The variability of juvenile dispersal in an opportunistic raptor

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Abstract

in the second year, 70% decreased their mean distance to the nest, 65% decreased their hourly travelled
distances, and 50% decreased their dislu travelled distances. On the other hand, the Red Kites usually
combined wander The juvenile dispersal of raptors is a crucial stage that stretches from parental independence to the establishment of the first breeding area. Between 2012 and 2020, 44 juvenile red kites *Milvus milvus* from the Spanish breeding population were tagged using GPS telemetry to study their dispersal. Juveniles left the parental breeding area at the end of their first summer and performed wandering movements throughout the Iberian Peninsula, returning to the parental breeding area the following year, repeating the same pattern until they settled in their first breeding area. We analyzed the mean distance from the nest, the maximum reached distances, and the travelled distances (daily and hourly) during the first two years of dispersal and compared them. Despite the high individual variability, variables describing the dispersal movements of juveniles showed a decreasing trend during the second dispersal year: 80 % of individuals reached a shorter maximum distance in the second year, 70% decreased their mean distance to the nest, 65% decreased their hourly travelled distances, and 50% decreased their daily travelled distances. On the other hand, the Red Kites usually combined wandering movements with establishment of temporary settlement areas (TSA). The average duration of settlement in the TSAs was 75 ± 40 days (up to 182 days) and were located at 182 ± 168 km from the nest. In those areas, juveniles used 781.0 \pm 1895.0 km² (KDE 95%). Some of the TSAs were used by several individuals, which suggests that these areas might be good targets for conservation in future management plans

Key words: red kite, landfills, spatial ecology, GPS telemetry, movement ecology, natal dispersal.

The juveniles from both sedentary and migratory birds, after becoming independent from their parents, disperse from their birth places for months or years (Newton 2010). This period, known as "juvenile dispersal" or "natal dispersal", comprises from the departure from the nest to the establishment of the first breeding area, when individuals become adults. Most of the movements performed by birds during this stage seem to occur randomly in all directions. Individuals range from one area to another, exploring territories and residing in temporary settlement areas, where food and clime are appropriate for a short period (Greenwood 1980, Ronce 2007, Newton 2010). Juvenile dispersal, despite being one of the most crucial stages in birds' life cycle (Gadgil 1971, Ferrer 1993, Soutullo *et al.* 2006, Penteriani & Del Mar Delgado 2009, Cadahía *et al.* 2010, Nemček *et al.* 2014, Urios *et al.* 2017), is often poorly known in comparison with breeding and migration events. However, juveniles' survival rates determine the settlement and distribution of later adults in their breeding areas, so high mortality rates during this period may affect population dynamics (Ferrer & Calderón 1990, Newton 1998, Penteriani & Del Mar Delgado 2009). Thus, the study of movements and settlements patterns of juveniles is indispensable for the correct design of management plans.

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1990, Newton 1998, Penteri Opportunistic raptors such as the Red Kite (*Milvus milvus*, Accipitridae) are very interesting species for the study of juvenile dispersal.This species is a partial migratory raptor from Western Palearctic (Aebischer & Scherler 2021; Panuccio, Mellone & Agostini 2021) and has high variability in its spatial strategies. Their breeding populations are mainly found in the west and center of Europe (BirdLife International 2021), but there are also more than 2,000 breeding pairs in Spain (Molina 2015). During the last decade, the Spanish breeding population has slightly increased (15.66% since 2004), but the species is far from being recovered in this country (Molina 2015), where it is categorized as "In Danger of Extinction" (MITECO 2020). Several reasons could explain this complicated situation in the southern limit of Red Kite's distribution. On one hand, they have a generalist diet, being facultative scavengers (García *et al.* 1998), feeding on carrion (Knott *et al.* 2009), so they are especially vulnerable to poison exposure (Monclus *et al*. 2018, Herrero-Villar *et al.* 2021, Viñuela *et al.* 2021). On the other hand, Spain is the limiting region in the southern distribution of the Red Kite (García-Macía *et al.* 2021, 2022), so its climatic conditions may also explain the decrease in the Spanish populations (Seoane *et al*. 2003, Wormworth & Mallon 2006). The rise of temperatures and global change may shift the distribution range of Red Kites towards the north, as it is happening with several species in the last decades (Visser *et al.* 2009, Heath *et al.* 2012, Martín *et al.* 2014).

Despite the importance for its conservation, the dispersal movements of the juvenile Spanish Red Kites are poorly known. This stage has been mainly studied in some European populations based on rings and field observations (Newton *et al.* 1989, Bustamante 1993, Newton *et al.* 1994), but also on GPS telemetry in the last decade (Nemček 2013, Vidal-Mateo 2019, Literák *et al.* 2020, Literák *et al*. 2022), which provides more accurate information (López-López *et al*. 2010, Urios *et al.* 2015, López-López 2016). However, juvenile Red Kites from the Spanish breeding population seem to have distinct characteristics, so there are still many open questions about their dispersal: How do movements evolve? What are the characteristics of their temporary settlement areas? Where are their first breeding areas located after dispersal? Are there sex differences in dispersal distances, as is the case for some Central European populations (Literák *et al*. 2022)? More studies focused on these issues are necessary to improve the knowledge of this species and have good tools for its conservation.

In this study, we analysed GPS telemetry data from 44 individuals of the Spanish breeding population, using dispersal periods from 7 months to 2 years. We hypothesize, first, that juveniles' movements of the Red Kite decrease during the second year of dispersal, includingless distances traveled and shorter mean distance to the nest. This process has been reported in other raptors, both during dispersal (Ferrer 1993) and migration (Mellone *et al.* 2013, García-Macía *et al.* 2021).Second, the juvenile dispersal throughout the Iberian Peninsula will not be sex dependent, because sex differences in the movement ecology of raptors often appear when individuals become adults (Newton *et al.* 1989, Walls & Kenward 1995, Real & Mañosa 2001, Balbontín & Ferrer 2009). Third, juveniles will tend to establish their breeding area near their natal nests. This philopatric behaviour has already been demonstrated in many other raptors (Bustamante 1993, Ferrer 1993, Newton *et al.* 1994, Negro *et al.* 1997, Literák *et al*. 2022). Finally,we hypothesize that there will be a tendency to

establish in TSAs, as often occurs in juvenile raptors (Ferrer 1993, Prommer & Bagyura 2007, Nemček *et al.* 2016, Vidal-Mateo 2019).

Materials and Methods

Area study and tagging

Forty-four chicks from the Spanish breeding population were tagged between 2012 and 2020 in different Spanish provinces (Figure IA), mainly from the center and north of the Iberian Peninsula, trying to follow the distribution of the species in that territory: Madrid (22), León (5), Soria (4), Ávila (3), Zamora (2), Toledo (2), Álava (2), Guipúzcoa (2), Valladolid (1), and Mallorca (1).

Individuals were tagged with GPS transmitters while they were in the nest, unable to fly freely, still able to be caught by hand. The GPS transmitter was fixed to the back of the juvenile with a non-abrasive material, a full body Teflon harness, tied with a cotton thread (Garcelon 1985, Thaxter *et al.* 2014, García *et al.* 2021). During the tagging process, the individuals were weighed and ringed, and a blood sample was taken for sex determination by DNA (Ellegren 1996). The average body weight of the nestlings was 886.6 ± 70.9 g (mean *± SD*; range = 720-1045 g). The transmitters weight was between 2-3% of the bird's body weight, which is within the recommended range (< 5%; Kenward 2001). Birds were returned to the nest within less than 30 min after capture.

We used the following transmitter models: 21-22 g SAKER GPS-GSM (Ecotone Telemetry; *n =* 19), 25 g OrniTrack-E25 GPS-GSM and 30 g OrniTrack-30 GPS-GSM (Ornitela; *n =* 13 and *n =* 8, respectively), PTT-100 30g solar-powered Argos/GPS (Microwave Telemetry Inc.; *n =* 2), and 27 g Anitra GPS-GSM (Anitra; *n =* 1). Different positioning frequencies were programmed in the transmitter models. Ecotone tags obtained locations every 1-2 hours. The Ornitela and Anitra models collected locations every 5-10 minutes. Microwave tags collected positions every 1 to 3 hours. The transmitters started to record locations at sunrise and stop their activity at nightfall, frequently from 7:00 to 19:00, but they were adjusted depending on the hours of daylight in the different seasons.

Sampling

Complete data for the first dispersal year were available for 44 individuals (20 females, 15 males, and 9 individuals with undetermined sex), and for the second dispersal year were available for 21 individuals (13 females, 7 males, and 1 individual with undetermined sex).

Note 12, Conjugator (2, 7) and 2 manuscripts). The mean that the mean that the respective has the mean that the beaution of the single mean that the mean that the beaution of the single mean that the mean that is the beau We considered the first two years of dispersal, for two reasons. On the one hand, it is the most common age to breed in this species. The Red Kite starts breeding at 2-4 years up to 7 years (Newton *et al.* 1989, Mougeot *et al.* 2011), although exceptional cases of successful breeding at 1 year have been reported (Evans *et al.* 1998). However, the most frequent age in the Spanish population for breeding is two years old (SEO/BirdLife own database). On the other hand, the dynamics of the individuals throughout the year (departure from natal nest-dispersal movements-return to natal nest; more details in the Results section) allowed the comparison between natural years. We only considered complete "years" of dispersal, that is, continuous data recorded from the departure from the nest until they return in the next year. These "years" of dispersal, based on own data (see Results section), could last from 7 to 14 months, due to variability in the departure and arrival dates. Therefore, the dispersal years lasted from summer to summer, from departure from the nest until the return to it. Individuals dead before the first arrival to the nest, or with serious transmitters failures, were discarded.

Spatial parameters and statistical analysis

We calculated the mean and maximum distance to nest, the daily travelled distances, and the hourly travelled distance. The maximum distance was calculated as the difference between the nest and the farthest point

reached by the individual during the year, following a straight-line distance. The daily distance was calculated by selecting the last location of each day, the closest as possible to midnight to assure a stable point during the resting of the individual (Limiñana *et al.* 2012). The hourly travelled distance was obtained dividing the distance between two locations, previously filtered at 1-hour frequency, to homogenize the data provided by the different transmitters (García-Macía *et al*. 2021). We did not include locations close to the nest, so parameters were calculated after departure from the nest, both in the first and second year of dispersal. To study the differences in dispersal movements within the same individual, Mann-Whitney-Wilcoxon tests (MWW; data followed a non-normal distribution) were performed (Vidal-Mateo 2019). We compared the first and the second years of dispersal for each individual separately, using the mean distances to the nest, the daily travelled distances, and the hourly travelled distances. We summarized the results by providing percentages of individuals that decreased, increased, or kept these values constant. We also performed a Repeatability test, that is, the intraclass correlation coefficient of the variable "year", using the LMM method with the *rptR* package (Stoffel, Nakagawa & Schielzeth, 2017). Repeatability is the proportion of the variation between measurements that is due to consistent differences between the measures (Harper 1994).

We analysed the differences in dispersal movements in relation to sex. We used the mean distances to the nest, the daily travelled distances, and the hourly travelled distances as response variables in three different linear mixed models (LMM). "Sex" was included as fixed factor in all models. In addition, "individual" identity and "period" were included as random effects. Data from the Balearic Islands individual was excluded from LMM analyses, due to the geographic limitations of islands for dispersal, where they are expected to travel lower distances than in the Iberian Peninsula (Panuccio 2011).

individuals that decreased, increase, increase loss constant. We also performed a Repeatability that the mean standard increase correlation coefficient of the variable "year", using the LMM method with the grifs
package (S The establishment of the first breeding area was considered when individuals ended their dispersal movements, settling during the breeding season of the second year and remained in the same area for, at least, the whole summer. No ground truthing of the breeding events took place, so we presumed they were settled to breed. Not all individuals ended their dispersal movements after the second year, so those were not considered in these analyses. We also did not consider individuals that died, or had transmitters failures, during the summer of the second year. Therefore, only individuals with complete data for the two years were considered, that is, those ones that reached at least the month of July of their second year, as breeding comprises from March to July (Vidal-Mateo, 2019).

Temporary settlement areas (TSAs) of the individuals were determined and analysed. Earlier studies used different criteria to define TSAs (Balbotin & Ferrer 2009, Nemček *et al.* 2016, Vidal-Mateo 2019). However, based on our data, we considered TSA as a zone where one or more juveniles settled for a period longer than 30 days and travelled less than 10 km per day. TSA for different periods and/or individuals whose mean locations were separated by less than 15 km were considered as the same TSA. We estimated the home range area for each TSA with the 95 %, 75 % and 50 % Kernel Density Estimation (KDE) (Worton 1989), excluding a few individuals with gaps and transmitters failures during the settlement period.

All statistical analyses were performed using R software version 4.0.3 (R Core Team 2020), and the significance level was established at p < 0.05. The QGIS software version 3.12 was used to visualize the positioning and routes followed by the individuals, the TSA locations, as well as for the creation of maps. Values were expressed as mean ± standard deviation (*SD*).

Results

Differences between first and second year of dispersal

Red kites in our study moved across the entire Iberian Peninsula, in all directions, without a clear pattern due to their wandering movements and geographic variability of their nests (Figure IA). They abandoned the nest after their first summer as chicks. Then, they performed wandering movements, usually combined with establishments in TSAs. Finally, they returned to the parental breeding area the next summer. If individuals lived for more than a year, they repeated these dynamics during the second one, until they established in their first breeding area in their third summer. Four individuals

continued their dispersal during the third year, but they died after a few months, so we considered this information was incomplete, and we excluded them from the analyses.Not a single individual left the Iberian Peninsula during its dispersal period.

The dispersal for all juveniles started between 26 May and 26 July, with an average date on the 7 July, showing a high variability. The duration of the dispersal periods (departure from the natal nest and later return to it) was 314 ± 63 days (mean *± SD*), from a minimum of 210 days (7 months) up to 403 days (13 months). There was high individual variability during the dispersal period regarding mean distance to nest, travelled distances, and maximum distance reached. The first year of dispersal (*n =* 43) had higher and more variable values of these parameters than the second one (*n =* 21; Figure 1), but there was an overlap between the values of one and another year.

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However, we found differences in pairwise comparisons between the first and second disper

year of each individual, showing a decreasing However, we found differences in pairwise comparisons between the first and second dispersal year of each individual, showing a decreasing trend of the mentioned variables (Table 1, Figure 2). 80 % of individuals reached a shorter maximum distance during the second year, 70% decreased their mean distance to nest, 65% decreased their hourly travelled distances, and 50% decreased their daily travelled distances. The remaining individuals either kept constant or increased these variables (Figure IIA). The repeatability values (intra-class correlation coefficients for the variable "individual") varied between the different variables: mean distance to nest (*r =* 0.744 ± 0.110), maximum distance to nest (*r =* 0.570 ± 0.19), hourly travelled distance (r= 0.596 ± 0.142), and daily travelled distance (*r =* 0.176 ± 0.165).

On the other hand, there were no significant differences between males and females (*n =* 57) for any variable (Figure IIB): mean distance to the nest (*df =* 45.42; *P =* 0.199), maximum distance to the nest (*df =* 42.77; *P =* 0.120), daily travelled distance (*df =* 34.44; *P =* 0.464) and hourly travelled distance (*df =* 47.60; *P =* 0.394).

First breeding area

Coinciding with the breeding season of their second year, nine individuals (*n =*13 with complete data until that date) settled with breeding expectations at 29.1 ± 29.8 km from their natal nests (range = 5 – 91 km; Table 2). Four individuals settled at less than 10 km from their natal nest, three individuals at 25 – 30 km, and two individuals at more than 60 km. The four remaining individuals continued their dispersal across the territory and did not settle.

Temporary settlement areas (TSA)

We identified 76 temporary settlement areas (Table IIIA, Figure 3). Forty-one individuals (95 %) established at least one TSA during their dispersal (according to our definition criteria). Individuals settled in 1.85 ± 1.04 different TSAs (range = 1-6) considering their complete dispersal period (until end of data, death, or end of dispersal). The duration of settlement in each TSA was 75 ± 40 days (up to 182 days; Table A1; Figure 6). The TSAs were located at 182 ± 168 km from the nest, with high variability between individuals. The closest TSA was at 13 km from the nest, and the farthest was at 653 km (Table IA). Nevertheless, the highest proportion of TSAs were located at less than 100 km from the nest (Figure 5). The TSAs surface presented high variability (Figure 6, Table IIIB): 95% KDE estimated 781.0 ± 1895.0 km² (range = 7.4 – 14244.6), 75% KDE estimated 241.9 \pm 462.8 km² (range = 1.1 – 3189.9), and 50% KDE estimated 93.7 \pm 180.8 km² (range = 0.1 – 1214.9).

Many juveniles coincided spatially during their settlement. We found ten common TSA (Figure 4), where between 2 and 4 individuals settled. 54 % of the individuals in our study settled in one of these common areas at least once during their dispersal. They were distributed in the provinces of Madrid, Toledo, Badajoz, and Córdoba. There were five TSAs located on the outskirts of the city of Madrid, close to urban areas, with landscapes of plains with trees patches and clearings. Two of these points were found near landfills and

dumps. In Toledo, the two areas were surrounded by wide areas of fields and dispersed tree crops. In Badajoz, one TSA was 2 km from an urban area with dispersed trees and frequent crops, while the second TSA was more isolated, in a landscape with frequent forest patches and a few tree crops. The TSA of Córdoba was in a landscape with clearings and groves, close to a residue treatment plant.

DISCUSSION

Here, we studied the dispersal period of juvenile Spanish Red Kites with a large sample size.We found differences in their behavior between the first and second year of dispersal, and identified their temporary settlement areas (TSA), which could be one of the keys for the conservation of this raptor.

movements (Carter 2007, Newton 2010), variables describing the dispersal movements of juverilles show
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distances to the Despite the high individual variability, expected due to the wandering and exploratory nature of these movements (Carter 2007, Newton 2010), variables describing the dispersal movements of juveniles showed a decreasing trend during the second dispersal year, which was characterized by shorter journeys and smaller distances to the nest than the first year. On the other hand, the repeatability values showed that birds with long dispersal distances and hourly travelled distances in the first year also showed similar trends relative to other individuals in the second year. Therefore, this could indicate that Red Kites often explore similar territories during their second year, but they do it in a more efficient manner. In the first year, the Red Kites performed far and long-lasting journeys, reaching higher distances from the nest to scout the potential areas of interest, including TSAs. During the second year, juveniles had a better knowledge of the territory, including the areas with the best conditions for settlement. Thus, individuals reduced the energetic cost of their trips and flew directly to the best areas discovered in the previous year. Other studies on juvenile dispersal of raptors showed the same trend. In the case of the Spanish imperial eagle *Aquila adalberti*, the exploratory flights decreased after the 4th month of dispersal, due to acquired knowledge of the dispersal area (Ferrer 1993). Furthermore, the immature raptors, including the Red Kite, take more time and travel longer distances during migration than adults (Mellone *et al*. 2013, Sergio *et al.* 2014, García-Macía *et al.* 2021), also indicating that raptors improve their movements as they reach adulthood.

Juvenile dispersal of the Spanish Red Kites was not influenced by sex, which is supported by earlier studies in this species and other raptors (Newton *et al.* 1989, Walls & Kenward 1995, Real & Mañosa 2001, Balbontín & Ferrer 2009). However, Literák *et al.* (2022) demonstrated that females Red Kites from Central European populations travelled higher distances than males, including both migratory and dispersal movements, which was already pointed by some studies with other raptors (Forsman *et al.* 2002, Whitfield *et al.* 2009, Millsap *et al.* 2014). These differences in movement lengths may be no appreciable in the Spanish breeding population due to the geographical limits of the Iberian Peninsula. These individuals do not migrate long distances, as Central European Red Kites do, and do not leave the Iberian Peninsula during their dispersal, so both males and females can only reach small distances. Therefore, the geographic limits of the explorable territories might do that sex differences during juvenile dispersal of birds are much smaller or statistically unnoticeable.

We identified a philopatric tendency on the juvenile Red Kites. Most of the juveniles settled with breeding expectations after their second year of dispersal near their natal nests. This philopatric behavior has been found in the Red Kite and other raptors (Bustamante 1993, Ferrer 1993, Newton *et al.* 1994, Negro *et al.* 1997, Literák *et al*. 2022), and could have negative consequences in terms of population recovery and expansion of their distribution area. Short dispersal when establishing the first nest hinders the colonization of new or abandoned breeding areas, impeding the expansion of the species (Heuck *et al.* 2013). However, our results may not be conclusive about this issue, because we did not confirm breeding by field monitoring and did not have breeding performance data. Therefore, so further studies about the philopatric behaviour in the Spanish Red Kites should be made, including specific data of the first breeding of juveniles.

The establishment of TSAs is fundamental to understand the juvenile dispersal of Red Kites, as well as the main threats they face, and the possible conservation measures. Our results showed that juvenile Red Kites stopped during their dispersive phase in temporary settlement areas (TSA), where they spent at least one month. These TSA have already been described in previous literature, in which the time and space criteria differ depending on the authors (Ferrer 1993, Prommer & Bagyura 2007, Nemček *et al.* 2016, Vidal-Mateo

2019). Not all the TSA of the Iberian Peninsula can be established based on our individuals, but the large sample size in our study allows us to draw some relevant conclusions. The common factor of the shared TSA described in this study is a landscape formed by open spaces with pastures and disperse tree patches, which offer the best conditions for prey detectability and roosting places (Ontiveros *et al.* 2005). Also, all the TSAs were located at the surroundings of urban nuclei, at distances lower than 5 km, which has been lately reported because of low food availability in their traditional settlement areas at countryside agricultural areas (McIntyre *et al.* 2009, Nicolai *et al.* 2017).

residues is attractive to both Red Kites and other animals such as rodents. Similarly occurs with a founth Taylor
which instead of a landfill, is located near a cattle facility, which also acts as a food source and a juven Landfills and similar facilities were also located near TSAs. Three of the ten TSA shared by several individuals were placed near landfills and residues treatment plants (<10 km). These landfills offer regular food supply, which facilitates the settlement of juvenile individuals. Thus, the concentration of carrion and organic residues is attractive to both Red Kites and other animals such as rodents. Similarly occurs with a fourth TSA, which instead of a landfill, is located near a cattle facility, which also acts as a food source and a juveniles' area of interest for settlement. Poisoning by pollutants in the Red Kite has been widely studied, being considered one of the biggest threats to the conservation of this species (Viñuela *et al.* 2021). There are many studies that demonstrated intoxication after feeding on middens (Viñuela 1999, Monclús *et al.* 2018). However, there are not many reports of poisoning near landfills, except for one individual found in Spain (Herrero-Villar *et al.* 2021). However, our results suggest that landfills may be a risk factor for the conservation of this species, so further specific studies are necessary to find out this issue.

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References

Aebischer A*,* Scherler P, 2021. *Der Rotmilan: Ein Greifvogel im Aufwind.* Germany: Haupt Verlag

- Balbontín J*,* Ferrer M, 2009. Movements of juvenile Bonelli's eagles *Aquila fasciata* during dispersal. *Bird Study* 56(1): 86–95.
- BirdLife International, 2021. *Red Kite (*Milvus milvus*) – BirdLife species factsheet.* BirdLife International. Available at: http://datazone.birdlife.org/species/factsheet/red-kite-milvus-milvus/refs
- Bustamante J, 1993. Post-fledging dependence period and development of flight and hunting behaviour in the red kite *Milvus milvus. Bird Study* 40(3): 181–188.
- Cadahía L, López-López P, Urios V*,* Negro JJ, 2010. Satellite telemetry reveals individual variation in juvenile Bonelli's eagle dispersal areas. *Eur J Wildl Res* 56(6): 923–930.
- Carter I, 2007. *The Red Kite*. 1st edn. England: Arlequin Publications.
- Ellegren H, 1996. First gene on the avian W chromosome (CHD) provides a tag for universal sexing of nonratite birds. *Proc R Soc B* 263: 1635-1641.
- Evans IM, Cordero P*,* Parkin DT, 1998. Successful breeding at one year of age by Red Kites *Milvus milvus* in southern England. *Ibis* 140(1): 53–57.
- Ferrer M, 1993. Juvenile dispersal behaviour and natal philopatry of a long-lived raptor, the Spanish imperial eagle *Aquila adalberti. Ibis* 135(2): 132–138.
- Ferrer M*,* Calderón J, 1990. The Spanish imperial eagle *Aquila adalberti* C. L. Brehm 1861 in Doñana National Park (South West Spain): A study of population dynamics. *Biol Conserv* 51(2): 151–161.
- Forsman ED, Anthony RG, Reid JA, Loschl PJ, Sovern SG et al., 2002. Natal and breeding dispersal of northern spotted owls. *J Wildl Manag* 66(4 SUPPL. 149): 1–35.
- Gadgil M, 1971. Dispersal: population consequences and evolution. *Ecology* 52: 253-261.
- Garcelon DK, 1985. *Mounting Backpack Telemetry Packages on Bald Eagles.* Arcata: Institute for Wildlife **Studies**
- García V, Iglesias-Lebrija JJ*,* Moreno-Opo R, 2021. Null effects of Garcelon harnessing method and transmitter type on soaring raptors. *Ibis* 163 (3): 899-912.
- Cadahía L. López-López P. Urios V. Negro JJ. 2010. Satellite telemetry reveals individual variation in juve

Bonelli's eagle dispersal areas. *Eur J Wildl Res* 56(6): 923–930.

Carter 1, 2007. The Red Kite. 1³ edn. Engl García JT, Viñuela J*,* Sunyer C, 1998. Geographic variation of the winter diet of the red kite *Milvus milvus* in the Iberian Peninsula. *Ibis* 140(2): 302–309.
- García-Macía J, De la Puente J, Bermejo A, Raab R*,* Urios V, 2022. High variability and dual strategy in the wintering red kites *Milvus milvus*. *Diversity* 14: 117.
- García-Macía J, Vidal-Mateo J, De la Puente J, Bermejo A, Raab R et al., 2021. Seasonal differences in migration strategies of red kites *Milvus milvus* wintering in Spain. *J Ornithol*
- Greenwood PJ, 1980. Mating systems, philopatry and dis- persal in birds and mammals. *Anim Behav* 28: 1140– 1162
- Harper DGC, 1994. Some comments on the repeatability of measurements. *Ring Migr* 15: 84-90.
- Heath J, Steenhof K*,* Foster M, 2012. Shorter migration distances associated with higher winter temperatures suggest a mechanism for advancing nesting phenology of American kestrels *Falco sparverius*. *J Avian Biol* 43(4): 376–384.
- Heuck C, Brandl R, Albrecht J*,* Gottschalk TK, 2013. The potential distribution of the Red Kite in Germany. *J Ornithol* 154(4): 911-921.
- Herrero-Villar M, Sánchez-Barbudo IS, Camarero PR, Taggart MA*,* Mateo R, 2021. Increasing incidence of barbiturate intoxication in avian scavengers and mammals in Spain. *Environ Pollut* 284: 117452.
- Kenward RE, 2001. *A Manual for Wildlife Radio Tagging. 1st edn.* Cambridge: Academic Press.
- Knott J, Newbery P*,* Barov B, 2009. *Species Action Plan for the Red Kite Milvus milvus in the European Union*. European Commission. Available at: http://ec.europa.eu/environment/nature/conservation/wildbirds/action_plans/docs/milvus_milvus.pdf.
- Limiñana R, Romero M, Mellone U*,* Urios V, 2012. Mapping the migratory routes and wintering areas of Lesser Kestrels *Falco naumanni*: new insights from satellite telemetry. *Ibis* 154: 389-399.
- Literák I, Raab R, Petretto M, Škrábal J, Spakovszky P et al., 2020. Diverse natal dispersal in four sibling red kites originating from Austria, including wintering in Tunisia. *Biologia* 75(9): 1399–1407.
- Literák I, Raab R, Škrábal J, Vyhnal S, Dostál M et al., 2022. Dispersal and philopatry in Central European red kites *Milvus milvus. J Ornithol* https://doi.org/10.1007/s10336-021-01950-5
- López-López P, 2016. Individual-based tracking systems in Ornithology: welcome to the era of big data. *Ardeola* 63(1): 103–136.
- López-López P, Limiñana R, Mellone U*,* Urios V, 2010. From the Mediterranean Sea to Madagascar: Are there ecological barriers for the long-distance migrant Eleonora's falcon? *Landscape Ecol* 25: 803-813.
- Martín B, Onrubia A*,* Ferrer MA, 2014. Effects of climate change on the migratory behavior of the common buzzard *Buteo buteo. Clim Res* 60(3): 187–197.
- Mattsson BJ, Mateo-Tomás P, Aebischer A, Rösner S, Kunz F et al., In Review. Enhancing monitoring and transboundary collaboration for conserving migratory species under global change: the priority case of the red kite *Milvus milvus*. *Glob Change Biol*
- McIntyre CL, Douglas DC*,* Adams LG, 2009. Movements of juvenile Gyrfalcons from western and interior Alaska following departure from their natal areas. *J Raptor Res* 43(2): 99–109.
- Knot. J, Newbery P, Barrov B, 2009. Species Action Plan for the Red Kite Milvus milvus in the European Unity-leven commission. Available at:

Imagino Commission. Available at:

Imagino Laty Concept control in the European Mellone U, López-López P, Limiñana R, Piasevoli G*,* Urios V, 2013. The trans-equatorial loop migration system of Eleonora's falcon: Differences in migration patterns between age classes, regions and seasons. *J Avian Biol* 44(5): 417–426.
- Millsap BA, Harmata AR, Stahlecker DW*,* Mikesic DG, 2014. Natal dispersal distance of bald and golden eagles originating in the coterminous United States as inferred from band encounters. *J Raptor Res* 48(1): 13–23.
- MITECO, 2020. Situación actual del Listado de Especies Silvestres en Régimen de Protección Especial y Catálogo Español de Especies Amenazadas. MITECO – Gobierno de España. https://www.miteco.gob.es/es/biodiversidad/ temas/conservacion-de-especies/especies-proteccionespecial/ce-proteccion-listado-situacion.aspx.
- Molina, B, 2015. *El milano real en España. III Censo nacional. Población invernante y reproductora en 2014 y métodos de censo.* Madrid, Spain: SEO/Birdlife.
- Monclús L, Ballesteros-Cano R, De La Puente J, Lacorte S*,* Lopez-Bejar M, 2018. Influence of persistent organic pollutants on the endocrine stress response in free-living and captive red kites *Milvus milvus*. *Environ Pollut* 242: 329-337.
- Mougeot F, Garcia JT*,* Viñuela J, 2011. Breeding biology, behaviour, diet and conservation of the red kite *Milvus milvus*, with particular emphasis on Mediterranean populations. In: Zuberogoitia I, Martínez JE eds. *Ecology and Conservation of European Dwelling Forest Raptors and Owls*. Vizcaya, Spain: Diputación Foral de Vizcaya, 190-204.
- Negro JJ, Hiraldo F*,* Donazar JA, 1997. Causes of natal dispersal in the lesser kestrel: Inbreeding avoidance or resource competition? J *Anim Ecol* 66(5): 640.
- Nemček V, 2013. Movements of a juvenile red kite *Milvus milvus* in the border zone of Austria, Slovakia and the Czech Republic. *Slovak Raptor J* 7(1): 43–48.
- Nemček V, Chavko J*,* Deutschová L, 2014. Movement of satellite-tracked juvenile saker falcons *Falco cherrug* in SW Slovakia. *Slovak Raptor J* 8(2): 97–103.
- Nemček V, 2013. Movements of a juvenile red kite *Miltus miltus* in the border zone of Austria, Shovakia a

the Czech Republic. *Slovak Raptor* J 7(1): 43–48.

Nemček V, Chavka, Deutschero's 1, 2014. Movement of satellite-Nemček V, Uhrin M, Chavko J, Deutschova L, Maderič B et al., 2016. Habitat structure of temporary settlement areas of young Saker falcon *Falco cherrug* females during movements in Europe. *Acta Ornithol* 51(1): 93– 103.
- Newton I, 1998. *Population Limitation in Birds*. 1st edn. Cambridge, United Stated: Academic Press.
- Newton I, 2010. *Bird Migration.* Glasgow: William Collins.
- Newton I, Davis PE*,* Davis JE, 1989. Age of first breeding, dispersal and survival of red kites *Milvus milvus* in Wales. *Ibis* 131(1): 16–21.
- Newton I, Davis PE*,* Moss D, 1994. Philopatry and population growth of red kites *Milvus milvus* in Wales. *Proc R Soc B* 257(1350): 317–323. https://doi.org/10.1098/rspb.1994.0132
- Nicolai B, Mammen U*,* Kolbe M, 2017. Long-term changes in population and habitat selection of red kite *Milvus milvus* of the region with the highest population density. *Vogelwelt* 137: 194–197.
- Ontiveros D, Pleguezuelos JM*,* Caro J, 2005. Prey density, prey detectability and food habits: The case of Bonelli's eagle and the conservation measures. *Biol Conserv* 123(1): 19–25.
- Panuccio M, 2011. Across and around a barrier: migration ecology of raptors in the Mediterranean basin. *Scientifica Acta* 5(1): 27–36.
- Panuccio M, Mellone U*,* Agostini N, 2021. *Migration Strategies of Birds Of Prey in Western Palearctic*. CRC Press: USA.

Penteriani V*,* Del Mar Delgado M, 2009. Thoughts on natal dispersal. *J Raptor Res* 43(2): 90–98.

Prommer M*,* Bagyura J, 2007. Dangerous journeys of sakers of the Carpathian basin. *Tracker News* 8(2): 4–7.

- R Core Team, 2020. *R: A Language and Environment for Statistical Computing*. Vienna: R Foundation for Statistical Computing. https://www.R-project.org/.
- Real J*,* Mañosa S, 2001. Dispersal of juvenile and immature Bonelli's eagles in northeastern Spain. *J Raptor Res* 35(1): 9–14.
- Ronce O, 2007. How does it feel to be like a rolling stone? Ten questions about dispersal evolution. *Annu Rev Ecol Evol Syst* 38:231–253.
- Seoane J, Viñuela J, Díaz-Delgado R, Bustamante J, 2003. The effects of land use and climate on red kite distribution in the Iberian Peninsula. *Biol Conserv* 111(3): 401-414.
- Sergio F, Tanferna A, De Stephanis R, Jiménez LL, Blas J et al., 2014. Individual improvements and selective mortality shape lifelong migratory performance. *Nature* 515(7527): 410–413.
- Soutullo A, Urios V, Ferrer M*,* Peñarrubia SG, 2006. Dispersal of golden eagles *Aquila chrysaetos* during their first year of life. *Bird Study* 53(3): 258–264.
- Stoffel MA, Nakagawa S*,* Schielzeth H, 2017. rptR: repeatability estimation and variance decomposition by generalized linear mixed-effects models. *Methods Ecol Evol* 8: 1639-1644.
- Thaxter CB, Ross-Smith VH, Clark JA, Clark NA, Conway GJ et al., 2014. A trial of three harmes attachmomethods and their suitability for long-term use on lesser black-bucked gulls and great skans. *King Migran* 29(2): 65-7 Thaxter CB, Ross-Smith VH, Clark JA, Clark NA, Conway GJ et al., 2014. A trial of three harness attachment methods and their suitability for long-term use on lesser black-backed gulls and great skuas. *Ring Migr* 29(2): 65-76.
- Urios V, Muñiz-López R*,* Vidal-Mateo J, 2017. Juvenile dispersal of harpy eagles *Harpia harpyja* in Ecuador. *J Raptor Res* 51(4): 439–445.
- Urios V, Romero M, Mellone U, 2015. *The Use of Satellite Telemetry for the Study of the Movement Ecology of Raptors.* Alicante, Spain: Publicaciones de la Universidad de Alicante.
- Vidal-Mateo J, 2019. *Ecología del movimiento de aves rapaces migradoras*. PhD thesis. University of Alicante. Alicante.
- Vidal-Mateo J, De La Puente J, Bermejo A*,* Urios V, 2021. Red Kite *Milvus Milvus*. In: Panuccio M, Mellone U, Agostini A eds. *Migration Strategies of birds of Prey in Western Paleartic*. Boca Raton CRC Press, 160- 164.
- Viñuela J, De la Puente J*,* Bermejo A, 2021. Milano real *Milvus milvus.* In: López-Jiménez N ed. *Libro Rojo de las Aves de España:* Madrid SEO/BirdLife, 446-455.
- Viñuela J, Martí R*,* Ruiz A, 1999. *El Milano Real en España. Monografía nº6.* Madrid: SEO/BirdLife.
- Visser M, Perdeck A, Van Balen J*,* Both C, 2009. Climate change leads to decreasing bird migration distances. *Glob Change Biol* 15(8): 1859–1865.
- Walls S*,* Kenward R, 1995. Movements of radio-tagged common buzzards *Buteo buteo* in their first year. *Ibis* 137(2): 177–182.
- Whitfield DP, Douse A, Evans RJ, Grant J, Love J et al., 2009. Natal and breeding dispersal in a reintroduced population of white-tailed eagles *Haliaeetus albicilla*. *Bird Study* 56(2): 177–186.
- Worton BJ, 1989. Kernel methods for estimating the utilization distribution in home-range studies. *Ecology* 70 (1): 164-168.
- Wormworth J*,* Mallon K, 2006. *Bird Species and Climate Change: The Global Status Report*. *version 1.0.* Australia: WWF.

Table 1. Distance from the nest, daily distance, hourly travelled distance, and maximum distance of the two years of the dispersal of all individuals. Mann-Whitney-Wilcoxon test p-values are also shown.

Table 2. Location of the breeding areas for the individuals that settled to breed at the end of the second year. The relative distance to their natal nest is presented, with the duration of the settlement periods.

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Figure 1. Comparison between first and second dispersal periods according to the mean distance to nest (km), maximum distance to nest (km), daily travelled distance (km/day) and hourly travelled distance (km/h)

Figure 2. Example of two individuals' tracks (Yelmo and Millau) which reduced their dispersal movements between the first and second year of dispersal. Temporary Settlement Areas are represented by a white pentagon, while natal nest is represented by a yellow star .

Figure 3. Locations of all the Temporary Settlement Areas (TSA) where tagged red kites settled during their juvenile dispersal.

Figure 4. Distribution of the TSAs (Temporary Settlement Area) shared by more than one individual.

Figure 5 Histograms of the mean distances from the nest (A), and duration settlement (B) of Temporary Settlement Areas identified for the juvenile red kites.

Figure 6. Kernel Density Estimators (95 %, 75 % and 50 %) for a Temporary Settlement Area (TSA) of Millau in the province of Badajoz during its first year of dispersal. This area is representative because forest, agricultural, urban, river and montane patches are found inside the home range of this red kite.

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