

1 **Continental-scale niche differentiation of dominant topsoil archaea in**
2 **drylands**

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4 Running title: Topsoil archaeal niche differentiation in drylands

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29 **Summary**

30 Archaea represent a diverse group of microorganisms often associated with extreme environments.
31 However, an integrated understanding of biogeographical patterns of the specialist *Haloarchaea*
32 and the potential generalist ammonia-oxidizing archaea (AOA) across large-scale environmental
33 gradients remains limited. We hypothesize that niche differentiation determines their distinct
34 distributions along environmental gradients. To test the hypothesis, we use a continental-scale
35 research network including 173 dryland sites across northern China. Our results demonstrate that
36 *Haloarchaea* and AOA dominate topsoil archaeal communities. As hypothesized, *Haloarchaea*
37 and AOA show strong niche differentiation associated with two ecosystem types mainly found in
38 China's drylands (i.e., deserts vs. grasslands), and they differ in the degree of habitat
39 specialization. The relative abundance and richness of *Haloarchaea* are higher in deserts due to
40 specialization to relatively high soil salinity and extreme climates, while those of AOA are greater
41 in grassland soils. Our results further indicate a divergence in ecological processes underlying the
42 segregated distributions of *Haloarchaea* and AOA. *Haloarchaea* are governed primarily by
43 environmental-based processes while the more generalist AOA are assembled mostly via
44 spatial-based processes. Our findings add to existing knowledge of large-scale biogeography of
45 topsoil archaea, advancing our predictive understanding on changes in topsoil archaeal
46 communities in a drier world.

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55 **Originality-Significance Statement**

56 This work contributes to the current limited understanding on the large-scale biogeography of
57 *Haloarchaea* and ammonia-oxidizing archaea (AOA) in drylands. Our study is original and
58 significant, as it reveals a strong niche differentiation between these two dominant topsoil archaeal
59 groups, primarily driven by habitat specialization associated with contrasting ecosystem types (i.e.,
60 deserts and grasslands) rather than small-scale microsites (i.e., bare ground and vegetated areas).
61 Moreover, this work also provides new insights into the community assembly processes
62 underpinning the distinct biogeographical patterns of *Haloarchaea* and AOA. It reveals that the
63 distribution of *Haloarchaea* is mainly determined by environmental-based processes, while AOA
64 are more influenced by spatial-based processes. These observations are important to understand
65 future changes in key soil microbial taxa in drylands under climate change.

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79 **Introduction**

80 Archaea is one of the most abundant and diverse groups of organisms on Earth, which can
81 colonize a wide range of environments (Chaban *et al.*, 2006). Moreover, there is a growing
82 appreciation of the astonishing diversity, life history, and metabolic versatility of archaea, as well
83 as their contributions to global biogeochemistry, particularly the carbon (C) and nitrogen (N)
84 cycles (Offre *et al.*, 2013; Baker *et al.*, 2020). Despite their ecological importance, however,
85 biogeographical patterns of archaea at large spatial scales remain relatively unexplored,
86 particularly compared with those of bacteria and fungi. This limit of knowledge is therefore
87 hampering our understanding of their ecology in natural ecosystems.

88 Previous large-scale studies on bacterial and fungal biogeography have collectively shown
89 that diversity and community composition of these organisms are strongly partitioned by
90 ecosystem types, biomes or geographic regions associated with particular environmental
91 conditions. That is, there is a strong role for niche differentiation, which can be attributed to
92 environmental filtering and/or biotic interactions (Fierer and Jackson, 2006; Lozupone and Knight,
93 2007; Tedersoo *et al.*, 2014; Bahram *et al.*, 2018; Delgado-Baquerizo *et al.*, 2018; Egidi *et al.*,
94 2019). Strong evidence, from experimental work and local studies, also support the existence of
95 distinct adaptive strategies and metabolic lifestyles within contrasting archaeal types (Walsh *et al.*,
96 2005; Jiang *et al.*, 2007; Ke *et al.*, 2014; Rath *et al.*, 2019). However, empirical evidence from
97 environmental gradients at large spatial scales is still limited.

98 We posit that niche differentiation can explain the distinct distributions of archaeal groups
99 with contrasting traits (e.g., specialists vs. generalists) across large-scale environmental gradients
100 (Barberán *et al.*, 2012). For example, *Haloarchaea* are well-known specialist archaea inhabiting a
101 specific and restricted range of environmental conditions (Auguet *et al.*, 2010). Conversely,
102 ammonia-oxidizing archaea (AOA) affiliated with the phylum *Thaumarchaeota* represent one of
103 the most abundant and ubiquitous groups of archaea dominating most soil and aquatic systems
104 (Francis *et al.*, 2005; Leininger *et al.*, 2006; Auguet *et al.*, 2010; Bates *et al.*, 2011). In particular,
105 AOA have been suggested to be resistant to conditions of low energy availability (e.g., water and
106 nutrient) such as those often found in arid ecosystems (Valentine, 2007; Adair and Schwartz, 2008;
107 Bates *et al.*, 2011; Trivedi *et al.*, 2019). As such, AOA appear to be generalist archaea showing

108 broader environmental tolerances than *Haloarchaea* (Erguder *et al.*, 2009). Recent studies provide
109 evidence for niche differentiation between AOA and their bacterial counterparts
110 (ammonia-oxidizing bacteria; AOB) in dryland soils driven by the presence of perennial plants
111 (Delgado-Baquerizo *et al.*, 2013a; Trivedi *et al.*, 2019), yet whether niche partitioning also occurs
112 between AOA and *Haloarchaea* remains poorly understood.

113 Given the ecological significance of *Haloarchaea* and AOA, a comprehensive investigation
114 of the ecological preferences of these archaeal groups across large-scale natural ecosystems is
115 needed. We argue that *Haloarchaea* and AOA may differ in niche breadth and thus should show
116 contrasting distributions across large-scale environmental gradients. *Haloarchaea*, as specialists,
117 are expected to thrive in the extremes of environmental gradients, leading to restricted
118 distributions as previously observed in communities of plants (Berdugo *et al.*, 2019), invertebrates
119 (Pandit *et al.*, 2009), birds (Devictor *et al.*, 2008), and bacterioplankton (Logares *et al.*, 2013). In
120 contrast, AOA are expected to be generalists with greater niche breadth. This is further expected to
121 cause a tradeoff such that AOA are relatively weak competitors under environmental conditions
122 where *Haloarchaea* are favoured, but AOA may inhabit a larger habitat range than *Haloarchaea*
123 (*sensu* Fournier *et al.*, 2020). However, given that AOA can tolerate a wide range of
124 environmental conditions (Erguder *et al.*, 2009) and are likely to coexist with *Haloarchaea* in
125 some local environments (Auguet *et al.*, 2010; Shi *et al.*, 2016; Rath *et al.*, 2019), it is not
126 immediately clear how the biogeographies of these two archaeal groups will be related to each
127 other across large-scale environmental gradients.

128 To test the hypotheses stated above, we use a continental-scale research network including
129 173 dryland sites across northern China (Fig. S1). These study sites cover > 3 500 000 km² and
130 span a distance of 4 000 km from northeast to northwest of China, representing a wide range of
131 environmental gradients including plant attributes, climatic conditions, and soil properties (Hu *et al.*,
132 2021). Drylands across northern China are highly heterogeneous environments that are
133 typically formed by two major ecosystem types (i.e., deserts and grasslands) (Chinese Academy of
134 Sciences, 2001) (Fig. S1).

135 Several previous studies have suggested that ecosystem type plays a crucial role in
136 structuring the composition of soil microbial communities from local to global scales (Bates *et al.*,

137 2011; Chu *et al.*, 2011; Griffiths *et al.*, 2011; Delgado-Baquerizo *et al.*, 2018; Egid *et al.*, 2019).
138 Different ecosystem types reflect different environmental conditions; most directly relevant to the
139 questions herein, the transitions from grasslands to deserts may represent an increase in the degree
140 of habitat specialization of soil archaeal communities (Fig. S2). Specifically, deserts are generally
141 considered as more extreme ecosystems that have higher soil salinity and pH and their climates are
142 characterized by higher temperature and aridity and lower precipitation, as compared to grasslands.
143 We therefore posit that niche differentiation between *Haloarchaea* and AOA in drylands across
144 northern China is associated with ecosystem type (Fig. S2). More specifically, *Haloarchaea* are
145 expected to be specialized to thrive in more extreme desert ecosystems but have a more restricted
146 distribution in grasslands. Hence, we expect a stronger influence of environmental-based
147 processes (e.g., environmental selection) over *Haloarchaea* (Pandit *et al.*, 2009; Logares *et al.*,
148 2013; Liao *et al.*, 2016). Conversely, AOA are expected to be found in both deserts and grasslands
149 but prefer the latter in which fitness advantages of *Haloarchaea* are largely reduced. Here, a
150 stronger role of spatial-based processes (e.g., probabilistic dispersal) is expected because AOA, as
151 generalists, are thought to be less influenced by environmental selection (Pandit *et al.*, 2009;
152 Langenheder and Székely, 2011; Liao *et al.*, 2016) (Fig. S2). The coexistence of these two distinct
153 ecosystem types in China's drylands provides a unique opportunity to test our hypotheses under
154 field conditions, and to address a critical research gap in microbial ecology.

155

156 **Results**

157 *Taxonomic composition of topsoil archaea*

158 After correcting for the differences in sequencing depth, we retained 1 871 archaeal operational
159 taxonomic units (OTUs). *Haloarchaea* and *Thaumarchaeota* dominated the topsoil archaeal
160 communities in drylands across northern China, together accounting for 57% of the OTUs and 89%
161 of 16S rRNA gene sequences (Fig. 1a). *Thermoplasmata*, deep-sea hydrothermal vent
162 *Euryarchaeota* group (DSHVEG) and *Nanohaloarchaeota* were relatively diverse, encompassing
163 between 4.5 and 22% of the OTUs, but less abundant (~10.5% of the sequences). Other high-rank
164 archaeal taxa, including *Methanomicrobia*, *Lokiarchaeota*, *Aenigmarchaeota*, *Bathyarchaeota*,
165 *Archaeoglobi* and *Methanobacteria*, represented a minor fraction of the archaeal communities

166 (~1.5% of the OTUs and < 0.05% of the sequences).

167 Haloarchaeal OTUs can be assigned to three major orders, with *Halobacteriales* being most
168 abundant, followed by *Natrialbales* and *Haloferacales* (Fig. 1b,c). Thaumarchaeotal OTUs fell
169 into three identifiable orders belonging to three AOA lineages: *Nitrosopumilales* of Group I.1a,
170 *Nitrososphaerales* of Group I.1b, and *Nitrosocaldales* of HWCG-III, among which
171 *Nitrososphaerales* accounted for > 99.9% of thaumarchaeotal sequences (Fig. 1b,c). Our network
172 analyses showed that haloarchaeal and thaumarchaeotal OTUs tended to co-occur with others
173 belonging to the same taxa and generated two independent clusters that had extremely few
174 connections with each other (Figs. 1d, S3). Such a contrasting co-occurrence pattern indicates that
175 OTUs affiliated with these two dominant archaeal taxa may differ in ecological preferences.

176

177 *Niche breadth and ecological distributions of Haloarchaea and Thaumarchaeota*

178 Both the Levins' index and the tolerance index were significantly higher for thaumarchaeotal
179 OTUs than those of *Haloarchaea* (Mann-Whitney U test, both $P < 0.001$; Fig. 2a,b). Moreover,
180 both the relative abundance and richness of *Haloarchaea* were much higher in deserts than in
181 grasslands (Mann-Whitney U test, both $P < 0.001$; Fig. 2c–f). Conversely, *Thaumarchaeota* were
182 distributed widely in both deserts and grasslands, but were more abundant and diverse in the latter
183 (Mann-Whitney U test, both $P < 0.001$; Fig. 2c,d,g,h). Consistent results were observed for the
184 sub-dataset (i.e., study sites without missing environmental data; see also Experimental
185 Procedures) (Mann-Whitney U test, all $P < 0.01$; Fig. S4).

186 To ensure that merging data across vegetated and non-vegetated microsites (i.e., vegetated
187 areas and bare ground; see also Experimental Procedures) was not influencing our results, we
188 examined the niche breadth of *Haloarchaea* and *Thaumarchaeota* for these two potentially
189 different niches separately. Further, we compared both the relative abundance and richness of
190 *Haloarchaea* and *Thaumarchaeota* between the two microsites. Our results showed consistently
191 higher values of the Levins' index and the tolerance index for *Thaumarchaeota* than those of
192 *Haloarchaea* in both the microsites (Mann-Whitney U test, all $P < 0.001$); however, the two
193 indices had no differences between vegetated and bare ground microsites for both the archaeal
194 groups (Mann-Whitney U test, all $P > 0.05$) (Fig. S5). Moreover, we also showed similar relative

195 abundance and richness of *Haloarchaea* and *Thaumarchaeota* between vegetated areas and bare
196 ground (Mann-Whitney U test, all $P > 0.05$; Fig. 2c,d,i-l). The similarities in both niche breadth
197 and distribution patterns across the two microsites can be, at least partially, attributed to the
198 relatively homogeneous soil properties and nutrient contents between the microsites (Fig. S6).

199

200 *Environmental predictors of distributions of Haloarchaea and Thaumarchaeota*

201 Random forest models identified electrical conductivity, a common indicator of salinity (soil
202 dissolved salt), as the most important and significant predictor for the distributions of
203 *Haloarchaea* and *Thaumarchaeota* (Fig. S7). Ecosystem type (deserts = 1 vs. grasslands = 0) was
204 also identified as a major predictor in all models. Partial Spearman's correlations further
205 confirmed positive relationships between ecosystem type and the relative abundance and richness
206 of *Haloarchaea* after controlling for spatial influences (Fig. 3). Accordingly, we also found
207 negative relationships of ecosystem type with both thaumarchaeotal attributes. Given that deserts
208 ($4576.2 \pm 1594.3 \mu\text{S cm}^{-1}$; $N = 82$) had a significantly higher soil electrical conductivity than
209 grasslands ($300.4 \pm 112.2 \mu\text{S cm}^{-1}$; $N = 51$) (Fig. S8), our results indicated that changes in soil
210 salinity associated with different ecosystem types largely explain the changes in the large-scale
211 distribution patterns of *Haloarchaea* and *Thaumarchaeota* (Fig. 3). Similarly, higher relative
212 abundance and richness of *Thaumarchaeota* in grasslands could be partially attributed to grassland
213 soils having relatively higher total N [$0.22 \pm 0.02 \text{ g kg}^{-1}$ ($N = 82$) and $1.47 \pm 0.18 \text{ g kg}^{-1}$ ($N = 51$)
214 for deserts and grasslands, respectively] and ammonium [$3.11 \pm 0.59 \text{ mg kg}^{-1}$ ($N = 82$) and $4.76 \pm$
215 0.51 mg kg^{-1} ($N = 51$) for deserts and grasslands, respectively] —in these nutrient poor
216 ecosystems—, and plant species richness [5.5 ± 0.5 ($N = 82$) and 24 ± 2 ($N = 51$) for deserts and
217 grasslands, respectively] (Figs. S8, S9). These three variables all showed positive associations
218 with thaumarchaeotal richness and relative abundance, which is likely due to the direct
219 mechanistic connection *Thaumarchaeota* have with N cycling (Fig. 3).

220 Furthermore, the contrasting distribution patterns of *Haloarchaea* and *Thaumarchaeota*
221 between deserts and grasslands could also be mediated by certain climatic factors (Fig. 3). For
222 instance, mean annual temperature (MAT), mean diurnal temperature range (MDTR),
223 isothermality and aridity, which were higher in desert ecosystems (Fig. S9), were positively

224 associated with haloarchaeal attributes. Conversely, negative correlations were found between
225 both thaumarchaeotal attributes and MDTR, isothermality and aridity. Interestingly, our results
226 revealed strong and negative correlations among haloarchaeal and thaumarchaeotal attributes (Fig.
227 3).

228

229 *Ecological preferences of Haloarchaea and Thaumarchaeota*

230 Our mapping effort based on results from four ecological clusters of haloarchaeal and
231 thaumarchaeotal OTUs sharing ecological preferences further supported a strong influence of
232 niche differentiation (Figs. 4a,b, 5a,b). *Haloarchaea* were assigned exclusively to the desert and
233 high-salinity clusters (331 and 561 OTUs, respectively), whereas *Thaumarchaeota* were included
234 in all clusters but more frequently present in the grassland (80 vs. 57 OTUs) and low-salinity (122
235 vs. 2 OTUs) clusters. Our continental maps accounted for 70 to 94% of the spatial variation in
236 these clusters. Also, strong relationships of ecosystem type and electrical conductivity with the
237 relative abundance of corresponding clusters indicated that these were reasonably well-defined
238 and ecologically meaningful clusters (Figs. 4a,b, 5a,b). Overall, our maps provided predictions on
239 the regions where the four ecological clusters were expected to be most abundant. As expected, an
240 opposing spatial distribution pattern was found between the desert and grassland clusters and
241 between the high- and low-salinity clusters. For example, the high-salinity cluster showed a
242 narrower distribution compared with the desert cluster and was particularly abundant in the most
243 arid desert regions (e.g., Taklamakan, Kumtag and Qaidam Deserts) where their soils are often
244 hypersaline. However, the low-salinity cluster revealed a similar distribution with the grassland
245 cluster and had greater relative abundance in moist grassland regions (e.g., central and eastern
246 Inner Mongolia) known for their relatively low-salinity soils. Our network analyses showed that
247 nodes within the same ecological clusters were more connected, indicating that haloarchaeal and
248 thaumarchaeotal OTUs sharing a particular ecological preference tended to co-occur with each
249 other (Figs. 4c,d, 5c,d, S10).

250

251 *Ecological processes governing the assembly of Haloarchaea and Thaumarchaeota*

252 Relative contributions of major ecological processes structuring community assembly differed

253 between *Haloarchaea* and *Thaumarchaeota* (Fig. 6). *Haloarchaea* were associated with a greater
254 influence of homogeneous selection (41.8%) than *Thaumarchaeota* (3.5%). By contrast, a greater
255 contribution from ecological drift enabled by dispersal limitation was observed for
256 *Thaumarchaeota* (58.6%) than for *Haloarchaea* (23.9%). In sum, *Haloarchaea* were largely
257 driven by environmental-based processes, as expected for specialists, while *Thaumarchaeota* were
258 primarily assembled by spatial-based processes, as expected for generalists. Also, consistent
259 results were observed when the vegetated and bare ground microsites were analyzed separately.
260 That is, the relative influences of different community assembly processes were similar between
261 the two microsites for both *Haloarchaea* and *Thaumarchaeota* (Fig. S11). Together with our
262 results above (Figs. 2c,d,i-l, S5), the consistency in assembly processes across the two microsites
263 further indicates that merging data across the microsites should not affect the conclusions of this
264 study.

265

266 **Discussion**

267 Our results showed that topsoil archaeal communities across China's drylands were dominated by
268 *Haloarchaea* and AOA, and aligned with our hypotheses and suggested that *Haloarchaea* tended
269 to be specialists governed by environmental-based processes whereas AOA were generalists
270 governed by spatial-based processes. We also observed that relative to grasslands, desert soils had
271 higher relative abundance and richness of *Haloarchaea* and lower relative abundance and richness
272 of AOA. Our findings provided strong evidence for niche differentiation between these two
273 archaeal groups across large-scale environmental gradients, primarily driven by habitat
274 specialization associated with contrasting ecosystem types.

275 Our results indicate that niche specialization of *Haloarchaea* to salt is likely a key
276 determinant of the clear differentiation of *Haloarchaea* and AOA observed between deserts and
277 grasslands, which can be linked to their highly specialized osmoadaptive strategy. The most
278 prevalent strategy for osmoadaptation performed by *Haloarchaea* is "salt-in", which involves the
279 accumulation of potassium ions to balance osmotic pressure and maintain their requirements for
280 salts to maintain growth and structural stability (Oren, 2008). Because this strategy requires

281 extensive adaptation of the intracellular enzymatic machinery to high salt concentrations (Oren,
282 2008, 2013), *Haloarchaea* generally can't thrive in low-salinity environments but thrive only
283 under extreme conditions with high salinity (Oren, 1994; Walsh *et al.*, 2005; Jiang *et al.*, 2007;
284 Auguet *et al.*, 2010; Rath *et al.*, 2019; Zhang *et al.*, 2019). Niche specialization to salt endows
285 *Haloarchaea* with a competitive advantage over the more generalist AOA in deserts, where high
286 soil salinity selectively enriched *Haloarchaea* which then may outcompete AOA due to their
287 relatively low-salinity tolerance (*sensu* Fournier *et al.*, 2020). Salinity-driven competitive
288 suppression was observed previously between *Haloarchaea* and *Thaumarchaeota* and other
289 archaeal taxa (Walsh *et al.*, 2005; Jiang *et al.*, 2007; Shi *et al.*, 2016; Rath *et al.*, 2019).
290 Competitive suppression associated with salinity is further supported by the strong negative
291 correlations between haloarchaeal and thaumarchaeotal attributes along the salinity gradient across
292 our study sites (see Fig. S12). Furthermore, nitrification is generally inhibited at high salt
293 concentrations (Oren, 1994), thereby putting AOA at a further competitive disadvantage in desert
294 soils. We infer that the lower relative abundance and richness of AOA in deserts is likely due to
295 both competitive interactions and direct negative effects of soil salinity on the availability of
296 ammonium, the main inorganic N source for AOA (Mosier and Francis, 2008; Moin *et al.*, 2009;
297 Delgado-Baquerizo *et al.*, 2013a).

298 Grasslands contrast with deserts in many ways, such as grassland soils having lower salinity,
299 higher N availability, and supporting higher plant diversity. The lower salinity in grassland soils
300 likely changes the competitive dynamics to facilitate AOA, with additional influences from N
301 availability and plants. Specifically, in the transitions from deserts to grasslands as salinity
302 decreases and N availability increases, AOA become more competitive over *Haloarchaea*. We
303 must note that, however, as N availability keeps increasing AOA will be outcompeted by AOB
304 (Delgado-Baquerizo *et al.*, 2013a; Trivedi *et al.*, 2019). Thus, our results suggest that the
305 relationship between N availability and the relative abundance and richness of AOA may be
306 unimodal and further modified by salinity. In general, AOA tend to dominate in soils with
307 ammonium concentrations below 15 mg kg⁻¹, whereas AOB become more competitive when
308 ammonium concentrations are above 100 mg kg⁻¹ (Pester *et al.*, 2011). However, for this study the
309 N availability in grasslands is just high enough to favor AOA but not so high to select for AOB

310 (Fig. S8). Moreover, soil AOA have been found to have a specific association with plant roots
311 (Simon *et al.*, 2000; Ke *et al.*, 2013, 2014), and their abundance and diversity have been reported
312 to be influenced greatly by root exudates (Sliwinski and Goodman, 2004; Simon *et al.*, 2005;
313 Chen *et al.*, 2008), indicating a profound plant–soil–AOA interaction. Because plant species
314 generally differ in both the composition and amount of root exudates (Grayston *et al.*, 1998),
315 individual plant species may select for specific AOA communities in soil via bottom-up forces
316 (Nicol *et al.*, 2005). As a result, the higher plant diversity of grasslands may combine with lower
317 salinity and higher N availability, relative to deserts, to facilitate higher relative abundance and
318 richness of soil AOA.

319 Further, we emphasize that the observed ecosystem type-associated niche differentiation may
320 not be independent of climate, which is a major aspect of the Earth system that affects the
321 structure and functioning of drylands (Maestre *et al.*, 2016; Berdugo *et al.*, 2020). Our results
322 indicate that climatic factors could mediate the degree of habitat specialization of *Haloarchaea* to
323 desert ecosystems indirectly via influences on soil salinity. For example, the high temperature, low
324 precipitation and extreme temperature fluctuations typical of desert ecosystems could increase
325 salinity via soil drying and accumulations of inorganic minerals (e.g., halite, gypsum, and
326 carbonates) (Pointing and Belnap, 2012; Delgado-Baquerizo *et al.*, 2013b; Genderjahn *et al.*,
327 2018). Our results support this, as soil salinity was found to be positively correlated with MAT,
328 MDTR, isothermality and aridity across our studied sites (see Fig. S13). On the other hand, we
329 found negative correlations of MDTR, isothermality and aridity with total soil N, ammonium, and
330 plant species richness (see Fig. S13). These results coincide with previous studies indicating that
331 both increasing aridity and temperature variability decreased soil N availability and plant diversity
332 in drylands worldwide (Delgado-Baquerizo *et al.*, 2013b, 2016a; Zhang *et al.*, 2018; Berdugo *et*
333 *al.*, 2020), and also explain the observed negative associations of aridity and temperature
334 variability with the relative abundance and richness of AOA. Taken together, our results suggest
335 that future changes in temperature and/or precipitation regimes could modify the distributions of
336 the main topsoil archaeal groups in drylands across northern China.

337 One goal of understanding drivers of microbial community composition is to use that
338 knowledge to better understand differences in the spatial distributions of different microbial taxa

339 and ecologically coherent groups. To this end, we used high-resolution maps to visualize the
340 spatial distributions of *Haloarchaea* and AOA sharing ecological preferences for deserts and
341 grasslands as well as for high or low salinity. Just like those atlases for preferences of dominant
342 soil microbial taxa at continental and global scales (Delgado-Baquerizo *et al.*, 2018; Egidi *et al.*,
343 2019; Jiao *et al.*, 2019), our results suggest that there are predictable clusters of co-occurring
344 dominant archaeal groups in dryland soils across northern China. As hypothesized, our continental
345 atlases and co-occurrence networks confirm the importance of ecosystem type and specialization
346 to specific environmental factors in determining the biogeographical patterns of *Haloarchaea* and
347 AOA in drylands.

348 Given the identities of *Haloarchaea* and AOA as specialists and generalists, respectively, the
349 distinct continental atlases also imply that their distribution patterns could be underpinned by
350 contrasting community assembly processes (Pandit *et al.*, 2009; Hanson *et al.*, 2012; Stegen *et al.*,
351 2013, 2015). Our results confirm this, and are consistent with earlier findings that habitat
352 specialists were more strongly structured by environmental factors, whereas the distribution of
353 habitat generalists was more influenced by spatial-based processes (Pandit *et al.*, 2009;
354 Langenheder and Székely, 2011; Logares *et al.*, 2013; Liao *et al.*, 2016; Malard *et al.*, 2019).
355 These results also confirm that *Haloarchaea* have strict requirements for specific environmental
356 conditions (e.g., high salinity and extreme climates) (Oren, 1994), while AOA have a wide
357 environmental tolerance and high resistance to environmental stresses and therefore could
358 colonize diverse ecological niches (Erguder *et al.*, 2009).

359 Moreover, we argue that the segregated distributions of *Haloarchaea* and AOA should be
360 taken into consideration when predicting changes in the functioning of drylands in a changing
361 world, given their strong links with multiple ecosystem functions including organic matter
362 decomposition, C fixation, N cycling and climate regulation (Chaban *et al.*, 2006; Offre *et al.*,
363 2013; Baker *et al.*, 2020; Hu *et al.*, 2021). For example, AOA may contribute to rapid nutrient
364 cycles and therefore to high plant productivity in grassland ecosystems (Wardle *et al.*, 2004; Offre
365 *et al.*, 2013; Trivedi *et al.*, 2019), whereas *Haloarchaea* may contribute to the slow decomposition
366 of recalcitrant organic matter and therefore to soil C sequestration in desert ecosystems (Oren,
367 1994; Wardle *et al.*, 2004; Chaban *et al.*, 2006; Auguet *et al.*, 2010).

368 The fact that *Haloarchaea* and AOA did not show niche differentiation across small-scale
369 vegetated and bare ground microsites is surprising, as the two microsites are typically considered
370 as contrasting niches that are reported to support different soil microbial communities as well as
371 the segregated distributions of AOA and AOB (Delgado-Baquerizo *et al.*, 2013a; Trivedi *et al.*,
372 2019). Dryland plant patches often can accumulate and capture soil resources under their canopies,
373 leading to the formation of the so-called fertile islands, which largely contribute to variation in the
374 abundance, diversity, and composition of soil microbial communities compared to adjacent bare
375 areas (Ochoa-Hueso *et al.*, 2018). However, we found no significant differences in soil salinity, N
376 availability, or other soil nutrients between the two microsites (see Fig. S6). Therefore, the lack of
377 clear differentiation between *Haloarchaea* and AOA across the two microsites can be partially
378 explained by the lack of the fertile island phenomenon and the relative homogeneity in soil
379 properties between the microsites studied here.

380 In conclusion, our results support the hypothesis of the ecosystem type-associated niche
381 differentiation for *Haloarchaea* and AOA, which are the two dominant topsoil archaeal groups
382 found in drylands across northern China. This indicates that the relative contributions of these
383 archaeal groups to ecosystem functioning are likely to vary systematically across China's drylands,
384 and potentially beyond. This understanding of the habitat preferences of *Haloarchaea* and AOA is
385 critical for predicting the responses of their spatial distributions to ongoing global change and the
386 resulting ecological consequences. Given the more deterministic assembly of *Haloarchaea*
387 compared with that of AOA, we argue that environmental impacts on the biogeography and
388 functional contribution of *Haloarchaea* will be more predictable than for AOA.

389

390 **Experimental Procedures**

391 **Field survey and sampling**

392 Field data and soil samples were collected from 173 dryland sites located in five provinces of
393 northern China (Fig. S1). Vegetation survey and soil sampling were conducted between June and
394 September from 2015 to 2017 according to the well-established standardized protocols as
395 described by Hu *et al.* (2021) (see Supporting Experimental Procedures for more details). In brief,

396 6–21 topsoil samples (0–20 cm depth) per site were collected under the canopies of the dominant
397 plant species and in bare areas devoid of vegetation, and 1 155 samples were taken and analyzed
398 in total.

399 To test the niche-differentiation hypothesis proposed (Fig. S2), we classified all study sites
400 into deserts and grasslands, the two most representative ecosystem types occurring in drylands
401 across northern China, using China's vegetation atlas at a scale of 1:1 000 000 (Chinese Academy
402 of Sciences, 2001) combined with our field observation. Each site grouped into a certain
403 ecosystem type was considered as a replicate. Totally, we had 112 replicates collected in deserts
404 and 61 collected in grasslands (Fig. S1).

405

406 **Molecular analyses**

407 Topsoil archaeal diversity and community composition were assessed using Illumina-based
408 sequencing. Genomic DNA from each of topsoil samples ($N = 1\ 155$) was extracted from 0.5 g of
409 defrosted soils using the PowerSoil[®] DNA Isolation Kit following manufacturer's protocols, and
410 pooled at site level, ultimately resulting in 173 composite DNA samples under the canopies of
411 vegetation and in bare ground, respectively.

412 The primer pair ARC344F (5'-ACGGGGYGCAGCAGGCGCGA-3') and ARC915R
413 (5'-GTGCTCCCCGCCAATTCCT-3') was used to amplify the V3–V5 regions of the archaeal
414 16S rRNA gene. While amplicon sequencing can always introduce some biases, the primer pair
415 used here is a set of archaeal universal primers and to our best knowledge has not been reported to
416 be biased towards any archaeal taxa. We therefore infer that the higher relative abundance of
417 *Thaumarchaeota* than *Haloarchaea* observed in this study (Fig. 1a) is not expected to be the result
418 of a bias in the primers towards the *Thaumarchaeota*. PCR reaction system and thermal cycling
419 conditions are described in Supporting Experimental Procedures. Amplicons were gel-purified
420 using the AxyPrep DNA Gel Extraction Kit and paired-end sequenced (2×300 bp) on an Illumina
421 MiSeq platform at the Majorbio Bio-pharm Technology Co., Ltd. (Shanghai, China). Resulting
422 sequences were processed by using the QIIME package (Caporaso *et al.*, 2010) (see Supporting
423 Experimental Procedures for more details of the bioinformatic analyses). OTUs not assigned at
424 kingdom level or identified as non-archaeal organisms were removed and the OTU table was

425 subsequently rarefied to the lowest number of sequences (10 707) found within an individual
426 sample. Our resampled dataset included a total of 1 871 archaeal OTUs. Raw sequencing data of
427 the archaeal 16S rRNA gene have been deposited in the NCBI Sequence Read Archive under
428 BioProject accession number PRJNA608843.

429

430 **Soil physicochemical analyses**

431 All topsoil samples ($N = 1\ 155$) were subjected to physicochemical analyses of pH, electrical
432 conductivity, organic C, total N, total phosphorus (P), available P, ammonium and nitrate contents
433 using standard laboratory methods as described in Supporting Experimental Procedures. It should
434 be noted that topsoil samples were collected under the canopies of vegetation and in bare ground
435 in each site; thus, the site-level measurements of all the soil physicochemical properties analyzed
436 here were obtained by using a weighted average of the mean values observed in vegetated areas
437 and bare ground by their respective cover (Maestre *et al.*, 2012; Delgado-Baquerizo *et al.*, 2013b).

438

439 **Data acquisition from public databases**

440 For all the sites surveyed, we gathered 20 standardized climatic variables (Table S1) related to
441 temperature, precipitation and solar radiation from the Worldclim database (Fick and Hijmans,
442 2017). We obtained the aridity index and potential evapotranspiration from the Global Aridity
443 Index and Potential Evapotranspiration Climate database
444 (<https://doi.org/10.6084/m9.figshare.7504448.v3>). Note that we used 1 – aridity index as our
445 surrogate of aridity level in all cases. In addition, we used normalized difference vegetation index
446 (NDVI) as a metric for net aboveground primary productivity as explained in recent studies in
447 drylands (Delgado-Baquerizo *et al.*, 2016b, 2018). NDVI data were obtained from the moderate
448 resolution imaging spectroradiometer aboard NASA's Terra satellites
449 (<https://neo.sci.gsfc.nasa.gov/>). We used the average NDVI values during our sampling dates as a
450 proxy for net aboveground primary productivity at the site level as described by
451 Delgado-Baquerizo *et al.* (2016b, 2018). Finally, we obtained information on soil texture
452 (percentage of clay, silt and sand) from the SoilGrids system (<https://soilgrids.org/>).

453

454 **Pre-selection of multicollinearity free environmental variables**

455 After data collection, we maximized the number of study sites without missing environmental data,
456 ultimately generating a sub-dataset ($N = 133$) that included 82 desert sites and 51 grassland sites.
457 For this sub-dataset, we excluded those environmental variables that were strongly correlated with
458 each other (Pearson's $r > 0.8$) within either climatic or soil variables as they may cause
459 multicollinearity problems in subsequent analyses (Maestre *et al.*, 2012; Delgado-Baquerizo *et al.*,
460 2017). Based on these analyses, we retained 9 out of 22 climatic variables (Fig. S14) and 11 out of
461 12 soil variables (Fig. S15) that were not strongly correlated with the rest ($r < 0.8$) for our
462 statistical analyses.

463

464 **Phylogenetic analyses**

465 We constructed a phylogenetic tree for OTUs belonging to the *Haloarchaea* and *Thaumarchaeota*
466 to visualize their taxonomic affiliations (Fig. 1b). Representative sequences from haloarchaeal (N
467 = 863) and thaumarchaeotal ($N = 208$) OTUs were multiple aligned using the MEGA7 program
468 (Kumar *et al.*, 2016) and filtered to remove uninformative regions. An approximately
469 maximum-likelihood tree was then constructed with the FastTree 2 (Price *et al.*, 2010) based on
470 the generalized time-reversible model of nucleotide evolution and visualized with the interactive
471 tool Evolview (Subramanian *et al.*, 2019).

472

473 **Network analyses**

474 We conducted correlation network analyses to explore the co-occurrence pattern of haloarchaeal
475 and thaumarchaeotal OTUs (Fig. 1d). For co-occurrence network construction, we first calculated
476 all possible Spearman's rank correlations (ρ) among OTUs. We focused only on positive
477 correlations as they provide information on haloarchaeal and thaumarchaeotal OTUs that may
478 have similar ecological preferences (Delgado-Baquerizo *et al.*, 2018). We considered a
479 co-occurrence to be robust if the Spearman's correlation coefficient (ρ) was both > 0.65 and
480 statistically significant ($P < 0.00001$) (Delgado-Baquerizo *et al.*, 2018). The final network was
481 visualized with the interactive platform gephi (<https://gephi.org/>). We then tested whether

482 haloarchaeal and thaumarchaeotal OTUs tend to co-occur with others within the same taxa or not.
483 To achieve this, we fitted a power law model to determine the degree distributions of *Haloarchaea*
484 and *Thaumarchaeota* in the constructed co-occurrence network (Banerjee *et al.*, 2018) (Fig. S3).

485

486 **Statistical analyses**

487 We estimated niche breadth for each of haloarchaeal and thaumarchaeotal OTUs using the Levins'
488 index (Levins, 1968) and the tolerance index (Dolédec *et al.*, 2000). The Levins' index estimates
489 niche breadth in terms of the number of sites where an OTU was detected, while the tolerance
490 index estimates niche breadth in terms of the environmental (including plant species richness,
491 NDVI, and the multicollinearity free climatic and soil variables as determined above) tolerance of
492 OTUs. These indices were calculated using the *spaa* and *ade4* packages in R
493 (<http://cran.r-project.org/>), and their differences between haloarchaeal and thaumarchaeotal OTUs
494 were evaluated using the non-parametric Mann-Whitney U test (Figs. 2a,b, S4a). Low values of
495 the indices correspond to specialists and high values to generalists. Further, we used the
496 Mann-Whitney U test to compare the differences in both the relative abundance and richness of
497 *Haloarchaea* and *Thaumarchaeota* between deserts and grasslands (Figs. 2e–h, S4d–g). These
498 analyses were conducted for both the full dataset and the sub-dataset, except that the tolerance
499 index was calculated exclusively for the sub-dataset.

500 Next, we performed random forest models to identify the major statistically significant
501 predictors of the relative abundance and richness of *Haloarchaea* and *Thaumarchaeota* (Fig. S7).
502 Variables used for random forest models comprised the multicollinearity free environmental
503 variables (Figs. S14, S15), plant species richness, NDVI and ecosystem type (coded as a
504 categorical variable with two levels: 1 = deserts and 0 = grasslands). These analyses were carried
505 out using the R package *rfPermute*. Then, we conducted non-parametric partial Spearman's rank
506 correlations to explore the net associations of haloarchaeal and thaumarchaeotal attributes with
507 ecosystem type, climatic factors, plant attributes and soil properties after accounting for their
508 simultaneous correlations with space (latitude and longitude). We included exclusively in these
509 analyses those variables that were identified as main statistically significant predictors of
510 haloarchaeal and thaumarchaeotal attributes from random forest models (Fig. S7). Furthermore,

511 we also considered the partial correlations among haloarchaeal and thaumarchaeotal attributes.
512 Information on partial Spearman's correlations was visualized using the R package corrplot (Fig.
513 3). To facilitate the interpretation of results for partial correlation analyses, we used the
514 Mann-Whitney U test to assess the differences in climatic factors, plant attributes and soil
515 properties between deserts and grasslands (Figs. S8, S9).

516 To further evaluate our ecosystem type-associated niche-differentiation hypothesis (Fig. S2),
517 we conducted Spearman's correlations ($P < 0.05$) to group haloarchaeal and thaumarchaeotal
518 OTUs into two ecological clusters with shared habitat preferences for deserts and grasslands.
519 Similarly, given the key role of soil salinity in determining the niche differentiation (Figs. 3, S7,
520 S8), we identified another two ecological clusters for high salinity and low salinity. We then
521 calculated the standardized (z -score) relative abundance of each of the four ecological clusters per
522 sample. We further used the Cubist regression models to predict the continental-scale distributions
523 of relative abundance of the four ecological clusters in drylands across northern China as
524 described previously (Delgado-Baquerizo *et al.*, 2018) (see also Supporting Experimental
525 Procedures for more details). Predictions on the continental-scale distributions of the four
526 ecological clusters were done on an approximate 10 km resolution grid (Figs. 4a,b, 5a,b). The
527 accuracy of our predictions (R^2) was evaluated using a training subset (75%) and a test subset
528 (25%) of the data. These analyses were performed using the R package Cubist, and all maps were
529 visualized in ArcGIS 10.2. We also constructed correlation networks to elucidate the
530 co-occurrence pattern of haloarchaeal and thaumarchaeotal OTUs sharing ecological preferences
531 (Figs. 4c,d, 5c,d) and evaluated whether those OTUs within a particular ecological cluster tend to
532 co-occur more often (Fig. S10), as described above. These analyses were performed exclusively
533 for the sub-dataset.

534 Finally, we inferred the relative influences of major environmental-based and spatial-based
535 processes on the community assembly of *Haloarchaea* and *Thaumarchaeota* using a
536 null-modeling-based quantitative framework (Stegen *et al.*, 2012, 2013, 2015) (Fig. 6a; see also
537 Supporting Experimental Procedures for detailed principles and calculations). The relative
538 influences of these processes were estimated separately for *Haloarchaea* and *Thaumarchaeota*
539 (Fig. 6b,c). Before analyses, the OTU tables were rarefied to 500 and 5 000 sequences per sample

540 for *Haloarchaea* and *Thaumarchaeota*, respectively.

541 We must note, however, that the vegetated and bare ground microsites typically found in
542 drylands may represent distinct niches and therefore harbor significantly different soil microbial
543 communities (Delgado-Baquerizo *et al.*, 2013a; Trivedi *et al.*, 2019). Thus, merging these two
544 small-scale contrasting microsites could affect the large-scale patterns of the ecosystem
545 type-associated niche differentiation of *Haloarchaea* and *Thaumarchaeota*. To address this, we
546 also evaluated the niche breadth, ecological distributions, and relative importance of those
547 community assembly processes of the two archaeal groups for vegetated areas and bare ground
548 separately. This set of analysis revealed no significant differences in all of these aspects between
549 the two microsites for both *Haloarchaea* and *Thaumarchaeota* (Figs. 2c,d,i-1, S5, S11).
550 Combining data across the two microsites should not, therefore, influence the conclusions of our
551 study.

552

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571

572 **Conflict of Interest**

573 The authors declare no conflict of interest.

574

575 **Data Availability Statement**

576 The data that support the findings of this study are available from the corresponding author upon
577 reasonable request.

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791 **Figure legends**

792 **Fig. 1 Taxonomic composition of topsoil archaea in drylands across northern China. a**

793 Percentage of OTUs and relative abundance of 16S rRNA gene sequences representing high-rank
794 archaeal taxa. **b** Phylogenetic tree of haloarchaeal and thaumarchaeotal OTUs. Taxonomic
795 affiliations of *Haloarchaea* and *Thaumarchaeota* are presented at the end of the branch (Ring 1).
796 Branches are colored by these two archaeal taxa (blue for *Haloarchaea* and red for
797 *Thaumarchaeota*). **c** Taxonomic composition of *Haloarchaea* and *Thaumarchaeota*.
798 *Nitrososphaerales* affiliated with the lineage Group I.1b accounts for > 99.9% of thaumarchaeotal
799 sequences. The remaining haloarchaeal and thaumarchaeotal OTUs assigned to the least abundant
800 taxa are not shown on bars (including *Nitrosocaldales* < 0.001%, *Nitrosopumilales* = 0.02%,
801 unidentified *Haloarchaea* < 0.001%, and unidentified *Thaumarchaeota* = 0.01%). **d**
802 Co-occurrence network with nodes (OTUs) colored by *Haloarchaea* and *Thaumarchaeota*. The
803 network diagram includes 1 071 nodes with 9 167 significant edges (ecological connections
804 among OTUs). The size of the nodes is proportional to the number of edges they contain.

805 **Fig. 2 Niche breadth of *Haloarchaea* and *Thaumarchaeota* and their ecological distributions**

806 **in deserts and grasslands as well as in bare ground and vegetated areas. a, b** Boxplots of the
807 niche breadth for each of haloarchaeal and thaumarchaeotal OTUs estimated using the Levins'
808 index and the tolerance index, respectively. **c, d** Relative abundance (**c**) and percentage of OTUs
809 (**d**) of 16S rRNA gene sequences representing the *Haloarchaea*, *Thaumarchaeota* and the
810 remaining archaea in deserts, grasslands, bare ground and vegetated areas, respectively. **e–l**
811 Boxplots of the relative abundance and richness of *Haloarchaea* and *Thaumarchaeota* in deserts
812 and grasslands (**e–h**) as well as in bare ground and vegetated areas (**i–l**). **a–l** These results are
813 shown for the full dataset, except that the tolerance index is calculated for the sub-dataset.
814 Significant differences between boxes are determined using the non-parametric Mann-Whitney U
815 test. Significance levels are: ns, not significant ($P > 0.05$); *** $P < 0.001$.

816 **Fig. 3 Partial Spearman's correlation matrix of the main determinants of ecological**

817 **distributions of *Haloarchaea* and *Thaumarchaeota*.** Partial Spearman's correlations of the
818 relative abundance and richness of *Haloarchaea* and *Thaumarchaeota* with ecosystem type,
819 climatic factors, plant attributes and soil properties after controlling for spatial influence (latitude

820 and longitude). The variables that are identified as significant predictors of haloarchaeal and
821 thaumarchaeotal attributes from random forest models are included in these analyses. The size and
822 color of the squares are proportional to partial correlation coefficients as shown in the scale bar.
823 Blue and red colors represent significant positive and negative correlations, respectively. na, not
824 applicable (non-significant predictors of haloarchaeal and thaumarchaeotal attributes from random
825 forest models); ns, not significant ($P > 0.05$); ATR, annual temperature range; MTWQ, mean
826 temperature of wettest quarter; PS, precipitation seasonality; PDM, precipitation of driest month;
827 PSR, plant species richness.

828 **Fig. 4 Predicted distributions and co-occurrence patterns of haloarchaeal and**
829 **thaumarchaeotal OTUs with shared habitat preferences. a, b** Predicted continental-scale
830 distributions of relative abundance of the two ecological clusters of haloarchaeal and
831 thaumarchaeotal OTUs sharing habitat preferences for deserts and grasslands in drylands across
832 northern China, respectively. R^2 denotes the percentage of variation explained by the models. The
833 scale bar depicts the standardized relative abundance (z -score) of each ecological cluster. The inset
834 shows relationship between ecosystem type and the z -score of each ecological cluster. **c, d**
835 Co-occurrence networks with nodes colored by the ecological clusters and taxa, respectively. The
836 size of the nodes is proportional to the number of edges they contain.

837 **Fig. 5 Predicted distributions and co-occurrence patterns of haloarchaeal and**
838 **thaumarchaeotal OTUs with shared salinity preferences. a, b** Predicted continental-scale
839 distributions of relative abundance of the two ecological clusters of haloarchaeal and
840 thaumarchaeotal OTUs sharing environmental preferences for high salinity and low salinity,
841 respectively. The inset shows relationship between the standardized relative abundance (z -score)
842 of each ecological cluster and soil salinity, indicated as log-transformed electrical conductivity. **c,**
843 **d** Co-occurrence networks with nodes colored by the salinity clusters and taxa, respectively. **a–d**
844 Other details as in Fig. 4.

845 **Fig. 6 Relative influences of ecological processes mediating the community assembly of**
846 ***Haloarchaea* and *Thaumarchaeota*. a** The null-modeling-based analytical framework for
847 estimating relative influences of multiple environmental-based and spatial-based processes
848 structuring microbial community assembly proposed by Stegen *et al.* (2012, 2013, 2015). β NTI,

849 β -nearest taxon index; RC_{bray} , Bray–Curtis-based Raup–Crick. **b, c** Relative contributions of the
850 five ecological processes governing community turnover of *Haloarchaea* and *Thaumarchaeota*,
851 respectively.