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Sources of dissolved organic carbon in small volcanic mountainous tropical rivers, examples from Guadeloupe (French West Indies)

Emily Lloret, Céline Dessert, Heather L. Buss, Carine Chaduteau, Sylvain Huon, Patrick Alberic, Marc F. Benedetti

Abstract

In the tropical zone, small watersheds are affected by intense meteorological events. These events play an important role in the erosion of soils and therefore on the sources of organic carbon in small tropical rivers. We studied the geochemistry of two soils on Basse-Terre Island (French West Indies, FWI): ferralitic soil and Andosol. The two studied soils are very similar in terms of soil organic matter (SOM) and soil solution parameters. The total organic carbon (TOC) and total nitrogen (TN) contents vary between 1.7 and 92 g kg⁻¹ and between 0.1 and 5.5 g kg⁻¹, respectively, with the highest concentrations observed in the topsoil. The C/N ratios are relatively constant throughout the soil profiles (ca. 12). The carbon isotopic composition of SOM varies between -27.3‰ and -22.7‰ and presents an enrichment with increasing depth of soil profiles. Dissolved organic carbon (DOC) concentrations in soil solutions, varying from 3.2 to 91.3 mg L⁻¹, are similar for the both extraction used in lab (with milliQ water and Ca(NO₃)₂) but are higher than those measured in soil solutions sampled from lysimeters (0.65–1.46 mg L⁻¹). The isotopic compositions of DOC obtained by extractions and SOM are comparable, with δ¹³C values ranging from -28.6‰ to -25.8‰. The DOC sampled from lysimeters is systematically depleted in ¹³C compared to DOC obtained by extractions, with δ¹³C values of -33.8‰ to -30.6‰. The enrichment of δ¹³C of SOM through the soil profiles is either consistent with the carbon isotopic fractionation of SOM by decomposing organisms, or the differential mineralization of both labile and stable carbon stocks in soils. DOC concentrations in stream waters vary between 0.46 and 5.75 mg L⁻¹, and are generally lower during low water level than floods. The isotopic compositions of DOC in the rivers range from -38.9‰ to -27.2‰, with δ¹³C values, which are more depleted in ¹³C during low water level than flood events. The δ¹³C_{DOC} of water river samples and soil solutions obtained by extraction and collected with lysimeters demonstrates that the DOC in rivers derives essentially from both the lixiviation of the soil surface layers during floods and groundwater flow during low water levels. Lixiviation of soil surface layers can be boosted by significant increases of intensity and duration of meteorological events and can strongly favor the release of surface soil organic matter in rivers and the impoverishment in nutrients of soil surface layers.

Keywords : Organic carbon sources; SOM; Soil solution; DOC; Small tropical rivers

1. Introduction

Rivers represent the major export of carbon from continents to oceans, with a global carbon flux (dissolved and particulate, organic and inorganic carbon) of 400–900 Mt yr⁻¹ (Hedges, J.I., et al., 1997, Schlünz, B. and Schneider, R.R., 2000 and Aitkenhead-Peterson, J.A., et al., 2003). The organic contribution (dissolved and particulate) represents about 40% of the global continental carbon flux (Hedges, J.I., et al., 1997, Schlünz, B. and Schneider, R.R., 2000 and Aitkenhead-Peterson, J.A., et al., 2003). Locally, the contribution of organic carbon could be higher than that of inorganic carbon. This is notably observed for large tropical river systems (Huang et al., 2012), which have particular hydrological processes (Wohl et al., 2012), for which the dissolved form can represent more than 60% of the total organic carbon export (Schlesinger, W.H. and Melack, J.M., 1981, Hope, D., et al., 1994, Ludwig, W., et al., 1996a and Ludwig, W

1996a, 1996b). For small rivers directly connected to ocean basins with limited floodplains and mangroves, particulate organic carbon (POC) export is usually higher than dissolved organic carbon (DOC) export (Alvarez-Cobelas et al., 2012). Moreover, in mountainous tropical islands (e.g., Taiwan, New Zealand, Guadeloupe), the majority part of the organic carbon export occurs during flood events (Hilton et al., 2008a; Fujii et al., 2009; Bass et al., 2011; Lloret et al., 2013). Accordingly, the sources of organic carbon exported from the catchments vary with the hydrological stage, in particular for small rivers which can be sustained by quick or ground flow (Lloret et al., 2011).

The major sources of organic carbon in rivers are either autochthonous (DOC is directly produced in the river by biological activity) or allochthonous (detrital inputs) (e.g., Aitkenhead-Peterson et al., 2003; Mulholland, 2003; Finlay and Kendall, 2007). In the case of allochthonous sources, the main input is due to soil erosion. During erosive rainfall events, organic carbon contained in soil organic matter (SOM) can be intensively either leached or eroded and transferred to aquatic ecosystems in dissolved and particulate forms, respectively (Lal, 2004). During low water stages, DOC input to rivers can also be completed by soil water infiltration. With 1400–1500 Gt of carbon, the SOM pool is one of the major carbon surface reservoirs (Schlesinger, 1977; Gregory et al., 1999) and stores twice the quantity of carbon contained in the vegetation. Accordingly, changes in SOM abundance and composition have important effects on the global carbon cycle as well as on soil ecosystems.

The organic matter amounts in soils are essentially regulated by the biological production and the decay rates of organic compounds. Plant residues (litter) falling onto soil surface are gradually decomposed through physical fragmentation, faunal and micro-faunal interactions

(decomposers), mineralization, and humus formation (Baudin et al., 2007). The overall processes induce variations in the concentration and in the isotopic composition of total organic carbon (TOC) through the soil profile. The decomposition of litter and SOM induces a differential decrease of carbon and nitrogen contents in soils. As they are not recycled at the same rate, the C/N ratio also varies through the soil profile. They are typically higher in litterfall and decrease in SOM during humification (Snowdon et al., 2005; Boström et al., 2007; Yang et al., 2010). Selective degradation changes the stable C isotope ratios of SOM. Surface litter remains more depleted in ^{13}C than SOM (Lichtfouse et al., 1995) that is progressively enriched in ^{13}C with soil depth (Balesdent and Mariotti, 1996; Amiotte-Suchet et al., 1999, 2007). Highlighting the variations in concentrations and isotopic compositions of TOC (in solid and solute fractions) in soils could help to track the origin of organic carbon in small tropical rivers with respect to their hydrological stages (low or high water stages).

In this way, Guadeloupean soils and rivers provide an ideal opportunity to investigate the interactions between soil and river pools and fluxes (e.g., Rivé et al., 2013). The island of Guadeloupe, which presents surface soil horizons thoroughly enriched in organic matter (10–15%; Colmet-Daage and Bernard, 1979), is frequently impacted by high intensity meteorological events such as tropical storms and cyclones (Zahibo et al., 2007) that accentuate soil erosion (Waterloo et al., 2006; Dawson et al., 2008; Hilton et al., 2008b; Lloret et al., 2011; Allemand et al., 2014). In order to determine the DOC sources according to different hydrological stages of Guadeloupean rivers, the present study focuses on the characterization (especially concentrations, and isotopic compositions) of organic carbon in SOM and soil solutions through soil profiles of two small watersheds with contrasted settings.

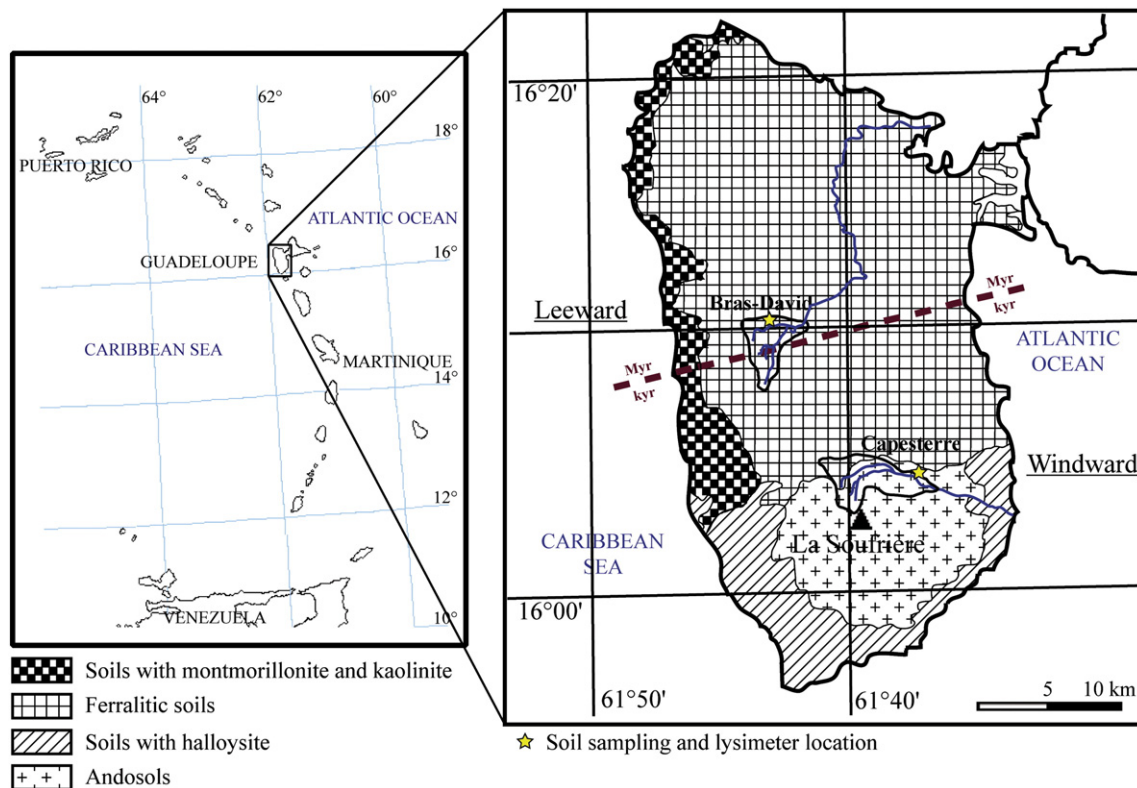


Fig. 1. Lesser Antilles map with the location of the Guadeloupe, French West Indies. The inset map shows Basse-Terre Island with simplified pedology and the location of the two studied watersheds (Bras-David and Capesterre).

2. Study areas

Guadeloupe is a part of the Lesser Antilles volcanic arc (Fig. 1), generated by the subduction of the North American plate beneath the Caribbean plate. The volcanic island of Basse-Terre, part of the Guadeloupe archipelago, belongs to the central segment of the arc (e.g., Feuillet et al., 2011). Basse-Terre Island is characterized by old geological formations in its northern part and increasingly recent ones near the present-day volcano: La Soufrière, which is located in its southern part (Samper et al., 2007, 2009). Our study is focused on two watersheds with different soil and river characteristics (Table 1), one located in the northern part of the island and another in the southern part (Fig. 1). These two watersheds are located in the National Park of Guadeloupe, and are monitored within the framework of the French Critical Zone Observatory OBSERA (INSU-CNRS) observatory devoted to the study of weathering and erosion in the French West Indies. This observatory belongs to the French network of monitored watersheds (RBV supported by INSU-CNRS and AllEnvi).

2.1. Climate and hydrology

Basse-Terre Island is characterized by a wet tropical climate, with a mean annual temperature and humidity of ca. 23 °C and 75%, respectively (Plaisir et al., 2003). The average annual precipitation of the last decade ranges from 1500 to 7000 mm yr⁻¹, depending on the topography (Clergue et al., 2015; Dessert et al., 2015). Climate is characterized by two seasons: a dry season from January to June and a wet and rainy season from July to December (representing about 60% of the total annual precipitation, MétéoFrance data). During the wet season, hurricanes and tropical depressions produce heavy rainfall events (Zahibo et al., 2007), which play a major role for soil weathering and erosion. The spatial distribution of precipitation is strongly influenced by easterly winds and topography. The Bras-David and the Capesterre watersheds are located on the windward coast influenced by easterly winds and high annual precipitation, between 2000 and 4300 mm yr⁻¹ (Météo France data). Runoff represents about 60% of the precipitation. The discharges of the studied rivers are monitored by DREAL (French Water Survey agency; <http://www.hydro.eaufrance.fr>).

2.2. Vegetation cover, geology, and soils

Located in the National Park of Guadeloupe, the vegetation cover of the two studied watersheds is mainly tropical rainforest and is dominated by C3 photosynthetic pathway plants, including *Dichapetalaceae* (*Tapura latifolia*), *Euphorbiaceae* (*Richeria grandis*, *Amanoa caribaea*), *Burseraceae* (*Dacryodes excelsa*), *Sapotaceae* (*Pouteria pallida*), and *Annonaceae* (*Guatteria caribaea*) (Rousteau et al., 1994; Rousteau, 1996).

The geological basement of the Bras-David watershed is composed of Pleistocene andesitic and dacitic (Samper et al., 2007, 2009), and covered by thick ferrallitic soils (> 15 m; Colmet-Daage and Bernard, 1979). These soils were previously studied (Henriet et al., 2008; Buss et al., 2010; Sak et al., 2010; Opfergelt et al., 2012; Clergue et al., 2015) and consist of highly weathered volcanoclastic debris flows containing rocky clasts at various stages of weathering. Clays, dominantly halloysite, represent about 75% wt of the mineralogical content and nonclays are almost entirely Fe(III)-hydroxides and quartz/cristobalite. The Capesterre watershed is underlain by andesitic rocks linked to late Pleistocene volcanism (Samper et al., 2007). Because of the steep slopes of the recent volcanic rocks (49%; Plaisir et al., 2003), soils in this region are dominated by thin Andosols (<1 m; Colmet-Daage and Bernard, 1979; Cattani et al., 2007; Cabidoche et al., 2009).

3. Materials and methods

3.1. Sample collection

3.1.1. Litter and soil material sampling

Two soil cores, representative of the two main soil types (ferrallitic and Andosol), were sampled (Fig. 1) to analyze the TOC and total nitrogen (TN) concentrations and the stable carbon isotope composition ($\delta^{13}\text{C}$) of litter and SOM. The soil profiles were sampled (during the installation of lysimeters) every ~15 or ~30 cm (1/2 or 1 ft), with a hand-auger at 0.15–0.91 m and 0.15–12.50 m soil depths for Capesterre and Bras-David watersheds, respectively. In addition, larger amounts of topsoil layers (from 0 to 30 cm) were also sampled every 5–10 cm in pits. All soil samples were air-dried, sieved to 2 mm, split with a sample divider, and homogenized by grinding at the USGS Menlo Park Campus (California, USA). Litter material from the two watersheds was collected in plastic bags and was rapidly air-dried to avoid mould development, then ground.

3.1.2. Soil solution sampling

Vadose zone pore waters were collected approximately monthly during 2007 and 2008 from 5 cm diameter nested porous-cup suction water samplers (Soil Moisture Inc., Santa Barbara, CA) (referred to as lysimeters hereafter). These lysimeters were placed in hand-augered holes at depths from 0.15 to 0.91 m and from 0.15 to 12.50 m for Capesterre and Bras-David sites, respectively. Soil solution characteristics were obtained by extracting soils with milliQ water following the procedure in Bardy (2008) or using a $\text{Ca}(\text{NO}_3)_2$ solution at 4.6×10^{-3} M according to Zsolnay (2003). The solution/soil ratio and the time of extraction are reported in Table 2. Soil solution samples used for DOC concentrations and $\delta^{13}\text{C}$ measurements were filtered through glass fiber filters (GF/F Whatman® by Schleicher & Schuell cutoff 0.7 μm) and acidified with concentrated H_3PO_4 (85%) in pre-cleaned and pre-combusted glass bottles, and stored at 4 °C in the dark.

Table 1
Watershed characteristics.

Sites	Latitude	Longitude	Area, km ²	Elevation, m	Age of bedrock ^a , Myrs	Slopes ^b				Vegetation ^c		
						0–24%	25–48%	49–99%	>99%	Thickets	Altimountain Forest	Rainforest
						% surface				% surface		
Bras-David	N16°10'33.6"	W61°41'34.8"	11.3	228–1088	1.460	38	48		0		35	65
Capesterre	N16°04'18.0"	W61°36'34.1"	16.6	200–1342	0.554	18	32	33	5	33	39	29

^a Samper et al. (2007).

^b Plaisir et al. (2003)

^c Rousteau et al. (1994); Rousteau (1996).

Table 2
Characteristics of milliQ water and Ca(NO₃)₂ extractions.

Extraction	milliQ water Bardy (2008)	Ca(NO ₃) ₂ Zsolnay (2003)
Solution/soil (mL g ⁻¹ of dry soil)	3	2
Extraction time	12 h	10 min
Centrifugation time	20 min	10 min

3.1.3. River sampling

Pristine water samples were collected upstream of any anthropogenic activities at the outlet of each watershed (Fig. 1). Surface water was sampled manually from 2007 to 2010 at different hydrological stages corresponding to low water and flood levels (Lloret et al., 2011, 2013; Dessert et al., 2015). An automatic water sampler, ISCO-6712, was set up on the Capesterre River to sample the largest number of extreme meteorological and hydrological events. It allowed the sampling of 27 flood events, including 5 extreme events during which up to 24 samples were taken (from every 15 min to 2 h). Water samples used for the measurement of DOC concentrations and δ¹³C measurements were filtered, acidified and stored as for soil solution sampling.

3.2. Analytical methods

TOC and TN concentrations of litter and SOM were measured using an elemental analyzer (Thermo Scientific Flash 2000), after checking that samples are carbonates free (controlled under acid vapor). The detection limit was <5%.

The δ¹³C_{TOC} values were measured twice:

- at the Equipe de Géochimie des Isotopes Stables (LGIS, IGP, France), using an IR-MS Finnigan MAT 253 coupled with an elemental analyzer Thermo Flash EA 1112 Series.
- at the Institut d'Ecologie et des Sciences de l'Environnement de Paris (iEES, UPMC, France), using an IR-MS Sira 10 Fisons coupled with an elemental analyzer Carlo-Erba Na-1500 NC.

Results are reported in per mil (‰) relative to the Pee Dee Belemnite (PDB) standard with a precision of 0.3‰. The results of the two labs were very similar (±0.2‰), and data reported in this study are averages of these two measurements.

The DOC concentrations from soil solutions or river waters were measured using a Shimadzu TOC-VCSH analyzer (Sugimura and Suzuki, 1988). The detection limit was 0.24 mg L⁻¹ and the precision was 2%. The δ¹³C_{DOC} was measured at the "Institut des Sciences de la Terre d'Orléans" (ISTO, France) with an IR-MS Delta V Advantage coupled with a LC-Isolink interface (both Thermo Scientific) and an HPLC system serving as a pump for the carrier flow. Aliquots of 100 μL of filtered and acidified sample water were directly injected in bulk mode and monitored in continuous flow (Albéric et al., 2010). The standards used for the δ¹³C_{DOC} measurements were the internal standard from ISTO (NaHCO₃, -4.3‰), a benzoic acid (-25.7‰), USGS-40 (L-glutamic acid, -26.389‰), IAEA-C8 (oxalic acid, -18.3‰), and IAEA-C6 (sucrose, -10.8‰). The precision of DOC isotopic compositions was 0.3‰.

Spectrophotometric analyses of liquid samples were conducted with a dual beam Evolution 600 UV/Vis Thermo Scientific spectrophotometer. An aliquot of each acidified water sample collected for the DOC measurements was used for the spectroscopic characterization of the dissolved organic matter (DOM). Samples were placed in a 1 cm quartz window "cuvette" and scanned from 200 to 600 nm, with 1 nm resolution (Chin et al., 1994). The percentages of aromatic carbon and the hydrophobic fraction of DOM were estimated from the UV-Vis absorbance

data, with empirical relationships (Chin et al., 1994). Different wavelengths were selected (254, 270, and 280 nm) to highlight the markers of organic matter. The absorbance at these different wavelengths defined as SUVA (Specific UV absorbance) is calculated with the following formula:

$$SUVA_i = \frac{\text{Absorbance at wavelength } i \text{ (cm}^{-1}\text{)}}{\text{DOC concentration (mg L}^{-1}\text{)}} \quad (\text{i.e. } SUVA_{254}, SUVA_{270}, SUVA_{280}) \quad (1)$$

4. Results

4.1. Organic carbon and total nitrogen distributions through soil profiles

4.1.1. TOC and TN in litter and SOM

For the Bras-David site, the mean organic carbon and nitrogen contents in the litter layers are 395 ± 3 g kg⁻¹ and 12.8 ± 0.2 g kg⁻¹, respectively, corresponding to a C/N (or TOC/TN) ratio of 30.8 ± 0.6 (Table 3). For Capesterre, they averaged 409 ± 6 g kg⁻¹ and 15.9 ± 0.1 g kg⁻¹, respectively, corresponding to a C/N ratio of 25.8 ± 0.5 (Table 4). The TOC content of the Bras-David ferralitic soil decreases within the top two meters ranging 92 to 1.7 g kg⁻¹ (Table 3). For soil depths between 2.0 and 12.5 m, the TOC content is <3.7 g kg⁻¹ (range: ca. 0.9–3.7). Similarly, TN contents are higher within the top two meters (range: ca. 0.1–5.5 g kg⁻¹), a decrease to 0.1 g kg⁻¹ for deeper horizons. Consequently, the C/N ratio remains relatively constant through the soil profile (ca. 12.7 ± 3.2) (Table 3). For the Capesterre soil, the highest TOC and TN contents are found in the topsoil (0.00–0.20 m depth) where they reached more than 23.6 g kg⁻¹ and 2.3 g kg⁻¹, respectively, then they decrease progressively to 5.6 g kg⁻¹ and 0.4 g kg⁻¹, respectively, in 0.20–0.91 m soil horizons (Table 4). As in the Bras-David ferralitic soil, the C/N ratio in the Capesterre Andosol remains constant through the soil profile (ca. 11.4 ± 1.4) (Table 4).

The TOC and TN stocks for the upper 30 cm were calculated from concentrations (c), dry densities (d = 0.650 and 0.875 10³ kg m⁻³ for 0–15 cm and 15–30 cm, respectively; Buss et al., 2010) and thicknesses (t) of each horizon, using the following formula:

$$\text{Stock} = c \times d \times t \quad (2)$$

TOC and TN stocks in the Bras-David ferralitic soil are almost 11,600 ± 7500 and 800 ± 450 Mg km⁻², respectively, and in the Capesterre Andosol are almost 11,200 ± 6000 and 1000 ± 500 Mg km⁻², respectively.

The carbon isotopic composition of SOM for the Bras-David ferralitic soil and the Capesterre Andosol varies between -27.3‰ and -24.2‰ (Tables 3 and 4) and is representative of degraded and humified ¹³C-enriched organic matter derived from trees and understory C3 photosynthetic pathway plants (Deines, 1980; Balesdent et al., 1993) that compose the Guadeloupean rainforest. Additional evidence is pictured by the even more ¹³C-depleted (less degraded) composition of litterfall for both soils (ca. -29.0‰ to -29.7‰, Tables 3 and 4).

4.1.2. DOC in soil solutions

The DOC concentrations in soil solutions obtained by milliQ water and Ca(NO₃)₂ extractions are similar and follow the same trends through the two soil profiles (Tables 3 and 4). For the Bras-David ferralitic soil, the DOC concentrations decrease with depth from 91.3 to 5.2 mg L⁻¹ within the upper two meters and remain constant below 2 m with a value of 5.8 ± 1.5 mg L⁻¹ (Table 3). For the Capesterre Andosol, the DOC concentrations range from 3.2 to 28.9 mg L⁻¹ (Table 4), exhibiting the highest values within the upper 30 cm and decreasing with depth.

The DOC concentrations in soil solutions sampled from lysimeters are lower than those obtained by soil extractions, but the trends through the two soil profiles are similar for both techniques (Tables 3

Table 3

Particulate concentrations (TOC, TN), C/N ratios, and isotopic compositions of TOC ($\delta^{13}\text{C}_{\text{TOC}}$) for soil solid fraction, and dissolved concentration (DOC), isotopic composition of DOC ($\delta^{13}\text{C}_{\text{DOC}}$) and percentage of aromaticity for liquid fractions obtained by lysimeters and lab extractions (water and $\text{Ca}(\text{NO}_3)_2$) for Bras-David soil.

Depth, cm	Solid fraction				Liquid fraction								
	TOC, g kg ⁻¹	TN, g kg ⁻¹	C/N	$\delta^{13}\text{C}_{\text{TOC}}$, ‰	Lysimeters			Water extractions			Ca(NO ₃) ₂ extractions		
					DOC, mg L ⁻¹	$\delta^{13}\text{C}_{\text{DOC}}$, ‰	Aromaticity, %	DOC, mg L ⁻¹	$\delta^{13}\text{C}_{\text{DOC}}$, ‰	Aromaticity, %	DOC, mg L ⁻¹	$\delta^{13}\text{C}_{\text{DOC}}$, ‰	
Litter	394.50	12.80	30.8	-29.7									
0-5	91.92	5.48	16.8									77.92	
5-10	48.66	3.91	12.5										
10-20	17.95	1.26	14.3										
20-25	17.29	1.23	14.1										
15	57.59	4.01	14.4	-27.3	1.62	-31.4	21.7	91.31	-26.4	9.2			-28.6
30	32.61	2.30	14.2	-27.2	0.95	-33.1	32.1	55.81	-26.3	7.7		57.77	-28.2
61	8.79	0.53	16.5	-26.3	1.69	-31.2	17.6	8.66	-27.3	9.9		12.14	-25.8
91	7.71	0.49	15.6	-26.6	0.88	-33.1	18.0	8.25	-28.4	11.8		10.62	-26.7
122	3.96	0.24	16.8	-26.2	0.69	-32.6	24.2	6.81	-27.8	12.6		7.94	-26.5
152	2.32	0.15	15.0	-26.5	0.61	-33.8	31.7	6.50	-27.9	12.9		7.50	-26.6
183	1.72	0.13	13.4	-25.6	0.77	-33.0	24.0	3.69	-27.8	19.5		4.66	-26.0
213	1.89	0.13	14.2	-25.7				5.21		13.7		5.90	-26.6
244	3.71	0.24	15.6	-26.5				6.76		11.4		6.73	-25.8
274	2.41	0.16	14.6	-26.8	0.60	-33.2	23.3	5.88	-27.4	13.0		6.82	-25.9
305	3.34	0.36	9.3	-25.8				3.95	-27.9	15.8		5.50	-26.6
344					0.60	-33.2	24.0						
366	1.68	0.13	13.0	-24.8				4.35	-28.1	18.1		4.95	-26.6
427	2.93	0.32	9.1	-25.9				5.32	-28.2	14.8		4.52	-27.1
457					0.62	-33.8	23.2						
488	3.72	0.24	15.5	-24.6				4.72	-28.1	15.8		4.34	-26.8
549	1.91	0.15	13.0	-24.7				7.04		12.8		5.55	-27.1
610	1.29	0.14	9.4	-26.5				6.85	-28.2	13.2			-27.0
671	1.72	0.18	9.7	-25.8				5.44		15.3		6.73	-26.9
732								4.36		14.5		4.22	-26.9
792	1.30	0.13	9.9	-25.7				2.92		20.2		8.33	-27.6
823					0.66	-32.9	28.2						
853	1.73	0.18	9.4	-25.8				4.45	-28.2	17.6		5.32	-26.9
914	1.41	0.14	10.4	-26.4				4.56		15.3		7.01	-27.2
975	0.88	0.09	9.6	-24.7				3.09	-28.1	17.8		5.64	-27.2
1036	1.14	0.10	11.8	-25.1				5.67		13.2		6.12	-27.4
1097	1.00	0.10	9.6	-27.3				5.41		13.0		7.32	-27.6
1158	0.99	0.10	10.3	-24.2				4.84		14.4		6.13	-26.8
1219	2.51	0.34	7.3	-26.3				6.92		11.9		7.71	-27.2
1250	2.70	0.31	8.7	-25.4	0.51		32.6	6.65	-28.2	13.5		10.69	-27.5

and 4). For the Bras-David ferrallitic soil, the DOC concentrations in soil solutions sampled by lysimeters decrease with depth from 1.46 to 0.65 mg L⁻¹ within the upper two meters and remain constant (ca. 0.69 ± 0.08 mg L⁻¹) below 2 m (Table 3). For the Capesterre Andosol, the DOC concentrations in soil solutions decrease with depth from 1.45 to 0.88 mg L⁻¹ (Table 4).

For the two soil profiles, the isotopic compositions of DOC obtained by extractions and SOM are comparable, with $\delta^{13}\text{C}$ values ranging from -28.6‰ to -26.3‰ and from -28.3‰ to -25.8‰ for the soil solutions of Bras-David ferrallitic soil and the Capesterre Andosol, respectively

(Tables 3 and 4). The DOC sampled from lysimeters is, however, systematically depleted in ¹³C compared to DOC obtained by extractions, with $\delta^{13}\text{C}$ values of -33.8‰ to -31.2‰ and -33.1‰ to -30.6‰ for the Bras-David ferrallitic soil and the Capesterre Andosol, respectively (Tables 3 and 4).

The percentage of aromatic organic carbon, or “aromaticity” of the dissolved organic matter (DOM) was calculated by averaging the values following the relationships for various DOM types from Chin et al. (1994); Peuravuori and Pihlaja (1997), and Weishaar et al. (2003). The DOM is more aromatic in soil solution sampled from lysimeters

Table 4

Particulate concentrations (TOC, TN), C/N ratios and isotopic compositions of TOC ($\delta^{13}\text{C}_{\text{TOC}}$) for soil solid fraction, and dissolved concentration (DOC), isotopic composition of DOC ($\delta^{13}\text{C}_{\text{DOC}}$), and percentage of aromaticity for liquid fractions obtained by lysimeters and lab extractions (water and $\text{Ca}(\text{NO}_3)_2$) for Capesterre soil.

Depth, cm	Solid fraction				Liquid fraction								
	TOC, g kg ⁻¹	TN, g kg ⁻¹	C/N	$\delta^{13}\text{C}_{\text{TOC}}$, ‰	Lysimeters			Water extractions			Ca(NO ₃) ₂ extractions		
					DOC, mg L ⁻¹	$\delta^{13}\text{C}_{\text{DOC}}$, ‰	Aromaticity, %	DOC, mg L ⁻¹	$\delta^{13}\text{C}_{\text{DOC}}$, ‰	Aromaticity, %	DOC, mg L ⁻¹	$\delta^{13}\text{C}_{\text{DOC}}$, ‰	
Litter	408.60	15.90	25.8	-29.0									73.97
0-5	74.29	6.19	12.0										
5-10	53.86	5.14	10.5										
10-15	39.05	3.65	10.7										
15-20	23.60	2.34	10.1										
20-30	20.55	1.92	10.7										
15	25.50	2.64	9.7	-26.6	1.75	-30.6	16.4	28.90	-26.9	7.9		28.38	-25.8
30	26.85	2.54	10.6	-26.9	0.78	-32.9	24.3	24.02	-27.7	7.2		28.02	-26.5
46	10.29	0.92	11.2	-26.4				7.40		9.3		5.88	-26.4
61	7.39	0.61	12.1	-25.4	0.58	-33.1	32.5	6.13	-28.2	14.2		4.59	-26.2
76	5.83	0.40	14.6	-25.2				4.79		11.6		2.95	-26.5
91	5.63	0.44	12.9	-25.7	0.76	-32.8	19.3	6.35	-28.3	11.8		3.24	-26.4

Table 5 Dissolved concentration (DOC), isotopic composition of DOC ($\delta^{13}\text{C}_{\text{DOC}}$), and percentage of aromaticity of DOC in stream waters of Bras-David and Capesterre rivers during low water and flood levels from 2007 to 2010 (Lloret et al., 2011, 2013).

Sites	Low water level			Flood		
	DOC, mg L ⁻¹	$\delta^{13}\text{C}_{\text{DOC}}$, ‰	Aromaticity, %	DOC, mg L ⁻¹	$\delta^{13}\text{C}_{\text{DOC}}$, ‰	Aromaticity, %
Bras-David	N = 46 0.46 to 5.31	N = 5 -38.9 to -30.4	N = 26 18.3 to 100.4	N = 8 0.69 to 3.23	N = 4 -30.9 to -29.4	N = 4 15.8 to 73.7
Capesterre	N = 175 0.46 to 3.75	N = 25 -34.4 to -27.3	N = 6 33.1 to 54.4	N = 243 0.96 to 5.75	N = 20 -30.5 to -27.2	N = 2 35.3 to 39.1

than those obtained by milliQ water extractions. In the lysimeter samples, the aromaticity is 18% to 33% and 16% to 33% for the Bras-David ferrallitic soil and the Capesterre Andosol, respectively. In milliQ water extractions, the aromaticity is 8% to 20% and 7% to 14%, for the Bras-David ferrallitic soil and the Capesterre Andosol, respectively (Tables 3 and 4). Similarly, we have also observed that the percentage of hydrophobic organic carbon (calculated following methods from Chin et al. (1994), and Weishaar et al. (2003)) for the two soil profiles is more significant for solutions collected from lysimeters than from extractions.

4.2. DOC in river waters

The DOC concentrations for the two studied Guadeloupean rivers were reported in studies of Lloret et al. (2011, 2013) and are summarized in Table 5. These rivers exhibit comparable DOC concentrations, between 0.46 and 5.75 mg L⁻¹. During low water level, the mean DOC concentrations are 0.76 ± 0.28 mg L⁻¹ and 1.40 ± 0.60 mg L⁻¹ for the Bras-David River and the Capesterre River, respectively. During flood events, the mean DOC concentrations are similar for both rivers, with 2.71 ± 1.37 mg L⁻¹ for the Bras-David River and 2.24 ± 0.66 mg L⁻¹ for the Capesterre River.

The isotopic compositions of DOC in the rivers range from -38.9‰ to -27.2‰, with $\delta^{13}\text{C}$ values, which are more depleted in ¹³C during low water level than flood events (Table 5).

For both rivers, the aromaticity ranges from 18.3% to 100.4%, but no differences were observed between low water level and floods (Table 5). Again, the percentage of hydrophobic organic carbon for both rivers is not related to hydrological levels, and ranges from 50.6% to 139.0%.

5. Discussion

The following discussion will first focus on the characteristics of the Guadeloupean SOM, and then on the comparison between soil solutions sampled by lysimeters and obtained by extractions. Finally, we will address the question of the sources of dissolved organic carbon according to the hydrological stages.

5.1. Production and evolution of SOM in volcanic tropical soils

For both sites, litter exhibits rather similar TOC (ca. 400 g kg⁻¹) and TN (ca. 14.5 g kg⁻¹) contents (Tables 3 and 4), which are close to those measured in other tropical forests (Schwartz, 1993). The C/N ratios for litter samples (25.8–30.8, Tables 3 and 4) are lower than those usually observed for litterfall in tropical forests (38–43; Schwartz, 1993), indicating that fresh organic matter had likely partly decayed in litterfall before incorporation into soil organic matter.

The TOC and TN contents and the C/N ratios are also similar for both soils (Tables 3 and 4) and are comparable to values measured in pristine forest soils in the East Kalimantan Province, Indonesia (Fujii et al., 2009), humid tropical primary forest soils in Sumatra (van Noordwijk et al., 1997), and ferrallitic soils under tropical rainforest originate from different countries around the world like Brazil, Indonesia, Zaire (Kauffman et al., 1998). Several inferences can be made from these results:

- The decreases of TOC and TN contents and C/N ratios between litter and soils and with soil depth (Tables 3 and 4) reflect the ongoing decomposition or humification processes of organic matter in the soils.
- The low C/N ratios of the two Guadeloupean soils indicate a high degree of decomposition (Kauffman et al., 1998).
- The relatively constant C/N ratios through the two soil profiles indicate that organic carbon and nitrogen are degraded at a similar rate.

For the upper 30 cm of soil profile, the carbon stocks in the Guadeloupean ferrallitic soil and Andosol (11,600 and 11,200 Mg km⁻², respectively) are two times higher than values found for primary wet tropical forests (e.g., Baudin et al., 2007; Ngo et al., 2013). However, they are comparable to estimates given for soils of Martinique (13,000 Mg km⁻²; Blanchart and Bernoux, 2005). Moreover, Martinique and Guadeloupe islands are covered by a tropical rainforest with similar characteristics (e.g., climate, soil cover). These high carbon contents are probably due to soils formed with similar volcanic rocks coming from the same volcanic arc.

Litters are more depleted in ¹³C than the upper 15 cm of both soils (Tables 3 and 4), a relationship which is generally observed worldwide (e.g., Ladyman and Harkness, 1980; Bellanger et al., 2004). During the decay of fresh organic matter (litter), the decomposers preferentially use the ¹²C, which results in ¹³C enrichment of the remaining SOM. Accordingly, SOM is typically ¹³C-enriched by 1.5–4.3‰ relative to homogenous plant and litter constituents (Lichtfouse et al., 1995).

For both soil profiles, the $\delta^{13}\text{C}$ values of SOM ($\delta^{13}\text{C}_{\text{SOM}}$) increased slightly (1.5‰) with depth within the top meter. This ¹³C enrichment is negatively correlated with the TOC content. These results are similar to those observed by Balesdent et al. (1993); Balesdent and Mariotti (1996), for mainland French soils, and for other soils in equilibrium with C3 plants under tropical forests in Amazon basin (Andreux et al., 1990; Desjardins, 1991; Desjardins et al., 1994; Koutika et al., 1997). The observed pattern of ¹³C enrichment over the top meter is typically interpreted as the stable carbon isotopic fractionation of SOM by decomposing organisms (Mariotti and Balesdent, 1990; Mariotti, 1991; Boutton, 1996) and is supported by the correlation between $\delta^{13}\text{C}_{\text{TOC}}$ and LnC for both soil profiles (Fig. 2a). This enrichment is controlled by different factors such as the molecular characteristics of the initial material (Agren et al., 1996), the degree of SOM decay (Blair et al., 1985), and the diversity of decomposers (Andrews et al., 2000). Alternatively, a differential mineralization of two carbon stocks in the soils, one labile and one more stable (Mariotti, 1991), may be involved, as highlighted by the correlation between $\delta^{13}\text{C}_{\text{TOC}}$ and 1/C for both sites (Fig. 2b). Finally, this trend may also result from the leaching of SOM by groundwater flow, which tends to preferentially remove the ¹²C and leave a ¹³C-enriched SOM residue (Kaiser et al., 2001). The extent of carbon isotopic variations of these two Guadeloupean soils may reflect the following: (1) the high clay content typical of tropical soils (Schulte and Ruhayat, 1998; Buss et al., 2010), (2) high net primary productivity, and so high soil microbial activity (Baudin et al., 2007), and (3) high annual precipitation resulting in high percolation fluxes and high adsorption potential (Shen, 1999; Neff and Asner, 2001). For the Capesterre Andosol, the dominant process that can be put forward is likely carbon isotopic fractionation during SOM decay (better correlation coefficient

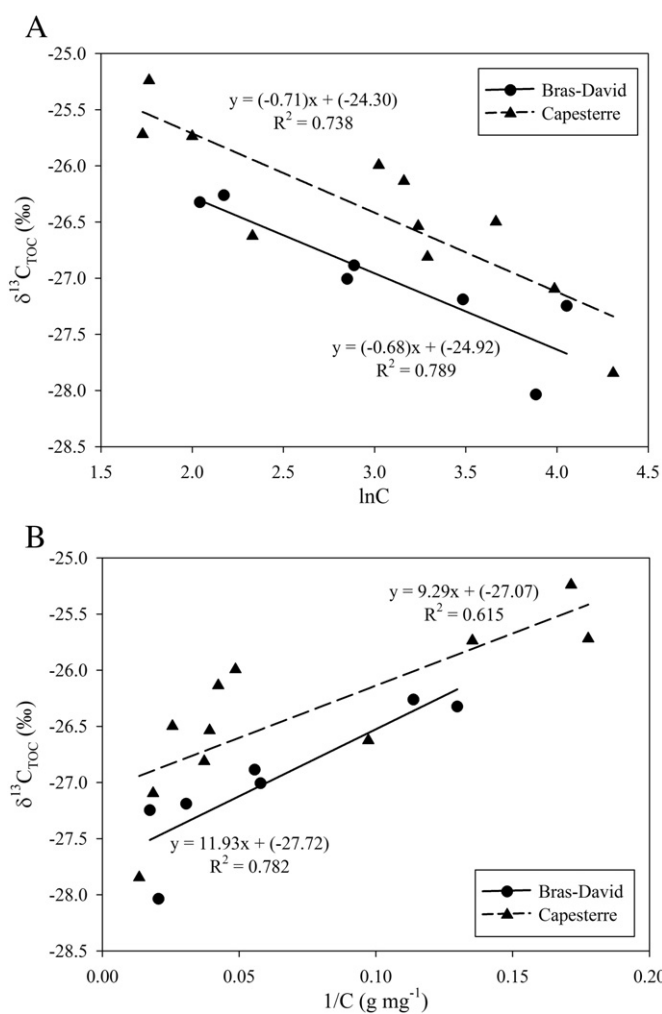


Fig. 2. Carbon isotopic composition of soil organic matter plotted as a function of $\ln C$ (a) and as a function of $1/C$ ratio (b), with C the organic carbon content in soil organic matter.

than for mineralization, Fig. 2). However, for the Bras-David ferrallitic soil, it is hard to solve what is the dominant process because the correlation coefficients (Fig. 2) are quite similar for the two processes of carbon isotopic fractionation origin (SOM decay or mineralization).

5.2. DOC production and cycling in tropical volcanic soils

The DOC concentrations obtained by milliQ water and $\text{Ca}(\text{NO}_3)_2$ extractions are very similar and follow the same decreasing trend as the TOC (Tables 3 and 4). These similarities could be explained by the extraction technique, which allow extracting the same fraction of dissolved organic matter. For the upper 30 cm, the Capesterre Andosol contains less DOC than the Bras-David ferrallitic soil, probably due to the enrichment of organic carbon in the surface layers of the Bras-David ferrallitic soil relative to the Capesterre Andosol. Moreover, ferrallitic soils are generally more acidic than Andosols (Colmet-Daage and Lagache, 1965, average $\text{pH } 4.85 \pm 0.23$ and 5.74 ± 0.19 in soil solution of Bras-David soil and Capesterre soil, respectively) and thereby more effective at dissolving organic carbon. The dissolved organic matter in soil solutions collected by lysimeters could reflect the mobile fraction of organic matter (Zsolnay, 2003), whereas the extractions would extract both the mobile and immobile (complexed or adsorbed) fractions (Zsolnay, 2003), explaining the higher DOC concentrations for the extractions (Tables 3 and 4).

The extracted soil solutions are generally more depleted in ^{13}C than SOM, except in the upper 30 cm (Tables 3 and 4). This likely reflects the

preferential adsorption of carboxylic groups which are enriched in ^{13}C , leaving the soil solution depleted in ^{13}C (Kaiser et al., 2001; Garten et al., 2000). There is, however, a carbon isotopic difference in the DOC between the two types of extraction. For both soils, the extracted milliQ water solutions are generally more depleted in ^{13}C than the extracted $\text{Ca}(\text{NO}_3)_2$ solutions (Tables 3 and 4). This observation shows that different fractions of dissolved organic matter are extracted by these two methods. Conversely, for both sites, increasing $\delta^{13}\text{C}$ of SOM is observed with decreasing $\delta^{13}\text{C}$ of DOC in lysimeter solutions (Tables 3 and 4). This trend was also observed by Ludwig et al. (2000) and attributed either to the preferential decay of labile substances or to the selective adsorption of carbon. During SOM decay, the soil solution produced is depleted in ^{13}C relative to SOM (Agren et al., 1996; Amiotte-Suchet et al., 2007). According to the second process, the preferential adsorption of the carboxylic groups enriched in ^{13}C (Kaiser et al., 2001) would also leave a soil solution depleted in ^{13}C (Garten et al., 2000).

The proportion of aromatic DOC is higher in soil solutions collected by lysimeters than in those extracted with milliQ water (Tables 3 and 4). In fact, dissolved organic matter in soil solutions obtained by milliQ water extraction can be represented by fulvic acids that are enriched in ^{13}C and less aromatic. The dissolved organic matter in soil solutions collected by lysimeters can be represented by humic acids (depleted in ^{13}C and more aromatic) (Nissenbaum and Kaplan, 1972; Flexor and Volkoff, 1977).

5.3. DOC sources in small tropical volcanic rivers

The DOC in stream waters results from the mixing of two sources, which are represented by soil solutions collected from lysimeters and by extractions (Fig. 3). During flood stage (Lloret et al., 2011, 2013), the streams are essentially fed by lixiviation of soil surface layers (represented by extracted soil solutions). During low water levels (Lloret et al., 2011, 2013), the streams are fed by groundwater (represented by soil solution collected from lysimeters).

Lixiviation of soil surface layers can be boosted by significant increases of intensity and duration of meteorological events, as has been

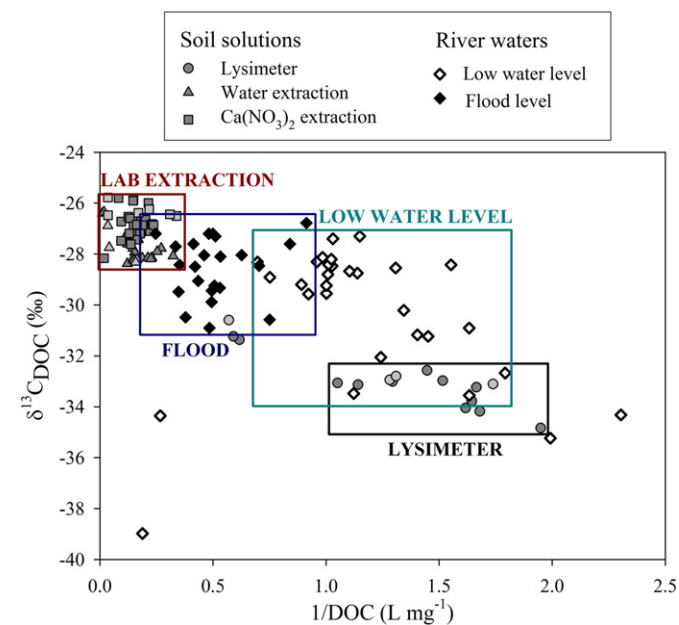


Fig. 3. Carbon isotopic composition of DOC in studied rivers for two hydrological stages (low water level and flood) and in soil solutions sampled from lysimeters and obtained by lab extractions. Soil solutions (lysimeters and lab extractions) are represented by dark grey and light grey symbols for Bras-David and Capesterre, respectively.

observed for several decades (Emanuel, 2005; Webster et al., 2005). This can strongly favor the following:

- the release of top soil organic matter into rivers and increase in the DOC yields, which are similar to large rivers (Lloret et al., 2011, 2013). This terrestrial organic matter can be degraded by microbial organisms, preserved in continental margins and/or transported offshore (Hedges et al., 1997; Benner, 2004).
- the impoverishment of soil surface layers in nutrients can affect the long-term stability of the Guadeloupean tropical rainforest ecosystem.

For instance, the average soil organic carbon residence time is 580 years (Lloret et al., 2013), and an increase of extreme meteorological events will decrease this residence time and consequently release more labile organic matter in oceans.

Due to the geomorphic characteristics of Guadeloupean watersheds (strong relief, direct connection to the oceans), the meteorological events lead to specific hydrological processes as flash floods, which are limited in duration. Consequently, more than 2/3 of the year, these rivers are at the low water level stage (Lloret et al., 2013) and DOC comes from groundwater.

The two river samples, collected during low water levels, which have very low $\delta^{13}\text{C}$ and low 1:DOC ratio (Fig. 3), can be attributed to additional sources, likely primary production in the rivers ($\delta^{13}\text{C}$ of phytoplankton in natural river waters is -35‰ to -25‰, Kendall et al., 2001) and the mineralization of leaves falling directly into the rivers ($\delta^{13}\text{C}$ of leaves is lower than of litter or soil, thus < -25.8‰), respectively. This interpretation is supported by the observation of a biofilm on river stones during the low water periods.

6. Conclusion

This study shows clearly that there are different sources of DOC according to hydrological stages of the Guadeloupean rivers. These sources are deeply related to processes occur in soils. TOC and TN decrease with depth in the bulk soil, whereas the C/N ratio remains relatively constant through the soil profile. The enrichment of $\delta^{13}\text{C}$ of SOM with depth through the soil profile is consistent with the carbon isotopic fractionation of SOM by decomposing organisms, or the differential mineralization of both labile and stable carbon stocks in soils, or the leaching of SOM by groundwater flow.

The $\delta^{13}\text{C}_{\text{DOC}}$ of river water samples and soil solutions obtained by extraction and collected from lysimeters demonstrates that the DOC in rivers derives essentially from both the lixiviation of the soil surface layers during floods and from groundwater during low water levels. The dissolved organic matter in extracted soil solutions, which is represented by fulvic acids, is more enriched in ^{13}C and less aromatic than that from lysimeters, which is represented by humic acids. Another source with very low $\delta^{13}\text{C}_{\text{DOC}}$ and high 1/DOC ratios may also feed the rivers during low water levels. This additional source is likely to be either the internal primary production of the rivers or the mineralization of leaves that fall into the rivers. Further study of the $\delta^{13}\text{C}$ of biofilms formed on river stones during low water periods may help to better constrain this other source.

Therefore, three main sources of DOC in rivers were identified in the particular case of small tropical mountainous watersheds. As it was demonstrated by Lloret et al. (2013), the DOC yields of Guadeloupean rivers are close to those of large tropical rivers, like the Amazon or the Orinoco. Consequently, identification of several DOC sources is a crucial point to better understand global carbon cycle.

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