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## Projecting the impact of air pollution on child stunting in India – synergies and trade-offs between climate change mitigation, ambient air quality control, and clean cooking access

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# 1 Projecting the impact of air pollution on child stunting in 2 India – synergies and trade-offs between climate change 3 mitigation, ambient air quality control, and clean cooking 4 access

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## Abstract

Many children in India face the double burden of high exposure to ambient (AAP) and household air pollution (HAP), both of which can affect their linear growth. Although climate change mitigation is expected to decrease AAP, climate policies could increase the cost of clean cooking fuels. Here, we develop a static microsimulation model to project the air pollution-related burden of child stunting in India up to 2050 under four scenarios combining climate change mitigation (2°C target) with national policies for AAP control and subsidised access to clean cooking. We link data from a nationally representative household survey, satellite-based estimates of fine particulate matter (PM<sub>2.5</sub>), multi-dimensional demographic projection and PM<sub>2.5</sub> and clean cooking access projections from an integrated assessment model. We find that the positive effects on child linear growth from reductions in AAP under the 2°C Paris Agreement target could be fully offset by the negative effects of climate change mitigation through reduced clean cooking access. Targeted AAP control or subsidised access to clean cooking could shift this trade-off to result in net benefits of 2.8 (95% uncertainty interval [UI]: 1.4, 4.2) or 6.5 (UI: 6.3, 6.9) million cumulative prevented cases of child stunting between 2020-50 compared to business-as-usual. Implementation of integrated climate, air quality, and energy access interventions has a synergistic impact, reducing cumulative number of stunted children by 12.1 (UI: 10.7, 13.7) million compared to business-as-usual, with the largest health benefits experienced by the most disadvantaged children and geographic regions. Findings underscore the importance of complementing climate change mitigation efforts with targeted air quality and energy access policies to concurrently deliver on carbon mitigation, health and air pollution and energy poverty reduction goals in India.

**Key words:** stunting, air pollution, India, climate change mitigation, co-benefits, projection

Abbreviation	Definition
AAP	Ambient air pollution
CCA	Clean cooking access
GAINS	Greenhouse-Gas Air Pollution
	Interaction and Synergies
HAP	Household air pollution
HAZ	Height-for-age z score
IAM	Integrated assessment model
LPG	Liquefied petroleum gas
MAD	Minimum acceptable diet
MFR	Maximum feasible reduction
NFHS	National family health survey
NPi	National policy implementation
OR	Odds ratio
PM2.5	Fine particulate matter
UI	Uncertainty interval

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## 94 Introduction

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96 The 25 million children born annually in India are exposed to some of the highest  
97 levels of ambient air pollution (AAP) in the world, several-fold greater than current  
98 WHO guidelines. With 56% of households in the country relying on highly polluting  
99 solid fuels to meet household energy needs, many children bear the double health  
100 burden of both high AAP and household air pollution (HAP) (International Institute  
101 for Population Sciences (IIPS) and ICF 2017). Air pollution (both AAP and HAP) is  
102 currently recognised as the second leading risk factor for disease burden and mortality  
103 in India, surpassed only by malnutrition (IHME 2019). A recent India Disease Burden  
104 study attributed 4.8% (95% uncertainty interval [UI] 3.6-6.0) of all under-5 deaths in  
105 the country to AAP and 4.0% (95% UI 3.0-5.1) to HAP (India State-Level Disease  
106 Burden Initiative Child Mortality Collaborators 2020). Exposure to air pollutants in-  
107 utero and early in life can be especially detrimental for children's health because of  
108 their biological vulnerability and rapid development resulting in a range of adverse  
109 health outcomes (Perera 2017; Backes *et al* 2013). These include adverse birth  
110 outcomes, such as low birth weight and pre-term birth; respiratory diseases such as  
111 pneumonia, asthma and bronchitis; and impaired cognitive and neurological  
112 development. In addition to these well-established health outcomes, there is  
113 accumulating evidence that in-utero and early life exposure to AAP and HAP are also  
114 associated with child linear growth retardation (Zhu *et al* 2015; Yuan *et al* 2019; Bruce  
115 *et al* 2013; Pun *et al* 2021; Boamah-Kaali *et al* 2021).

116 Stunting, defined as being too short for one's age, is a largely irreversible linear  
117 growth impairment that can have severe long-lasting impacts on child health and  
118 human capital formation. In childhood, stunting is associated with poor cognitive  
119 development (Poveda *et al* 2021), higher risk of mortality, and susceptibility to  
120 infectious diseases such as pneumonia and diarrhoea. Later in life stunting can lead to  
121 lower productivity and earnings and increased risk of metabolic diseases (Prendergast  
122 and Humphrey 2014). Although the biological mechanisms underlying the effects of  
123 air pollution on stunting are yet to be fully understood, it is recognised that these start  
124 during the in-utero period. Particles or their components can reach beyond the lungs  
125 of pregnant women to induce systemic inflammation or oxidative stress, leading to  
126 poor foetal growth (Backes *et al* 2013). Postnatally, environmental exposure to air  
127 pollution may compound the adverse effects of poor nutrition and pathogens on  
128 immune development and function, resulting in a cycle of recurrent disease and  
129 malnutrition (Dewey and Mayers 2011). More specifically, recurrent respiratory  
130 infections caused by air pollution may lead to suppressed appetite, impaired  
131 absorption of nutrients, increased nutrient loss, and diversion of nutrients towards  
132 immune response and away from growth (Dewey and Mayers 2011). Several  
133 observational studies from India, where child undernutrition is among the highest in

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3 134 the world, reported consistent associations between early-life exposure to AAP (Singh  
4 135 *et al* 2019; Spears *et al* 2019) and use of polluting cooking fuels (Islam *et al* 2021;  
5 136 Fenske *et al* 2013; Tielsch *et al* 2009) and child stunting. The epidemiological  
6 137 evidence linking AAP and HAP with prenatal (small for gestational age) or postnatal  
7 138 (low height-for-age Z-score) stunting has been summarised by several meta-analyses  
8 139 (Zhu *et al* 2015; Yuan *et al* 2019; Bruce *et al* 2013; Pun *et al* 2021). According to the  
9 140 most recent pooled estimates a 10  $\mu\text{g}/\text{m}^3$  increase in ambient  $\text{PM}_{2.5}$  over the entire  
10 141 pregnancy increased the odds of prenatal stunting by 8% (95% CI: 3–13%), while  
11 142 postnatal exposure to solid versus clean cooking fuel increased the risk of postnatal  
12 143 stunting by 19% (95% CI: 10-29%)(Pun *et al* 2021).

13 144 Previous studies have shown that reductions in greenhouse gas emissions in line with  
14 145 climate change mitigation targets can bring substantial AAP improvements and health  
15 146 benefits in India, so-called co-benefits (West *et al* 2013; Silva *et al* 2016; Vandyck *et*  
16 147 *al* 2018 ; Markandya *et al* 2018; Sampedro *et al* 2020; Chowdhury *et al* 2018), and  
17 148 even more so when combined with stricter national measures for air quality control  
18 149 (Dimitrova *et al* 2021). However, scenario analysis from six different Integrated  
19 150 Assessment Models (IAMs), which quantified the interactions between climate  
20 151 change mitigation and energy access, suggest that stringent climate policy might  
21 152 significantly slow down the transition to clean cooking fuels by affecting energy  
22 153 prices (McCollum *et al* 2018). Thus, climate change mitigation may have opposing  
23 154 effects on the levels of AAP and HAP exposure and the associated health and  
24 155 developmental outcomes for future generations of children.  
25 156

26 157 Most assessments of air pollution health co-benefits from climate change mitigation  
27 158 to date have focused on mortality outcomes and on adult populations. These estimates  
28 159 obscure the full health burden from air pollution by not including morbidity impacts.  
29 160 The potential lifelong consequences for future generations of children are still very  
30 161 poorly reflected in the literature, even though these populations will bear a  
31 162 disproportionate amount of the disease burden from environmental change.  
32 163 Furthermore, existing health co-benefit projections are based on comparative risk  
33 164 assessment or life table methods, which do not allow for more detailed analysis of  
34 165 health inequalities across different socio-demographic groups and geographical areas.  
35 166 Lastly, the health co-benefits literature has largely focused on single exposure  
36 167 pathways and rarely considered concurrent effects of multiple exposures. A few  
37 168 microsimulation models analysing health outcomes under air pollution control have  
38 169 been developed for some high-income countries (Symonds *et al* 2019, Pimpin *et al*  
39 170 2018), but models focusing on health co-benefits from climate change mitigation and  
40 171 on Low and Middle-Income countries are lacking. We addressed these research gaps  
41 172 by investigating for the first time how the synergies and trade-offs between climate  
42 173 change mitigation, targeted ambient  $\text{PM}_{2.5}$  control and energy access support policies  
43 174 could affect future child linear growth in India. We employed a static microsimulation  
44 175 with a soft link to an IAM and a multi-dimensional demographic projection, which

176 allowed us to incorporate population-specific exposure response functions, explore  
177 differential impacts across population groups and geographical areas, and consider  
178 simultaneous effects of indoor and outdoor air pollution.

## 180 **Methods**

### 181 **Study design**

182 Our analysis proceeded in two stages. First, we examined the associations between  
183 early-life exposure to ambient PM<sub>2.5</sub> and polluting cooking fuel and stunting in a large  
184 dataset of children under-5 years in India. In the second stage, we developed a static  
185 microsimulation model of child stunting based on the following input data (1)  
186 National Family Health Survey (NFHS) data; (2) a multi-dimensional population  
187 projection; and (3) projections of ambient PM<sub>2.5</sub> concentrations, clean fuel use and  
188 per-capita income levels from an IAM (IIASA 2021). We projected the prevalence  
189 of child stunting at local level (district and urban/rural residence) and for distinct  
190 population groups under four scenarios combining climate change mitigation, air  
191 quality control, and policies to support clean cooking access (CCA). A detailed  
192 description of the data sources and methods is provided in the appendix.

### 193 **Stage one: Epidemiological analysis**

194 *Observed population data* We used nationally representative anthropometric and  
195 household data of children under-5 from India's 2015-16 NFHS (NFHS-4, also known  
196 as the 2015–16 India Demographic Health Survey (DHS)). NFHS is a nation-wide,  
197 multi-round, two-stage stratified survey conducted in a representative sample of  
198 women of reproductive age (International Institute for Population Sciences (IIPS) and  
199 ICF 2017). Using NFHS's child anthropometric data, we defined stunting as height-  
200 for-age z score (HAZ) two standard deviations below the median of the WHO Child  
201 Growth Standards.

202 **Baseline ambient PM<sub>2.5</sub> and clean cooking access data** We retrieved high resolution  
203 annual average PM<sub>2.5</sub> concentrations (0.01° x 0.01°) for the period 2009-2016 from  
204 the Atmospheric Composition Analysis Group (Hammer *et al* 2020). Each child was  
205 assigned average PM<sub>2.5</sub> exposure in-utero based on their date of birth, pregnancy  
206 duration and the geo-location of their household cluster.

207 As a proxy of exposure to HAP we used the type of primary cooking fuel of the  
208 households reported in the survey. We assumed households used the same fuel at the  
209 time of birth of the child as reported at the time of interview as previous studies have  
210 shown that cooking fuel transitions are relatively slow (Van Der Kroon *et al* 2013).  
211 We analysed the effect on child stunting of cooking with clean cooking fuels  
212 (electricity, liquefied petroleum gas (LPG), natural gas and biogas) compared to high-  
213 polluting fuels (kerosene, coal, charcoal, wood, straw, crop waste and dung).

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3 214 *Statistical analysis* We estimated the effect of PM<sub>2.5</sub> exposure in-utero and type of  
4 215 cooking fuel using logistic regression, with a random intercept for administrative  
5 216 district to account for clustering. Based on the literature, we identified and adjusted  
6 217 for the following confounders: age and sex of the child, age, education and caste of  
7 218 the mother, urban-rural residence and household income category (based on the  
8 219 household wealth index as shown in the next section).

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12 220 We included a penalized spline for child age in months and interaction terms between  
13 221 PM<sub>2.5</sub> in-utero with the child's sex, urban-rural residence, maternal education,  
14 222 household income category and caste, and interaction terms between clean fuel use  
15 223 with the child's sex and caste in order to account for differential vulnerabilities to air  
16 224 pollution across different socio-demographic groups. The analysis was performed  
17 225 with R (version 3.6.1), using the package *mgcv* (Wood 2011).

18  
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21 226 We performed a series of model specification checks by including a larger set of  
22 227 covariates in the model, adjusting for seasonality, estimating effects of life-course  
23 228 PM<sub>2.5</sub> exposure (i.e. in-utero and after birth). We also conducted a sensitivity analysis  
24 229 based on a subsample of the data to explore potential residual confounding by  
25 230 nutrition as operationalised by a Minimum Acceptable Diet (MAD) index (appendix  
26 231 1.1).

## 232 **Stage two: Projections**

233 233 *Scenarios* We developed four hypothetical pathways for India to deliver on the Paris  
234 234 Agreement target and compared them to a reference scenario (Table 1). "NPI  
235 235 (National Policy implementation) without access policy" specifies a business-as-usual  
236 236 pathway of global greenhouse gasemissions based on currently announced climate  
237 237 policies through 2030, current AAP legislation and no additional support for CCA.  
238 238 We explored four mitigation pathways, which assumed the implementation of a  
239 239 carbon price of US\$40 per ton CO<sub>2</sub> equivalent in the year 2020 that increased at the  
240 240 social discount rate through until the end of the century. These pathways were  
241 241 consistent with a >66% chance of limiting global mean temperature increases to 2°C  
242 242 relative to pre-industrial levels throughout the end of the century. The four mitigation  
243 243 pathways differ only with respect to the AAP control and compensatory energy access  
244 244 policies implemented at the national level. The 2°C scenarios assume compliance with  
245 245 current air pollution legislation only, while the 2°C MFR (Maximum Feasible  
246 246 Reduction) scenarios model implementation of additional end-of-pipe national air  
247 247 quality control measures in industrial, power generation, household, and agricultural  
248 248 sectors. The "no access" scenarios assume no counterbalancing price support policies  
249 249 on clean fuels and stoves, while the two "access" scenarios model a universal subsidy  
250 250 covering 15% of the cost of LPG cooking stoves and 75% of the cost of LPG fuel.

251 251 The AAP and CCA scenarios were developed independently in the MESSAGE-  
252 252 GLOBIOM global energy-economy IAM framework (IIASA 2021) based on the same  
253 253 national CO<sub>2</sub> budget constraints and projections of population growth, urbanisation



254 and various regionalised economic activities. The AAP projections were generated  
 255 within the Greenhouse-Gas Air Pollution Interaction and Synergies (GAINS) module,  
 256 while the clean access transitions were modelled within the Access household fuel-  
 257 choice module of MESSAGE-GLOBIOM. More details on the climate-energy  
 258 modelling and the linkages of the different modules can be found elsewhere (Cameron  
 259 *et al* 2016; Purohit *et al* 2019).

260

Scenario	Climate Change Mitigation	Ambient Air Pollution Control	Clean Cooking Access
<b>NPi without access policy</b>	National Policies for climate, energy, environment and development until 2030, no climate policy after 2030.	Current air pollution legislation	No additional clean cooking access support policy
<b>2° C without access policy</b>	National Policies until 2020, after which mitigation measures in line with a >66% chance of staying below 2°C throughout 21st century.		No additional clean cooking access support policy
<b>2° C with access policy</b>			15% LPG cooking stove & 75% LPG cost subsidies available to all households
<b>2° C MFR without access policy</b>		Maximum Feasible Reduction (MFR) of air pollution	No additional clean cooking access support policy
<b>2° C MFR with access policy</b>			15% LPG cooking stove & 75% LPG cost subsidies available to all households

261 **Table 1: Scenarios description**

262 NPi – National Policy Implementation, MFR – Maximum Feasible Reduction

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264 *Static microsimulation* For each year and scenario, we generated datasets with  
 265 individuals with identical characteristics to those in the stage one dataset. We applied  
 266 a reweighting procedure to reproduce the changes in the demographic characteristics  
 267 of children under-5 over time (age, sex, subnational state, urban/rural place of

268 residence, maternal education) as forecasted by a multi-dimensional demographic  
269 projection for India (Samir *et al* 2018). The population projection assumes a  
270 continuation of past demographic trends, leading to a decline in fertility and in child  
271 mortality, improvement in educational attainment and increase in urbanisation (Table  
272 S2). In each simulated dataset we altered the individual PM<sub>2.5</sub> exposure during  
273 pregnancy, the income category and the household primary cooking fuel based on  
274 projections from the IAM, keeping other covariates fixed.

275 Under each scenario for the period 2010-2050, gridded annual mean PM<sub>2.5</sub>  
276 concentrations from the GAINS model were matched with the simulated datasets  
277 based on the geographic coordinates of NFHS-4 clusters.

278 Data on changes in income levels and uptake of clean cooking fuels from the  
279 MESSAGE-Access household fuel-choice model were available for the whole of  
280 India and for four socio-economic groups based on rural-urban residence and daily  
281 per-capita expenditure threshold (Purchasing Power Parity of \$2 per day in rural and  
282 \$5 per day in urban areas) (Cameron *et al* 2016). We translated aggregate level  
283 projections into individual cooking fuel choices based on several assumptions. First,  
284 we assumed the same rate of change in income and uptake of cleaner cooking fuels  
285 for all regions. Second, for each future year and scenario we generated an indicator of  
286 household income level based on the household wealth index from NFHS-4 and the  
287 projected population distribution in each income category from the IAM (appendix  
288 1.2). Third, we ranked fuel preferences following the theory of the “energy ladder”  
289 and assumed that as households’ economic status improves they tend to gradually shift  
290 to cleaner fuels (Van Der Kroon *et al* 2013) (appendix 1.2). To account for the  
291 importance of socio-demographic factors in determining household fuel choice, we  
292 conditioned transition to clean cooking on maternal educational level. We used the  
293 regression model specified in the epidemiologic analysis (Stage one) to predict the  
294 probability of stunting under the specified scenarios for each individual in the dataset.  
295 The adjusted sampling weights were then applied to estimate the stunting prevalence  
296 in the population under each scenario.

297 We performed posterior simulations to derive 95% uncertainty intervals (UIs)  
298 (appendix 1.3). Lack of confidence bounds in the projections of ambient PM<sub>2.5</sub>, access  
299 to clean cooking fuels, income, and population change limited our ability to  
300 incorporate these uncertainties in our final estimates. We performed a sensitivity  
301 analysis by re-running the simulations after calibration of modelled PM<sub>2.5</sub>  
302 concentrations in GAINS with those from the Atmospheric Composition Analysis  
303 Group (appendix 1.4).

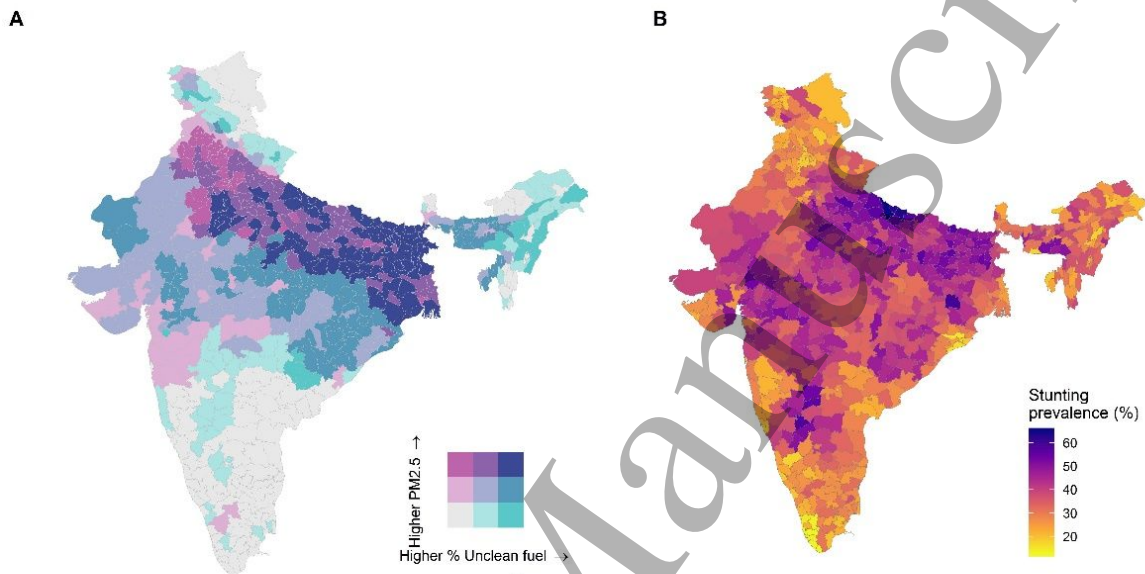
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## 305 **Results**

### 306 **Epidemiological analysis**

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3 307 We included 203,870 children from the NFHS-4 in our final sample, after removing  
4 308 missing observations, children that died or changed location since birth. Summary  
5 309 statistics for the exposure variables and other covariates by stunting status are  
6 310 presented in Table S1. Children were on average exposed to 73.6  $\mu\text{g}/\text{m}^3$   $\text{PM}_{2.5}$  in-  
7 311 utero, while 67% of them lived in households without CCA. There were large regional  
8 312 variations in ambient and household air pollution exposure as well as in stunting  
9 313 prevalence (Figure 1).  
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316 **Figure 1: Map of India showing (a) children's district-level mean ambient  $\text{PM}_{2.5}$**   
317 **exposure in-utero ( $\mu\text{g}/\text{m}^3$ ) and households using polluting cooking fuel (%) in 2015,**  
318 **and (b) district-level prevalence of stunting among children under-5 in 2015.**

319 All values are weighted using sampling weights of NFHS-4.

320

321 After adjustment for confounders, in-utero exposure to ambient  $\text{PM}_{2.5}$  significantly  
322 increased the odds of child stunting (OR: 1.04, 95% CI: 1.03-1.05 per 10  $\mu\text{g}/\text{m}^3$   
323 increase in  $\text{PM}_{2.5}$ ), while clean compared to polluting cooking fuel decreased the odds  
324 of stunting (OR: 0.81, 95% CI: 0.79-0.84) (Table S3). We observed modification of  
325 the effect of in-utero  $\text{PM}_{2.5}$  exposure on stunting by sex, residence (urban/rural),  
326 maternal education, caste, and household income (p interaction term < 0.05). The  
327 effect of CCA was modified by sex and caste (p interaction term < 0.05). In particular,  
328 female children, those living in urban areas, born to less educated mothers, belonging  
329 to more disadvantaged castes and to lower income households were more susceptible  
330 to the harmful effects of  $\text{PM}_{2.5}$  on linear growth. Conversely, the beneficial effects of  
331 CCA on child stunting were more pronounced for children who were female and did  
332 not belong to socially disadvantaged castes. Similar to Spears *et al* (2019) we found  
333 no evidence of a non-linear association between  $\text{PM}_{2.5}$  in-utero exposure and child

stunting. Adjusting for additional covariates including month of birth to account for seasonal variation in exposures had minimal effect on the exposure effect estimates (Figure S1). Adjusting for MAD among children aged 6-23 months did not change observed associations between ambient PM<sub>2.5</sub> in-utero and clean cooking fuel use with child stunting (Table S4). While MAD was associated with lower odds of child stunting in models without adjustment for socio-economic status (Table S4, model 3), the association was fully attenuated after adjusting for urban residence and socio-economic status (Table S4, model 2), suggesting nutritional status is not an important confounder in models with already adjusted for socio-economic status and urban residence. As we used annual PM<sub>2.5</sub> data in the analysis, we could not test the effect of PM<sub>2.5</sub> exposure in different trimester periods on child stunting. In-utero exposure to ambient PM<sub>2.5</sub> was more strongly associated with child linear growth than life-course exposure (in-utero and after birth) (Figure S1).

### Projections of impacts on stunting

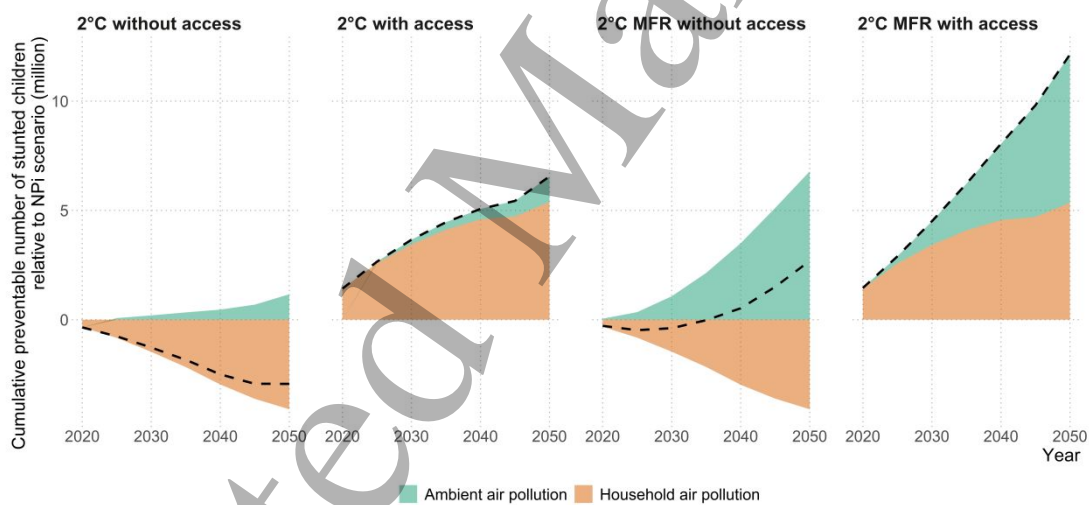
Projected in-utero ambient PM<sub>2.5</sub> exposure and the share of population with CCA by residence and year are shown in Table 2. Under most scenarios average in-utero PM<sub>2.5</sub> exposure was projected to decrease and CCA to increase over time relative to the baseline year both for rural and urban areas. The largest reductions in ambient PM<sub>2.5</sub> were observed in scenarios where climate change mitigation was accompanied by end-of-pipe AAP controls, while population access to clean cooking was maximised in scenarios with additional access support policies. The projected characteristics of children under-5 were identical across all scenarios (Table S2).

Scenario	Year	Average in-utero ambient PM <sub>2.5</sub> (µg/m <sup>3</sup> )		Share of children living in households with CCA (%)	
		Rural	Urban	Rural	Urban
	2015	45	58	17	73
NPI without access policy	2030	50	61	53	90
	2050	57	73	65	95
2° C without access policy	2030	48	59	36	80
	2050	49	60	49	90
2° C with access policy	2030	48	59	77	96
	2050	49	60	90	97
2° C MFR without access policy	2030	39	48	36	80
	2050	22	30	49	90
2° C MFR with access policy	2030	39	48	77	96
	2050	22	30	90	97

**Table 2: Baseline and projected exposure variables according to scenario and year**  
 The 2015 values for CCA are calculated based on NFHS-4 data, applying sampling weights. Ambient PM<sub>2.5</sub> concentrations and future CCA are based on the GAINS and MESSAGE-Access modelled data, respectively, applying adjusted sampling weights to account for changes in demographics, urbanisation and maternal education over time.

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362 Figure 2 and Table S5 show the cumulative (2020-50) preventable number of stunted  
 363 children over time under each intervention scenario compared to NPi and  
 364 disaggregated by the contribution of changes in AAP and HAP. In the 2°C scenario  
 365 without access policy, the increase in child stunting from higher HAP (+ 4 million) is  
 366 larger than the reduction in the burden from AAP (-1.2 million), leading to an overall  
 367 higher cumulative number of stunted children compared to NPi (2.9 million, UI: 2.8,  
 368 3.0). However, accompanying the 2°C mitigation efforts with additional AAP control  
 369 or CCA support is projected to reduce the overall burden of child stunting from air  
 370 pollution compared to NPi. Implementation of national policies for maximum feasible  
 371 reduction of AAP can help prevent 2.8 (UI: 1.4, 4.2) million cases of child stunting  
 372 between 2020-50, while compensatory subsidies for LPG cooking fuel and stoves can  
 373 avert growth faltering in 6.5 (UI: 6.3, 6.9) million children. The joint implementation  
 374 of the two policies along with mitigation efforts had synergistic effects for child  
 375 growth, i.e. yielded greater health benefits than the sum of health benefits from  
 376 individual implementation, and prevented linear growth impairment in 12.1 (UI: 10.7,  
 377 13.7) million children compared to NPi. Sensitivity analysis with calibrated ambient  
 378 PM<sub>2.5</sub> data did not notably affect the final results (Table S6 and S12, S13).

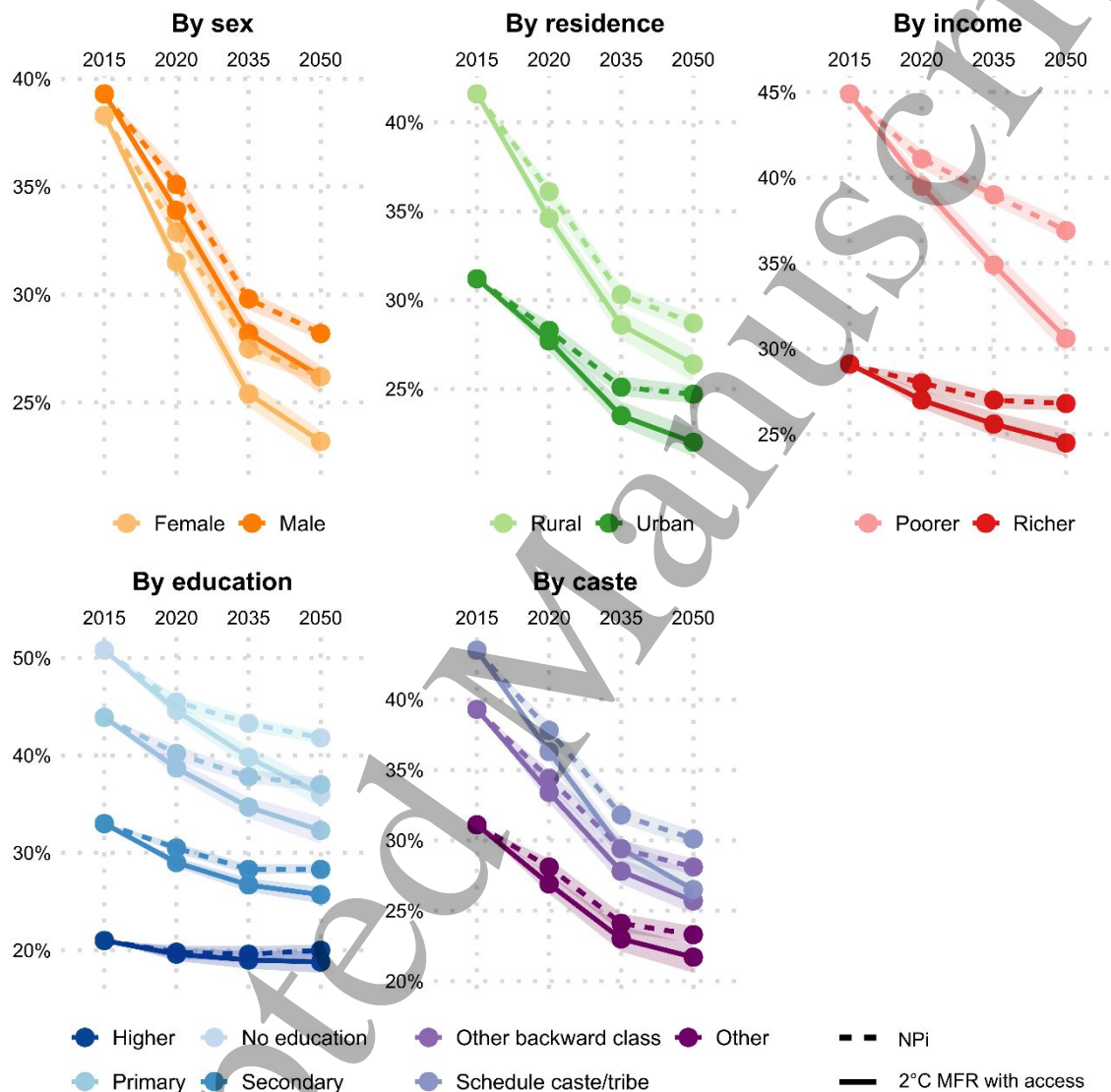


379

380 **Figure 2: Cumulative preventable number of stunted children (in million) from**  
 381 **changes in household air pollution (orange), ambient air pollution (green) and**  
 382 **household and ambient air pollution combined (dashed black line) according to**  
 383 **mitigation scenario and year relative to NPi scenario.**

384 The benefits of the most aspirational scenario (2°C MFR with access policy)  
 385 compared to NPi differed by population groups (Figure 3 and Tables S7-S11). While  
 386 all children benefited from improvements in indoor and outdoor air quality under the  
 387 2°C MFR with access policy scenario compared to NPi, child linear growth improved  
 388 the most among more disadvantaged groups with the highest prevalence of stunting  
 389 in 2015. Larger difference in the prevalence of child stunting in 2050 between the 2°C

390 MFR with access policy and the NPi were estimated for children living in poorer  
 391 households (– 6.3% compared to – 2.3% for richer households), belonging to a  
 392 scheduled caste or tribe (– 3.6% compared to – 1.6% for those from other castes) or  
 393 having an uneducated mother (– 5.8% compared to – 2.2% for those with highest  
 394 maternal education). The benefits of the 2°C MFR with access policy scenario in 2050  
 395 were similar for both sexes and for urban and rural residents, thus only marginally  
 396 reducing existing disparities in child stunting among these groups.

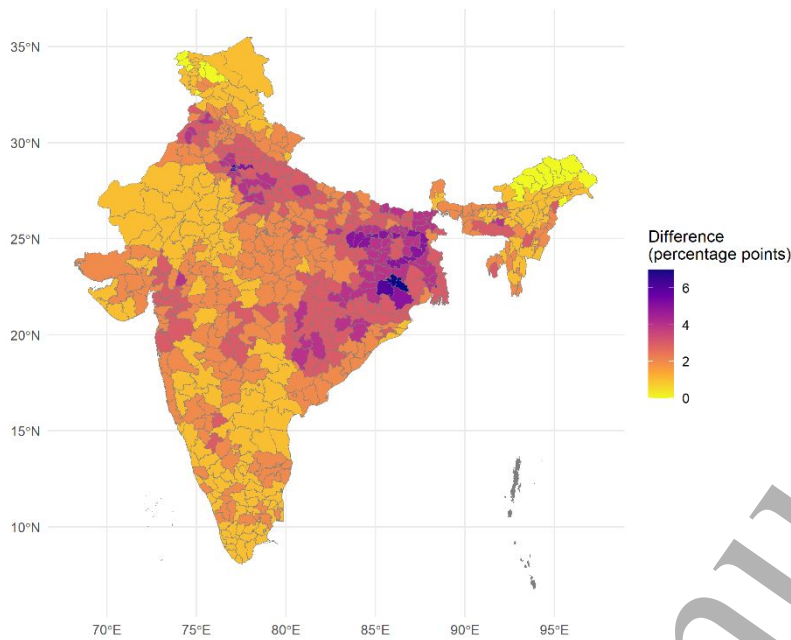


397

398 **Figure 3: Projected trends in stunting prevalence (children under-5) by population**  
 399 **sub-group under NPi and 2°C MFR with access policy scenarios**

400 Similarly, implementation of the 2°C MFR with access policy scenario was projected  
 401 to reduce stunting prevalence in the districts with the highest burden of child stunting  
 402 in 2015, especially in North-eastern India and around the Indo-Gangetic Plain (Figure  
 403 4). In 2050, largest reductions in the prevalence of child stunting were recorded in the  
 404 Purbi Singhbhum and Saraikela Kharsawan districts in Jharkhand (– 6%) and in

405 districts within the National Capital Territory of Delhi (– 7%), almost three times  
 406 higher than the India average (– 2.5%) (Figure 4).



407

408 **Figure 4: Percent difference in projected prevalence of child stunting in 2050**  
 409 **between the 2°C MFR with access policy and NPI scenarios according to**  
 410 **administrative district.**

#### 411 Discussion

412 We used a static microsimulation model to assess the potential impacts of changes in  
 413 AAP and HAP on child linear growth impairment in India under four policy scenarios  
 414 for delivering on the Paris Agreement climate change mitigation target. Our analysis  
 415 resulted in several key findings. First, the slower transition to clean cooking fuels  
 416 under climate change mitigation could fully cancel out projected benefits for child  
 417 linear growth due to reduced AAP without additional policies. Second, net benefits  
 418 for health could still occur if stringent climate policy were complemented by well-  
 419 designed national end-of-pipe air quality control or universal clean cooking access  
 420 subsidies. These policies could prevent stunting in 2.8 (UI: 1.4, 4.2) million and 6.5  
 421 (UI: 6.3, 6.9) million children between 2020-2050, respectively, compared to  
 422 business-as-usual. Third, optimal results for child growth can be achieved when  
 423 mitigation action is combined with both complementary policies (stunting avoided in  
 424 12.2 (UI: 10.7, 13.7) million children). This policy pathway could also provide an  
 425 opportunity to reduce inequalities in health and human capital early in life by  
 426 benefiting the most underprivileged children – those with lowest household income,  
 427 maternal education and social status. We estimated that implementation of integrated  
 428 climate, air quality and energy access policies would help reduce stunting in locations  
 429 where it is currently most prevalent (e.g. Indo-Gangetic Plain, north-east). Due to the  
 430 high concentrations of ambient PM<sub>2.5</sub> and high levels of poverty and reliance on

431 polluting cooking fuels, children in these regions would particularly benefit from the  
432 combined ambient air pollution controls and clean cooking access policies.

433 We used a novel health impact modelling approach, which allowed for an in-depth  
434 assessment of complex population-environment dynamics and multiple exposure  
435 pathways on human health not captured by comparative risk assessment methods.  
436 Compared to other modelling approaches that allow for comprehensive evaluation of  
437 the distributional effects of policies including dynamic microsimulations and agent-  
438 based models, a particular advantage of static microsimulation is the more modest  
439 modelling and computational requirements. We identified a number of socio-  
440 economic effect modifiers for the two exposure variables in the first stage of the  
441 analysis. The static microsimulation approach allowed us to reflect these heterogenous  
442 individual effects in the health impact assessment without the full computational  
443 burden of a dynamic microsimulation. By using a re-weighting procedure, we  
444 accounted for changes in many important socio-demographic characteristics of the  
445 population – age, sex, urban residence, region and maternal education – without  
446 having to perform a multidimensional demographic projection. The combination of  
447 static microsimulation with integrated assessment models and demographic  
448 projections offers a flexible and efficient approach for meeting the policy demand for  
449 projections that assess long-term health impacts and differential population  
450 vulnerabilities related to climate change.

451 Our findings, which differentiate impacts across multiple population subgroups and  
452 regions, could inform more targeted national- or local-level efforts to improve air  
453 quality and clean cooking access. However, policy makers may also have other policy  
454 tools at their disposal for increasing clean energy access apart from universal LPG  
455 price support, such as microfinance or clean cooking subsidies targeting only the most  
456 vulnerable. The health gains of modelled policies will depend on effective  
457 enforcement and overcoming legal, financial, social, behavioural and other barriers to  
458 their sustained implementation (Peng *et al* 2020, Malakar *et al* 2018, Sharma *et al*  
459 2020). The Indian Government provides subsidies for LPG consumption and  
460 connection through the Pratyaksh Hanstantrit Labh Yojana and Pradhan Mantri  
461 Ujjwala Yojana programmes. Despite the success of the two programmes in rapidly  
462 increasing LPG adoption, they have been less effective in ensuring sustained use,  
463 especially among low-income rural households (Kar *et al* 2019, Sharma *et al* 2021).  
464 Poor and socially marginalised households face major obstacles in accessing LPG  
465 subsidies as a result of their informal living situation, precarious income, limited  
466 access to information and physical and social isolation (Neto-Bradley *et al* 2021,  
467 Saxena and Bhattacharya 2018). Therefore, designing government support with  
468 consideration to the specific needs and constraints of different types of households  
469 will be important for realising the equity benefits of the modelled interventions.

470 Our analysis has a number of limitations. First, although the ambient PM<sub>2.5</sub> and the  
471 clean cooking access projections in our model were developed within the same IAM,



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3 472 they were not fully integrated. The effect of clean energy uptake on ambient  $PM_{2.5}$   
4 473 exposure was not considered, leading to a possible underestimation in the reductions  
5 474 in ambient  $PM_{2.5}$ . Chowdhury *et al* (2019) showed that complete mitigation of  
6 475 biomass emissions from cooking in 2015 would have reduced ambient  $PM_{2.5}$   
7 476 concentrations in India by 17.5%. The likely underestimation in our analysis would  
8 477 be smaller since the difference in clean cooking access in our mitigation scenarios  
9 478 with and without access in 2050 was 29 % rather than 100 %. Ambient  $PM_{2.5}$   
10 479 reductions from end-of-pipe air quality control on indoor air quality were not reflected  
11 480 in the MFR scenarios since we used CCA as a proxy of indoor air pollution exposure.  
12 481 The adoption of more efficient biomass cookstoves modelled in the MFR scenarios  
13 482 was also not implemented as NFHS does not include data on type of cooking stove.  
14 483 However, this effect is likely to be small as improved biomass cookstoves have  
15 484 resulted in minimal health benefits (Sambandam *et al* 2015). Second, we did not  
16 485 explicitly model fuel stacking due to lack of data on use of multiple fuels in NFHS-4.  
17 486 Fuel stacking is a well-documented behavioural response to volatile fuel supplies and  
18 487 prices, household incomes, or a result of cultural preferences (Van Der Kroon *et al*  
19 488 2013). Accounting for fuel stacking would likely lead to somewhat smaller estimated  
20 489 benefits of CCA policies on child stunting given that some households might not use  
21 490 clean fuels exclusively. Third, projected trends in average daily per-capita income and  
22 491 clean fuel use were available only at aggregate level from the IAM. Differences in  
23 492 trends in average per-capita income and CCA across states in our model thus only  
24 493 reflect disparities in 2015. As higher resolution energy, population and income  
25 494 projections from IAMs and demographic models become available in the future, more  
26 495 refined geographical variations in health impacts could be assessed. Fourth, our  
27 496 scenarios did not explicitly consider expansion of electrification. Our analysis focused  
28 497 on LPG intervention scenarios due to the dominance of LPG in national policy plans  
29 498 for expansion of clean cooking access in India and because historically electricity has  
30 499 rarely been used for cooking purposes in South Asia. While this could change in the  
31 500 future, in the short to medium term, electricity is unlikely to become a dominant means  
32 501 of meeting cooking energy needs. While we do not explicitly model electricity access,  
33 502 the effect on indoor air quality from cooking with electricity or LPG is assumed to be  
34 503 identical. Fifth, our sensitivity analysis did not indicate that nutrition (measured as  
35 504 MAD) was a confounder of the AAP or HAP stunting associations conditional on  
36 505 household income and other socioeconomic variables. However, MAD was based  
37 506 only on feeding practices on the day or evening preceding the survey and may not  
38 507 fully reflect longer-term nutritional status and we cannot rule out possible residual  
39 508 confounding. Finally, the population, energy and income projections in our model do  
40 509 not reflect the catastrophic effects that COVID-19 has had on population health, the  
41 510 economy, and clean energy access. Although the full impacts of the crisis are still to  
42 511 be fully evaluated, research suggests that the pandemic might slow down the transition  
43 512 to clean cooking fuels and other development objectives (Ravindra *et al* 2021;

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3 513 Pachauri *et al* 2021) and affect global investments in emission reductions (Reilly *et al*  
4 514 2021).

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6 515 Future extensions of this modelling approach could focus on incorporating dynamic  
7 516 feedback effects and behavioural responses such as the effects of air pollution on child  
8 517 survival over time or the influence of child stunting on educational attainment and  
9 518 adult survival later in life. In addition, an extension of this analysis could evaluate the  
10 519 balance of costs between scenarios. Both the end-of-pipe air quality measures and the  
11 520 CCA subsidies presented here entail additional policy costs besides mitigation  
12 521 finance. However, previous research has shown that avoided premature mortality  
13 522 through climate change mitigation or MFR of ambient air pollution in India will  
14 523 considerably outweigh the potential implementation costs (Sanderson *et al* 2013;  
15 524 Markandya *et al* 2018). Additional finance to cover subsidies for universal CCA could  
16 525 be mobilised through effort-sharing international climate regimes (Cameron *et al*  
17 526 2016). The anticipated improvements in child linear growth, both through air pollution  
18 527 co-benefits and through avoided impacts from climate change via income and food  
19 528 prices (Lloyd *et al* 2018), represent a human capital investment, which is likely to  
20 529 bring substantial savings through higher productivity, reduced morbidity, work  
21 530 absenteeism and associated health care costs. Future studies should consider other  
22 531 pathways through which climate change can impact child health in India. The adverse  
23 532 effects associated with increases in rainfall, heat stress and extreme weather events  
24 533 (floods/draughts) on child linear growth have been well documented (Dimitrova and  
25 534 Muttarak 2020, Phalkey *et al* 2015, Cooper *et al* 2019, Tusting *et al* 2020, Muttarak  
26 535 and Dimitrova 2019, Baker 2020, Belesova *et al* 2019). Some of these effects are  
27 536 mediated by altered patterns of water- and vector-borne infections, quality and  
28 537 quantity of crops, food prices and household income (Myers *et al* 2017, Phalkey *et al*  
29 538 2015). Considering these multiple causal pathways, stringent climate change  
30 539 mitigation is likely to bring much larger benefits to child linear growth than those  
31 540 quantified in this study. Previous studies have also reported a strong social gradient  
32 541 for some of these effects, with the poorest and least educated being most affected  
33 542 (Dimitrova and Muttarak 2020). More detailed consideration of the timing, magnitude  
34 543 and equity implications of the multiple impacts of climate change on child health,  
35 544 apart from energy access and outdoor air quality, would be important for a more  
36 545 accurate comparison of the trade-offs and benefits of mitigation action in India.

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