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Oksuz, Duygu P.

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1 **Bird taxonomic and functional responses to land abandonment in**
2 **wood-pastures**

3

4 **Authors' names:**

5 Duygu P. Oksuz^{a,*}, Jorge M. Palmeirim^a, Ricardo A. Correia^{b,c,d}

6

7 **Affiliations:**

8 ^acE3c–Centre for Ecology, Evolution and Environmental Changes, Departamento de Biologia
9 Animal, Faculdade de Ciências, Universidade de Lisboa, Lisboa, Portugal

10

11 ^bDBIO & CESAM-Centre for Environmental and Marine Studies, University of Aveiro, Campus
12 Universitário de Santiago, Aveiro, Portugal

13

14 ^cHelsinki Lab of Interdisciplinary Conservation Science (HELICS), Department of Geosciences and
15 Geography, University of Helsinki, Helsinki, Finland

16

17 ^dHelsinki Institute for Sustainability Science (HELSUS), University of Helsinki, Helsinki, Finland

18

19 ***Corresponding author:**

20 Duygu P. Oksuz

21 ^cE3c–Centre for Ecology, Evolution and Environmental Changes, Departamento de Biologia
22 Animal, Faculdade de Ciências, Universidade de Lisboa, Lisboa, Portugal.

23 dpoksuz@fc.ul.pt +351915665912

24 **ORCID ID of authors:**

25 **Duygu P. Oksuz**

26 ORCID ID: 0000-0002-4782-9815

27 **Jorge M. Palmeirim**

28 ORCID ID: 0000-0003-4734-8162

29 **Ricardo A. Correia**

30 ORCID ID: 0000-0001-7359-9091

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46 **Abstract**

47 Wood-pastures are socio-ecological systems covering vast areas in Europe. Although used for
48 grazing and production of various forest goods, wood-pastures harbour a rich biodiversity and are
49 usually considered as High Nature Value Farmlands. However, socio-economic pressures are
50 driving the transformation of these valuable landscapes from multi-functional, heterogeneous
51 habitats to homogeneous areas through either intensification or land abandonment. We investigated
52 how changes in management intensity influence the taxonomic diversity, functional diversity and
53 functional composition of birds in these landscapes using generalized linear models. In contrast to
54 taxonomic diversity, functional diversity decreased significantly towards shrub-dominated and less
55 heterogeneous areas related to the abandonment of grazing and/or understory management
56 practices. Grassland and generalist species, and associated guilds such as granivores, ground-nesters
57 and ground-foragers are almost absent less managed areas. On the other hand, shrub-dominated
58 areas favour forest species, particularly understory/canopy foragers and arboreal nesters, although
59 the forest guild is still well-represented in actively managed, heterogeneous areas. Our results
60 indicate the abandonment of wood-pasture management affects the prevalence of grassland and
61 generalist species, leading to functional diversity loss and potentially reduced ecosystem
62 functioning. We suggest non-intensive, active management is needed to maintain habitat
63 heterogeneity and canopy openness, enhancing trait diversity in wood-pastures.

64 **Keywords:** functional diversity; breeding birds; wood-pasture; management intensity;
65 montado/dehesa; socio-ecological systems

66

67 **1. Introduction**

68 Wood-pastures are social-ecological landscapes that have been shaped by various land-use
69 regimes prevailing in much of Europe (Hartel and Plieninger 2014). They usually have high
70 economic value provided by a multi-functional management that may include livestock grazing,

71 cork extraction, timber production, crop cultivation and tree pruning for firewood and charcoal
72 (Moreno et al. 2018). Many wood-pastures conciliate economic value with the maintenance of rich
73 biodiversity, and as such are considered High Nature Value Farmlands (Pinto-Correia and Ribeiro
74 2012). This richness is particularly evident in wood-pasture landscapes which are spatially
75 heterogeneous due to the availability of habitats such as riparian galleries, hedgerows, shrubby
76 patches and olive orchards (Leal et al. 2016; Oksuz et al. 2020). The maintenance of the multiple
77 values of wood-pastures depends on a balanced management of the landscape (Plieninger et al.
78 2015). However, recent social and economic changes in Europe are transforming traditional wood-
79 pasture management through intensification or abandonment (Bergmeier and Roellig 2014), putting
80 the balance between natural and economic values at risk.

81 Land-use intensification induces habitat homogenization and causes biodiversity loss in
82 wood-pastures (Flynn et al. 2009). Land abandonment also affects many wood-pastures (Godinho et
83 al. 2016) although its consequences are comparatively less studied than intensification (Estel et al.
84 2015). Land abandonment, which may have natural (Rey Benayas et al. 2007) and socio-economic
85 drivers (Levers et al. 2018), often leads to areas with little use beyond extensive grazing or even a
86 complete ceasing of all exploitation, resulting in less heterogeneous habitats with dense shrubby
87 vegetation (Rey Benayas et al. 2010). These landscape changes can influence biodiversity and
88 ecosystem processes, although the effects vary depending on the taxon and geographical region
89 (Queiroz et al. 2014). For instance, the loss of open areas can reduce habitat heterogeneity and,
90 consequently, the diversity of feeding and nesting resources for birds (Sirami et al. 2009). The
91 subsequent loss of bird species changes the trait composition of assemblages, which may in turn
92 affect ecosystem functioning and services (Hooper et al. 2005). Thus, assessing how trait
93 composition responds to land abandonment, particularly those traits that simultaneously influence
94 species responses and functions (i.e. “response and effect traits”; Luck et al. 2012; Díaz et al. 2013),
95 can reveal how biodiversity and ecosystem functions may be affected by this type of land use

96 change (Mouillot et al. 2013). This knowledge is crucial to inform managers and decision-makers
97 about potential negative consequences of such changes and to develop adequate strategies to
98 minimize them (Wood et al. 2015).

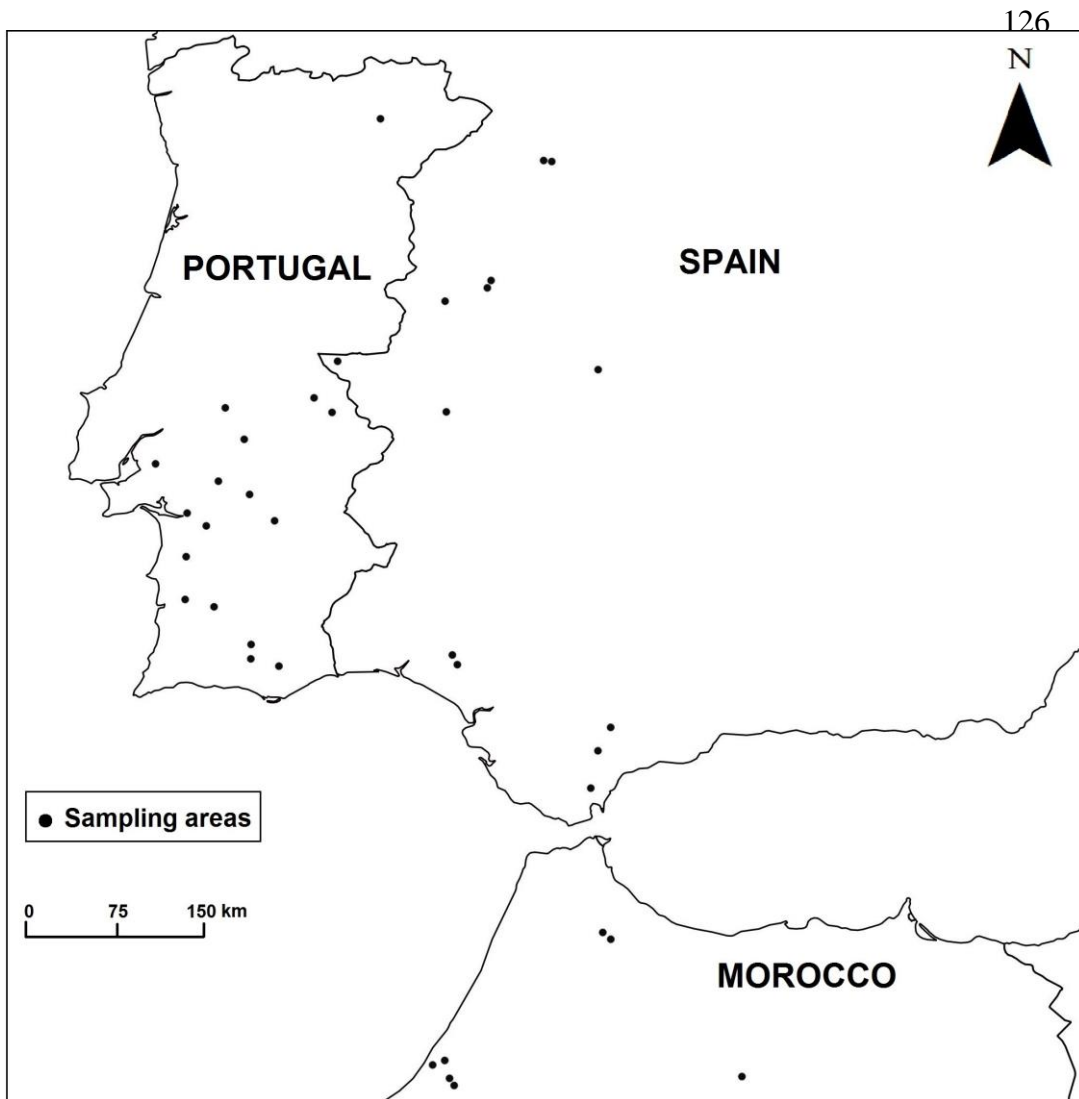
99 In this study, we aimed to explore how changes in habitat structure driven by different
100 management intensities influence bird taxonomic diversity, functional diversity and trait
101 assemblages in wood-pastures across the Iberian Peninsula and North Africa. Birds, being good
102 indicators of environmental change (Sekercioglu 2006), are likely to represent key ecosystem
103 services in wood-pastures, including seed dispersal, pest regulation or pollination (Pons and Pausas
104 2007; Ceia and Ramos 2016). We related taxonomic and functional indices to habitat structure,
105 characterized by a set of variables including understory cover and height, tree density and habitat
106 heterogeneity, as indicators of a management intensity gradient ranging from active management
107 towards land abandonment. Specifically, we aimed to answer the following questions: i) How do
108 changes in management intensity gradient influence the taxonomic and functional diversity of birds
109 in wood-pastures? ii) Which specific traits drive the responses of birds to the habitat along the
110 management intensity gradient?

111 **2. Methods**

112 **2.1 Study Area**

113 Our study encompassed much of Portugal, Spain and Morocco, in areas dominated by a
114 wood-pasture system distributed across the western Mediterranean, known as *montado* in Portugal,
115 *dehesa* in Spain and *azaghar* in some regions of North Africa (Fig. 1). The woody plant
116 composition of the sampled areas is dominated by cork oak (*Quercus suber*), sometimes co-
117 occurring with other oaks (e.g. Algerian oak *Q. canarensis*, holm oak *Q. rotundifolia*, Pyrenean oak
118 *Q. pyrenaica*), pine trees (e.g. stone pine *Pinus pinea*, maritime pine *P. pinaster*), and other tree and
119 shrub species (e.g. mastic tree *Pistacia lentiscus*, wild olive trees *Olea europaea* var. *sylvestris*,

120 strawberry tree *Arbutus unedo*, etc.). The most common management practices in the study area are
121 livestock grazing with various degrees of intensity, cork extraction, cropping and pruning
122 (Berrahmouni et al. 2007; Moreno et al. 2018). Moderate to intensive grazing is prevalent in
123 lowland wood-pastures, whereas shrub encroachment tends to occur in more rugged areas (Bugalho
124 et al. 2009; Bugalho et al. 2011). The annual average temperature ranges between 11-18°C and the
125 annual rainfall is 410-910 mm in the study area (<http://www.worldclim.org/>).



141
142 **Fig. 1** Map of the study areas located in Portugal, Spain and North Africa. Circles represent the
143 sampled wood-pastures in Portugal (N=17), Spain (N=13) and in Morocco (N=7).

144

145 **2.2 Bird sampling**

146 Bird sampling was performed during the spring of 2011 using five-minute bird point counts
147 (Bibby et al. 2005). In total, thirty-seven wood-pastures with a minimum of 50 hectares and at least
148 10 km apart, were sampled across Iberia (17 in Portugal, 13 in Spain) and North Africa (7 in
149 Morocco). We were able to investigate the biodiversity patterns of birds across Europe and North
150 Africa given the similarity of the bird species assemblages in both Europe and North Africa wood-
151 pastures (Correia et al. 2015a). Fifteen sampling stations were set up in each of these wood-pasture
152 areas, at least 200 m apart and 100 m or more from the edge. Each station was visited twice, once
153 during the early half (1 April to 15 May) and once during the late half (16 May to 20 June) of the
154 breeding season. The same observer performed all the counts, always during periods of peak bird
155 activity (controls in the morning and in the late afternoon for each area) and avoiding rainy and
156 windy conditions and areas recently harvested for cork (Godinho and Rabaça 2011). All birds
157 detected visually or acoustically were recorded and their distance to the observer was estimated.
158 Birds detected more than 100 m away from the observer and over-flying birds were excluded from
159 the analysis since they may not be relevant to the studied habitat. The total abundance of each
160 species in each area was defined as the maximum sum of individuals detected in the fifteen
161 counting stations for any of the two controls, as this represents the minimum number of birds
162 present in that area (Bibby et al. 2005). Bird abundance data is presented in Online Resource 1.

163

164 **2.3 Trait data**

165 We obtained data on six response and effect traits (Luck et al. 2012; Hevia et al. 2016) for
166 54 recorded species to analyse the functional diversity and functional composition of bird
167 assemblage. The traits considered for analysis include: habitat guild, feeding guild, foraging strata,
168 nest type, wing aspect ratio and body mass (Online Resource 2). These traits were selected based on

169 a priori hypotheses regarding their role in determining bird responses to habitat structure (Online
170 Resource 3).

171 **2.4 Environmental data**

172 Habitat structure was characterized with a set of variables including understory cover and
173 height, tree density and habitat heterogeneity (Online Resource 4). These habitat variables were
174 estimated in the same circle plots where bird census was performed. These variables were selected
175 as indicators of a management gradient ranging from areas with active human management (e.g.
176 regular grazing, shrub removal) to sparsely used areas (e.g. occasional grazing, no shrub removal)
177 resembling a process of land abandonment. Herb cover and shrub cover (% of ground cover), herb
178 height (in 5 cm classes up to 25 cm) and shrub height (in 25 cm classes up to 150 cm) were visually
179 estimated by the same observer. Tree density (number of trees per hectare) and canopy cover (%
180 ground cover covered by tree crowns) were estimated visually using aerial images available from
181 Google Earth v7.1 for the year 2011. A habitat heterogeneity index was obtained using six variables;
182 herb cover, herb height, shrub cover, shrub height, tree density and canopy cover applying the
183 formula of “max.value–min.value/mean value” for each habitat variable. The resulting values were
184 summed to obtain a single heterogeneity value for each wood-pasture (Rotenberry and Wiens 1980).
185 Finally, we performed a Principal Components Analysis (PCA) combining variables of herb and
186 shrub cover, herb and shrub height, tree density and habitat heterogeneity to reduce the number of
187 variables and avoid collinearity (Dormann et al. 2013). The first and the second axis of the PCA
188 were used to represent the main habitat management gradients in the subsequent analysis (Table A1,
189 Online Resource 5). The first PCA axis mostly represents ground vegetation structure ranging from
190 actively managed herb dominated and more heterogeneous areas to shrub dominated and less
191 heterogeneous areas, where the management intensity is lower. The second PCA axis represents a
192 gradient from sparsely treed and heterogeneous areas to densely treed and less heterogeneous areas.

193 Shrub cover and height are lower in these densely treed areas due to more intensive management.
194 (Figure A1, Online Resource 5).

195 **2.5 Data Analysis**

196 We estimated bird taxonomic diversity using species richness and the Shannon diversity
197 index, and estimated functional diversity with multi-trait functional richness (FRic), functional
198 dispersion (FDis) and functional evenness (FEve) indices (Villéger et al. 2008). Before each index
199 was calculated, we assessed the correlation between traits using Spearman correlations to avoid
200 biases in calculation of functional diversity indices (Lepš et al. 2006). There were no correlations
201 between traits ($p>0.05$), so they were all included in calculations and given equal weight. We also
202 calculated the community weighted means (CWM) index of each trait to test how individual trait
203 composition responded to changes in habitat structure.

204 All analyses were performed in R, version 3.5.2 (R Core Team 2019). Species richness and
205 Shannon index were calculated using the “vegan” package (Oksanen et al. 2016), and functional
206 diversity and functional composition indices with package “FD” (Laliberté et al. 2014). Later, we
207 tested the relationship between biodiversity indices and the management intensity gradients
208 represented by PCA axes using generalized linear models. Adjusted R-squared values were
209 calculated for each model using “rsq” function in “rsq” package (Zhang 2018). Additionally, we
210 performed a Detrended Correspondence Analysis (DCA) using “vegan” package to confirm that we
211 were able to test the biodiversity of birds across Europe and North Africa (Fig. A2 and Table A2,
212 Online Resource 5). All figures were produced using “ggplot2” (Wickham 2016).

213 **3. Results**

214 **3.1 Taxonomic diversity**

215 There were no evident changes in species richness or Shannon diversity in relation to the
216 first axis representing management intensity ($p>0.05$). Both species richness and diversity
217 decreased towards more densely treed and less heterogeneous areas (second axis representing
218 management intensity), but neither trend was significant ($p>0.05$) (Fig. A3 and Table A3, Online
219 Resource 5).

220

3.2 Functional diversity

221 Functional diversity showed a significant response to management, generally decreasing
222 towards lower management intensity. FRic showed a nearly-significant decrease in relation to PCA
223 axis1 ($p=0.08$), and thus towards areas with less management, where shrub cover and shrub height
224 are higher (Fig. 2a). Fdis ($p<0.001$; $R^2=-0.29$) and FEve ($p\leq 0.01$, $R^2=-0.21$) also decreased towards
225 shrub-dominated and less heterogeneous areas (Figs. 2b and 2c), and in both cases the decrease was
226 significant. No significant associations were detected between FD indices and the second PCA axis
227 ($p>0.05$). All diversity data is available in Table A4, Online Resource 5.

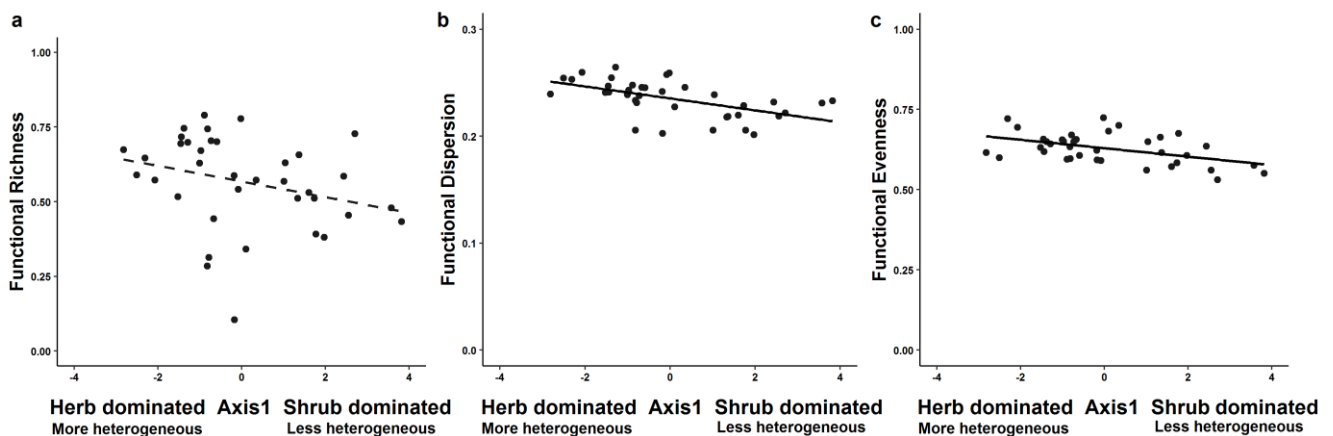


Fig. 2 Relation of bird functional diversity to habitat structure, represented by the first PCA axis.

All variables decrease towards shrub dominated, less managed areas, but the trend is significant

only in (b) and (c), represented by solid trend lines. See Table A4, Online Resource 5 for test statistics of linear models.

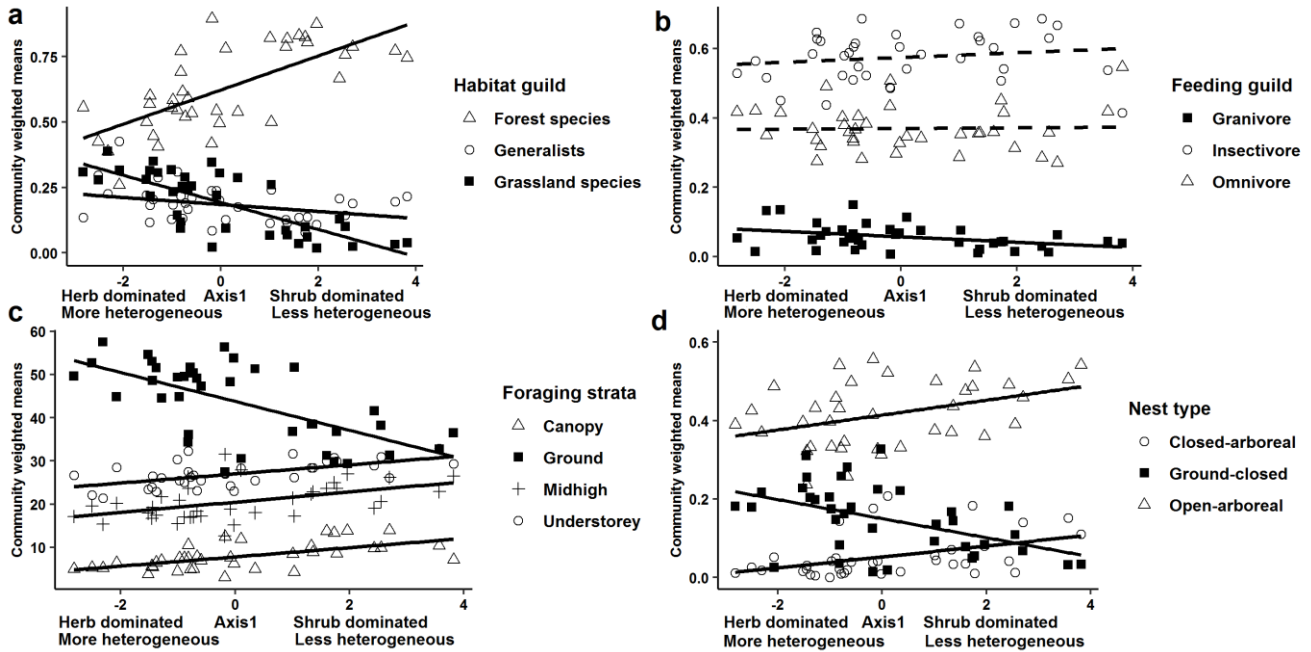
3.3 Functional composition

228 We observed significant variations in feeding guild, foraging strata, nest type and habitat use
229 traits of birds along the management gradient represented by the first PCA axis (Table A5, Online
230 Resource 5). CWM of grassland species ($p < 0.001$; $R^2 = -0.56$) and generalists ($p \leq 0.05$; $R^2 = -0.08$)
231 significantly decrease towards shrub dominated areas, while forest species ($p < 0.001$; $R^2 = 0.46$)
232 show the opposite trend. However, it should be noted that forest species are still well represented in
233 more open and heterogeneous areas, where they compose approximately half of the observed bird
234 communities, whereas grassland specialist species are often absent in shrub dominated areas (Fig.
235 3a).

236 CWM of granivores ($p \leq 0.05$; $R^2 = -0.11$) decreases significantly towards less managed,
237 shrub-dominated areas. However, the CWM of omnivore and insectivore species ($p > 0.05$) did not
238 vary significantly in relation to any of the habitat variables, indicating their relative abundance
239 remains constant independently of habitat structure (Fig. 3b). The relative abundance of ground-
240 foragers ($p < 0.001$; $R^2 = -0.39$) also decreases with higher shrub cover in areas tending towards
241 abandonment, while species foraging in the canopy ($p < 0.001$; $R^2 = 0.34$), midhigh ($p \leq 0.01$; $R^2 = 0.20$)
242 and understory ($p < 0.001$; $R^2 = 0.35$) benefit from the abundant woody vegetation available in these
243 areas (Fig. 3c). Closed-ground nesters ($p \leq 0.05$; $R^2 = -0.21$) are better represented in open and
244 heterogeneous areas maintained by active human management. On the other hand, closed-arboreal
245 ($p \leq 0.01$; $R^2 = 0.15$) and open-arboreal ($p \leq 0.05$; $R^2 = 0.12$) nesting birds are more prevalent in less
246 managed areas with higher shrub cover and tree density (Fig. 3d).

247 Only wing aspect ratio ($p \leq 0.05$; $R^2 = -0.08$) decreased with intensive shrub management, as
248 suggested by the significant negative relation to the second PCA axis (Fig. A4, Online Resource 5).

249 We did not observe significant variations in body mass, tree/cavity and bare ground nesting types in
 250 relation to habitat variables (Table A5, Online Resource 5).



251
 252 **Fig. 3** Community weighted means (CWM) of functional traits across the management gradient.
 253 CWM of generalists and grassland species (a) granivores (b) ground-foragers (c) closed-ground
 254 nesters (d) decreases towards shrub dominated, less managed areas, while forest species (a), canopy
 255 and understory foragers (c) and arboreal nesters (d) benefit from shrublands. Solid and dashed lines
 256 indicate significant ($p \leq 0.05$) and non-significant trends, respectively. See Table A5, Online
 257 Resource 5 for test statistics of linear models.

258
 259 **4. Discussion**

260 **4.1 Land abandonment leads to decreases in bird functional diversity, but not in taxonomic**
 261 **diversity**

262 Our results show that changes in vegetation structure and habitat heterogeneity linked to
263 land abandonment influence trait-level diversity of birds during the breeding season, but do not
264 influence taxonomic diversity. Neither species richness nor Shannon diversity metrics responded to
265 changes in habitat structure, indicating the presence of diverse species assemblages across the
266 management gradient (Fig. A3, Table A3, Online Resource 5). However, we found a significant
267 decline of functional dispersion and evenness toward less managed areas (Fig. 2), suggesting that
268 trait assemblages are more similar and less evenly distributed in wood-pastures within areas
269 undergoing active management (Hillebrand et al. 2008; Crowder et al. 2010). Studies report
270 contrasting trends in species and trait-level responses to various management strategies, suggesting
271 the need to explore different dimensions of biodiversity to understand the complex relation of
272 species to the ecosystem (Devictor et al. 2010). Our results support this view and suggest more
273 pronounced trait-level responses to changes in wood-pasture structure than those observed at the
274 species-level, mostly due to the loss of traits of grassland birds (Fig. 3a). Variations in niche
275 structure due to different management strategies are likely to explain the observed patterns, which
276 can benefit or impair species with specific traits without necessarily leading to changes in
277 taxonomic diversity (Gagic et al. 2015).

278 Land abandonment in wood-pastures is often characterised by a reduction in grazing and the
279 absence of activities commonly used to improve grazing potential, namely shrub removal, resulting
280 in higher tree and shrub densities, denser canopies and a decrease in habitat heterogeneity (Castro
281 and Freitas 2009; Oldén et al. 2017). Grassland species usually depend on open mosaics of habitats
282 to feed and nest (Reino et al. 2010), therefore the presence of denser and taller woody vegetation is
283 likely to reduce the availability of suitable conditions for these species (Sirami et al. 2007; Spitzer
284 et al. 2008). Our results suggest this is indeed the case in wood-pastures as the relative abundance
285 of grassland species decreased towards areas with denser and taller shrubby vegetation (Fig. 3a).
286 Grassland bird populations have suffered severe declines in recent decades all across Europe mostly

287 because of land-use intensification (Butler et al. 2010), but our findings underline land
288 abandonment may be another factor negatively affecting this guild. Traits such as ground nesting,
289 ground feeding and granivory are often available only in open areas, also decreased towards
290 abandoned areas (Fig. 3). These changes are likely due to the presence of dense shrubby vegetation
291 in these areas, which can substantially restrict seed availability, predator detectability and the
292 availability of suitable ground foraging sites for these species (Vickery and Arlettaz 2012).

293 In contrast, our results show forest species may benefit from land abandonment (Fig. 3a) and
294 similar results observed in wood-pastures across Europe (Sirami et al. 2009; Nikolov et al. 2011;
295 Jakobsson et al. 2018). Dense shrublands are known to be suitable for forest species such as
296 *Erithacus rubecula*, *Sylvia atricapilla* and *Troglodytes troglodytes* by providing feeding resources
297 and reducing predation (Santana et al. 2012). Furthermore, the availability of nesting sites for
298 arboreal nesters is likely to increase in densely wooded areas, as our results present (Fig. 3). We
299 also observed significant differences in birds linked to different foraging strata (Fig. 3), which is
300 informative of the way birds use the habitat under different management strategies (Martin and
301 Possingham 2005). The abundance of understory and canopy foragers increases whereas ground
302 foragers decrease in less managed areas, suggesting that land abandonment leads to changes in
303 different foraging strata use.

304 There may be important consequences to wood-pasture ecosystem dynamics related to
305 vertical shift of birds towards higher foraging strata in less managed areas. For example, many oak
306 pest species spend at least part of their life-cycle on or under the ground (Ceia and Ramos 2016).
307 Ground foraging birds may thus play an important role controlling these pests, but this role may be
308 limited in areas with dense ground vegetation, as our results further suggest.

309 We also observed that bird assemblages in more homogeneous areas with higher tree density
310 and less shrubs tend to feature more birds with a lower wing aspect ratio (Fig. A4, Online Resource

311 5). Lower wing aspect ratio represent shorter and more rounded wings allowing for better
312 manoeuvrability in densely treed habitats (Vanhooydonck et al. 2009) but are less suited for longer
313 distance flights.

314 **4.2 Active management maintaining habitat heterogeneity supports functionally diverse bird** 315 **communities in wood-pastures**

316 Our results provide evidence for the important role of management in maintaining bird
317 functional diversity in wood-pastures. Extensive grazing acts to maintain the characteristic habitat
318 heterogeneity of wood-pastures which is crucial to provide diverse feeding and nesting resources
319 for birds (Tews et al. 2004; Leal et al. 2019) particularly during the breeding season (Mag and Ódor
320 2015). Grassland species are especially dependent on resources that are only available in more open
321 and heterogeneous areas, which are usually sustained by extensive grazing (Reino et al. 2010).
322 Generalists also seem to increase with active management, whereas forest species dominate the
323 assemblage in the denser habitats that results from land abandonment. However, we underline that
324 forest species were also present in open managed areas, where they comprised approximately half
325 of the bird community, whereas grassland species are often absent in densely vegetated areas (Fig.
326 3a). This result indicates that by maintaining habitat heterogeneity, actively managed areas can
327 provide a wide range of niches, inclusively for most forest species.

328 While active management seems necessary to maintain heterogeneous wood-pasture
329 landscapes, the intensity of management also plays a key role. There is evidence that in highly
330 exploited systems, reducing management intensity level can be advantageous providing more
331 suitable and connected habitats for species narrower niche requirements (Queiroz et al. 2014). On
332 the other hand, studies focusing on multiple species responses to land abandonment also report the
333 potential threats of this change in land-use types to the overall biodiversity value of managed
334 habitats (Horák et al. 2018a). Specifically, land abandonment may ultimately lead to the functional

335 homogenization of biological communities in Mediterranean habitats (Clavero and Brotons 2010)
336 due to a loss of habitat heterogeneity, as our results also indicate.

337 It is obvious that both habitat intensification and abandonment can have adverse
338 consequences for the biodiversity of wood-pastures, whose preservation depends on the
339 maintenance of habitat heterogeneity emerging from multi-purpose management strategies
340 (Mönkkönen et al. 2014; Roellig et al. 2016). As more traditional management strategies struggle to
341 maintain the economic sustainability of wood-pastures, the challenge for the future is to find simple
342 and inexpensive management strategies that conciliate economic and natural values in wood-
343 pastures.

344 Ultimately, our results propose that maintaining active management may lead to higher
345 response trait diversity of birds in wood-pastures. Lastly, we should underline that the traits we
346 analysed are both “response and effect traits”, implying that many of the trait-level changes
347 observed may also reflect greater effect on the ecosystem services (de Bello et al. 2010; Torralba et
348 al. 2016).

349 **5. Conclusions**

350 We show substantial shifts in bird assemblage functional diversity and composition, in
351 response to vegetation structure and habitat heterogeneity changes associated with land
352 abandonment in wood-pastures. These changes seem to be unfavourable for grassland and
353 generalist species with life-history traits such as ground nesting, granivory and ground feeding,
354 ultimately leading to loss of bird functional diversity and potentially hampering the range of
355 ecosystem services that birds can provide. In many regions, wood-pastures are undergoing a
356 transformation from multi-functional heterogeneous systems to more homogeneous habitats, due to
357 a reduction of management resulting from multiple socio-economic pressures (Hartel et al. 2015).
358 The abandonment of traditional management, and the subsequent encroachment of woody

359 vegetation, are increasingly threatening the biodiversity of many European habitats (Queiroz et al.
360 2014). Balancing the natural and economic values of human-modified habitats is one of the main
361 challenges in improving conservation efforts (Graves et al. 2007; Landis 2017) and our results
362 suggest that maintaining non-intensive, active management may be a sustainable answer to this
363 challenge in wood-pastures.

364

365 **Declarations**

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370 **Conflicts of interest**

371 The authors have no relevant financial or non-financial interests to disclose.

372 **Ethics approval**

373 Not applicable.

374 **Consent to participate**

375 Not applicable.

376 **Consent for publication**

377 Not applicable.

378 **Availability of data and material**

379 All data generated or analysed during this study is included in Supplementary Information files
380 (Online Resources 1 and 2).

381 **Code availability**

382 Not applicable.

383 **Authors' contributions**

384 **Duygu P. Oksuz:** Conceptualization, Data Curation, Formal analysis, Investigation, Methodology,
385 Writing-original draft, Writing-review&editing.

386 **Jorge M. Palmeirim:** Conceptualization, Methodology, Supervision, Funding
387 acquisition, Writing-review&editing.

388 **Ricardo A. Correia:** Conceptualization, Data Curation, Investigation, Methodology, Funding
389 acquisition, Supervision, Writing-review&editing.

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393

394 **References**

395 Bergmeier E, Roellig M (2014) Diversity, threats and conservation of European wood-pastures. In:
396 Plieninger T, Hartel T (ed) European wood-pastures in transition: A Social-Ecological Approach,
397 Routledge, London, pp 19–38. <https://doi.org/10.4324/9780203797082>

398 Berrahmouni N, Escuté X, Regato P, Stein C (2007) Beyond cork: a wealth of resources for people
399 and nature. Lessons from the Mediterranean. WWF Mediterranean, Rome.

400 Bibby CJ, Burgess ND, Hill DA, Mustoe SH (2015) Bird census techniques. Elsevier Academic
401 Press, London. <https://doi.org/10.1016/c2009-0-03531-4>

402 Bugalho MN, Plieninger T, Aronson J, Ellatifi M (2009) Open woodlands: A diversity of uses (and
403 overuses). In: Aronson J, Pereira JS, Pausas JG (ed) Cork Oak woodlands on the edge. Ecology,
404 adaptive management, and restoration, Society for Ecological Restoration International, Island
405 Press, Washington D.C., USA, pp 33–47. <https://doi.org/10.1111/j.1526-100x.2010.00701.x>

406 Bugalho MN, Caldeira MC, Pereira JS, Aronson J, Pausas JG (2011) Mediterranean Cork Oak
407 Savannas require human use to sustain biodiversity and ecosystem services. *Front Ecol Environ*
408 9(5):278–286. <https://doi.org/10.1890/100084>

409 Butler SJ, Boccaccio L, Gregory RD, Voříšek P, Norris K (2010) Quantifying the impact of land-use
410 change to European farmland bird populations. *Agric Ecosyst Environ* 137:348–357.
411 <https://doi.org/10.1016/j.agee.2010.03.005>

412 Castro H, Freitas H (2009) Aboveground biomass and productivity in the Montado: from
413 herbaceous to shrub dominated communities. *J Arid Environ* 73:506–511.
414 <https://doi.org/10.1016/j.jaridenv.2008.12.009>

415 Ceia RS, Ramos JA (2016) Birds as predators of cork and holm oak pests. *Agrofor Syst* 90(1):159–
416 176. <https://doi.org/10.1007/s10457-014-9749-7>

417 Clavero M, Brotons L (2010) Functional homogenization of bird communities along habitat
418 gradients: accounting for niche multidimensionality. *Glob Ecol Biogeogr* 19(5):684–696.
419 <https://doi.org/10.1111/j.1466-8238.2010.00544.x>

420 Correia RA, Franco AMA, Palmeirim JM (2015a) Role of the Mediterranean Sea in differentiating
421 European and North African woodland bird assemblages. *Community Ecol* 16:106–114.

422 Crowder DW, Northfield TD, Strand MR, Snyder WE (2010) Organic agriculture promotes
423 evenness and natural pest control. *Nature* 466:109–112. <https://doi.org/10.1038/nature09183>

424 de Bello F, Lavorel S, Díaz S, et al. (2010) Towards an assessment of multiple ecosystem processes
425 and services via functional traits. *Biodivers Conserv* 19:2873–2893.
426 <https://doi.org/10.1007/s10531-010-9850-9>

427 Devictor V, Mouillot D, Meynard C, Jiguet F, Thuiller W, Mouquet N (2010) Spatial mismatch and
428 congruence between taxonomic, phylogenetic and functional diversity: the need for integrative
429 conservation strategies in a changing world. *Ecol Lett* 13:1030–1040.
430 <https://doi.org/10.1111/j.1461-0248.2010.01493.x>

431 Díaz S, Purvis A, Cornelissen JHC, Mace GM, Donoghue MJ, Ewers RM, Jordano P, Pearse WD
432 (2013) Functional traits, the phylogeny of function, and ecosystem service vulnerability. *Ecol Evol*
433 3:2958–75. <https://doi.org/10.1002/ece3.601>

434 Dormann CF, Elith J, Bacher S, Buchmann C, et al. (2013) Collinearity: a review of methods to
435 deal with it and a simulation study evaluating their performance. *Ecography* 36(1):27–46.
436 <https://doi.org/10.1111/j.1600-0587.2012.07348.x>

437 Estel S, Kuemmerle T, Alcántara C, Levers C, Prishchepov A, Hostert P (2015) Mapping farmland
438 abandonment and recultivation across Europe using MODIS NDVI time series. *Remote Sens*
439 *Environ* 163:312–325. <https://doi.org/10.1016/j.rse.2015.03.028>

440 Flynn DF, Gogol-Prokurat M, Nogeire T, Molinari N, Richers BT, Lin BB, DeClerck F (2009) Loss
441 of functional diversity under land use intensification across multiple taxa. *Ecol Lett* 12(1):22–33.
442 <https://doi.org/10.1111/j.1461-0248.2008.01255.x>

443 Gagic V, Bartomeus I, Jonsson T, Taylor A, Winqvist C, Fischer C, Slade EM, Steffan-Dewenter I,
444 Emmerson M, Potts SG, Tscharrntke T, Weisser W, Bommarco R (2015) Functional identity and
445 diversity of animals predict ecosystem functioning better than species-based indices. *Proc R Soc*
446 *Lond B Biol Sci* 282(1801):20142620. <http://dx.doi.org/10.1098/rspb.2014.2620>

447 Godinho C, Rabaça JE (2011) Birds like it Corky: the influence of habitat features and management
448 of ‘montados’ in breeding bird communities. *Agrofor Syst* 82:183–195.
449 <https://doi.org/10.1007/s10457-010-9345-4>

450 Godinho S, Guiomar N, Machado R et al. (2016) Assessment of environment, land management,
451 and spatial variables on recent changes in montado land cover in southern Portugal. *Agrofor Syst*
452 90(1):177–192, <http://dx.doi.org/10.1007/s10457-014-9757-7>.

453 Graves AR, Burgess PJ, Palma JHN, Herzog F, Moreno G, Bertomeu M, Dupraz C, Liagre F,
454 Keesman K, van der Werf W, de Nooy AK, van den Briel JP (2007) Development and application of
455 bio-economic modelling to compare silvoarable, arable, and forestry systems in three European
456 countries. *Ecol Eng* 29:434–449. <http://doi:10.1016/j.ecoleng.2006.09.018>

457 Hartel T, Plieninger T (2014) The social and ecological dimensions of wood-pastures. In: Plieninger
458 T, Hartel T (ed) *European wood-pastures in transition: A Social-Ecological Approach*, Routledge,
459 London, pp 3–18 <https://doi.org/10.4324/9780203797082>

460 Hartel T, Plieninger T, Varga A (2015) Wood-pastures in Europe In: Kirby KJ, Watkinsi C (ed)
461 *Europe's changing woods and forests: From wildwood to managed landscapes*, Chapter 5, vol.432.
462 CABI, Wallingford, pp 63–76. <https://doi.org/10.1079/9781780643373.0061>

463 Hevia V, Martin-Lopez B, Palomo S, Garcia-Llorente M, de Bello F, Gonzalez JA (2016) Trait-
464 based approaches to analyse links between the drivers of change and ecosystem services:
465 synthesizing existing evidence and future challenges. *Ecol Evol* 7:831–844.
466 <https://doi.org/10.1002/ece3.2692>

467 Hillebrand H, Bennett DM, Cadotte MW (2008) Consequences of dominance: a review of evenness
468 effects on local and regional ecosystem processes. *Ecology* 89:1510–1520.
469 <https://doi.org/10.1890/07-1053.1>

470 Hooper DU, Chapin FS, Ewel JJ, Hector A, et al. (2015) Effects of biodiversity on ecosystem
471 functioning: a consensus of current knowledge. *Ecol Monogr* 75:3–35. <https://doi.org/10.1890/04->
472 0922

473 Horák J, Pavlíček J, Kout J, Halda JP (2018a) Winners and losers in the wilderness: response of
474 biodiversity to the abandonment of ancient forest pastures. *Biodivers Conserv* 27:3019–3029.
475 <https://doi.org/10.1007/s10531-018-1585-z>.

476 Jakobsson S, Wood H, Ekroos J, Lindborg R (2018) Contrasting multi-taxa functional diversity
477 patterns along vegetation structure gradients of wooded pastures. In: Jakobsson S (2018) *Wooded or*
478 *treeless pastures?: Linking policy, farmers' decisions and biodiversity*. Dissertation, Stockholm
479 University.

480 Laliberté E, Legendre P, Shipley B (2014) FD: Measuring functional diversity from multiple traits,
481 and other tools for functional ecology. R package version 1:0–12.
482 <https://cran.rproject.org/web/packages/FD/FD.pdf>

483 Landis DA (2017) Designing agricultural landscapes for biodiversity-based ecosystem services.
484 *Basic Appl Ecol* 18:1–12. <https://doi.org/10.1016/j.baae.2016.07.005>

485 Leal AI, Rainho A, Martins RC et al. (2016) Modelling future scenarios to improve woodland
486 landscapes for birds in the Mediterranean. *J Nat Conserv* 30:103–112.
487 <https://doi.org/10.1016/j.jnc.2016.02.001>

488 Leal AI, Acácio M, Meyer CFJ et al. (2019) Grazing improves habitat suitability for many ground
489 foraging birds in Mediterranean wooded grasslands. *Agric Ecosyst Environ* 270–271:1–8.
490 <https://doi.org/10.1016/j.agee.2018.10.012>

491 Lepš J, de Bello F, Lavorel S, Berman S (2006) Quantifying and interpreting functional diversity of
492 natural communities: practical considerations matter. *Preslia* 78:481–501.

493 Levers C, Schneider M, Prishchepov AV, Estel S, Kuemmerle T (2018) Spatial Variation in
494 Determinants of Agricultural Land Abandonment in Europe. *Sci Total Environ* 644:95–111.
495 <https://doi.org/10.1016/j.scitotenv.2018.06.326>

496 Luck GW, Lavorel S, McIntyre S, Lumb K (2012) Improving the application of vertebrate trait-
497 based frameworks to the study of ecosystem services. *J Anim Ecol* 81:1065–1076.
498 <https://doi.org/10.1111/j.1365-2656.2012.01974.x>

499 Mag Z, Ódor P (2015) The effect of stand-level habitat characteristics on breeding bird assemblages
500 in Hungarian temperate mixed forests. *Community Ecology* 16:156–166.
501 <https://doi.org/10.1556/168.2015.16.2.3>

502 Martin TG, Possingham HP (2005) Predicting the impact of livestock grazing on birds using
503 foraging height data. *J Appl Ecol* 42:400–408. <https://doi.org/10.1111/j.1365-2664.2005.01012.x>

504 Moreno G, Aviron S, Berg S, et al. (2018) Agroforestry systems of high nature and cultural value in
505 Europe: provision of commercial goods and other ecosystem services. *Agrofor Syst* 92:877–891.
506 <https://doi.org/10.1007/s10457-017-0126-1>

507 Mouillot D, Graham NAJ, Villéger S, Mason NWH, Bellwood DR (2013) A functional approach
508 reveals community responses to disturbances. *Trends Ecol Evol* 28:167–177.
509 <https://doi.org/10.1016/j.tree.2012.10.004>

510 Mönkkönen M, Juutinen A, Mazziotta A, Miettinen K, Podkopaev D, Reunanen P, Salminen H,
511 Tikkanen OP (2014) Spatially dynamic forest management to sustain biodiversity and economic
512 returns. *J Environ Manage* 134:80–89. <https://doi.org/10.1016/j.jenvman.2013.12.021>

513 Nikolov SC, Demerdzhiev DA, Popgeorgiev GS, Plachiyski DG (2011) Bird community patterns in
514 sub-Mediterranean pastures: the effects of shrub cover and grazing intensity. *Anim Biodivers*
515 *Conserv* 34(1):11–21.

516 Oksanen O, Blanchet FG, Kindt R, Legendre P, McGlenn D (2016) *vegan: Community Ecology*
517 *Package*. R Package Version 2.3–0. Available from: <https://github.com/vegandevs/vegan>.

518 Oksuz DP, Aguiar CA, Tápiá S, et al. (2020) Increasing biodiversity in wood-pastures by protecting
519 small shrubby patches. *For. Ecol. Manage.* 464:118041.
520 <https://doi.org/10.1016/j.foreco.2020.118041>

521 Oldén A, Komonen A, Tervonen K, Halme P (2017) Grazing and abandonment determine different
522 tree dynamics in wood-pastures. *Ambio* 46(2):227–236. <https://doi.org/10.1007/s13280-016-0821-6>.

523 Pinto-Correia T, Ribeiro S (2012) HNV in 35 countries in Europe: Portugal. In: Oppermann R,
524 Beaufoy G, Jones G (ed) *High Nature Value Farming in Europe. 35 European countries-experiences*
525 *and perspectives*. Verlag Regionalkultur, Heidelberg, pp 336–345.
526 <https://doi.org/10.1659/mrd.mm126>

527 Pons J, Pausas JG (2007) Acorn dispersal estimated by radio-tracking. *Oecologia* 153(4):903–911.
528 <https://doi.org/10.1007/s00442-007-0788-x>

529 Plieninger T, Hartel T, Martín-lópez B, Beaufoy G, Bergmeier E, Kirby K, Montero MJ, Moreno G,
530 Oteros-Rozas E, Uytvanck JV (2015) Wood-pastures of Europe: Geographic coverage, social-
531 ecological values, conservation management, and policy implications. *Biol Conserv* 190:70–79.
532 <https://doi.org/10.1016/j.biocon.2015.05.014>

533 Queiroz C, Beilin R, Folke C, Lindborg R (2014) Farmland abandonment: threat or opportunity for
534 biodiversity conservation? A global review. *Front Ecol Environ* 12:288–296.
535 <https://doi.org/10.1890/120348>

536 R Core Team (2019) *R: A language and environment for statistical computing*. R Foundation for
537 Statistical Computing, Vienna, Austria. <https://www.R-project.org/>.

538 Reino L, Porto M, Morgado R et al. (2010) Effects of changed grazing regimes and habitat
539 fragmentation on Mediterranean grassland birds. *Agric Ecosyst Environ* 138:27–34.
540 <https://doi.org/10.1016/j.agee.2010.03.013>

541 Rey Benayas JM, Martins A, Nicolau JM, Schulz JJ (2007) Abandonment of agricultural land: an
542 overview of drivers and consequences. *CAB reviews: Perspectives in agriculture, veterinary science,*
543 *nutrition and natural resources*, 2 (57). <https://doi.org/10.1079/PAVSNNR20072057>

544 Rey Benayas JM, Galván I, Carrascal LM (2010) Differential effects of vegetation restoration in
545 Mediterranean abandoned cropland by secondary succession and pine plantations on bird
546 assemblages. *For Ecol Manage* 260(1):87–95. <https://doi.org/10.1016/j.foreco.2010.04.004>

547 Roellig M, Sutcliffe LME, Sammul M, von Wehrden H, Newig J, Fischer J (2016) Reviving wood-
548 pastures for biodiversity and people: A case study from western Estonia. *Ambio* 45:185–195.
549 <https://doi.org/10.1007/s13280-015-0719-8>

550 Rotenberry JT, Wiens JA (1980) Habitat Structure, Patchiness, and Avian Communities in North
551 American Steppe Vegetation: A Multivariate Analysis. *Ecology* 61:1228–1250.
552 <https://doi.org/10.2307/1936840>

553 Santana J, Porto M, Gordinho L, Reino L, Beja P (2012) Long-term responses of Mediterranean
554 birds to forest fuel management. *J Appl Ecol* 49:632–643. <https://doi.org/10.1111/j.1365-2664.2012.02141.x>

556 Sekercioglu CH (2006) Ecological significance of bird populations. *Handbook of the Birds of the*
557 *World* 11, Lynx Edicions, Spain, pp 15–51.

558 Sirami C, Brotons L, Martin JL (2007) Vegetation and songbird response to land abandonment:
559 from landscape to census plot. *Divers Distrib* 13(1):42–52. <https://doi.org/10.1111/j.1472-4642.2006.00297.x>

561 Sirami C, Seymour C, Midgley G, Barnard P (2009) The impact of shrub encroachment on savanna
562 bird diversity from local to regional scale. *Divers Distrib* 15:948–957.
563 <https://doi.org/10.1111/j.1472-4642.2009.00612.x>

564 Spitzer L, Konvička M, Beneš J, Tropek R, Tuf IH, Tufová J (2008) Does closure of traditionally
565 managed open woodlands threaten epigeic invertebrates? Effects of coppicing and high deer
566 densities. *Biol Conserv* 141:827–837. <https://doi.org/10.1016/j.biocon.2008.01.005>

567 Tews J, Brose U, Grimm V, Tielborger K, Wichmann MC, Schwager M, Jeltsch F (2004) Animal
568 species diversity driven by habitat heterogeneity/diversity: the importance of keystone structures. *J*
569 *Biogeogr* 31:79–92. <https://doi.org/10.1046/j.0305-0270.2003.00994.x>.

570 Torralba M, Fagerholm N, Burgess PJ, Moreno G, Plieninger T (2016) Do European agroforestry
571 systems enhance biodiversity and ecosystem services? A meta-analysis. *Agric Ecosyst Environ*
572 230:150–161. <http://dx.doi.org/10.1016/j.agee.2016.06.002>

573 Vanhooydonck B, Herrel A, Gabela A, Podos J (2009) Wing shape variation in the medium ground
574 finch (*Geospiza fortis*): an ecomorphological approach. *Biol J Linn Soc Lond* 98:129–138.
575 <https://doi.org/10.1111/j.1095-8312.2009.01269.x>

576 Vickery JA, Arlettaz R (2012) The importance of habitat heterogeneity at multiple scales for birds
577 in European agricultural landscapes. In: Fuller RJ (ed) *Birds and Habitat: Relationships in Changing*
578 *Landscapes*, Cambridge University Press, Cambridge, pp 177–204.
579 <https://doi.org/10.1017/cbo9781139021654.009>

580 Villéger S, Mason NWH, Mouillot D (2008) New multidimensional functional diversity indices for
581 a multifaceted framework in functional ecology. *Ecology* 89(8):2290–2301.
582 <https://doi.org/10.1890/07-1206.1>

583 Wickham H (2016) *ggplot2: Elegant Graphics for Data Analysis*. Springer–Verlag, New York.
584 <https://doi.org/10.1111/j.1541-0420.2011.01616.x>

585 Wood SA, Karp DS, DeClerck F, Kremen C, Naeem S, Palm CA (2015) Functional traits in
586 agriculture: agrobiodiversity and ecosystem services. *Trends Ecol Evol* 30:531–539.
587 <https://doi.org/10.1016/j.tree.2015.06.013>