ENVIRONMENTAL BIOTECHNOLOGY 9 (2) 2013, 65-71; doi:10.14799/ebms193

# Assessment of mutagenic activity of methyl- and phenylphenanthrenes based on *Salmonella* test and micronucleus test

# Katarzyna Rudnicka<sup>1</sup>, Sebastian Tejs<sup>1</sup>, Kinga A. Budzikur<sup>1</sup>, Danuta Mielżyńska-Švach<sup>2</sup>, Ewa Jakimiuk<sup>3</sup>, Anna Chachaj<sup>4</sup>, Maciej Góra<sup>4</sup>, Kamila Żelazna<sup>1</sup>, Michał K. Łuczyński<sup>5</sup>

<sup>1</sup> Department of Environmental Biotechnology, University of Warmia and Mazury in Olsztyn, Sloneczna 45G, 10-718 Olsztyn, Poland

<sup>2</sup> Institute of Occupational Medicine and Environmental Health, Kościelna 13, 41-200 Sosnowiec, Poland

<sup>3</sup> Department of Veterinary Prevention and Feed Hygiene, University of Warmia and Mazury in Olsztyn, Oczapowskiego 14, 10-817 Olsztyn, Poland

<sup>4</sup> Department of Organic Chemistry, Jagiellonian University, Ingardena 3, 30-060 Kraków, Poland

<sup>5</sup> Department of Chemistry, University of Warmia and Mazury in Olsztyn, Plac Łódzki 4, 10-957 Olsztyn, Poland

Corresponding author: Katarzyna Rudnicka, Department of Environmental Biotechnology, University of Warmia and Mazury in Olsztyn, Sloneczna 45G, 10-718 Olsztyn, Poland; Phone: +48 089 5234144; E-mail: katarzyna.rudnicka@uwm.edu.pl

Key words: Ames test, in vitro, micronucleus test, mutagenicity, phenanthrenes

Received in February 2013. Published in April 2014.

## ABSTRACT

Polycyclic aromatic hydrocarbons (PAHs) are widely spread environmental pollutants mainly originating from anthropogenic sources such as fossil fuel combustion, industries, and others. Although a large body of literature exists on the toxicity and carcinogenicity of PAHs, primarily benzo[*a*]pyrene, toxicity data for phenanthrene deriveratives are very limited. The main aim of the experiment was to investigate if there exists correlation between molecular structure and mutagenic activity of four phenanthrene, derivatives: 1-methylphenanthrene, 4-methylphenanthrene, 1-phenylphenanthrene, and 4-phenylphenanthrene. An Ames assay using two strains of histidine dependent *Salmonella* Typhimurium (TA98 and TA100) was conducted to assess the

# INTRODUCTION

Polycyclic aromatic hydrocarbons (PAHs) are widely spread environmental pollutants mainly originating from anthropogenic sources such as fossil fuel combustion, industries, and others, however, vehicular traffic has been one of the major sources of air pollution (Gaffney and Marley 2009; Ravindra et al. 2008). PAHs which are released into the atmosphere are redistributed between the gaseous

© UNIVERSITY OF WARMIA AND MAZURY IN OLSZTYN

mutagenic activity of studied compounds both in the presence (+S9) and in the absence (-S9) of an exogenous source of metabolic activation. The compounds were also tested in an *in vitro* chromosome aberration assay in which V-79 cells were exposed to the phenanthrene derivatives investigated both in the presence and in the absence of metabolic activation. The phenylphenanthrenes showed no mutagenic effect. These compounds occasionally induced significant decrease in the number of revertants in the Ames test. The greatest mutagenic effects were observed for 1-methylphenanthrene after metabolic activation (+S9). In the micronucleus test the greatest mutagenic effect was observed for 4-methylphenanthrene also in the presence of metabolic activation system. The results obtained are comparable to those reported earlier for the methylphenanthrenes.

and particulate phases. Low-molecular-weight (LMW) PAHs tend to be more concentrated in the vapour phase, while those with higher molecular weight (HMW) are often associated with particulates (Bidleman 1988; Harner and Bidleman 1998; Pankow 1998). Dry and wet deposition and volatilization from the water and soil are the main processes that contribute to air/water and air/soil exchange of PAHs (Godish 1991). These molecules are predominantly found in the environment surrounding urban zones with high vehicular density and/or presence of industrialized areas (Godish 1991; Macdonald et al. 2005).

PAH toxicity ranges from non-toxic through extremely toxic to humans. Many PAHs are mutagenic (Mielzynska et al. 2006), carcinogenic (Bostrom et al. 2002), and/or teratogenic (Singh et al. 2008).

Phenanthrene and its derivatives can be derived from coal tar. Phenanthrene is ubiquitous in the environment as a product of incomplete combustion of fossil fuels and wood (Wrobel and Reinhardt 2003). It appears in the exhaust gas produced in diesel engines (Levsen 1988) and has been identified in ambient air, surface and drinking water, and in foods (IARC 1983; U.S. EPA 1988). Based on data from animal bioassays, U.S. EPA (1987, 1993) has placed phenanthrene in weight-of-evidence group D, not classifiable as to human carcinogenicity. The compound accumulates in the tissues and is toxic to the aquatic form of liverwort *Riccia fluitans* (Burrit 2008).

Data regarding the gastrointestinal or pulmonary absorption of phenanthrene in humans are not available. However, data from structurally related PAHs suggest that phenanthrene would be absorbed readily from the gastrointestinal tract and lungs. In general, PAHs are highly lipid-soluble and can pass across epithelial membranes (U.S. EPA 1987). Although many PAHs accumulate in body fat, Bock and Dao (1961) found little phenanthrene in the perirenal and mammary fat of rats administered the compound by gavage 24 hours earlier. Phenanthrene itself is one of the PAH model molecules because of its low toxicity, as well as its structural and physical similarity to other PAHs (Dabestani and Ivanov 1999).

Among others, phenanthrenes and their derivatives are the most common PAHs widely distributed in the whole environment. Although an extensive body of literature exists regarding the toxicity and carcinogenicity of PAHs, primarily benzo[*a*]pyrene, toxicity data for phenanthrene and their deriveratives are very limited. Phenanthrene is a relatively good substrate for cytochrome P-450 monooxygenases (Nordqvist et al. 1981). Studies in the early 1980-ties have shown that for different strains of *Salmonella* Thyphimurium (TA97, TA98, TA100) phenanthrene, as well as its metabolites, was not mutagenic (Sakai et al. 1985). Mutagenic activity of some methyl derivatives of phenanthrene has been previously reported (LaVoie et al. 1981). In 1953, attempts were made to study 2-phenylphenanthrene derivatives, which were characterized by a slight activity or did not show any biological activity. Even in small quantities these compounds occurring together with other fuel components, can cause mutagenic or carcinogenic effects (Levsen 1988). The carbon skeleton of the phenanthrene molecule is an excellent model for the study of correlations of PAH structure and substitution pattern with its genotoxic activity (Łuczyński et al. 2005).

In this paper, from the large group of methyl- and phenylphenanthenes, four compounds were selected because it is likely that the structure of their molecules promotes their conversion by organisms to mutagenic active metabolites.

## MATERIAL AND METHODS

#### Chemistry

1-methylphenanthrene, 4-methylphenanthrene, 1-phenylphenanthrene and 4-phenylphenanthrene were synthesized because they were not available commercially. All chemicals were synthesized at the Faculty of Chemistry in Jagiellonian University by the research group of Chemistry of Carbocyclic Compounds. Purity of samples was confirmed by NMR spectroscopy and was greater then 95%. Gas chromatography (GC-FID) analysis showed no impurities. Yield of isolated phenananthrene derivatives was within range of 81-95%. Chemical structure of these compounds is in Figure 1.

Synthesis scheme for methyl-, and phenylphenanthrene is shown in Figure 2.

The methyl and phenylphenanthrenes were obtained according to modified synthesis published by Haworth (1932) except for the first two steps leading to naphthylbutyric acids (Figure 2 steps 4 and 5) which remained unmodified. Compared to the original procedure significant improvements were achieved at the cyclisation step. Heating of 4 and 5 in anhydrous methanesulfonic acid led directly to the production of phenanthrenones 6 and 7 with yields exceeding 60%. This

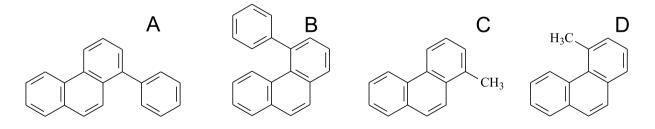


Figure 1. The structure of tested chemicals: A, 1-phenylphenanthrene; B, 4-phenylphenanthrene; C, 1-methylphenanthrene; D, 4-methylphenanthrene.

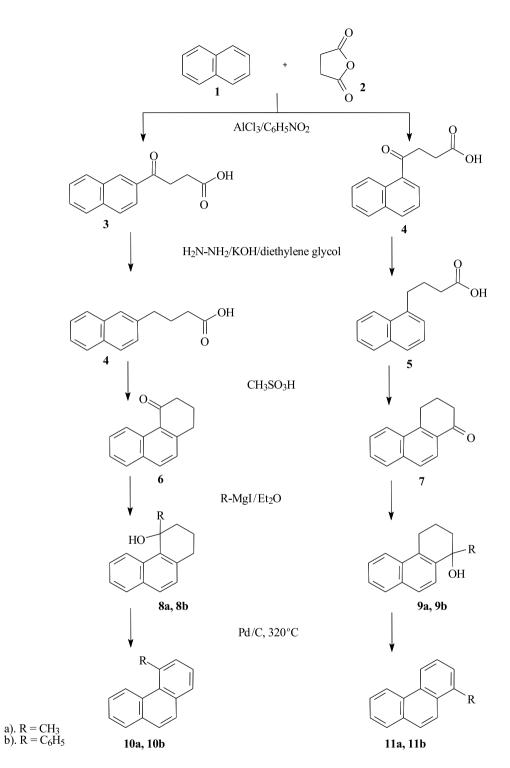


Figure 2. Scheme of synthesis of methyl- and phenylphenanthrene (Rospondek et al. 2009).

procedure allow to avoid the conversion of the naphthylbutyric acids to appropriate acid chlorides and further cyclisation in presence of SnCl<sub>4</sub>. Reaction of phenylmagnesium or methylmagnesium iodides with 1,2,3,4-tetrahydrophenanthrenones **6** and **7** resulted in formation of phenanthrols **8** and **9**. Heating of isolated phenanthrols with 10% palladium on charcoal under slow stream of argon caused their simultaneous dehydration and aromatization to methyl- and phenylphenanthrenes **10** and **11**. Melting points and spectroscopic data of isolated methyland phenylphenanthrenes were identical with those previously obtained (Rospondek et al. 2009).

# The Salmonella test

Histidine dependent strains TA98 and TA100 of *Salmonella* Typhimurium have been purchased from TRINOVA Biochem GmbH Germany.

The Ames mutagenicity test was conducted with two bacterial strains of S. Typhimurium (strains TA98 and TA100) with and without metabolic activation ( $\pm$ S9 mix). The S9 microsome fraction is induced with Aroclor 1254 (Trinova Biochem GmbH Germany) and is able to stimulate eukaryotic processes that cannot be conduced by microorganisms themselves. The strains were cultured as described by Maron and Ames (1983) and Mortelmans and Zeiger (2000). Confirmation of the genotypes of the tester strains was routinely carried out, including crystal violet, UV and ampicillin sensitivities. The number of spontaneous revertants, induced in the negative controls (solvent, S9) and following exposure to such diagnostic mutagens as NQNO  $(10\mu g \cdot plate^{-1})$  and B[a]P  $(10\mu g \cdot plate^{-1})$  were measured in each experiment. In these studies, the tested compound was dissolved in 80µL of DMSO and added to 2.5mL molten top agar (at 42°C) with 18h of nutrient broth culture of appropriate strain of S. Typhimurium and 0.5mL of S9 mix. The final mixture was poured onto minimal glucose agar plates. Mutations from histidine-dependent to histidineindependent bacteria were assessed 48h after plating by counting the colonies of bacteria in the Petri dishes. The compounds were assayed in triplicate at each dose level based on the solubility of the compounds.

#### In vitro micronucleus test (MN)

Cell line V-79 was purchased from Leibniz Institute DSMZ-German Collection of Microorganisms and Cell Cultures, Germany.

*In vitro* micronucleus test was conducted with V-79 cell line according to ISO 21427-2:2006. It was performed on 4-well plates (1mL of medium) and consisted of four stages. Test conditions for all steps were the same: temperature 37°C, humidity 95% and 5% carbon dioxide atmosphere. For all the steps culture medium consisted of 90% of MEM (Minimal Essential Medium), 10% of FBS (Fetal Bovine Serum) and 1% of antibiotics solution (Amphotericin B, Streptomycin, Penicillin) (Rudnicka et al. 2012).

The pre-incubation of cell cultures was carried out with use of minimal essential medium supplemented with 10% of fetal bovine serum, 1% antibiotics. At the beginning of the test, cell density was 5-8·10<sup>4</sup> cells per well. Incubation was carried out for 48 hours. Next stage was divided into two variants. In the variant without metabolic fraction (-S9) the cell cultures after pre-incubation were covered with fresh medium supplemented with culture appropriate concentrations of test substances, based on the toxicity according to OECD guideline for testing chemicals (maximal volume of tested chemical if dissolved in DMSO should be 1% of volume of the well), with negative and positive controls. Then the cells were incubated for 24 hours. The second variant was preceded by incubation of the test substances, with negative and positive controls, with addition of S9 (+S9) (in this variant FBS should not be added to culture medium) and the appropriate volume of culture medium for 1 hour. After incubation, the solutions obtained were filtered on sterile syringe filters ( $0.22\mu m$ ). Cell cultures were covered with solutions of test substances. Incubation was performed for 24h. After incubation with the respective chemicals culture medium was removed from the solutions of test substances. Then the cultures were covered with fresh medium supplemented with cytochalasin B at a concentration of 3µg·mL<sup>-1</sup>. Cells were then incubated for 24h. At the end of the test, the Giemza staining protocol was employed. Thereafter probes were closed with use of DPX mounting medium and counted with use of light microscope (Fenech 2006; Fenech et al. 1999; Phelps et al. 2003).

#### Statistical data analysis

The method of least squares regression analysis and analysis of variance were used. Doses of test compounds were treated as an independent variable, and the number of revertants induced in strains of S. Typhimurium (TA98 and TA100) as the dependent variable. The number of revertants in the corresponding negative control was treated as a response to zero dose. In the event of a toxic effect or appearance of the same response with increasing doses, the analysis was performed again with the rejection of the results for the highest doses. In order to compare the mutagenic potency of the tested compounds their mutagenic rates (MR) were calculated. MR is a relative measure of the mutagenic potency and is defined as the ratio of the number of revertants observed at a given dose to the average number of revertants observed in the negative control. A positive mutagenic response is indicated when MR values are  $\geq 2$ .

For the *in vitro* micronucleus test data from the experiments were analyzed using Student t-test, considering number of micronuclei observed in 1000 cells at each concentration level, including the control. A type I error rate of 0.05 was used in the statistical analyses. A positive effect was declared for a chemical when p-value was less than or equal to 0.05.

Rudnicka et al.

# RESULTS

## Salmonella test

None of the four studied compounds showed mutagenic activity in the absence of a metabolic activation (-S9).

For S. Typhimurium strain TA98, only 1-methylphenanthrene showed a distinct mutagenic effect in the presence of S9. The highest activity appeared at a concentration of  $50\mu$ g·plate<sup>-1</sup> and was associated with MR value of 11.25. The number of revertants at the  $200\mu$ g·plate<sup>-1</sup> dose level was significantly lower, but there was still a significant mutagenic effect observed (MR=8.44) (Figure 3). 4-methylphenanthrene showed weak toxic effect for strain TA98 (Figure 4).

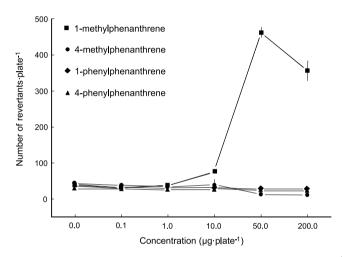
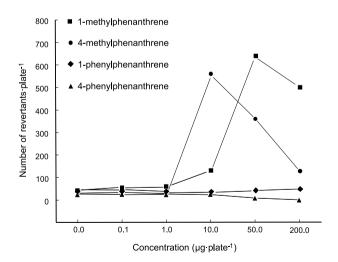


Figure 3. Number of revertants caused by phenanthrene derivatives and their metabolites (+S9) in the *Salmonella* Typhimurium strain TA98, mean values  $\pm 0.95$  confidence interval.



In the case of strain TA100, a mutagenic effect occurred in the variant of the enzyme activation. 1-methylphenanthrene induced higher number of revertants (Figure 4) what indicated mutagenic activity of this compound. The highest effect was caused by the dose of  $50\mu$ g·plate<sup>-1</sup> (MR=4.13). The highest number of revertants for 4-methylphenanthrene appeared at the concentration of  $10\mu$ g·plate<sup>-1</sup> (MR=3.56).

#### In vitro micronucleus test

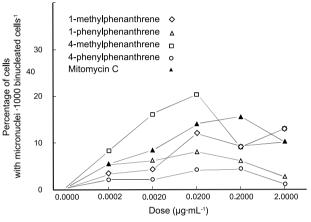


Figure 5. Percentage of micronuclei among 1000 binucleated cells induced by phenanthrene derivatives in V-79 cells (-S9), mean values±0.95 confidence interval (spontaneous micronulei 0.3%).

All examined methyl- and phenylphenanthrenes showed evidence of mutagenic effects at all concentrations (Figure 5, 6). These chemicals induced significant increase in number of

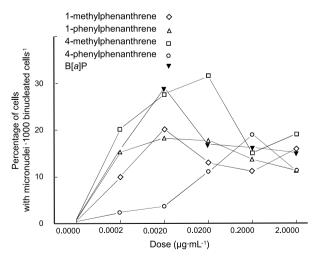


Figure 4. Number of revertants caused by phenanthrene derivatives and their metabolites (+S9) in the *Salmonella* Typhimurium strain TA100, mean values  $\pm 0.95$  confidence interval.

Figure 6. Percentage of micronuclei among 1000 binucleated cells induced by phenanthrene derivatives in V-79 cells (+S9), mean values  $\pm 0.95$  confidence interval (spontaneous micronulei 0.3%).

micronuclei in V-79 cells, but the effect was more pronounced with S9 metabolic activation. The greatest effect was observed for metabolites of 4-methylphenanthrene. These compounds at a dose of  $0.02\mu$ g·mL<sup>-1</sup> induced the maximum number, 317 cells with micronuclei per 1000 binucleated cells (MN/1000BC). At the same dose, metabolites of 1-phenylphenanthrene caused 178 MN/1000 BC. The weaker response at the same dose appeared for metabolites of 1-methylphenanthrene (120 MN/1000BC) and 4-phenylphenanthrene (115 MN/1000 BC), but there was still a mutagenic effect observed.

## DISCUSSION

The main aim of the experiment was the examination of potential mutagenic activity of four phenanthrene derivatives and their metabolites. In their pure form, PAHs, like phenanthrene, cannot be considered as strong carcinogens (Korytko et al. 2000), but they can become highly mutagenic as a result of their metabolism, especially by P-450 enzyme systems (Ioannides 2007).

The mutagenicity and in vitro metabolism of selected methylphenanthrenes have been previously evaluated. The monomethyl isomers which were mutagenic toward S. Typhimurium were 1- and 9-methylphenanthrenes. Among the disubstituted phenanthrenes assayed for mutagenicity, 1,4-dimethylphenanthrene was active in the presence of metabolic activation (LaVoie et al. 1981). In our studies we confirmed the mutagenic activity of 1-methylphenanthrene in the Salmonella test, which was observed for both tested strains: TA98 and TA100 after metabolic activation (+S9). This result indicates that metabolites of this compound can cause both frameshift and base-pair substitution point mutation. Metabolites of 4-methylphenanthrene showed mutagenic effects only for S. Typhimurium strain TA100. In the case of this chemical the mutagenic effect appeared at a lower concentration compared to 1-methylphenanthrene, what may suggest greater mutagenic activity of 4-methylphenanthrene. This may result from the position of the methyl group, located in position 4 (i.e., close to "bay-region"). In this position, the methyl substituent can form a "fjord-region", which has a higher affinity to DNA and is able to create DNA-adducts (LaVoie et al. 1981).

Phenylphenanthrenes and their metabolites were not mutagenic to both *Salmonella* strains. This was probably due to the fact that the test compounds did not form DNA adducts and did not lead to the formation of mutations. Mutagenicity potential of 1-methylphenanthrene is associated with inhibition of 9,10-dihydrodiol formation. LaVoie et al. (1981) studies of the methylphenanthrenes metabolism confirm that for the non-mutagenic derivatives, the of 9,10-dihydrodiol is a major metabolite. *In vivo* and *in vitro* studies indicate that metabolism of phenantrene proceeds via epoxidation at the 1-2, 3-4, and 9-10 carbons, with dihydrodiols as primary metabolites of which the 9,10-dihydrodiol is the major component. The 9-10-, 1,2-, and 3,4-dihydrodiols of phenanthrene were identified unaltered or as glucuronic acid conjugates in the urine of rats and rabbits that received intraperitoneal injections of phenanthrene (Boyland and Sims 1962; Boyland and Wolf 1950). Also the glucuronic acid conjugates of 1-, 2-, 3-, and 4-hydroxyphenanthrene, 1,2-dihydroxyphenanthrene, and 3,4-dihydroxyphenanthrene were identified (Boyland and Sims 1962). *In vitro* studies with guinea pig, rat, and mouse liver preparations also identified the 9-10-, 1,2-, and 3,4-dihydrodiols. Further oxidative metabolism to the 9,10-oxide and 1,2-diol-3,4-epoxide of phenanthrene has also been reported (Chaturapit and Holder 1978; Nordqvist et al. 1981; Sims 1970).

The results of the micronucleus test showed that metabolites of 4-methylphenanthrene caused formation of more micronuclei then 1-methylphenanthrene. These results indicate that 4-methylphenanthrene may be considered a stronger mutagen relative to 1-methylphenanthrene, a result trends to echo that of the Salmonella test. An increased frequency of micronuclei formation in the case of 4-methylphenanthrene may result from the position of methyl substituent which is close to the "bay region". It can be concluded that the presence of such a substituent in the "bay region" increases the mutagenic activity of phenanthrene derivatives. Methyl group in position 1 does not produce the same strong effect as for 4-methylphenanthrene. This results from the fact that the methyl group is a small substituent which does not prevent the access of metabolic enzymes to the "K-region". In contrast, greater mutagenic effect of phenyl derivatives was observed for 1-phenylphenanthrene than that for 4-phenylphenanthrene. The reason for theses results could be due to the difference in the location of the phenyl substituent. For 1-phenylphenanthrene it is close to the "K-region" what may result in the formation of diolepoxides.

Parent phenanthrene has a reduced mutagenic effect on bacteria and mammalian cells relative to 1- and 4-methylphenanthrenes (Wood et al. 1979). This difference in the type of interaction may result from differences in the molecular structure of the studied compounds. Substituents in the molecule may influence interactions between the molecule and the genetic material.

To our knowledge, there are no reports concerning particular metabolic pathways of 1-phenylphenanthrene and 4-phenylphenanthrene, so further experiments on this subject would be of interest.

## ACKNOWLEDGEMENTS

This work was supported by Ministry of Science and Higher Education: project N N305 323136 and 0536/R/TO2/ 2007/03.

Rudnicka et al.

## REFERENCES

- Bidleman, T.F. 1988. Atmospheric processes: wet and dry deposition of organic compounds are controlled by their vapor-particle partitioning. Environmental Science and Technology 22: 361–367.
- Bock, F.G., T.L. Dao. 1961. Factors affecting the polynuclear hydrocarbon level in rat mammary glands. Cancer Research 21: 1024-1029.
- Bostrom, C.E., P. Gerde, A. Hanberg, B. Jernstrom, C. Johansson, T. Kyrklund, A. Rannug, M. Tornqvis, K. Victorin, R. Westerholm. 2002. Cancer risk assessment, indicators, and guidelines for polycyclic aromatic hydrocarbons in the ambient air. Environment Health Perspectives 110: 451–488.
- Boyland, E., P. Sims. 1962. Metabolism of polycyclic compounds. The metabolism of phenanthrene in rabbits and rats: Dihydrodihydroxy compounds and related glucosiduronic acids. Biochemical Journal 84: 571-582.
- Boyland, E., G. Wolf. 1950. Metabolism of polycyclic compounds. 6. Conversion of phenanthrene into dihydroxydihydrophenanthrene. Biochemical Journal 47: 64-69.
- Burrit, D.J. 2008. The polycyclic aromatic hydrocarbon phenanthrene causes oxidative stress and alters polyamine metabolism in the aquatic liverwort *Riccia fluitans* L. Plant Cell and Environment 31: 1416-1431.
- Chaturapit, S., G.M. Holder. 1978. Studies on the hepatic microsomal metabolism of (14C)phenanthrene. Biochemical Pharmacolgy 27: 1865-1872.
- Dabestani R., I.N. Ivanov. 1999. A compilation of physical, spectroscopic and photophysical properties of polycyclic aromatic hydrocarbons. Photochemistry and Photobiology 70: 10-34.
- Fenech, M. 2006. Cytokinesis-block micronucleus assay evolves into a "cytome" assay of chromosomal instability, mitotic dysfunction and cell death. Mutation Research 600: 58-66.
- Fenech, M., J. Crott, J. Turner, S. Brown. 1999. Necrosis, apoptosis, cytostasis and DNA damage in human lymphocytes measured simultaneously within the cytokinesis-block micronucleus assay: description of the method and results for hydrogen peroxide. Mutagenesis 14: 605-612.
- Gaffney, J.S., N.A. Marley. 2009. The impacts of combustion emissions on air quality and climate – from biofuels and beyond. Atmospheric Environment 43: 23–36.
- Godish, T. 1991. Air Quality. 480p. Lewis Publishers, Michigan.
- Harner, T., T.F. Bidleman. 1998. Octanol-air partition coefficient (KOA) for describing particle/gas partitioning of aromatic compounds in urban air. Environmental Science & Technology 32: 1494–1502.
- Haworth, R.D. 1932. Haworth phenanthrene synthesis. Journal of the Chemical Society 29: 96-219.
- IARC (International Agency for Research on Cancer). 1983. IARC Monographs on the Evaluation of Carcinogenic Risks to Humans. Polynuclear Aromatic Compounds. Part 1. Chemical, Environmental and Experimental Data, Vol. 32. pp. 419-430. World Health Organization, Lyon, France.
- Ioannides, C. 2007. Bioactivation of chemicals by cytochromes P450. Environmental Biotechnology 3: 1-9.
- Korytko, P.J., F.W. Quimby, J.G. Scott. 2000. Metabolism of phenanthrene by house fly CYP6D1 and dog liver cytochrome P450. Journal of Biochemical and Molecular Toxicology 14: 20-25.
- LaVoie, E.T., L. Tulley-Freiler, V. Bedenko, D. Hoffman. 1981. Mutagenicity, tumor-initiating activity, and metabolism of methylphenanthrenes. Cancer Research 41: 3441--3447.
- Levsen, K. 1988. Mass spectrometry in environmental organic analysis. Organic Mass Spectrometry 23: 406–415.
- Łuczyński, M.K., M. Góra, P. Brzuzan, J. Wilamowski, B. Kozik. 2005. Oxidative metabolism, mutagenic and carcinogenic properties of some polycyclic aromatic hydrocarbons. Environmental Biotechnology 1: 16-28.

- Macdonald, R.W., T.T. Harner, J. Fyfe. 2005. Recent climate change in the Arctic and its impact on contaminant pathways and interpretation of temporal trend data. Science of The Total Environment 342: 5–86.
- Maron, D.M., B.N. Ames. 1983. Revised methods for the Salmonella mutagenicity test. Mutation Research 113: 173-215.
- Mielzynska, D., E. Siwinska, L. Kapka, K. Szyfter, L.E. Knudsen, D.F. Merlo. 2006. The influence of environmental exposure to complex mixtures including PAHs and lead on genotoxic effects in children living in Upper Silesia, Poland. Mutagenesis 21: 295–304.
- Mortelmans, K., E. Zeiger. 2000. The Ames Salmonella/microsome mutagenicity assay. Mutation Research 455: 29-60.
- Nordqvist, M., D.R. Thakker, K.P. Vyas, H. Yagi, W. Levin, D.E. Ryan, P.E. Thomas, A.H. Conney, D.M. Jerina. 1981. Metabolism of chrysene and phenanthrene to bayregion diol epoxides by rat liver enzymes. Molecular Pharmacology 19: 168-178.
- Pankow, J.F. 1998. Further discussion of the octanol/air partition coefficient K-oa as a correlating parameter for gas/particle partitioning coefficients. Atmosphere Environment 32: 1493-1497.
- Phelps, J.B., W.P. Hoffman, M.L. Garriott. 2003. Relative percent cell survival and positive response in the *in vitro* micronucleus test. Mutation Research 537: 115-116.
- Ravindra, K., R. Sokhia, R. Van Griekenet. 2008. Atmospheric polycyclic aromatic hydrocarbons: source attribution, emission factors and regulation. Atmospheric Environment 42: 2895–2921.
- Rospondek, M.J., L. Marynowski, A. Chachaj, M. Góra. 2009. Novel aryl polycyclic aromatic hydrocarbons: Phenylphenanthrene and phenylanthracene identification, occurrence and distribution in sedimentary rocks. Organic Geochemistry 40: 986–1004.
- Rudnicka, K., M.K. Łuczyński, M. Góra 2012. Genotoxicity of cyclopentha[c]phenanthrene and its two derivatives based on *in vitro* micronucleus test. Environmental Biotechnology 8: 63-66.
- Sakai, M., D. Yoshida, S. Mizusaki. 1985. Mutagenicity of polycyclic aromatic hydrocarbons and quinones on *Salmonella* Typhimurium TA97. Mutation Research 156: 61-67.
- Sims, P. 1970. Qualitative and quantitative studies on the metabolism of a series of aromatic hydrocarbons by rat-liver preparations. Biochemical Pharmacology 19: 795-818.
- Singh, V.K., D.K. Patel, S. Ram, N. Mathur, M.K. Siddiqui, J.R. Mehari. 2008. Blood levels of polycyclic aromatic hydrocarbons in children of Lucknow, India. Archives of Environmental Contamination and Toxicology 54: 348–354.
- U.S. EPA (U.S. Environmental Protection Agency). 1987. Health and Environmental Effects Profile for Phenanthrene. Prepared by the Environmental Criteria and Assessment Office, Office of Health and Environmental Assessment, U.S. Environmental Protection Agency, Cincinnati, OH, for the Office of Solid Waste and Emergency Response. http://rais.ornl.gov/guidance/epa\_hh.html, accessed April 2014.
- U.S. EPA (U.S. Environmental Protection Agency). 1988. Drinking Water Criteria Document for Polycyclic Aromatic Hydrocarbons (PAHs). Prepared by the Environmental Criteria and Assessment Office, Office of Health and Environmental Assessment, U.S. Environmental Protection Agency, Cincinnati, OH, for the Office of Drinking Water. 28p. ECAO-CIN-D010
- U.S. EPA (U.S. Environmental Protection Agency). 1993. Integrated Risk Information System (IRIS). Environmental Criteria and Assessment Office, Office of Health and Environmental Assessment, Cincinnati, OH. Emergency Response. http://rais.ornl.gov/guidance/epa\_hh.html, accessed April 2014.
- Wood, A.W., R.L. Chang, W. Levin, D.E. Ryan, P.E. Thomas, He Duck Mah, J.M. Karie, H. Yagi, D.M. Jerina, A.H. Conney. 1979. Mutagenicity and tumorgenicity of phenanthrene and chrysene epoxides and diol epoxides. Cancer Research 39: 4069-4077.
- Wrobel, C., T. Reinhardt. 2003. Review of Potential Air Emissions from Burning Polyethylene Plastic Sheeting with Piled Forest Debris. URS Corporation, Seattle, WA.