as sensor baseline drift, sensitivity to variations in meteorological parameters such as ambient temperature and relative humidity, instrument measurement artifact needs to be addressed and accounted for as outlined succinctly by Morawska and colleagues [28]. Extensive quality assurance and quality control (QA and QC) studies conducted by the South Coast Air Quality Management District (South Coast AQMD) on BlueSky sensors have shown a good record of performance and evaluations with the sensor showing a moderate to strong $PM_{2.5}$ (0.66 < R^2 < 0.78) correlation with other Federal Equivalent Method (FEM) instruments like FEM GRIMM and FEM Teledyne API T640 in the field [29].

2.3. Statistical Data Analysis. Data from these low-cost sensors were inputted in Microsoft Excel (2021) for calculating hourly concentrations and time series. The time series depicts $PM_{2.5}$ concentrations through the study period hours, while also comparing data from each site. Spearman's Rho correlations were calculated with SPSS for MacOS (SPSS, Inc., Chicago, IL) to estimate temporal variability in $PM_{2.5}$ concentrations across the eleven locations. Visual data analysis from R programming (RStudio, Inc., Boston, MA) was used to demonstrate box plot hourly variability. The boxes are the interquartile ranges (75th and 25th), the whiskers show the minimum and maximum values, the outliers are shown in asterisk. The median is indicated by the black line inside the boxes and the diamond in the boxes indicate the mean.

Spatial variation in $PM_{2.5}$ concentrations in each site are analyzed with the performance of Coefficient of Divergence (COD) analysis was performed to study the spatial variation between the various study sites [30–32]. The COD provides a degree of uniformity between two simultaneously sampled sites, *j* and *k* by the following equation:

$$\text{COD}_{jk} = \sqrt{\frac{1}{p} \sum_{i=1}^{p} \left[\frac{x_{ij} - x_{i,k}}{x_{ij} + x_{i,k}} \right]^2},$$
 (1)

where the number of observations is indicated by p and x_{ij} is the *i*th concentration measured at site *j* over the sampling period. COD values less than 0.20 indicate the two observed sites are spatially homogeneous in terms of the pollutant concentration. COD equal to greater than 0.20 establishes spatial heterogeneity in the pollutant concentrations or significant differences between the two simultaneously sampled sites [31].

3. Results and Discussion

3.1. $PM_{2.5}$ Concentrations. Time series was plotted for the data available from the evening of June 03, 2021 (19:00 hours) till the evening of July 06, 2021, (19:00 hours) as

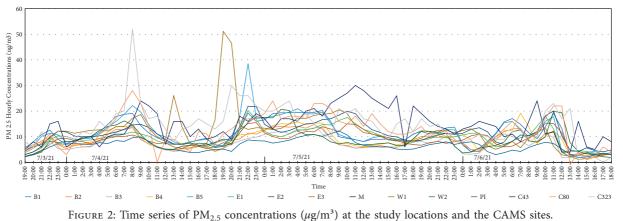


TABLE 1: Basic statistics of hourly PM_{2.5} concentrations during the study period at the various study locations and the CAMS sites.

Date	Sites	B1	B2	B3	B4	B5	E1	E2	E3	W1	W2	PI	C43	C80	C1046
	М	10.6	7.5	7.0	8.0	10.9	8.4	9.5	7.4	11.9	6.2	8.4	11.83	10.78	14.83
	STD	3.9	2.7	2.7	3.0	4.4	2.3	3.1	2.2	4.8	2.0	3.8	5.16	6.48	9.02
July 03, 19:00-July 04, 18:59 hours	MAX	19.1	13.6	14.0	14.6	22.2	11.8	14.8	10.7	26.1	9.9	16.2	24.00	28.00	52.00
	MIN	5.7	3.8	3.7	4.5	4.7	3.0	3.0	2.3	4.4	2.2	2.2	5.00	3.00	6.00
	Median	9.4	6.9	6.1	7.1	9.7	8.8	10.0	7.5	12.2	6.5	7.7	5.5	6	7.4
	М	13.6	9.6	8.8	10.4	14.8	12.5	14.1	11.1	17.2	7.9	10.8	19.54	16.54	18.17
	STD	4.9	3.5	2.9	3.6	7.1	3.2	4.1	2.7	10.6	2.3	4.7	5.96	3.82	5.38
July 04, 19:00-July 05, 18:59 hours	MAX	20.3	14.5	13.9	15.7	38.5	18.8	21.9	15.5	51.2	12.3	20.7	30.00	24.00	30.00
	MIN	6.7	4.9	5.1	5.5	6.3	7.3	8.3	6.8	8.3	4.3	6.3	6.00	11.00	9.00
	Median	11.8	8.9	8.3	9.7	15.2	12.1	13.9	11.3	14.2	7.8	8.6	6.65	7.7	6.9
	М	9.7	7.5	6.4	8.6	9.4	8.8	9.7	8.0	7.6	5.3	8.3	12.08	9.67	10.50
	STD	5.1	3.9	3.2	4.5	4.8	3.7	4.1	3.6	4.3	2.1	4.9	4.83	6.03	6.35
July 05, 19:00-July 06, 18:59 hours	MAX	19.1	15.3	10.8	19.3	18.2	17.7	20.1	15.3	13.8	8.0	19.9	24.00	22.00	23.00
·	MIN	2.6	1.9	1.9	2.3	2.0	3.4	4.1	2.8	1.2	1.8	1.9	4.00	0.00	0.00
	Median	11.5	8.5	7.1	9.9	10.5	9.2	10.3	7.8	7.5	5.5	7.8	4.4	6.9	7.9

TABLE 2: Meteorological parameters during the study period at the various CAMS sites.

		July 03, 19:00—July 04, 18:				July 04	, 19:00)—July	05, 18:	July 05, 19:00—July 06, 18:			
Met parameters	CAMS sites		59 I	hours		59 hours				59 hours			
		Mean	STD	Min	Max	Mean	STD	Min	Max	Mean	STD	Min	Max
	C43	30.06	3.59	25.44	35.72	29.28	2.62	26.11	33.67	26.98	1.89	24.44	32.17
	C80	29.17	2.20	26.72	32.78	29.02	2.30	25.72	32.44	26.95	2.28	22.67	30.06
Temp (°C)	C323	30.18	1.24	28.72	32.00	30.25	1.04	28.94	32.00	28.69	2.26	24.11	31.11
	C1023	29.51	2.67	26.33	33.89	29.37	2.71	26.06	33.44	26.58	2.36	22.39	31.11
	C1046	29.92	3.37	25.78	35.17	29.44	3.08	25.83	34.00	26.73	2.14	23.56	31.89
	C43	2.83	0.90	1.74	4.96	3.09	0.80	1.21	5.27	2.08	1.03	0.27	3.84
	C80	2.97	1.07	1.52	4.74	3.32	1.56	0.13	5.10	2.98	1.44	0.13	5.36
RWS (m/s)	C323	3.38	1.07	1.65	4.96	3.36	0.70	2.32	4.87	3.50	0.72	1.52	4.78
	C1023	3.02	1.09	1.52	5.23	3.66	1.17	1.70	5.54	2.73	0.81	1.52	4.34
	C1046	2.69	0.84	1.30	4.34	3.03	1.20	1.21	5.14	1.97	1.25	0.22	4.51
Solar	C43	0.43	0.52	0.00	1.35	0.26	0.31	0.00	0.87	0.04	0.06	0.00	0.19
Radiation (Langleys per minute)	C80	0.33	0.45	0.00	1.39	0.31	0.39	0.00	1.20	0.07	0.12	0.00	0.42

shown in Figure 2. Hourly basic statistics for $PM_{2.5}$ for the entire duration are shown in Table 1 and the meteorological parameters are shown in Table 2. The weather conditions during the study period were stable with the mean temperature in the region around 30°C. The resultant wind speed also ranged from 1.97 to 3.36 m/s across the five CAMS sites. The time series shows the average hourly data $PM_{2.5}$ in $\mu g/$

m³. The time series expressions are estimated as hourly averages to better interpret any easily identifying variations. The four-day duration is labeled accordingly on the *x* axis. An evident increase in spike starting the evening of July 4th at approximately 20:00 hours is obvious with high concentrations lingering till the morning hours of July 5. The sensor W1, a location in the town of Weslaco, recorded the highest

hourly $PM_{2.5}$ concentrations $(17.2 \pm 10.6 \,\mu g/m^3)$ compared to the other locations from July 04, 19:00 hours to July 05, 18:59 hours. In contrast, W2 site recorded the lowest $PM_{2.5}$ concentrations $(7.9 \pm 2.3 \,\mu g/m^3)$ for the same time frame.

In the city of Brownsville, across the five sampling sites one can observe a slight increase in mean PM_{2.5} concentrations on the day of Independence. For example, site B1 showed the mean concentration of $13.6 \pm 4.9 \,\mu\text{g/m}^3$ for between July 04, 19:00 hours- July 05, 18:59 hours. For the same time frame, site B3 had the lowest concentration $(8.8 \pm 2.9) \mu g/m^3$ and site B5 had the highest $(14.8 \pm 7.1) \mu g/m^3$ m³. In the city of Edinburg, hourly concentrations ranged from $11.1 \pm 2.7 \,\mu\text{g/m}^3$ at E3 to $14.1 \pm 4.1 \,\mu\text{g/m}^3$ at E2 for the same study period. Similarly, the sensor at Port Isabel (PI) showed a mean of $10.8 \pm 4.7 \,\mu\text{g/m}^3$. The PM_{2.5} concentrations reverted to control values starting July 05. As is obvious from Tables 1 and 2, the mean concentrations varied from the lowest at W2 $(5.3 \pm 2.1 \,\mu\text{g/m}^3)$ to the highest at B1 $(9.7 \pm 5.1 \,\mu\text{g/m}^3)$ and E2 $(9.7 \pm 4.1 \,\mu\text{g/m}^3)$. It is also important to mention here a striking similarity in the pollutant concentrations a day prior and after the 4th of July Independence Day celebrations as is obvious from Table 1.

Site W1 is located in a residential area as mentioned before and it is quite probable that the huge amount of fireworks lit in this neighborhood by the residents resulted in an hourly maximum of $51.2 \,\mu g/m^3$ on the Independence Day. For that same period, the maximum recorded at site W2 is only 12.3 μ g/m³. This makes sense because site W2 is located at one of the buildings of the Weslaco Police Department and there were no fireworks reported in this neighborhood. This same trend is observed in the city of Brownsville. Sites B1 and B5 were in residential neighborhoods where we expect people to light firecrackers. Hence the maximum concentration recorded at these two sites were $20.3 \,\mu\text{g/m}^3$ (B1) and $38.5 \,\mu\text{g/m}^3$ (B5). This is in sharp contrast to site B4 (adjacent to the interstate highway) where the maximum concentration was recorded as $15.5 \,\mu g/m^3$, as well as site B3 which is the University campus. These findings, therefore, suggest the role of fireworks in the high pollutant concentration observed during this festive period.

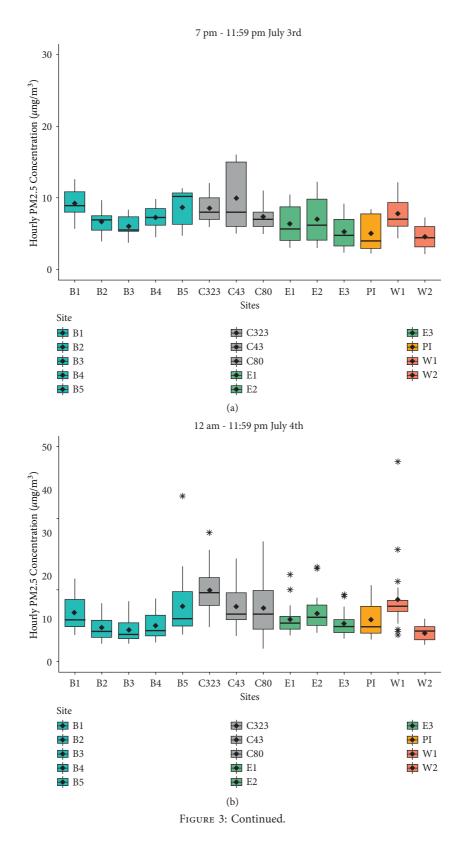
The three CAMS sites also exhibited higher $PM_{2.5}$ concentrations between July 04 and July 05. The mean concentration at C43 (Mission) was $19.54 \pm 5.96 \,\mu g/m^3$, C80 (Brownsville), $16.54 \pm 3.82 \,\mu g/m^3$, and C1046 (Edinburg), $18.17 \pm 5.38 \,\mu g/m^3$. The maximum hourly concentration recorded at C43 and C1046 for this period is $30 \,\mu g/m^3$. For the next 24-hour period, the $PM_{2.5}$ values at the three CAMS sites ranged from $9.67 \pm 6.03 \,\mu g/m^3$ (C80) to $12.08 \pm 4.83 \,\mu g/m^3$ (C43) suggesting a decrement in the pollutant concentrations after the firework festive period. These values between July 05 and July 06 (post celebrations) reflect the same low concentration trend as was observed between July 03 and July 04 (pre celebrations, $10.78 \pm 6.48 \,\mu g/m^3$ at C80 to $14.83 \pm 9.02 \,\mu g/m^3$ at C1046).

The United States Environmental Protection Agency is mandated by the United States Congress, under the auspices of the Clean Air Act of 1970 and its amendments thereof in 1990 and onwards, to set National Ambient Air Quality Standards (NAAQS) for the criteria air pollutants-one of which is $PM_{2.5}$. Currently, the 24-hour mean average for $PM_{2.5}$ is $35 \,\mu g/m^3$. Based on our research findings, we can posit that none of the study sites during the three-day study period exceeded the $PM_{2.5}$ NAAQS standards. However, it would be prudent to mention here that the daily mean $PM_{2.5}$ standard by the World Health Organization is $15 \,\mu g/m^3$ and our study sites are just within this stricter standard as well during the study period.

Figure 3 displays the box plots for the hourly $PM_{2.5}$ concentrations across the eleven sites and the three CAMS sites from July 03 to July 06. The color coordination for the box plots is as followed: blue for Brownsville, gray for the CAMS sites, green for Edinburg, gold for Port Isabel, and salmon for Weslaco. Box plots are helpful to detect any noticeable outliers, which are seen during for $PM_{2.5}$ concentrations on the Independence Day. The high values on this day and the outliers accentuate the role play by firework displays across most of the sites. Outliers are also observed for July 05 and July 06 for Port Isabel, close to South Padre Island-a major U.S. tourist destination that organizes massive firework displays.

3.2. Spatial and Temporal Variations in $PM_{2.5}$ Hourly Concentrations. COD matrix for PM_{2.5} concentrations across the eleven sites and the three CAMS sites are shown in Table 3. COD values equal to greater than 0.2 are in italic and bold. In Brownsville, spatial uniformity in PM_{2.5} concentrations are observed by values less than 0.2 for the following pairings: B1-B4 (0.15), B1-B5 (0.15), B2-B3 (0.1), B2-B4 (0.15), and B2-B5 (0.19). Values are equal to and greater than 0.2 suggest a slight spatial heterogeneity between sites B1-B2 (0.21), B1-B3 (0.26), B3-B4 (0.2), B3-B5 (0.23), and B4–B5 (0.21). Spatial variation is more pronounced between the CAMS site in Brownsville (C80) and the five Brownsville study sites. B1 (0.32), B2 (0.45), B3 (0.47), B4 (0.37), and B5 (0.42). These COD values suggest that central ambient sites such as TCEQ CAMS may not be an accurate representation of the actual PM_{2.5} exposure burden in any urban setting. The two sampling sites at Weslaco exhibit a COD value 0.3 indicating spatial nonuniformity. In the city of Edinburg, spatial homogeneity was observed across the three sampled sites with COD values less than 0.2. However, the three Edinburg sites demonstrated spatial nonuniformity in the $PM_{2.5}$ concentrations when compared with the CAMS C43 site in Mission (E1: 0.27, E2: 0.23, and E3: 0.30). Port Isabel sampling site also exhibited spatial heterogeneity in PM_{2.5} pollutant concentrations with all the sampled sites except for B1 (0.19) and B4 (0.17). The COD value (0.32) between PI site and C323 site in Port Isabel also exhibit slight spatial nonhomogeneity. These COD findings demonstrate the importance of intra- and interurban sampling of criteria air pollutants like fine particulate matter.

Spearman's rho correlation coefficients were computed between all the study and CAMS sites for fine particulate matter and are shown in Table 4. The correlation coefficients were all statistically significant at the 0.01 level. A stronger relationship across two simultaneously sample sites would have a higher value, whereas a weaker relationship would



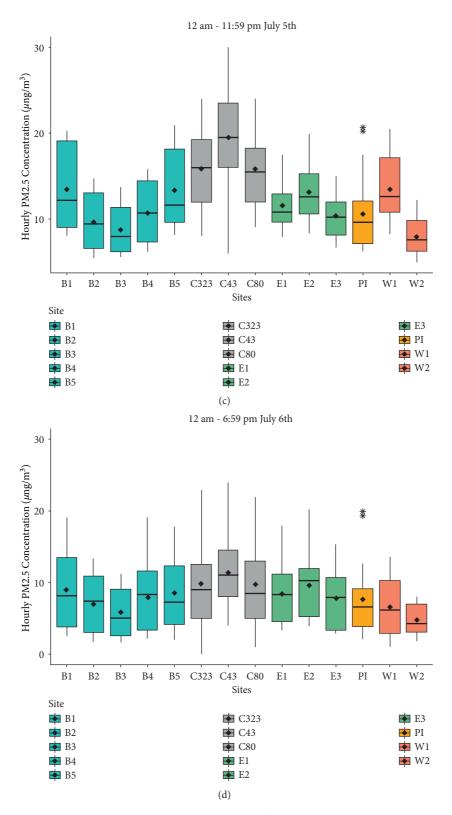


FIGURE 3: Box plots of hourly average concentrations of PM2.5 (µg/m³) from the study locations and the three CAMS sites.

indicate a lower coefficient. All the site pairings yielded a statistically significant and positive correlations. Across the city of Brownsville, all the sites were very strongly correlated with r > 0.8. Similarly, the three Edinburg sites were very

robustly correlated with r > 0.9. On similar lines, sites W1 and W2 were correlated with each other (r = 0.77). The Port Isabel site was positively correlated with all the sites exhibiting a very strong relationship with site (B1, r = 0.84)

TABLE 3: PM2.5 COD values between the study locations and the CAMS sites.

PM _{2.5}	B2	B3	B4	B5	E1	E2	E3	W1	W2	PI	C43	C80	C323
B1	0.21	0.26	0.15	0.15	0.23	0.19	0.22	0.26	0.3	0.19	0.29	0.32	0.27
B2		0.1	0.15	0.19	0.2	0.27	0.23	0.31	0.24	0.21	0.37	0.45	0.37
B3			0.2	0.23	0.24	0.31	0.25	0.33	0.23	0.23	0.40	0.47	0.40
B4				0.21	0.23	0.22	0.19	0.27	0.23	0.17	0.33	0.37	0.33
B5					0.19	0.22	0.26	0.28	0.33	0.24	0.30	0.42	0.30
E1						0.15	0.15	0.25	0.25	0.24	0.27	0.33	0.32
E2							0.13	0.23	0.28	0.23	0.23	0.28	0.27
E3								0.25	0.19	0.2	0.30	0.35	0.32
W1									0.3	0.31	0.29	0.40	0.29
W2										0.24	0.40	0.41	0.41
PI											0.36	0.29	0.32
C43												0.34	0.32
C80													0.28

TABLE 4: Correlation Coefficients for PM_{2.5} concentrations between the study locations and the CAMS sites.

	B1	B2	B3	B4	B5	E1	E2	E3	W1	W2	PI	C43	C80	C323
B1	1													
B2	0.923**	1												
B3	0.922**	0.955**	1											
B4	0.919**	0.937**	0.861**	1										
B5	0.940**	0.915**	0.934**	0.844**	1									
E1	0.522**	0.519**	0.524**	0.531**	0.606**	1								
E2	0.564**	0.496**	0.512**	0.521**	0.601**	0.954**	1							
E3	0.527**	0.437**	0.451**	0.500**	0.510**	0.915**	0.952**	1						
W1	0.416**	0.384**	0.424^{**}	0.409**	0.478^{**}	0.652**	0.637**	0.557**	1					
W2	0.639**	0.572**	0.583**	0.627**	0.667**	0.867**	0.874^{**}	0.795**	0.772**	1				
PI	0.839**	0.777**	0.743**	0.803**	0.767**	0.485**	0.525**	0.522**	0.352**	0.602**	1			
C43	0.275^{*}	0.246^{*}	0.292*	0.233*	0.361**	0.640**	0.607**	0.557**	0.474^{**}	0.519**	00.178	1		
C80	0.820**	0.701**	0.752**	0.716**	0.784**	0.638**	0.693**	0.686**	0.523**	0.661**	0.704**	0.506**	1	
C323	0.510**	0.411**	0.455**	0.399**	0.519**	0.463**	0.547**	0.508**	0.487**	0.501**	0.601**	0.292*	0.680**	1

** Correlation is significant at the 0.01 level (2-tailed), N=71-72 for all pairs.

and weakly correlated with site W1 (r = 0.35). The CAMS site C80 was also very strongly correlated with all the study locations (0.506 < r < 0.820). C323 was, however, weakly correlated with C43 (r = 0.292). All the study sites were weakly to moderately correlated with C43 (0.233 < r < 0.640) except for the Port Isabel study location (r = 0.178). The correlation coefficient values suggest that across the lower RGV region landscape ubiquity in temporal increase or decrease of PM_{2.5} concentrations is observed.

4. Conclusions

To the best of our knowledge, the findings from this study are the first in the lower RGV region of South Texas, characterizing fine particulate matter pollution across eleven sites before, during, and after the 4th of July Independence Day celebrations. These celebrations are accompanied by the display of massive fireworks display resulting in short-term increase in particulate matter pollution. Whereas, our study findings do not exceed the US NAAQS 24-hr mean PM_{2.5} concentrations, even a short-term exposure of a few hours to high PM_{2.5} levels can result in respiratory health issues such a wheezing, tightness of chest, and asthma aggravation and subsequent asthma attacks. Findings from this study may, therefore, contribute toward the future formulation of policies for firework display at the local level. This study demonstrates the usefulness of using low-cost sensors to characterize intraurban spatial variability in fine particulate matter concentrations as well as address fine-scale changes in $PM_{2.5}$ pollution, which is inadequately captured by federal and state fixed continuous ambient monitoring sites.

Data Availability

The data used to support the findings of this study are available from the corresponding author upon request.

Conflicts of Interest

The authors declare that they have no conflicts of interest.

Authors' Contributions

A.U.R., D.W., K.S., and O.T. conceived and designed the study. E.M. implemented the study. A.U.R., O.T., and E.M. supervised the data collection. E.M. analyzed the data and

wrote the initial draft of the manuscript. A.U.R. edited and prepared the final draft. D.W., K.S., and O.T. gave valuable comments on the initial draft. All authors provided valuable comments and ideas while drafting the manuscript. All authors read and approved the final draft of the manuscript.

Acknowledgments

The authors express their gratitude to various field staff for help with the deployment of air quality instruments. Funding for this research work was graciously provided by the UTRGV under the new faculty startup funds provided to A.U.R. Graduate assistantship for E.M. was kindly provided by School of Earth, Environment, and Marine Sciences, UTRGV.

References

- G. Beig, D. M. Chate, S. D. Ghude et al., "Evaluating population exposure to environmental pollutants during Deepavali fireworks displays using air quality measurements of the SAFAR network," *Chemosphere*, vol. 92, no. 1, pp. 116–124, 2013.
- [2] C. C. Lin, "A review of the impact of fireworks on particulate matter in ambient air," *Journal of the Air & Waste Man*agement Association, vol. 66, no. 12, pp. 1171–1182, 2016.
- [3] C. Wu, G. Wang, J. Wang et al., "Chemical characteristics of haze particles in Xi'an during Chinese Spring Festival: impact of fireworks burning," *Journal of Environmental Sciences*, vol. 71, pp. 179–187, 2018.
- [4] X. Zhang, H. Shen, T. Li, and L. Zhang, "The effects of fireworks discharge on atmospheric PM2.5 concentration in the Chinese lunar new year," *International Journal of Environmental Research and Public Health*, vol. 17, no. 24, p. 9333, 2020.
- [5] J. Chen and G. Hoek, "Long-term exposure to pm and allcause and causespecific mortality: a systematic review and meta-analysis," *Environment International*, vol. 143, Article ID 105974, 2020.
- [6] B. Liu, C. Wu, N. Ma et al., "Vertical profiling of fine particulate matter and black carbon by using unmanned aerial vehicle in Macau, China," *Science of the Total Environment*, vol. 709, Article ID 136109, 2020.
- [7] A. Mousavi, Y. Yuan, S. Masri, G. Barta, and J. Wu, "Impact of 4th of july fireworks on spatiotemporal PM2.5 concentrations in California based on the PurpleAir sensor network: implications for policy and environmental justice," *International Journal of Environmental Research and Public Health*, vol. 18, no. 11, p. 5735, 2021.
- [8] X.-Y. Zheng, P. Orellano, H.-L. Lin, M. Jiang, and W.-J. Guan, "Short-term exposure to ozone, nitrogen dioxide, and sulphur dioxide and emergency department visits and hospital admissions due to asthma: a systematic review and metaanalysis," *Environment International*, vol. 150, Article ID 106435, 2021.
- [9] R. E. Connolly, Q. Yu, Z. Wang et al., "Long-term evaluation of a low-cost air sensor network for monitoring indoor and outdoor air quality at the community scale," *Science of the Total Environment*, vol. 807, Article ID 150797, 2022.
- [10] M. C. Kelley, M. M. Brown, C. B. Fedler, and K. Ardon-Dryer, "Long-term measurements of PM2.5 concentrations in lubbock, Texas," *Aerosol and Air Quality Research*, vol. 20, no. 6, pp. 1306–1318, 2020.

- [11] J. Li, T. Xu, X. Lu et al., "Online single particle measurement of fireworks pollution during Chinese New Year in Nanning," *Journal of Environmental Sciences*, vol. 53, pp. 184–195, 2017.
- [12] B. Duo, Y. Zhang, L. Kong et al., "Individual particle analysis of aerosols collected at Lhasa City in the Tibetan Plateau," *Journal of Environmental Sciences*, vol. 29, pp. 165–177, 2015.
- [13] S. Tiwari, D. M. Chate, M. K. Srivastava et al., "Statistical evaluation of PM10 and distribution of PM1, PM2.5, and PM10 in ambient air due to extreme fireworks episodes (Deepawali festivals) in megacity Delhi," *Natural Hazards*, vol. 61, no. 2, pp. 521–531, 2011.
- [14] C. Gouder and S. Montefort, "Potential impact of fireworks on respiratory health," *Lung India*, vol. 31, no. 4, pp. 375–379, 2014.
- [15] C. Hickey, C. Gordon, K. Galdanes et al., "Toxicity of particles emitted by fireworks," *Particle and Fibre Toxicology*, vol. 17, no. 1, p. 28, 2020.
- [16] W. Li, Z. Shi, C. Yan, L. Yang, C. Dong, and W. Wang, "Individual metal-bearing particles in a regional haze caused by firecracker and firework emissions," *Science of the Total Environment*, vol. 443, pp. 464–469, 2013.
- [17] Y. Wang, G. Zhuang, C. Xu, and Z. An, "The air pollution caused by the burning of fireworks during the lantern festival in Beijing," *Atmospheric Environment*, vol. 41, no. 2, pp. 417–431, 2007.
- [18] L. Yao, D. Wang, Q. Fu et al., "The effects of firework regulation on air quality and public health during the Chinese Spring Festival from 2013 to 2017 in a Chinese megacity," *Environment International*, vol. 126, pp. 96–106, 2019.
- [19] M. Bencardino, V. Andreoli, J. Castagna et al., "Airborne particles during a firework festival in belvedere M.mo, southwestern Italian Coast," *Open Journal of Air Pollution*, vol. 7, no. 2, pp. 156–180, 2018.
- [20] D. J. Seidel and A. N. Birnbaum, "Effects of Independence Day fireworks on atmospheric concentrations of fine particulate matter in the United States," *Atmospheric Environment*, vol. 115, pp. 192–198, 2015.
- [21] J. Zhang, L. Yang, J. Chen et al., "Influence of fireworks displays on the chemical characteristics of PM2.5 in rural and suburban areas in Central and East China," *Science of the Total Environment*, vol. 578, pp. 476–484, 2017.
- [22] Y. Zhang, J. Wei, A. Tang, A. Zheng, Z. Shao, and X. Liu, "Chemical characteristics of PM2.5 during 2015 spring festival in beijing, China," *Aerosol and Air Quality Research*, vol. 17, no. 5, pp. 1169–1180, 2017.
- [23] A. Joly, A. Smargiassi, T. Kosatsky et al., "Characterisation of particulate exposure during fireworks displays," *Atmospheric Environment*, vol. 44, no. 34, pp. 4325–4329, 2010.
- [24] X. Pang, L. Chen, K. Shi et al., "A lightweight low-cost and multipollutant sensor package for aerial observations of air pollutants in atmospheric boundary layer," *Science of the Total Environment*, vol. 764, Article ID 142828, 2021.
- [25] B. Feenstra, V. Papapostolou, S. Hasheminassab et al., "Performance evaluation of twelve low-cost PM2.5 sensors at an ambient air monitoring site," *Atmospheric Environment*, vol. 216, Article ID 116946, 2019.
- [26] "2021 demographics," 2021, https://ucd.thehcn.net/ demographicdata?id=281259§ionId=935.
- [27] BlueSky Air Quality Monitor Operation and Maintenance Manual, TSI Inc, Shoreview, MN, USA, 2021, https://tsi.com/ getmedia/a9299d7b-de37-4177-ab3b-488dfbfd2d07/BlueSky_ Op_Maint_Manual_6013929?ext=.pdf.
- [28] L. Morawska, P. K. Thai, X. Liu et al., "Applications of lowcost sensing technologies for air quality monitoring and

exposure asseesment: how far have they gone?" *Environment International*, vol. 116, pp. 286–299, 2018.

- [29] "AQMD air quality sensor performance evaluation center," 2021, https://www.aqmd.gov/aq-spec/evaluations/summarypm.
- [30] M. A. Krudysz, S. J. Dutton, G. L. Brinkman et al., "Intracommunity spatial variation of size-fractionated organic compounds in Long Beach, California," *Air Quality, Atmosphere & Health*, vol. 2, pp. 69–88, 2009.
- [31] J. P. Pinto, A. S. Lefohn, and D. S. Shadwick, "Spatial variability of PM2.5 in urban areas in the United States," *Journal* of the Air & Waste Management Association, vol. 54, no. 4, pp. 440–449, 2004.
- [32] A. U. Raysoni, J. A. Sarnat, S. E. Sarnat et al., "Binational school-based monitoring of traffic-related air pollutants in el paso, Texas (USA) and ciudad juarez, chihuahua (Mexico)," *Environmental Pollution*, vol. 159, no. 10, pp. 2476–2486, 2011.