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Christopher F. Bonzek Virginia Institute of Marine Science

James Gartland Virginia Institute of Marine Science

Debra J. Gauthier Virginia Institute of Marine Science

Robert J. Latour Virginia Institute of Marine Science

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ANNUAL REPORT

Data collection and analysis in support of single and multispecies stock assessments in Chesapeake Bay:

The Chesapeake Bay Multispecies Monitoring and Assessment Program

Prepared for:

Virginia Marine Resources Commission

and

U.S. Fish & Wildlife Service

For Sampling During:

Calendar Year 2021 and Previous Years

Project Number:

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Prepared by:

Christopher F. Bonzek James Gartland Debra J. Gauthier Robert J. Latour, Ph.D

School of Marine Science College of William and Mary Virginia Institute of Marine Science Gloucester Point, VA 23062

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Introduction

Historically, fisheries management has been based on the results of single-species stock assessment models that focus on the interplay between exploitation level and sustainability. There currently exists a suite of standard and accepted analytical frameworks (e.g., virtual population analysis (VPA), biomass dynamic production modeling, delay difference models, etc.) for assessing the stocks, projecting future stock size, evaluating recovery schedules and rebuilding strategies for overfished stocks, setting allowable catches, and estimating fishing mortality or exploitation rates. A variety of methods also exist to integrate the biological system and the fisheries resource system, thereby enabling the evaluation of alternative management strategies on stock status and fishery performance. These well-established approaches have specific data requirements involving biological (life history), fisheries-dependent, and fisheries-independent data (Table 1). From these, there are two classes of stock assessment or modeling approaches used in fisheries: partial assessment based solely on understanding the biology of a species, and full analytical assessment including both biological and fisheries data.

| Data Category | Assessment Type | Data Description |
|---------------------------|-----------------|--|
| Biological / Life History | Partial | Growth (length / weight) |
| | | Maturity schedule |
| | | Fecundity |
| | | Partial recruitment schedules |
| | | Longevity |
| | | Life history strategies (reproductive and |
| | | behavioral) |
| Fishery-Dependent Data | Analytical | Catch, landings, and effort |
| | | Biological characterization of the harvest |
| | | (size, sex, age) |
| | | Gear selectivity |
| | | Discards/bycatch |
| Fishery-Independent Data | Analytical | Biological characterization of the |
| | | population (size, sex, age) |
| | | |
| | | Mortality rates |
| | | Estimates of annual juvenile recruitment |

Table 1. Summary of biological, fisheries-dependent and fisheries-independent data requirements for single-species analytical stock assessment models.

Although single-species assessment models are valuable and informative, a primary shortcoming is that they generally fail to consider the ecology of the species under management (e.g., habitat requirements, response to environmental change), ecological interactions (e.g., predation, competition), and technical interactions (e.g., discards, bycatch) (NMFS 1999, Link 2002a,b). Inclusion of ecological processes into fisheries management plans is now strongly recommended (NMFS 1999) and in some cases even mandated (NOAA 1996). Multispecies assessment models have been developed to move towards an ecosystem-based approach to fisheries management (Hollowed et al. 2000, Whipple et al. 2000, Link 2002a,b). Although such models are still designed to yield information about sustainability, they are structured to do so by incorporating the effects of ecological processes among interacting populations.

In recent years, the number and type of multispecies models designed to provide insight about fisheries questions has grown significantly (Hollowed et al. 2000, Whipple et al. 2000, Collie et al. 2014, Heymans et al. 2016). While this growth has been fueled primarily by the need to better inform fisheries policy makers and managers, recent concerns about effects of fishing on the structure of ecosystems have also prompted research activities on multispecies modeling and the predator-prey relationships that are implied. From a theoretical perspective, basing fisheries stock assessments on multispecies rather than single-species models certainly appears to be more appropriate, since multispecies approaches allow a greater number of the processes that govern population abundance to be modeled. However, this increase in realism leads to an increased number of model parameters, which in turn, creates the need for additional types of data.

In the Chesapeake Bay region, there has been a growing interest in ecosystem-based fisheries management, as evidenced by the recent development of fisheries steering groups (e.g., ASMFC multispecies committee), the convening of technical workshops (Miller et al. 1996; Houde et al. 1998) development of planning processes for implementing ecosystem planning (C.B. Fisheries Ecosystem Advisory Panel, 2006) and most recently a Scientific and Technical Advisory Committee sponsored workshop title "Assessing the Chesapeake Bay Forage Base" at which ChesMMAP data were the principal source of data for numerous species (Ihde et al., 2015).

If either single-species or ecosystem-based management plans are to be developed and maintained, they must be based on sound stock assessments. In the Chesapeake Bay region, however, the data needed to perform single and multispecies assessments are typically either partially available or nonexistent. The Chesapeake Bay Multispecies Monitoring and Assessment Program (ChesMMAP) was developed to assist in filling these data gaps, and ultimately to support bay-specific stock assessment modeling activities at both single and multispecies scales. While no single gear or monitoring program can collect all of the data necessary for both types of assessments, ChesMMAP was designed to maximize the biological and ecological information collected for several recreationally, commercially, and ecologically important species in the bay.

In general, ChesMMAP is fishery-independent monitoring survey that uses a bottom trawl designed to sample late juvenile-to-adult fishes in the mainstem of Chesapeake Bay. This program currently provides data on relative abundance, length, weight, sex ratio, maturity, age, and trophic interactions for several important fish species that inhabit the bay seasonally. Among the research agencies in the Chesapeake Bay region, only VIMS has a program focused on multispecies issues involving the late juvenile and adult (i.e., harvested) components of the exploited fish species that seasonally inhabit the bay. The Multispecies Research Group (MRG) is also responsible for executing the nearshore trawl survey for the Northeast Area Monitoring and Assessment Program (NEAMAP), as well as the VIMS Elasmobranch Longline Survey and the Striped Bass Spawning Stock Monitoring and Tagging program. In this report, we summarize the ChesMMAP field, laboratory, and data analysis activities through the 2021 sampling year.

A new ChesMMAP task included during recent segments was initial evaluation of a potential new sampling gear system. This system includes a one-half size (200 x 12cm fishing circle) version of the same trawl net in use for the Northeast Fisheries Science Center (NEFSC) and NEAMAP surveys (400 x 12cm fishing circle). Scale model flume tank testing occurred during an earlier segment, initial field testing took place during 2009-2010 and the first comparative (to the existing gear) field trials took place in 2010-2011. Due to previously unanticipated upgrades and replacement plans for the *R/V Bay Eagle* it was determined that the most prudent course of action was to delay further testing during the current

segment (fully explained in Methods below). In late 2018 VIMS took delivery of the *Bay Eagle* replacement vessel, the *R/V Virginia* and beginning in June 2019 the new fishing system was implemented for ChesMMAP surveys.

The MRG has been attempting to steadily improve its online presence and provide stakeholders, scientists, and managers with ready access to significant parts of the ChesMMAP (and other monitoring surveys conducted by the group) data bases. Several elements of particular significance have been made accessible in recent years:

- Introduction Text, photos and drill-down links regarding the ChesMMAP program at <u>https://www.vims.edu/research/departments/fisheries/programs/mrg_draft/chesmmap/index.php</u>.
- Abundance Indices All measures of relative abundance and most other analyses presented in this report are also available online at www.vims.edu/fisheries/chesmmapindices/index.php
- Food Habits Summaries A variety of user-selectable summarizations of fish diet information, from either the predator or the prey point of view, are available at http://www.vims.edu/fisheries/fishfood.
- Station-Specific Catches GIS style representations of tow-specific catch information for ChesMMAP (and other) data with user-selected data filters are at: www.vims.edu/fisheries/fao/index.php.
- An interactive Infogram report located at: <u>https://infogram.com/2021-chesmmap-1h7k230rwqg1g2x</u>

These links as well as much more information about ChesMMAP and other programs conducted by the MRG are available at <u>http://www.vims.edu/fisheries/mrg</u>.

The following Tasks are addressed in this report:

- Task 1 Conduct research cruises
- Task 2 Synthesize data for single species analyses
- Task 3 Quantify trophic interactions for multispecies analyses
- Task 4 Estimate abundance
- Task 5 Continue evaluation of alternative sampling gear.

Methods

Task 1 – Conduct research cruises

The timing of the cruises was chosen so as to coincide with the seasonal abundances of fishes in the bay. Prior to 2019 the ChesMMAP sampling protocol was to conduct five 80-station surveys per year, one each in March, May, July, September and November, though this protocol was occasionally interrupted by funding shortfalls and/or logistical hurdles (e.g. vessel breakdowns). This sampling schedule changed in 2019 due to a combination of increased costs associated with the *R/V Virginia* and a decreasing budget. The result was that the equivalent of only 3 sampling cruises could be conducted. However, in considering the annual pattern of fish abundances and in examining the subsets of the data used for the various species' abundance indices, an alternative approach was implemented. In the early season (March) cruise none of the data from sampling in Virginia are used for any abundance indices. Likewise, in late season sampling (November), data for only one species in Maryland strata are used. Rather than settling for 3 full cruises we now sample in March, June, September and November, with the March and November trips sampling only in the upper (Maryland) and lower (Virginia) regions respectively. While not ideal, we can still sample during the entire spring/summer/fall annual cycle and this pattern was used during the four research cruises in calendar year 2021.

The *R/V Bay Eagle*, a 19.8m aluminum hull, twin diesel vessel owned and operated by VIMS, served as the sampling platform for all cruises between March 2002 and April 2019. Fishes (and select invertebrates) were collected using a 13.7m (headrope length), two-bridle, four-seam bottom trawl manufactured by *Reidar's Manufacturing Inc.* of New Bedford, MA. The top belly, bottom belly, and side panels of the net are constructed of 15.2cm stretch mesh (2.6mm diameter twine), and the cod-end is constructed of 7.6cm stretch mesh (1.6mm diameter twine). The bridles (legs) of the net are 6.1m and connected directly to 1.3m x 0.8m steel-V trawl doors weighing 71.8kg each. The trawl net is deployed with a single-warp system using 9.5mm (dia.) steel main cable and a 37.6m bridle constructed of 7.9mm stainless steel wire rope.

Beginning with the June 2019 trip, sampling was and, in the future, will be conducted onboard VIMS' new *R/V Virginia*, a 28.3m steel hull vessel with twin diesels tied to a single controllable-pitch propeller and a separate bow thruster for station-holding. The biota is sampled using a 200 x 12cm 3-bridle 4-seam trawl using Thyboron Type IV 44" steel doors attached to a double-warp system controlled by Rapp-Hydema's Pentagon[©] winch control software. The gear is a half-scale duplicate of the trawling system used both by the NEAMAP Mid-Atlantic survey and the NEFSC.

Between March 2002 and April 2019, the goal was to sample 80 sites throughout the mainstem of Chesapeake Bay during each cruise. Sampling sites were selected using a stratified random design. The bay was stratified by dividing the mainstem into five regions of 30 latitudinal minutes each (the upper and lower regions being slightly smaller and larger than 30 minutes, respectively). Regions were numbered 1 through 5 from north to south. Regions 1-3 coincide with the Maryland portion of the bay and regions 4-5 correspond with Virginia waters (note that due to the irregular state boundary it is possible that stations in the very southernmost portion of Region 3 may actually be in Virginia and likewise stations in the northernmost reaches of Region 4 may be north of the state border). Within each region, three depth strata ranging from 3.0 m-9.1 m, 9.1 m-15.2 m, and >15.2 m were defined. A grid of 1.9 km² cells was superimposed over the mainstem, where each cell represented a potential sampling location. The number of stations sampled in each region and in each stratum was proportional to the surface area of water represented. Stations were sampled without replacement and those north of Pooles Island (latitude 39° 17') have not been sampled since July 2002 due to repeated loss of trawl gear.

Knowing that significant changes would be coming to the survey with the change in research vessel and sampling gear, we took the opportunity to also consider changes in the survey stratification. Analyses revealed that the prior design was over-stratified, with small numbers of samples coming from small strata but being over-represented in the design due to the criterion of sampling at least three stations from every stratum. Both the number of Regions and the number of depth strata were reduced. The prior three Regions corresponding to the Maryland portion of the bay were condensed to two and similarly the number of depth strata in each Region was reduced from three (described in the preceding paragraph) to two (<=12.2m, >12.2m). Thus, the total number of strata sampled during any cruise was reduced from 14 (there was no deep stratum in Region 1) to 8. Regions are now described as Regions A (upper Maryland), B (lower Maryland), C (upper Virginia) and D (lower Virginia) and depth strata are similarly named A (shallow) and B (deep). While it may be somewhat confusing to use a similar labeling

system for both the Region and the depths these conventions provide a clear distinction from the previous classifications.

For sampling aboard the *Bay Eagle*, tows were normally conducted in the same general direction as the tidal current (pilot work conducted using the net monitoring gear in November 2001 indicated that the survey gear performed most consistently when towed with the current rather than against the current). The net was generally deployed at a 4:1 scope, which refers to the cable length: water depth ratio. For shallow stations, however, the 120ft. bridle wires were always fully deployed, implying that the scope ratio could be quite high in these particular situations. The target tow speed was 3.0 kts but occasionally varied depending on wind and tidal conditions. Based on data collected from the net monitoring gear, tow speed and scope were adjusted to ensure that the net maintained expected geometry. Tows were 20 minutes in duration, unless obstructions or other logistical issues forced a tow to be shortened (if the duration of a tow was at least 10 minutes, it was considered valid). Computer software was used to record data from the net monitoring gear (i.e., wingspread and headrope height) as well as a continuous GPS stream during each tow. On occasions when the monitoring gear failed or was not deployed, the trawl geometry was assumed to follow cruise averages and beginning and ending tow coordinates were recorded by hand from the vessel's GPS system. These same general parameters are held true for sampling on the *Virginia*.

Task 2 – Synthesize data for single species analyses

Once onboard, the catch from each tow was sorted and measured by species and, if distinct modal length classes within a particular species were evident, by size-class. A subsample of each species/sizeclass was further processed for individual weight determination, stomach contents, ageing, and determination of sex and maturity stage. At each sampling location, in addition to these biological data, water temperature, salinity, and dissolved oxygen readings were recorded electronically at approximately 1m intervals.

Single-species assessment models typically require information on (among others) age-, length-, and weight-structure, sex ratio, and maturity stage. Data were synthesized to characterize annual lengthand age-frequency distributions. Analytical computer programs to characterize each of the assessmentrelated data elements (length, weight, age, sex, maturity) were developed to allow for the summarization of these characteristics across a variety of spatial and temporal scales (e.g., by year, season, or region of the bay) for each species.

Task 3 – Quantify trophic interactions for multispecies analyses

In addition to the population-level information described under Task 2, multispecies assessment models require information on predator-prey interactions across broad seasonal and spatial scales. In general, these procedures involve examining the stomach contents of predators and identifying each prey item to the lowest possible taxonomic level. As such, stomach samples were collected and preserved in the field and were processed at VIMS following standard diet analysis procedures (Hyslop 1980). Several diet indices were calculated to identify the main prey types for each species sampled by the ChesMMAP Survey: percent weight, percent number, and percent frequency-of-occurrence.

Both percent weight and percent number are offered in this report. In the food habits figures presented for each species, prey types are ordered first in decreasing percentage by weight order by major taxa (e.g. fish, crustaceans, molluscs, etc.) and within each taxon by decreasing percentage for each species or subgroup. To make comparisons between percent by weight vs. by number readily accomplished, the same color scheme of major taxa is maintained in the succeeding percent by number figure though the

taxa order (again by decreasing percentage), as well as species or subgroup order within each taxon are allowed to vary.

These indices can be coupled with the information generated from Task 2 and age-, length-, and sexspecific diet characterizations can be developed for each species. Characterizing spatial and temporal variability in these diets is also possible using ChesMMAP data.

As noted above, several diet index values were calculated to identify the main prey in the diet of predators in the mainstem Chesapeake Bay. Since trawl collections essentially yield a cluster of fish at each sampling location, these indices were calculated using a cluster sampling estimator (Buckel et al. 1999).

Specifically, the contribution of each prey type to the diet by weight ($(%Q_k)$) is given by:

$$\mathcal{O}_{k} = rac{\sum\limits_{i=1}^{n} M_{i} q_{ik}}{\sum\limits_{i=1}^{n} M_{i}}$$
 ,

where

$$q_{ik}=\frac{w_{ik}}{w_i}*100\,,$$

and where *n* is the number of clusters (species/size-class combinations) of the predator of interest sampled, M_i is the number of individuals of this predator species represented in cluster *i*, w_i is the total weight of all prey items encountered in the stomachs of that predator sampled from cluster *i*, and w_{ik} is the total weight of prey type *k* in those stomachs.

Task 4 – Estimate abundance

Time-series of abundance information are standard products developed from the basic catch data of a fishery independent monitoring survey. For each species sampled by the ChesMMAP Survey, a variety of relative abundance trends can be generated according to year, season, and location within Chesapeake Bay.

Absolute abundance estimates (i.e. minimum trawlable abundance) can be generated for each species by combining abundance data with area swept by the trawl and gear efficiency. Area swept was calculated for each tow by multiplying tow distance (provided by GPS) by average wingspread (provided by net monitoring gear). Gear efficiency estimates, gained through hydroacoustic data collection as described in previous project reports, have been estimated for two species common in ChesMMAP catches (Atlantic Croaker and White Perch) and results were published (Hoffman et al. 2009).

While minimum total or absolute abundance estimates are important for certain bioenergetics and ecosystem level analyses, fishery assessments typically depend upon relative abundance indices from surveys as indicators of abundance. Previous ChesMMAP progress reports have presented an evolving series of relative and absolute abundance estimates. Still another new step in the evolution of those indices was introduced in the 2011 report. Specifically, for species for which identifiable (from analysis of hard parts) age cohorts are present in ChesMMAP samples, age-specific indices of abundance based

on ChesMMAP-developed age-length keys (ALK) were offered and those estimates are presented again this year, based on improved ALKs.

Development of ChesMMAP-specific ALKs was required due to the multiple annual sampling events (i.e. bi-monthly cruises) and inter-cruise growth. Such specific growth information has not been previously available for most species in Chesapeake Bay and could only be accomplished now, as ChesMMAP sample sizes became large enough after several years of field sampling and laboratory ageing efforts.

The methodology employed to develop the ALKs was modified during the 2019 segment and new ALKs were used for the indices presented in this report. Previously ChesMMAP ALK's represented data pooled over several years and were developed for each survey month (March, May, July, Sept., Nov.) using those pooled data. While providing a larger sample size this method resulted in a decreased ability to resolve year-to-year variability in age structure. The updated ALKs use year-specific data but in-year cruise data are pooled over two seasons labeled Spring and Summer. For most species, March and May/June trips are pooled as the Spring season and those between July through November as Summer. For a few species data indicated that the July trip should be included with the Spring season.

Once the ALKs were established for each season, all non-aged measured specimens were assigned to length bins, the total number of specimens captured within each length bin at each station was summed (full-workup specimens which had been aged remained in the assigned age class) and the season-specific age-at-length proportions applied to those sums. From this, the total number of age-specific fish captured at each station was determined. That number was then fed into the index calculation algorithm (below). For age-specific biomass indices, the average weight of specimens within each length bin within each age-class was calculated, then multiplied by the calculated (as above) number within the length bin to estimate total weight. Similarly, that figure was then processed through the index calculation algorithm. This method to calculate age-specific abundance differs somewhat from that employed by analysts at the NEFSC in which the proportion-at-age is applied to the overall index for each year. The methodology employed in this report has a slight disadvantage in that, due primarily to the transformations and back-transformations, the sum of the age-specific indices is not equal to the overall abundance index. It has the advantage however that it allows normal calculation of confidence limits on the age-specific indices.

For this report, only geometric mean abundance indices are presented. Arithmetic indices as offered in previous reports are rarely statistically valid. Delta-lognormal indices, model-based indices, and other methods of calculating relative abundance are being explored and will likely replace the geometric mean indices in future reports, on a species-by-species basis.

Abundance index calculations presented here are calculated according to:

- Raw catch data used for each species index are restricted by month, region, and depth strata such that only those strata with maximum catch-per-unit-effort for that species are used. The methods used to determine these species-specific restrictions were described in a previous progress report (Bonzek et al. 2009). For a small number of species these limiting parameters were updated in a previous segment report and were somewhat modified again in 2019 as a result of the restratification of the survey sampling frame.
- 2. Geometric Mean: Using the restricted data, annual geometric mean catch per area swept indices for each species for all ages combined, were calculated according to the formula:

$$\mathbf{I} = \exp\left\{\sum_{i=1}^{n} \left(\log\left(\frac{C}{a} + \mathbf{1}\right)\right) \times w\right\} - \mathbf{1}$$

where: I = Index

C = number or biomass caught at a station a = area swept at a station i = ith stratum n = number of strata w = stratum weight

Task 5 – Evaluation of alternative sampling gear

As discussed in previous project reports, personnel associated with the ChesMMAP Trawl Survey worked in conjunction with *Reidar's Manufacturing, Inc.* to design a survey trawl that could serve as a replacement for the sampling net currently used by this program. Specifically, a three-bridle, four-seam, 200 x 12cm (fishing circle) bottom trawl has been developed. This net is identical in design to that used to sample the near shore coastal ocean by the NEAMAP Trawl Survey, and is nearly-identical to that used by the NEFSC Bottom Trawl Survey. The three-bridle, four-seam net developed for this program is half of the size (i.e., 200 x 12cm fishing circle net for ChesMMAP vs. 400 x 12cm fishing circle net for NEAMAP and NEFSC) of those used by NEAMAP and NEFSC. Again, flume trials conducted on model trawls in December 2009 indicated that the 200 x 12cm net may be a more appropriate sampling gear than the current two-bridle four-seam, semi-balloon bottom trawl used by ChesMMAP, as the optimal configuration and performance consistency of the alternate net appeared to be superior to that of the current gear.

In October 2018, after approximately six-years of development and construction, VIMS took delivery of the *R/V Virginia* as a replacement for the aging *R/V Bay Eagle*. Most survey groups faced with a change in sampling platform, perform extensive calibration studies prior to fully implementing the new gear. In contrast, MRG chose to immediately switch to the new vessel/gear while concurrently performing the calibrations. This will result in a short period (2-3 years) in which the data from the two stanzas cannot be compared but we were confident that the new survey gear is so superior to the original one that such a delay is worth the cost. The *Virignia* was not yet Coast Guard certified in time for the April 2019 cruise but was put into service for full time ChesMMAP sampling beginning with the June 2019 trip.

Again, in contrast with many survey groups, MRG chose not to conduct calibration tows during regular survey operations. This choice was made to assure that no side-by-side vessel effects would bias the survey data.

Results

Task 1 – Conduct Research Cruises

Cruise dates and the numbers of stations completed during each survey since 2002 are shown in Table 2. For years 2002-2004 the target number of stations per cruise was 90 and since 2005 that target number has been 80 (extensive analyses of data collected through 2004 revealed that the target number could be decreased by 10 stations per cruise with little effect on survey precision, but that decreases below 80 do have a significant negative effect on precision). Examination of the data presented in Table 2 reveals that as experience has been gained and survey procedures improved, the number of calendar days per cruise has decreased from an average of 11-13 days down to 9-11 (or even fewer days if we are fortunate to have a good weather window). Likewise, the number of actual work days has decreased from a range of 8-10 down to 7-8. As the survey only pays vessel costs on days actually worked, this increased efficiency has resulted in significant cost savings (note however that some of these efficiencies have likely resulted from an overall decrease in the number of fish caught, described below). For 2019 the decrease from five 80-station trips to two full trips (June and September) plus two half trips (April and November) decreased the total number of work days from approximately 40 to 28, which closely matched the budgeted request. It should be noted however that due to the slower cruising speed of the Virginia (~8kt) compared to the Bay Eagle (11-12kt) and to the much higher catch rates and therefore station processing times with the '200' net, the average number of stations completed per day decreased from 9-10 to 7-8.

In mid-2008 we gained the ability to plot previous successful tow tracks onto electronically displayed overlays of selected sampling cells for each cruise. In difficult trawling areas, which are very common in Chesapeake Bay, by approximately retracing a successful tow track it becomes much less likely that the trawl gear will 'hang up' and/or be significantly damaged. This has resulted both in a further increase in efficiency (much less time is spent retrieving 'hung' gear so more time is spent sampling) and a decrease in the number of nets requiring major repair or replacement. Both of these elements offer further cost savings.

After reaching a maximum during the third survey year (2004), the total number of specimens sampled annually generally declined through 2018 but rose by approximately an order of magnitude in 2019 and 2020 presumably due primarily to the change in sampling gear (Table 3). Total samples collected and processed reached a time series low in 2011 (which represented a 55% decrease in total catch compared to 2004, with comparable levels of total sampling effort) and then another low in 2012, though without a March 2012 cruise. However, even if the March cruise yielded catch rates comparable to other recent years, the total number of specimens captured in 2012 would still be a time series low value. Catch rates increased somewhat in 2013 but declined to a previously unseen low value in 2014 of only 11,000 fish. Between 2015 and 2018 overall catch numbers remained generally low with the exception of 2016 when approximately 28,000 total fish were captured.

Concerns as to whether this decrease in catch is due to actual changes in species abundance or is an artifact of unknown sampling effects were examined in the previous segment reports (Bonzek et al., 2010 and 2011). Those analyses revealed that much of the decrease in total catch can be attributed to declines in measured abundance of a single species, Atlantic Croaker. Catch rates of other commonly abundant species, (e.g. Spot, Weakfish, Summer Flounder) have also declined when compared to the mid-2000s. There is still some uncertainty in the investigators' minds as to whether these declines represent real biological abundance in Chesapeake Bay or are a sampling artifact.

As anticipated, catch rates for most species increased substantially coinciding with use of the '200' net. For some species much of the increase is due to catching a broader size range, especially on the smaller end, but the increase is very large for almost every species. For those species in which either the average yearly catch between 2014 and 2018 or the 2019-2021 average catch was greater than 100 specimens (36 species, with the three categories of blue crabs counting just once) the 2019-2021 average total catch was lower for give species, ranging between 1.7% and 66.8% lower. For the remaining 31 species the increase in total catch ranged between 16.8% and 38,900%. Of the species for which abundance indices are included in this report the difference in total catch ranged between -67% (Northern Puffer) to over 2,000% (Black Seabass and Weakfish) and averaged a 763% increase (Table 4). It should be noted that the April 2019 cruise was conducted using the *Bay Eagle* sampling system so these values likely represent an underestimate of the difference in catch rates.

For the species for which abundance indices are presented, the vast majority of ageing structures (i.e. otoliths, opercles, etc.) and stomach samples preserved have been analyzed (Table 3). Currently, most of the otolith and stomach samples which remain to be processed represent species which are either of relatively minor management interest (e.g. oyster toadfish otoliths), which involve significantly different preparation and analysis techniques (e.g. elasmobranch vertebrae), which are particularly difficult to analyze (e.g. Atlantic menhaden stomachs), or which currently have no accepted processing protocols (e.g., Butterfish sampled from inshore waters).

| Year | Cruise | Begin Date | End Date | Stations Completed | Calendar Days | Work Days | Year | Cruise | Begin Date | End Date | Stations Completed | Calendar Days | Work Days |
|------|-----------------------|------------|------------|-----------------------|------------------|--------------|------|-----------------------|--------------|-----------------|-----------------------|------------------|--------------|
| 2002 | March | 3/29/2002 | 4/16/2002 | 50 | 19 | 8 | 2012 | March | No cruise du | le to vessel re | | Days | Days |
| 2002 | May | 5/20/2002 | | 80 | 9 | 8 | 2012 | May | 5/26/2012 | | 80 sources | 8 | 8 |
| | July | 7/8/2002 | | 77 | 9 | 8 | | July | 7/9/2012 | | 79 | | 8 |
| | September | 9/13/2002 | | 76 | 10 | 10 | | September | 9/3/2012 | | 80 | | 8 |
| | November | | 11/10/2002 | 70 | 10 | 9 | | November | | 11/17/2012 | 72 | | 8 |
| 2003 | March | 3/24/2003 | | 69 | 14 | 8 | 2013 | March | 3/20/2012 | | 80 | | 7 |
| 2003 | May | 5/20/2003 | | 29 | 4 | 8 | 2013 | May | 6/4/2013 | | 80 | | 7 |
| | July | 6/30/2003 | | 87 | 11 | 8 | | July | 7/8/2013 | | 80 | | 8 |
| | September | 9/30/2003 | | 73 | 9 | 8 | | September | 9/3/2013 | | 80 | | 7 |
| | November | 10/28/2003 | | 75 | 9 | 9 | | November | | 11/22/2013 | 79 | | 9 |
| 2004 | March | 3/20/2003 | | 90 | 12 | 9 | 2014 | March | 3/20/2014 | | 79 | | 7 |
| 2004 | May | 5/17/2004 | | 90 | 12 | 10 | 2014 | May | 5/29/2014 | | 80 | | 7 |
| | July | 7/1/2004 | | 59 | 10 | 10 | | July | 7/16/2014 | | 80 | | 10 |
| | | | | 80 | | 8 | | | | | 80 | | 8 |
| | September November | 9/2/2004 | | 80 | 14 14 | 10 | | September November | 9/3/2014 | | 80 | | 8 |
| 2005 | | | 11/10/2004 | 80 | 14 | | 2015 | | | 11/16/2014 | 80 | | 10 |
| 2005 | March | 3/16/2005 | | 80 | 9 | 8 | 2015 | March May | 3/19/2015 | | 80 | | |
| | May | 5/2/2005 | | | | | | | 5/28/2015 | | | | 8 |
| | July | 7/1/2005 | | 80 76 | 12 | 8 | | July | 7/7/2015 | | 80 | | 6 |
| | September | 9/8/2005 | | | 11 | 8 | | September | 8/29/2015 | | 80 | | |
| | November | 10/31/2005 | | 80 | 10 | 9 | | November | | 11/18/2015 | 80 | | 8 |
| 2006 | March | 3/23/2006 | | 80 | 9 | 8 | 2016 | March | 3/14/2016 | | | | 7 |
| | May | 5/15/2006 | | 80 | 11 | 8 | | May | 5/25/2016 | | 80 | | 7 |
| | July | 6/28/2006 | | 73 | 16 | 7 | | July | 7/6/2016 | | 80 | | 6 |
| | September | 8/30/2006 | | 70 | 15 | 8 | | September | 8/29/2016 | | 80 | | 7 |
| | November | 10/30/2006 | | 74 | 9 | 8 | | November | | 11/19/2016 | | | 9 |
| 2007 | March | 3/13/2007 | | 77 | 11 | 8 | 2017 | March | 3/20/2017 | | 79 | 9 | 8 |
| | May | 5/9/2007 | | 77 | 15 | 9 | | May | 5/15/2017 | | 80 | | 8 |
| | July | 7/2/2007 | 7/10/2007 | 78 | 9 | 9 | | July | 7/6/2017 | | 79 | | 7 |
| | September | | | | | | | September | 9/6/2017 | | 80 | | 8 |
| | November | | 11/12/2007 | 77 | 14 | 8 | | November | 11/12/2017 | 11/29/2017 | 80 | 18 | 9 |
| 2008 | March | 3/17/2008 | | 80 | 10 | 8 | 2018 | March | | | | | |
| | May | 5/20/2008 | | 78 | 8 | 8 | | May | 5/22/2018 | | 80 | | 7 |
| | July | 6/28/2008 | | 80 | 10 | 7 | | July | 7/2/2018 | | 80 | 7 | 7 |
| | September | 9/2/2008 | | 80 | 10 | 7 | | September | 9/4/2018 | | 57 | 7 | 5 |
| | November | | 11/11/2008 | 80 | 13 | 8 | | November | | 11/17/2018 | 80 | | 7 |
| 2009 | March | 3/16/2009 | 3/26/2009 | 80 | 11 | 7 | 2019 | April | 4/2/2019 | 4/5/2019 | 35 | 4 | 4 |
| | May | | | 0 | 0 | 0 | | June | 5/31/2019 | | 80 | | 9 |
| | July | 7/14/2009 | | 80 | 7 | 7 | | September | 9/3/2019 | 9/14/2019 | 79 | 12 | 10 |
| | September | 9/2/2009 | | 80 | 11 | 8 | | November | 11/19/2019 | 11/23/2019 | 45 | 5 | 5 |
| | November | 11/3/2009 | 11/10/2009 | 78 | 8 | 7 | | | | | | | |
| 2010 | March | 3/22/2010 | 3/31/2010 | 79 | 10 | 7 | 2020 | March | 3/10/2020 | 3/14/2020 | 35 | 5 | 5 |
| | May | 5/22/2010 | 5/28/2010 | 79 | 7 | 7 | | June | 6/17/2020 | 6/26/2020 | 80 | 10 | 10 |
| | July | 7/6/2010 | 7/9/2010 | 45 | 4 | 4 | | September | 9/4/2020 | 9/11/2020 | 80 | 8 | 8 |
| | September | 8/31/2010 | 9/11/2010 | 80 | 12 | 8 | | November | 11/10/2020 | 11/15/2020 | 45 | 6 | 6 |
| | November | 11/2/2010 | 11/15/2010 | 79 | 14 | 8 | | | | | | | |
| 2011 | March | 3/22/2011 | 3/30/2011 | 80 | 9 | 7 | 2021 | March | 3/15/2021 | 3/18/2021 | 35 | 4 | 4 |
| | May | 5/26/2011 | 6/1/2011 | 79 | 7 | 7 | | June | 6/1/2021 | 6/8/2021 | 80 | 8 | 8 |
| | July | 7/7/2011 | 7/13/2011 | 79 | 7 | 7 | | September | 8/30/2021 | 9/10/2021 | 80 | 11 | 9 |
| | September | 9/1/2011 | | 79 | 8 | 8 | | November | | 11/13/2021 | 45 | 5 | 5 |
| | November | | 11/10/2011 | 78 | 9 | 8 | | | | | | | |

Table 2. Cruise dates and number of stations completed during ChesMMAP cruises 2002-2021.

Table 3. Number of specimens collected, measured and processed for age determination and diet composition information from ChesMMAP, 2002 – 2021.

| Year | Fish Collected | Fish Measured | Otoliths Collected | Otoliths Processed | Stomachs Collected | Stomachs Processed |
|-------|----------------|------------------|-----------------------|-----------------------|-----------------------|-----------------------|
| 2002 | 32,014 | 23,602 | 5,657 | 4,503 | 4,875 | 3,041 |
| 2003 | 30,914 | 20,819 | 4,246 | 3,058 | 3,767 | 2,423 |
| 2004 | 47,619 | 31,242 | 5,483 | 4,290 | 4,721 | 3,325 |
| 2005 | 45,201 | 36,906 | 6,358 | 5,006 | 5,359 | 3,429 |
| 2006 | 43,957 | 31,247 | 5,416 | 4,230 | 4,403 | 3,503 |
| 2007 | 30,895 | 22,129 | 4,284 | 3,276 | 3,671 | 2,868 |
| 2008 | 26,302 | 19,598 | 4,209 | 3,048 | 3,678 | 3,432 |
| 2009 | 22,051 | 15,698 | 3,228 | 2,264 | 2,729 | 2,643 |
| 2010 | 26,333 | 20,565 | 4,003 | 2,677 | 3,424 | 3,237 |
| 2011 | 21,182 | 16,397 | 3,429 | 2,017 | 2,742 | 2,525 |
| 2012 | 17,328 | 14,954 | 2,496 | 1,519 | 2,014 | 1,732 |
| 2013 | 21,369 | 14,623 | 2,739 | 1,646 | 1,939 | 1,375 |
| 2014 | 11,316 | 7,807 | 1,740 | 988 | 893 | 837 |
| 2015 | 22,981 | 13,011 | 2,502 | 1,439 | 1,137 | 1,126 |
| 2016 | 23,924 | 18,042 | 3,341 | 1,517 | 1,606 | 1,587 |
| 2017 | 21,246 | 16,516 | 3,169 | 1,625 | 1,887 | 1,863 |
| 2018 | 18,806 | 13,150 | 2,343 | 1,093 | 1,273 | 1,264 |
| 2019 | 200,637 | 78,995 | 5,094 | 3,078 | 2,738 | 2,709 |
| 2020 | 272,276 | 87,766 | 3,486 | 2,225 | 1,652 | 1,631 |
| 2021 | 163,248 | 73,754 | 4,181 | 2,653 | 2,135 | 2,107 |
| Total | 1,099,599 | 576,821 | 77,404 | 52,152 | 56,643 | 46,657 |

Table 4. Average yearly catch for 2014 through 2018 compared to the average yearly catch for 2019 through 2021 for which either value exceeded 100 specimens. Species for which abundance indices are reported are highlighted.

| | Average | Average | Percent | | Average | Average | Percent |
|-------------------------|-------------|-------------|------------|--------------------|-------------|-------------|------------|
| Species | 2014 - 2018 | 2019 - 2021 | Difference | Species | 2014 - 2018 | 2019 - 2021 | Difference |
| Atlantic Brief Squid | 27.8 | 1,092 | 3,828.1 | Inshore Lizardfish | 5.6 | 330 | 5,798.8 |
| Atlantic Croaker | 1,314.6 | 18,269 | 1,289.7 | Kingfish Sp. | 289.0 | 3,682 | 1,174.2 |
| Atlantic Cutlassfish | 10.8 | 2,589 | 23,875.3 | Longfin Squid | 5.3 | 126 | 2,300.0 |
| Atlantic Menhaden | 769.2 | 898 | 16.8 | Mantis Shrimp | 192.2 | 189 | (1.7) |
| Atlantic Moonfish | 84.8 | 708 | 734.9 | Northern Puffer | 267.0 | 89 | (66.8) |
| Atlantic Spadefish | 80.2 | 334 | 316.5 | Northern Searobin | 692.8 | 347 | (50.0) |
| Atlantic Thread Herring | 24.6 | 9,604 | 38,940.7 | Pigfish | 3.6 | 110 | 2,964.8 |
| Alewife | 16.0 | 102 | 535.9 | Scup | 305.4 | 962 | 215.1 |
| Banded Drum | 5.4 | 166 | 2,967.9 | Silver Perch | 100.6 | 2,781 | 2,664.4 |
| Bay Anchovy | 752.6 | 23,676 | 3,045.9 | Spot | 1,124.0 | 91,304 | 8,023.1 |
| Black Seabass | 21.4 | 489 | 2,183.5 | Spotted Hake | 51.4 | 398 | 673.3 |
| Blue Crab - ad. fem. | 1,826.0 | 790 | (56.7) | Star Drum | 1.0 | 374 | 37,266.7 |
| Blue Crab - juv. fem. | 104.6 | 126 | 20.2 | Striped Anchovy | 32.2 | 1,788 | 5,451.8 |
| Blue Crab - male | 303.0 | 151 | (50.3) | Striped Bass | 868.2 | 1,792 | 106.4 |
| Bluefish | 66.0 | 163 | 147.5 | Striped Burrfish | 108.4 | 57 | (47.7) |
| Butterfish | 328.6 | 1,671 | 408.5 | Summer Flounder | 101.8 | 295 | 189.8 |
| Clearnose Skate | 127.2 | 97 | (23.5) | Weakfish | 907.8 | 19,858 | 2,087.4 |
| Harvestfish | 784.4 | 1,507 | 92.2 | White Perch | 7,163.0 | 10,443 | 45.8 |
| Hogchoker | 26.8 | 607 | 2,164.9 | White Shrimp | 33.0 | 1,988 | 5,925.3 |
| | | | | Total | 18,926.3 | 199,951 | 956.5 |

Tasks 2-4 – Data Summaries

The data summaries in this report represent a subset of the biological and ecological analyses which could be calculated from the ChesMMAP data set. For those species which are well-sampled by the survey, overall abundance estimates are presented. Relative abundance index calculations were based on limiting the data used for each species to the months, regions, and depth strata of maximum abundance over all years (Table 5). Those limiting parameters have been updated for some species based on subsequent analyses conducted during 2010 and 2012 (but not presented here). For species for which age-specific indices can be calculated, those indices are shown in both graphical and tabular formats.

Length-frequency (for sexes combined and sex-specific for most species), age-frequency (for those species for which ageing has been substantially completed) and overall diet summaries are also presented. Age-frequency figures are given both in histogram format showing the 'raw' number at age expanded to the total catch (i.e. as if every specimen captured had been aged) and in standardized bubble plot format with the 'raw' figures standardized to the total number of minutes towed in a full ChesMMAP year if each cruise consisted of it full complement of completed stations (8,000 trawl minutes during 2002 – 2018, 4,800 trawl minutes subsequently). The bubble plots allow a representation of the age-specific abundance for all years simultaneously and can sometimes make it easier for the reader to follow large and small year-classes diagonally through the population.

Table 5. Selected months, regions, and depth strata data used for abundance indices for each species (cells colored in the lighter shade of blue not used for Age-0 index calculations).

| Atlantic Cro | bake | r | | | | | | | | | | | | | | | | | | | | | | |
|---|--------|-------|------|--------|-------|----------|---------------------|----------|------|----------|-------|------|-------|------------|------|--|--------|-----------|-------|------|--------|------|-------|------|
| 2002 - 2018 | | | | | | | | _ | | | - | | | | | 2019 and f | orwa | rd | | | | | | |
| | Ν | /larc | h | | May | v | July Sept. November | | | | | | | March June | | | ne | Sept. | | Nov. | | | | |
| Region | Shlw | Mid | Deep | | _ | Deep | Shlw | - | Deep | Shlw | Mid | Deep | Shlw | Mid | Deep | Region | Shlw | Deep | Shlw | Deep | Shlw | | Shlw | Deep |
| MD-North | | | | | | | | | | | | | | | | MD-North | | | | | | | | |
| MD-Mid | | | | | | | | | | | | | | | | MD-South | | | | | | | | |
| MD-South | | | | | | | | | | | | | | | | VA-North | | | | | | | | |
| VA-North | | | | | | | | | | | | | | | | VA-South | | | | | | | | |
| VA-South | | | | | | | | | | | | | | | | | | | | | | | | |
| | | | | | | | | | | | | | | | | | | | | | | | | |
| Black Sea B | | | | | | | | | | | | | | | | | | | | | | | | |
| 2002 - 2018 | | | | | | | r — | | | - | | | r | | | 2019 and f | | | | | - | | | |
| | Ν | /larc | h | | May | <u>/</u> | | July | (| | Sept | :. | No | vem | ber | | Ma | rch | Ju | ne | Se | pt. | No | ov. |
| Region | Shlw | Mid | Deep | Shlw | Mid | Deep | Shlw | Mid | Deep | Shlw | Mid | Deep | Shlw | Mid | Deep | Region | Shlw | Deep | Shlw | Deep | Shlw | Deep | Shlw | Deep |
| MD-North | | | | | | | | | | | | | | | | MD-North | | | | | | | | |
| MD-Mid | | | | | | | | | | | | | | | | MD-South | | | | | | | | |
| MD-South | | | | | | | | | | | | | | | | VA-North | | | | | | | | |
| VA-North | | | | | | | | | | | | | | | | VA-South | | | | | | | | |
| VA-South | | | | | | | | | | | | | | | | 1 | 1 | | - | - | | | | - |
| | | | | | | | | | | | | | | | | | | | | | | | | |
| Bluefish | | | | | | | | | | | | | | | | 2010 (| | | | | | | | |
| 2002 - 2018 | | _ | | | | | I | | | | _ | | | | | 2019 and f | T | | - | | | | | |
| <u> </u> | | /larc | | | May | _ | I | July | | | Sept | | | | ber | | | rch | | ne | Se | | No | - |
| Region | Shlw | Mid | Deep | Shlw | Mid | Deep | Shlw | Mid | Deep | Shlw | Mid | Deep | Shlw | Mid | Deep | Region | Shlw | Deep | Shlw | Deep | Shlw | Deep | Shlw | Deep |
| MD-North | | | | | | | | | | | | | | | | MD-North | | | | | | | | |
| MD-Mid | | | | | | | | | | | | | | | | MD-South | | | | | | | | |
| MD-South | | | | | | | | | | | | | | | | VA-North | | | | | | | | |
| VA-North | | | | | | | | <u> </u> | | | | | | | | VA-South | | | | | | | | |
| VA-South | | | | | | | | | | | | | | | | | - | | | | | | | _ |
| Buttorfish | | | | | | | | | | | | | | | | | | | | | | | | |
| Butterfish 2002 - 2018 | | | | | | | | | | | | | | | | 2019 and f | 0.000 | rd. | | | | | | |
| 2002 - 2018 | | | L. | | | | I | 1 I. | | 1 | C | | | | | 2015 and 1 | | | | | 6 | | | |
| De el eur | _ | /larc | | | May | | | July | - | | Sept | | | _ | ber | De si su | - | rch | | ne | Se | | No | - |
| Region | Shlw | Mid | Deep | Shlw | Mid | Deep | Shlw | Mid | Deep | Shlw | Mid | Deep | Shlw | Mid | Deep | Region | Shlw | Deep | Shlw | Deep | Shlw | Deep | Shlw | Deep |
| MD-North MD-Mid | | | | | | | | | | | | | | | | MD-North MD-South | | | | | | | | |
| MD-South | | | | | | | | - | | | | | | | | VA-North | - | | | | | | | |
| VA-North | | | | | | | | | | | | | | | | VA-North VA-South | | | | | | | | |
| VA-North VA-South | | | | | | | | | | | | | | | | VA-South | | | | | | | | |
| VA-South | | | | | | | | - | | | | | | | | 1 | 1 | | | | | | | |
| Kingfish | | | | | | | | | | | | | | | | | | | | | | | | |
| 2002 - 2018 | | | | | | | | | | | | | | | - i | 2019 and f | orwa | rd | | | | | | |
| | | /larc | h | | May | | | July | , | | Sept | ÷ | No | verr | ber | | | rch | In | ne | Se | nt | No | NV. |
| Region | | | Deep | | | Deep | Shlw | - | Deep | | | Deep | | _ | Deep | Region | | Deep | | Deep | Shlw | | Shlw | - |
| MD-North | 311100 | witu | Deep | 311100 | IVITU | Deep | 311100 | Ivitu | Deep | 311100 | IVITU | Deep | SIIIW | IVITU | Deep | MD-North | 311100 | Deep | SIIIW | Deep | 311100 | Deep | SIIIW | Deep |
| MD-Mid | | | | | | | | | | | | | | | | MD-South | | | | | | | | |
| MD-South | | | | | | | | | | | | | | | | VA-North | | | | | | | | |
| VA-North | | | | | | | | | | | | | | | | VA-South | | | | | | | | |
| VA-South | | | | | | | | | | | | | | | | | | | | | | | | |
| | | | | | | | | | | | | | | | | | | | | | | | | |
| Northern P | uffe | r | | | | | | | | | | | | | | | 1 | | | | | | | |
| 2002 - 2018 | | | | | | | | | | | | | | | | 2019 and f | orwa | rd | | | | | | |
| | Ν | /larc | h | | May | v | Γ | July | / | 1 | Sept | t. | No | vem | ber | | Ma | rch | Ju | ne | Se | pt. | No | ov. |
| | | _ | Deep | | | Deep | Shlw | - | Deep | | | Deep | | - | Deep | Region | | Deep | | Deep | Shlw | | Shlw | _ |
| Region | Shlw | Mid | | | | 1.10 | | | | | | 1.20 | | | | MD-North | | | | 1.00 | | | | |
| Region MD-North | Shlw | Mid | | | | | | | - | | | | | | | MD-South | | | | | | | | |
| MD-North | Shlw | Mid | | | | | | | | | | | | | | | | | | | | | | |
| MD-North MD-Mid | Shlw | Mid | | | | | | | | | | | | | | VA-North | | | | | | | | |
| MD-North | Shlw | Mid | | | | | | | | | | | | | | VA-North VA-South | | | | | | | | |
| MD-North MD-Mid MD-South | Shlw | Mid | | | | | | | | | | | | | | | | | | | | | | |
| MD-North MD-Mid MD-South VA-North | Shlw | Mid | | | | | | | | | | | | | | | | | | | | | | |
| MD-North MD-Mid MD-South VA-North | Shlw | Mid | | | | | | | | | | | | | | | | | | | | | | |
| MD-North MD-Mid MD-South VA-North VA-South | | Mid | | | | | | | | | | | | | | | orwa | rd | | | | | | |
| MD-North MD-Mid MD-South VA-North VA-South Scup | | Aaro | h | | May | | | July | / | | Sept | | No | vem | ber | VA-South | | rd rch | Ju | ne | Se | pt. | No |)v. |
| MD-North MD-Mid MD-South VA-North VA-South Scup 2002 - 2018 | | Aaro | h | | | Deep | Shlw | | Deep | | _ | Deep | | - | - | VA-South 2019 and f | Ma | | | ne | | | Nc | _ |
| MD-North MD-Mid MD-South VA-North VA-South Scup | | Aaro | | | | | Shlw | | | | _ | - | | - | ber | VA-South | Ma | rch | | | Se | | | _ |
| MD-North MD-Mid MD-South VA-North VA-South Scup 2002 - 2018 Region | | Aaro | | | | | Shlw | | | | _ | - | | - | - | 2019 and for Region | Ma | rch | | | | | | _ |
| MD-North MD-Mid MD-South VA-North VA-South Scup 2002 - 2018 Region MD-North | | Aaro | | | | | Shlw | | | | _ | - | | - | - | VA-South 2019 and fo Region MD-North | Ma | rch | | | | | | _ |
| MD-North MD-Mid MD-South VA-North VA-South Scup 2002 - 2018 Region MD-North MD-North | | Aaro | | | | | Shlw | | | | _ | - | | - | - | 2019 and for Region MD-North MD-South | Ma | rch | | | | | | _ |

Table 5. cont.

| Spot | | | | | | | | | | | | | | | | | | | | | | | | |
|--|---|-----------------------------|----------------------------------|----------|-------------------|------------------|----------|----------------------------|------------------|--------------------------------------|-------------|----------------------------------|--------------------|-------------------|----------------------|--|-------------|---|--------|-----------------------|----------------------------|------------------------|--------|-----------------------------|
| 2002 - 2018 | 3 | | | | | | | | | | | | | | - | 2019 and fo | orwar | ď | | | | | | |
| | | Лаго | h | | May | | | July | , | | Sept | t. | No | vem | ber | | March June | | | ne | Sept. | | Nov. | |
| Region | | - | Deep | | | Deep | Shlw | | Deep | | <u> </u> | Deep | - | | Deep | Region | | Deep | | Deep | | Deep | | Deep |
| MD-North | | | | | | | | | | | | | | | | MD-North | | | | | | | | |
| MD-Mid | | | | | | | | | | | | | | | | MD-South | | | | | | | | |
| MD-South | | | | | | | | | | | | | | | | VA-North | | | | | | | | |
| VA-North | | | | | | | | | | | | | | | | VA-South | | | | | | | | |
| VA-South | | | | | | | | | | | | | | | | | | | | | | | | |
| | | | | | | | | | | | | | | | | | | | | | | | | |
| Striped Ba | | 1arcl | 1 | | | | | | | | | | | | | | | | | | | | | |
| 2002 - 2018 | 3 | | | | | | | Mar | ch Ind | lices | | | | | | 2019 and fo | orwa | ď | | | | | - | |
| | N | Лаго | h | | May | / | | July | , | 2 | Sept | t. | No | vem | ber | | Ma | rch | Ju | ne | Se | pt. | No | ov. |
| Region | Shlw | Mid | Deep | Shlw | Mid | Deep | Shlw | Mid | Deep | Shlw | Mid | Deep | Shlw | Mid | Deep | Region | Shlw | Deep | Shlw | Deep | Shlw | Deep | Shlw | Deep |
| MD-North | | | | | | | | | | | | | | | | MD-North | | | | | | | | |
| MD-Mid | | | | | | | | | | | | | | | | MD-South | | | | | | | | |
| MD-South | | | | | | | | | | | | | | | | VA-North | | | | | | | | |
| VA-North | | | | | | | | | | | | | | | | VA-South | | | | | | | | |
| VA-South | | | | | | | | | _ | | | | | | | | | | | | | | | 1 |
| Chains of P | | | | | | | | | | | | | | | | | | | | | | | | |
| Striped Bas 2002 - 2018 | | ove | nper | | | | l | Ner | embe | l vr l m ri | lice | | | | | 2019 and fo | | d | | | | | | l |
| 2002 - 2018 | 1 | - | | | | | | | | | | | • • | | | 2019 and fo | | | - | | 6 | | | |
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The profiles that follow are organized first by species and then by type of analysis ('Task'). Each Task element (single-species stock parameter summarizations, trophic interaction summaries, and estimates of abundance) is included but is not labeled with a Task number and is not necessarily shown in Task number order.

For each species, the following data summaries are presented (note that some data/analyses may not be available for all species):

- 1) A table which summarizes the numbers and biomass captured and measured during each survey year as well as the numbers of ageing structure and stomach samples preserved and processed.
- 2) A series of GIS figures showing total abundance (kg / 10,000 m²) at each sampling site overlaid on the survey depth strata, for each cruise during the year.
- 3) Figures and tables presenting overall and age-specific (for appropriate species) area-sweptcorrected abundance indices by number and biomass, calculated using geometric means.
- 4) Length-frequency data by year, for sexes combined and separately.
- 5) Age-frequency distributions by year (for those species where appreciable numbers have been captured and otoliths have been processed) in both histogram and bubble plot format, as described above.
- 6) Diet analyses by weight and number, using all data collected and analyzed 2002-2020. For this report (and for presentation elsewhere), standardized categories of prey types (Fishes, Crustaceans, Molluscs, Worms, Misc.) have been developed for all ChesMMAP species. In each figure for each predator species, these categories are presented in decreasing order of importance and within each broad category specific prey types are shown also in decreasing order. Only those specific prey types greater than or equal to 1.0% of the overall diet are shown (unless the entire category is less than 1.0%). All other specific prey are lumped into a category called ' x other' (x = fishes, molluscs, etc.) which is distinct from unidentified prey types within the category. For the reader's convenience, the color scheme used for all species (e.g. red = crustaceans, light blue = fishes, etc.) is the same. This makes it relatively easy to compare figures across predator species or by weight/number within a species.

Species Data Summaries

Atlantic Croaker (Micropogonias undulatus)

<u>Abundance:</u> Atlantic Croaker is typically among the most abundant species in ChesMMAP survey catches, especially during the mid-year. During the years 2002 through 2007 at least 12,000 specimens totaling 2,600kg or more were captured. Between 2008 and 2012 no more than half that number were ever captured. In 2013 the number rose to about 9,000 specimens but the years 2014 through 2018 were the five lowest catch years in the 17-year time series with no more than 1,723 specimens in any year. In 2019 with the new sampling gear the total number and biomass captured rose by a factor of 10 to 11,685 fish weighing 920kg and then tripled to a time series high 34,291 specimens in 2020. However, these 34,000 fish only weighed a total of 1,817kg, indicating that the new sampling gear is far more efficient at capturing smaller specimens of this species. In 2021 the smallest number of Atlantic Croaker (8,832) were captured than in any of the years since the ChesMMAP gear change (Table 6).

The majority of fish are captured in regions 4 and 5 (Virginia). In years of higher abundance specimens are regularly captured in all survey regions and in June and September of 2020 and 2021 Atlantic Croaker were captured at the northernmost Maryland stations (Region A) but were relatively scarce in

Region B. Catches decline in November as this summer resident species leaves bay waters though in 2021 Croaker were still captured in about half of the November tows (Figure 1).

Through 2018 relative abundance indices were calculated using data collected during May, July and September, in Regions 4 and 5 from only the mid-depth and deep strata. With the 2019 restratification, samples used were from Regions C and D in June and September, and both depth strata (Table 5). As reflected in the trends of total catch, indices for all ages combined, both in numbers and biomass reveal low values in 2002 and 2003 that were followed by a period of high abundance throughout 2004-2007 then very low abundances from 2008 through 2018 (Table 7, Figure 2). Anecdotal information as well as trends in commercial and recreational landings suggests that this period of low abundance in ChesMMAP samples is representative of a coast wide phenomenon and may be related to cyclical abundances that have been observed in the past, though this continued period of very low abundance is concerning. Age-specific abundances are shown for ages 0 through 7+ and largely follow the same pattern as described above. Abundances along the coast as measured by the nearshore North East Area Monitoring and Abundance (NEAMAP) survey showed a steep rise in the fall of 2012 following into the spring of 2013 and those fish may be represented by the moderate rise in the Age-0 ChesMMAP index in 2013. Nearly all of the increase in abundance in 2019 through 2021 was in Age-0 fish. How much of this increase represents a true rise and how much is attributable to the increased efficiency of the sampling gear will be revealed when the full calibration data are available. For ages 2 and older the pattern of abundance generally follows that for overall abundance which indicates that to some extent at least, availability of this species to the ChesMMAP survey area (i.e. the proportion of the coastal stock that invades the bay during warm months) may play at least some role in determining abundance as estimated by ChesMMAP.

<u>Length and Age:</u> Specimens between 14mm and 499mm in total length (Figure 3) and between age 0 and 17 (Figure 4, Figure 5) appear in survey data. In recent years both the length frequency distribution and the age structure appear to have become increasingly truncated towards smaller and younger fish; most individuals range between 150mm and 350mm and ages 1-5. In most years, no particular pattern of differences in sex-specific length frequencies were observed though in 2018 the samples were heavily weighted towards females.

The length distribution of this species changes considerably year-to-year as year-classes of either extremely high or extremely low abundance move through the stock. For example, a highly abundant 2002 year-class was seen as a peak in the length-frequency histograms in each year from 2003 through 2007 and as a distinctly abundant year-class in the age-frequency figures even into 2008. There appears to be evidence of a mildly to highly successful year-class in 2006 which was still abundant in 2007 and 2008 and which was present in ChesMMAP samples until 2013, as 7-year olds. Conversely, the 2007 year-class appears to have been nearly absent in Chesapeake Bay and similarly was not abundant in 2008. In 2009 these two-year-old fish were the most abundant age-class but the number captured was very low compared with other years. A moderately abundant 2012 year-class was still present in 2018 in very low numbers. Though the original ChesMMAP sampling gear was not efficient at capturing age-0 Croaker, when substantial numbers are present it can be an indicator to expect larger numbers the following year as age-1s. In 2018 the second largest number of age-0 Croaker in the ChesMMAP time series was observed. In 2020 and 2021 the vast majority of specimens were age-0 and age-1 and only 2 full-processed specimens were older than age-2. This likely represents a shrinking of the age structure at low stock abundance.

Croakers to age 8 are not uncommon for this survey. During 2008, program personnel attended an Atlantic Croaker ageing workshop sponsored by the Atlantic States Marine Fisheries Commission. The consensus report from that workshop set a birth date of 1 January each year, as that date is the approximate mid-point of spawning in the southern portion (i.e., south of Cape Hatteras) of the species' range. Spawning north of Hatteras, including Virginia's waters, occurs several months earlier, and is often complete by early December. As a result, all Croaker ages in the ChesMMAP data base were adjusted down one year and in this survey it is now possible to capture fish classified as age-negative 1. This occurs when fish spawned in late summer and autumn of a given year are collected during the September or November cruises of that year. Those fish are not considered age-0 (or young-of-the year) until that upcoming January, so to place them in the correct year-class, they are assigned an age-negative 1. This phenomenon can cause a bit of confusion in interpreting year-and-age-specific indices of abundance and we hope to influence a change in the age-assignment protocol in the future.

<u>Diet:</u> Measured as prey biomass, various identifiable and non-identifiable polychaetes as well as other worm species represent the largest single prey type in the diet of Atlantic Croaker (39.8% by weight) followed by those prey classified as miscellaneous and unidentifiable items (33.0%). When measured by number though these two major categories are reversed (33.8% miscellaneous prey and 31.8% worms). The unidentified material is likely made up largely of worms and soft-bodied molluscs. Small bodied crustaceans (e.g. mysid shrimp) constitute the third major prey category by weight totaling 11.3% by weight and 22.9% by number. Several clam and mussel prey types contribute 10.5% and 8.7% of croaker diets by weight and number (fourth largest taxonomic category) respectively. Fishes constitute very minor amounts (5.4% W, 3.0% N) of Croaker diets (Figure 6).

Atlantic Croaker Tables and Figures:

Table 6. Atlantic Croaker sampling rates and preserved specimen analysis status by year.

Figure 1. Station specific biomass of Atlantic Croaker in Chesapeake Bay, 2021.

Table 7. Atlantic Croaker geometric mean indices of abundance, by number and biomass, overall and by age-class.

Figure 2. Atlantic Croaker geometric mean indices of abundance, by number and biomass, for all ages combined and by age-class.

Figure 3. Atlantic Croaker length-frequency in Chesapeake Bay 2002-2021, overall (A) and by sex (B).

Figure 4. Atlantic Croaker age-frequency by year, 2002-2021.

Figure 5. Atlantic Croaker age-frequency by year, 2002-2021 standardized to 8,000 (4,800 for 2019 and forward) annual trawl minutes.

Figure 6. Diet composition, expressed as percent by weight (A) and percent by number (B) of Atlantic Croaker collected during ChesMMAP cruises in 2002-2018 combined.

Black Sea Bass (Centropristis striata)

<u>Abundance</u>: Trawling is generally not considered to be a particularly effective method for sampling a structure-oriented species such as Black Sea Bass. Indeed the original ChesMMAP survey gear never captured more than 50 specimens in any survey year. However the '200' net caught 445, 507 and 514 total specimens in 2019, 2020 and 2021 respectively and the species is present in at least 50% of tows at its index stations (Table 8). Stations with the largest catches are concentrated along the edges of the various bay channels (Figure 7). In future years it is reasonable to expect that a reliable age-1 abundance index will result.

For purposes of calculating abundance indices, stations used include (Table 5):

- Through 2018: July, September, November Regions 4 and 5 All depth strata.
- 2019 and forward: June, September, November, Regions C and D Both depth strata.
- Age-0 indices do not use data from the June/July cruises.

Overall relative abundance indices expressed either in numbers or biomass exhibit nearly identical interannual patterns, indicating that the sizes of captured specimens are relatively constant year to year. A steady decline in measured abundances between 2002 and 2006 was followed by a period of fluctuating high and low values until 2011 when the index was in the middle range of the time series. In 2013 only two Black Sea Bass were captured and the indices found new time-series lows for both number and biomass. Between 2013 there was a definite upward trend which greatly accelerated in 2016 with the numerical index reaching a time-series high value and the biomass index at the third highest value in the series. In 2017 the indices fell to approximately average values and in 2018 fell again to approximately the time series low. The extremely high values for 2019 and 2020 are likely primarily due to the greater capture efficiency of the '200' net but this will have to be borne out during calibration studies (Table 9, Figure 8). Age-specific abundance indices indicate that age-1 fish are primarily available to ChesMMAP and between 2002 and 2018 follow a similar general downward trend, with occasional single-year upward ticks. The beginning of the new time series in 2019 indicates increasing abundance over three years. As catch rates for this species are low and inconsistent confidence limits on the abundance estimates between 2002 and 2018 are comparatively broad. With more consistent catches in the new sampling regime confidence bands should narrow.

<u>Length and Age:</u> Specimens captured in the survey tend to be relatively small (<250mm) and young (age-0 and age-1) though individuals up to 270mm total length have been sampled (Figure 9). Due to the small sizes of most individuals captured by ChesMMAP, the majority of specimens observed of this protogynous hermaphroditic species have been females. During 2012 the previous backlog of otolith samples was cleared and all otoliths collected have been assigned ages. Age-frequencies reveal that in most years the survey catches are dominated by age-0 and primarily age-1 specimens (Figure 10, Figure 11).

<u>Diet:</u> Though the size range of samples is limited, the diet data may be the most valuable ChesMMAP contribution for this species. Small bodied crustaceans (64.8% W, 71.9% N) contribute the highest portions of the diet, among identifiable prey. A variety of worms (12.4% W, 10.8% N), miscellaneous (9.4% W. 8.0% N), fishes (8.0% W, 6.0% N) and molluscs (3.2% W, 3.4% N) comprise the remainder of the diet. Amphipods, mysids, isopods and mud crabs constitute the largest identifiable prey species (Figure 12).

Black Sea Bass Tables and Figures:

Table 8. Black Sea Bass sampling rates and preserved specimen analysis status by year.

Figure 7. Station specific biomass Black Sea Bass in Chesapeake Bay, 2021.

Table 9. Black Sea Bass geometric mean indices of abundance, by number and biomass, overall and by age-class

Figure 8. Black Sea Bass geometric mean indices of abundance, by number and biomass, for all ages combined and by age-class.

Figure 9. Black Sea Bass length-frequency in Chesapeake Bay 2002-2021, overall (A) and by sex (B).

Figure 10. Black Sea Bass age-frequency by year, 2002-2021.

Figure 11. Black Sea Bass age-frequency by year, 2002-2021 standardized to 8,000 (4,800 for 2019 and forward) annual trawl minutes.

Figure 12. Diet composition, expressed as percent by weight (A) and percent by number (B) of black seabass collected during ChesMMAP cruises in 2002-2020 combined.

Bluefish (Pomatomus saltatrix)

<u>Abundance</u>: Due to the fast-swimming and pelagic nature of Bluefish, this species is not considered to be well sampled by ChesMMAP, though some useful assessment-related information can be generated from these survey data. Catch rates have increased substantially for small fish (primarily age-0 and age-1) with the new sampling gear. Between 2002 and 2018, no more than 126 (in 2010) and 125 (in 2015) Bluefish were captured during a ChesMMAP sampling year. The total number remained low with the change in sampling gear in 2019 but jumped to a series high 208 fish in 2020. The maximum biomass captured was about 32kg in 2003 due to a few large specimens which were captured that year. In 2016 through 2018 moderate numbers of specimens (36, 40, 85 respectively) were sampled. The past two years (2020, 2021) saw the largest number of Bluefish captured during the time series. Whether this is primarily due to increased abundance or increased efficiency of the sampling gear is currently unknown. Except for a very small number of mishandled or mislabeled specimens all preserved stomachs and otoliths have been fully processed (Table 10).

When captured, typically between just one and four specimens occur in a tow (Figure 13) though as many as 85 have been collected in a single sampling event. Bluefish are usually captured in either the shallow (10'-30') or mid-depth (30'-50') strata. Catches are typically highest late in the year, presumably as the young-of-the year fish are moving into deeper waters in preparation for outmigration from the bay.

Abundance is normally highest in regions 4 and 5 (C and D) but notable exceptions occur such as a single capture of 26 specimens in Region 1 during the September 2008 cruise (Bonzek et al. 2009).

Abundance indices are calculated using data from (Table 5):

- Through 2018: September and November cruises Regions 3 and 4 (this is an update from previous reports) all Depth strata.
- 2019 and forward: September and November cruises Regions C and D Depths A and B.

Abundance indices for all ages of Bluefish combined alternated between low and high values from 2002 to 2007, were consistently at time series lows between 2008 and 2011, exhibited a moderately rising pattern between 2012 and 2015, fell again in 2016 and rose slightly in both 2017 and 2018, fell in 2019 and rose substantially in 2020 and 2021 though the 2019 through 2021 data are not yet calibrated with previous years (Table 11, Figure 14). Patterns between indices by number and weight are very similar. As nearly all specimens captured are young-of-year fish, the age-0 index closely follows the pattern for the overall index.

<u>Length and Age</u>: Most individuals sampled in the survey are less than 350mm fork length and, due to the small number of specimens captured and to the protracted spawning season of this species, it is difficult to differentiate cohorts in length frequencies (Figure 15). No pattern of sexual differentiation by size has been observed. Nearly all ChesMMAP Bluefish are either age-0 or age-1 and in most years the majority of specimens captured are age-0 (Figure 16, Figure 17).

<u>Diet</u>: Diet data presented here are consistent with previous studies in showing that Bluefish are highly piscivorous (Figure 18). For the 442 specimens examined, which represent 268 clusters, Bay Anchovy constitute 53.4% of the diet by weight and 52.0% by number, while Spot (9.3% W, 5.8% N) are the other major identifiable fish prey. All fish species together represent 88.9 by weight and 83.0% by number. Crustaceans, mainly mysid shrimp at 4.4% W and 4.7% N, and sand shrimp (1.8% W, 5.9% N) represent most of the remainder.

Bluefish Tables and Figures:

Table 10. Bluefish sampling rates and preserved specimen analysis status by year.

Figure 13. Station specific biomass of Bluefish in Chesapeake Bay, 2021.

Table 11. Bluefish geometric mean indices of abundance, by number and biomass, overall and age-0.

Figure 14. Bluefish geometric mean indices of abundance, by number and biomass, for all ages combined and for ages 0 and 1.

Figure 15. Bluefish length-frequency in Chesapeake Bay 2002-2021, overall (A) and by sex (B).

Figure 16. Bluefish age-frequency by year, 2002-2021.

Figure 17. Bluefish age-frequency by year, 2002-2021 standardized to 8,000 (4,800 for 2019 and forward) annual trawl minutes.

Figure 18. Diet composition, expressed as percent by weight (A) and percent by number (B) of Bluefish collected during ChesMMAP cruises in 2002-2020 combined.

Butterfish (Peprilus triacanthus)

<u>Abundance</u>: Butterfish are moderately abundant in ChesMMAP survey tows with several hundred to over 2,600 specimens typically captured during any survey year (Table 12). Numerically, 2019 saw the fourth highest catch rate of the time series and 2020 resulted in the highest number recorded at 2,616 with a moderate decrease to 1,569 in 2021. Butterfish abundance follows a generally predictable annual pattern, building from near-zero during March, increasing (albeit low) abundance through the spring and summer, and reaching a maximum generally during the September and November cruises (Figure 19).

Abundance indices are generated from survey tows during the peak months of September and November in Regions 4 and 5 (C and D). During 2017 the depth strata used to calculate abundance for this species were reexamined. Previously only the mid-depth strata were included but the reexamination indicated that all three depth strata should be used (Table 5). Abundance indices generally varied without trend between 2002 and 2009 then followed a generally declining trend, reaching a time-series low value in 2018. At the request of assessment analysts at NEFSC indices for ageclasses up to age-4+ (previously only up to age-2+) are now generated. Based on 2019, 2020 and 2021 data it appears that the new trawl gear is far more efficient at capturing age-0 specimens (Table 13, Figure 20).

<u>Length and Age:</u> Yearly length frequency diagrams (Figure 21) appear to reveal at least two year-classes of varying strength present in the Chesapeake Bay fish during any given year, however this will require further analysis. This program (and others) has found Butterfish collected from estuarine areas extremely difficult to age. We are still investigating methods to obtain accurate age determinations from otolith samples. Pending the results of those efforts the otoliths collected during earlier survey years have not been processed (age-specific abundance indices were calculated using age-length keys derived from NEAMAP data).

<u>Diet:</u> Analyses of Butterfish stomachs from early program years revealed a high percentage of generally unidentifiable gelatinous zooplankton and other unidentifiable items. It was determined that further analyses of Butterfish diets were not an efficient use of resources and the decision was made to discontinue preservation and analysis of Butterfish stomachs.

Butterfish Tables and Figures:

Table 12. Butterfish sampling rates and preserved specimen analysis status by year.

Figure 19. Station specific biomass of Butterfish in Chesapeake Bay, 2021.

Table 13. Butterfish geometric mean indices of abundance, by number and biomass, overall and by ageclass.

Figure 20. Butterfish geometric mean indices of abundance, by number and biomass, for all ages combined and by age-class.

Figure 21. Butterfish length-frequency in Chesapeake Bay 2002-2021, overall.

Kingfish (Menticirrhus spp.)

The ranges of three closely related species, the Northern Kingfish (*Menticirrhus saxatilis*), the Southern Kingfish (*Menticirrhus americanus*), and the Gulf Kingfish (*Menticirrhus littoralis*) overlap in Chesapeake Bay. While some specimens are easily separable in the field, many are not. We have therefore adopted the practice of combining all of these specimens into a single category of Kingfish (*Menticirrhus spp.*). This practice is consistent with the manner in which these species are landed and reported in the fishery as well.

<u>Abundance</u>: Kingfish are moderately abundant in ChesMMAP tows with approximately 100-600 total specimens captured each year. In 2019, sampling with the '200' net total catch numbers increased approximately 10-fold to almost 3,900 specimens and increased still further in 2020 to almost 5,800 fish before declining (to a still third highest catch rate) to 1,409 in 2021 (Table 14). ChesMMAP catches for this species are almost exclusively in Regions 4 and 5 (C and D - lower bay), occur throughout the warm weather months and are often high into November (Figure 22).

Abundance indices are generated from these stations (Table 5):

- Through 2018: May, July, September, November Regions 4 and 5 All depth strata.
- 2019 and forward: June, September, November, Regions C and D Both depth strata.
- Age-0 indices do not use data from the May/June/July cruises.

Until 2010 it appeared that Kingfish had been on a nearly consistent increasing abundance trend throughout the survey years. However, between 2011 and 2014 a nearly seven-fold decline was observed in the indices back to levels observed at the beginning of the time series. This was followed by a slight increase in 2015 and then a very large uptick again in 2016 to time series high values for both numbers and biomass. In 2017 overall abundance decreased significantly but was still among the five highest values in the time series and in 2018 there was a slight increase to the fourth highest value in the time series. The very large increase in total numbers captured in 2019 - 2021 is reflected in the abundance indices though these data will be subjected to calibration factors when available. Age-specific ChesMMAP indices follow similar patterns with generally lower values through 2007, an increasing trend through 2010 or 2011, with a sharp decline in 2012, an uptick in 2013 and declines again in 2014, with a significant rise in 2016 and nearly stable values in 2017. The increase in overall abundance in 2018 was due to a large increase in age-0 fish while abundance for other ages were on par with 2017 values. The pre-calibrated 2019-2021 values were much higher for nearly all age-classes (Table 15, Figure 23).

<u>Length and Age</u>: Due to the relatively small number of specimens captured during early survey years and to the overlapping sizes-at-age, it is difficult to interpret length frequencies, though at least two cohorts are apparent in many years (Figure 24). No differential growth patterns between male and female Kingfish have been observed.

Specimens between ages 0 and 7 have been captured with most being age-4 or less. Year-classes of high (e.g. 2002) and low (e.g. 2004) abundance do seem to track through the stock from year to year, which indicates consistent survey sampling and otolith analysis. Relatively large numbers of age-0 and age-2 specimens were captured in 2009 but the number of age-3-and-older fish was very small. It is apparent that this species does not fully recruit to the *original* ChesMMAP sampling gear until at least age-1 and perhaps even age-2 but the new trawl appears to efficiently capture younger fish (Figure 25, Figure 26). As this species is not subjected to regular stock assessments the VIMS Multispecies Research

Group assigns it to lower level of priority for specimen processing so there is currently a backlog of unprocessed otoliths dating to 2012. Once true ages have been assigned the patterns observed in age-specific abundance indices will change.

<u>Diet:</u> The largest taxa of prey items in Kingfish stomachs are crustaceans (33.9% W, 40.8% N), primarily small shrimps and crabs. Worms and molluscs constitute the next largest portions (28.2% W, 20.0%N and 22.8% W, 22.4 %N respectively) of the diet, with fishes and miscellaneous items completing the diet (Figure 27).

Kingfish Tables and Figures:

Table 14. Kingfish sampling rates and preserved specimen analysis status by year.

Figure 22. Station specific biomass of Kingfish in Chesapeake Bay, 2021.

Table 15. Kingfish geometric mean indices of abundance, by number and biomass, overall and by ageclass.

Figure 23. Kingfish geometric mean indices of abundance, by number and biomass, for all ages combined and by age-class.

Figure 24. Kingfish length-frequency in Chesapeake Bay 2002-2021, overall (A) and by sex (B).

Figure 25. Kingfish age-frequency by year, 2002-2021.

Figure 26. Kingfish age-frequency by year, 2002-2021 standardized to 8,000 (4,800 for 2019 and forward) annual trawl minutes

Figure 27. Diet composition, expressed as percent by weight (A) and percent by number (B) of Kingfish collected during ChesMMAP cruises in 2002-2020 combined.

Northern Puffer (Sphoeroides maculatus)

<u>Abundance</u>: Abundance of Northern Puffer in ChesMMAP samples varies by an order of magnitude among years, with as many as 600 being captured in 2011 and as few as 41 in 2005 (Table 16). Typical patterns of abundance for this species in the survey are minimal numbers in spring and early summer, and a peak in abundance during the September and/or November cruises, perhaps as the summer residents are migrating toward offshore wintering grounds. Catches are consistently greatest in Regions 4 and 5 (C and D), though the species is common into Region 3 (Figure 28). Catch rates with the new sampling gear were comparable to those in previous years. As catches in the survey are patchy, estimates of abundance for this species are of unknown reliability.

Abundance indices are generated from these stations (Table 5):

- Through 2018: September, November Regions 4 and 5 All depth strata.
- 2019 and forward: September, November, Regions C and D Both depth strata.

The peak year for the relative abundance indices from survey data have (both in numbers and biomass) was in 2007, though the total number of specimens captured that year was merely average. This lone

high value likely was the result of restricted sampling that year which probably artificially inflated the index. Other years have varied without apparent trend though there was a modest increase in both 2015 and 2016, a decline in 2017 and a modest rise in 2018 and fall values in the last three years (Table 17, Figure 29).

<u>Length and Age</u>: Specimens between approximately 50mm and 270mm total length have been captured, though most individuals measured between 100mm and 250mm. The length composition varies year to year, likely as a result of varying year-classes entering and leaving the bay stock. There may be evidence that the new trawl gear samples some number of smaller specimens than were previously observed (Figure 30). However, as this is not a high priority species and as standard ageing protocols have not been established, ageing of vertebrae has not been attempted. The largest individuals captured have generally been females but there appears to be no overall pattern of differential growth between sexes.

<u>Diet:</u> Molluscs (34.5% W, 29.1% N), miscellaneous taxa including unidentified material (31.2%W, 31.0%N) and crustaceans, primarily small crab species (26.6%W, 30.0%N) constitute approximately equal parts of the diets of Northern Puffer. Worms (7.1% W, 9.1% N) make up nearly all of the remainder with fish tissue contributing less than 1% by both weight and number (Figure 31).

Northern Puffer Tables and Figures:

Table 16. Northern Puffer sampling rates and preserved specimen analysis status by year.

Figure 28. Station specific biomass of Northern Puffer in Chesapeake Bay, 2021.

Table 17. Northern Puffer geometric mean indices of abundance, by number and biomass, overall.

Figure 29. Northern Puffer geometric mean indices of abundance, by number and biomass, for all ages combined.

Figure 30. Northern Puffer length-frequency in Chesapeake Bay 2002-2021, overall (A) and by sex (B).

Figure 31. Diet composition, expressed as percent by weight (A) and percent by number (B) of Northern Puffer collected during ChesMMAP cruises in 2002-2020 combined.

Scup (Stenotomus chrysops)

<u>Abundance</u>: Total yearly captures of Scup are highly variable, probably as a result both of actual coast wide abundance and availability to the survey gear (Table 18). For example, in 2015 a time series (pre-2019) peak of almost 1,000 scup were captured by ChesMMAP, due mainly to a small number of large catches during the September cruise but in 2016 the total number caught was only 65. In 2017 only 25 Scup were present, which was the second-lowest number in the time series but catches rebounded in 2018 with 386 total specimens captured, the fifth highest total number during the survey years. Use of the '200' net during 2019 yielded over 1,100 scup, the most of any survey year, followed by 626 fish in 2020 and a new time series high number of 1,135 in 2021. Survey catches of Scup are typically rare during spring through early summer and nearly always reach a peak in September before declining again in November as fish leave bay waters (Figure 32). The species is most abundant in Regions 4 and 5 (C and D) and is rarely captured north of Region 4 (now Region C). It is important to note that 2007 data

are limited due to cancellation of the September cruise. Scup are typically most abundant in shallow strata (10'-30') and mid-depth strata (30'-50').

For purposes of calculating abundance indices, stations used include (Table 5):

- Through 2018: July, September, November Regions 4 and 5 Shallow and mid-depth strata.
- 2019 and forward: June, September, November, Regions C and D Both depth strata.
- Age-0 indices do not use data from the June/July cruises.

Discerning trends over the time series is problematic due to the difficulty in interpreting 2007 data when the September cruise was cancelled resulting from a budget shortfall. Geometric mean indices for both number and biomass indicate moderate abundance through 2007 then a sharp decline in 2008 followed by a two year upward trend toward a time series high in 2010. Following time-series low values between 2011 and 2014 there was a substantial increase in 2015 followed by another low value in 2016, a lower value still in 2017 but substantial increase in 2018. The apparent large increase in 2019 is likely to be largely due to the change in survey trawl gear but this analysis will have to wait for the completion of calibration studies (Table 19, Figure 33). As nearly all specimens captured by ChesMMAP are either age-0 or age-1, the age-2+ survey index should be interpreted with care.

<u>Length and Age</u>: Most specimens captured in the survey are less than 200mm fork length and at least two year-classes are apparent in length data (Figure 34). Due to the small size and sexual immaturity of the majority of Scup sampled by ChesMMAP, sex cannot be determined in the field for large numbers of specimens so sex-specific length frequencies do not display any discernible pattern of differences in sex ratios at size.

Nearly all specimens captured are either age-0 or age-1, so it is difficult to discern whether year-class abundance can be followed through time in age frequency figures (Figure 35, Figure 36). Both the length frequency and age distribution were similar in 2019 - 2021 to those in previous years.

<u>Diet:</u> By weight, worm species constitute the major portion (45.4%) of identifiable items in Scup stomachs but represent only 27.2% of prey by number (Figure 37). Unidentifiable prey (likely largely constituted of worms and other soft-bodied prey) also make up a large portion (19.2% W, 15.9% N). At 15.5% by weight, crustaceans (primarily mysids and amphipods) are also a major prey source, and at 30.8% represent the second largest taxon in Scup diets when measured by number.

Scup Tables and Figures:

Table 18. Scup sampling rates and preserved specimen analysis status by year.

Figure 32. Station specific biomass of Scup in Chesapeake Bay, 2021.

Table 19. Scup geometric mean indices of abundance, by number and biomass, overall and by age-class.

Figure 33. Scup geometric mean indices of abundance, by number and biomass, for all ages combined and by age-class.

Figure 34. Scup length-frequency in Chesapeake Bay 2002-2021, overall (A) and by sex (B).

Figure 35. Scup age-frequency by year, 2002-2021.

Figure 36. Scup age-frequency by year, 2002-2021 standardized to 8,000 (4,800 for 2019 and forward) annual trawl minutes.

Figure 37. Diet composition, expressed as percent by weight (A) and percent by number (B) of Scup collected during ChesMMAP cruises in 2002-2020 combined.

Spot (Leiostomus xanthurus)

<u>Abundance:</u> Spot are typically among the most abundant species in the survey during all cruises except March. Prior to 2014 between 2,000 and 11,500 Spot (115kg-1000kg) were captured annually during the 12 survey years. However, in 2014 only 939 specimens were sampled and in 2015 this figure fell further to only 401 individuals. While the total number captured rose to over 1,000 in 2016 this was still the third lowest total during the time series, with moderate increases to 1,586 fish in 2017 and 1,635 in 2018. Deployment of the new sampling gear in 2019 resulted in at least a 10 fold increase in numbers sampled and then a further doubling of in numbers in 2020 and still very high numbers in 2021 (Table 20). This species is well distributed throughout the bay, though concentrations are highest in regions 4 and 5 (Figure 38).

Overall abundances for the time series were on a generally rising trend between 2002 and 2006 then followed a downward trajectory until 2019, reaching time series lows between 2014 and 2018 (Table 28, Figure 39). This pattern does not follow the trend in coastal and regional landings (both commercial and recreational) which have been erratic but generally flat during the overlapping time series and this phenomenon deserves further attention. Age-specific indices are given for ages 0 through 2+ though since relatively few specimens older than age-1 are captured; the age-2+ index is of unknown reliability. These indices largely follow the same pattern as described for all ages combined except that the age-1 index reached its peak in 2007 rather than 2006 indicating that the large 2006 year-class was still abundant one year later. The considerable rise in abundance indices in 2019 through 2021 may be difficult to interpret once the calibration experiments and analyses are completed due to the very large increased capture rate of small/younger fish with the new sampling system.

<u>Length and Age</u>: Individuals between 100mm and 250mm are most common in the survey, with a smaller number of specimens up to 300mm occasionally captured (Figure 40). The largest individuals are most often captured in regions 2 or 3. No pattern of differential growth rates between the sexes is apparent.

Nearly all fish in the survey are either age-0 or age-1 with the oldest fish (5 total specimens) captured at age-4 (Figure 41, Figure 42). As discussed above, even though the age distribution of this species in Chesapeake Bay is not wide, the relative numbers of smaller vs. larger specimens can vary significantly year to year. This likely represents both changes in relative year-class strength and the numbers and sizes of specimens invading the bay each year. Much of the very large increase in catch of this species with the new trawling system appears to come in smaller, Age-0, specimens.

<u>Diet:</u> Not surprisingly, given the bottom-feeding habit of this species, the largest single prey type is 'unidentified material' (29.3% W, 24.7% N). In total 'miscellaneous' items (those which do not fit into one of the other major taxa) constitute 42.3% by weight and 40.0% by number of Spot diets. This is followed by worms (34.6% W, 28.4% N) which for the most part were not identifiable to specific taxa.

Crustaceans (10.2% W, 19.7% N), principally mysids and amphipods and molluscs (primarily clams) at 8.5% by weight and 8.1% by number, were also major portions of the diet for Spot (Figure 43).

Spot Tables and Figures:

Table 20. Spot sampling rates and preserved specimen analysis status by year.

Figure 38. Station specific biomass of Spot in Chesapeake Bay, 2021.

Table 21. Spot geometric mean indices of abundance, by number and biomass, overall and by age-class.

Figure 39. Spot geometric mean indices of abundance, by number and biomass, for all ages combined and by age-class.

Figure 40. Spot length-frequency in Chesapeake Bay 2002-2021, overall (A) and by sex (B).

Figure 41. Spot age-frequency by year, 2002-2021.

Figure 42. Spot age-frequency by year, 2002-2021 standardized to 8,000 (4,800 for 2019 and forward) annual trawl minutes.

Figure 43. Diet composition, expressed as percent by weight (A) and percent by number (B) of Spot collected during ChesMMAP cruises in 2002-2020 combined.

Striped Bass (Morone saxatilis)

<u>Abundance</u>: Striped Bass are typically captured in relatively high numbers each survey year with as many as 2,200 (weighing almost 1,000kg) sampled in 2005, 2,259 captured in 2019 and 2,201 in 2020 with the new trawl gear (Table 22). Intra-annual patterns of abundance for Striped Bass typically follow a consistent pattern. Large numbers of spawning migrants are captured during the March cruise, followed by lower numbers in May/June as the spawners leave the bay. Fewer captures occur in July and September, and higher numbers are encountered again in November as fish school before leaving the bay for offshore wintering grounds. Most Striped Bass are captured in regions 1 - 3 (A and B - Maryland waters) but the species occurs regularly in samples from all bay locations. In March, catches are high in all depth strata, but in other survey months catch rates are greatest in waters less than 50' (Figure 44).

Two sets of abundance indices have been calculated for this species: one using data from the March cruise which assesses abundance of the spring spawning stock, and one using data from November which characterizes the number of summer residents as they school together in the fall. Slightly different station sets are used for these two indices: in March all stations in Regions 1-3 (A and B) are included while in November all available stations (pre-2019: Regions 1 – 5 2019 and forward: Regions C and D) contribute to the index calculations (Table 5). Calculation of the November index may need to be reexamined as Regions A and B are no longer sampled in November.

March abundance for all ages combined, as measured both by number and biomass, was highest in 2004, 2008, 2013 and 2016, otherwise varying within a fairly narrow range in most years. After three low index values in 2009-2011 (no March cruise was conducted in 2012 due to vessel unavailability) a

significant rise was seen in 2013 due mainly to high values for age-3 and age-4 fish. This pattern generally held for age-specific abundance as well except that for age-1 and age-2 fish 2003 was also a year of high abundance. As most of the specimens captured in March are assumed to be reproductive migrants, it follows that in years of high overall abundance that all age classes would be present (Table 23, Figure 45). In 2018 no March cruise was conducted due to a funding shortfall. Note that the March/April 2019 cruise was conducted using the original sampling gear and the *R/V Bay Eagle* so the large spike in March abundance due to the new sampling system wasn't seen until 2020.

November abundance indices (summer residents) show high values in 2004 (more so in numbers than in biomass), 2006, 2014 and 2016. In 2011 through 2013 abundance turned upwards to mid-level values after a brief decline over the preceding two years then rose substantially in 2014, fell again in 2015 and rose again in 2016 and declined in both 2017 and 2018. Again, the same general pattern is seen in age-specific indices though variations do exist. The uptick in 2011-2014 appears to be due mainly to a larger number of age-2, age-3, and age-4+ specimens captured. The increases in 2016 indices appear to be mainly due to larger numbers of smaller/younger specimens (Table 24, Figure 46). Going forward, the November Striped Bass will be affected by the change in the annual sampling schedule, as no samples will be collected in the Regions A and B (Maryland) in November. This species and White Perch are the only species for which such an effect will occur.

New abundance indices which include data from all months and all regions, using a newly developed methodology, have been submitted to the ASMFC Striped Bass Technical Committee and ChesMMAP indices were included in the 2018 Benchmark Assessment.

<u>Length and Age:</u> Most specimens captured in the survey are about 600mm fork length or less (ages 1 – 7). The largest individuals approach 1000mm and are captured during spring spawning. Due to the relatively long-lived nature of this species, the varying life history scenarios for different portions of the stock and associated variable growth rates, along with variable young-of-year recruitment, it is difficult to differentiate year-classes within length-frequency histograms (Figure 47). However, age distribution figures (Figure 48, Figure 49) readily reveal year-class strength (high peaks during one year tend to follow into succeeding years, as do low abundances) which generally correspond to strong and weak year-classes as measured by the Maryland and Virginia young-of-year beach seine surveys. The largest fish captured tend to be migrating females and many 'resident' male fish are captured up to about 50cm. The oldest specimens yet sampled by the survey, age-20, were captured in 2008, 2010, and 2015 (1988, 1990 and 1995 year-classes, respectively). The most recent years may be exhibiting a contraction of the age distribution.

<u>Diet:</u> Results of diet analyses from this study differ appreciably from previous studies using specimens from Chesapeake Bay (Figure 50). Fishes comprise the largest taxonomic group in the diet (63.2% W, 57.1% N), with crustaceans the next most abundant (17.0% W vs. 26.1% N) due to consumption of a large number of small bodied mysid shrimp and amphipods. Among fish species, this survey consistently finds that Bay Anchovy contributes the highest proportion by weight (33.0%) with Atlantic menhaden second (15.9%). Mysid shrimp (7.3% W, 12.2%N) and amphipods (3.4% W, 6.2% N) combined constitute large portions of the diet, a sharp contrast to previous studies; and worms make up the only other major prey type (11.7% W, 9.9% N). These differences from previous diet studies are likely the result both of sampling methodological differences (the broad temporal and geographic scale of ChesMMAP as well as the trawl gear used compared to many studies which were limited in temporal or geographical scale or which use capture methodologies which yield a narrower size range) and analytical/mathematical differences in calculating percentages in the diet. In brief, this study calculates fish diets using cluster-

sampling theory and analytical methods whereas previous studies are thought to have used the assumption of simple random sampling of fish. The cluster method moderates the effect of a relatively small number of large predator specimens with large prey in the stomachs (e.g. Atlantic menhaden) as compared to a large number of smaller specimens with a significantly different diet.

Striped Bass Tables and Figures:

Table 22. Striped Bass sampling rates and preserved specimen analysis status by year.

Figure 44. Station specific biomass of Striped Bass in Chesapeake Bay, 2021.

Table 23. Striped Bass (March) geometric mean indices of abundance, by number and biomass, overall and by age-class.

Figure 45. Striped Bass (March) geometric mean indices of abundance, by number and biomass, for all ages combined and by age-class.

Table 24. Striped Bass (November) geometric mean indices of abundance, by number and biomass, overall and by age-class.

Figure 46. Striped Bass (November) geometric mean indices of abundance, by number and biomass, for all ages combined and by age-class.

Figure 47. Striped Bass length-frequency in Chesapeake Bay 2002-2021, overall (A) and by sex (B).

Figure 48. Striped Bass total age-frequency, 2002-2021.

Figure 49. Striped Bass age-frequency by year, 2002-2021 standardized to 8,000 (4,800 for 2019 and forward) annual trawl minutes.

Figure 50. Diet composition, expressed as percent by weight (A) and percent by number (B) of Striped Bass collected during ChesMMAP cruises in 2002-2020 combined.

Summer Flounder (Paralichthys dentatus)

<u>Abundance</u>: Though capture numbers (and biomass) have been lower in recent years, Summer Flounder are a primary target species for the survey with several hundred being sampled in most years (up to about 1,000 specimens weighing 450kg). While the numbers captured in 2019 through 2021 using the updated sampling gear was significantly higher than it had been in recent years it was within the range captured in other survey years (Table 25). The typical intra-annual pattern of abundance for summer flounder shows catches increasing monthly throughout the sample year, with highest catches in September and/or November (Figure 51). Summer flounder are most abundant in Regions 4 and 5 (C and D) but are common in Regions 2 and 3 (~ Region B) as well. A slightly higher catch rate is exhibited for mid-depth (30' - 50') and deep (>50') stations than in shallow (10' - 30') waters. The highest catches of summer flounder often occur along the eastern portions of Virginia waters but this is not an absolute.

For purposes of calculating abundance indices, stations used include (Table 5):

- Through 2018: September, November Regions 4 and 5 All depth strata.
- 2019 and forward: September, November, Regions C and D Both depth strata.

Abundance indices have varied considerably over the time but exhibit a consistent downward trend since 2006, reaching time series low values in 2012 through 2018 (Table 26, Figure 52). This is consistent with recent stock assessment updates. In 2019 the uncalibrated index rose substantially, though within the historical range and declined to time-series average values in 2020 and 2021.

Age-specific indices were calculated for ages 0 through 7+ (changed from 4+ previously in order to coincide with the current stock assessment strategy). Age-0 fish reached time series high values in 2006 and 2007 while most other year-classes were most abundant one or two years earlier. As these abundant young of year do not seem to result in higher abundance one or two years later perhaps specific individuals of this species do not reinvade the Chesapeake Bay each year. The increase in overall abundance in 2019 through 2021 appears to be mostly due to increase in the age-0 indices. It appears that the new sampling gear is more efficient at capturing smaller specimens when compared to the original ChesMMAP gear.

<u>Length and Age</u>: Fish which measure between approximately 20cm and 50cm total length are most prevalent in survey samples though fish as large as 760mm have been captured (Figure 53). In several years a large number of fish under 30cm (mostly age-0) can be differentiated in length-frequency graphs. This species is known to exhibit sexually dimorphic growth patterns (Dery 1981) and this is demonstrated in the sex-specific length plots. The vast majority of ChesMMAP specimens larger than 35cm and nearly all individuals larger than 40cm are females.

Most fish in the survey are age-5 and under, and the oldest fish yet captured are three specimens at age-12. In age classes older than age-2 it appears to be more difficult, compared to other species, to follow abundance trends of particular year-classes in successive years (Figure 54, Figure 55). This could be the result of differential migration patterns among different sized fish or of fishery preferences and/or regulations. As well as the declining abundance estimates described above, the Chesapeake portion of the Summer Flounder stock appear to have constricted somewhat in the age distribution in recent years. Since approximately 2007, as total captures have decreased and the age composition of the Chesapeake Summer Flounder has also compressed. In 2021 no fish older than age-1 were seen.

<u>Diet:</u> As measured by percent weight, fishes comprise a majority (55.0%) of summer flounder diets (Figure 56) in the survey, with the primary prey being Bay Anchovy (18.5%), Weakfish (7.9%), and Spot (7.0%), with crustaceans (41.2%) only slightly lower; as measured by number, crustaceans constitute about three-fifths of the diet (57.5%) with the main prey types being mysid shrimp (41.0%), sand shrimp (5.8%), and mantis shrimp (5.1%). The high prevalence of fish in summer flounder stomachs, especially for larger individuals, leads to the conclusion that this species should be considered a top predator in Chesapeake Bay along with Striped Bass, Bluefish, and Weakfish (Latour et al. 2008). It is noteworthy that by percent weight as measured by this survey, in Chesapeake Bay summer flounder are more highly piscivorous than are striped bass and are nearly on par with Weakfish in this characteristic.

Summer Flounder Tables and Figures:

Table 25. Summer Flounder sampling rates and preserved specimen analysis status by year.

Figure 51. Station specific biomass of Summer Flounder in Chesapeake Bay, 2021.

Table 26. Summer Flounder geometric mean indices of abundance, by number and biomass, overall and by age-class.

Figure 52. Summer Flounder geometric mean indices of abundance, by number and biomass, for all ages combined and by age-class.

Figure 53. Summer Flounder length-frequency in Chesapeake Bay 2002-2021, overall (A) and by sex (B).

Figure 54. Summer Flounder total age-frequency, 2002-2021.

Figure 55. Summer Flounder age-frequency by year, 2002-2021 standardized to 8,000 (4,800 for 2019 and forward) annual trawl minutes.

Figure 56. Diet composition, expressed as percent by weight (A) and percent by number (B) of Summer Flounder collected during ChesMMAP cruises in 2002-2020 combined.

Weakfish (Cynoscion regalis)

<u>Abundance</u>: Weakfish is among the most abundant species in survey samples and until 2010 most years resulted in 1,000 to 3,500 (75kg – 550kg) total captures. In the last years using the original ChesMMAP trawl gear numbers dropped and only 172 Weakfish were sampled in 2014 though this rose to 688 specimens in 2015 and again reached the 1,000 mark in 2016, declined to slightly less than 1,000 in 2017 and rose again to over 1,600 in 2018. In 2019 with the new sampling gear catches increased 5 to 10 times over previous levels to near 19,000 fish weighing over 1,300kg and rose still further in 2020 to 23,685 fish which totaled 1,300kg and dropped only moderately to 16,901 fish (1,044kg) in 2021 (Table 27). Catches are typically low in March but by May fish have begun to migrate into the bay and remain abundant in the survey throughout the rest of the year. Peak catches are usually in September and decline somewhat in November as fish begin their late fall migration out of the bay (Figure 57). Catches are typically higher in mid-depth (30' - 50') and deep (>50') stations than at shallow ones (10' - 30').

For purposes of calculating abundance indices, stations used include (Table 5):

- Through 2018: July, September, November Regions 4 and 5 All depth strata.
- 2019 and forward: June, September, November, Regions C and D Both depth strata.

Consistent with recent coast wide trends (ASMFC Weakfish Technical Committee, 2009), overall abundance for this species increased between 2002 and 2005 and then steadily declined over the next several years. However, after reaching a time series low in 2008 a slight upward tick was found in the successive two years but a sharp decline was seen again in 2011 through 2015 with a slight-to-moderate uptick in 2016, another decrease in 2017 and nearly flat values in 2018. Indices for 2019 increased remarkably (presumably due mainly to the change in trawl gear) rose again in 2020 before falling back to 2019 levels in 2021 (Table 28, Figure 58). As the vast majority of Weakfish sampled by ChesMMAP (and presumably present in the bay) in recent years have been either age-0 or age-1, the age specific abundances for these age classes tends to follow the same pattern as the overall indices.

<u>Length and Age</u>: Most Weakfish captured by the survey are between 100mm and 350mm total length. Minimum and maximum sizes found during the survey are 23mm and 616mm respectively. The length distributions in 2019 through 2021 were similar to that in other survey years though at much higher numbers (Figure 59). With only a few exceptions, most fish captured over 400mm were sampled during the first two years of the survey (2002 and 2003). Likewise, the age structure of Chesapeake Bay Weakfish has compressed over the past several years, with few individuals older than age-2 captured in recent years and almost none older than age-3 (Figure 60, Figure 61). In this survey, and others, each sampling year seems to result in (what appear to be) reasonable numbers of young fish but very few of these specimens are captured in successive years as older fish.

<u>Diet:</u> Fishes (68.3%), primarily Bay Anchovy (40.5%), comprise a majority of prey types in the Weakfish diet as measured by biomass ingested followed by crustaceans at 25.6%. Other taxa constitute only minor parts of the biomass diet (Figure 62). Notably, Weakfish account for 1.6% of prey in the diet of Weakfish, by weight. As measured by number, fishes comprise 53.3% of the diet with crustaceans following at 39.9%, dominated by mysid shrimp at 21.8%. Bay Anchovy are 31.3% of the diet by number. The relatively low percent of Atlantic menhaden seen in the survey stomach samples (less than the 1.0% threshold to appear in the current analysis), when compared to earlier studies, may be due to the truncation of the size range of Weakfish in Chesapeake Bay as well as the broad geographic and temporal scale of this survey and due to the cluster sampling analytical methodology as explained for Striped Bass above.

Weakfish Tables and Figures:

Table 27. Weakfish sampling rates and preserved specimen analysis status by year.

Figure 57. Station specific biomass of Weakfish in Chesapeake Bay, 2021.

Table 28. Weakfish geometric mean indices of abundance, by number and biomass, overall and by ageclass.

Figure 58. Weakfish geometric mean indices of abundance, by number and biomass, for all ages combined and by age-class.

Figure 59. Weakfish length-frequency in Chesapeake Bay 2002-2021, overall (A) and by sex (B).

Figure 60. Weakfish total age-frequency, 2002-2021.

Figure 61. Weakfish age-frequency by year, 2002-2021 standardized to 8,000 (4,800 for 2019 and forward) annual trawl minutes.

Figure 62. Diet composition, expressed as percent by weight (A) and percent by number (B) of Weakfish collected during ChesMMAP cruises in 2002-2020 combined.

White Perch (Morone americana)

<u>Abundance</u>: White Perch are extremely abundant in survey samples throughout each year in Regions 1 and 2 (Region A and upper Region B) and are common into Region 3 (Region B and lower Region A - Table 29, Figure 63). Due to this species' concentration in the shallow waters of Region 1, catches are highest in the shallowest strata (10' - 30'), followed by the mid-depth strata (30' - 50'), with this species rarely seen in samples from the deepest stations (>50'). Interpretation of abundance indices for this

species must account for the fact that ChesMMAP samples only a portion of the range of the species and catches can be significantly influenced by salinity.

As with Striped Bass, indices of abundance are presented for both the spring (March) spawning population and for the fall (November) when fish again school together. For both the March and November indices, data only from the shallow and mid-depth stations in Regions 1 and 2 (~Region A) are included. Interestingly, these two sets of indices show nearly opposing trends in abundance. The March indices (Table 30, Figure 64), measured either by number or biomass, show relatively flat abundance in all years except for peak values (about 4-5 times higher than other values) in 2007 and 2008, with a significant uptick in 2013 and downward points in 2014 and 2015 then up to record high values in 2016. Meanwhile, the November indices (Table 31, Figure 65) fluctuate without trend through 2006, and then reach time series lows in 2007 and 2008, followed by a steady upward trend with a distinct decline in 2012 and 2013 then upticks in 2014 and 2015 with slight declines in 2016. If it is assumed that the peaks in March abundance in 2007 and 2008 reflected a high abundance of spawners then it could well make sense that the stock increased for several of the succeeding years. Due to the cessation of November sampling in new Regions A and B the November White Perch indices will terminated or perhaps replaced in future reports.

<u>Length and Age</u>: All White Perch of sizes greater than approximately 150mm fork length are well sampled in the survey (Figure 66). Due to the relatively small maximum size, long life, and slow growth rates it is difficult to separate year-classes of this species using length-frequency. The peak of abundance in 2007 and 2008 samples was at a smaller size than during previous years. It appears that more females are sampled by ChesMMAP than are males and that females reach a slightly larger maximum size than do males.

This species is not well sampled by the survey until approximately age-4 (Figure 67, Figure 68); however past that age the survey appears to adequately represent all age classes. Specimens as old as 19 years have been captured. The species age distribution appears to be regulated by the relative success of each year-class. Year-class specific peaks in abundance can be easily followed during successive years in survey samples (e.g., 1993, 1996, 2000, 2003, 2011 year-classes).

<u>Diet:</u> Amphipods represents the largest single prey category by both weight and number (21.1% W, 32.5% N) in White Perch stomachs among identifiable prey, and crustaceans (33.7% W, 48.4% N) constitute the largest identifiable taxon, followed by a number of other small crustacean prey. Worms (26.1% W, 19.1% N), primarily *Nereis* clam worms (9.6% W, 7.2% N) and other polychaetes (14.8% W, 10.1% N), are the second most abundant prey, followed by a variety of mollusc species, (14.5% W, 13.0% N). Notably, a small number of Bay Anchovy (2.2% W, 1.7% N) are present in White Perch stomachs (Figure 69).

White Perch Tables and Figures:

Table 29. White Perch sampling rates and preserved specimen analysis status by year.

Figure 63. Station specific biomass of White Perch in Chesapeake Bay, 2021.

Table 30. White Perch geometric mean indices of abundance for March by number and biomass, overall and by age class.

Figure 64. White Perch geometric mean indices of abundance, by number and biomass, for all ages combined and by age-class for March.

Table 31. White Perch geometric mean indices of abundance for November by number and biomass, overall and by age class.

Figure 65. White Perch geometric mean indices of abundance, by number and biomass, for all ages combined and by age-class for November.

Figure 66. White Perch length-frequency in Chesapeake Bay 2002-2021, overall (A) and by sex (B).

Figure 67. White Perch total age-frequency, 2002-2021.

Figure 68. White Perch age-frequency by year, 2002-2021 standardized to 8,000 (4,800 for 2019 and forward) annual trawl minutes.

Figure 69. Diet composition, expressed as percent by weight (A) and percent by number (B) of White Perch collected during ChesMMAP cruises in 2002-2020 combined.

Task 5 – Evaluation of alternative sampling gear

As noted in several previous proposals and annual reports, ChesMMAP has implemented a change to its sampling gear to a bottom trawl that is the same design and half of the size of that used by the NEAMAP and NEFSC surveys. This net is often called the '200' net which refers to the number of 12cm meshes around the mouth (200 x 12cm). This style of net achieves a greater headline height, more consistent geometry, and better bottom contact relative to typical survey trawls. The original ChesMMAP net typically fishes at about 8m of wingspread and 1m of headrope height. In flume tank tests and in the field the '200' net has a similar wingspread but more than double the headrope height. In short, it is expected to provide a much better sample of the demersal and semi-pelagic biota.

After a six-year design and construction process, VIMS took delivery of its new research vessel (*R/V Virginia*) in October 2018. This vessel is specifically outfitted to deploy, fish and retrieve the '200' net and to allow efficient processing of the catch. The *R/V Bay Eagle* will remain in the VIMS fleet until at least 2022 which should allow adequate time for calibrations. Following delivery, MRG made several important decisions regarding sampling aboard the new vessel:

New Sampling

The *R/V Virginia* is considerably larger than the *R/V Bay Eagle* (93'LOA v. 65'LOA) and more expensive to operate and it will require both a larger vessel crew and scientific crew, both of which increase daily costs. Further, we have been notified by VMRC that in coming years we should expect fewer available Wallop-Breaux funds. In combination, these facts necessitate a decrease in the number of days-at-sea available for ChesMMAP. Initially MRG planned to decrease the number of annual trips from five bimonthly cruises (March, May, July, September, November) to three seasonal ones (Spring, Summer, Fall). However, further thought and analysis resulted in a choice to conduct two full bay-wide cruises (June and September) and two 'half-cruises' in March and November. While this strategy departs somewhat from our goal of synoptic sampling across the ecosystem, prior ChesMMAP data reveal that few fish stocks of interest are utilizing the lower (Virginia) bay in March or the upper (Maryland) bay in November.

Just two sets of abundance indices (November Striped Bass and White Perch) will be impacted by this choice.

- Full trips will continue sampling at up to 80 sites per trip. The goal of the March and November trips will be 35 and 45 stations respectively.
- Rather than wait 2 or more years to complete calibration studies and then commence official sampling with the new vessel/gear combination we chose to implement the new sampling platform as soon as it was fully available (June 2019). We are convinced that the new system is far superior to the previous one in many respects and it is worth a temporary break in availability of calibrated survey data to begin collecting better data immediately.
- We made a choice to *not* conduct side-by-side calibration tows during regular survey operations with the new vessel. Rather, calibration tows will be conducted in between survey dates. This assures that survey data are not affected by the presence of another fishing vessel in the immediate vicinity.
- Once calibrations are complete, we will calibrate the older data series to the newer, rather than the converse.
- In the future we plan to implement collection, at a random subset of sites, an increased amount of synoptic data. Along with the newer bottom trawling gear we will also deploy a mid-water trawl to sample pelagic species which are currently under-represented in our data. As a part of a graduate student/proof of concept project several years ago we established that fishing such a net from the older vessel was possible and we gathered considerable data on how to properly design and fish such a trawl.
- At this same subset of stations, we plan to take bottom 'grabs' to sample the benthos as well as a plankton net with which we especially hope to sample mysid shrimp, which previous ChesMMAP food habits data have shown to be very important, but unmonitored, prey items for many other economically and ecologically valuable fish species. These so called 'Ponar' grabs were commenced in 2020 at a subset of stations. Results will be summarized in future reports.
- Over time, this synoptic data collection system will provide for development of much improved multi-species and ecosystem models of Chesapeake Bay. Both of these last two increases in sampling will be dependent on additional funding, either from VMRC or other sources.

Calibration:

While vessel availability due to scheduling or breakdowns of either the *Virginia* or the *Bay Eagle* and/or availability of MRG personnel due to other commitments somewhat constrained the availability of sampling days in CY 2019, three calibration trips were made in June, July, and December 2019. A total of 61 paired tows were completed in Virginia waters. In 2020 four calibration trips in March (Maryland), June, September and November were completed with a total of 135 paired tows. Four trips (March, June, August, November) were also made in 2021 with a total of 171 paired tows. So far in 2022 an additional 35 calibration pairs have been completed. This brings the 3+ year total to 381. Absent significant interruptions (e.g. vessel breakdowns) we expect calibration trips to be completed in 2022 or early 2023.

These trips were made soon after regularly scheduled ChesMMAP cruises were completed and concentrated effort in 'index' strata for target (i.e. managed) species, in locations where high concentrations of these fishes had been found during survey operations.

Side-by-side tows were completed as simultaneously as possible with the two vessels within approximately one quarter mile of one another (Figure 70). All deployment, retrieval and catch

processing procedures were identical to those employed during regular surveys except that all specimens were processed using the 'deck' protocol wherein each species/modal size group sample was weighed in whole and then individual fish were measured. No 'full workup' biological sampling occurred though in the December trip when large numbers of Striped Bass were encountered a small number of them were tagged as part of the Interstate Tagging Program.

A complete description of the calibration methodology will be reported upon completion of the calibration studies.

Preliminary Results:

For descriptive purposes only, a set of preliminary results are presented. For any given species the number of tows in which both vessels encountered that species is limited. Further, the analysis methodology shown here is for a 'first look' only and does not represent the methods which will be used to develop final calibration coefficients. Though simple regression parameters are included in these results they must not be used to adjust any data presented elsewhere in this report.

For each species two types of figures are shown:

A linear regression of tow-specific Virginia catch on Bay Eagle catch, in numbers. When final calibration coefficients are calculated we intend for Bay Eagle data to be adjusted to Virginia data but for these preliminary analyses the Virginia data are shown as the independent variable. Raw catch numbers for each vessel are adjusted to a 'standard' area swept value for that vessel. For species for which visual examination of the data revealed obvious modal size cohorts, analyses were performed separately for each cohort.

Each plot contains several elements:

- Black dots represent tows in which both vessels encountered the species.
- Blue dots signify tows in which the *Virginia* captured the species but the *Bay Eagle* did not.
- Red dots are tows in which the species was seen in the *Bay Eagle* catch but not in the *Virginia* catch.
- The black line is the one-to-one line where points would appear if catch rates were equal.
- The green 'Non-Zero Reg' line is the linear regression including only those tows in the species was seen by both vessels. The regression parameters are shown in the inset box.
- Some figures include points which are circled in yellow. These points represent tows in which one vessel or the other encountered an extreme number of fish and these points may skew the regression line. In these cases, a second yellow regression line ('Edited Reg') and related parameters are shown.
- An overlaid comparison of length frequencies for each species, for all tows combined, with no editing or restricting of data points. Consistent with the color coding above, *Virginia* data are in blue and *Bay Eagle* data are in reddish hues.

As expected, for most species catches on the R/V Virginia using the new '200' net are substantially larger and the length frequencies are much broader than those using the original sampling system. Many of the additional fish are on the smaller end of the length scale.

Calibration Figures:

Figure 70. The R/V Bay Eagle and the R/V Virginia preparing to conduct a calibration tow (A) and the R/V Virginia travelling towards the next site upon completion of a tow (B).

Figure 71. Preliminary comparison of catch rates during calibration tows for Atlantic Croaker, 0 – 125mm.

Figure 72. Preliminary comparison of catch rates during calibration tows for Atlantic Croaker, 126=mm.

Figure 73. Comparison of length frequency distributions during calibration tows for Atlantic Croaker 0 – 125mm.

Figure 74. Comparison of length frequency distributions during calibration tows for Atlantic Croaker 126+mm.

Figure 75. Preliminary comparison of catch rates during calibration tows for Black Sea Bass, all sizes.

Figure 76. Comparison of length frequency distributions during calibration tows for Black Sea Bass.

Figure 77. Preliminary comparison of catch rates during calibration tows for Butterfish, 0 – 60mm.

Figure 78. Preliminary comparison of catch rates during calibration tows for Butterfish, 61+mm.

Figure 79. Comparison of length frequency distributions during calibration tows for Butterfish, 0 – 60mm (A) and 61+mm (B).

Figure 80. Preliminary comparison of catch rates during calibration tows for Kingfish, 0 – 200mm.

Figure 81. Preliminary comparison of catch rates during calibration tows for Kingfish, 201+mm.

Figure 82. Comparison of length frequency distributions during calibration tows for Kingfish, 0 – 200mm (A) and 200+mm (B).

Figure 83. Preliminary comparison of catch rates during calibration tows for Scup, all sizes.

Figure 84. Comparison of length frequency distributions during calibration tows for Scup.

Figure 85. Preliminary comparison of catch rates during calibration tows for Spot, 0 – 140mm.

Figure 86. Preliminary comparison of catch rates during calibration tows for Spot, 141+mm.

Figure 87. Comparison of length frequency distributions during calibration tows for Spot, 0 – 140mm (A) and 140+mm (B).

Figure 88. Preliminary comparison of catch rates during calibration tows for Striped Bass, 0- 200mm.

Figure 89. Preliminary comparison of catch rates during calibration tows for Striped Bass, 201-300mm.

Figure 90. Preliminary comparison of catch rates during calibration tows for Striped Bass, 301+mm.

Figure 91. Comparison of length frequency distributions during calibration tows for Striped Bass 0-200mm (A), 201-30mm (B) and 300+mm (C).

Figure 92. Preliminary comparison of catch rates during calibration tows for Summer Flounder, all sizes.

Figure 93. Comparison of length frequency distributions during calibration tows for Summer Flounder.

Figure 94. Preliminary comparison of catch rates during calibration tows for Weakfish, all sizes.

Figure 95. Comparison of length frequency distributions during calibration tows for Weakfish.

Figure 96. Preliminary comparison of catch rates during calibration tows for White Perch, 0 – 190mm.

Figure 97. Preliminary comparison of catch rates during calibration tows for White Perch, 0 – 190mm.

Figure 98. Comparison of length frequency distributions during calibration tows for White Perch, 0 – 190mm (A) and 191+mm (B).

Water Quality

Bottom temperature (Figure 99), salinity (Figure 100), dissolved oxygen (Figure 101) readings, interpolated among sample locations for each trip, reveal varying physical conditions both within and among years. For each parameter three figures are presented; A) two-dimensional interpolation of 2021 readings, B) similar interpolation including data from all sample years, and C) the difference between the two. In March the upper portion of the bay was somewhat cooler, fresher and with slightly higher oxygen readings than in previous years. In June, bottom temperatures were still generally lower than average, salinities were generally near average, as was dissolved oxygen except in larger portions of Maryland waters where summer hypoxic readings were beginning to show. Similar conditions held in September except that the hypoxic zones were even larger. Virginia recordings in November were close to average with no exceptional readings.

Using readings of the water profiles at each sampling station, three dimensional interpolations of the same parameters for each Region were also calculated (Figure 102, Figure 103, Figure 104). While these summarizations collapse three-dimensional data from fairly large portions of the bay into simpler two-dimensional figures, they do provide a measure of understanding of the complex and varying seasonal and geographical conditions in Chesapeake Bay.

Figure 99. Interpolated Chesapeake Bay bottom water temperature by cruise for 2021 (A), averaged over 2002 through 2021 (B), and 2021 deviation from average (C).

Figure 100. Interpolated Chesapeake Bay bottom salinity by cruise for 2021 (A), averaged over 2002 through 2021 (B), and 2021 deviation from average (C).

Figure 101. Interpolated Chesapeake Bay bottom dissolved oxygen by cruise for 2021 (A), averaged over 2002 through 2021 (B), and 2021 deviation from average (C).

Figure 102. Interpolated bi-monthly water temperature profiles in Chesapeake Bay, 2021.

Figure 103. Interpolated bi-monthly salinity profiles in Chesapeake Bay, 2021.

Figure 104. Interpolated bi-monthly dissolved oxygen profiles in Chesapeake Bay, 2021.

<u>Appendix</u>

Abundance data summaries for a selection of common species which are not considered as recreational species for funding and management purposes are provided in the Appendix. The species are blue crab – males and mature females separately, and clearnose skate.

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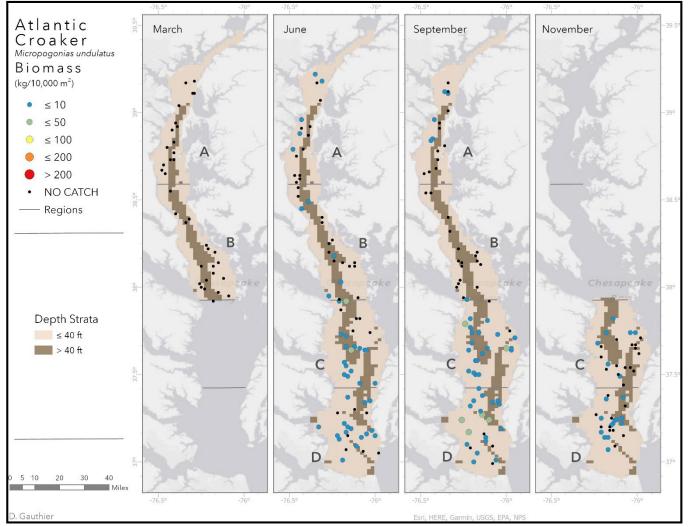
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| Year | Number Caught | Biomass Caught (kg) | Presence at Index Stations (%) | Number Measured | Age Specimens | Ages Read | Stomach Specimens | Stomachs Analyzed |
|------|------------------|------------------------|--------------------------------------|--------------------|------------------|--------------|----------------------|----------------------|
| 2002 | 12,689 | 2,834.0 | 57.6 | 7,082 | 1,126 | 1,126 | 1,104 | 93 |
| 2003 | 12,217 | 2,850.3 | 68.2 | 5,721 | 548 | 548 | 542 | 62 |
| 2004 | 20,394 | 5,330.5 | 74.8 | 8,850 | 717 | 717 | 702 | 254 |
| 2005 | 13,281 | 3,184.8 | 74.8 | 7,757 | 716 | 716 | 704 | 261 |
| 2006 | 14,878 | 3,486.6 | 79.0 | 8,904 | 854 | 854 | 834 | 750 |
| 2007 | 12,678 | 1,963.6 | 68.5 | 5,974 | 526 | 526 | 523 | 506 |
| 2008 | 6,260 | 1,031.3 | 53.3 | 3,070 | 480 | 480 | 460 | 454 |
| 2009 | 3,797 | 523.0 | 72.2 | 3,250 | 369 | 369 | 361 | 358 |
| 2010 | 3,243 | 454.3 | 55.2 | 2,355 | 322 | 322 | 317 | 310 |
| 2011 | 5,187 | 605.5 | 57.8 | 2,776 | 322 | 322 | 291 | 287 |
| 2012 | 2,448 | 152.9 | 42.2 | 1,998 | 312 | 312 | 280 | 269 |
| 2013 | 8,971 | 655.1 | 45.9 | 3,684 | 282 | 282 | 237 | 231 |
| 2014 | 1,449 | 143.3 | 24.4 | 620 | 111 | 111 | 73 | 71 |
| 2015 | 1,723 | 167.4 | 36.3 | 1,402 | 160 | 160 | 110 | 107 |
| 2016 | 919 | 90.6 | 27.4 | 551 | 113 | 113 | 69 | 69 |
| 2017 | 1,318 | 92.9 | 31.1 | 1,037 | 247 | 247 | 190 | 188 |
| 2018 | 1,164 | 51.6 | 20.7 | 455 | 88 | 88 | 56 | 56 |
| 2019 | 11,685 | 919.7 | 84.4 | 5,792 | 354 | 354 | 233 | 230 |
| 2020 | 34,291 | 1,816.6 | 84.4 | 6,970 | 303 | 303 | 194 | 191 |
| 2021 | 8,832 | 552.9 | 81.1 | 3,849 | 316 | 316 | 205 | 205 |

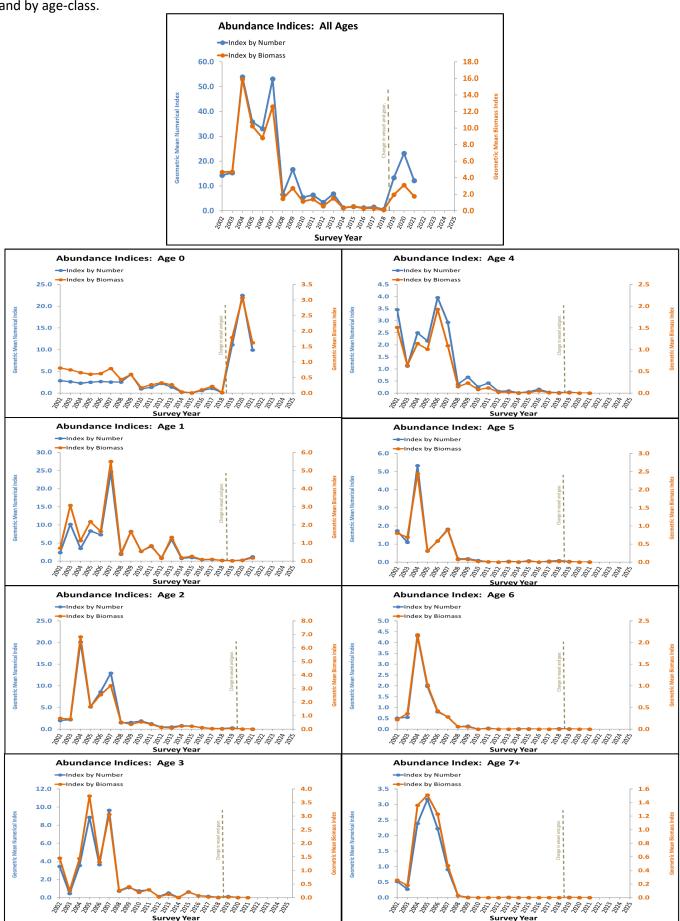
Table 6. Atlantic Croaker sampling rates and preserved specimen analysis status by year (shaded rows represent a change in vessel and trawl gear beginning in June 2019).

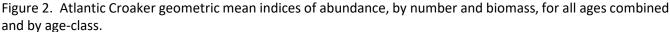
Figure 1. Station specific biomass of Atlantic Croaker in Chesapeake Bay, 2021.



| Table 7. | Atlantic Croaker | geometric mean ir | ndices of a | ibundance, l | by number a | and biomass, | overall and by age-class. |
|----------|------------------|-------------------|-------------|--------------|-------------|--------------|---------------------------|
| | | | | | | | |

| | Age | n | | erical Ind | | | nass Ind | | Year A | Age n | | rical Ind | | | nass Ind | |
|----------------------|-----|----------|--------------|---------------|--------------|-------------|--------------|--------------|--------------|------------|------------|--------------|-------------------|------------|--------------|------------|
| | AII | 70 | 7.9 | Index 14.3 | UCI 25.1 | 2.8 | Index 4.7 | UCI 7.4 | 2002 | 4 70 | 2.0 | Index 3.5 | <u>UCI</u> 5.8 | 0.9 | Index 1.5 | UCI 2.4 |
| 2003 2004 | | 48 77 | 7.5 33.8 | 15.3 53.9 | 30.3 85.7 | 2.5 10.3 | 4.7 15.9 | 8.5 24.1 | 2003 2004 | 48 | 0.5 | 1.1 2.5 | 2.0 3.6 | 0.3 | 0.6 | 1.2 1.6 |
| 2005 2006 | | 77 74 | 22.6 20.5 | 35.9 33.0 | 56.6 52.8 | 6.8 5.8 | 10.2 8.8 | 15.2 13.2 | 2005 2006 | 77 | 1.4 2.4 | 2.2 4.0 | 3.1 6.3 | 0.7 | 1.0 1.9 | 1.4 2.9 |
| 2007 | | 52 | 26.5 | 53.1 | 105.3 | 7.1 | 12.6 | 21.7 | 2007 | 52 | 1.8 | 2.9 | 4.5 | 0.7 | 1.1 | 1.6 |
| 2008 2009 | | 76 52 | 3.9 9.6 | 6.6 16.6 | 10.8 28.2 | 0.9 | 1.5 | 2.2 | 2008 2009 | 76 52 | 0.2 | 0.4 | 0.6 | 0.1 | 0.2 | 0.3 0.4 |
| 2010 | | 78 | 3.3 | 5.4 | 8.5 | 0.7 | 1.1 | 1.7 | 2010 | 78 | 0.1 | 0.3 | 0.4 | 0.0 | 0.1 | 0.1 |
| 2011 2012 | | 78 78 | 3.8 1.9 | 6.4 3.4 | 10.4 5.7 | 0.9 | 1.4 0.6 | 2.1 0.9 | 2011 2012 | 78 78 | 0.2 | 0.4 | 0.7 | 0.1 | 0.1 | 0.2 |
| 2013 | | 78 78 | 3.6 | 6.8 | 12.3 | 0.9 | 1.5 | 2.3 | 2013 | 78 78 | 0.0 | 0.1 | 0.2 | 0.0 | 0.0 | 0.0 |
| 2014 2015 | | 78 | 0.6 | 1.3 | 2.2 2.9 | 0.1 | 0.4 | 0.6 0.8 | 2014 2015 | 78 | 0.0 | 0.1 | 0.1 | 0.0 | 0.0 | 0.0 0.0 |
| 2016 2017 | | 78 78 | 0.6 0.8 | 1.2 | 2.1 2.5 | 0.1 | 0.3 | 0.5 | 2016 2017 | 78 78 | 0.0 | 0.2 | 0.3 | 0.0 | 0.1 | 0.1 |
| 2018 | | 78 | 0.3 | 0.5 | 0.8 | 0.0 | 0.1 | 0.2 | 2018 | 78 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 |
| 2019 2020 | | 90 90 | 8.4 12.7 | 13.3 23.1 | 21.0 41.3 | 1.3 1.9 | 1.9 3.1 | 2.8 4.9 | 2019 2020 | 90 90 | 0.0 0.0 | 0.0 | 0.1 | 0.0 0.0 | 0.0 0.0 | 0.0 0.0 |
| 2021 | | 90 | 7.7 | 12.2 | 18.9 | 1.2 | 1.8 | 2.5 | 2021 | 90 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 2022 2023 | | - | | | | | | | 2022 2023 | | | | | | | |
| 2024 2025 | | | | | | | | | 2024 2025 | | | | | | | |
| 2002 | | 70 | 1.5 | 2.8 | 4.9 | 0.4 | 0.8 | 1.3 | 2002 | 5 70 | 1.0 | 1.7 | 2.8 | 0.5 | 0.8 | 1.2 |
| 2003 2004 | | 48 77 | 1.2 1.3 | 2.6 2.3 | 4.9 3.6 | 0.3 | 0.8 | 1.3 1.0 | 2003 2004 | 48 | 0.4 | 1.1 5.3 | 2.0 8.0 | 0.3 | 0.7 | 1.3 3.5 |
| 2005 | | 77 | 1.4 | 2.5 | 4.1 | 0.4 | 0.6 | 0.9 | 2005 | 77 | 0.4 | 0.6 | 0.9 | 0.2 | 0.3 | 0.5 |
| 2006 2007 | | 74 52 | 1.6 | 2.7 | 4.2 4.9 | 0.4 | 0.6 0.8 | 0.9 | 2006 2007 | 74 52 | 0.7 | 1.2 | 1.7 2.9 | 0.4 | 0.6 | 0.8 |
| 2008 | | 76 | 1.5 | 2.5 | 4.1 | 0.3 | 0.4 | 0.6 | 2008 | 76 | 0.1 | 0.2 | 0.3 | 0.0 | 0.1 | 0.1 |
| 2009 2010 | | 52 78 | 2.7 0.6 | 4.3 | 6.7 1.5 | 0.4 | 0.6 | 0.8 | 2009 2010 | 52 78 | 0.1 | 0.2 | 0.3 | 0.0 | 0.1 | 0.1 |
| 2011 | | 78 | 0.8 | 1.4 | 2.2 | 0.1 | 0.3 | 0.4 | 2011 | 78 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 2012 | | 78 78 | 1.2 0.8 | 2.2 | 3.7 | 0.2 | 0.3 | 0.5 | 2012 2013 | 78 78 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 0.0 |
| 2014 | | 78 | 0.1 | 0.2 | 0.4 | 0.0 | 0.0 | 0.1 | 2014 | 78 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 2015 2016 | | 78 78 | 0.0 | 0.0 | 0.1 1.0 | 0.0 | 0.0 | 0.0 | 2015 2016 | 78 78 | 0.0 | 0.1 | 0.2 | 0.0 | 0.0 | 0.1 0.0 |
| 2017 2018 | | 78 78 | 0.6 | 1.1 | 1.8 0.2 | 0.1 | 0.2 | 0.4 | 2017 2018 | 78 78 | 0.0 | 0.1 | 0.1 | 0.0 | 0.0 | 0.0 |
| 2019 | | 90 | 6.7 | 11.1 | 18.1 | 1.1 | 1.8 | 2.6 | 2018 | 90 | 0.0 | 0.0 | 0.2 | 0.0 | 0.0 | 0.0 |
| 2020 | | 90 90 | 12.3 6.0 | 22.5 9.9 | 40.4 15.9 | 1.9 1.1 | 3.1 | 4.8 2.3 | 2020 2021 | 90 | 0.0 0.0 | 0.0 | 0.0 0.0 | 0.0 0.0 | 0.0 | 0.0 0.0 |
| 2021 | | 90 | 6.0 | 9.9 | 15.9 | ±.± | 1.6 | 2.3 | 2021 | 90 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 2023 2024 | | ~ | | | | | | | 2023 2024 | | | | | | | |
| 2025 | | ~ | | | | | | | | | | | | | | |
| 2002 | | 70 48 | 1.5 5.1 | 2.4 10.2 | 3.6 19.5 | 0.5 | 0.7 | 1.0 5.3 | 2002 | 6 70 48 | 0.3 | 0.5 | 0.8 | 0.1 | 0.2 | 0.3 |
| 2004 | | 77 | 2.1 | 3.5 | 5.5 | 0.7 | 1.1 | 1.7 | 2004 | 77 | 2.8 | 4.3 | 6.5 | 1.4 | 2.2 | 3.2 |
| 2005 2006 | | 77 74 | 5.0 4.8 | 8.3 7.3 | 13.4 11.0 | 1.4 | 2.2 | 3.2 | 2005 2006 | 77 | 1.3 0.5 | 2.0 | 2.9 1.2 | 0.7 | 1.0 0.4 | 1.5 0.6 |
| 2007 | | 52 | 12.2 | 24.6 | 48.7 | 3.1 | 5.5 | 9.2 | 2007 | 52 | 0.3 | 0.6 | 0.9 | 0.1 | 0.3 | 0.5 |
| 2008 2009 | | 76 52 | 1.1 4.5 | 1.9 8.2 | 2.9 14.3 | 0.2 | 0.4 | 0.7 | 2008 2009 | 76 52 | 0.0 | 0.1 | 0.2 | 0.0 | 0.1 | 0.1 |
| 2010 | | 78 | 1.6 | 2.7 | 4.2 | 0.3 | 0.5 | 0.8 | 2010 | 78 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 2011 2012 | | 78 78 | 2.5 0.5 | 4.0 0.9 | 6.3 1.5 | 0.5 | 0.9 | 1.2 0.3 | 2011 2012 | 78 78 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 0.0 |
| 2013 2014 | | 78 78 | 3.2 0.3 | 6.0 0.7 | 10.6 1.3 | 0.8 | 1.3 0.2 | 2.1 0.3 | 2013 2014 | 78 78 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 0.0 |
| 2014 | | 78 | 0.4 | 1.0 | 1.3 | 0.1 | 0.2 | 0.5 | 2014 | 78 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 2016 | | 78 78 | 0.1 | 0.3 | 0.7 | 0.0 | 0.1 | 0.2 | 2016 2017 | 78 78 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 2018 | | 78 | 0.1 | 0.2 | 0.3 | 0.0 | 0.0 | 0.1 | 2018 | 78 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 |
| 2019 2020 | | 90 90 | 0.1 0.1 | 0.1 0.2 | 0.2 0.4 | 0.0 0.0 | 0.0 0.0 | 0.0 | 2019 2020 | 90 90 | 0.0 0.0 | 0.0 0.0 | 0.0 0.0 | 0.0 0.0 | 0.0 0.0 | 0.0 0.0 |
| 2021 | | 90 | 0.9 | 1.2 | 1.6 | 0.1 | 0.2 | 0.2 | 2021 | 90 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 2022 | | ~ | | | | | | | 2022 | | | | | | | |
| 2024 | | | | | | | | | 2024 | | - | | | | | |
| 2025 | 2 | 70 | 1.2 | 2.0 | 3.1 | 0.5 | 0.8 | 1.2 | 2002 | 7+ 70 | 0.3 | 0.5 | 0.8 | 0.1 | 0.3 | 0.4 |
| 2003 | | 48 | 1.2 | 2.2 | 3.6 | 0.4 | 0.7 | 1.2 | 2003 | 48 | 0.1 | 0.3 | 0.5 | 0.0 | 0.2 | 0.3 |
| 2004 2005 | | 77 77 | 11.7 3.5 | 20.1 5.1 | 34.1 7.5 | 4.2 | 6.8 1.7 | 10.7 2.3 | 2004 2005 | 77 | 1.5 2.1 | 2.4 | 3.6 4.7 | 0.9 | 1.4 1.5 | 2.0 2.2 |
| 2006 | | 74 | 5.4 | 8.6 | 13.4 | 1.7 | 2.5 | 3.7 | 2006 | 74 | 1.3 | 2.2 | 3.5 | 0.7 | 1.2 | 1.9 |
| 2007 | | 52 76 | 6.9 0.8 | 13.0 1.5 | 23.6 2.5 | 1.9 0.3 | 3.2 | 5.0 0.8 | 2007 | 52 76 | 0.5 | 0.9 | 1.5 0.1 | 0.2 | 0.5 | 0.8 0.1 |
| 2009 | | 52 | 0.9 | 1.6 | 2.5 | 0.2 | 0.4 | 0.6 | 2009 | 52 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 2010 2011 | | 78 78 | 1.1 0.7 | 1.9 1.3 | 3.1 | 0.3 | 0.5 | 0.8 | 2010 2011 | 78 78 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 0.0 |
| 2012 | | 78 | 0.2 | 0.5 | 0.8 | 0.0 | 0.1 | 0.2 | 2012 | 78 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 2013 2014 | | 78 78 | 0.3 | 0.5 | 0.8 1.4 | 0.0 | 0.1 | 0.1 | 2013 2014 | 78 78 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 0.0 |
| 2015 2016 | | 78 78 | 0.3 | 0.7 | 1.2 0.7 | 0.1 | 0.2 | 0.4 | 2015 2016 | 78 78 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 2017 | | 78 | 0.0 | 0.2 | 0.3 | 0.0 | 0.1 | 0.1 | 2017 | 78 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 2018 2019 | | 78 90 | 0.1 | 0.1 | 0.2 0.4 | 0.0 | 0.0 | 0.0 | 2018 2019 | 78 90 | 0.0 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 0.0 |
| 2020 | | 90 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 2020 | 90 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 2021 2022 | | 90 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 2021 2022 | 90 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 2023 | | | | | | | | | 2023 | | | | | | | |
| 2024 2025 | | | | | | | | | 2024 | | | | | | | |
| 2002 | | 70 | 2.0 | 3.4 | 5.7 | 0.9 | 1.5 | 2.3 | _ | | | | | | | |
| 2003 2004 | | 48 77 | 0.2 | 0.4 | 0.8 5.3 | 0.1 | 0.3 | 0.5 | | | | | | | | |
| 2005 2006 | | 77 74 | 5.5 2.3 | 8.9 3.6 | 13.9 5.5 | 2.4 0.9 | 3.7 | 5.6 1.9 | _ | | | | | | | |
| 2006 | | 52 | 2.3 5.4 | 9.7 | 5.5 16.8 | 1.9 | 3.1 | 4.7 | | | | | | | | |
| 2008 | | 76 52 | 0.4 0.6 | 0.7 | 1.2 1.8 | 0.1 | 0.3 | 0.4 0.6 | _ | | | | | | | |
| 2010 | | 78 | 0.4 | 0.8 | 1.2 | 0.2 | 0.2 | 0.3 | | | | | | | | |
| 2011 2012 | | 78 78 | 0.5 | 0.9 | 1.4 0.2 | 0.1 | 0.3 | 0.5 | | | | | | | | |
| 2013 | | 78 | 0.3 | 0.5 | 0.9 | 0.0 | 0.0 | 0.2 | | | | | | | | |
| 2014 2015 | | 78 78 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | | | | | | | | |
| 2016 | | 78 | 0.0 | 0.2 | 0.5 | 0.0 | 0.1 | 0.2 | | | | | | | | |
| 2017 | | 78 78 | 0.0 | 0.1 | 0.3 | 0.0 | 0.0 | 0.1 | | | | | | | | |
| 2011 0 | | 90 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | | | | | | | | |
| 2018 2019 | | | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | _ | | | | | | | |
| 2019 2020 | | 90 | | 0.0 | 0.0 | 0.0 | | | | | | | | | | |
| 2019 | | 90 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | | | | | | | | |
| 2019 2020 2021 | | | | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | | | | | | | | |





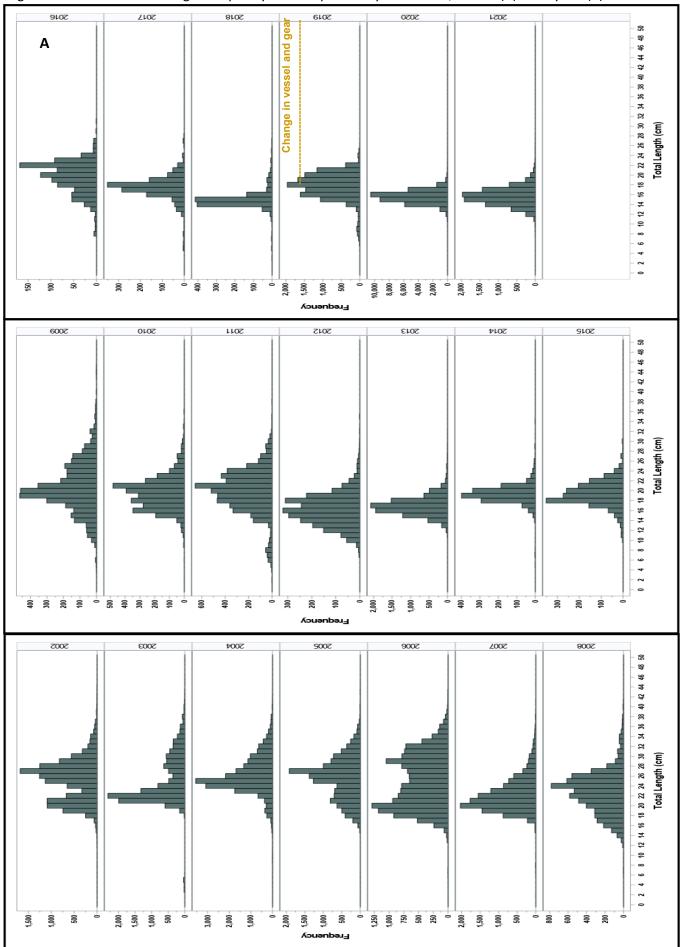
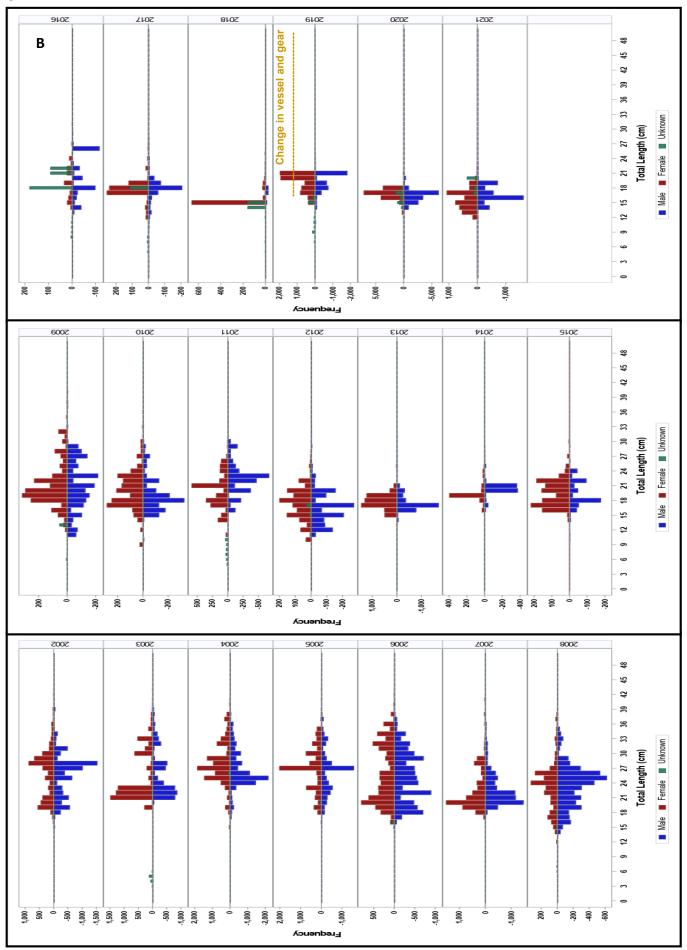
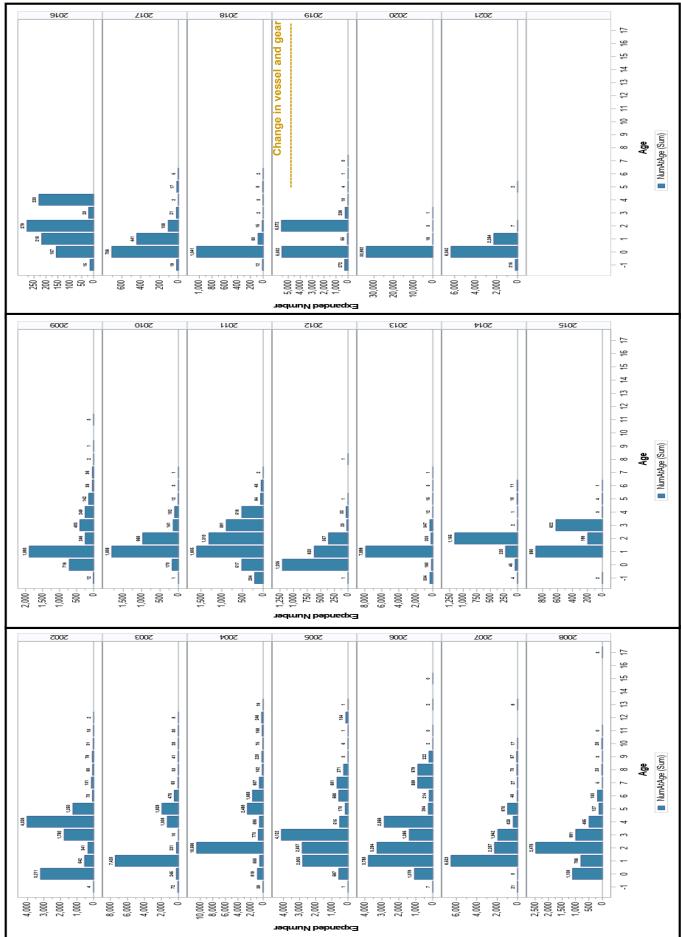


Figure 3. Atlantic Croaker length-frequency in Chesapeake Bay 2002-2021, overall (A) and by sex (B).

Figure 3. continued.







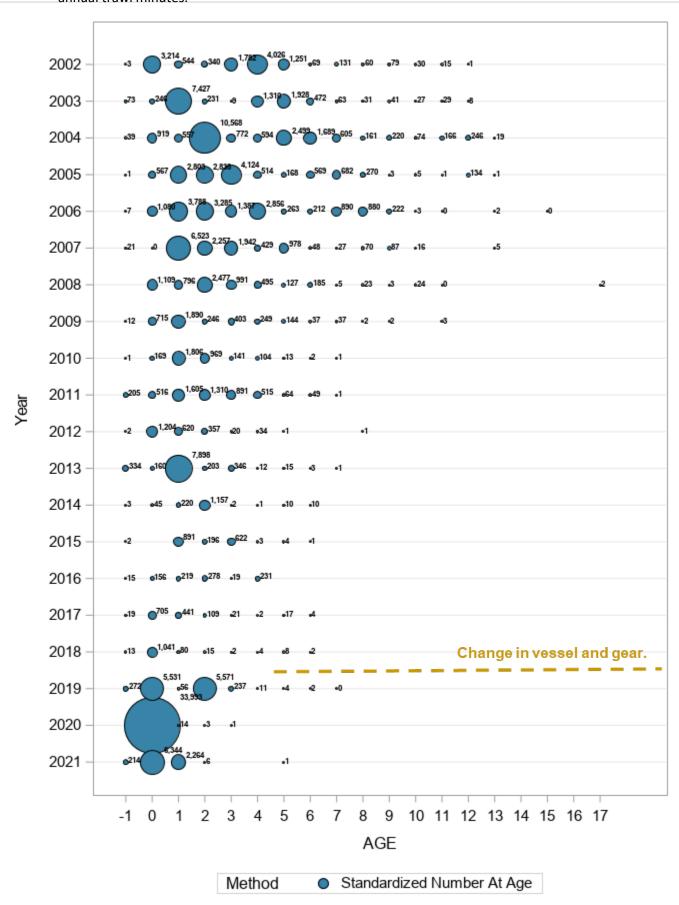


Figure 5. Atlantic Croaker age-frequency by year, 2002-2021 standardized to 8,000 (4,800 for 2019 and forward) annual trawl minutes.

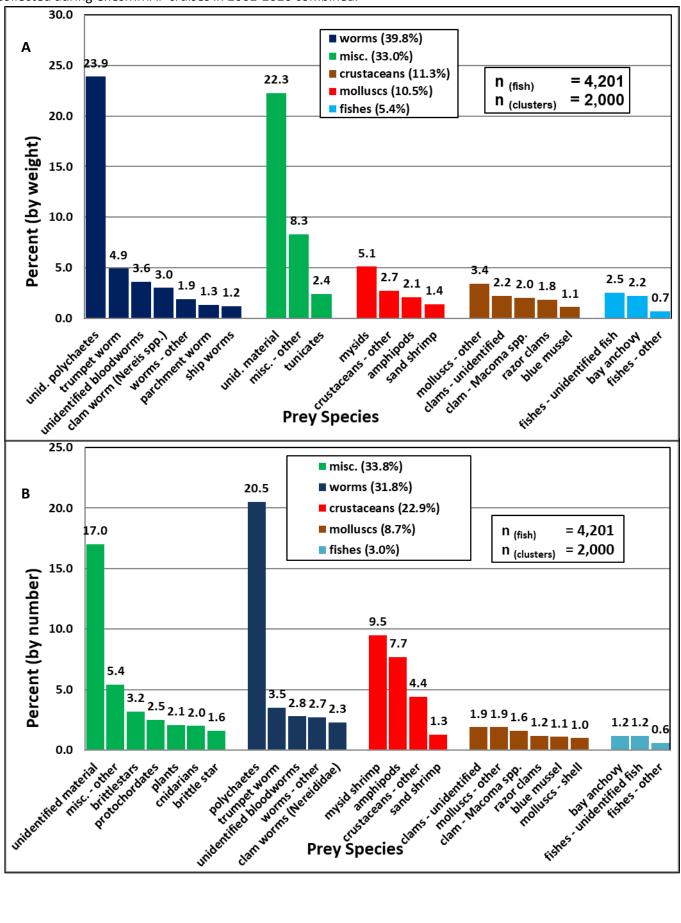
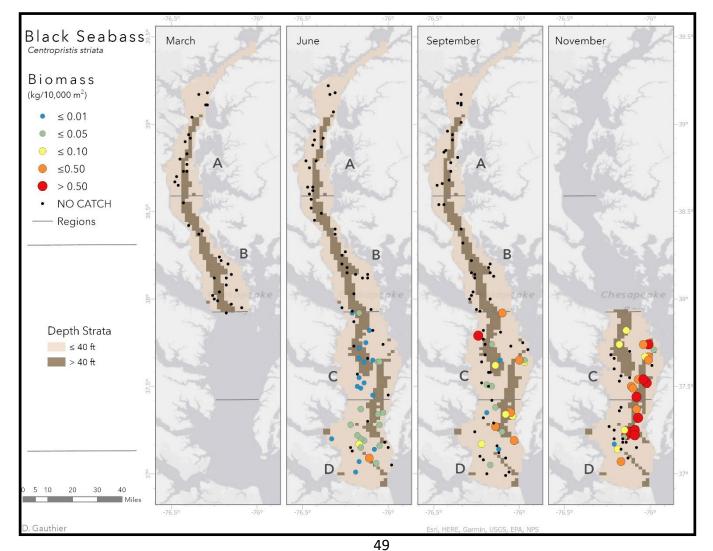


Figure 6. Diet composition, expressed as percent by weight (A) and percent by number (B) of Atlantic Croaker collected during ChesMMAP cruises in 2002-2020 combined.

| Year | Number Caught | Biomass Caught (kg) | Presence at Index Stations (%) | Number Measured | Age Specimens | Ages Read | Stomach Specimens | Stomachs Analyzed |
|------|------------------|------------------------|--------------------------------------|--------------------|------------------|--------------|----------------------|----------------------|
| 2002 | 50 | 4.4 | 14.8 | 50 | 48 | 48 | 46 | 46 |
| 2003 | 42 | 5.0 | 15.3 | 42 | 32 | 32 | 31 | 31 |
| 2004 | 14 | 2.2 | 7.8 | 14 | 14 | 14 | 14 | 14 |
| 2005 | 13 | 1.7 | 5.3 | 13 | 13 | 13 | 13 | 12 |
| 2006 | 22 | 1.7 | 6.7 | 22 | 17 | 17 | 16 | 16 |
| 2007 | 30 | 1.8 | 13.6 | 30 | 30 | 30 | 29 | 28 |
| 2008 | 34 | 2.2 | 5.9 | 34 | 28 | 28 | 26 | 25 |
| 2009 | 35 | 2.0 | 14.1 | 35 | 35 | 35 | 35 | 34 |
| 2010 | 23 | 0.6 | 8.9 | 23 | 23 | 23 | 22 | 22 |
| 2011 | 23 | 1.4 | 9.7 | 23 | 23 | 23 | 21 | 21 |
| 2012 | 9 | 0.4 | 2.3 | 9 | 9 | 9 | 8 | 7 |
| 2013 | 2 | 0.1 | 1.5 | 2 | 2 | 2 | 1 | 1 |
| 2014 | 11 | 0.6 | 3.7 | 11 | 11 | 11 | 8 | 8 |
| 2015 | 11 | 0.5 | 5.9 | 11 | 11 | 11 | 9 | 9 |
| 2016 | 42 | 2.0 | 16.3 | 42 | 42 | 42 | 30 | 30 |
| 2017 | 35 | 1.3 | 7.4 | 35 | 34 | 34 | 22 | 22 |
| 2018 | 8 | 0.4 | 1.5 | 8 | 8 | 8 | 4 | 4 |
| 2019 | 445 | 11.1 | 51.1 | 445 | 209 | 209 | 148 | 147 |
| 2020 | 507 | 16.7 | 60.7 | 507 | 256 | 256 | 192 | 190 |
| 2021 | 514 | 19.7 | 57.8 | 514 | 263 | 263 | 179 | 178 |

Table 8. Black Sea Bass sampling rates and preserved specimen analysis status by year (shaded rows represent a change in vessel and trawl gear beginning in June 2019).

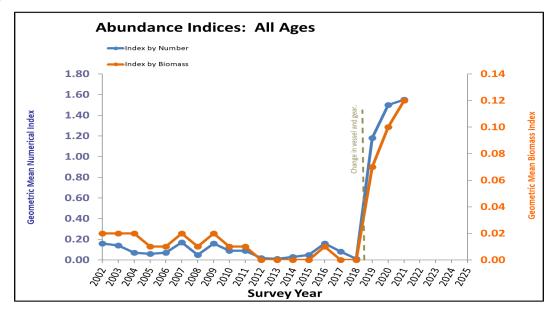
Figure 7. Station specific biomass Black Sea Bass in Chesapeake Bay, 2021.

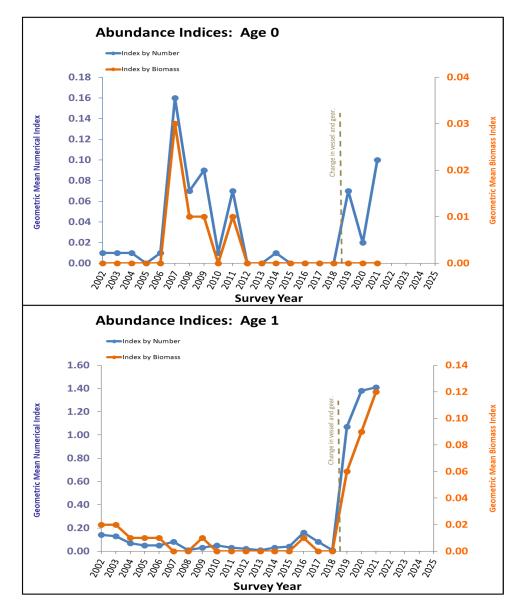


| Year | Age | n | Num | nerical Ind | lex | Bio | mass Inc | lex | Year | Age | n | Nun | nerical Inc | lex | Bio | mass Inc | dex |
|------|-----|-----|------|-------------|------|------|----------|------|------|-----|-----|------|-------------|------|------|----------|------|
| | 0- | | LCI | Index | UCI | LCI | Index | UCI | | 0- | | LCI | Index | UCI | LCI | Index | UCI |
| 2002 | All | 122 | 0.08 | 0.16 | 0.23 | 0.01 | 0.02 | 0.03 | 2002 | 1 | 122 | 0.07 | 0.14 | 0.21 | 0.01 | 0.02 | 0.02 |
| 2003 | | 149 | 0.07 | 0.14 | 0.21 | 0.01 | 0.02 | 0.04 | 2003 | | 149 | 0.06 | 0.13 | 0.19 | 0.01 | 0.02 | 0.03 |
| 2004 | | 127 | 0.03 | 0.07 | 0.12 | 0.00 | 0.02 | 0.03 | 2004 | | 127 | 0.02 | 0.07 | 0.11 | 0.00 | 0.01 | 0.03 |
| 2005 | | 131 | 0.01 | 0.06 | 0.10 | 0.00 | 0.01 | 0.02 | 2005 | | 131 | 0.01 | 0.05 | 0.10 | 0.00 | 0.01 | 0.02 |
| 2006 | | 120 | 0.01 | 0.07 | 0.14 | 0.00 | 0.01 | 0.02 | 2006 | | 120 | 0.00 | 0.05 | 0.10 | 0.00 | 0.01 | 0.02 |
| 2007 | | 88 | 0.07 | 0.17 | 0.27 | 0.01 | 0.02 | 0.03 | 2007 | | 88 | 0.01 | 0.08 | 0.16 | 0.00 | 0.00 | 0.01 |
| 2008 | | 135 | 0.01 | 0.05 | 0.10 | 0.00 | 0.01 | 0.02 | 2008 | | 135 | 0.00 | 0.01 | 0.02 | 0.00 | 0.00 | 0.00 |
| 2009 | | 135 | 0.08 | 0.16 | 0.24 | 0.01 | 0.02 | 0.03 | 2009 | | 135 | 0.00 | 0.03 | 0.06 | 0.00 | 0.01 | 0.01 |
| 2010 | | 135 | 0.03 | 0.09 | 0.14 | 0.00 | 0.01 | 0.01 | 2010 | | 135 | 0.01 | 0.05 | 0.09 | 0.00 | 0.00 | 0.01 |
| 2011 | | 134 | 0.04 | 0.09 | 0.15 | 0.00 | 0.01 | 0.02 | 2011 | | 134 | 0.00 | 0.03 | 0.06 | 0.00 | 0.00 | 0.01 |
| 2012 | | 129 | 0.00 | 0.02 | 0.05 | 0.00 | 0.00 | 0.00 | 2012 | | 129 | 0.00 | 0.02 | 0.05 | 0.00 | 0.00 | 0.00 |
| 2013 | | 134 | 0.00 | 0.01 | 0.03 | 0.00 | 0.00 | 0.00 | 2013 | | 134 | 0.00 | 0.01 | 0.02 | 0.00 | 0.00 | 0.00 |
| 2014 | | 135 | 0.00 | 0.03 | 0.06 | 0.00 | 0.00 | 0.00 | 2014 | | 135 | 0.00 | 0.03 | 0.06 | 0.00 | 0.00 | 0.00 |
| 2015 | | 135 | 0.02 | 0.05 | 0.09 | 0.00 | 0.00 | 0.01 | 2015 | | 135 | 0.01 | 0.04 | 0.07 | 0.00 | 0.00 | 0.01 |
| 2016 | | 135 | 0.09 | 0.16 | 0.24 | 0.01 | 0.01 | 0.02 | 2016 | | 135 | 0.09 | 0.16 | 0.24 | 0.01 | 0.01 | 0.02 |
| 2017 | | 135 | 0.02 | 0.08 | 0.14 | 0.00 | 0.00 | 0.01 | 2017 | | 135 | 0.02 | 0.08 | 0.14 | 0.00 | 0.00 | 0.01 |
| 2018 | | 135 | 0.00 | 0.01 | 0.02 | 0.00 | 0.00 | 0.00 | 2018 | | 135 | 0.00 | 0.01 | 0.02 | 0.00 | 0.00 | 0.00 |
| 2019 | | 135 | 0.85 | 1.18 | 1.56 | 0.05 | 0.07 | 0.10 | 2019 | | 135 | 0.79 | 1.07 | 1.40 | 0.05 | 0.06 | 0.08 |
| 2020 | | 135 | 1.11 | 1.50 | 1.96 | 0.08 | 0.10 | 0.13 | 2020 | | 135 | 1.03 | 1.38 | 1.79 | 0.07 | 0.09 | 0.12 |
| 2021 | | 135 | 1.16 | 1.55 | 2.00 | 0.09 | 0.12 | 0.16 | 2021 | | 135 | 1.06 | 1.41 | 1.81 | 0.08 | 0.12 | 0.15 |
| 2022 | | | | | | | | | 2022 | | | | | | | | |
| 2023 | | | | | | | | | 2023 | | | | | | | | |
| 2024 | | | | | | | | | 2024 | | | | | | | | |
| 2025 | | | | | | | | | 2025 | | | | | | | | |
| 2002 | 0 | 75 | 0.00 | 0.01 | 0.02 | 0.00 | 0.00 | 0.00 | | | | | | | | | |
| 2003 | | 101 | 0.00 | 0.01 | 0.02 | 0.00 | 0.00 | 0.00 | | | | | | | | | |
| 2004 | | 92 | 0.00 | 0.01 | 0.03 | 0.00 | 0.00 | 0.01 | | | | | | | | | |
| 2005 | | 86 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | | | | | | | | | |
| 2006 | | 79 | 0.00 | 0.01 | 0.04 | 0.00 | 0.00 | 0.00 | | | | | | | | | |
| 2007 | | 44 | 0.03 | 0.16 | 0.30 | 0.01 | 0.03 | 0.05 | | | | | | | | | |
| 2008 | | 90 | 0.01 | 0.07 | 0.14 | 0.00 | 0.01 | 0.03 | | | | | | | | | |
| 2009 | | 90 | 0.02 | 0.09 | 0.16 | 0.00 | 0.01 | 0.02 | | | | | | | | | |
| 2010 | | 90 | 0.00 | 0.01 | 0.03 | 0.00 | 0.00 | 0.00 | | | | | | | | | |
| 2011 | | 89 | 0.02 | 0.07 | 0.12 | 0.00 | 0.01 | 0.01 | | | | | | | | | |
| 2012 | | 84 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | | | | | | | | | |
| 2013 | | 89 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | | | | | | | | | |
| 2014 | | 90 | 0.00 | 0.01 | 0.03 | 0.00 | 0.00 | 0.00 | | | | | | | | | |
| 2015 | | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | | | | | | | | | |
| 2016 | | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | | | | | | | | | |
| 2017 | | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | | | | | | | | | |
| 2018 | | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | | | | | | | | | |
| 2019 | | 90 | 0.02 | 0.07 | 0.12 | 0.00 | 0.00 | 0.01 | | | | | | | | | |
| 2020 | | 90 | 0.00 | 0.02 | 0.05 | 0.00 | 0.00 | 0.00 | | | | | | | | | |
| 2021 | | 90 | 0.04 | 0.10 | 0.17 | 0.00 | 0.00 | 0.01 | | | | | | | | | |
| 2022 | | | | | | | | | | | | | | | | | |
| 2023 | | | | | | | | | | | | | | | | | |
| 2024 | | | | | | | | | | | | | | | | | |
| 2025 | | | | | | | | | | | | | | | | | |

Table 9. Black Sea Bass geometric mean indices of abundance, by number and biomass, overall and by age-class.

Figure 8. Black Sea Bass geometric mean indices of abundance, by number and biomass, for all ages combined and by age-class.





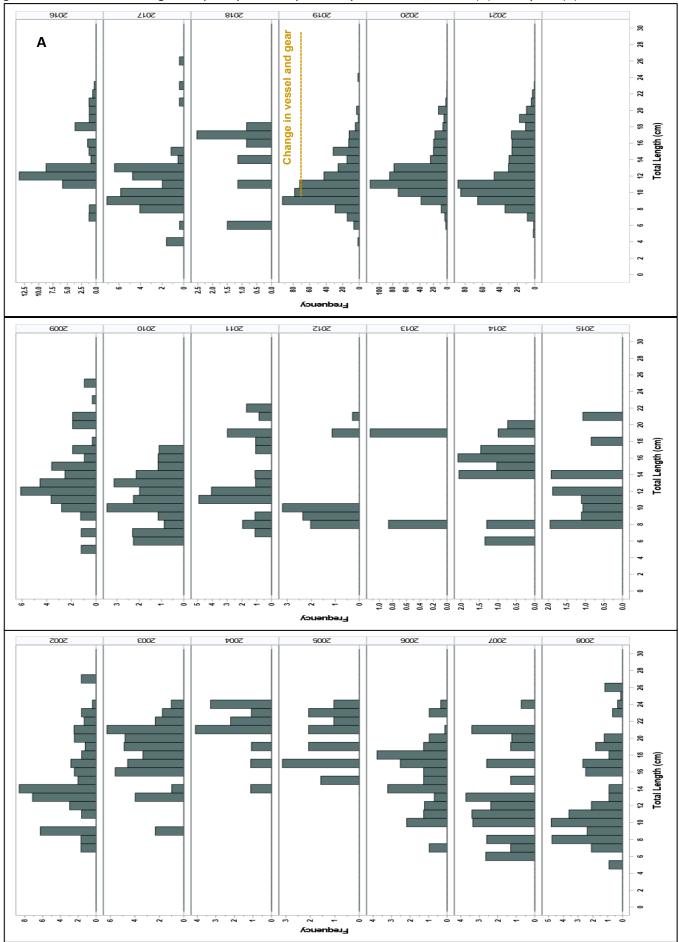
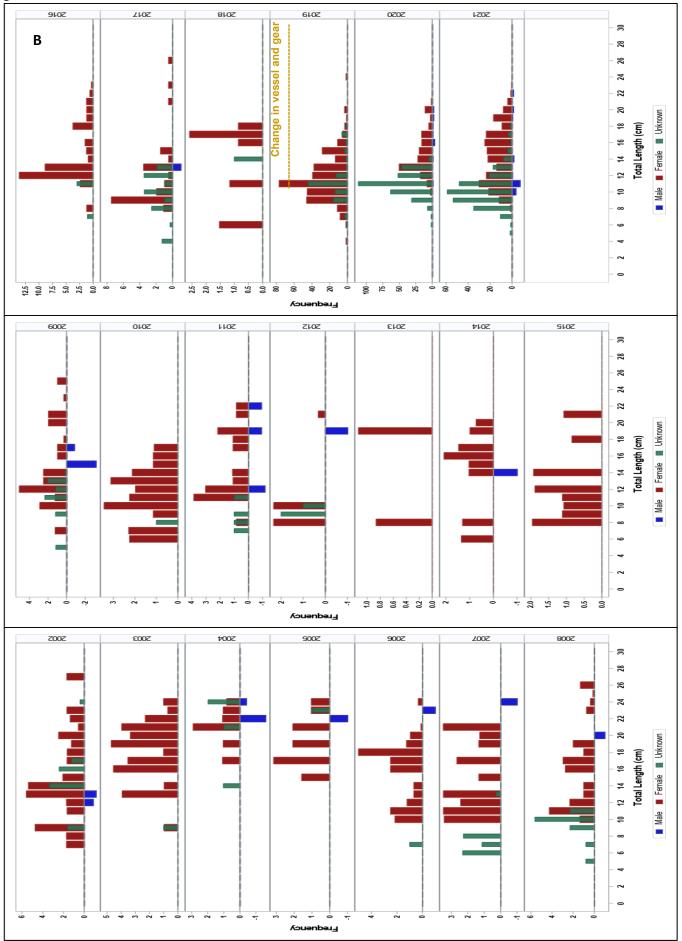
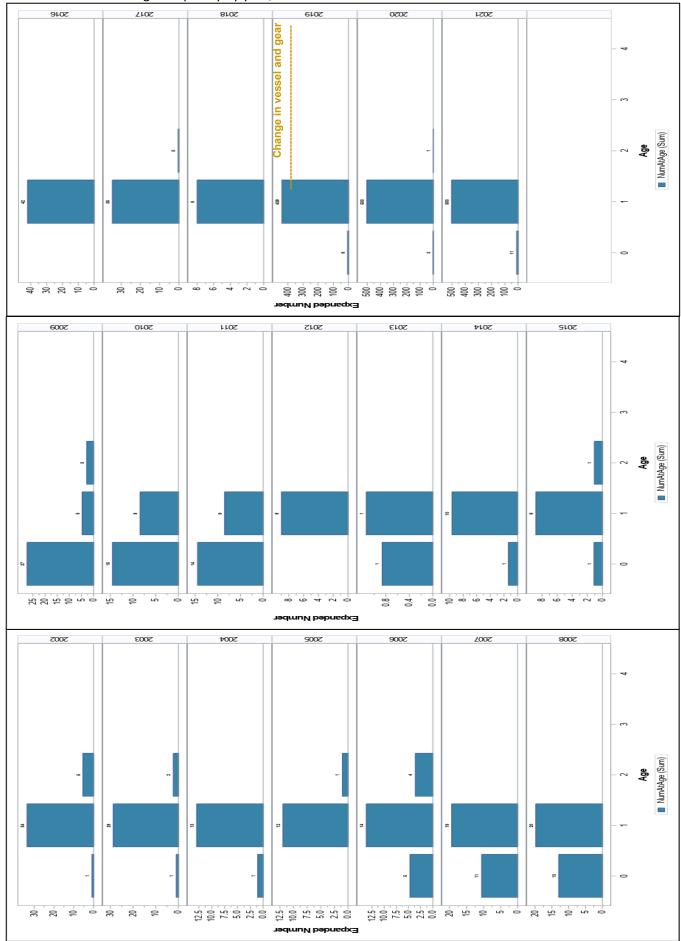




Figure 9. cont.







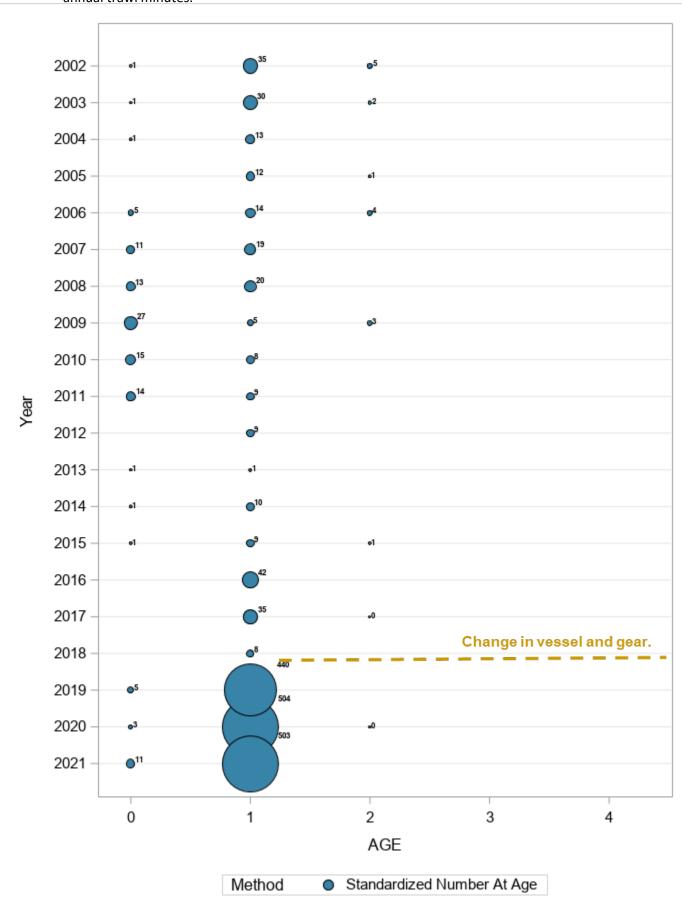


Figure 11. Black Sea Bass age-frequency by year, 2002-2021 standardized to 8,000 (4,800 for 2019 and forward) annual trawl minutes.

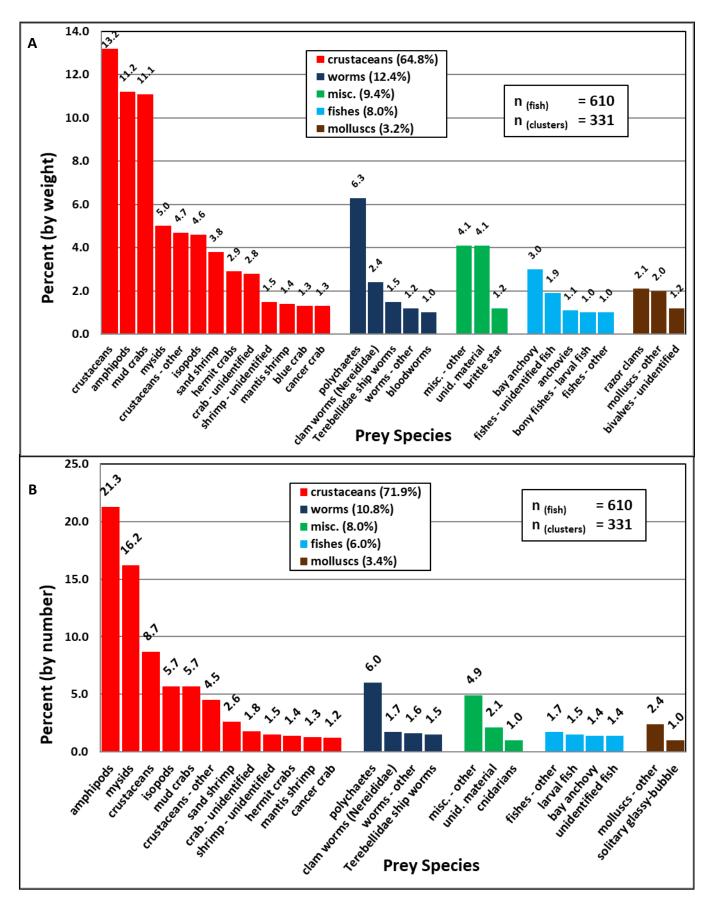


Figure 12. Diet composition, expressed as percent by weight (A) and percent by number (B) of black seabass collected during ChesMMAP cruises in 2002-2020 combined.

| | Number | Biomass | Presence at Index Stations | Number | Age | | Stomach | Stomachs |
|------|--------|-------------|-------------------------------|----------|-----------|-----------|-----------|----------|
| Year | Caught | Caught (kg) | (%) | Measured | Specimens | Ages Read | Specimens | Analyzed |
| 2002 | 34 | 10.7 | 10.3 | 34 | 34 | 34 | 24 | 22 |
| 2003 | 114 | 31.7 | 24.8 | 114 | 74 | 74 | 63 | 62 |
| 2004 | 28 | 10.0 | 11.8 | 28 | 27 | 27 | 22 | 22 |
| 2005 | 108 | 22.2 | 16.1 | 108 | 71 | 71 | 60 | 60 |
| 2006 | 23 | 5.5 | 9.9 | 23 | 23 | 23 | 17 | 17 |
| 2007 | 58 | 18.2 | 31.8 | 58 | 50 | 50 | 44 | 44 |
| 2008 | 52 | 15.8 | 8.0 | 52 | 27 | 27 | 14 | 13 |
| 2009 | 11 | 2.3 | 4.8 | 11 | 11 | 11 | 9 | 9 |
| 2010 | 126 | 20.2 | 4.0 | 82 | 30 | 30 | 13 | 12 |
| 2011 | 8 | 2.3 | 4.9 | 8 | 8 | 8 | 7 | 6 |
| 2012 | 17 | 4.0 | 6.7 | 17 | 17 | 17 | 12 | 12 |
| 2013 | 32 | 5.4 | 7.3 | 32 | 32 | 32 | 26 | 26 |
| 2014 | 44 | 5.9 | 12.8 | 44 | 39 | 39 | 26 | 25 |
| 2015 | 125 | 18.5 | 12.8 | 125 | 49 | 49 | 28 | 28 |
| 2016 | 36 | 9.8 | 7.2 | 36 | 36 | 36 | 19 | 19 |
| 2017 | 40 | 6.6 | 7.2 | 40 | 31 | 31 | 20 | 20 |
| 2018 | 85 | 8.4 | 12.7 | 85 | 41 | 41 | 24 | 24 |
| 2019 | 35 | 6.4 | 10.5 | 35 | 33 | 33 | 14 | 14 |
| 2020 | 208 | 23.2 | 22.4 | 208 | 97 | 97 | 54 | 54 |
| 2021 | 247 | 23.9 | 22.4 | 247 | 122 | 122 | 81 | 80 |

Table 10. Bluefish sampling rates and preserved specimen analysis status by year (shaded rows represent a change in vessel and trawl gear beginning in June 2019).

Figure 13. Station specific biomass of Bluefish in Chesapeake Bay, 2021.

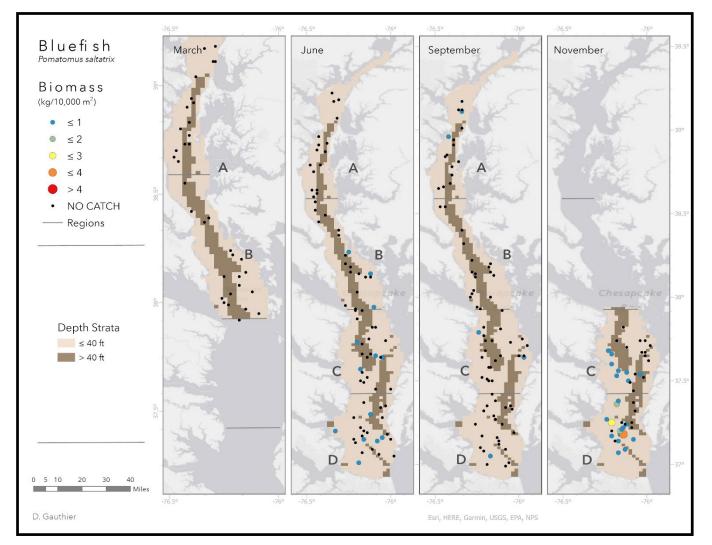
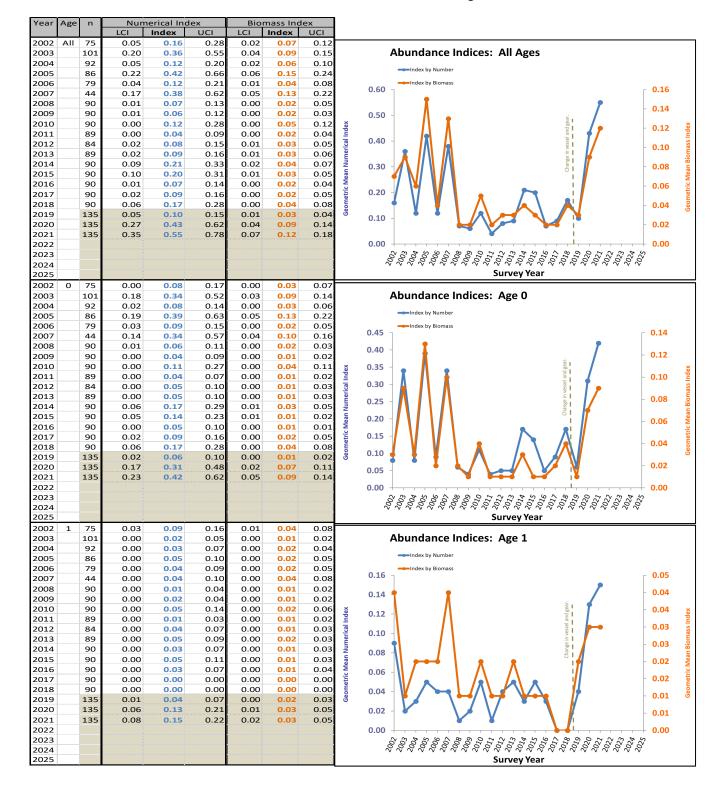


Table 11. Bluefish geometric mean indices of abundance, by number and biomass, overall and age-0.

Figure 14. Bluefish geometric mean indices of abundance, by number and biomass, for all ages combined and for ages 0 and 1.



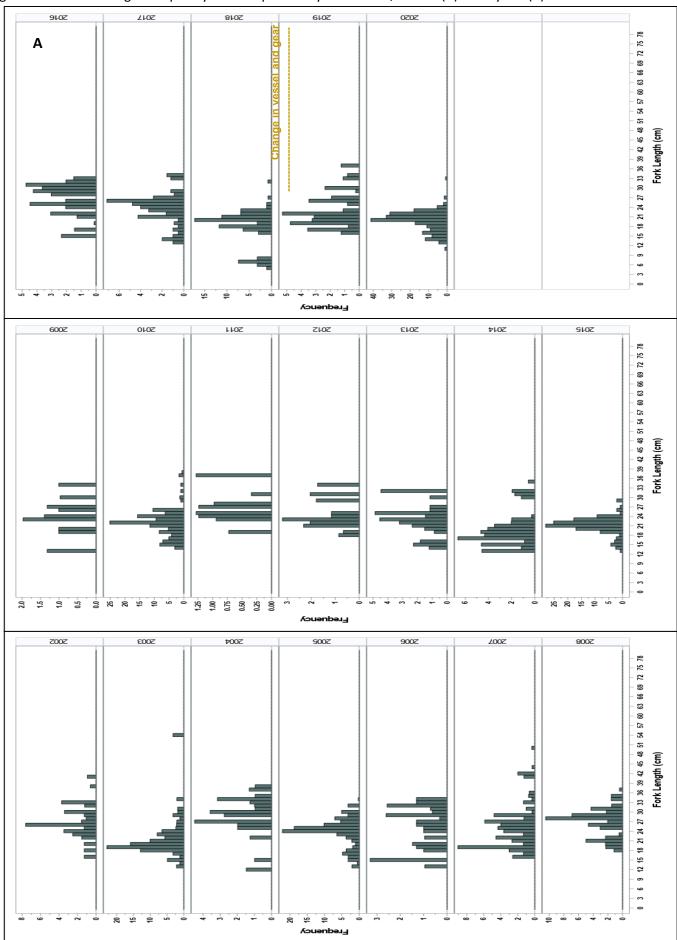




Figure 15. cont.

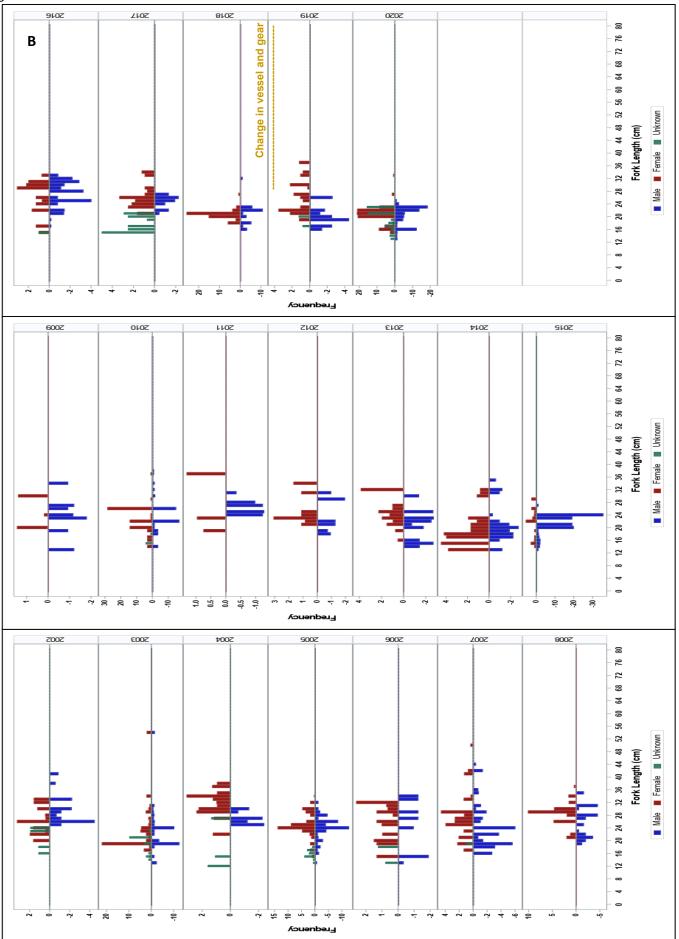
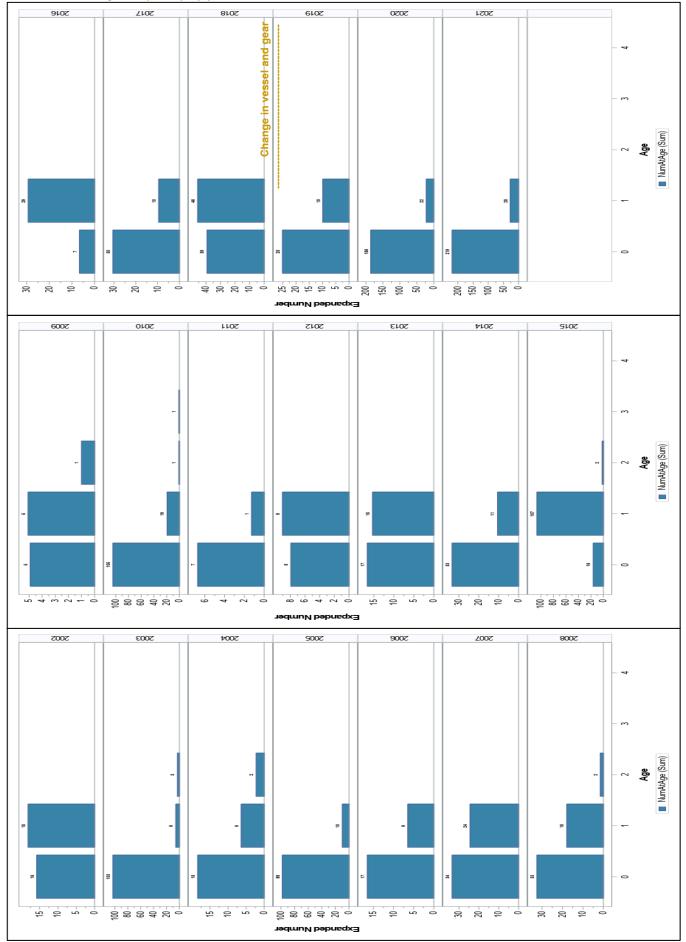


Figure 16. Bluefish age-frequency by year 2002-2021.



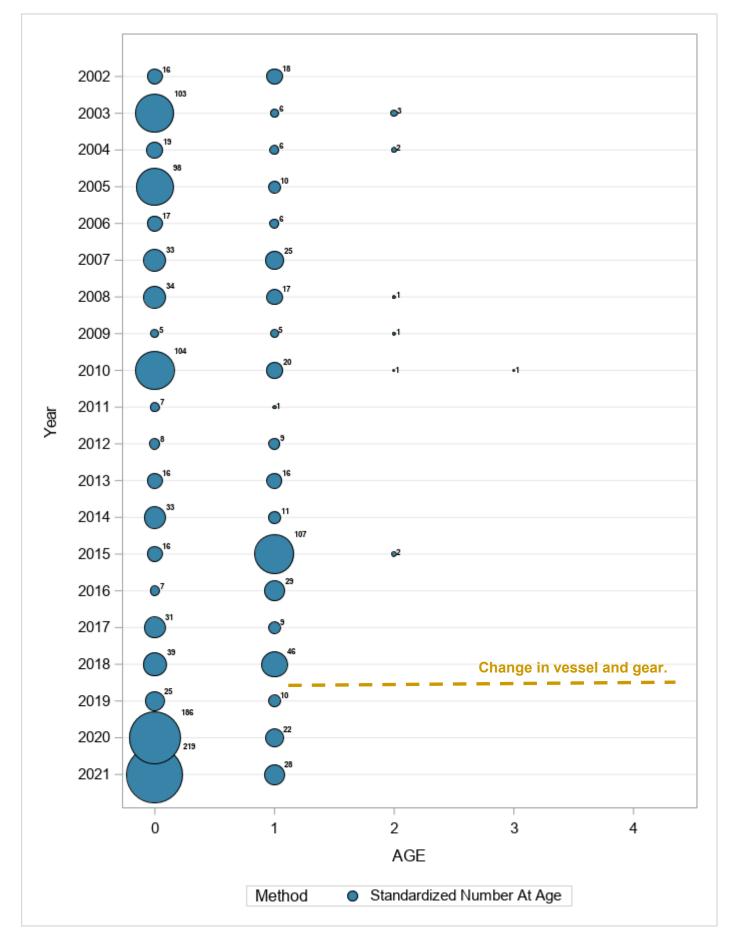


Figure 17. Bluefish age-frequency by year, 2002-2021 standardized to 8,000 (4,800 for 2019 and forward) annual trawl minutes.

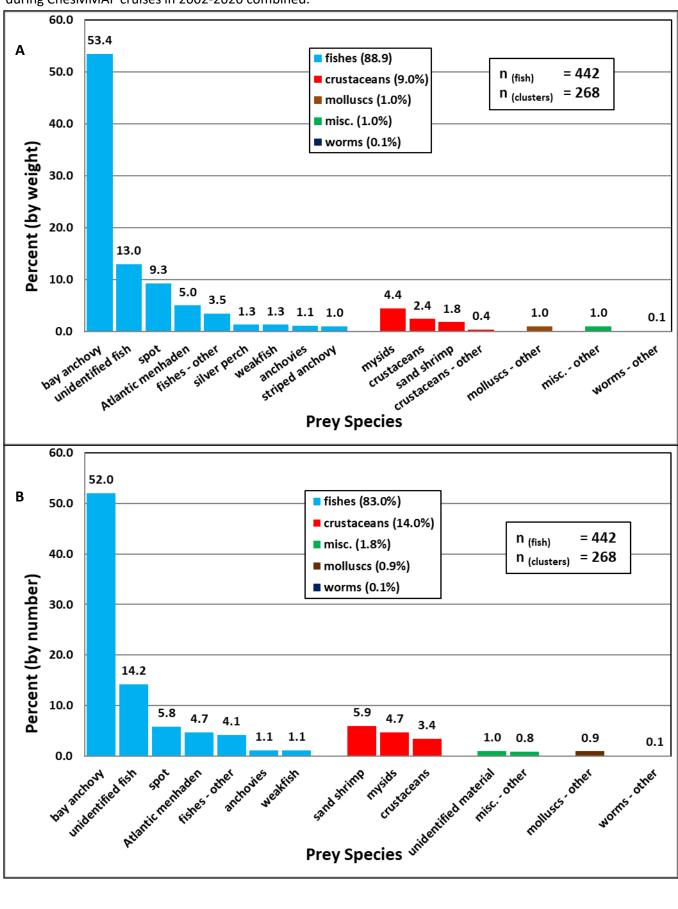
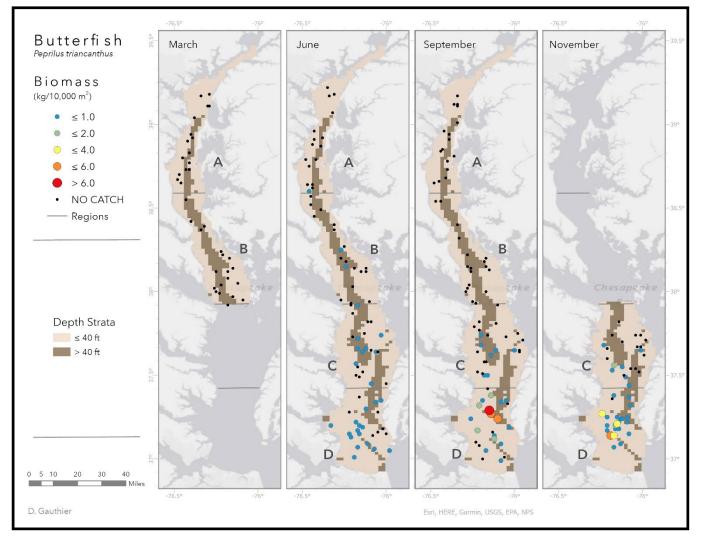


Figure 18. Diet composition, expressed as percent by weight (A) and percent by number (B) of Bluefish collected during ChesMMAP cruises in 2002-2020 combined.

| | Number | Biomass | Presence at Index Stations | Number | Age | | Stomach | Stomachs |
|------|--------|-------------|-------------------------------|----------|-----------|-----------|-----------|----------|
| Year | Caught | Caught (kg) | (%) | Measured | Specimens | Ages Read | Specimens | Analyzed |
| 2002 | 310 | 18.3 | 44.0 | 310 | 170 | 0 | 168 | 158 |
| 2003 | 1,000 | 57.4 | 62.7 | 1,000 | 334 | 0 | 334 | 17 |
| 2004 | 1,133 | 113.4 | 55.9 | 1,071 | 316 | 0 | 316 | 1 |
| 2005 | 693 | 48.0 | 57.0 | 693 | 294 | 0 | 293 | 0 |
| 2006 | 634 | 43.7 | 62.0 | 634 | 3 | 0 | 1 | 0 |
| 2007 | 204 | 18.8 | 47.7 | 204 | 0 | 0 | 0 | 0 |
| 2008 | 318 | 22.0 | 37.8 | 318 | 2 | 0 | 0 | 0 |
| 2009 | 415 | 18.7 | 55.6 | 415 | 0 | 0 | 0 | 0 |
| 2010 | 429 | 21.9 | 36.7 | 429 | 0 | 0 | 0 | 0 |
| 2011 | 366 | 22.5 | 44.9 | 366 | 0 | 0 | 0 | 0 |
| 2012 | 991 | 65.3 | 35.7 | 991 | 0 | 0 | 0 | 0 |
| 2013 | 220 | 9.6 | 29.2 | 220 | 1 | 0 | 0 | 0 |
| 2014 | 409 | 20.2 | 36.7 | 409 | 0 | 0 | 0 | 0 |
| 2015 | 402 | 25.6 | 21.1 | 402 | 0 | 0 | 0 | 0 |
| 2016 | 300 | 23.3 | 28.9 | 300 | 0 | 0 | 0 | 0 |
| 2017 | 408 | 21.8 | 36.7 | 408 | 0 | 0 | 0 | 0 |
| 2018 | 124 | 6.7 | 20.0 | 124 | 0 | 0 | 0 | 0 |
| 2019 | 828 | 39.9 | 35.6 | 828 | 0 | 0 | 0 | 0 |
| 2020 | 2,616 | 75.6 | 61.1 | 1,876 | 0 | 0 | 0 | 0 |
| 2021 | 1,569 | 73.9 | 57.8 | 1,569 | 0 | 0 | 0 | 0 |

Table 12. Butterfish sampling rates and preserved specimen analysis status by year (shaded rows represent a change in vessel and trawl gear beginning in June 2019).

Figure 19. Station specific biomass of Butterfish in Chesapeake Bay, 2021.



| Year | | | | | | | | | | | | | | | | | |
|--------------|-----|----------|--------------|---------------------|--------------|--------------|-------------------|--------------|------------------|-----|----------|------------|---------------------|--------------|------------|-------------------|--------------|
| rear | Age | n | LCI | nerical In Index | dex UCI | Bior LCI | nass Inc Index | dex UCI | Year | Age | n | Nun LCI | nerical In Index | dex UCI | Bio LCI | mass Inc Index | lex UCI |
| 2002 | All | 75 | 0.71 | 1.19 | 1.80 | 0.08 | 0.15 | 0.23 | 2002 | 2 | 75 | 0.14 | 0.26 | 0.39 | 0.01 | 0.03 | 0.05 |
| 2002 | | 101 | 1.54 | 2.27 | 3.21 | 0.18 | 0.26 | 0.36 | 2002 | ~ | 101 | 0.33 | 0.48 | 0.64 | 0.01 | 0.05 | 0.07 |
| 2004 | | 92 | 1.52 | 2.25 | 3.18 | 0.29 | 0.43 | 0.58 | 2004 | | 92 | 0.56 | 0.82 | 1.14 | 0.10 | 0.17 | 0.23 |
| 2005 | | 86 | 1.74 | 2.64 | 3.83 | 0.23 | 0.34 | 0.46 | 2005 | | 86 | 0.44 | 0.65 | 0.90 | 0.06 | 0.09 | 0.12 |
| 2006 | | 79 | 1.44 | 2.29 | 3.44 | 0.16 | 0.30 | 0.45 | 2006 | | 79 | 0.29 | 0.53 | 0.81 | 0.03 | 0.09 | 0.15 |
| 2007 | | 44 | 0.54 | 0.97 | 1.52 | 0.06 | 0.15 | 0.25 | 2007 | | 44 | 0.11 | 0.26 | 0.44 | 0.01 | 0.04 | 0.07 |
| 2008 | | 90 | 0.77 | 1.15 | 1.60 | 0.10 | 0.17 | 0.24 | 2008 | | 90 | 0.35 | 0.54 | 0.75 | 0.04 | 0.08 | 0.12 |
| 2009 2010 | | 90 90 | 1.32 0.65 | 1.91 1.11 | 2.65 1.69 | 0.11 0.09 | 0.16 | 0.22 0.26 | 2009 2010 | | 90 90 | 0.30 | 0.45 | 0.61 | 0.02 | 0.04 | 0.05 0.14 |
| 2010 | | 89 | 0.83 | 1.11 | 1.09 | 0.09 | 0.17 | 0.20 | 2010 | | 89 | 0.38 | 0.85 | 0.39 | 0.04 | 0.03 | 0.14 |
| 2012 | | 84 | 0.49 | 0.86 | 1.32 | 0.09 | 0.18 | 0.24 | 2012 | | 84 | 0.11 | 0.33 | 0.54 | 0.01 | 0.05 | 0.08 |
| 2013 | | 89 | 0.33 | 0.62 | 0.97 | 0.03 | 0.08 | 0.12 | 2013 | | 89 | 0.05 | 0.11 | 0.18 | 0.00 | 0.01 | 0.02 |
| 2014 | | 90 | 0.55 | 0.89 | 1.30 | 0.05 | 0.11 | 0.17 | 2014 | | 90 | 0.03 | 0.08 | 0.13 | 0.00 | 0.01 | 0.01 |
| 2015 | | 90 | 0.32 | 0.63 | 1.01 | 0.05 | 0.12 | 0.18 | 2015 | | 90 | 0.10 | 0.24 | 0.39 | 0.01 | 0.04 | 0.07 |
| 2016 | | 90 | 0.43 | 0.72 | 1.08 | 0.07 | 0.14 | 0.21 | 2016 | | 90 | 0.00 | 0.01 | 0.03 | 0.00 | 0.00 | 0.00 |
| 2017 | | 90 | 0.61 | 1.03 | 1.57 | 0.07 | 0.13 | 0.20 | 2017 | | 90 | 0.00 | 0.04 | 0.08 | 0.00 | 0.01 | 0.02 |
| 2018 2019 | | 90 | 0.13 | 0.28 | 0.45 | 0.00 | 0.04 | 0.07 | 2018 2019 | | 90 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2019 | | 90 90 | 0.51 3.08 | 0.86 4.99 | 1.29 7.80 | 0.07 0.32 | 0.14 0.48 | 0.22 0.67 | 2019 | | 90 | 0.05 | 0.09 0.05 | 0.14 0.08 | 0.00 | 0.01 | 0.02 |
| 2020 | | 90 | 1.92 | 2.83 | 4.03 | 0.32 | 0.48 | 0.55 | 2020 | | 90 | 0.03 | 0.18 | 0.08 | 0.00 | 0.02 | 0.03 |
| 2022 | | 50 | 1.52 | | | 0.20 | | 0.00 | 2022 | | 50 | 0.111 | | 0125 | 0.01 | | 0.000 |
| 2023 | | | | | | | | | 2023 | | | | | | | | |
| 2024 | | | | | | | | | 2024 | | | | | | | | |
| 2025 | | | | | | | | | 2025 | | | | | | | | |
| 2002 | 0 | 75 | 0.30 | 0.54 | 0.81 | 0.02 | 0.05 | 0.08 | 2002 | 3 | 75 | 0.03 | 0.08 | 0.12 | 0.00 | 0.01 | 0.02 |
| 2003 | | 101 | 0.72 | 1.09 | 1.53 | 0.06 | 0.09 | 0.13 | 2003 | | 101 | 0.10 | 0.15 | 0.20 | 0.01 | 0.02 | 0.02 |
| 2004 | | 92 86 | 0.31 0.64 | 0.45 | 0.61 | 0.02 | 0.04 | 0.05 0.10 | 2004 2005 | | 92 86 | 0.27 | 0.41 | 0.57 | 0.04 | 0.08 | 0.11 |
| 2005 2006 | | 80 79 | 0.64 | 1.00 | 1.40 | 0.04 | 0.07 | 0.10 | 2005 | | 80 79 | 0.16 | 0.24 | 0.33 | 0.02 | 0.03 | 0.05 |
| 2000 | | 44 | 0.13 | 0.27 | 0.42 | 0.04 | 0.02 | 0.10 | 2000 | | 44 | 0.03 | 0.09 | 0.34 | 0.01 | 0.04 | 0.00 |
| 2008 | | 90 | 0.07 | 0.15 | 0.24 | 0.00 | 0.01 | 0.01 | 2008 | | 90 | 0.17 | 0.30 | 0.44 | 0.03 | 0.05 | 0.07 |
| 2009 | | 90 | 0.15 | 0.27 | 0.41 | 0.01 | 0.01 | 0.02 | 2009 | | 90 | 0.10 | 0.17 | 0.25 | 0.01 | 0.02 | 0.02 |
| 2010 | | 90 | 0.10 | 0.21 | 0.33 | 0.00 | 0.01 | 0.02 | 2010 | | 90 | 0.15 | 0.28 | 0.41 | 0.01 | 0.03 | 0.05 |
| 2011 | | 89 | 0.32 | 0.53 | 0.76 | 0.02 | 0.05 | 0.08 | 2011 | | 89 | 0.04 | 0.09 | 0.15 | 0.00 | 0.01 | 0.02 |
| 2012 | | 84 | 0.04 | 0.15 | 0.27 | 0.00 | 0.02 | 0.05 | 2012 | | 84 | 0.02 | 0.04 | 0.07 | 0.00 | 0.00 | 0.01 |
| 2013 | | 89 | 0.23 | 0.44 | 0.69 | 0.02 | 0.04 | 0.07 | 2013 | | 89 90 | 0.01 | 0.02 | 0.04 | 0.00 | 0.00 | 0.01 |
| 2014 2015 | | 90 90 | 0.34 | 0.60 | 0.90 | 0.02 | 0.06 0.01 | 0.10 0.02 | 2014 2015 | | 90 | 0.00 | 0.01 | 0.02 | 0.00 | 0.00 | 0.00 0.00 |
| 2015 | | 90 | 0.07 | 0.13 | 0.20 | 0.00 | 0.01 | 0.02 | 2015 | | 90 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 |
| 2017 | | 90 | 0.23 | 0.43 | 0.66 | 0.02 | 0.04 | 0.06 | 2017 | | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2018 | | 90 | 0.09 | 0.19 | 0.30 | 0.00 | 0.02 | 0.03 | 2018 | | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2019 | | 90 | 0.33 | 0.57 | 0.86 | 0.03 | 0.07 | 0.11 | 2019 | | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2020 | | 90 | 2.67 | 4.36 | 6.81 | 0.28 | 0.43 | 0.60 | 2020 | | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2021 | | 90 | 1.28 | 1.94 | 2.77 | 0.14 | 0.23 | 0.34 | 2021 | | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2022 | | | | | | | | | 2022 | | | | | | | | |
| 2023 | | | | | | | | | 2023 | | | | | | | | |
| 2024 2025 | | | | | | | | | 2024 2025 | | | | | | | | |
| 2023 | 1 | 75 | 0.45 | 0.76 | 1.12 | 0.04 | 0.09 | 0.14 | 2023 | 4+ | 75 | 0.00 | 0.01 | 0.01 | 0.00 | 0.00 | 0.00 |
| 2002 | | 101 | 0.91 | 1.34 | 1.85 | 0.09 | 0.14 | 0.14 | 2002 | | 101 | 0.00 | 0.01 | 0.01 | 0.00 | 0.00 | 0.00 |
| 2004 | | 92 | 0.87 | 1.27 | 1.76 | 0.15 | 0.23 | 0.31 | 2004 | | 92 | 0.05 | 0.12 | 0.19 | 0.01 | 0.02 | 0.03 |
| 2005 | | 86 | 1.06 | 1.59 | 2.25 | 0.13 | 0.20 | 0.27 | 2005 | | 86 | 0.02 | 0.04 | 0.06 | 0.00 | 0.01 | 0.01 |
| 2006 | | 79 | 0.85 | 1.35 | 2.00 | 0.09 | 0.18 | 0.27 | 2006 | | 79 | 0.01 | 0.04 | 0.08 | 0.00 | 0.01 | 0.01 |
| 2007 | | 44 | 0.35 | 0.66 | 1.03 | 0.03 | 0.10 | 0.17 | 2007 | | 44 | 0.00 | 0.03 | 0.06 | 0.00 | 0.01 | 0.01 |
| 2008 2009 | | 90 90 | 0.31 0.90 | 0.50 | 0.72 | 0.03 | 0.05 | 0.07 | 2008 2009 | | 90 90 | 0.04 | 0.09 | 0.14 | 0.01 | 0.01 | 0.02 |
| 2009 | | 90 90 | 0.90 | 1.33 0.57 | 1.86 0.87 | 0.07 0.03 | 0.11 | 0.15 0.12 | 2009 | | 90 90 | 0.02 | 0.04 0.04 | 0.06 0.06 | 0.00 | 0.00 | 0.01 |
| 2010 | | 89 | 0.32 | 0.37 | 1.10 | 0.03 | 0.10 | 0.12 | 2010 | | 89 | 0.01 | 0.04 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2012 | | 84 | 0.40 | 0.71 | 1.07 | 0.07 | 0.14 | 0.21 | 2012 | | 84 | 0.00 | 0.00 | 0.00 | | 0.00 | 0.00 |
| 2013 | | 89 | 0.16 | 0.33 | 0.51 | 0.01 | 0.03 | 0.06 | 2013 | | 89 | 0.00 | 0.00 | 0.00 | | 0.00 | 0.00 |
| 2014 | | 90 | 0.23 | 0.40 | 0.60 | 0.02 | 0.05 | 0.09 | 2014 | | 90 | 0.00 | 0.00 | 0.00 | | 0.00 | 0.00 |
| 2015 | | 90 | 0.25 | 0.50 | 0.80 | 0.04 | 0.08 | 0.13 | 2015 | | 90 | 0.00 | 0.00 | 0.00 | | 0.00 | 0.00 |
| 2016 | | 90 | 0.40 | 0.67 | 1.01 | 0.06 | 0.13 | 0.19 | 2016 | | 90 | 0.00 | 0.00 | 0.00 | | 0.00 | 0.00 |
| 2017 | | 90 | 0.43 | 0.76 | 1.15 | 0.05 | 0.10 | 0.14 | 2017 | | 90 | 0.00 | 0.00 | 0.00 | | 0.00 | 0.00 |
| 2018 2019 | _ | 90 90 | 0.03 | 0.15 | 0.28 | 0.00 | 0.03 | 0.06 0.14 | 2018 2019 | | 90 90 | 0.00 | 0.00 | 0.00 | | 0.00 | 0.00 |
| 2019 | | 90 | 0.21 | 0.43 | 1.35 | 0.04 | 0.09 | 0.14 | 2019 | | 90 | 0.00 | 0.00 | 0.00 | | 0.00 | 0.00 |
| 2020 | | 90 | 0.77 | 1.17 | 1.66 | 0.03 | 0.20 | 0.29 | 2020 | | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2022 | | | | | | | | | 2022 | | | | | | | | |
| 2023 | | | | | | | | | 2023 | | | | | | | | |
| 2024 | | | | | | | | | 2024 | | | | | | | | |
| 2025 | | | | | | | | | 2025 | | | | | | | | |

Table 13. Butterfish geometric mean indices of abundance, by number and biomass, overall and by age-class.

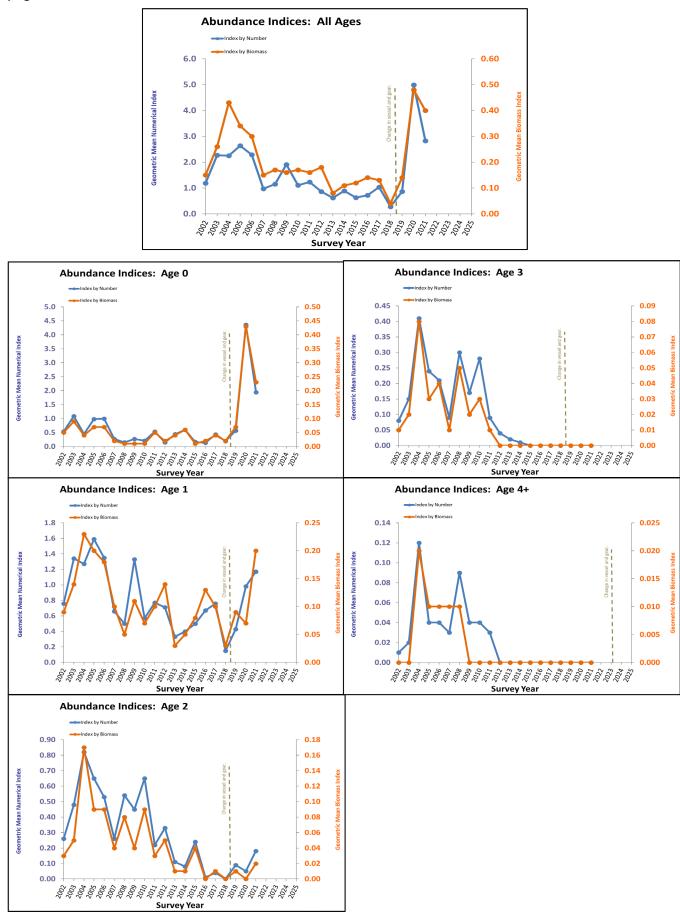
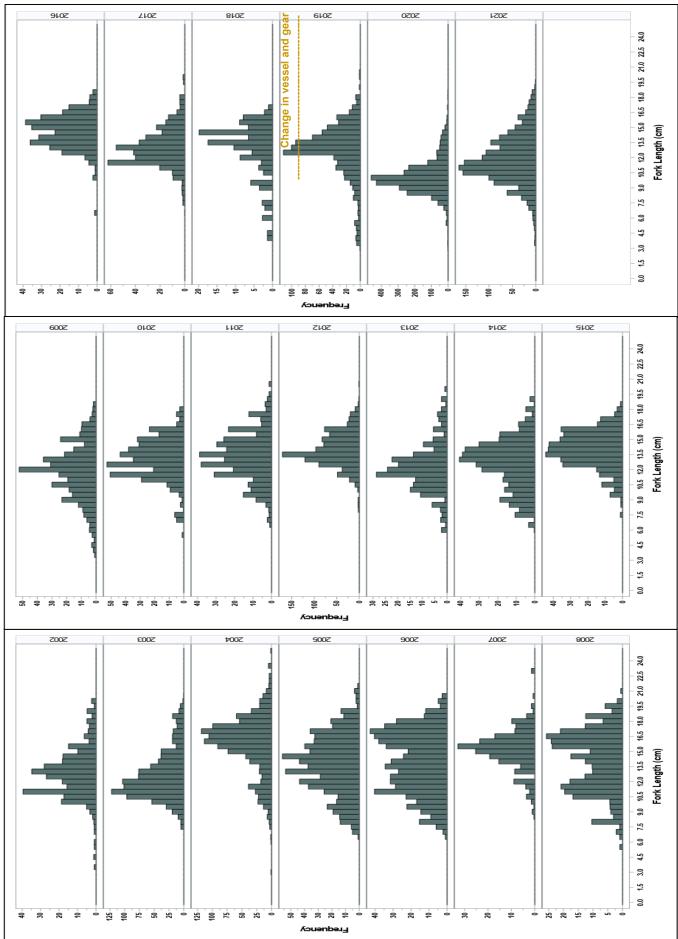
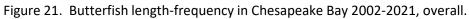


Figure 20. Butterfish geometric mean indices of abundance, by number and biomass, for all ages combined and by age-class.

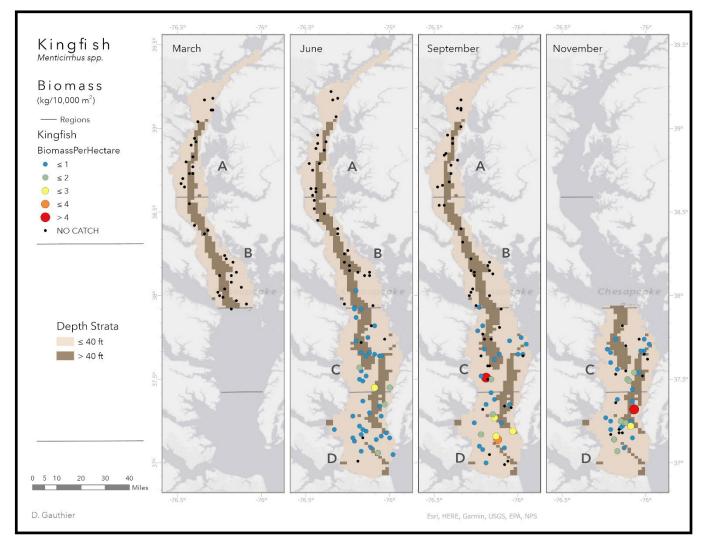




| | Number | Biomass | Presence at Index Stations | Number | Age | | Stomach | Stomachs |
|------|--------|---------|-------------------------------|----------|-----------|-----------|-----------|----------|
| Year | Caught | Caught | (%) | Measured | Specimens | Ages Read | Specimens | Analyzed |
| 2002 | 143 | 18.5 | 22.4 | 143 | 91 | 91 | 87 | 79 |
| 2003 | 68 | 19.2 | 12.9 | 68 | 55 | 55 | 55 | 50 |
| 2004 | 67 | 16.0 | 14.0 | 67 | 55 | 55 | 50 | 48 |
| 2005 | 86 | 15.3 | 19.3 | 86 | 72 | 72 | 69 | 68 |
| 2006 | 120 | 24.1 | 26.1 | 120 | 94 | 94 | 84 | 83 |
| 2007 | 122 | 17.7 | 25.6 | 122 | 88 | 88 | 78 | 76 |
| 2008 | 333 | 62.6 | 21.7 | 300 | 113 | 113 | 97 | 97 |
| 2009 | 195 | 24.8 | 36.3 | 195 | 152 | 152 | 135 | 134 |
| 2010 | 447 | 82.5 | 35.8 | 447 | 231 | 231 | 206 | 199 |
| 2011 | 336 | 55.7 | 32.4 | 336 | 176 | 175 | 155 | 155 |
| 2012 | 148 | 24.6 | 25.9 | 148 | 114 | 0 | 96 | 92 |
| 2013 | 165 | 32.1 | 24.0 | 165 | 106 | 0 | 77 | 77 |
| 2014 | 76 | 14.2 | 12.8 | 76 | 57 | 0 | 39 | 36 |
| 2015 | 156 | 24.1 | 19.4 | 156 | 112 | 0 | 61 | 60 |
| 2016 | 613 | 80.1 | 42.8 | 613 | 265 | 0 | 166 | 163 |
| 2017 | 361 | 55.2 | 30.6 | 361 | 198 | 0 | 138 | 137 |
| 2018 | 239 | 39.0 | 37.2 | 239 | 167 | 0 | 104 | 104 |
| 2019 | 3,871 | 435.9 | 71.9 | 2,904 | 331 | 0 | 217 | 213 |
| 2020 | 5,767 | 579.1 | 88.1 | 3,163 | 282 | 0 | 192 | 191 |
| 2021 | 1,409 | 188.9 | 75.6 | 1,409 | 264 | 0 | 181 | 177 |

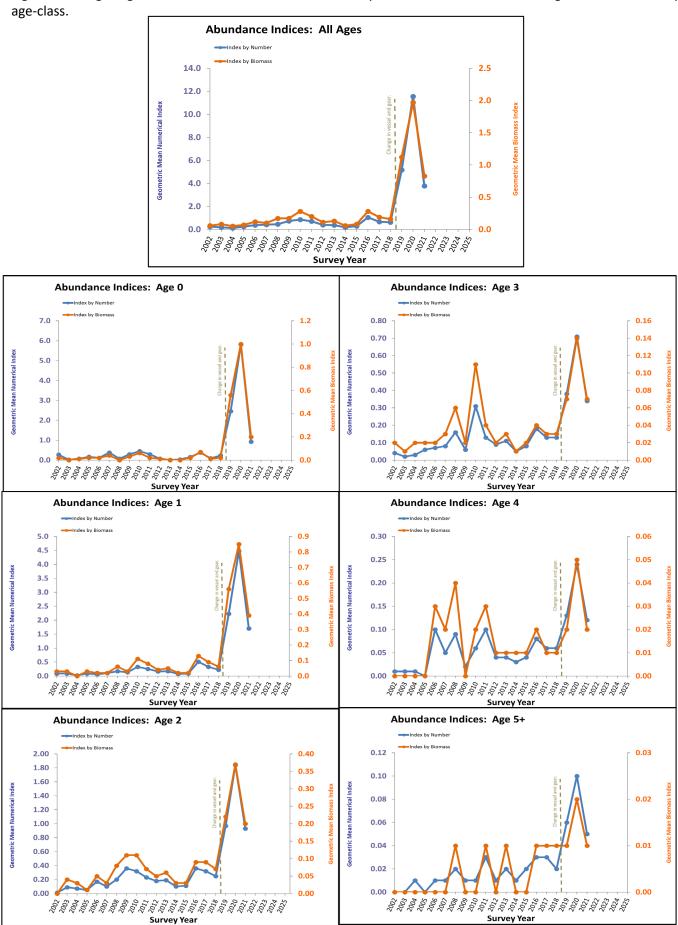
Table 14. Kingfish sampling rates and preserved specimen analysis status by year (shaded rows represent a change in vessel and trawl gear beginning in June 2019).

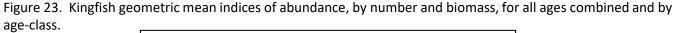
Figure 22. Station specific biomass of Kingfish in Chesapeake Bay, 2021.

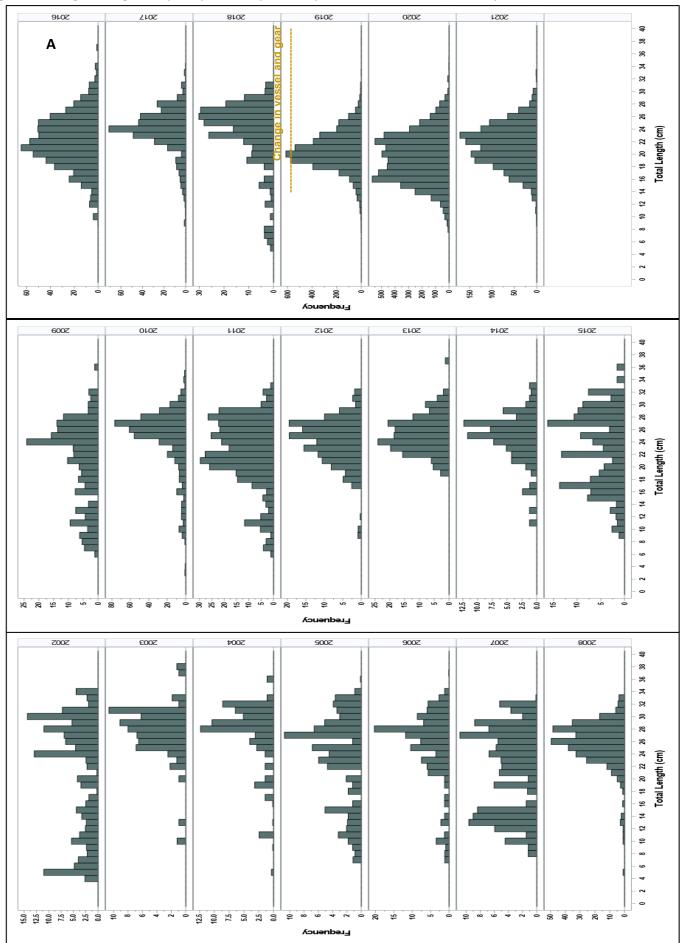


| Tab | ole 15. | Kir | ngfish geometric mea | n indices of abund | ance, b | y num | ber an | d biomass, ov | verall and by age-class. |
|-----|---------|-----|----------------------|--------------------|---------|-------|--------|-----------------|--------------------------|
| Ve | | | Numerical Index | Riemass Index | Vear | A | , î | Numerical Index | Biomass Index |

| Year | Age | n | | merical In | | | mass Inc | dex | Year | Age | n | Nur | nerical In | | Bio | mass Inc | lex |
|--------------|-----|------------|--------------|--------------|--------------|--------------|----------|--------------|--------------|-----|------------|-----------|------------|-----------|-----------|----------|--------------|
| | | | LCI | Index | UCI | LCI | Index | UCI | | | | LCI | Index | UCI | LCI | Index | UCI |
| 2002 | All | 165 | 0.16 | 0.25 | 0.35 | 0.03 | 0.06 | 0.09 | 2002 | 3 | 165 | 0.01 | 0.04 | 0.07 | 0.00 | 0.02 | 0.03 |
| 2003 2004 | | 149 177 | 0.10 | 0.18 | 0.28 | 0.04 | 0.08 | 0.12 0.08 | 2003 2004 | | 149 177 | 0.00 | 0.02 | 0.04 | 0.00 | 0.01 | 0.02 |
| 2004 | | 176 | 0.16 | 0.25 | 0.36 | 0.02 | 0.07 | 0.10 | 2004 | | 176 | 0.02 | 0.06 | 0.11 | 0.00 | 0.02 | 0.03 |
| 2006 | | 165 | 0.25 | 0.38 | 0.53 | 0.08 | 0.12 | 0.17 | 2006 | | 165 | 0.04 | 0.07 | 0.11 | 0.01 | 0.02 | 0.04 |
| 2007 | | 133 | 0.25 | 0.41 | 0.58 | 0.06 | 0.10 | 0.15 | 2007 | | 133 | 0.03 | 0.08 | 0.13 | 0.01 | 0.03 | 0.05 |
| 2008 | | 180 | 0.29 | 0.46 | 0.66 | 0.09 | 0.17 | 0.26 | 2008 | | 180 | 0.07 | 0.16 | 0.25 | 0.02 | 0.06 | 0.10 |
| 2009 2010 | | 135 179 | 0.50 | 0.73 | 1.00 1.16 | 0.11 0.18 | 0.17 | 0.23 0.38 | 2009 2010 | | 135 179 | 0.02 | 0.06 | 0.10 0.43 | 0.00 | 0.02 | 0.04 0.15 |
| 2010 | | 179 | 0.50 | 0.30 | 0.95 | 0.13 | 0.20 | 0.38 | 2010 | | 179 | 0.07 | 0.13 | 0.43 | 0.00 | 0.04 | 0.06 |
| 2012 | | 174 | 0.26 | 0.39 | 0.54 | 0.07 | 0.11 | 0.15 | 2012 | | 174 | 0.06 | 0.09 | 0.12 | 0.01 | 0.02 | 0.03 |
| 2013 | | 179 | 0.25 | 0.38 | 0.53 | 0.08 | 0.13 | 0.17 | 2013 | | 179 | 0.07 | 0.11 | 0.15 | 0.02 | 0.03 | 0.04 |
| 2014 | | 180 | 0.10 | 0.19 | 0.28 | 0.03 | 0.06 | 0.09 | 2014 | | 180 | 0.02 | 0.05 | 0.08 | 0.01 | 0.01 | 0.02 |
| 2015 2016 | | 180 180 | 0.18 0.75 | 0.29 | 0.41 1.37 | 0.05 | 0.08 | 0.12 0.37 | 2015 2016 | | 180 180 | 0.04 0.13 | 0.08 | 0.12 | 0.01 0.03 | 0.02 | 0.04 0.06 |
| 2010 | | 180 | 0.45 | 0.66 | 0.89 | 0.13 | 0.19 | 0.26 | 2010 | | 180 | 0.09 | 0.13 | 0.18 | 0.03 | 0.04 | 0.00 |
| 2018 | | 180 | 0.45 | 0.62 | 0.81 | 0.11 | 0.16 | 0.21 | 2018 | | 180 | 0.09 | 0.13 | 0.17 | 0.02 | 0.03 | 0.04 |
| 2019 | | 135 | 3.61 | 5.16 | 7.24 | 0.80 | 1.12 | 1.51 | 2019 | | 135 | 0.27 | 0.38 | 0.49 | 0.05 | 0.07 | 0.10 |
| 2020 | | 135 | 8.72 | 11.55 | 15.21 | 1.55 | 1.97 | 2.47 | 2020 | | 135 | 0.55 | 0.71 | 0.88 | 0.10 | 0.14 | 0.17 |
| 2021 2022 | | 135 | 2.86 | 3.78 | 4.91 | 0.65 | 0.83 | 1.03 | 2021 2022 | | 135 | 0.26 | 0.34 | 0.42 | 0.05 | 0.07 | 0.09 |
| 2023 | | | | | | | | | 2023 | | | | | | | | |
| 2024 | | | | | | | | | 2024 | | | | | | | | |
| 2025 | | | | | | | | | 2025 | | | | | | | | |
| 2002 | 0 | 75 | 0.14 | 0.27 | 0.40 | 0.00 | 0.02 | 0.05 | 2002 | 4 | 165 | 0.00 | 0.01 | 0.03 | 0.00 | 0.00 | 0.01 |
| 2003 2004 | | 101 92 | 0.00 | 0.02 | 0.04 | 0.00 | 0.00 | 0.01 | 2003 2004 | | 149 177 | 0.00 | 0.01 | 0.03 | 0.00 | 0.00 | 0.01 0.01 |
| 2004 | | 92 86 | 0.02 | 0.07 | 0.11 | 0.00 | 0.01 | 0.01 | 2004 | | 176 | 0.00 | 0.01 | 0.02 | 0.00 | 0.00 | 0.01 |
| 2005 | | 79 | 0.02 | 0.11 | 0.22 | 0.00 | 0.02 | 0.04 | 2005 | | 165 | 0.05 | 0.10 | 0.15 | 0.02 | 0.03 | 0.05 |
| 2007 | | 44 | 0.13 | 0.38 | 0.68 | 0.01 | 0.04 | 0.07 | 2007 | | 133 | 0.01 | 0.05 | 0.09 | 0.00 | 0.02 | 0.03 |
| 2008 | | 90 | 0.01 | 0.07 | 0.15 | 0.00 | 0.00 | 0.00 | 2008 | | 180 | 0.03 | 0.09 | 0.16 | 0.01 | 0.04 | 0.06 |
| 2009 2010 | | 90 90 | 0.13 | 0.29 | 0.48 | 0.00 | 0.03 | 0.06 | 2009 2010 | | 135 179 | 0.00 | 0.02 | 0.04 0.09 | 0.00 | 0.00 | 0.01 0.03 |
| 2010 | | 90 89 | 0.25 | 0.45 | 0.67 | 0.03 | 0.06 | 0.10 | 2010 | | 179 | 0.02 | 0.06 | 0.09 | 0.01 | 0.02 | 0.03 |
| 2011 | | 84 | 0.01 | 0.07 | 0.13 | 0.00 | 0.02 | 0.04 | 2011 | | 174 | 0.02 | 0.04 | 0.05 | 0.01 | 0.01 | 0.04 |
| 2013 | | 89 | 0.00 | 0.01 | 0.02 | 0.00 | 0.00 | 0.00 | 2013 | | 179 | 0.03 | 0.04 | 0.06 | 0.01 | 0.01 | 0.02 |
| 2014 | | 90 | 0.00 | 0.03 | 0.08 | 0.00 | 0.00 | 0.00 | 2014 | | 180 | 0.01 | 0.03 | 0.04 | 0.00 | 0.01 | 0.01 |
| 2015 | | 90 | 0.04 | 0.16 | 0.29 | 0.00 | 0.02 | 0.04 | 2015 | | 180 | 0.02 | 0.04 | 0.06 | 0.01 | 0.01 | 0.02 |
| 2016 2017 | | 90 90 | 0.20 | 0.38 | 0.59 0.18 | 0.03 | 0.07 | 0.11 0.03 | 2016 2017 | | 180 180 | 0.05 | 0.08 | 0.10 | 0.01 | 0.02 | 0.02 |
| 2018 | | 90 | 0.12 | 0.23 | 0.36 | 0.00 | 0.02 | 0.04 | 2018 | | 180 | 0.04 | 0.06 | 0.08 | 0.01 | 0.01 | 0.02 |
| 2019 | | 90 | 1.54 | 2.46 | 3.71 | 0.33 | 0.56 | 0.84 | 2019 | | 135 | 0.09 | 0.13 | 0.17 | 0.01 | 0.02 | 0.03 |
| 2020 | | 90 | 3.89 | 5.81 | 8.48 | 0.71 | 1.00 | 1.35 | 2020 | | 135 | 0.17 | 0.24 | 0.30 | 0.03 | 0.05 | 0.06 |
| 2021 | | 90 | 0.59 | 0.92 | 1.31 | 0.11 | 0.20 | 0.30 | 2021 | | 135 | 0.09 | 0.12 | 0.15 | 0.02 | 0.02 | 0.03 |
| 2022 2023 | | | | | | | | | 2022 2023 | | | | | | | | |
| 2023 | | | | | | | | | 2023 | | | | | | | | |
| 2025 | | | | | | | | | 2025 | | | | | | | | |
| 2002 | 1 | 165 | 0.04 | 0.09 | 0.14 | 0.01 | 0.03 | 0.05 | 2002 | 5+ | 165 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 |
| 2003 | | 149 | 0.04 | 0.09 | 0.15 | 0.01 | 0.03 | 0.06 | 2003 | | 149 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 |
| 2004 | | 177 | 0.00 | 0.02 | 0.03 | 0.00 | 0.00 | 0.01 | 2004 | | 177 | 0.00 | 0.01 | 0.02 | 0.00 | 0.00 | 0.01 |
| 2005 2006 | | 176 165 | 0.03 | 0.09 | 0.14 | 0.01 | 0.03 | 0.05 0.03 | 2005 2006 | | 176 165 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2000 | | 133 | 0.01 | 0.10 | 0.11 | 0.00 | 0.02 | 0.03 | 2000 | | 133 | 0.00 | 0.01 | 0.03 | 0.00 | 0.00 | 0.01 |
| 2008 | | 180 | 0.08 | 0.17 | 0.28 | 0.02 | 0.06 | 0.10 | 2008 | | 180 | 0.00 | 0.02 | 0.05 | 0.00 | 0.01 | 0.02 |
| 2009 | | 135 | 0.06 | 0.13 | 0.22 | 0.01 | 0.03 | 0.05 | 2009 | | 135 | 0.00 | 0.01 | 0.02 | 0.00 | 0.00 | 0.01 |
| 2010 | | 179 | 0.21 | 0.33 | 0.46 | 0.06 | 0.11 | 0.15 | 2010 | | 179 | 0.00 | 0.01 | 0.02 | 0.00 | 0.00 | 0.00 |
| 2011 2012 | | 179 174 | 0.14 0.09 | 0.25 | 0.38 | 0.04 | 0.08 | 0.13 0.06 | 2011 2012 | | 179 174 | 0.01 | 0.03 | 0.06 | 0.00 | 0.01 | 0.01 0.01 |
| 2012 | | 179 | 0.10 | 0.17 | 0.25 | 0.02 | 0.05 | 0.07 | 2012 | | 179 | 0.01 | 0.01 | 0.02 | 0.00 | 0.01 | 0.01 |
| 2014 | | 180 | 0.03 | 0.07 | 0.12 | 0.01 | 0.02 | 0.03 | 2014 | | 180 | 0.00 | 0.01 | 0.02 | 0.00 | 0.00 | 0.00 |
| 2015 | | 180 | 0.04 | 0.09 | 0.15 | 0.01 | 0.02 | 0.03 | 2015 | | 180 | 0.01 | 0.02 | 0.03 | 0.00 | 0.00 | 0.01 |
| 2016 | | 180 | 0.35 | 0.51 | 0.69 | 0.08 | 0.13 | 0.18 | 2016 | | 180 | 0.02 | 0.03 | 0.03 | 0.00 | 0.01 | 0.01 |
| 2017 2018 | | 180 180 | 0.22 | 0.33 | 0.46 | 0.05 | 0.09 | 0.12 | 2017 2018 | | 180 180 | 0.02 | 0.03 | 0.04 | 0.00 | 0.01 | 0.01 0.01 |
| 2018 | | 135 | 1.54 | 2.23 | 3.11 | 0.39 | 0.56 | 0.08 | 2018 | | 135 | 0.01 | 0.02 | 0.03 | 0.00 | 0.01 | 0.01 |
| 2020 | | 135 | 3.47 | 4.48 | 5.72 | 0.66 | 0.85 | 1.06 | 2020 | | 135 | 0.07 | 0.10 | 0.13 | 0.01 | 0.02 | 0.02 |
| 2021 | | 135 | 1.29 | 1.70 | 2.19 | 0.29 | 0.39 | 0.48 | 2021 | | 135 | 0.04 | 0.05 | 0.06 | 0.01 | 0.01 | 0.01 |
| 2022 | | | | | | | | | 2022 2023 | | | | | | | | |
| 2023 2024 | | | | | | | | | 2023 | | | | | | | | |
| 2024 | | | | | | | | | 2024 | | | | | | | | |
| 2002 | 2 | 165 | 0.00 | 0.01 | 0.02 | 0.00 | 0.00 | 0.01 | | | | | | | | | |
| 2003 | | 149 | 0.04 | 0.09 | 0.14 | 0.01 | 0.04 | 0.06 | | | | | | | | | |
| 2004 | | 177 | 0.02 | 0.07 | 0.11 | 0.01 | 0.03 | 0.05 | | | | | | | | | [|
| 2005 | | 176 | 0.02 | 0.05 | 0.08 | 0.01 | 0.01 | 0.02 | | | | | | | | | |
| 2006 2007 | | 165 133 | 0.09 | 0.17 | 0.25 | 0.03 | 0.05 | 0.08 0.05 | | | | | | | | | |
| 2007 | | 133 | 0.04 | 0.10 | 0.16 | 0.01 | 0.03 | 0.05 | | | | | | | | | |
| 2009 | | 135 | 0.23 | 0.36 | 0.51 | 0.07 | 0.11 | 0.15 | | | | | | | | | |
| 2010 | | 179 | 0.19 | 0.32 | 0.45 | 0.06 | 0.11 | 0.16 | | | | | | | | | |
| 2011 | | 179 | 0.14 | 0.23 | 0.34 | 0.04 | 0.07 | 0.11 | | | | | | | | | |
| 2012 2013 | | 174 179 | 0.12 | 0.18 | 0.25 | 0.03 | 0.05 | 0.06 0.08 | | | | | | | | | |
| 2013 | | 179 | 0.12 | 0.19 | 0.27 | 0.03 | 0.08 | 0.08 | | | | | | | | | |
| 2014 | | 180 | 0.06 | 0.11 | 0.16 | 0.01 | 0.03 | 0.05 | | | | | | | | | |
| 2016 | | 180 | 0.25 | 0.36 | 0.47 | 0.06 | 0.09 | 0.12 | | | | | | | | | |
| 2017 | | 180 | 0.21 | 0.32 | 0.45 | 0.05 | 0.09 | 0.12 | | | | | | | | | |
| 2018 | | 180 | 0.17 | 0.25 | 0.33 | 0.04 | 0.07 | 0.09 | | | | | | | | | |
| 2019 2020 | | 135 135 | 0.69 1.44 | 0.97 1.84 | 1.29 2.31 | 0.15 0.28 | 0.22 | 0.29 0.46 | | | | | | | | | |
| 2020 | | 135 | 0.73 | 1.84 0.93 | 2.31 1.16 | 0.28 | 0.37 | 0.46 | | | | | | | | | |
| 2021 | | | | | | | | | | | | | | | | | |
| 2023 | | | | | | | | | | | | | | | | | |
| 2024 | | | | | | | | | | | | | | | | | |
| 2025 | | | | | | | | | | | | | | | | | |







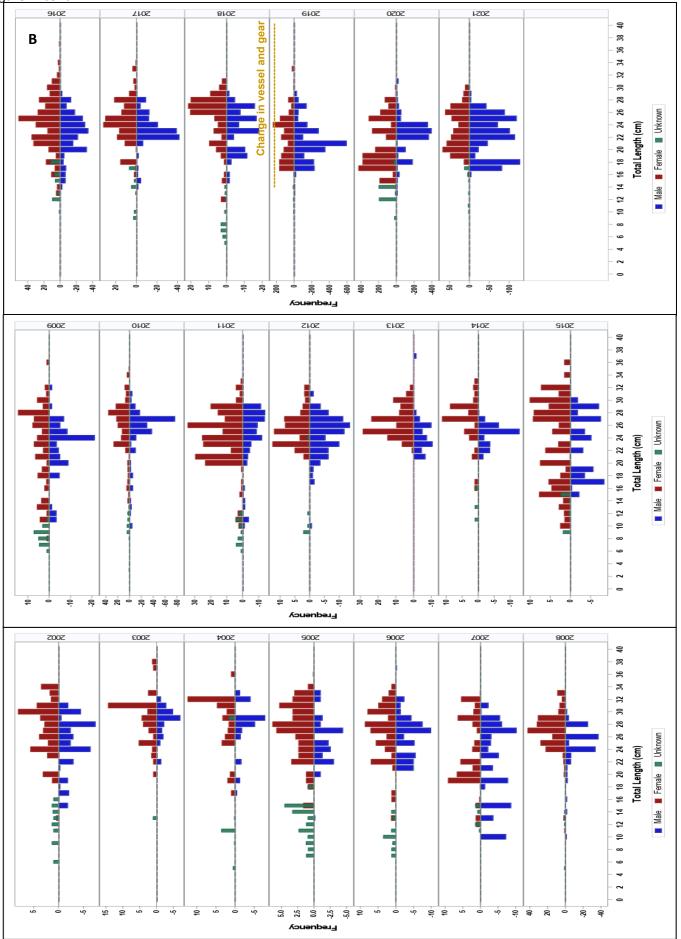
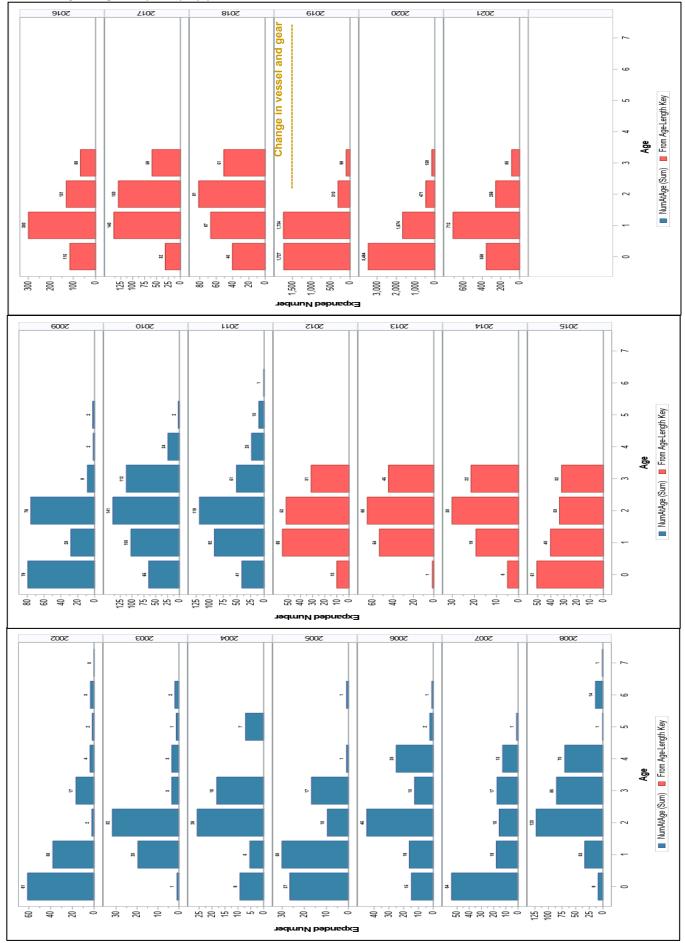


Figure 25. Kingfish age-frequency by year, 2002-2021.



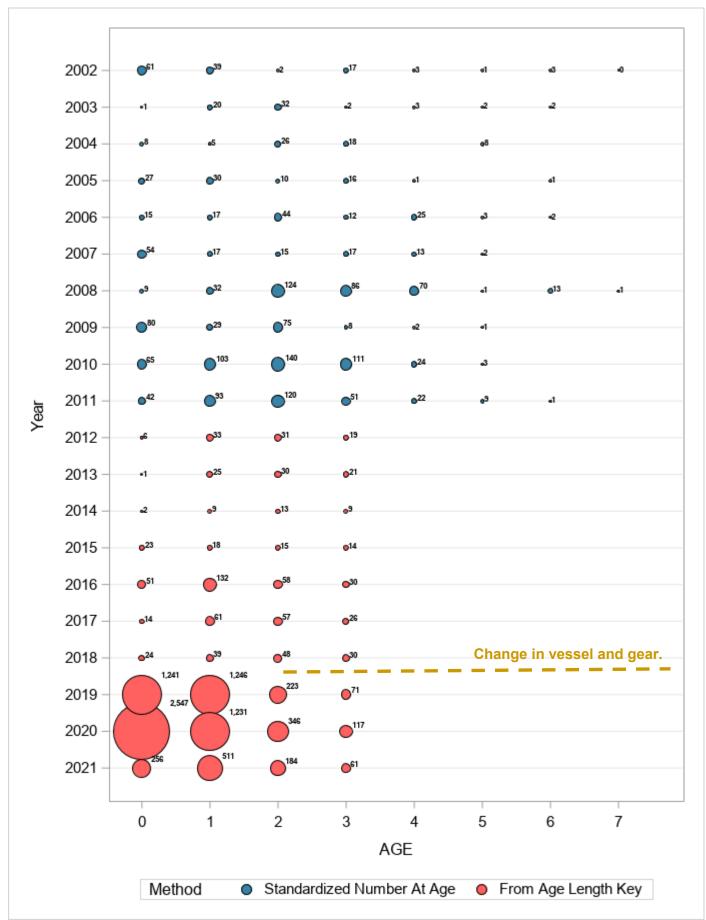


Figure 26. Kingfish age-frequency by year, 2002-2021 standardized to 8,000 (4,800 for 2019 and forward) annual trawl minutes.

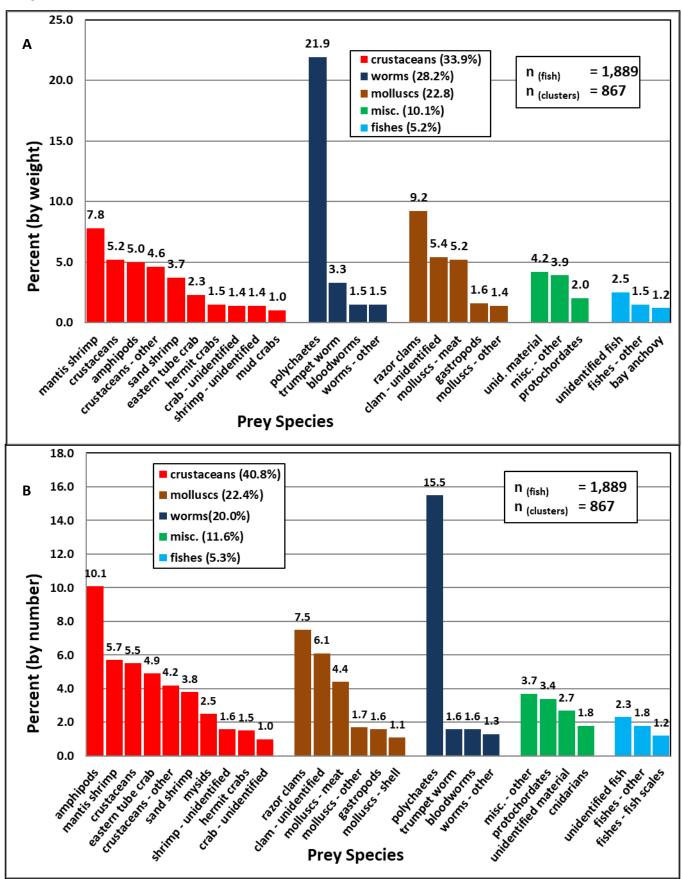
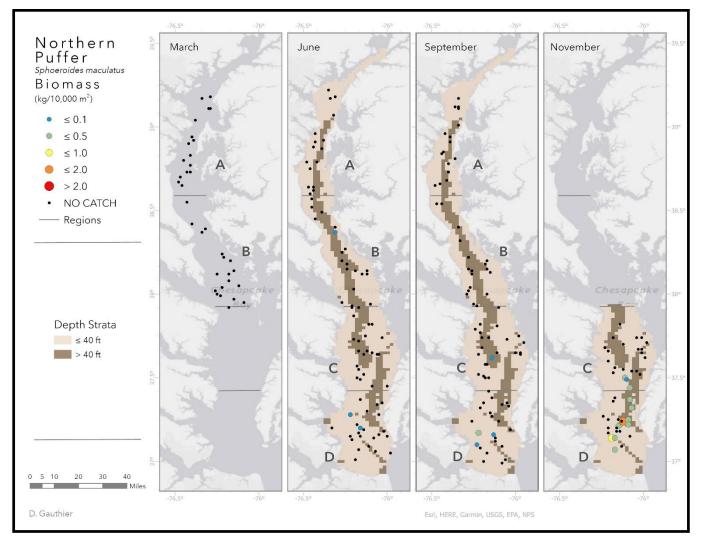


Figure 27. Diet composition, expressed as percent by weight (A) and percent by number (B) of Kingfish collected during ChesMMAP cruises in 2002-2020 combined.

| Maaa | Number | Biomass | Presence at Index Stations | Number | Age | A see Deed | Stomach | Stomachs |
|------|--------|-------------|-------------------------------|----------|-----------|------------|-----------|----------|
| Year | Caught | Caught (kg) | (%) | Measured | Specimens | Ages Read | Specimens | Analyzed |
| 2002 | 231 | 23.9 | 48.0 | 231 | 177 | 0 | 171 | 156 |
| 2003 | 225 | 32.5 | 36.3 | 225 | 100 | 0 | 92 | 91 |
| 2004 | 41 | 6.9 | 9.7 | 41 | 31 | 0 | 27 | 26 |
| 2005 | 131 | 13.7 | 25.6 | 131 | 84 | 0 | 84 | 83 |
| 2006 | 52 | 5.5 | 17.7 | 52 | 51 | 0 | 48 | 47 |
| 2007 | 155 | 19.8 | 75.0 | 155 | 127 | 0 | 124 | 124 |
| 2008 | 90 | 6.9 | 21.1 | 90 | 78 | 0 | 77 | 77 |
| 2009 | 76 | 7.2 | 24.4 | 76 | 69 | 0 | 68 | 67 |
| 2010 | 326 | 54.7 | 44.4 | 326 | 176 | 0 | 157 | 156 |
| 2011 | 614 | 55.0 | 50.6 | 614 | 247 | 0 | 238 | 236 |
| 2012 | 50 | 5.3 | 11.9 | 50 | 50 | 0 | 41 | 40 |
| 2013 | 63 | 4.2 | 15.7 | 63 | 61 | 0 | 55 | 52 |
| 2014 | 49 | 3.6 | 12.2 | 49 | 39 | 0 | 16 | 16 |
| 2015 | 290 | 44.1 | 36.7 | 290 | 157 | 0 | 54 | 54 |
| 2016 | 519 | 65.6 | 40.0 | 519 | 231 | 0 | 99 | 97 |
| 2017 | 231 | 22.4 | 25.6 | 231 | 148 | 0 | 116 | 116 |
| 2018 | 246 | 24.5 | 28.9 | 246 | 128 | 0 | 87 | 87 |
| 2019 | 143 | 13.6 | 22.2 | 143 | 99 | 0 | 77 | 75 |
| 2020 | 80 | 7.0 | 35.6 | 80 | 54 | 0 | 23 | 23 |
| 2021 | 43 | 5.0 | 18.9 | 43 | 34 | 0 | 23 | 23 |

Table 16. Northern Puffer sampling rates and preserved specimen analysis status by year (shaded rows represent a change in vessel and trawl gear beginning in June 2019).

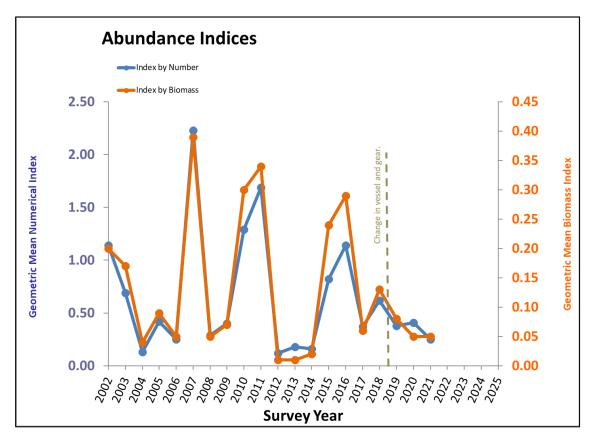
Figure 28. Station specific biomass of Northern Puffer in Chesapeake Bay, 2021.

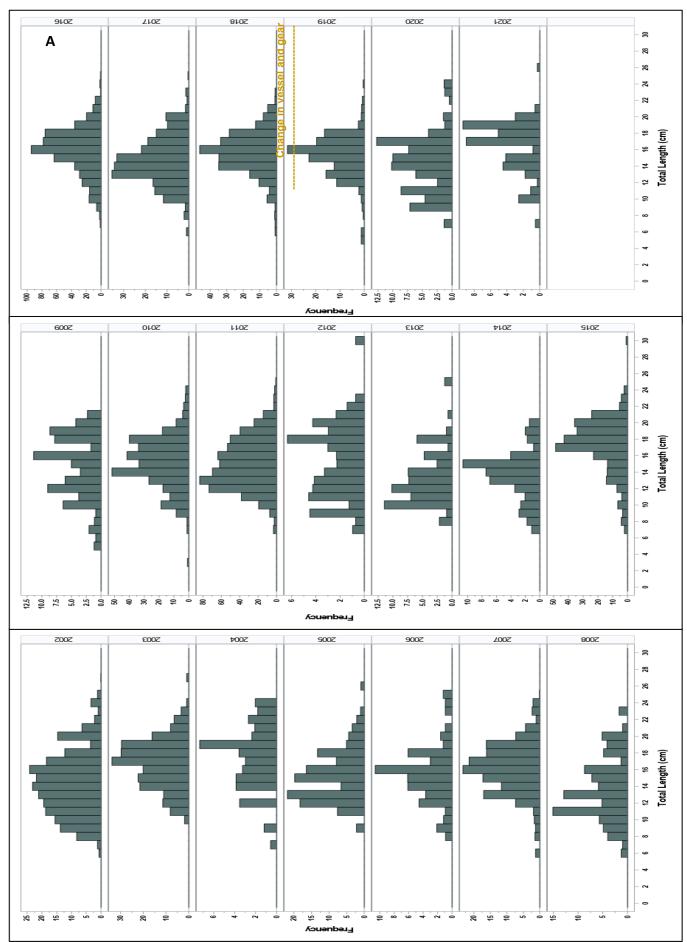


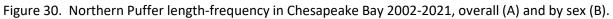
| Year | Age | n | Nur | merical Ind | ex | Bi | omass Inde | ≥x |
|-------|-------|-----|------|-------------|------|------|------------|------|
| . cu. | , .gc | | LCI | Index | UCI | LCI | Index | UCI |
| 2002 | All | 75 | 0.76 | 1.14 | 1.60 | 0.13 | 0.20 | 0.27 |
| 2003 | | 101 | 0.45 | 0.69 | 0.97 | 0.10 | 0.17 | 0.25 |
| 2004 | | 92 | 0.04 | 0.13 | 0.22 | 0.01 | 0.04 | 0.07 |
| 2005 | | 86 | 0.22 | 0.42 | 0.66 | 0.04 | 0.09 | 0.15 |
| 2006 | | 79 | 0.11 | 0.25 | 0.41 | 0.02 | 0.05 | 0.08 |
| 2007 | | 44 | 1.54 | 2.23 | 3.11 | 0.26 | 0.39 | 0.52 |
| 2008 | | 90 | 0.15 | 0.29 | 0.45 | 0.02 | 0.05 | 0.08 |
| 2009 | | 90 | 0.22 | 0.40 | 0.61 | 0.04 | 0.07 | 0.11 |
| 2010 | | 90 | 0.82 | 1.29 | 1.87 | 0.16 | 0.30 | 0.46 |
| 2011 | | 89 | 1.17 | 1.69 | 2.33 | 0.20 | 0.34 | 0.48 |
| 2012 | | 84 | 0.04 | 0.12 | 0.20 | 0.00 | 0.01 | 0.02 |
| 2013 | | 89 | 0.08 | 0.18 | 0.30 | 0.00 | 0.01 | 0.02 |
| 2014 | | 90 | 0.05 | 0.16 | 0.29 | 0.01 | 0.02 | 0.04 |
| 2015 | | 90 | 0.49 | 0.82 | 1.22 | 0.12 | 0.24 | 0.37 |
| 2016 | | 90 | 0.66 | 1.14 | 1.76 | 0.15 | 0.29 | 0.45 |
| 2017 | | 90 | 0.21 | 0.37 | 0.57 | 0.03 | 0.06 | 0.10 |
| 2018 | | 90 | 0.34 | 0.62 | 0.96 | 0.06 | 0.13 | 0.22 |
| 2019 | | 90 | 0.20 | 0.38 | 0.60 | 0.03 | 0.08 | 0.13 |
| 2020 | | 90 | 0.26 | 0.41 | 0.57 | 0.03 | 0.05 | 0.08 |
| 2021 | | 90 | 0.13 | 0.25 | 0.38 | 0.02 | 0.05 | 0.08 |
| 2022 | | | | | | | | |
| 2023 | | | | | | | | |
| 2024 | | | | | | | | |
| 2025 | | | | | | | | |

Table 17. Northern Puffer geometric mean indices of abundance, by number and biomass, overall.

Figure 29. Northern Puffer geometric mean indices of abundance, by number and biomass, for all ages combined.







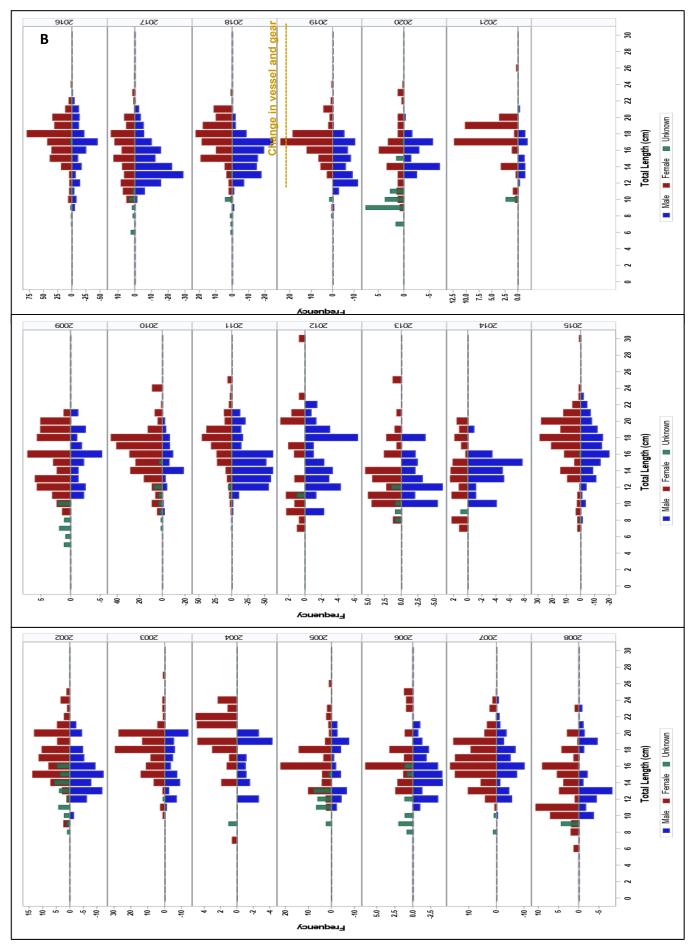
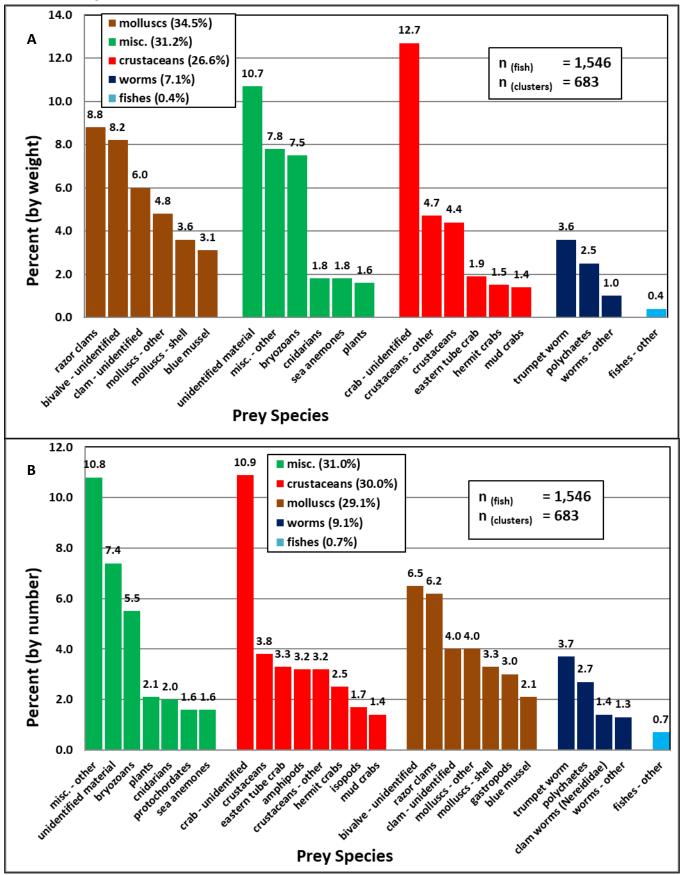


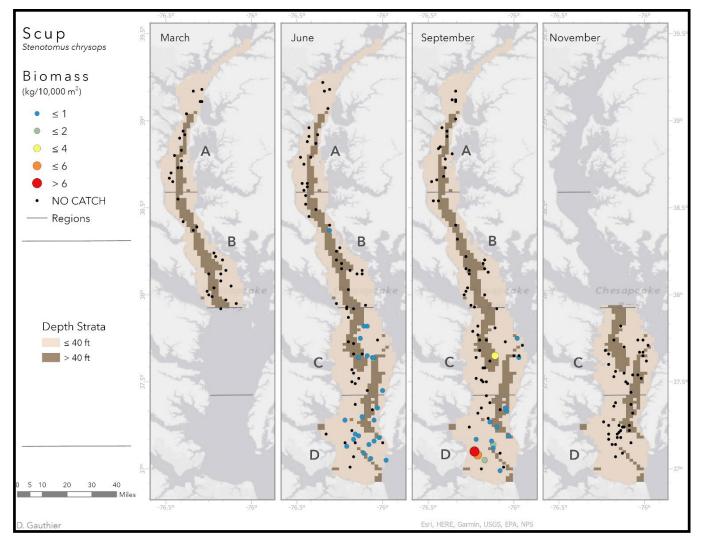
Figure 31. Diet composition, expressed as percent by weight (A) and percent by number (B) of Northern Puffer collected during ChesMMAP cruises in 2002-2020 combined.



| Year | Number Caught | Biomass Caught (kg) | Presence at Index Stations (%) | Number Measured | Age Specimens | Ages Read | Stomach Specimens | Stomachs Analyzed |
|------|------------------|------------------------|-----------------------------------|--------------------|------------------|--------------|----------------------|----------------------|
| 2002 | 107 | 7.8 | 17.9 | 84 | 40 | 40 | 39 | 34 |
| 2003 | 192 | 11.1 | 22.7 | 192 | 100 | 100 | 99 | 90 |
| 2004 | 475 | 26.0 | 42.7 | 475 | 155 | 155 | 150 | 142 |
| 2005 | 674 | 30.6 | 25.0 | 674 | 86 | 86 | 85 | 83 |
| 2006 | 317 | 12.7 | 30.1 | 317 | 115 | 115 | 112 | 111 |
| 2007 | 211 | 6.5 | 44.7 | 211 | 128 | 128 | 121 | 119 |
| 2008 | 56 | 4.1 | 11.1 | 56 | 42 | 0 | 42 | 42 |
| 2009 | 201 | 6.6 | 23.1 | 201 | 97 | 0 | 92 | 91 |
| 2010 | 853 | 29.2 | 29.1 | 653 | 126 | 0 | 125 | 123 |
| 2011 | 72 | 2.7 | 21.6 | 72 | 56 | 0 | 51 | 51 |
| 2012 | 12 | 0.4 | 3.6 | 12 | 12 | 0 | 12 | 12 |
| 2013 | 49 | 1.8 | 8.6 | 49 | 28 | 28 | 25 | 24 |
| 2014 | 63 | 2.6 | 7.7 | 63 | 26 | 26 | 19 | 19 |
| 2015 | 988 | 45.6 | 35.0 | 988 | 186 | 186 | 88 | 87 |
| 2016 | 65 | 2.0 | 9.4 | 65 | 40 | 40 | 20 | 20 |
| 2017 | 25 | 0.4 | 4.3 | 25 | 20 | 20 | 12 | 12 |
| 2018 | 386 | 12.2 | 28.2 | 386 | 94 | 94 | 58 | 58 |
| 2019 | 1,126 | 35.1 | 39.3 | 883 | 196 | 196 | 135 | 135 |
| 2020 | 626 | 18.7 | 22.2 | 626 | 34 | 34 | 23 | 23 |
| 2021 | 1,135 | 45.7 | 28.1 | 1,135 | 112 | 112 | 59 | 59 |

Table 18. Scup sampling rates and preserved specimen analysis status by year (shaded rows represent a change in vessel and trawl gear beginning in June 2019).

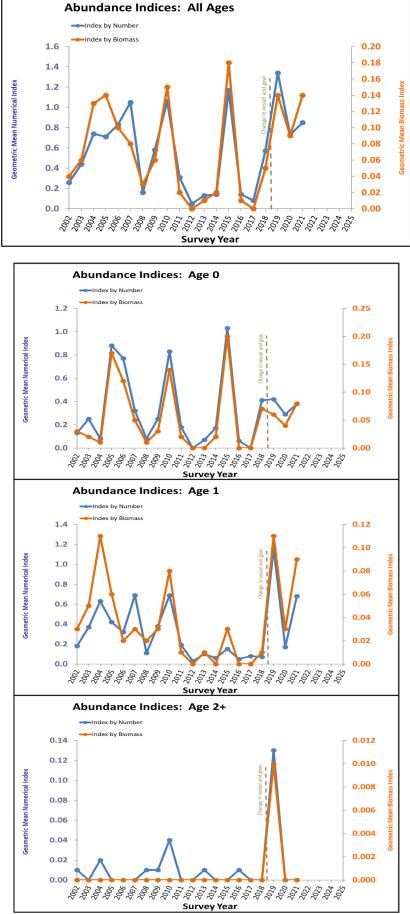
Figure 32. Station specific biomass of Scup in Chesapeake Bay, 2021.

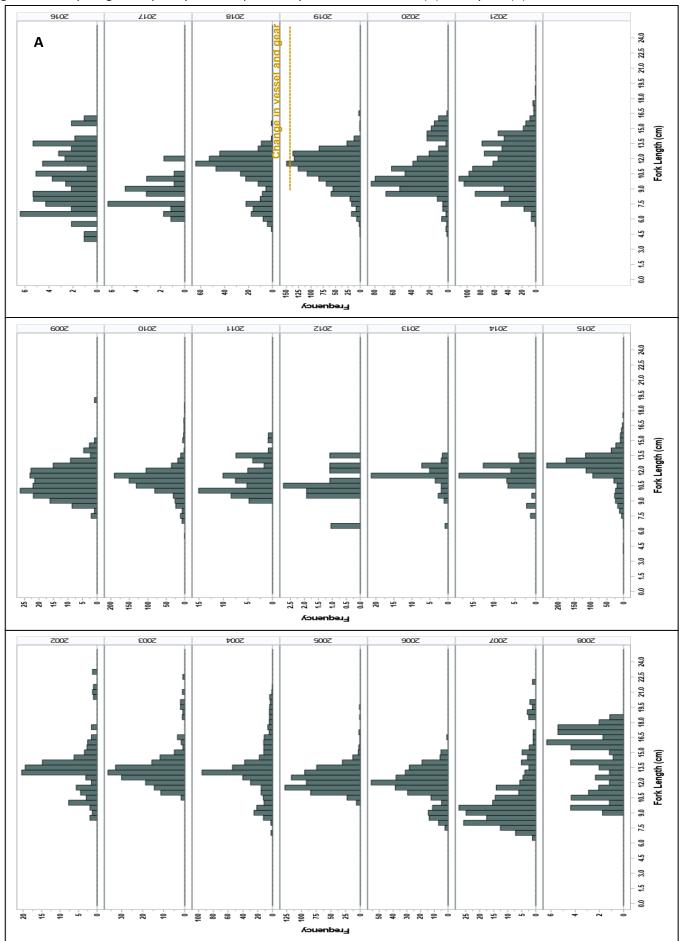


| Year | Age | n | Nur | nerical In | dex | Bio | mass Inc | dex | Year | Age | n | Nur | merical In | dex | Bio | mass Ind | dex |
|--------------|-----|-----|--------------|------------|------|------|----------|------|------|-----|-----|------|------------|------|------|----------|------|
| | 8. | | LCI | Index | UCI | LCI | Index | UCI | | | | LCI | Index | UCI | LCI | Index | UCI |
| 2002 | All | 106 | 0.12 | 0.26 | 0.41 | 0.01 | 0.04 | 0.08 | 2002 | 1 | 106 | 0.08 | 0.18 | 0.30 | 0.00 | 0.03 | 0.05 |
| 2003 | , | 127 | 0.26 | 0.44 | 0.65 | 0.03 | 0.06 | 0.09 | 2003 | - | 127 | 0.22 | 0.37 | 0.55 | 0.03 | 0.05 | 0.08 |
| 2003 | | 109 | 0.46 | 0.74 | 1.08 | 0.06 | 0.13 | 0.20 | 2003 | | 109 | 0.39 | 0.63 | 0.91 | 0.05 | 0.11 | 0.17 |
| 2004 | | 112 | 0.37 | 0.74 | 1.14 | 0.06 | 0.14 | 0.23 | 2004 | | 112 | 0.35 | 0.42 | 0.66 | 0.03 | 0.06 | 0.10 |
| 2005 | | 103 | 0.48 | 0.83 | 1.25 | 0.05 | 0.10 | 0.15 | 2005 | | 103 | 0.17 | 0.32 | 0.49 | 0.03 | 0.02 | 0.02 |
| 2000 | | 76 | 0.40 | 1.05 | 1.56 | 0.03 | 0.08 | 0.11 | 2000 | | 76 | 0.42 | 0.69 | 1.01 | 0.01 | 0.02 | 0.02 |
| 2007 | | 117 | 0.04 | 0.16 | 0.27 | 0.00 | 0.03 | 0.05 | 2007 | | 117 | 0.42 | 0.11 | 0.20 | 0.02 | 0.02 | 0.03 |
| 2008 | | 117 | 0.33 | 0.58 | 0.27 | 0.00 | 0.05 | 0.09 | 2008 | | 117 | 0.22 | 0.38 | 0.20 | 0.00 | 0.02 | 0.04 |
| 2009 | | 117 | 0.63 | 1.06 | 1.60 | 0.03 | 0.15 | 0.03 | 2003 | | 117 | 0.22 | 0.69 | 1.03 | 0.01 | 0.03 | 0.03 |
| 2010 | | 117 | 0.03 | 0.31 | 0.46 | 0.07 | 0.02 | 0.24 | 2010 | | 116 | 0.40 | 0.19 | 0.28 | 0.04 | 0.03 | 0.13 |
| 2011 | | 111 | 0.00 | 0.05 | 0.40 | 0.01 | 0.02 | 0.04 | 2011 | | 111 | 0.10 | 0.03 | 0.28 | 0.00 | 0.01 | 0.02 |
| 2012 | | 111 | 0.00 | 0.03 | 0.10 | 0.00 | 0.00 | 0.01 | 2012 | | 111 | 0.00 | 0.03 | 0.00 | 0.00 | 0.00 | 0.00 |
| | | 117 | 0.04 | 0.13 | 0.24 | 0.00 | 0.01 | 0.03 | 2013 | | 117 | 0.02 | | 0.19 | 0.00 | 0.01 | 0.02 |
| 2014 | | 117 | | 1.17 | 1.77 | 0.00 | | | | | 117 | 0.01 | 0.06 | 0.11 | 0.00 | 0.00 | 0.01 |
| 2015 | | 117 | 0.71 0.04 | 0.14 | 0.24 | 0.09 | 0.18 | 0.28 | 2015 | | 117 | 0.05 | 0.15 | 0.27 | 0.00 | 0.00 | 0.05 |
| 2016 | | | | | | | | | 2016 | | | | | | | | |
| 2017 | | 117 | 0.01 | 0.08 | 0.15 | 0.00 | 0.00 | 0.01 | 2017 | | 117 | 0.01 | 0.08 | 0.15 | 0.00 | 0.00 | 0.01 |
| 2018 | | 117 | 0.32 | 0.57 | 0.88 | 0.01 | 0.05 | 0.10 | 2018 | | 117 | 0.00 | 0.07 | 0.15 | 0.00 | 0.01 | 0.02 |
| 2019 | | 105 | 0.82 | 1.34 | 1.99 | 0.07 | 0.14 | 0.23 | 2019 | | 105 | 0.72 | 1.16 | 1.70 | 0.05 | 0.11 | 0.18 |
| 2020 | | 105 | 0.41 | 0.73 | 1.13 | 0.04 | 0.09 | 0.15 | 2020 | | 105 | 0.06 | 0.17 | 0.30 | 0.00 | 0.03 | 0.06 |
| 2021 | | 105 | 0.46 | 0.85 | 1.35 | 0.05 | 0.14 | 0.23 | 2021 | | 105 | 0.37 | 0.68 | 1.07 | 0.03 | 0.09 | 0.17 |
| 2022 | | | | | | | | | 2022 | | | | | | | | |
| 2023 | | | | | | | | | 2023 | | | | | | | | |
| 2024 | | | | | | | | | 2024 | | | | | | | | |
| 2025 | | | | | | | | | 2025 | | | | | | | | |
| 2002 | 0 | 65 | 0.00 | 0.13 | 0.29 | 0.00 | 0.03 | 0.09 | 2002 | 2+ | 106 | 0.00 | 0.01 | 0.03 | 0.00 | 0.00 | 0.00 |
| 2003 | | 85 | 0.12 | 0.25 | 0.39 | 0.01 | 0.02 | 0.04 | 2003 | | 127 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2004 | | 80 | 0.00 | 0.08 | 0.17 | 0.00 | 0.01 | 0.03 | 2004 | | 109 | 0.00 | 0.02 | 0.04 | 0.00 | 0.00 | 0.00 |
| 2005 | | 73 | 0.40 | 0.88 | 1.52 | 0.07 | 0.17 | 0.29 | 2005 | | 112 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 |
| 2006 | | 68 | 0.34 | 0.77 | 1.33 | 0.05 | 0.12 | 0.20 | 2006 | | 103 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2007 | | 38 | 0.12 | 0.32 | 0.57 | 0.02 | 0.05 | 0.08 | 2007 | | 76 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2008 | | 78 | 0.02 | 0.08 | 0.16 | 0.00 | 0.01 | 0.02 | 2008 | | 117 | 0.00 | 0.01 | 0.02 | 0.00 | 0.00 | 0.00 |
| 2009 | | 78 | 0.09 | 0.25 | 0.43 | 0.00 | 0.03 | 0.06 | 2009 | | 117 | 0.00 | 0.01 | 0.02 | 0.00 | 0.00 | 0.00 |
| 2010 | | 78 | 0.39 | 0.83 | 1.41 | 0.05 | 0.14 | 0.23 | 2010 | | 117 | 0.01 | 0.04 | 0.06 | 0.00 | 0.00 | 0.00 |
| 2011 | | 77 | 0.06 | 0.18 | 0.31 | 0.00 | 0.02 | 0.03 | 2011 | | 116 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2012 | | 72 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2012 | | 111 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2013 | | 77 | 0.00 | 0.07 | 0.14 | 0.00 | 0.00 | 0.01 | 2013 | | 116 | 0.00 | 0.01 | 0.03 | 0.00 | 0.00 | 0.00 |
| 2014 | | 78 | 0.03 | 0.17 | 0.33 | 0.00 | 0.02 | 0.05 | 2014 | | 117 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2015 | | 78 | 0.49 | 1.03 | 1.76 | 0.08 | 0.20 | 0.34 | 2015 | | 117 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2016 | | 78 | 0.00 | 0.06 | 0.12 | 0.00 | 0.00 | 0.01 | 2016 | | 117 | 0.00 | 0.01 | 0.03 | 0.00 | 0.00 | 0.00 |
| 2017 | | 78 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2017 | | 117 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2018 | | 78 | 0.14 | 0.41 | 0.75 | 0.01 | 0.07 | 0.13 | 2018 | | 117 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2019 | | 70 | 0.15 | 0.42 | 0.74 | 0.01 | 0.06 | 0.11 | 2019 | | 105 | 0.04 | 0.13 | 0.23 | 0.00 | 0.01 | 0.03 |
| 2020 | | 70 | 0.09 | 0.29 | 0.52 | 0.01 | 0.04 | 0.08 | 2020 | | 105 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2021 | | 70 | 0.11 | 0.38 | 0.72 | 0.01 | 0.08 | 0.15 | 2021 | | 105 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2022 | | | | | | | | | 2022 | | | | | | | | |
| | | | | | | | | | 2023 | | | | | | | | |
| 2023 | | | | | | | | | | | | | | | | | |
| 2023 2024 | | | | | | | | | 2024 | | | | | | | | |

Table 19. Scup geometric mean indices of abundance, by number and biomass, overall and by age-class.

Figure 33. Scup geometric mean indices of abundance, by number and biomass, for all ages combined and by ageclass. Abundance Indices: All Ages





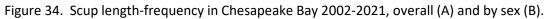


Figure 34. cont.

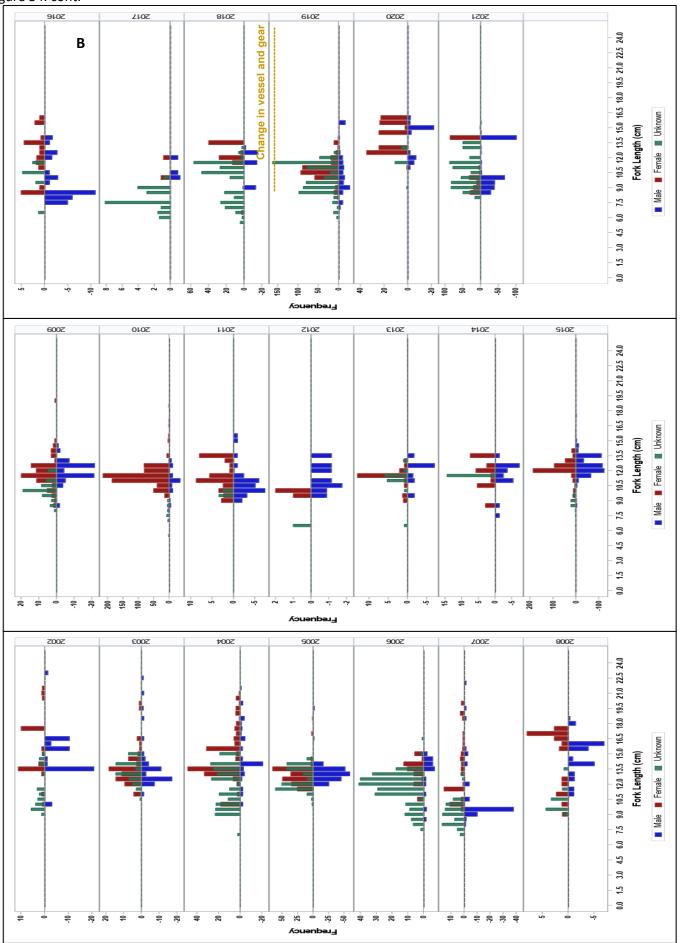
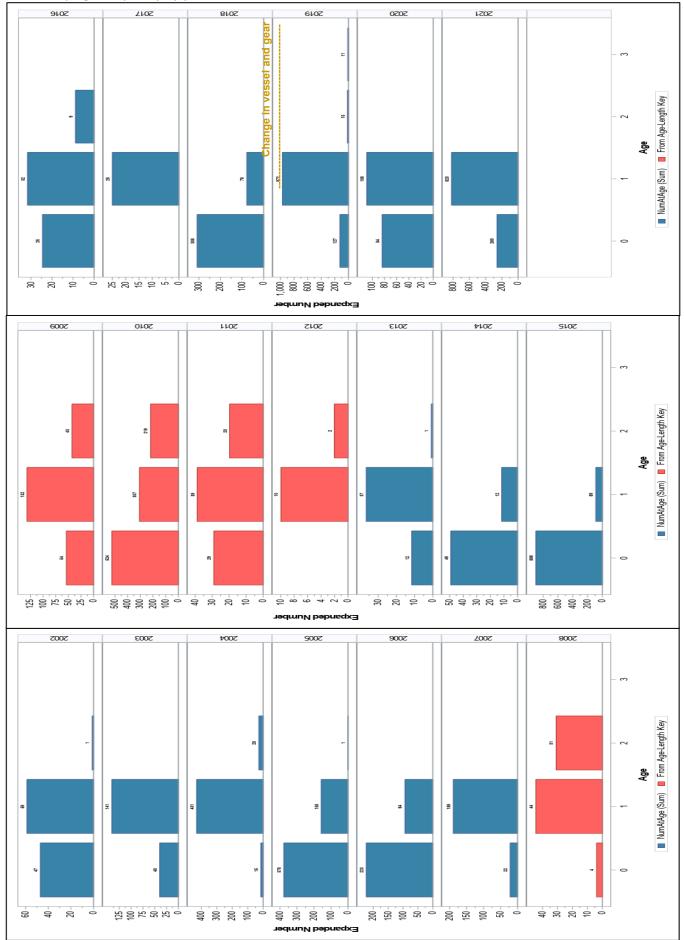


Figure 35. Scup age-frequency by year, 2002-2021.



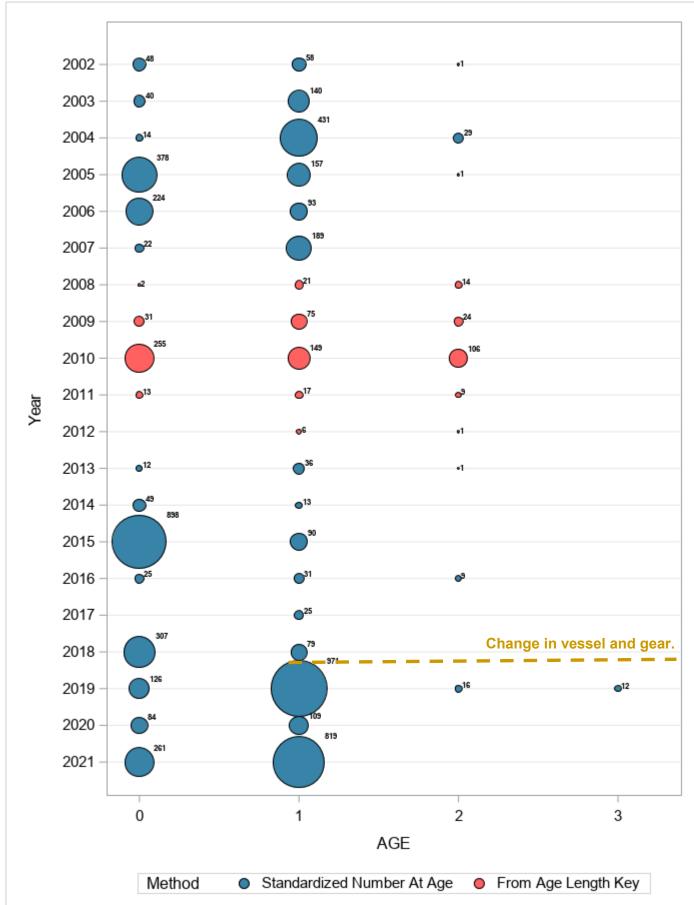


Figure 36. Scup age-frequency by year, 2002-2021 standardized to 8,000 (4,800 for 2019 and forward) annual trawl minutes.

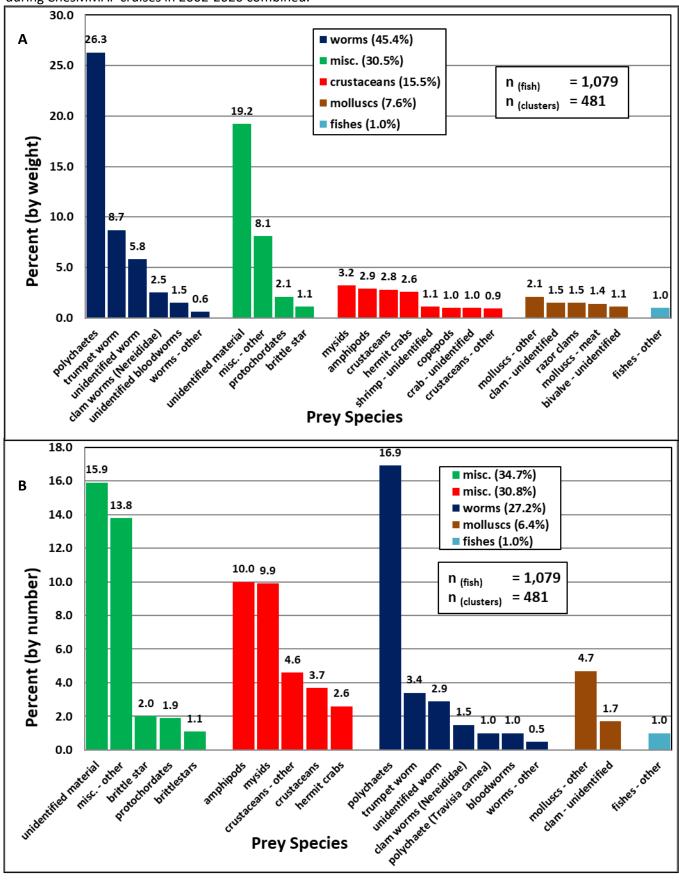
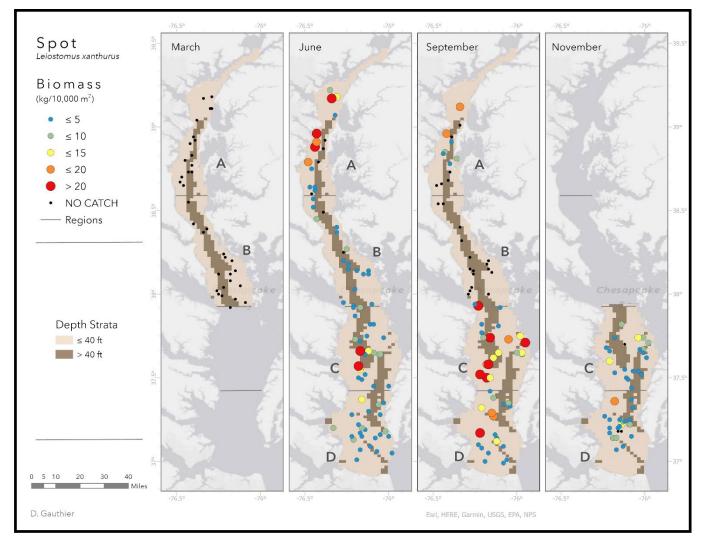


Figure 37. Diet composition, expressed as percent by weight (A) and percent by number (B) of Scup collected during ChesMMAP cruises in 2002-2020 combined.

| Year | Number Caught | Biomass Caught (kg) | Presence at Index Stations (%) | Number Measured | Age Specimens | Ages Read | Stomach Specimens | Stomachs Analyzed |
|------|------------------|------------------------|-----------------------------------|--------------------|------------------|--------------|----------------------|----------------------|
| 2002 | 3,122 | 443.2 | 49.8 | 3,034 | 672 | 672 | 647 | 19 |
| 2003 | 4,081 | 568.8 | 51.7 | 3,102 | 414 | 395 | 396 | 4 |
| 2004 | 4,131 | 419.6 | 64.4 | 4,089 | 619 | 619 | 578 | 18 |
| 2005 | 11,561 | 1,011.2 | 73.2 | 10,690 | 1,030 | 1,030 | 979 | 3 |
| 2006 | 7,080 | 700.4 | 71.0 | 6,439 | 680 | 656 | 632 | 7 |
| 2007 | 5,729 | 462.8 | 72.3 | 5 <i>,</i> 396 | 626 | 626 | 602 | 4 |
| 2008 | 6,256 | 417.5 | 63.3 | 5,197 | 785 | 785 | 742 | 735 |
| 2009 | 5,191 | 682.6 | 47.1 | 3,481 | 465 | 449 | 447 | 442 |
| 2010 | 6,744 | 255.3 | 67.2 | 6,336 | 687 | 687 | 652 | 623 |
| 2011 | 2,867 | 278.0 | 39.0 | 2,867 | 352 | 352 | 320 | 315 |
| 2012 | 2,161 | 114.5 | 35.9 | 1,758 | 345 | 345 | 259 | 253 |
| 2013 | 4,087 | 316.0 | 44.4 | 3 <i>,</i> 430 | 428 | 428 | 289 | 280 |
| 2014 | 939 | 117.3 | 23.3 | 939 | 188 | 188 | 89 | 88 |
| 2015 | 401 | 54.0 | 15.4 | 401 | 102 | 102 | 11 | 11 |
| 2016 | 1,059 | 67.2 | 27.1 | 835 | 167 | 167 | 43 | 40 |
| 2017 | 1,586 | 116.4 | 26.8 | 1,586 | 213 | 213 | 105 | 101 |
| 2018 | 1,635 | 77.0 | 32.7 | 1,635 | 204 | 204 | 101 | 99 |
| 2019 | 67,938 | 3,529.2 | 78.4 | 22,694 | 556 | 556 | 229 | 227 |
| 2020 | 132,547 | 6,173.8 | 89.3 | 34,056 | 370 | 370 | 134 | 133 |
| 2021 | 73,427 | 3,428.0 | 84.9 | 21,513 | 686 | 686 | 283 | 276 |

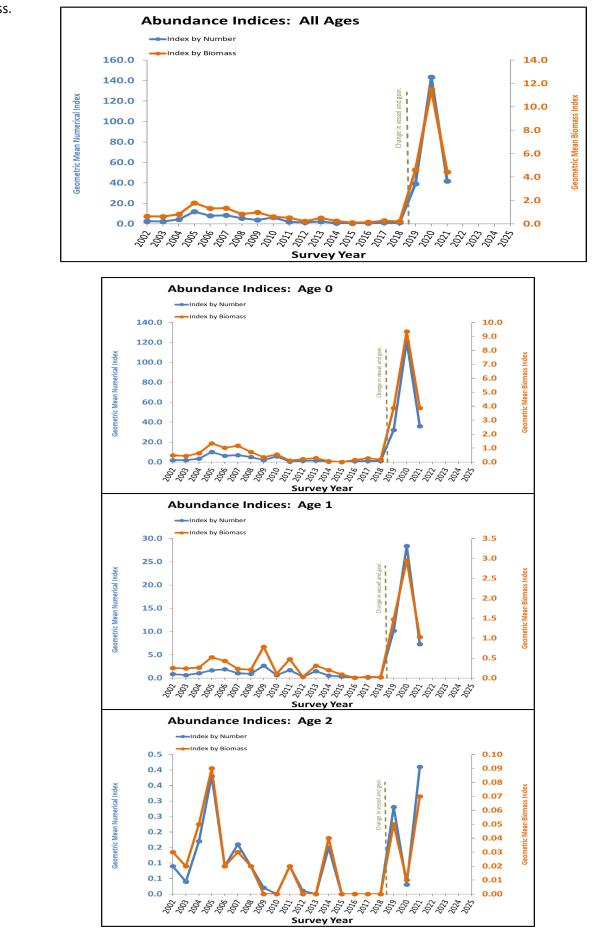
Table 20. Spot sampling rates and preserved specimen analysis status by year (shaded rows represent a change in vessel and trawl gear beginning in June 2019).

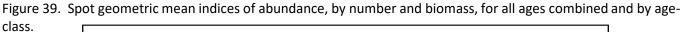
Figure 38. Station specific biomass of Spot in Chesapeake Bay, 2021.



| Year | | | | nerical In | | | mass Inc | | , Year | | | | nerical In | | 1 | mass In | dov |
|--------------|-----|------|------------|------------|------------|------------|----------|------------|--------------|-----|-----|-------|--------------|-------|------|---------|------|
| Teal | Age | | LCI | Index | UCI | LCI | Index | UCI | Teal | Age | | LCI | Index | UCI | LCI | Index | UCI |
| 2002 | All | 227 | 1.9 | 2.5 | 3.3 | 0.5 | 0.6 | 0.8 | 2002 | 1 | 227 | 0.63 | 0.86 | 1.12 | 0.17 | 0.25 | 0.34 |
| 2002 | All | 240 | 1.9 | 2.5 | 3.1 | 0.5 | 0.6 | 0.8 | 2002 | 1 | 240 | 0.03 | 0.80 | 0.85 | 0.17 | 0.25 | 0.34 |
| 2003 | | 2240 | 3.3 | 4.2 | 5.3 | 0.5 | 0.8 | 1.0 | 2003 | | 240 | 0.40 | 1.02 | 1.30 | 0.14 | 0.24 | 0.30 |
| 2004 | | 235 | 9.3 | 12.0 | 15.4 | 1.5 | 1.8 | 2.2 | 2004 | | 235 | 1.24 | 1.67 | 2.18 | 0.19 | 0.20 | 0.55 |
| 2005 | | 235 | 9.5 5.9 | 7.8 | 10.3 | | 1.8 | 1.6 | 2005 | | 235 | 1.24 | 1.87 | 2.18 | 0.38 | 0.52 | 0.67 |
| 2008 | | 155 | 6.1 | 8.4 | 10.5 | 1.1 1.0 | 1.3 | 1.0 | 2008 | | 155 | 0.73 | 1.00 | 1.40 | 0.52 | 0.43 | 0.35 |
| | | 240 | 4.3 | 5.6 | 7.2 | 0.7 | 0.8 | 1.7 | 2007 | | 240 | 0.73 | | 1.40 | 0.10 | 0.23 | 0.31 |
| 2008 2009 | | 240 | 4.5 2.9 | 3.9 | 5.1 | 0.7 | 1.0 | 1.0 | 2008 | | 240 | 1.95 | 0.87 2.66 | 3.55 | 0.14 | 0.21 | 1.03 |
| 2009 | | 204 | 4.6 | 6.2 | 8.2 | 0.7 | 0.6 | 0.8 | 2009 | | 204 | 0.45 | 0.60 | 0.78 | 0.08 | 0.78 | 0.14 |
| 2010 | | 204 | 1.4 | 1.9 | 2.5 | 0.5 | 0.5 | 0.8 | 2010 | | 204 | 1.26 | 1.68 | 2.19 | 0.08 | 0.11 | 0.14 |
| | | 230 | 0.9 | | 2.5 1.8 | 0.4 | 0.5 | | | | 230 | 0.16 | | | 0.35 | | |
| 2012 | | 231 | 1.7 | 1.3 | 3.0 | 0.2 | 0.2 | 0.3 0.6 | 2012 2013 | | 231 | 1.08 | 0.24 | 0.32 | 0.02 | 0.03 | 0.04 |
| 2013 | | | | 2.3 | | | | | | | | | 1.46 | 1.90 | | 0.31 | |
| 2014 | | 240 | 0.5 | 0.7 | 1.0 | 0.2 | 0.3 | 0.4 | 2014 | | 240 | 0.35 | 0.54 | 0.75 | 0.12 | 0.20 | 0.29 |
| 2015 | | 240 | 0.2 | 0.3 | 0.5 | 0.0 | 0.1 | 0.2 | 2015 | | 240 | 0.17 | 0.29 | 0.43 | 0.04 | 0.09 | 0.15 |
| 2016 | | 240 | 0.6 | 0.8 | 1.0 | 0.1 | 0.2 | 0.2 | 2016 | | 240 | 0.02 | 0.05 | 0.07 | 0.00 | 0.01 | 0.01 |
| 2017 | | 239 | 0.8 | 1.1 | 1.5 | 0.2 | 0.3 | 0.4 | 2017 | | 239 | 0.07 | 0.12 | 0.19 | 0.01 | 0.03 | 0.05 |
| 2018 | | 217 | 0.8 | 1.1 | 1.5 | 0.1 | 0.2 | 0.3 | 2018 | | 217 | 0.13 | 0.19 | 0.25 | 0.01 | 0.02 | 0.03 |
| 2019 | | 204 | 26.9 | 39.1 | 56.5 | 3.5 | 4.6 | 6.0 | 2019 | | 204 | 7.52 | 10.19 | 13.70 | 1.15 | 1.47 | 1.85 |
| 2020 | | 205 | 103.4 | 143.3 | 198.4 | 9.0 | 11.5 | 14.6 | 2020 | | 205 | 21.27 | 28.36 | 37.70 | 2.35 | 2.95 | 3.67 |
| 2021 | | 205 | 30.7 | 41.8 | 56.7 | 3.5 | 4.4 | 5.6 | 2021 | | 205 | 5.67 | 7.29 | 9.30 | 0.80 | 1.03 | 1.28 |
| 2022 | | | | | | | | | 2022 | | | | | | | | |
| 2023 | | | | | | | | | 2023 | | | | | | | | |
| 2024 | | | | | | | | | 2024 | | | | | | | | |
| 2025 | | 227 | | 1.0 | 2.5 | 0.4 | 0.5 | 0.6 | 2025 | 2 | 450 | 0.02 | 0.00 | 0.44 | 0.00 | 0.00 | 0.00 |
| 2002 | 0 | 227 | 1.4 | 1.9 | 2.5 | 0.4 | 0.5 | 0.6 | 2002 | 2 | 153 | 0.03 | 0.09 | 0.14 | 0.00 | 0.03 | 0.06 |
| 2003 | | 240 | 1.4 | 1.8 | 2.3 | 0.3 | 0.4 | 0.5 | 2003 | | 150 | 0.01 | 0.04 | 0.08 | 0.00 | 0.02 | 0.04 |
| 2004 | | 224 | 2.7 | 3.4 | 4.3 | 0.5 | 0.6 | 0.7 | 2004 | | 139 | 0.08 | 0.17 | 0.26 | 0.02 | 0.05 | 0.09 |
| 2005 | | 235 | 7.8 | 10.0 | 12.7 | 1.1 | 1.3 | 1.6 | 2005 | | 156 | 0.27 | 0.38 | 0.50 | 0.06 | 0.09 | 0.13 |
| 2006 | | 217 | 4.7 | 6.3 | 8.2 | 0.8 | 1.0 | 1.3 | 2006 | | 143 | 0.04 | 0.09 | 0.14 | 0.01 | 0.02 | 0.04 |
| 2007 | | 155 | 5.0 | 6.9 | 9.5 | 0.9 | 1.2 | 1.5 | 2007 | | 78 | 0.06 | 0.16 | 0.27 | 0.01 | 0.03 | 0.05 |
| 2008 | | 240 | 3.8 | 4.9 | 6.4 | 0.6 | 0.7 | 0.9 | 2008 | | 160 | 0.04 | 0.09 | 0.14 | 0.01 | 0.02 | 0.03 |
| 2009 | | 238 | 1.4 | 1.8 | 2.3 | 0.3 | 0.3 | 0.4 | 2009 | | 160 | 0.00 | 0.02 | 0.04 | 0.00 | 0.00 | 0.01 |
| 2010 | | 204 | 4.2 | 5.7 | 7.5 | 0.4 | 0.5 | 0.7 | 2010 | | 125 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 |
| 2011 | | 236 | 0.3 | 0.5 | 0.6 | 0.1 | 0.1 | 0.1 | 2011 | | 158 | 0.05 | 0.09 | 0.13 | 0.01 | 0.02 | 0.03 |
| 2012 | | 231 | 0.9 | 1.2 | 1.7 | 0.1 | 0.2 | 0.3 | 2012 | | 159 | 0.00 | 0.01 | 0.03 | 0.00 | 0.00 | 0.00 |
| 2013 | | 239 | 1.0 | 1.4 | 1.8 | 0.2 | 0.3 | 0.4 | 2013 | | 160 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 |
| 2014 | | 240 | 0.2 | 0.3 | 0.4 | 0.0 | 0.1 | 0.1 | 2014 | | 160 | 0.08 | 0.15 | 0.23 | 0.02 | 0.04 | 0.06 |
| 2015 | | 240 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 2015 | | 160 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2016 | | 240 | 0.5 | 0.8 | 1.0 | 0.1 | 0.2 | 0.2 | 2016 | | 160 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2017 | | 239 | 0.7 | 1.1 | 1.5 | 0.2 | 0.3 | 0.3 | 2017 | | 159 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2018 | | 217 | 0.8 | 1.1 | 1.5 | 0.1 | 0.2 | 0.2 | 2018 | | 137 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2019 | | 204 | 22.2 | 32.1 | 46.3 | 3.0 | 3.9 | 5.0 | 2019 | | 159 | 0.16 | 0.28 | 0.40 | 0.02 | 0.05 | 0.09 |
| 2020 | | 205 | 88.2 | 120.9 | 165.5 | 7.5 | 9.4 | 11.7 | 2020 | | 160 | 0.00 | 0.03 | 0.07 | 0.00 | 0.01 | 0.02 |
| 2021 | | 205 | 26.2 | 35.8 | 48.8 | 3.0 | 3.9 | 4.9 | 2021 | | 160 | 0.28 | 0.41 | 0.56 | 0.04 | 0.07 | 0.09 |
| 2022 | | | | | | | | | 2022 | | | | | | | | |
| 2023 | | | | | | | | | 2023 | | | | | | | | |
| 2024 | | | | | | | | | 2024 | | | | | | | | |
| 2025 | | | | | | | | | 2025 | | | | | | | | |

Table 21. Spot geometric mean indices of abundance, by number and biomass, overall and by age-class.





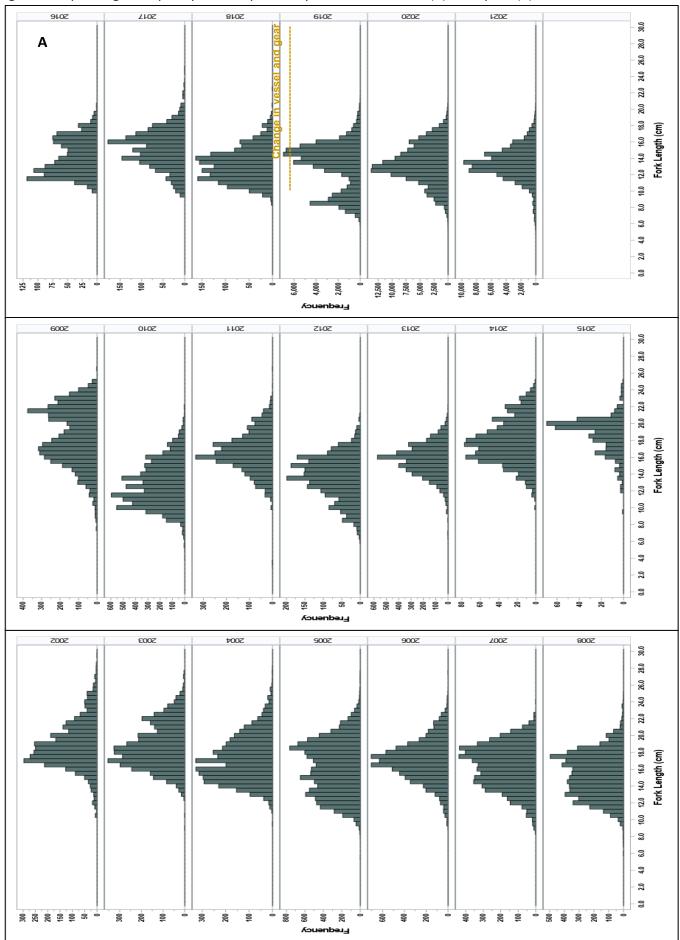


Figure 40. Spot length-frequency in Chesapeake Bay 2002-2021, overall (A) and by sex (B).

Figure 40. cont.

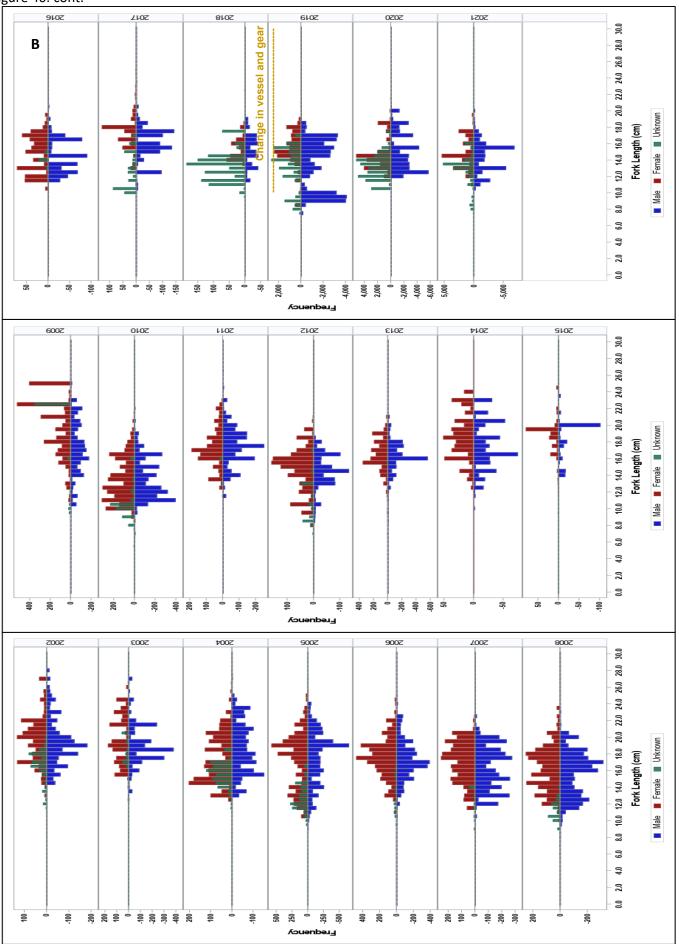
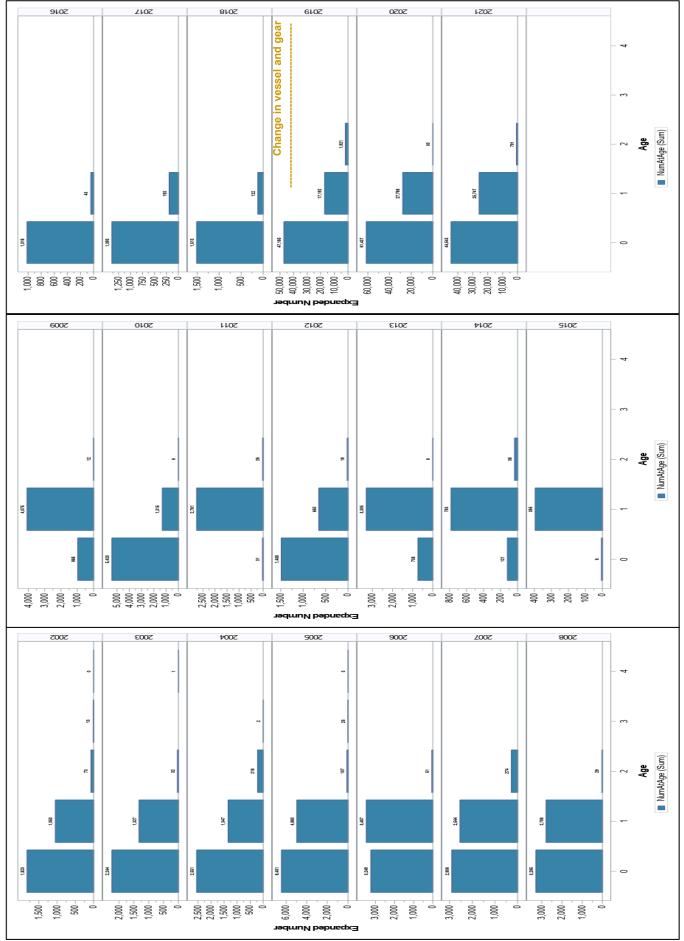
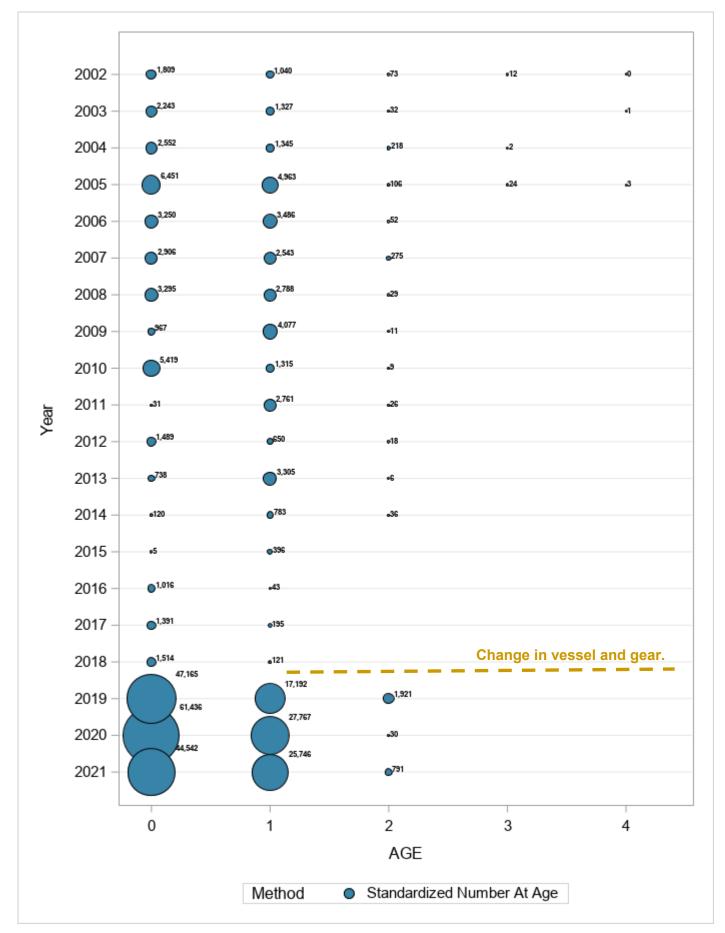
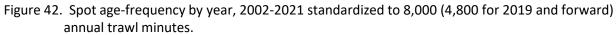


Figure 41. Spot age-frequency by year, 2002-2021.







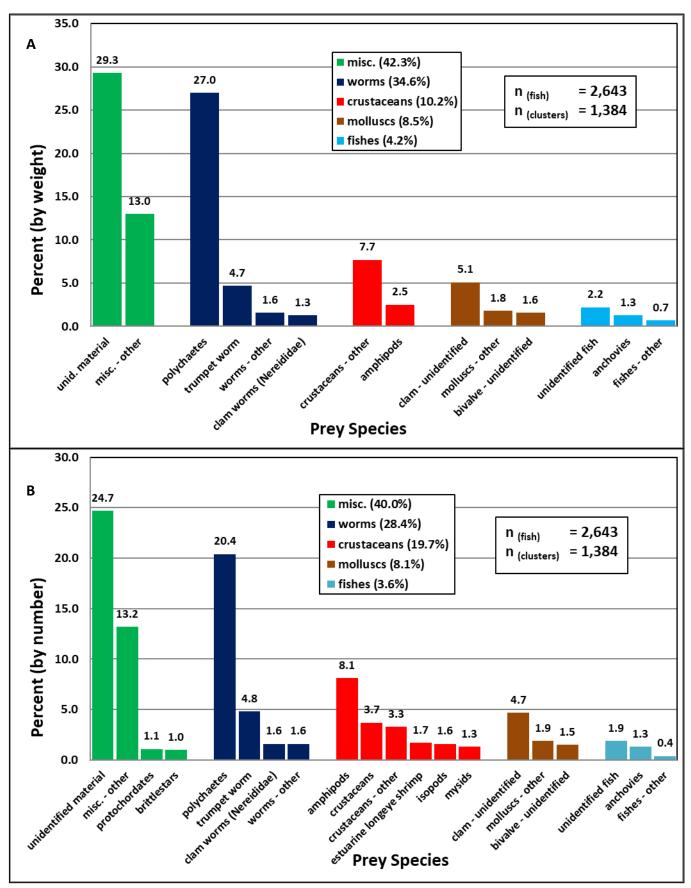
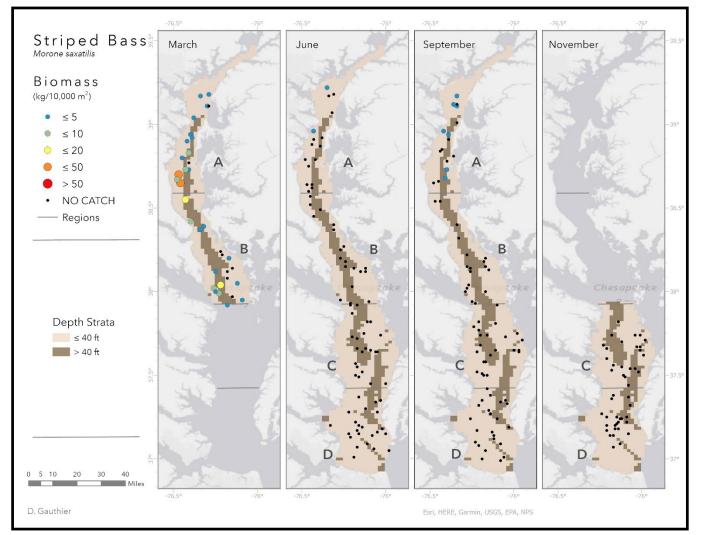


Figure 43. Diet composition, expressed as percent by weight (A) and percent by number (B) of Spot collected during ChesMMAP cruises in 2002-2020 combined.

| Year | Number | Biomass | Presence at Index Stations (%) | Number Measured | Age Specimens | Ages Read | Stomach Specimens | Stomachs |
|------|--------|-------------|-----------------------------------|--------------------|------------------|--------------|----------------------|----------|
| | Caught | Caught (kg) | . , | | • | | • | Analyzed |
| 2002 | 495 | 313.9 | 40.2 | 495 | 337 | 337 | 248 | 230 |
| 2003 | 765 | 710.1 | 55.6 | 765 | 501 | 501 | 367 | 354 |
| 2004 | 918 | 668.9 | 66.7 | 918 | 590 | 590 | 476 | 468 |
| 2005 | 2,245 | 982.4 | 63.5 | 1,919 | 724 | 724 | 528 | 513 |
| 2006 | 911 | 839.1 | 60.6 | 911 | 535 | 535 | 412 | 407 |
| 2007 | 579 | 423.4 | 47.3 | 579 | 389 | 389 | 246 | 241 |
| 2008 | 472 | 476.9 | 52.2 | 472 | 380 | 380 | 317 | 309 |
| 2009 | 315 | 243.1 | 37.2 | 315 | 198 | 198 | 152 | 149 |
| 2010 | 288 | 285.4 | 29.2 | 288 | 205 | 205 | 147 | 144 |
| 2011 | 284 | 224.9 | 46.9 | 284 | 237 | 237 | 178 | 178 |
| 2012 | 935 | 330.5 | 52.8 | 935 | 257 | 257 | 197 | 196 |
| 2013 | 695 | 482.3 | 50.9 | 695 | 373 | 373 | 259 | 124 |
| 2014 | 578 | 355.8 | 39.1 | 578 | 255 | 255 | 186 | 183 |
| 2015 | 718 | 398.5 | 38.3 | 718 | 319 | 319 | 133 | 131 |
| 2016 | 1,266 | 530.2 | 70.4 | 1,266 | 534 | 534 | 280 | 279 |
| 2017 | 1,466 | 829.0 | 43.0 | 1,313 | 426 | 426 | 270 | 268 |
| 2018 | 313 | 157.2 | 35.0 | 313 | 173 | 173 | 100 | 100 |
| 2019 | 2,559 | 679.0 | 50.0 | 1,134 | 265 | 265 | 200 | 200 |
| 2020 | 2,201 | 412.1 | 50.0 | 1,432 | 300 | 300 | 137 | 134 |
| 2021 | 1,881 | 528.8 | 33.8 | 1,400 | 195 | 195 | 104 | 102 |

Table 22. Striped Bass sampling rates and preserved specimen analysis status by year (shaded rows represent a change in vessel and trawl gear beginning in June 2019).

Figure 44. Station specific biomass of Striped Bass in Chesapeake Bay, 2021.



| Year | Age | n | Nup | nerical In | dex | Bio | mass In | dex | Year | Age | n | Nur | nerical In | dex | Rio | mass Ind | ex |
|--------------|-----|----------|-------------|--------------|------------|------------|------------|------------|--------------|-----|----------|------------|------------|------------|-----|------------|------------|
| rear | Age | | LCI | Index | UCI | LCI | Index | UCI | lear | Age | | LCI | Index | UCI | LCI | Index | UCI |
| 2002 | All | 28 | 0.0 | 0.4 | 1.0 | 0.0 | 0.4 | 0.8 | 2002 | 3 | 28 | 0.0 | 0.1 | 0.2 | 0.0 | 0.0 | 0.1 |
| 2003 | | 34 | 2.5 | 4.1 | 6.3 | 1.3 | 2.5 | 4.3 | 2003 | | 34 | 0.8 | 1.4 | 2.3 | 0.6 | 1.1 | 1.9 |
| 2004 | | 40 | 5.0 | 6.6 | 8.5 | 3.4 | 4.7 | 6.4 | 2004 | | 40 | 1.9 | 2.7 | 3.7 | 1.4 | 2.0 | 2.8 |
| 2005 2006 | | 35 35 | 2.7 1.6 | 4.4 2.6 | 6.9 4.1 | 1.4 1.3 | 2.3 2.3 | 3.5 3.7 | 2005 2006 | | 35 35 | 0.3 1.1 | 0.6 1.8 | 1.0 2.8 | 0.2 | 0.3 1.2 | 0.5 1.7 |
| 2000 | | 33 | 2.5 | 3.8 | 5.6 | 1.5 | 2.2 | 3.7 | 2000 | | 33 | 0.8 | 1.3 | 1.9 | 0.6 | 0.9 | 1.7 |
| 2008 | | 35 | 3.7 | 5.1 | 6.9 | 2.9 | 4.5 | 6.9 | 2008 | | 35 | 1.3 | 1.9 | 2.7 | 0.8 | 1.2 | 1.6 |
| 2009 | | 35 | 0.8 | 1.4 | 2.2 | 0.8 | 1.5 | 2.4 | 2009 | | 35 | 0.0 | 0.2 | 0.3 | 0.0 | 0.1 | 0.2 |
| 2010 | | 34 | 0.7 | 1.3 | 2.1 | 0.7 | 1.4 | 2.6 | 2010 | | 34 | 0.4 | 0.7 | 1.1 | 0.3 | 0.6 | 0.9 |
| 2011 | | 35 | 0.9 | 1.5 | 2.4 | 0.6 | 1.2 | 1.9 | 2011 | | 35 | 0.1 | 0.4 | 0.6 | 0.1 | 0.3 | 0.5 |
| 2012 | | 35 | 2.1 | 2.4 | F 2 | 2.1 | 2.4 | F 2 | 2012 | | 25 | 0.4 | 0.9 | 1.2 | 0.2 | 0.7 | 1.1 |
| 2013 2014 | | 35 35 | 2.1 0.5 | 3.4 1.4 | 5.3 2.7 | 2.1 0.3 | 3.4 0.8 | 5.2 1.5 | 2013 2014 | | 35 35 | 0.4 | 0.8 | 1.3 2.4 | 0.3 | 0.7 | 1.1 1.4 |
| 2014 | | 35 | 0.3 | 0.8 | 1.7 | 0.0 | 0.4 | 1.0 | 2014 | | 35 | 0.0 | 0.2 | 0.4 | 0.2 | 0.1 | 0.3 |
| 2016 | | 35 | 3.3 | 4.9 | 7.0 | 1.7 | 2.4 | 3.4 | 2016 | | 35 | 0.2 | 0.4 | 0.7 | 0.1 | 0.3 | 0.4 |
| 2017 | | 34 | 1.0 | 1.8 | 3.1 | 0.6 | 1.0 | 1.5 | 2017 | | 34 | 0.2 | 0.6 | 1.0 | 0.1 | 0.3 | 0.5 |
| 2018 | | | | | | | | | 2018 | | | | | | | | |
| 2019 | | 35 | 0.3 | 0.6 | 0.9 | 0.2 | 0.6 | 1.1 | 2019 | | 35 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.1 |
| 2020 | | 35 35 | 10.4 5.2 | 18.4 10.8 | 32.3 | 3.6 | 5.9 4.4 | 9.3 8.4 | 2020 | | 35 35 | 0.8 | 1.5 | 2.5 | 0.4 | 0.8 | 1.3 2.7 |
| 2021 2022 | | 35 | 5.2 | 10.8 | 21.4 | 2.1 | 4.4 | 8.4 | 2021 2022 | | 35 | 1.5 | 3.1 | 5.9 | 0.7 | 1.5 | 2.7 |
| 2022 | | | | | | | | | 2022 | | | | | | | | |
| 2024 | | | | | | | | | 2024 | | | | | | | | |
| 2025 | | | | | | | | | 2025 | | | | | | | | |
| 2002 | 1 | 28 | 0.0 | 0.1 | 0.3 | 0.0 | 0.1 | 0.2 | 2002 | 4+ | 28 | 0.0 | 0.1 | 0.2 | 0.0 | 0.1 | 0.2 |
| 2003 | | 34 | 0.0 | 0.1 | 0.1 | 0.0 | 0.0 | 0.0 | 2003 | | 34 | 0.0 | 0.2 | 0.3 | 0.0 | 0.1 | 0.3 |
| 2004 | | 40 35 | 0.1 | 0.3 | 0.6 | 0.0 | 0.1 | 0.1 | 2004 2005 | | 40 | 0.8 | 1.1 | 1.4 | 0.8 | 1.1 0.2 | 1.5 |
| 2005 2006 | | 35 35 | 0.0 | 0.1 | 0.3 | 0.0 | 0.0 | 0.1 | 2005 | | 35 35 | 0.1 | 0.3 | 0.5 0.4 | 0.1 | 0.2 | 0.4 0.7 |
| 2000 | | 33 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 2000 | | 33 | 0.7 | 1.1 | 1.6 | 0.1 | 0.8 | 1.2 |
| 2008 | | 35 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 2008 | | 35 | 0.6 | 0.9 | 1.4 | 0.5 | 0.9 | 1.4 |
| 2009 | | 35 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 2009 | | 35 | 0.2 | 0.4 | 0.6 | 0.2 | 0.4 | 0.7 |
| 2010 | | 34 | 0.0 | 0.1 | 0.2 | 0.0 | 0.0 | 0.1 | 2010 | | 34 | 0.0 | 0.1 | 0.2 | 0.0 | 0.1 | 0.2 |
| 2011 | | 35 | 0.0 | 0.1 | 0.1 | 0.0 | 0.0 | 0.0 | 2011 | | 35 | 0.2 | 0.5 | 1.0 | 0.2 | 0.5 | 0.9 |
| 2012 2013 | | 35 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 2012 2013 | | 35 | 0.5 | 0.8 | 1.3 | 0.4 | 0.7 | 1.1 |
| 2013 | | 35 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 2013 | | 35 | 0.1 | 0.3 | 0.5 | 0.4 | 0.2 | 0.3 |
| 2015 | | 35 | 0.0 | 0.3 | 0.6 | 0.0 | 0.1 | 0.2 | 2015 | | 35 | 0.0 | 0.4 | 1.1 | 0.0 | 0.3 | 0.8 |
| 2016 | | 35 | 0.2 | 0.5 | 1.0 | 0.1 | 0.2 | 0.3 | 2016 | | 35 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.1 |
| 2017 | | 34 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 2017 | | 34 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.1 |
| 2018 | | 25 | 0.0 | 0.1 | 0.1 | 0.0 | | 0.1 | 2018 | | 25 | 0.1 | 0.0 | 0.1 | 0.1 | 0.2 | 0.4 |
| 2019 2020 | | 35 35 | 0.0 | 0.1 | 0.1 | 0.0 0.3 | 0.0 | 0.1 0.8 | 2019 2020 | | 35 35 | 0.1 | 0.2 | 0.4 0.6 | 0.1 | 0.2 | 0.4 0.6 |
| 2020 | | 35 | 0.3 | 0.9 | 1.7 | 0.1 | 0.2 | 0.3 | 2020 | | 35 | 0.0 | 0.3 | 0.8 | 0.1 | 0.4 | 0.9 |
| 2022 | | | | | | | | | 2022 | | | | | | | | |
| 2023 | | | | | | | | | 2023 | | | | | | | | |
| 2024 | | | | | | | | | 2024 | | | | | | | | |
| 2025 | | | | | | | | | 2025 | | | | | | | | |
| 2002 | 2 | 28 | 0.0 | 0.3 | 0.7 | 0.0 | | 0.5 | | | | | | | | | |
| 2003 2004 | | 34 40 | 1.2 0.4 | 1.9 0.7 | 2.8 1.0 | 0.6 | 1.0 0.4 | 1.5 0.5 | | | | | | | | | |
| 2004 | | 35 | 1.5 | 2.7 | 4.4 | 0.2 | 1.3 | 2.0 | | | | | | | | | |
| 2006 | | 35 | 0.3 | 0.6 | 1.0 | 0.1 | 0.3 | 0.5 | | | | | | | | | |
| 2007 | | 33 | 0.8 | 1.3 | 1.9 | 0.4 | 0.6 | 0.8 | | | | | | | | | |
| 2008 | | 35 | 0.1 | 0.3 | 0.5 | 0.0 | 0.1 | 0.2 | | | | | | | | | |
| 2009 | | 35 24 | 0.3 | 0.6 | 0.9 | | 0.4 | 0.6 | | | | | | | | | |
| 2010 2011 | | 34 35 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 0.5 | | | | | | | | | |
| 2011 | | 33 | 0.2 | 0.5 | 1.0 | 0.1 | 0.0 | 0.5 | | | | | | | | | |
| 2013 | | 35 | 0.5 | 1.0 | 1.7 | 0.3 | 0.6 | 1.0 | | | | | | | | | |
| 2014 | | 35 | 0.0 | 0.1 | 0.3 | 0.0 | 0.0 | 0.1 | | | | | | | | | |
| 2015 | | 35 | 0.0 | 0.3 | 0.6 | | 0.1 | 0.4 | | | | | | | | | |
| 2016 | | 35 | 1.4 | 2.3 | 3.6 | 0.7 | 1.0 | 1.5 | | | | | | | | | |
| 2017 2018 | | 34 | 0.5 | 1.2 | 2.2 | 0.3 | 0.6 | 0.8 | | | | | | | | | |
| 2018 | | 35 | 0.0 | 0.1 | 0.3 | 0.0 | 0.1 | 0.2 | | | | | | | | | |
| 2020 | | 35 | 4.9 | 9.1 | 16.3 | 1.3 | 2.2 | 3.6 | | | | | | | | | |
| 2021 | | 35 | 2.8 | 6.0 | 11.8 | 1.1 | 2.3 | 4.2 | | | | | | | | | |
| 2022 | | | | | | | | | | | | | | | | | |
| 2023 | | | | | | | | | | | | | | | | | |
| 2024 2025 | | | | | | | | | | | | | | | | | |
| 2025 | | | | | | | | | | | | | | | | | |

Table 23. Striped Bass (March) geometric mean indices of abundance, by number and biomass, overall and by age-class.

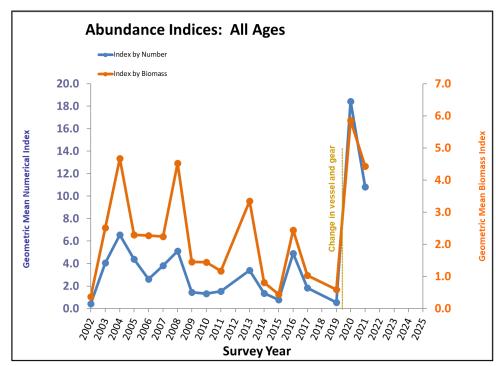
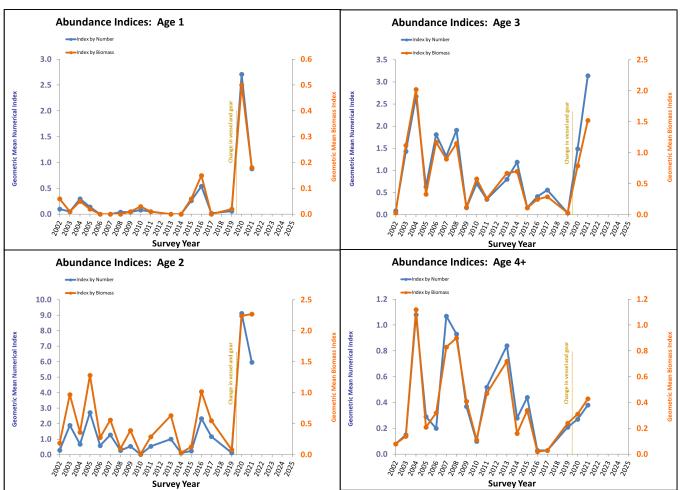


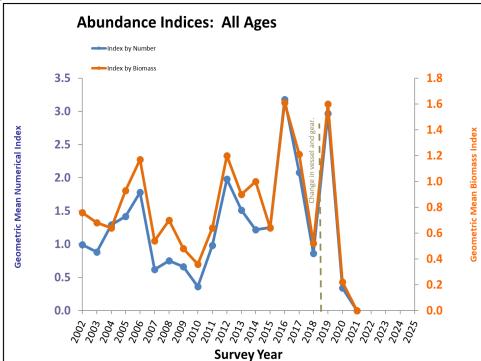
Figure 45. Striped Bass (March) geometric mean indices of abundance, by number and biomass, for all ages combined and by age-class.

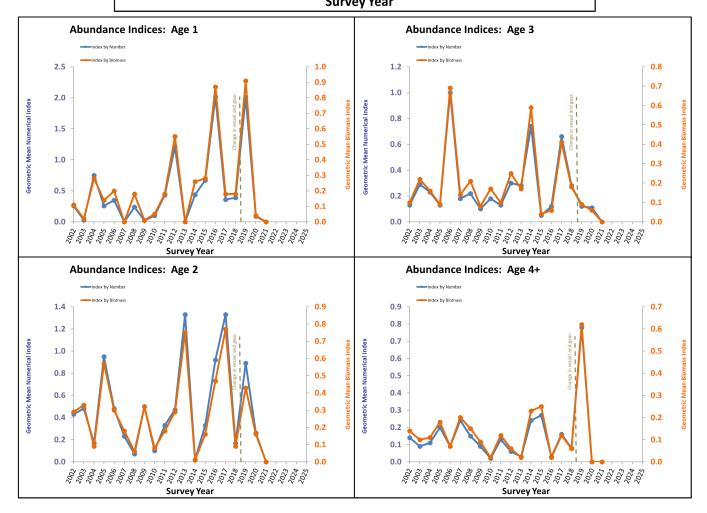


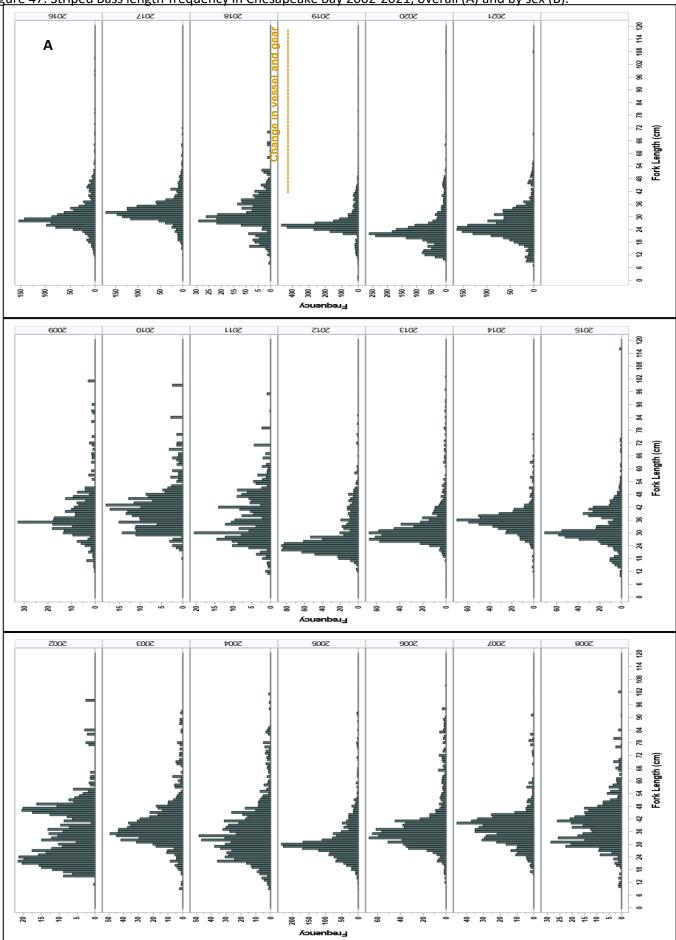
| Year | | lass. | Nur | nerical In | dex | Bic | mass In | dex | Year | Age | n | Num | erical In | dex | Bio | mass Inc | dex |
|--------------|-----|----------|------------|------------|------------|------------|------------|------------|--------------|-----|----------|-----|------------|------------|-----|----------|------------|
| | | | LCI | Index | UCI | LCI | Index | UCI | | | | LCI | Index | UCI | LCI | Index | UCI |
| 2002 | All | 74 | 0.7 | 1.0 | 1.4 | 0.5 | 0.8 | 1.1 | 2002 | 3 | 74 | 0.1 | 0.1 | 0.2 | 0.0 | 0.1 | 0.2 |
| 2003 | | 90 | 0.6 | 0.9 | 1.3 | 0.4 | | 1.0 | 2003 | | 90 | 0.2 | 0.3 | 0.4 | 0.1 | 0.2 | 0.3 |
| 2004 2005 | | 85 79 | 0.9 0.9 | 1.3 1.4 | 1.7 2.1 | 0.4 | 0.6 | 0.9 1.4 | 2004 2005 | | 85 79 | 0.1 | 0.2 | 0.4 | 0.1 | 0.2 | 0.3 0.2 |
| 2005 | | 74 | 1.2 | 1.4 | 2.1 | 0.0 | 1.2 | 1.4 | 2005 | | 74 | 0.6 | 1.0 | 1.5 | 0.4 | 0.7 | 1.1 |
| 2007 | | 77 | 0.3 | 0.6 | 1.0 | 0.3 | 0.5 | 0.9 | 2007 | | 77 | 0.1 | 0.2 | 0.3 | 0.0 | 0.1 | 0.3 |
| 2008 | | 80 | 0.5 | 0.8 | 1.1 | 0.4 | 0.7 | 1.1 | 2008 | | 80 | 0.1 | 0.2 | 0.4 | 0.1 | 0.2 | 0.3 |
| 2009 | | 78 | 0.4 | 0.7 | 1.0 | 0.2 | 0.5 | 0.8 | 2009 | | 78 | 0.0 | 0.1 | 0.2 | 0.0 | 0.1 | 0.2 |
| 2010 | | 79 | 0.2 | 0.4 | 0.6 | 0.1 | 0.4 | 0.6 | 2010 | | 79 | 0.0 | 0.2 | 0.3 | 0.0 | 0.2 | 0.3 |
| 2011 | | 78 | 0.7 | 1.0 | 1.4 | 0.4 | 0.6 | 0.9 | 2011 | | 78 | 0.0 | 0.1 | 0.2 | 0.0 | 0.1 | 0.2 |
| 2012 2013 | | 72 79 | 1.2 0.9 | 2.0 1.5 | 3.0 2.3 | 0.8 0.6 | 1.2 0.9 | 1.8 1.3 | 2012 2013 | | 72 79 | 0.1 | 0.3 | 0.5 0.4 | 0.1 | 0.3 | 0.4 0.3 |
| 2013 | | 80 | 0.9 | 1.5 | 1.9 | 0.6 | 1.0 | 1.5 | 2013 | | 80 | 0.1 | 0.3 | 1.2 | 0.1 | 0.2 | 1.0 |
| 2015 | | 80 | 0.8 | 1.3 | 1.9 | 0.4 | | 0.9 | 2015 | | 80 | 0.0 | 0.1 | 0.1 | 0.0 | 0.0 | 0.1 |
| 2016 | | 80 | 2.3 | 3.2 | 4.3 | 1.2 | 1.6 | 2.1 | 2016 | | 80 | 0.1 | 0.1 | 0.2 | 0.0 | 0.1 | 0.1 |
| 2017 | | 80 | 1.3 | 2.1 | 3.2 | 0.7 | 1.2 | 1.8 | 2017 | | 80 | 0.4 | 0.7 | 1.0 | 0.2 | 0.4 | 0.6 |
| 2018 | | 80 | 0.6 | 0.9 | 1.2 | 0.3 | 0.5 | 0.7 | 2018 | | 80 | 0.2 | 0.3 | 0.4 | 0.1 | 0.2 | 0.3 |
| 2019 | | 45 | 1.5 | 3.0 | 5.2 | 0.8 | 1.6 | 2.7 | 2019 | | 45 | 0.0 | 0.1 | 0.3 | 0.0 | 0.1 | 0.2 |
| 2020 | | 45 | 0.0 | 0.3 | 0.8 | 0.0 | 0.2 | 0.5 | 2020 | | 45 | 0.0 | 0.1 | 0.3 | 0.0 | 0.1 | 0.2 |
| 2021 2022 | | 45 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 2021 2022 | | 45 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 2022 | | | | | | | | | 2022 | | | | | | | | |
| 2024 | | | | | | | | | 2023 | | | | | | | | |
| 2025 | | | | | | | | | 2025 | | | | | | | | |
| 2002 | 1 | 74 | 0.1 | 0.3 | 0.4 | 0.0 | 0.1 | 0.2 | 2002 | 4+ | 74 | 0.0 | 0.1 | 0.3 | 0.0 | 0.1 | 0.2 |
| 2003 | | 90 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 2003 | | 90 | 0.1 | 0.1 | 0.1 | 0.0 | 0.1 | 0.2 |
| 2004 | | 85 | 0.5 | 0.8 | 1.0 | 0.2 | 0.3 | 0.4 | 2004 | | 85 | 0.0 | 0.1 | 0.2 | 0.0 | 0.1 | 0.2 |
| 2005 | | 79 | 0.1 | 0.3 | 0.4 | 0.1 | 0.1 | 0.2 | 2005 | | 79 | 0.1 | 0.2 | 0.3 | 0.1 | 0.2 | 0.3 |
| 2006 | | 74 77 | 0.2 | 0.4 | 0.5 | 0.1 | 0.2 | 0.3 | 2006 | | 74 | 0.0 | 0.1 | 0.2 | 0.0 | 0.1 | 0.2 |
| 2007 2008 | | 80 | 0.0 | 0.0 | 0.0 0.4 | 0.0 | 0.0 | 0.0 0.3 | 2007 2008 | | 77 80 | 0.1 | 0.2 0.2 | 0.4 | 0.1 | 0.2 | 0.3 0.3 |
| 2008 | | 78 | 0.0 | 0.0 | 0.4 | 0.0 | 0.2 | 0.0 | 2008 | | 78 | 0.0 | 0.2 | 0.3 | 0.0 | 0.1 | 0.3 |
| 2010 | | 79 | 0.0 | 0.1 | 0.2 | 0.0 | | 0.1 | 2010 | | 79 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 2011 | | 78 | 0.3 | 0.4 | 0.6 | 0.1 | 0.2 | 0.3 | 2011 | | 78 | 0.0 | 0.1 | 0.2 | 0.0 | 0.1 | 0.2 |
| 2012 | | 72 | 0.7 | 1.2 | 1.9 | 0.3 | 0.6 | 0.8 | 2012 | | 72 | 0.0 | 0.1 | 0.1 | 0.0 | 0.1 | 0.1 |
| 2013 | | 79 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 2013 | | 79 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 |
| 2014 | | 80 | 0.3 | 0.4 | 0.7 | 0.2 | 0.3 | 0.4 | 2014 | | 80 | 0.1 | 0.2 | 0.4 | 0.1 | 0.2 | 0.4 |
| 2015 | | 80 | 0.4 | 0.7 | 1.0 | 0.2 | 0.3 | 0.4 | 2015 | | 80 | 0.1 | 0.3 | 0.4 | 0.1 | 0.3 | 0.4 |
| 2016 2017 | | 80 80 | 1.3 0.1 | 2.0 0.4 | 2.9 0.6 | 0.6 0.1 | 0.9 | 1.2 0.3 | 2016 2017 | | 80 80 | 0.0 | 0.0 0.2 | 0.1 0.3 | 0.0 | 0.0 | 0.0 0.2 |
| 2018 | | 80 | 0.1 | 0.4 | 0.6 | 0.1 | 0.2 | 0.3 | 2017 | | 80 | 0.0 | 0.1 | 0.3 | 0.0 | 0.1 | 0.2 |
| 2019 | | 45 | 0.9 | 2.0 | 3.9 | 0.4 | | 1.7 | 2019 | | 45 | 0.4 | 0.8 | 1.3 | 0.3 | 0.6 | 1.0 |
| 2020 | | 45 | 0.0 | 0.1 | 0.2 | 0.0 | 0.0 | 0.1 | 2020 | | 45 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 2021 | | 45 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 2021 | | 45 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 2022 | | | | | | | | | 2022 | | | | | | | | |
| 2023 | | | | | | | | | 2023 | | | | | | | | |
| 2024 2025 | | | | | | | | | 2024 2025 | | | | | | | | |
| 2023 | 2 | 74 | 0.3 | 0.4 | 0.6 | 0.2 | 0.3 | 0.4 | 2025 | | | | | | | | |
| 2002 | - | 90 | 0.3 | 0.4 | 0.0 | 0.2 | 0.3 | 0.4 | | | | | | | | | |
| 2003 | | 85 | 0.1 | 0.2 | 0.3 | 0.0 | | 0.2 | | | | | | | | | |
| 2005 | | 79 | 0.5 | 1.0 | 1.5 | 0.3 | 0.6 | 0.9 | | | | | | | | | |
| 2006 | | 74 | 0.3 | 0.5 | 0.8 | 0.2 | 0.3 | 0.5 | | | | | | | | | |
| 2007 | | 77 | 0.1 | 0.2 | 0.4 | 0.0 | 0.2 | 0.3 | | | | | | | | | |
| 2008 | | 80 | 0.0 | 0.1 | 0.1 | 0.0 | 0.1 | 0.1 | | | | | | | | | |
| 2009 2010 | | 78 79 | 0.3 | 0.5 | 0.8 0.2 | 0.2 | 0.3 | 0.5 0.2 | | | | | | | | | |
| 2010 | | 79 78 | 0.0 | 0.1 | 0.2 | 0.0 | | 0.2 | | | | + | | | | | |
| 2011 | | 72 | 0.2 | 0.5 | 0.5 | 0.1 | | 0.3 | | | | + | | | | | |
| 2013 | | 79 | 0.8 | 1.3 | 2.0 | 0.5 | 0.8 | 1.1 | | | | | | | | | |
| 2014 | | 80 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | | | | | | | | | |
| 2015 | | 80 | 0.2 | 0.3 | 0.5 | 0.1 | 0.2 | 0.3 | | | | | | | | | |
| 2016 | | 80 | 0.7 | 0.9 | 1.2 | 0.3 | | 0.6 | | | | | | | | | |
| 2017 | | 80 | 0.8 | 1.3 | 2.1 | 0.5 | 0.8 | 1.2 | | | | | | | | | |
| 2018 2019 | | 80 45 | 0.1 | 0.2 | 0.3 1.6 | 0.0 | 0.1 | 0.2 | | | | | | | | | |
| 2019 | | 45 45 | 0.4 | 0.9 | 0.6 | 0.1 | | 0.8 | | | | + | | | | | |
| 2020 | | 45 | 0.0 | 0.0 | 0.0 | 0.0 | | 0.4 | | | | | | | | | |
| 2022 | | | 0.0 | | 0.0 | 5.5 | | 0.0 | | | | | | | | | |
| 2023 | | | | | | | | | | | | | | | | | |
| 2024 | | | | | | | | | | | | | | | | | |
| 2025 | | | | | | | | | | | | | | | | | |

Table 24. Striped Bass (November) geometric mean indices of abundance, by number and biomass, overall and by age-class.

Figure 46. Striped Bass (November) geometric mean indices of abundance, by number and biomass, for all ages combined and by age-class.







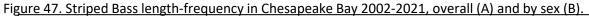
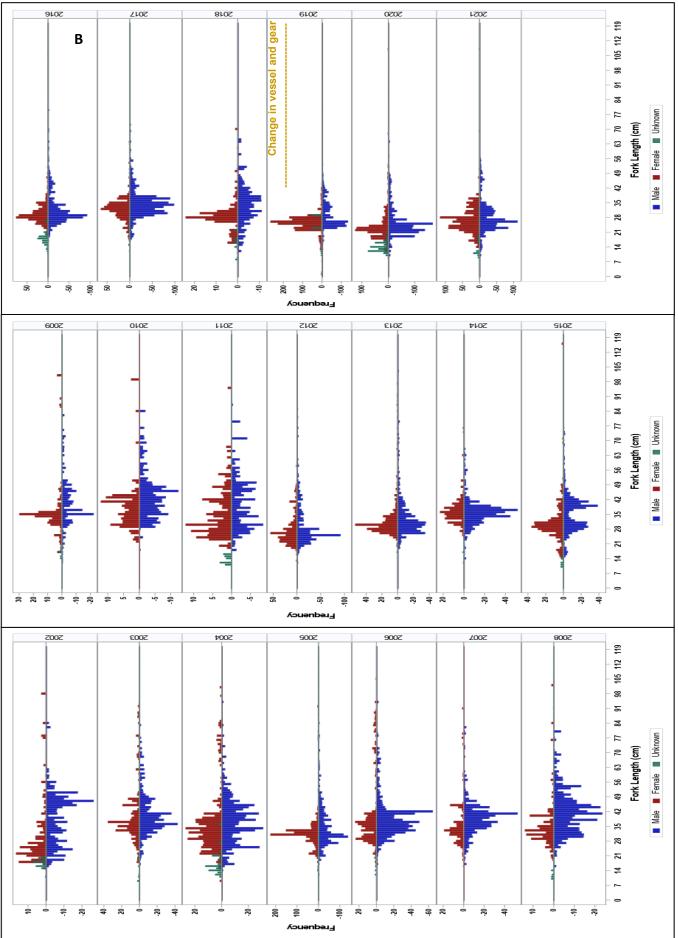
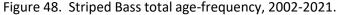
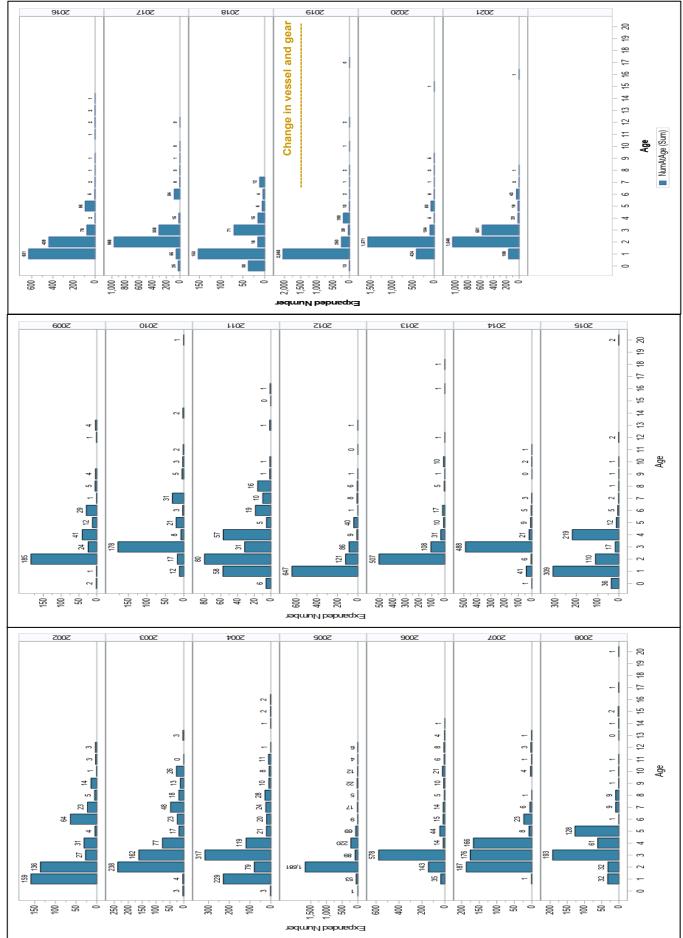


Figure 47. cont.







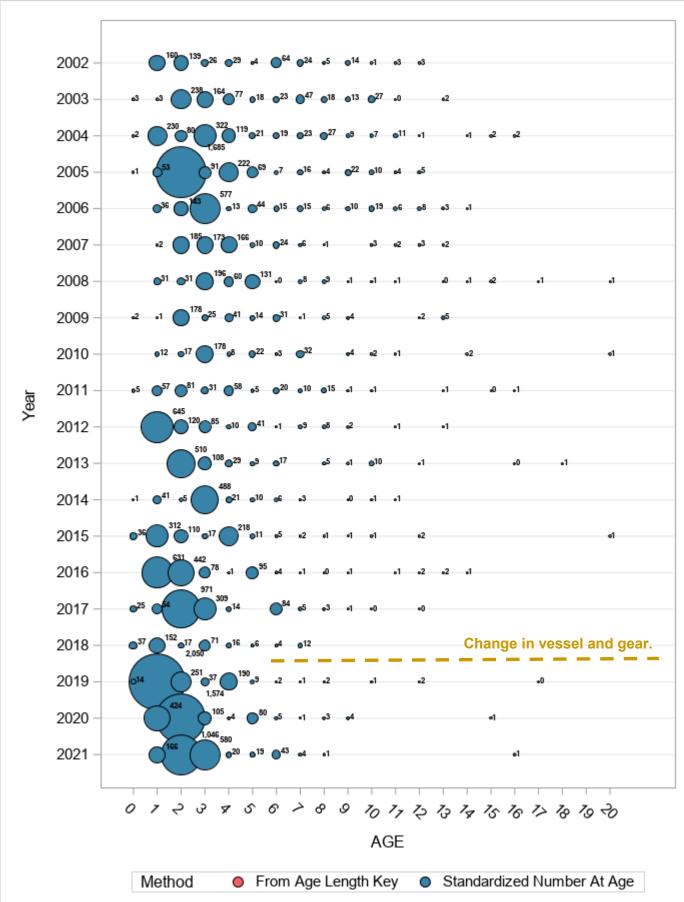


Figure 49. Striped Bass age-frequency by year, 2002-2021 standardized to 8,000 (4,800 for 2019 and forward) annual trawl minutes.

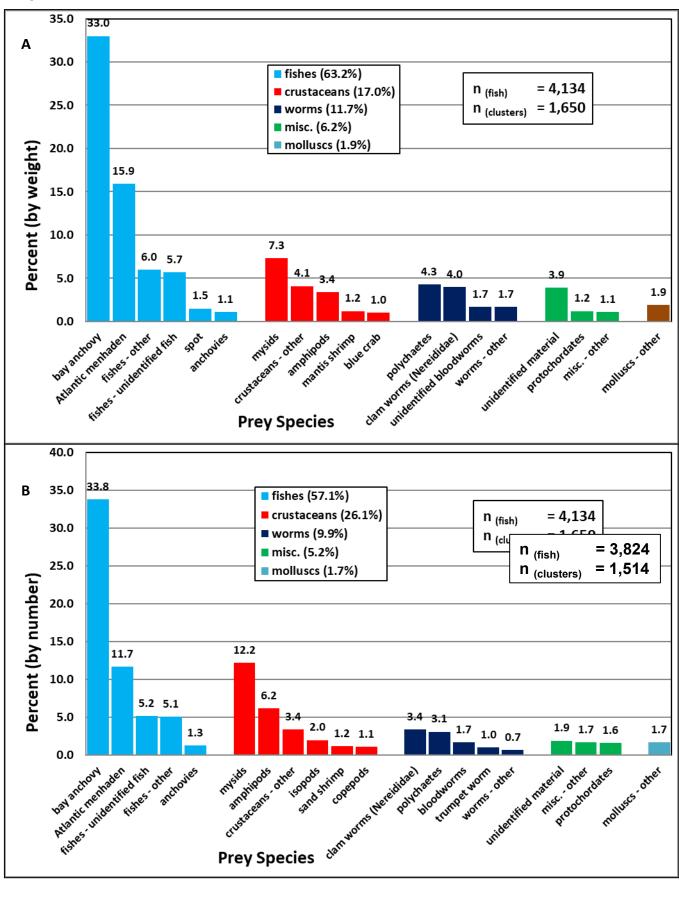
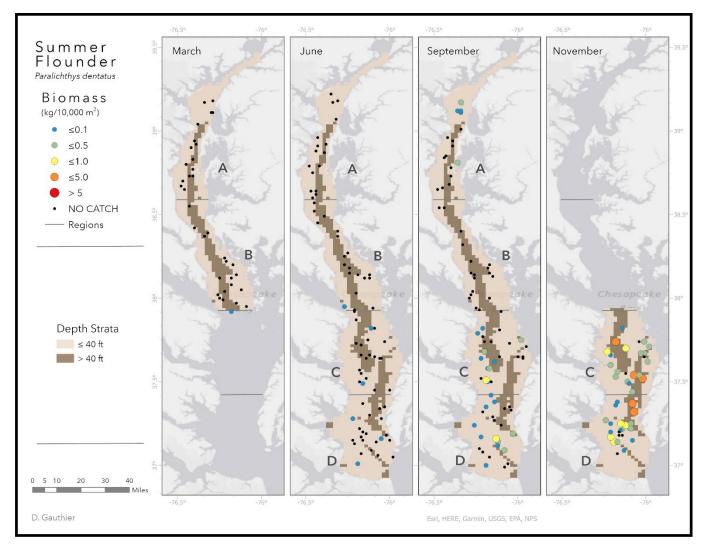


Figure 50. Diet composition, expressed as percent by weight (A) and percent by number (B) of Striped Bass collected during ChesMMAP cruises in 2002-2020 combined.

| | Number | Biomass | Presence at Index | Number | Age | Ages | Stomach | Stomachs |
|------|--------|-------------|-------------------|----------|-----------|------|-----------|----------|
| Year | Caught | Caught (kg) | Stations (%) | Measured | Specimens | Read | Specimens | Analyzed |
| 2002 | 770 | 430.5 | 84.0 | 770 | 649 | 649 | 425 | 409 |
| 2003 | 563 | 341.5 | 67.6 | 562 | 441 | 441 | 325 | 316 |
| 2004 | 728 | 309.7 | 72.0 | 728 | 565 | 565 | 377 | 372 |
| 2005 | 759 | 386.7 | 89.5 | 759 | 669 | 669 | 420 | 409 |
| 2006 | 932 | 453.1 | 88.6 | 932 | 755 | 755 | 444 | 430 |
| 2007 | 567 | 259.1 | 81.8 | 563 | 489 | 489 | 317 | 313 |
| 2008 | 638 | 280.9 | 77.8 | 638 | 543 | 543 | 354 | 346 |
| 2009 | 393 | 187.1 | 66.7 | 393 | 369 | 369 | 243 | 239 |
| 2010 | 385 | 180.0 | 67.8 | 385 | 354 | 354 | 215 | 209 |
| 2011 | 211 | 125.3 | 62.9 | 211 | 208 | 208 | 111 | 107 |
| 2012 | 92 | 33.4 | 31.0 | 92 | 91 | 91 | 57 | 53 |
| 2013 | 110 | 35.7 | 33.7 | 110 | 107 | 107 | 51 | 45 |
| 2014 | 63 | 16.7 | 30.0 | 63 | 63 | 63 | 40 | 40 |
| 2015 | 129 | 41.9 | 35.6 | 129 | 127 | 127 | 72 | 72 |
| 2016 | 77 | 21.8 | 30.0 | 77 | 77 | 77 | 40 | 39 |
| 2017 | 135 | 35.0 | 28.9 | 135 | 128 | 128 | 85 | 85 |
| 2018 | 105 | 26.5 | 15.6 | 105 | 96 | 96 | 44 | 44 |
| 2019 | 623 | 78.7 | 90.0 | 623 | 385 | 385 | 220 | 216 |
| 2020 | 286 | 42.0 | 64.4 | 286 | 215 | 215 | 105 | 105 |
| 2021 | 267 | 23.6 | 67.8 | 267 | 185 | 185 | 82 | 82 |

Table 25. Summer Flounder sampling rates and preserved specimen analysis status by year (shaded rows represent a change in vessel and trawl gear beginning in June 2019).

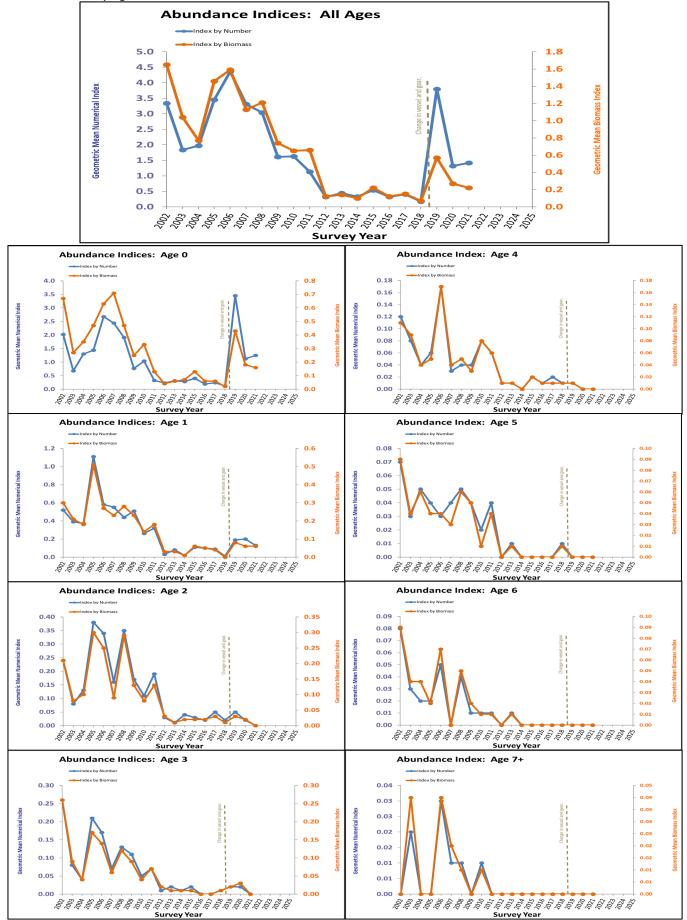
Figure 51. Station specific biomass of Summer Flounder in Chesapeake Bay, 2021.



| Table 26. Summer Flounder geometric mean indices of abundance, by number and biomass, overall and by | |
|--|--|
| age-class. | |

| Year Age | n | Nun | nerical Inc | | Bio LCI | mass In | | Year | Age | n | Num | nerical Inc | | Bio | mass Inc | | Year | Age | n | Nun | nerical Inc | | | mass Ind | |
|--------------|----------|--------------|---------------|--------------|------------|---------------|--------------|--------------|-----|-----------|--------------|-------------|--------------|------|---------------|--------------|--------------|-----|----------|------|---------------|--------------|-------------|---------------|--------------|
| 2002 All | 75 | 2.48 | Index 3.34 | UCI 4.42 | 1.21 | Index 1.65 | UCI 2.16 | 2002 | 2 7 | 75 | 0.12 | 0.24 | UCI 0.36 | 0.10 | Index 0.21 | UCI 0.32 | 2002 | 5 | 75 | 0.02 | Index 0.07 | UCI 0.12 | LCI 0.02 | Index 0.09 | UCI 0.17 |
| 2002 711 | 101 | 1.35 | 1.84 | 2.44 | 0.76 | 1.04 | 1.36 | 2002 | | .01 | 0.03 | 0.08 | 0.13 | 0.03 | 0.08 | 0.13 | 2002 | 5 | 101 | 0.02 | 0.03 | 0.06 | 0.02 | 0.04 | 0.08 |
| 2004 | 92 | 1.49 | 1.97 | 2.55 | 0.59 | 0.77 | 0.98 | 2004 | | 92 | 0.07 | 0.13 | 0.20 | 0.05 | 0.10 | 0.16 | 2004 | | 92 | 0.01 | 0.05 | 0.09 | 0.01 | 0.06 | 0.10 |
| 2005 | 86 | 2.78 | 3.45 | 4.25 | 1.16 | 1.46 | 1.80 | 2005 | 8 | 86 | 0.25 | 0.38 | 0.52 | 0.19 | 0.30 | 0.41 | 2005 | | 86 | 0.00 | 0.04 | 0.07 | 0.00 | 0.04 | 0.07 |
| 2006 | 79 | 3.42 | 4.36 | 5.52 | 1.23 | 1.59 | 2.01 | 2006 | | 79 | 0.20 | 0.34 | 0.48 | 0.14 | 0.25 | 0.37 | 2006 | | 79 | 0.00 | 0.03 | 0.06 | 0.00 | 0.04 | 0.09 |
| 2007 | 44 | 2.35 | 3.31 | 4.53 | 0.81 | 1.13 | 1.52 | 2007 | | 44 | 0.06 | 0.16 | 0.27 | 0.03 | 0.09 | 0.15 | 2007 | | 44 | 0.00 | 0.04 | 0.09 | 0.00 | 0.03 | 0.06 |
| 2008 | 90 | 2.25 | 3.04 | 4.01 | 0.87 | 1.21 | 1.60 | 2008 | | 90 | 0.22 | 0.35 | 0.50 | 0.17 | 0.29 | 0.42 | 2008 | | 90 | 0.01 | 0.05 | 0.10 | 0.01 | 0.06 | 0.10 |
| 2009 | 90 | 1.19 | 1.61 | 2.10 | 0.54 | 0.74 | 0.96 | 2009 | | 90 | 0.08 | 0.17 | 0.26 | 0.06 | 0.13 | 0.20 | 2009 | _ | 90 | 0.00 | 0.04 | 0.08 | 0.00 | 0.05 | 0.10 |
| 2010 2011 | 90 89 | 1.22 0.84 | 1.63 1.13 | 2.12 1.46 | 0.48 | 0.65 | 0.83 0.86 | 2010 2011 | | 90 89 | 0.05 0.09 | 0.11 | 0.18 0.29 | 0.03 | 0.08 | 0.13 0.21 | 2010 2011 | | 90 89 | 0.00 | 0.02 | 0.04 0.07 | 0.00 | 0.01 0.04 | 0.03 0.08 |
| 2011 | 84 | 0.84 | 0.32 | 0.47 | 0.48 | 0.66 | 0.86 | 2011 | | 89 84 | 0.09 | 0.19 | 0.29 | 0.00 | 0.13 | 0.21 | 2011 | | 84 | 0.00 | 0.04 | 0.07 | 0.00 | 0.04 | 0.08 |
| 2012 | 89 | 0.28 | 0.44 | 0.63 | 0.07 | 0.14 | 0.21 | 2012 | | 89 | 0.00 | 0.01 | 0.03 | 0.00 | 0.01 | 0.03 | 2013 | | 89 | 0.00 | 0.01 | 0.02 | 0.00 | 0.01 | 0.00 |
| 2014 | 90 | 0.21 | 0.33 | 0.47 | 0.05 | 0.10 | 0.15 | 2014 | | 90 | 0.00 | 0.04 | 0.07 | 0.00 | 0.02 | 0.05 | 2014 | | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2015 | 90 | 0.35 | 0.54 | 0.75 | 0.13 | 0.22 | 0.31 | 2015 | g | 90 | 0.00 | 0.03 | 0.07 | 0.00 | 0.02 | 0.04 | 2015 | | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2016 | 90 | 0.20 | 0.32 | 0.45 | 0.07 | 0.12 | 0.18 | 2016 | g | 90 | 0.00 | 0.02 | 0.05 | 0.00 | 0.02 | 0.04 | 2016 | | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2017 | 90 | 0.24 | 0.40 | 0.58 | 0.08 | 0.15 | 0.21 | 2017 | | 90 | 0.00 | 0.05 | 0.09 | 0.00 | 0.03 | 0.07 | 2017 | | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2018 | 90 | 0.08 | 0.17 | 0.27 | 0.02 | 0.07 | 0.11 | 2018 | | 90 | 0.00 | 0.02 | 0.04 | 0.00 | 0.01 | 0.04 | 2018 | | 90 | 0.00 | 0.01 | 0.02 | 0.00 | 0.01 | 0.04 |
| 2019 | 90 | 3.02 | 3.80 | 4.74 | 0.45 | 0.57 | 0.70 | 2019 | | 90 | 0.01 | 0.05 | 0.09 | 0.00 | 0.03 | 0.06 | 2019 | | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2020 | 90 90 | 0.94 | 1.32 | 1.77 | 0.19 | 0.27 | 0.36 | 2020 | | 90 90 | 0.00 | 0.02 | 0.05 | 0.00 | 0.02 | 0.04 | 2020 | | 90 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2021 2022 | 90 | 1.00 | 1.42 | 1.94 | 0.15 | 0.22 | 0.30 | 2021 2022 | 5 | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2021 2022 | - | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2022 | | | | | | | | 2022 | | | | | | | | | 2022 | | | | | | | | |
| 2024 | | | | | | | | 2024 | | | | | | | | | 2024 | | | | | | | | |
| 2025 | | | | | | | | 2025 | | | | | | | | | 2025 | | | | | | | | |
| 2002 0 | 75 | 1.43 | 2.02 | 2.74 | 0.49 | 0.67 | 0.87 | 2002 | 3 7 | 75 | 0.13 | 0.26 | 0.41 | 0.12 | 0.26 | 0.41 | 2002 | 6 | 75 | 0.02 | 0.08 | 0.14 | 0.03 | 0.09 | 0.17 |
| 2003 | 101 | 0.47 | 0.68 | 0.91 | 0.19 | 0.27 | 0.35 | 2003 | 1 | .01 | 0.04 | 0.08 | 0.13 | 0.04 | 0.09 | 0.15 | 2003 | | 101 | 0.00 | 0.03 | 0.06 | 0.00 | 0.04 | 0.07 |
| 2004 | 92 | 0.91 | 1.30 | 1.76 | 0.25 | 0.35 | 0.46 | 2004 | | 92 | 0.01 | 0.04 | 0.08 | 0.00 | 0.04 | 0.07 | 2004 | | 92 | 0.00 | 0.02 | 0.05 | 0.00 | 0.04 | 0.07 |
| 2005 | 86 | 1.08 | 1.44 | 1.86 | 0.37 | 0.47 | 0.58 | 2005 | | 86 | 0.11 | 0.21 | 0.31 | 0.09 | 0.17 | 0.25 | 2005 | | 86 | 0.00 | 0.02 | 0.05 | 0.00 | 0.02 | 0.06 |
| 2006 | 79 | 2.02 | 2.68 | 3.50 | 0.50 | 0.63 | 0.77 | 2006 | | 79 | 0.09 | 0.17 | 0.26 | 0.08 | 0.14 | 0.21 | 2006 | | 79 | 0.00 | 0.05 | 0.09 | 0.01 | 0.07 | 0.14 |
| 2007 2008 | 44 90 | 1.72 | 2.44 1.91 | 3.35 2.56 | 0.53 | 0.71 | 0.91 0.61 | 2007 2008 | | 44 90 | 0.01 | 0.07 | 0.14 0.22 | 0.01 | 0.06 | 0.12 0.19 | 2007 2008 | | 44 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 0.05 | 0.00 0.10 |
| 2008 | 90 | 0.53 | 0.77 | 1.05 | 0.34 | 0.25 | 0.34 | 2008 | | 90 | 0.00 | 0.13 | 0.22 | 0.03 | 0.09 | 0.15 | 2008 | | 90 | 0.00 | 0.04 | 0.03 | 0.00 | 0.03 | 0.10 |
| 2005 | 90 | 0.73 | 1.04 | 1.40 | 0.24 | 0.33 | 0.43 | 2010 | | 90 | 0.01 | 0.05 | 0.10 | 0.00 | 0.04 | 0.07 | 2010 | | 90 | 0.00 | 0.01 | 0.02 | 0.00 | 0.01 | 0.03 |
| 2011 | 89 | 0.19 | 0.32 | 0.46 | 0.08 | 0.13 | 0.18 | 2011 | | 89 | 0.02 | 0.07 | 0.12 | 0.02 | 0.07 | 0.12 | 2011 | | 89 | 0.00 | 0.01 | 0.02 | 0.00 | 0.01 | 0.03 |
| 2012 | 84 | 0.11 | 0.23 | 0.35 | 0.02 | 0.04 | 0.07 | 2012 | 8 | 84 | 0.00 | 0.01 | 0.04 | 0.00 | 0.02 | 0.04 | 2012 | | 84 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2013 | 89 | 0.17 | 0.31 | 0.47 | 0.03 | 0.06 | 0.09 | 2013 | 8 | 89 | 0.00 | 0.02 | 0.04 | 0.00 | 0.01 | 0.03 | 2013 | | 89 | 0.00 | 0.01 | 0.02 | 0.00 | 0.01 | 0.04 |
| 2014 | 90 | 0.16 | 0.28 | 0.40 | 0.04 | 0.07 | 0.10 | 2014 | | 90 | 0.00 | 0.01 | 0.03 | 0.00 | 0.01 | 0.02 | 2014 | | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2015 | 90 | 0.25 | 0.40 | 0.56 | 0.08 | 0.13 | 0.19 | 2015 | | 90 | 0.00 | 0.02 | 0.04 | 0.00 | 0.01 | 0.02 | 2015 | | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2016 | 90 | 0.10 | 0.19 | 0.28 | 0.03 | 0.06 | 0.09 | 2016 | | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2016 | | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2017 2018 | 90 90 | 0.12 | 0.24 | 0.38 0.21 | 0.03 | 0.06 | 0.10 | 2017 2018 | | 90 90 | 0.00 | 0.00 | 0.00 0.02 | 0.00 | 0.00 | 0.00 | 2017 2018 | _ | 90 90 | 0.00 | 0.00 | 0.00 0.00 | 0.00 | 0.00 | 0.00 0.00 |
| 2018 | 90 | 2.70 | 3.45 | 4.35 | 0.01 | 0.02 | 0.04 | 2018 | | 90 | 0.00 | 0.01 | 0.02 | 0.00 | 0.01 | 0.03 | 2018 | | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2020 | 90 | 0.79 | 1.12 | 1.52 | 0.12 | 0.18 | 0.24 | 2020 | | 90 | 0.00 | 0.02 | 0.05 | 0.00 | 0.02 | 0.04 | 2020 | | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2021 | 90 | 0.85 | 1.25 | 1.74 | 0.10 | 0.16 | 0.23 | 2021 | | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2021 | | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2022 | | | | | | | | 2022 | | | | | | | | | 2022 | | | | | | | | |
| 2023 | | | | | | | | 2023 | | | | | | | | | 2023 | | | | | | | | |
| 2024 | | | | | | | | 2024 | | | | | | | | | 2024 | | | | | | | | |
| 2025 | _ | | | | | | | 2025 | | | | | | | | | 2025 | | | | | | | | |
| 2002 1 | 75 | 0.34 | 0.52 | 0.71 | 0.19 | 0.30 | 0.42 | 2002 | | 75 | 0.06 | 0.12 | 0.18 | 0.05 | 0.11 | 0.17 | 2002 | | | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2003 | 101 | 0.26 | 0.39 | 0.55 | 0.13 | 0.21 | 0.29 | 2003 | | .01 92 | 0.02 | 0.08 | 0.15 | 0.03 | 0.09 | 0.16 | 2003 | | 101 | 0.00 | 0.02 | 0.05 | 0.00 | 0.04 | 0.09 |
| 2004 2005 | 92 86 | 0.25 | 0.37 | 0.50 1.46 | 0.12 | 0.18 | 0.25 0.64 | 2004 2005 | | 92 86 | 0.01 | 0.04 | 0.08 0.10 | 0.01 | 0.04 | 0.08 | 2004 2005 | | 92 86 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 0.00 |
| 2005 | 79 | 0.02 | 0.58 | 0.77 | 0.18 | 0.27 | 0.37 | 2005 | | 79 | 0.01 | 0.00 | 0.26 | 0.01 | 0.03 | 0.10 | 2005 | | 79 | 0.00 | 0.03 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2007 | 44 | 0.31 | 0.55 | 0.83 | 0.12 | 0.23 | 0.35 | 2007 | | 44 | 0.00 | 0.03 | 0.08 | 0.00 | 0.04 | 0.11 | 2007 | | 44 | 0.00 | 0.01 | 0.04 | 0.00 | 0.02 | 0.05 |
| 2008 | 90 | 0.27 | 0.44 | 0.64 | 0.16 | 0.28 | 0.40 | 2008 | 9 | 90 | 0.01 | 0.04 | 0.08 | 0.01 | 0.05 | 0.09 | 2008 | | 90 | 0.00 | 0.01 | 0.02 | 0.00 | 0.01 | 0.01 |
| 2009 | 90 | 0.34 | 0.51 | 0.71 | 0.15 | 0.23 | 0.31 | 2009 | 9 | 90 | 0.01 | 0.04 | 0.08 | 0.00 | 0.03 | 0.06 | 2009 | | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2010 | 90 | 0.15 | 0.26 | 0.38 | 0.08 | 0.14 | 0.21 | 2010 | | 90 | 0.03 | 0.08 | 0.14 | 0.03 | 0.08 | 0.13 | 2010 | | 90 | 0.00 | 0.01 | 0.02 | 0.00 | 0.01 | 0.02 |
| 2011 | 89 | 0.18 | 0.32 | 0.48 | 0.09 | 0.18 | 0.28 | 2011 | | 89 | 0.00 | 0.06 | 0.12 | 0.00 | 0.06 | 0.12 | 2011 | | 89 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2012 | 84 | 0.00 | 0.03 | 0.07 | 0.00 | 0.03 | 0.05 | 2012 | | 84 | 0.00 | 0.01 | 0.03 | 0.00 | 0.01 | 0.02 | 2012 | | 84 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2013 2014 | 89 90 | 0.03 | 0.08 | 0.14 | 0.01 | 0.03 | 0.04 | 2013 2014 | | 89 90 | 0.00 | 0.01 | 0.02 | 0.00 | 0.01 | 0.04 | 2013 2014 | | 89 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 0.00 |
| 2014 2015 | 90 90 | 0.00 | 0.02 | 0.04 0.18 | 0.00 | 0.01 | 0.03 | 2014 | | 90 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2014 | | 90 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2015 | 90 | 0.03 | 0.11 | 0.18 | 0.02 | 0.00 | 0.09 | 2015 | | 90 | 0.00 | 0.02 | 0.03 | 0.00 | 0.02 | 0.04 | 2015 | | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2010 | 90 | 0.04 | 0.09 | 0.10 | 0.02 | 0.04 | 0.08 | 2010 | | 90 | 0.00 | 0.01 | 0.03 | 0.00 | 0.01 | 0.02 | 2010 | | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2018 | 90 | 0.00 | 0.01 | 0.03 | 0.00 | 0.00 | 0.01 | 2018 | | 90 | 0.00 | 0.01 | 0.02 | 0.00 | 0.01 | 0.03 | 2018 | | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2019 | 90 | 0.11 | 0.19 | 0.27 | 0.04 | 0.08 | 0.12 | 2019 | | 90 | 0.00 | 0.01 | 0.02 | 0.00 | 0.01 | 0.03 | 2019 | | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2020 | 90 | 0.12 | 0.20 | 0.29 | 0.03 | 0.06 | 0.09 | 2020 | 9 | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2020 | | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2021 | 90 | 0.06 | 0.13 | 0.20 | 0.02 | 0.06 | 0.09 | 2021 | | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | | 2021 | | 90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2022 | | | | | | | | 2022 | | | | | | | | | 2022 | | | | | | | | |
| 2023 | | | | | | | | 2023 | | | | | | | | | 2023 | | | | | | | | |
| 2024 | | | | | | | | 2024 | | | | | | | | | 2024 | - | | | | | | | |
| 2025 | | | | | | | | 2025 | | | | | | | | | 2025 | | | | | | | | |

Figure 52. Summer Flounder geometric mean indices of abundance, by number and biomass, for all ages combined and by age-class.



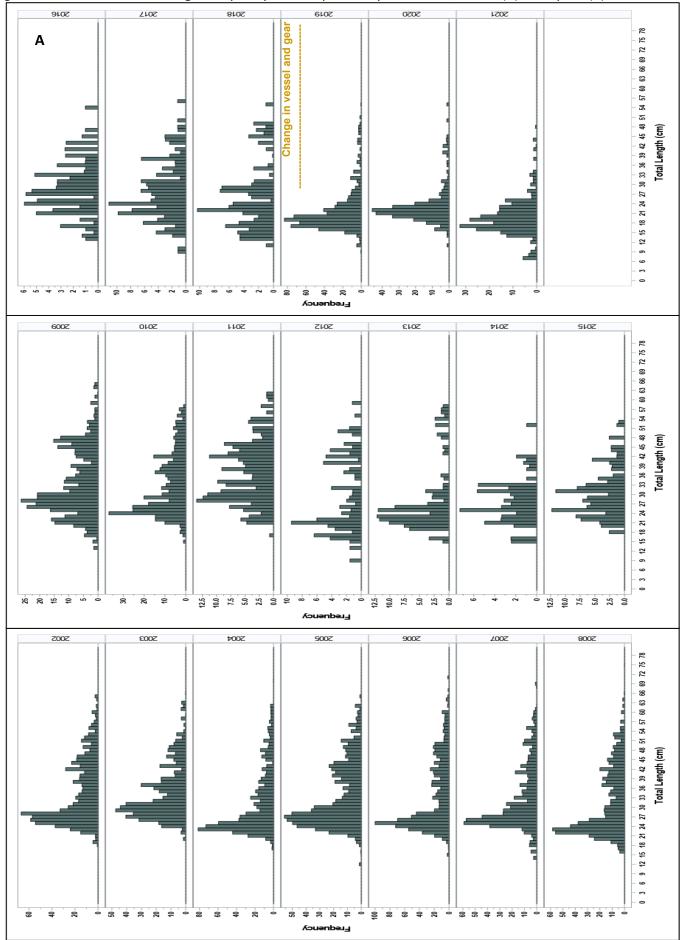
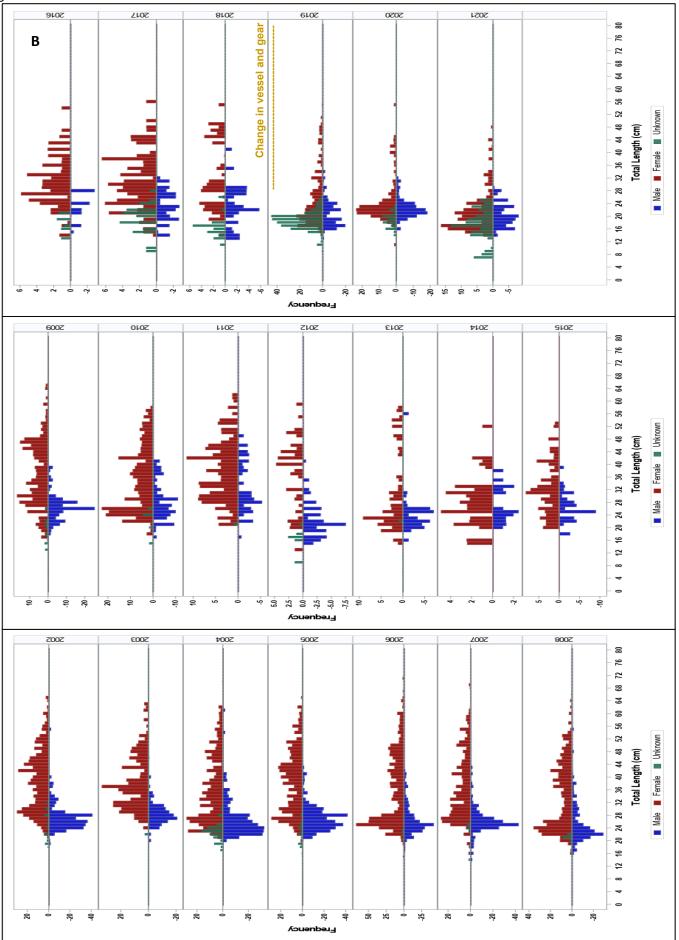


Figure 53. Summer Flounder length-frequency in Chesapeake Bay 2002-2021, overall (A) and by sex (B).

Figure 53. cont.



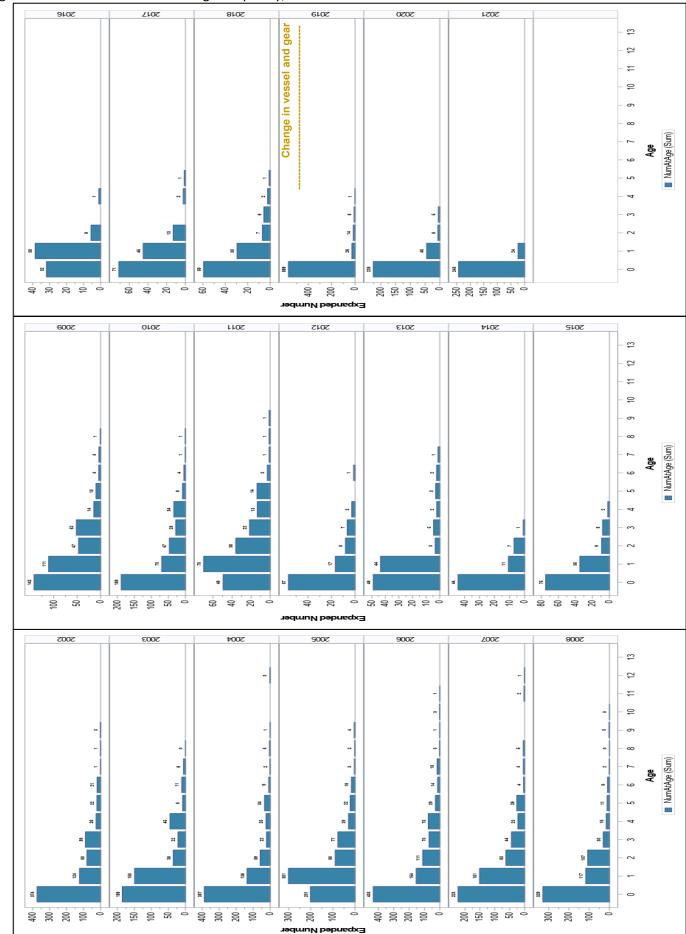


Figure 54. Summer Flounder total age-frequency, 2002-2021.

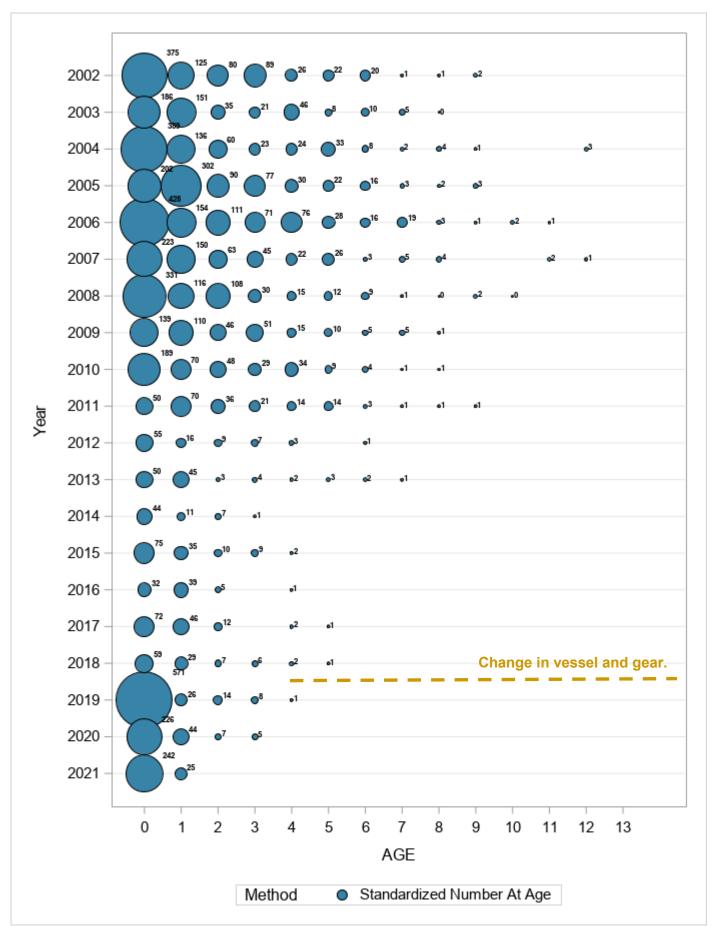


Figure 55. Summer Flounder age-frequency by year, 2002-2021 standardized to 8,000 (4,800 for 2019 and forward) annual trawl minutes.

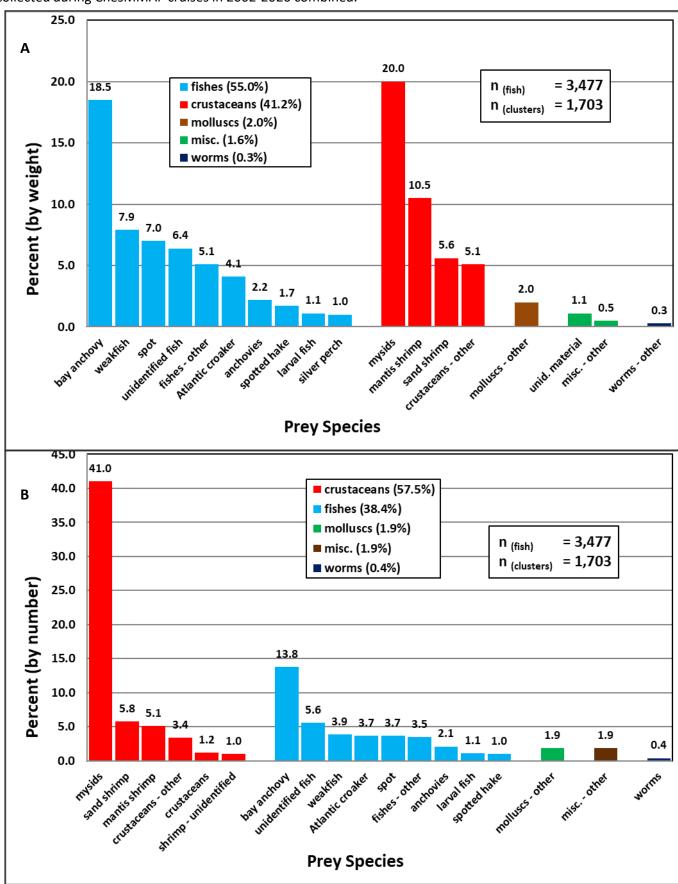
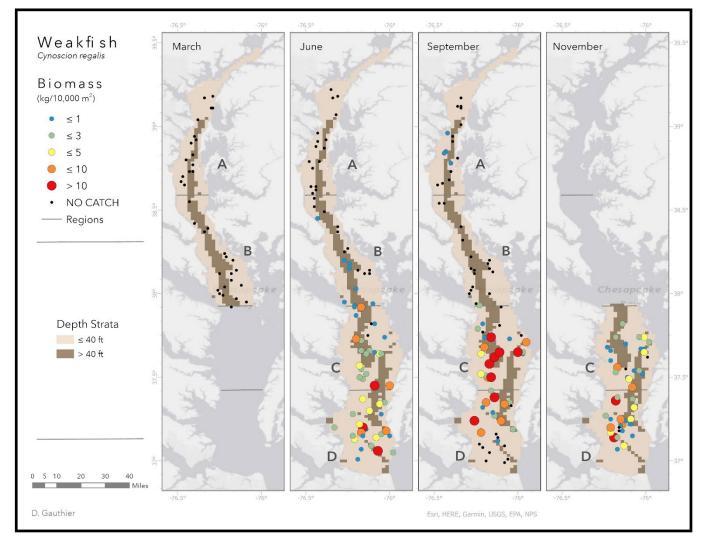


Figure 56. Diet composition, expressed as percent by weight (A) and percent by number (B) of Summer Flounder collected during ChesMMAP cruises in 2002-2020 combined.

| Year | Number Caught | Biomass Caught (kg) | Presence at Index Stations (%) | Number Measured | Age Specimens | Ages Read | Stomach Specimens | Stomachs Analyzed |
|------|------------------|------------------------|-----------------------------------|--------------------|------------------|--------------|----------------------|----------------------|
| 2002 | 1,734 | 304.7 | 47.5 | 1,692 | 803 | 803 | 607 | 583 |
| 2003 | 2,315 | 400.0 | 58.0 | 2,198 | 707 | 707 | 654 | 642 |
| 2004 | 3,851 | 561.9 | 69.5 | 3,551 | 1,108 | 1,108 | 901 | 889 |
| 2005 | 2,715 | 378.5 | 65.6 | 2,711 | 1,119 | 1,119 | 918 | 906 |
| 2006 | 1,476 | 159.5 | 60.8 | 1,462 | 728 | 728 | 561 | 554 |
| 2007 | 1,214 | 128.0 | 55.7 | 1,210 | 554 | 554 | 439 | 435 |
| 2008 | 812 | 83.8 | 42.2 | 812 | 368 | 368 | 330 | 324 |
| 2009 | 873 | 46.2 | 60.0 | 873 | 478 | 478 | 387 | 384 |
| 2010 | 1,207 | 76.8 | 60.7 | 1,207 | 607 | 607 | 542 | 531 |
| 2011 | 918 | 57.5 | 55.2 | 918 | 454 | 454 | 323 | 322 |
| 2012 | 886 | 72.2 | 35.7 | 886 | 328 | 328 | 260 | 256 |
| 2013 | 301 | 42.0 | 28.4 | 301 | 187 | 187 | 130 | 129 |
| 2014 | 172 | 8.6 | 23.0 | 172 | 126 | 126 | 72 | 72 |
| 2015 | 688 | 51.9 | 26.7 | 688 | 285 | 285 | 141 | 140 |
| 2016 | 1,115 | 91.2 | 38.5 | 1,115 | 281 | 281 | 143 | 141 |
| 2017 | 943 | 68.3 | 36.3 | 943 | 335 | 335 | 194 | 191 |
| 2018 | 1,621 | 61.5 | 43.7 | 1,621 | 273 | 273 | 173 | 171 |
| 2019 | 18,987 | 1,327.2 | 80.7 | 11,355 | 661 | 661 | 387 | 382 |
| 2020 | 23,685 | 1,305.0 | 90.4 | 10,855 | 372 | 372 | 171 | 168 |
| 2021 | 16,901 | 1,044.4 | 85.9 | 9,794 | 467 | 467 | 277 | 274 |

Table 27. Weakfish sampling rates and preserved specimen analysis status by year (shaded rows represent a change in vessel and trawl gear beginning in June 2019).

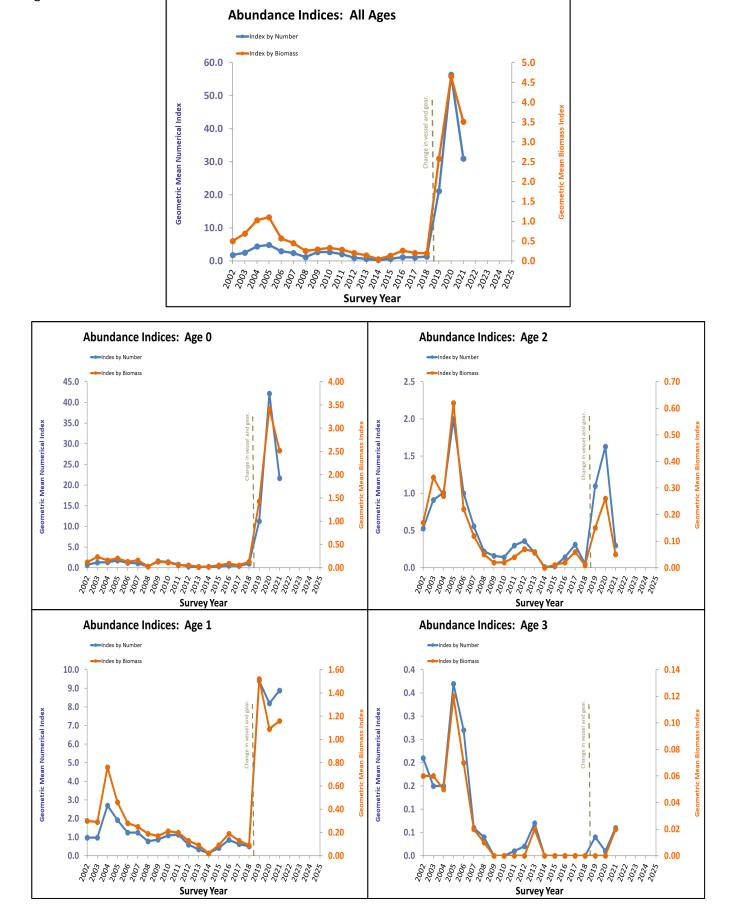
Figure 57. Station specific biomass of Weakfish in Chesapeake Bay, 2021.

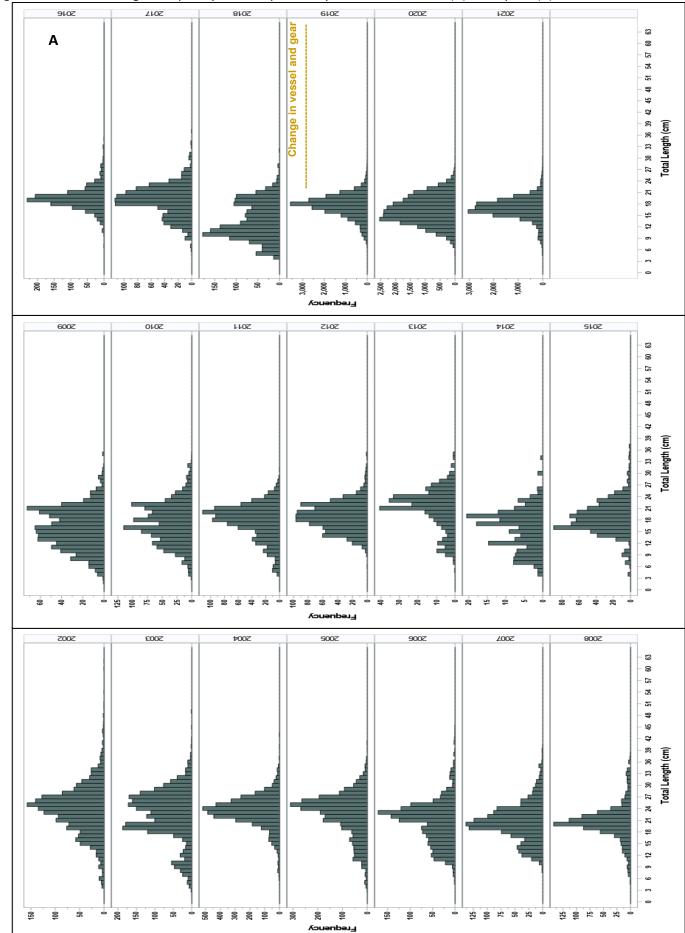


| Table 28. Weakfish geometric mear | indices of abundance, | by number and biomass | overall and by age-class. |
|-----------------------------------|---|-----------------------|---------------------------|
| | , | | |

| | | _ | | - | | | | | | | | | biomas | | | | |
|--------------|-----|------------|----------------|---------------------|----------------|--------------|--------------|--------------|------------------|-----|------------|--------------|---------------------|--------------|------|-------|------|
| Year | Age | n | LCI | merical In Index | dex UCI | LCI | Index | uci UCI | Year | Age | n | LCI | merical In Index | dex UCI | LCI | Index | UCI |
| 2002 | All | 122 | 1.26 | 1.81 | 2.50 | 0.34 | 0.50 | 0.69 | 2002 | 2 | 122 | 0.34 | 0.53 | 0.75 | 0.10 | 0.17 | 0.24 |
| 2003 | 7 | 149 | 1.82 | 2.53 | 3.41 | 0.50 | 0.69 | 0.91 | 2003 | - | 149 | 0.64 | 0.91 | 1.22 | 0.23 | 0.34 | 0.46 |
| 2004 | | 127 | 3.12 | 4.39 | 6.05 | 0.75 | 1.03 | 1.36 | 2004 | | 127 | 0.73 | 1.01 | 1.33 | 0.19 | 0.27 | 0.36 |
| 2005 | | 131 | 3.49 | 4.87 | 6.67 | 0.83 | 1.10 | 1.41 | 2005 | | 131 | 1.45 | 2.00 | 2.68 | 0.45 | 0.62 | 0.80 |
| 2006 | | 120 | 2.06 | 2.99 | 4.21 | 0.40 | 0.57 | 0.75 | 2006 | | 120 | 0.69 | 1.00 | 1.37 | 0.15 | 0.22 | 0.29 |
| 2007 | | 88 | 1.67 | 2.41 | 3.35 | 0.31 | 0.45 | 0.60 | 2007 | | 88 | 0.37 | 0.56 | 0.77 | 0.07 | 0.12 | 0.17 |
| 2008 | | 135 | 0.80 | 1.18 | 1.64 | 0.16 | 0.25 | 0.34 | 2008 | | 135 | 0.13 | 0.22 | 0.32 | 0.03 | 0.05 | 0.08 |
| 2009 2010 | | 135 135 | 1.92 1.99 | 2.66 2.65 | 3.60 3.46 | 0.21 | 0.29 | 0.38 0.41 | 2009 2010 | | 135 135 | 0.08 0.08 | 0.16 | 0.23 0.21 | 0.01 | 0.02 | 0.04 |
| 2010 | | 133 | 1.99 | 2.03 | 2.72 | 0.24 | 0.33 | 0.41 | 2010 | | 133 | 0.08 | 0.14 | 0.21 | 0.01 | 0.02 | 0.03 |
| 2011 | | 129 | 0.59 | 0.94 | 1.38 | 0.13 | 0.20 | 0.29 | 2011 | | 129 | 0.20 | 0.36 | 0.54 | 0.03 | 0.07 | 0.10 |
| 2013 | | 134 | 0.39 | 0.63 | 0.91 | 0.07 | 0.14 | 0.22 | 2013 | | 134 | 0.10 | 0.20 | 0.32 | 0.02 | 0.06 | 0.10 |
| 2014 | | 135 | 0.20 | 0.34 | 0.50 | 0.02 | 0.04 | 0.06 | 2014 | | 135 | 0.00 | 0.01 | 0.03 | 0.00 | 0.00 | 0.00 |
| 2015 | | 135 | 0.40 | 0.64 | 0.92 | 0.08 | 0.13 | 0.18 | 2015 | | 135 | 0.00 | 0.02 | 0.04 | 0.00 | 0.01 | 0.01 |
| 2016 | | 135 | 0.81 | 1.19 | 1.66 | 0.17 | 0.26 | 0.35 | 2016 | | 135 | 0.08 | 0.14 | 0.21 | 0.01 | 0.02 | 0.04 |
| 2017 | | 135 | 0.70 | 1.07 | 1.52 | 0.12 | 0.20 | 0.28 | 2017 | | 135 | 0.18 | 0.31 | 0.45 | 0.03 | 0.06 | 0.09 |
| 2018 | | 135 | 0.85 | 1.32 | 1.93 | 0.11 | 0.19 | 0.28 | 2018 | | 135 | 0.02 | 0.05 | 0.09 | 0.00 | 0.01 | 0.02 |
| 2019 | | 135 | 14.42 | 21.14 | 30.79 | 1.88 | 2.58 | 3.45 | 2019 | | 135 | 0.81 | 1.10 | 1.43 | 0.11 | 0.15 | 0.20 |
| 2020 2021 | | 135 135 | 40.51 21.78 | 56.41 30.99 | 78.41 43.93 | 3.68 2.72 | 4.65 3.51 | 5.82 4.46 | 2020 2021 | | 135 135 | 1.22 0.22 | 1.63 0.30 | 2.10 0.39 | 0.19 | 0.26 | 0.33 |
| 2021 | | 133 | 21.70 | 50.99 | 45.55 | 2.72 | 5.51 | 4.40 | 2021 | | 133 | 0.22 | 0.50 | 0.39 | 0.05 | 0.05 | 0.07 |
| 2022 | | | | | | | | | 2022 | | | | | | | | |
| 2024 | | | | | | | | | 2024 | | | | | | | | |
| 2025 | | | | | | | | | 2025 | | | | | | | | |
| 2002 | 0 | 122 | 0.49 | 0.72 | 1.00 | 0.08 | 0.12 | 0.17 | 2002 | 3 | 122 | 0.12 | 0.21 | 0.31 | 0.03 | 0.06 | 0.09 |
| 2003 | | 149 | 0.92 | 1.29 | 1.73 | 0.16 | 0.23 | 0.31 | 2003 | | 149 | 0.09 | 0.15 | 0.21 | 0.03 | 0.06 | 0.08 |
| 2004 | | 127 | 0.97 | 1.31 | 1.70 | 0.12 | 0.16 | 0.20 | 2004 | | 127 | 0.08 | 0.15 | 0.22 | 0.02 | 0.05 | 0.07 |
| 2005 | | 131 | 1.23 | 1.74 | 2.36 | 0.15 | 0.20 | 0.26 | 2005 | | 131 | 0.26 | 0.37 | 0.49 | 0.08 | 0.12 | 0.15 |
| 2006 | | 120 88 | 0.78 | 1.20 | 1.71 | 0.09 | 0.13 | 0.18 | 2006 | | 120 | 0.17 | 0.27 | 0.39 | 0.04 | 0.07 | 0.10 |
| 2007 2008 | | 88 135 | 0.65 0.20 | 1.05 0.33 | 1.54 0.47 | 0.10 | 0.16 | 0.23 | 2007 2008 | | 88 135 | 0.01 | 0.06 | 0.10 | 0.00 | 0.02 | 0.03 |
| 2008 | | 135 | 1.15 | 1.64 | 2.25 | 0.02 | 0.13 | 0.04 | 2008 | | 135 | 0.01 | 0.04 | 0.00 | 0.00 | 0.00 | 0.02 |
| 2010 | | 135 | 1.02 | 1.43 | 1.93 | 0.08 | 0.11 | 0.15 | 2010 | | 135 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2011 | | 134 | 0.56 | 0.81 | 1.10 | 0.04 | 0.06 | 0.08 | 2011 | | 134 | 0.00 | 0.01 | 0.02 | 0.00 | 0.00 | 0.00 |
| 2012 | | 129 | 0.19 | 0.35 | 0.52 | 0.02 | 0.05 | 0.07 | 2012 | | 129 | 0.00 | 0.02 | 0.04 | 0.00 | 0.00 | 0.01 |
| 2013 | | 134 | 0.09 | 0.20 | 0.31 | 0.01 | 0.02 | 0.02 | 2013 | | 134 | 0.02 | 0.07 | 0.13 | 0.00 | 0.02 | 0.05 |
| 2014 | | 135 | 0.12 | 0.24 | 0.36 | 0.01 | 0.02 | 0.03 | 2014 | | 135 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2015 | | 135 | 0.20 | 0.36 | 0.55 | 0.02 | 0.05 | 0.07 | 2015 | | 135 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2016 | | 135 | 0.36 | 0.55 | 0.78 | 0.05 | 0.09 | 0.13 | 2016 | | 135 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 |
| 2017 | | 135 135 | 0.30 | 0.46 | 0.65 | 0.03 | 0.05 | 0.07 | 2017 | | 135 135 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2018 2019 | | 135 | 0.60 | 0.98 | 1.45 16.20 | 0.06 | 0.13 | 0.19 | 2018 2019 | | 135 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2020 | | 135 | 30.10 | 42.15 | 58.87 | 2.72 | 3.42 | 4.27 | 2015 | | 135 | 0.00 | 0.01 | 0.07 | 0.00 | 0.00 | 0.01 |
| 2021 | | 135 | 14.96 | 21.64 | 31.12 | 1.93 | 2.52 | 3.23 | 2021 | | 135 | 0.02 | 0.06 | 0.10 | 0.01 | 0.02 | 0.04 |
| 2022 | | | | | | | | | 2022 | | | | | | | | |
| 2023 | | | | | | | | | 2023 | | | | | | | | |
| 2024 | | | | | | | | | 2024 | | | | | | | | |
| 2025 | | | | | | | | | 2025 | | | | | | | | |
| 2002 | 1 | 122 | 0.65 | 0.97 | 1.35 | 0.19 | 0.30 | 0.42 | | | | | | | ļ | | |
| 2003 | | 149 | 0.71 | 0.98 | 1.31 | 0.20 | 0.29 | 0.38 | | | | | | | ļ | | |
| 2004 2005 | | 127 | 1.86 | 2.70 | 3.80 | 0.53 | 0.76 | 1.02 0.59 | | | | | | | ļ | | |
| 2005 | | 131 120 | 1.40 0.84 | 1.92 1.25 | 2.56 1.76 | 0.34 0.19 | 0.46 | 0.59 | | | | | | | | | |
| 2000 | | 88 | 0.84 | 1.25 | 1.75 | 0.19 | 0.25 | 0.38 | | | | | | | | | |
| 2008 | | 135 | 0.49 | 0.77 | 1.11 | 0.11 | 0.19 | 0.27 | | | | | | | | | |
| 2009 | | 135 | 0.57 | 0.86 | 1.20 | 0.11 | 0.17 | 0.23 | | | | | | | | | |
| 2010 | | 135 | 0.79 | 1.09 | 1.45 | 0.15 | 0.21 | 0.28 | | | | | | | | | |
| 2011 | | 134 | 0.78 | 1.14 | 1.57 | 0.13 | 0.20 | 0.27 | | | | | | | | | |
| 2012 | | 129 | 0.34 | 0.59 | 0.88 | 0.07 | 0.13 | 0.19 | | | | | | | | | |
| 2013 | | 134 | 0.19 | 0.34 | 0.52 | 0.03 | 0.09 | 0.15 | | | | | | | | | |
| 2014 | | 135 | 0.05 | 0.13 | 0.21 | 0.01 | 0.02 | 0.03 | | | | | | | | - | |
| 2015 2016 | | 135 135 | 0.24 | 0.41 0.85 | 0.60 | 0.05 | 0.09 | 0.13 | | | | | | | | | |
| 2016 | | 135 | 0.56 0.38 | 0.85 | 0.91 | 0.12 | 0.19 | 0.27 | | | | | | | | | |
| 2017 | | 135 | 0.38 | 0.51 | 0.91 | 0.07 | 0.13 | 0.18 | | | | | | | | | |
| 2010 | | 135 | 6.38 | 9.41 | 13.69 | 1.10 | 1.52 | 2.03 | | | | | | | | | |
| 2020 | | 135 | 6.01 | 8.19 | 11.07 | 0.82 | 1.09 | 1.39 | | | | | | | | | |
| 2021 | | 135 | 6.70 | 8.89 | 11.70 | 0.93 | 1.16 | 1.42 | | | | | | | | | |
| 2022 | | | | | | | | | | | | | | | | | |
| 2023 | | | | | | | | | | | | | | | | | |
| 2024 | | | | | | | | | | | | | | | | | |
| 2025 | | | | | | | | | | | | | | | | | |

Figure 58. Weakfish geometric mean indices of abundance, by number and biomass, for all ages combined and by age-class.





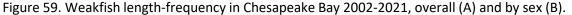


Figure 59. cont.

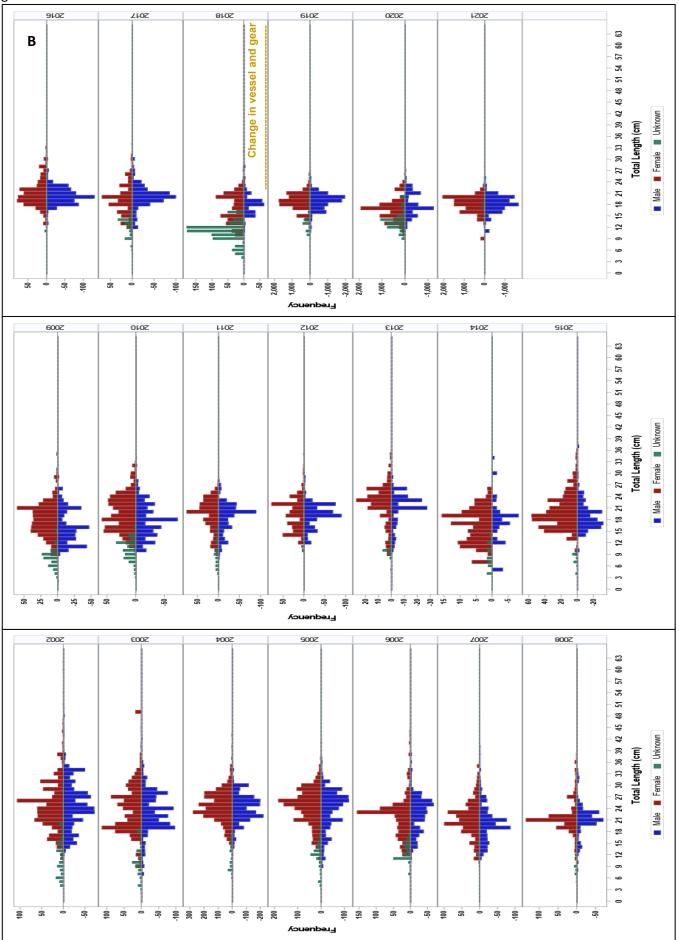
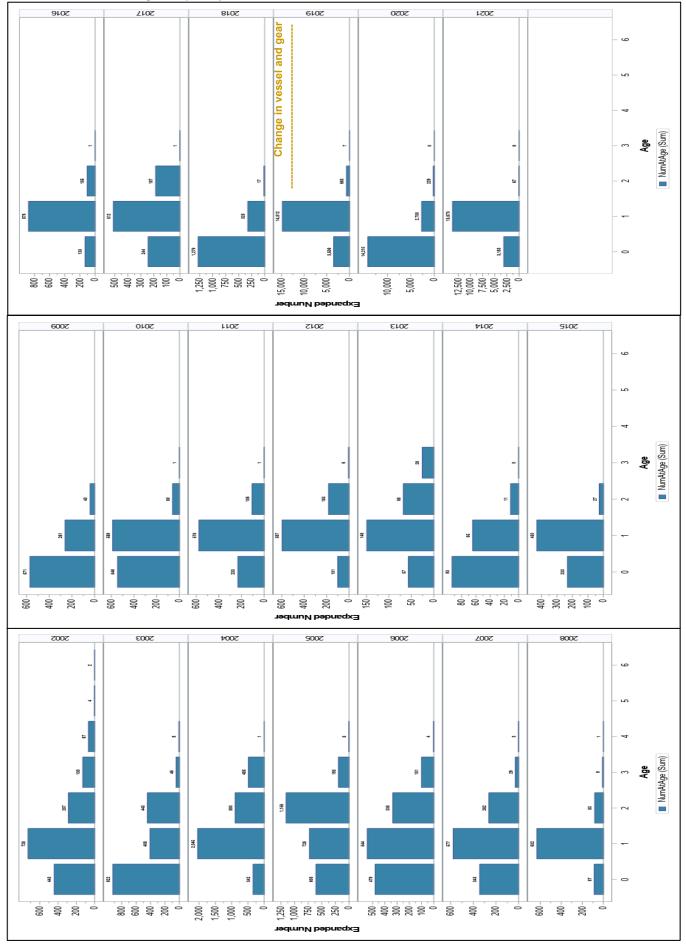


Figure 60. Weakfish total age-frequency, 2002-2021.



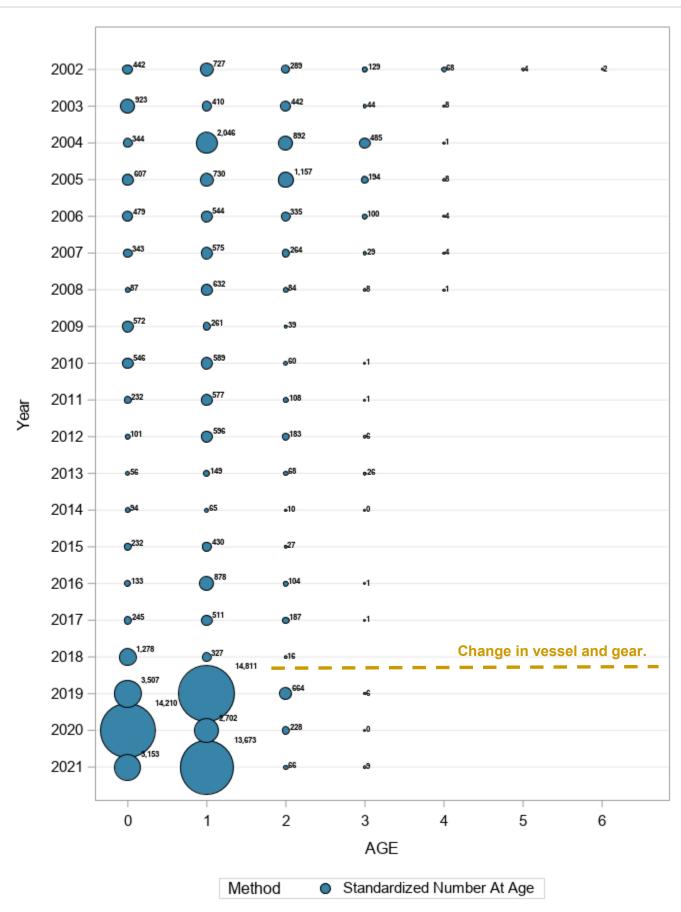


Figure 61. Weakfish age-frequency by year, 2002-2021 standardized to 8,000 (4,800 for 2019 and forward) annual trawl minutes.

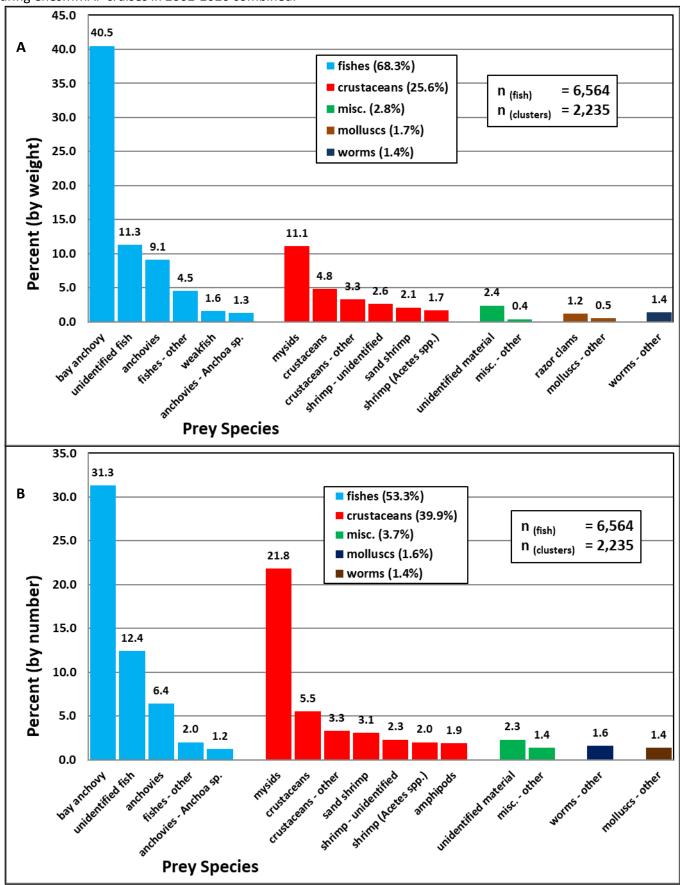
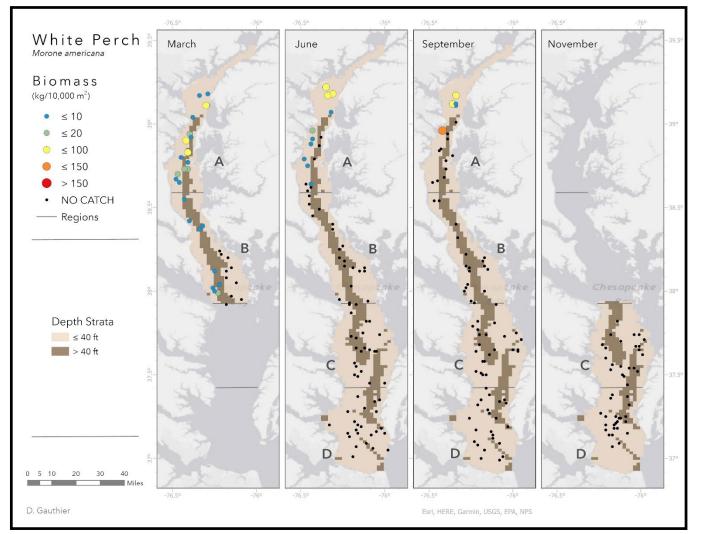


Figure 62. Diet composition, expressed as percent by weight (A) and percent by number (B) of Weakfish collected during ChesMMAP cruises in 2002-2020 combined.

| Year | Number Caught | Biomass Caught (kg) | Presence at Index Stations (%) | Number Measured | Age Specimens | Ages Read | Stomach Specimens | Stomachs Analyzed |
|------|------------------|------------------------|--------------------------------------|--------------------|------------------|-----------|----------------------|----------------------|
| 2002 | 6,625 | 996.6 | 50.0 | 4,020 | 552 | 552 | 471 | 401 |
| 2003 | 3,782 | 511.5 | 53.8 | 1,882 | 177 | 168 | 147 | 126 |
| 2004 | 11,021 | 1,727.4 | 66.7 | 6,677 | 356 | 356 | 270 | 267 |
| 2005 | 7,243 | 843.6 | 60.0 | 5,884 | 429 | 429 | 287 | 280 |
| 2006 | 11,980 | 1,611.0 | 60.7 | 5,899 | 385 | 385 | 263 | 254 |
| 2007 | 4,915 | 517.9 | 62.8 | 3,194 | 318 | 318 | 277 | 277 |
| 2008 | 2,924 | 340.1 | 52.5 | 2,360 | 260 | 257 | 227 | 224 |
| 2009 | 5,130 | 686.2 | 47.5 | 1,749 | 158 | 151 | 126 | 126 |
| 2010 | 2,996 | 453.6 | 50.8 | 1,627 | 207 | 207 | 158 | 156 |
| 2011 | 4,619 | 675.1 | 45.8 | 2,392 | 231 | 231 | 177 | 174 |
| 2012 | 3,737 | 459.9 | 58.1 | 2,423 | 151 | 151 | 111 | 109 |
| 2013 | 3,249 | 421.1 | 59.0 | 2,469 | 199 | 199 | 109 | 55 |
| 2014 | 3,208 | 341.6 | 55.7 | 1,844 | 153 | 153 | 94 | 92 |
| 2015 | 13,708 | 2,157.4 | 44.3 | 4,098 | 188 | 188 | 80 | 81 |
| 2016 | 7,165 | 979.5 | 55.7 | 2,935 | 208 | 208 | 104 | 103 |
| 2017 | 7,957 | 1,113.9 | 51.7 | 4,517 | 159 | 159 | 84 | 81 |
| 2018 | 3,777 | 522.7 | 75.0 | 2,131 | 102 | 102 | 47 | 46 |
| 2019 | 9,870 | 888.5 | 38.8 | 3,367 | 129 | 129 | 80 | 80 |
| 2020 | 15,945 | 1,580.0 | 40.0 | 3,128 | 93 | 93 | 43 | 42 |
| 2021 | 11,614 | 986.5 | 42.0 | 4,298 | 129 | 129 | 69 | 69 |

Table 29. White Perch sampling rates and preserved specimen analysis status by year (shaded rows represent a change in vessel and trawl gear beginning in June 2019).

Figure 63. Station specific biomass of White Perch in Chesapeake Bay, 2021.



| age o | | | | | | | | | | | _ | | | | | | |
|--------------|-----|----------|--------------|--------------|--------------|--------------|------------------|--------------|--------------|-----|----------|------|----------------------|--------------|------|--------------|--------------|
| Year | Age | n | LCI | nerical In | dex UCI | Bio LCI | mass In Index | UCI | Year | Age | n | LCI | nerical Inc Index | dex UCI | LCI | mass Inc | dex UCI |
| 2002 | All | 23 | 0.87 | 1.63 | 2.70 | 0.37 | 0.79 | 1.34 | 2002 | 7 | 23 | 0.29 | 0.72 | 1.30 | 0.07 | 0.29 | 0.56 |
| 2003 | | 27 | 1.20 | 2.52 | 4.63 | 0.51 | 1.05 | 1.78 | 2003 | | 27 | 0.31 | 0.72 | 1.24 | 0.09 | 0.22 | 0.37 |
| 2004 2005 | | 33 28 | 2.97 1.44 | 4.79 3.87 | 7.47 8.71 | 1.07 0.51 | 1.80 1.46 | | 2004 2005 | | 33 28 | 0.15 | 0.47 | 0.89 | 0.03 | 0.16 | 0.31 |
| 2005 | | 28 | 1.60 | 4.35 | 10.03 | 0.31 | 1.40 | 3.18 | 2005 | | 28 | 0.13 | 0.82 | 2.08 | 0.04 | 0.22 | 0.42 |
| 2007 | | 28 | 6.51 | 14.11 | 29.39 | 1.65 | 3.32 | 6.03 | 2007 | | 28 | 1.25 | 2.73 | 5.17 | 0.31 | 0.70 | 1.20 |
| 2008 | | 29 | 5.69 | 9.58 | 15.72 | 1.40 | 2.29 | 3.51 | 2008 | | 29 | 1.56 | 2.61 | 4.09 | 0.30 | 0.59 | 0.93 |
| 2009 2010 | | 29 28 | 1.30 1.25 | 1.91 2.43 | 2.67 4.23 | 0.37 0.37 | 0.63 | 0.93 | 2009 2010 | | 29 28 | 0.09 | 0.34 | 0.65 | 0.03 | 0.09 | 0.16 |
| 2010 | | 29 | 1.32 | 2.43 | 4.23 | 0.48 | 1.09 | 1.94 | 2010 | | 29 | 0.80 | 1.47 | 2.39 | 0.25 | 0.53 | 0.87 |
| 2012 | | | | | | | | | 2012 | | | | | | | | |
| 2013 | | 29 | 2.49 | 6.94 | 17.05 | 1.00 | 2.45 | 4.94 | 2013 | | 29 | 0.38 | 1.08 | 2.13 | 0.13 | 0.38 | 0.68 |
| 2014 2015 | | 29 29 | 3.23 0.45 | 4.96 1.24 | 7.40 2.46 | 0.88 0.16 | 1.37 0.48 | 2.00 0.89 | 2014 2015 | | 29 29 | 0.70 | 1.13 0.00 | 1.67 0.00 | 0.14 | 0.27 | 0.42 |
| 2016 | | 29 | 7.29 | 10.84 | 15.91 | 2.46 | 3.74 | 5.49 | 2016 | | 29 | 1.69 | 2.60 | 3.81 | 0.50 | 0.85 | 1.28 |
| 2017 | | 28 | 5.62 | 10.62 | 19.43 | 2.09 | 3.94 | 6.90 | 2017 | | 28 | 1.93 | 3.80 | 6.88 | 0.68 | 1.39 | 2.41 |
| 2018 2019 | | 25 | 1.47 | 2.81 | 4.89 | 0.43 | 0.84 | 1.37 | 2018 2019 | | 25 | 0.01 | 0.14 | 0.29 | 0.00 | 0.03 | 0.05 |
| 2019 | | 25 | 11.46 | 23.17 | 45.90 | 3.40 | 6.07 | 10.37 | 2019 | | 25 | 0.92 | 1.91 | 3.41 | 0.18 | 0.48 | 0.84 |
| 2021 | | 25 | 8.72 | 20.06 | 44.67 | 2.65 | 5.26 | | 2021 | | 25 | 3.07 | 5.70 | 10.02 | 0.72 | 1.32 | 2.13 |
| 2022 | | | | | | | | | 2022 | | | | | | | | |
| 2023 2024 | | | | | | | | | 2023 2024 | | | | | | | | |
| 2024 | | | | | | | | | 2024 | | | | | | | | |
| 2002 | 4 | 23 | 0.12 | 0.41 | 0.77 | 0.02 | 0.14 | | 2002 | 8 | 23 | 0.05 | 0.25 | 0.50 | 0.00 | 0.08 | 0.18 |
| 2003 | | 27 | 0.07 | 0.32 | 0.63 | 0.02 | 0.10 | | 2003 | | 27 | 0.33 | 0.77 | 1.35 | 0.12 | 0.30 | 0.52 |
| 2004 2005 | | 33 28 | 0.41 0.34 | 1.06 1.26 | 2.02 2.82 | 0.13 | 0.44 | 0.82 | 2004 2005 | | 33 28 | 0.98 | 1.63 0.69 | 2.49 1.40 | 0.36 | 0.68 | 1.07 0.44 |
| 2005 | | 28 | 0.34 | 0.76 | 2.82 | 0.12 | 0.43 | | 2005 | | 28 | 0.20 | 0.85 | 2.07 | 0.08 | 0.33 | 0.44 |
| 2007 | | 28 | 2.92 | 6.97 | 15.22 | 0.70 | 1.58 | 2.91 | 2007 | | 28 | 1.07 | 1.80 | 2.77 | 0.24 | 0.41 | 0.60 |
| 2008 | | 29 | 0.78 | 1.52 | 2.55 | 0.14 | 0.32 | 0.51 | 2008 | | 29 | 1.11 | 1.98 | 3.20 | 0.22 | 0.50 | 0.84 |
| 2009 2010 | | 29 28 | 0.09 | 0.27 | 0.48 | 0.00 | 0.07 | 0.15 | 2009 2010 | | 29 28 | 0.43 | 0.62 | 0.84 | 0.11 | 0.18 0.20 | 0.25 |
| 2010 | | 28 | 0.00 | 0.48 | 1.21 | 0.00 | 0.16 | 0.38 | 2010 | | 28 29 | 0.13 | 0.80 | 1.25 | 0.03 | 0.20 | 0.41 |
| 2012 | | | | | | | | | 2012 | | | | | | | | |
| 2013 | | 29 | 0.55 | 1.52 | 3.12 | 0.20 | 0.55 | 1.01 | 2013 | | 29 | 0.59 | 1.56 | 3.14 | 0.21 | 0.55 | 0.99 |
| 2014 2015 | | 29 29 | 1.23 0.30 | 1.84 0.82 | 2.63 1.56 | 0.27 0.09 | 0.50 | 0.77 | 2014 2015 | | 29 29 | 0.18 | 0.28 | 0.38 | 0.03 | 0.05 0.10 | 0.07 0.19 |
| 2016 | | 29 | 0.48 | 0.91 | 1.47 | 0.13 | 0.29 | 0.48 | 2016 | | 29 | 0.76 | 1.12 | 1.56 | 0.17 | 0.29 | 0.43 |
| 2017 | | 28 | 0.25 | 0.56 | 0.94 | 0.04 | 0.13 | 0.23 | 2017 | | 28 | 1.68 | 3.23 | 5.66 | 0.53 | 1.09 | 1.86 |
| 2018 | | 25 | 0.34 | 0.77 | 1.34 | 0.07 | 0.19 | 0.33 | 2018 | | 25 | 0.48 | 1.03 | 1.79 | 0.13 | 0.31 | 0.51 |
| 2019 2020 | | 25 | 1.25 | 3.21 | 6.88 | 0.37 | 0.19 | 1.73 | 2019 2020 | | 25 | 1.60 | 3.01 | 5.18 | 0.13 | 0.31 | 1.14 |
| 2021 | | 25 | 1.63 | 3.40 | 6.38 | 0.42 | 0.84 | 1.37 | 2021 | | 25 | 0.31 | 0.73 | 1.28 | 0.05 | 0.13 | 0.21 |
| 2022 | | | | | | | | | 2022 | | | | | | | | |
| 2023 2024 | | | | | | | | | 2023 2024 | | | | | | | | |
| 2024 | | | | | | | | | 2024 | | | | | | | | |
| 2002 | 5 | 23 | 0.14 | 0.48 | 0.92 | 0.02 | 0.18 | 0.36 | 2002 | 9 | 23 | 0.43 | 0.84 | 1.37 | 0.13 | 0.35 | 0.62 |
| 2003 | | 27 | 0.29 | 0.63 | 1.06 | 0.08 | 0.21 | 0.36 | 2003 | | 27 | 0.46 | 0.98 | 1.68 | 0.18 | 0.38 | 0.61 |
| 2004 2005 | | 33 28 | 0.29 | 0.80 1.36 | 1.50 2.79 | 0.08 0.16 | 0.29 | 0.55 | 2004 2005 | | 33 28 | 0.62 | 1.09 1.24 | 1.69 2.55 | 0.21 | 0.45 0.49 | 0.73 0.94 |
| 2005 | | 28 | 0.47 | 1.15 | 2.98 | 0.00 | 0.48 | 1.05 | 2005 | | 28 | 0.42 | 0.48 | 1.20 | 0.00 | 0.49 | 0.94 |
| 2007 | | 28 | 0.92 | 2.03 | 3.79 | 0.20 | 0.49 | 0.86 | 2007 | | 28 | 0.84 | 1.52 | 2.44 | 0.19 | 0.35 | 0.54 |
| 2008 | | 29 | 2.19 | 4.03 | 6.92 | 0.50 | 0.96 | | 2008 | | 29 | 0.14 | 0.28 | 0.45 | 0.02 | 0.04 | 0.07 |
| 2009 2010 | | 29 28 | 0.10 | 0.31 0.40 | 0.56 1.00 | 0.02 | 0.09 0.14 | 0.17 | 2009 2010 | | 29 28 | 0.36 | 0.66 | 1.03 1.51 | 0.10 | 0.19 0.29 | 0.29 |
| 2011 | | 29 | 0.26 | 0.96 | 2.07 | 0.09 | 0.36 | | 2011 | | 29 | 0.22 | 0.79 | 1.61 | 0.09 | 0.29 | 0.52 |
| 2012 | | | | | | | | | 2012 | | | | | | | | |
| 2013 2014 | | 29 29 | 0.34 | 0.93 | 1.77 0.80 | 0.11 | 0.30 | 0.52 | 2013 2014 | | 29 29 | 0.44 | 1.22 0.36 | 2.43 0.66 | 0.15 | 0.43 | 0.78 0.11 |
| 2014 | | 29 | 0.38 | 0.37 | 0.80 | 0.03 | 0.10 | | 2014 | | 29 | 0.12 | 0.38 | 0.88 | 0.02 | 0.08 | 0.03 |
| 2016 | | 29 | 3.85 | 5.86 | 8.71 | 1.30 | 2.06 | | 2016 | | 29 | 1.35 | 2.12 | 3.13 | 0.34 | 0.63 | 0.97 |
| 2017 | | 28 | 0.54 | 1.27 | 2.33 | 0.09 | 0.40 | 0.79 | 2017 | | 28 | 0.71 | 1.44 | 2.48 | 0.17 | 0.41 | 0.71 |
| 2018 2019 | | 25 | 0.31 | 0.78 | 1.43 | 0.07 | 0.21 | 0.38 | 2018 2019 | | 25 | 0.37 | 0.72 | 1.17 | 0.08 | 0.18 | 0.28 |
| 2019 | | 25 | 4.26 | 8.18 | 15.00 | 1.12 | 2.11 | 3.55 | 2019 | | 25 | 4.15 | 7.58 | 13.29 | 1.05 | 1.91 | 3.12 |
| 2021 | | 25 | 2.04 | 4.28 | 8.20 | 0.54 | 1.08 | | 2021 | | 25 | 0.31 | 0.73 | 1.28 | 0.05 | 0.13 | 0.21 |
| 2022 | | | | | | | | | 2022 | | | | | | | | |
| 2023 2024 | | | | | | | | | 2023 2024 | | | | | | | | |
| 2024 | | | | | | | | | 2024 | | | | | | | | |
| 2002 | 6 | 23 | 0.45 | 1.02 | 1.82 | 0.17 | 0.46 | | 2002 | 10+ | 23 | 0.07 | 0.33 | 0.65 | 0.00 | 0.12 | 0.26 |
| 2003 | | 27 | 0.50 | 1.09 | 1.90 | 0.19 | 0.42 | | 2003 | | 27 | 0.12 | 0.36 | 0.65 | 0.04 | 0.12 | 0.21 |
| 2004 2005 | | 33 28 | 0.65 | 1.24 0.96 | 2.05 2.08 | 0.21 | 0.49 | | 2004 2005 | | 33 28 | 0.62 | 1.21 0.98 | 2.01 2.05 | 0.24 | 0.54 | 0.92 |
| 2005 | | 28 | 0.31 | 1.41 | 3.44 | 0.01 | 0.47 | | 2005 | | 28 | 0.36 | 1.24 | 2.70 | 0.06 | 0.50 | 1.12 |
| 2007 | | 28 | 1.21 | 2.75 | 5.37 | 0.29 | 0.69 | 1.21 | 2007 | | 28 | 0.72 | 1.46 | 2.53 | 0.19 | 0.41 | 0.66 |
| 2008 | | 29 | 0.61 | 1.13 | 1.82 | 0.09 | 0.24 | | 2008 | | 29 | 1.15 | 1.92 | 2.96 | 0.24 | 0.44 | 0.67 |
| 2009 2010 | | 29 28 | 0.65 0.64 | 1.08 1.48 | 1.63 2.75 | 0.20 | 0.36 | 0.54 | 2009 2010 | | 29 28 | 0.04 | 0.20 | 0.37 1.45 | 0.01 | 0.04 0.25 | 0.08 0.60 |
| 2010 | | 29 | 0.51 | 1.07 | 1.83 | 0.13 | 0.36 | | 2010 | | 29 | 0.22 | 0.73 | 1.45 | 0.09 | 0.26 | 0.47 |
| 2012 | | | | | | | | | 2012 | | | | | | | | |
| 2013 | | 29 | 0.66 | 1.84 | 3.84 | 0.25 | 0.67 | | 2013 | | 29 | 1.55 | 4.05 | 8.98 | 0.63 | 1.50 | 2.82 |
| 2014 2015 | | 29 29 | 0.14 | 0.41 | 0.74 0.67 | 0.01 | 0.06 | | 2014 2015 | | 29 29 | 0.48 | 0.69 | 0.94 | 0.08 | 0.14 | 0.20 |
| 2015 | | 29 | 1.94 | 3.10 | 4.71 | 0.61 | 1.07 | | 2015 | | 29 | 0.99 | 1.61 | 2.43 | 0.25 | 0.47 | 0.74 |
| 2017 | | 28 | 3.13 | 5.92 | 10.59 | 1.12 | 2.16 | | 2017 | | 28 | 1.41 | 2.81 | 5.05 | 0.39 | 0.90 | 1.59 |
| 2018 2019 | | 25 | 0.24 | 0.55 | 0.95 | 0.05 | 0.13 | 0.22 | 2018 2019 | | 25 | 0.29 | 0.65 | 1.12 | 0.07 | 0.18 | 0.31 |
| 2019 | | 25 | 3.79 | 7.18 | 12.98 | 0.05 | 0.13 | 3.07 | 2019 | | 25 | 1.61 | 2.96 | 5.01 | 0.07 | 0.18 | 1.12 |
| 2021 | | 25 | 2.77 | 5.89 | 11.60 | 0.79 | 1.54 | | 2021 | | 25 | 2.81 | 6.16 | 12.44 | 0.77 | 1.59 | 2.79 |
| 2022 | | | | | | | | | 2022 | | | | | | | | |
| 2023 2024 | | | | | | | | | 2023 2024 | | | | | | | | |
| 2024 | | | | | | | | | 2024 | | | | | | | | |
| - | | | | | | | | | | | | | | | | | |

Table 30. White Perch geometric mean indices of abundance for March by number and biomass, overall and by age class.

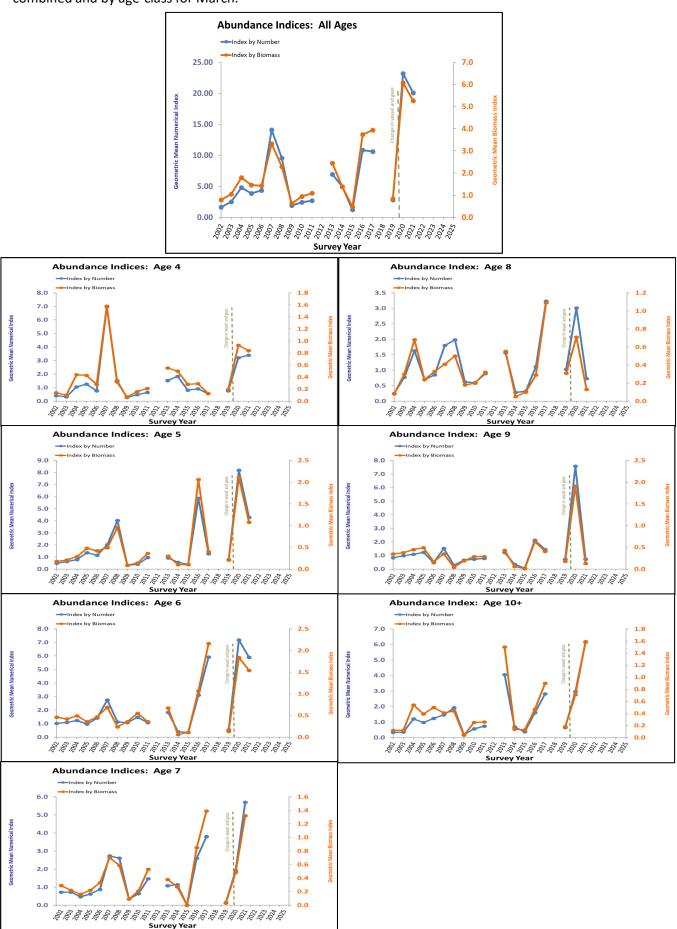
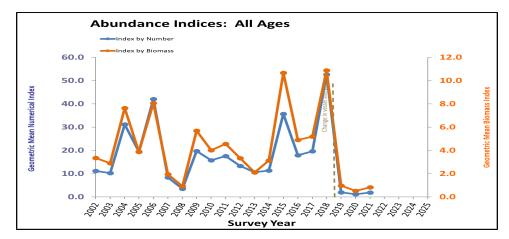


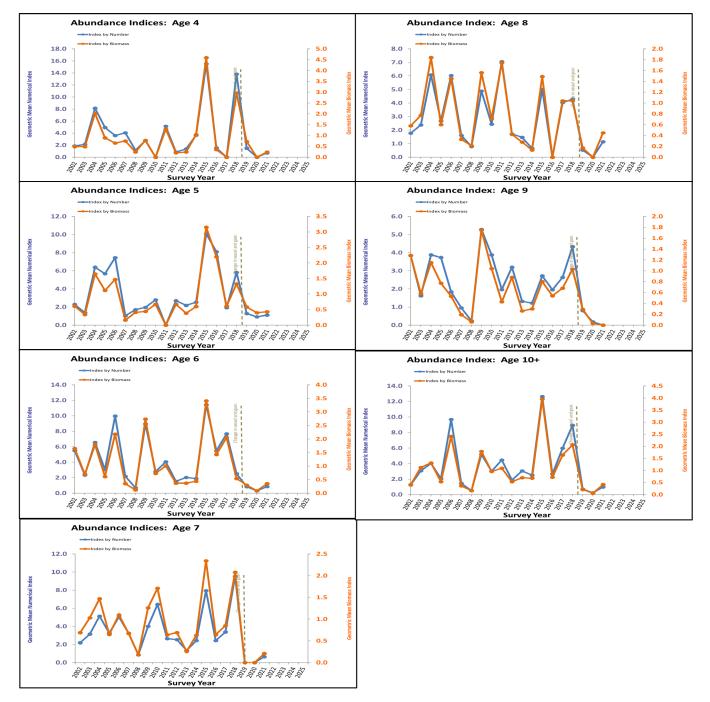
Figure 64. White Perch geometric mean indices of abundance, by number and biomass, for all ages combined and by age-class for March.

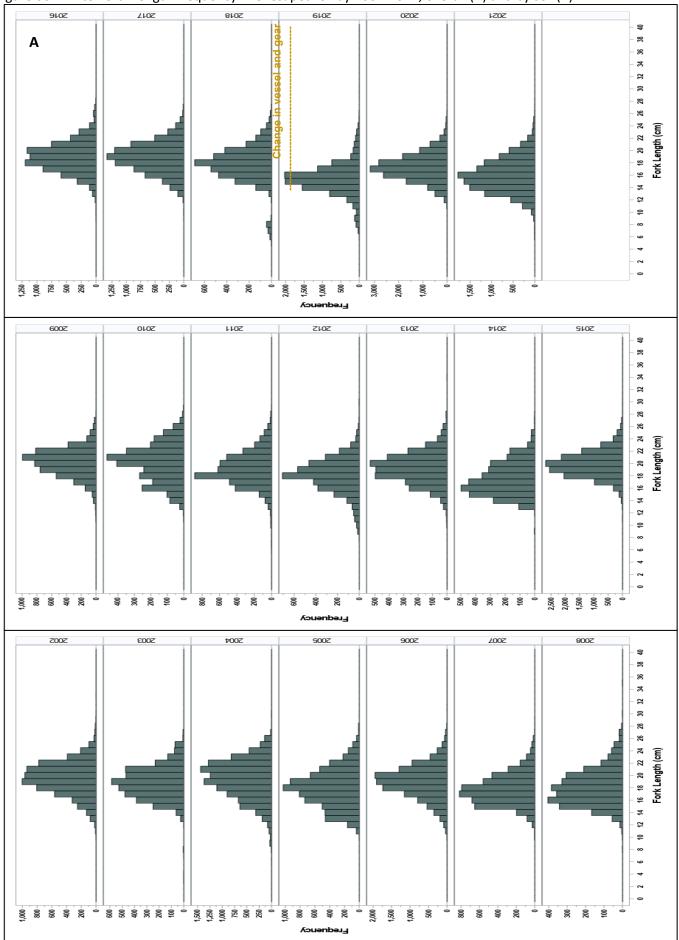
Table 31. White Perch geometric mean indices of abundance for fall by number and biomass, overall and by age class.

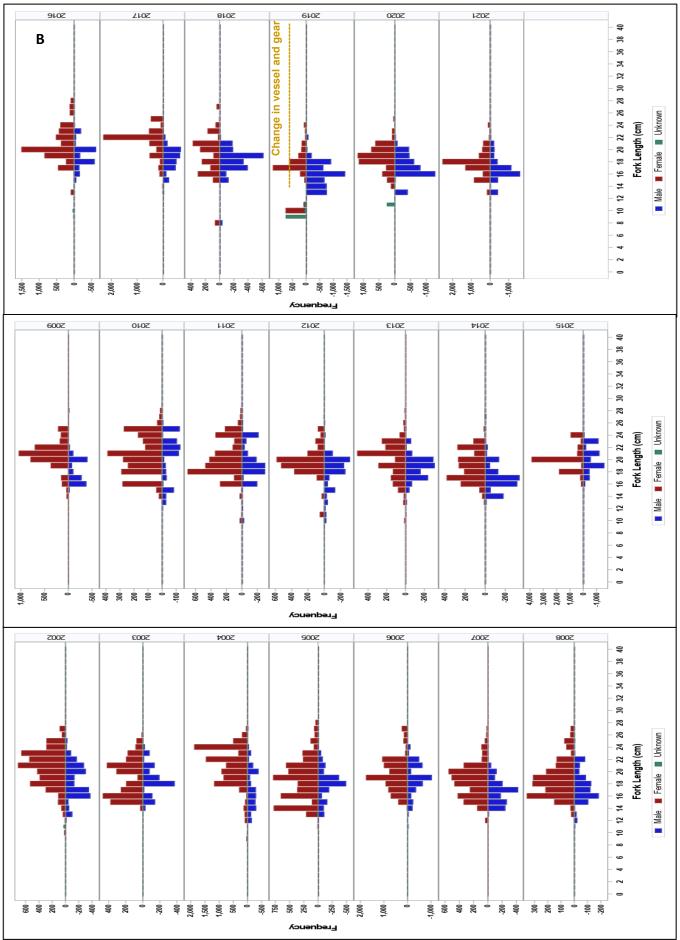
| Year | Age | n | Nur LCI | merical In | dex UCI | Bio | mass In Index | dex UCI | Year | Age | n | Nun LCI | nerical In Index | dex UCI | Bic LCI | mass In Index | dex UCI |
|--------------|-----|----------|-------------|--------------|---------------|------------|------------------|--------------|--------------|-----|----------|------------|---------------------|-------------|------------|------------------|------------|
| 2002 | All | 41 | 6.4 | 11.2 | 19.1 | 1.9 | 3.3 | 5.5 | 2002 | 7 | 41 | 1.2 | 2.2 | 3.6 | 0.4 | 0.7 | 1.1 |
| 2003 | | 24 | 4.0 | 10.3 | 24.4 | 1.0 | 2.9 | 6.6 | 2003 | | 24 | 1.0 | 3.1 | 7.6 | 0.3 | 1.0 | 2.3 |
| 2004 2005 | | 36 30 | 12.1 9.0 | 31.1 19.5 | 78.1 40.9 | 3.4 2.0 | 7.6 3.9 | 16.1 7.0 | 2004 2005 | | 36 30 | 2.2 | 5.1 3.3 | 10.6 5.7 | 0.6 | 1.5 | 2.7 |
| 2006 | | 28 | 16.5 | 42.1 | 105.1 | 3.5 | 8.1 | 17.1 | 2006 | | 28 | 2.2 | 5.0 | 10.6 | 0.4 | 1.1 | 2.1 |
| 2007 | | 15 | 3.8 | 8.3 | 17.3 | 0.9 | 2.0 | 3.6 | 2007 | | 15 | 1.6 | 3.2 | 5.9 | 0.4 | 0.7 | 1.1 |
| 2008 2009 | | 32 32 | 1.3 7.9 | 3.5 19.7 | 8.1 47.4 | 0.3 | 0.9 5.7 | 1.8 12.3 | 2008 2009 | | 32 32 | 0.3 | 0.9 | 1.7 9.1 | 0.0 | 0.2 | 0.4 |
| 2003 | | 31 | 6.1 | 15.8 | 38.7 | 1.7 | 4.0 | 8.3 | 2003 | | 31 | 2.5 | 6.4 | 14.9 | 0.4 | 1.5 | 3.4 |
| 2011 | | 30 | 6.6 | 17.6 | 44.5 | 2.1 | 4.6 | 9.0 | 2011 | | 30 | 1.3 | 2.6 | 4.8 | 0.3 | 0.6 | 1.1 |
| 2012 | | 31 | 5.1 | 13.3 | 32.4 | 1.4 | 3.3 | 6.7 | 2012 | | 31 | 1.0 | 2.5 | 5.4 | 0.3 | 0.7 | 1.3 |
| 2013 2014 | | 32 32 | 5.2 4.5 | 10.6 11.3 | 21.0 26.4 | 1.1 | 2.1 3.1 | 3.6 6.0 | 2013 2014 | | 32 32 | 0.6 | 1.4 | 2.5 5.1 | 0.1 | 0.3 | 0.5 |
| 2015 | | 32 | 14.9 | 35.7 | 83.6 | 5.0 | 10.7 | 21.7 | 2015 | | 32 | 3.5 | 7.9 | 16.9 | 1.1 | 2.3 | 4.3 |
| 2016 | | 32 | 8.2 | 17.9 | 37.8 | 2.5 | 4.9 | 8.9 | 2016 | | 32 | 1.1 | 2.4 | 4.6 | 0.2 | 0.6 | 1.2 |
| 2017 2018 | | 32 16 | 7.2 15.6 | 19.7 52.6 | 51.1 172.7 | 2.1 3.8 | 5.2 10.9 | 11.3 28.3 | 2017 2018 | | 32 16 | 1.4 3.3 | 3.4 9.5 | 6.9 24.7 | 0.3 | 0.9 | 1.6 4.6 |
| 2018 | | 24 | 0.2 | 2.0 | 6.6 | 0.1 | 1.0 | 28.3 | 2018 | | 24 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 2020 | | 25 | 0.0 | 1.2 | 3.6 | 0.0 | 0.5 | 1.3 | 2020 | | 25 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 2021 | | 25 | 0.2 | 1.8 | 5.6 | 0.1 | 0.8 | 2.1 | 2021 | | 25 | 0.0 | 0.6 | 1.6 | 0.0 | 0.2 | 0.5 |
| 2022 2023 | | | | | | | | | 2022 2023 | | | | | | | | |
| 2023 | | | | | | | | | 2023 | | | | | | | | |
| 2025 | | | | | | | | | 2025 | | | | | | | | |
| 2002 | 4 | 41 | 1.0 | 1.9 | 3.0 | 0.3 | 0.5 | 0.8 | 2002 | 8 | 41 | 1.0 | 1.8 | 2.9 | 0.3 | 0.6 | 0.9 |
| 2003 2004 | | 24 36 | 0.9 3.4 | 2.1 8.1 | 4.2 17.9 | 0.1 | 0.5 | 1.0 3.7 | 2003 2004 | | 24 36 | 0.8 | 2.4 6.1 | 5.3 13.0 | 0.2 | 0.8 | 1.6 3.5 |
| 2004 | | 30 | 2.6 | 5.0 | 8.8 | 0.4 | 0.9 | 1.5 | 2004 | | 30 | 1.3 | 2.7 | 4.8 | 0.3 | 0.6 | 1.0 |
| 2006 | | 28 | 1.6 | 3.6 | 7.0 | 0.3 | 0.7 | 1.1 | 2006 | | 28 | 2.6 | 6.0 | 12.8 | 0.5 | 1.5 | 2.9 |
| 2007 | | 15 | 2.0 | 4.1 | 7.4 | 0.3 | 0.8 | 1.4 | 2007 | | 15 | 0.4 | 1.6 | 3.9 | 0.1 | 0.3 | 0.7 |
| 2008 2009 | | 32 32 | 0.4 | 1.2 2.8 | 2.4 5.8 | 0.1 | 0.2 | 0.5 | 2008 2009 | | 32 32 | 0.2 | 0.8 4.9 | 1.7 11.2 | 0.0 | 0.2 | 0.4 |
| 2009 | | 31 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 2009 | | 31 | 0.8 | 2.4 | 5.5 | 0.2 | 0.7 | 1.4 |
| 2011 | | 30 | 2.1 | 5.1 | 11.0 | 0.6 | 1.3 | 2.4 | 2011 | | 30 | 3.3 | 7.1 | 14.3 | 0.9 | 1.7 | 2.9 |
| 2012 | | 31 | 0.3 | 0.9 | 1.8 | 0.1 | 0.2 | 0.4 | 2012 | | 31 | 0.6 | 1.7 | 3.5 | 0.1 | 0.4 | 0.8 |
| 2013 2014 | | 32 32 | 0.6 | 1.4 3.6 | 2.5 8.5 | 0.1 | 0.2 | 0.4 2.0 | 2013 2014 | | 32 32 | 0.6 | 1.5 | 2.7 1.2 | 0.1 | 0.3 | 0.5 |
| 2014 | | 32 | 7.4 | 15.5 | 31.7 | 2.3 | 4.6 | 8.4 | 2015 | | 32 | 2.1 | 5.0 | 10.5 | 0.6 | 1.5 | 2.8 |
| 2016 | | 32 | 0.6 | 1.5 | 3.0 | 0.2 | 0.4 | 0.6 | 2016 | | 32 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 2017 | | 32 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 2017 | | 32 | 1.6 | 4.0 | 8.7 | 0.4 | 1.0 | 1.9 |
| 2018 2019 | | 16 24 | 5.0 0.1 | 13.8 1.5 | 35.5 4.9 | 1.1 | 3.0 0.7 | 6.4 2.0 | 2018 2019 | | 16 24 | 1.5 | 4.3 0.5 | 10.4 1.3 | 0.4 | 1.0 0.2 | 2.0 |
| 2015 | | 25 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 2020 | | 25 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 2021 | | 25 | 0.0 | 0.7 | 1.8 | 0.0 | 0.2 | 0.6 | 2021 | | 25 | 0.1 | 1.1 | 3.1 | 0.0 | 0.5 | 1.2 |
| 2022 | | | | | | | | | 2022 | | | | | | | | |
| 2023 2024 | | | | | | | | | 2023 2024 | | | | | | | | |
| 2024 | | | | | | | | | 2024 | | | | | | | | |
| 2002 | 5 | 41 | 1.3 | 2.3 | 3.7 | 0.3 | 0.6 | 1.0 | 2002 | 9 | 41 | 2.2 | 3.8 | 6.4 | 0.7 | 1.3 | 2.1 |
| 2003 | | 24 | 0.4 | 1.4 | 3.0 | 0.1 | 0.3 | 0.7 | 2003 | | 24 | 0.4 | 1.6 | 3.9 | 0.1 | 0.6 | 1.3 |
| 2004 2005 | | 36 30 | 2.8 2.9 | 6.4 5.7 | 13.3 10.5 | 0.8 0.6 | 1.7 | 3.0 1.9 | 2004 2005 | | 36 30 | 1.7 2.0 | 3.9 3.7 | 7.8 6.5 | 0.5 | 1.2 0.8 | 2.2 |
| 2005 | | 28 | 3.2 | 7.5 | 16.2 | 0.6 | 1.5 | 2.8 | 2005 | | 28 | 0.6 | 1.8 | 4.0 | 0.4 | 0.5 | 1.2 |
| 2007 | | 15 | 0.2 | 1.0 | 2.3 | 0.0 | 0.2 | 0.3 | 2007 | | 15 | 0.2 | 1.0 | 2.2 | 0.0 | 0.2 | 0.4 |
| 2008 | | 32 | 0.5 | 1.7 | 3.8 | 0.1 | 0.4 | 0.8 | 2008 | | 32 | 0.0 | 0.2 | 0.5 | 0.0 | 0.1 | 0.1 |
| 2009 2010 | | 32 31 | 0.9 | 2.0 2.8 | 3.6 6.2 | 0.2 | 0.4 | 0.8 1.3 | 2009 2010 | | 32 31 | 1.8 1.5 | 5.3 3.9 | 12.9 8.4 | 0.6 | 1.8 1.0 | 3.8 |
| 2010 | | 30 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 2011 | | 30 | 1.0 | 2.0 | 3.4 | 0.4 | 0.4 | 0.7 |
| 2012 | | 31 | 1.2 | 2.7 | 5.3 | 0.3 | 0.7 | 1.2 | 2012 | | 31 | 1.2 | 3.2 | 7.0 | 0.3 | 0.9 | 1.7 |
| 2013 | | 32 | 1.1 | 2.2 | 3.9 | 0.1 | 0.4 | 0.7 | 2013 | | 32 | 0.6 | 1.3 | 2.4 | 0.1 | 0.3 | 0.4 |
| 2014 2015 | | 32 32 | 1.1 4.3 | 2.5 10.1 | 4.9 22.3 | 0.3 | 0.6 | 1.0 5.9 | 2014 2015 | | 32 32 | 0.5 | 1.2 | 2.3 5.7 | 0.1 | 0.3 | 0.5 |
| 2015 | | 32 | 3.7 | 8.1 | 16.7 | 1.0 | 2.2 | 4.1 | 2015 | | 32 | 0.9 | 2.0 | 3.7 | 0.2 | 0.5 | 1.0 |
| 2017 | | 32 | 0.6 | 1.9 | 4.5 | 0.2 | 0.6 | 1.2 | 2017 | | 32 | 1.1 | 2.6 | 5.4 | 0.2 | 0.7 | 1.3 |
| 2018 | | 16 | 1.9 | 5.8 | 14.9 | 0.4 | 1.3 | 3.0 | 2018 | | 16 | 1.5 | 4.3 | 10.5 | 0.4 | 1.0 | 2.1 |
| 2019 2020 | | 24 25 | 0.1 | 1.3 0.9 | 3.9 2.8 | 0.0 | 0.6 0.4 | 1.5 1.0 | 2019 2020 | | 24 25 | 0.0 | 0.8 | 2.1 0.4 | 0.0 | 0.3 | 0.7 |
| 2020 | | 25 | 0.0 | 1.1 | 3.0 | 0.0 | 0.4 | 1.0 | 2020 | | 25 | 0.0 | 0.2 | 0.4 | 0.0 | 0.0 | 0.1 |
| 2022 | | | | | | | | | 2022 | | | | | | | | |
| 2023 | | | | | | | | | 2023 | | | | | | | | |
| 2024 2025 | | | | | | | | | 2024 2025 | | | | | | | | |
| 2002 | 6 | 41 | 3.2 | 5.5 | 9.0 | 1.0 | 1.7 | 2.6 | 2002 | 10+ | 41 | 0.7 | 1.3 | 2.1 | 0.2 | 0.4 | 0.7 |
| 2003 | | 24 | 0.8 | 2.4 | 5.3 | 0.2 | 0.7 | 1.5 | 2003 | | 24 | 0.9 | 3.1 | 7.7 | 0.3 | 1.1 | 2.6 |
| 2004 | | 36 | 3.0 | 6.6 | 13.4 | 0.8 | 1.8 | 3.2 | 2004 | | 36 | 1.7 | 4.0 | 8.5 | 0.5 | 1.3 | 2.6 |
| 2005 2006 | | 30 28 | 1.5 4.1 | 3.1 10.0 | 5.4 22.7 | 0.3 | 0.6 2.2 | 1.0 4.4 | 2005 2006 | | 30 28 | 1.0 3.7 | 2.1 9.7 | 3.8 23.4 | 0.2 | 0.5 | 0.9 |
| 2006 | | 28 15 | 4.1 | 2.2 | 4.0 | 0.9 | 0.4 | 4.4 0.6 | 2006 | | 28 15 | 0.2 | 9.7 | 23.4 | 0.9 | 0.4 | 0.8 |
| 2008 | | 32 | 0.2 | 0.7 | 1.3 | 0.0 | 0.1 | 0.2 | 2008 | | 32 | 0.1 | 0.5 | 1.1 | 0.0 | 0.2 | 0.3 |
| 2009 | | 32 | 3.5 | 9.0 | 21.1 | 1.1 | 2.7 | 5.7 | 2009 | | 32 | 1.8 | 5.2 | 12.6 | 0.6 | 1.8 | 3.8 |
| 2010 2011 | | 31 30 | 1.0 1.8 | 2.8 4.1 | 6.3 8.3 | 0.2 | 0.7 | 1.4 1.8 | 2010 2011 | | 31 30 | 1.0 2.1 | 3.0 4.5 | 7.2 | 0.3 | 1.0 | 2.0 1.8 |
| 2011 | | 31 | 0.6 | 4.1 | 3.0 | 0.4 | 0.4 | 0.7 | 2011 | | 31 | 0.6 | 4.5 | 4.1 | 0.8 | 0.5 | 1.8 |
| 2012 | | 32 | 0.9 | 2.0 | 3.8 | 0.1 | 0.4 | 0.7 | 2013 | | 32 | 1.4 | 3.1 | 5.8 | 0.3 | 0.7 | 1.0 |
| 2014 | | 32 | 0.7 | 1.9 | 3.8 | 0.2 | 0.4 | 0.7 | 2014 | | 32 | 1.0 | 2.5 | 5.0 | 0.3 | 0.7 | 1.2 |
| 2015 | | 32 | 5.3 | 11.4 | 23.7 | 1.6 | 3.4 | 6.4 | 2015 | | 32 | 5.0 | 12.6 | 30.1 | 1.7 | 4.0 | 8.0 |
| 2016 2017 | | 32 32 | 2.7 3.0 | 5.5 7.7 | 10.5 17.8 | 0.7 | 1.4 2.1 | 2.4 4.0 | 2016 2017 | | 32 32 | 1.2 2.3 | 2.6 6.0 | 5.1 13.9 | 0.3 | 0.7 | 1.4 |
| 2018 | | 16 | 1.1 | 2.5 | 4.9 | 0.2 | 0.5 | 1.1 | 2018 | | 16 | 2.8 | 8.9 | 24.6 | 0.7 | 2.1 | 4.7 |
| 2019 | | 24 | 0.1 | 0.8 | 2.2 | 0.0 | 0.3 | 0.7 | 2019 | | 24 | 0.1 | 0.7 | 1.6 | 0.0 | 0.2 | 0.5 |
| 2020 | | 25 | 0.0 | 0.3 | 0.8 | 0.0 | 0.1 | 0.2 | 2020 | | 25 | 0.0 | 0.2 | 0.6 | 0.0 | 0.1 | 0.1 |
| 2021 | | 25 | 0.0 | 0.9 | 2.4 | 0.0 | 0.4 | 0.9 | 2021 2022 | | 25 | 0.0 | 1.0 | 2.9 | 0.0 | 0.4 | 1.1 |
| 2022 | - | _ | | | | | | | 2022 | | | | | | | | |
| 2022 2023 | | | | | | | | | | | | | | | | | |
| | | | | | | | | | 2024 | | | | | | | | |

Figure 65. White Perch geometric mean indices of abundance, by number and biomass, for all ages combined and by age-class for November.









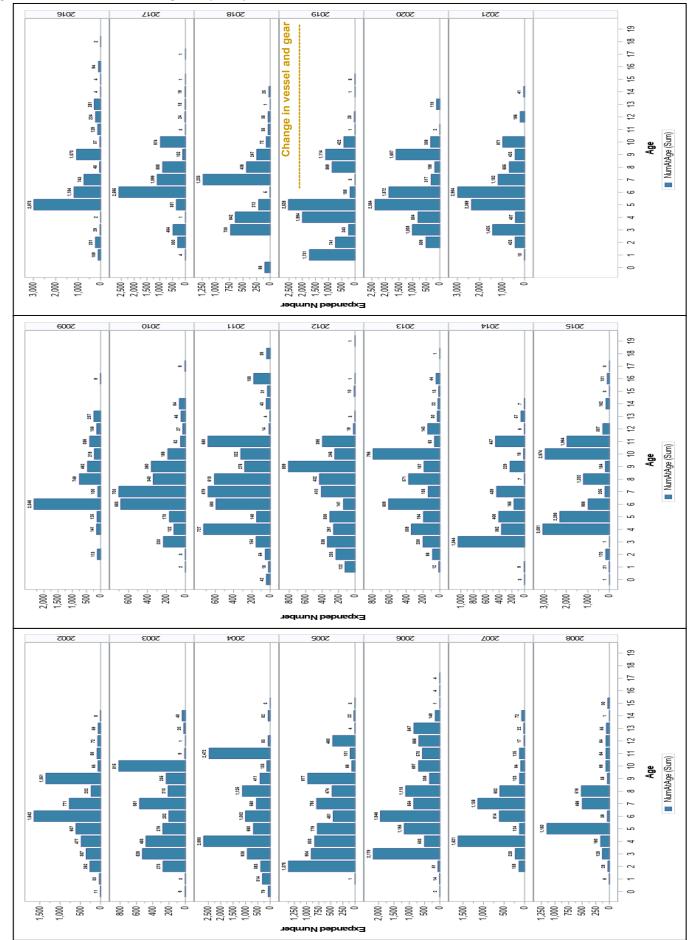


Figure 67. White Perch total age-frequency, 2002-2021.

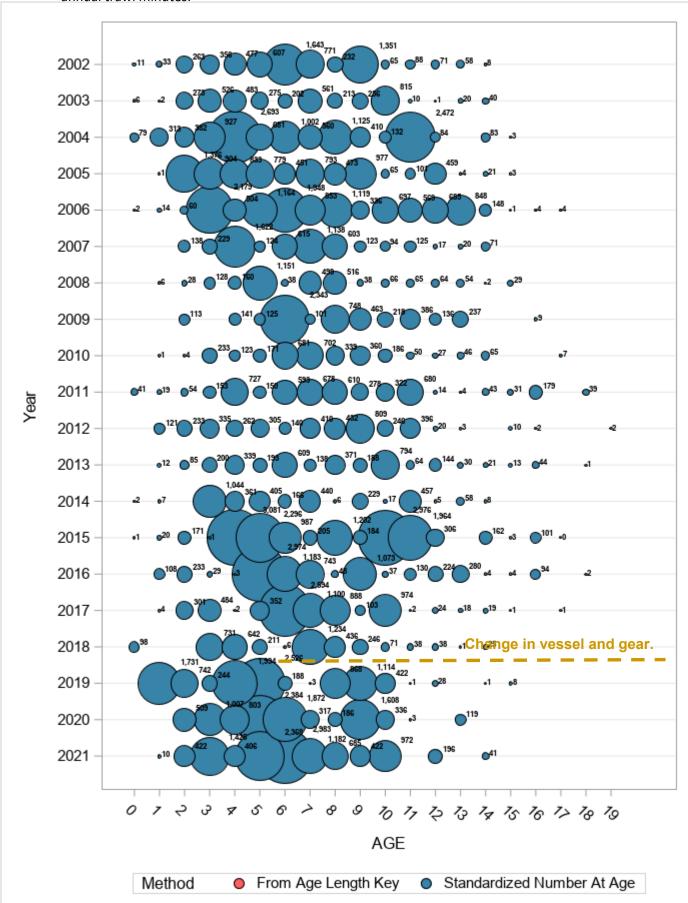


Figure 68. White Perch age-frequency by year, 2002-2021 standardized to 8,000 (4,800 for 2019 and forward) annual trawl minutes.

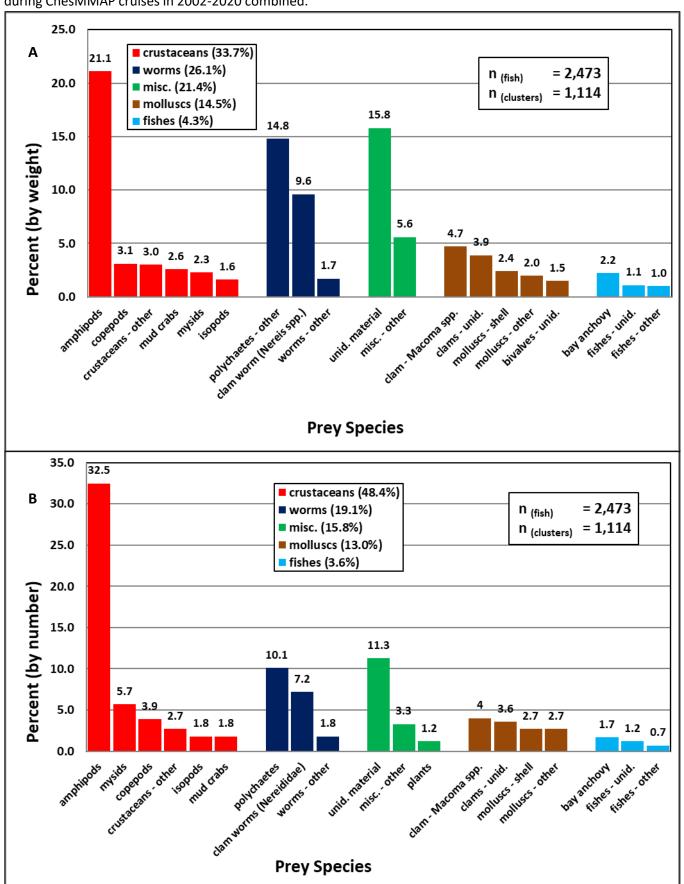


Figure 69. Diet composition, expressed as percent by weight (A) and percent by number (B) of White Perch collected during ChesMMAP cruises in 2002-2020 combined.

Figure 70. The R/V Bay Eagle and the R/V Virginia preparing to conduct a calibration tow (A) and the R/V Virginia travelling towards the next site upon completion of a tow (B).





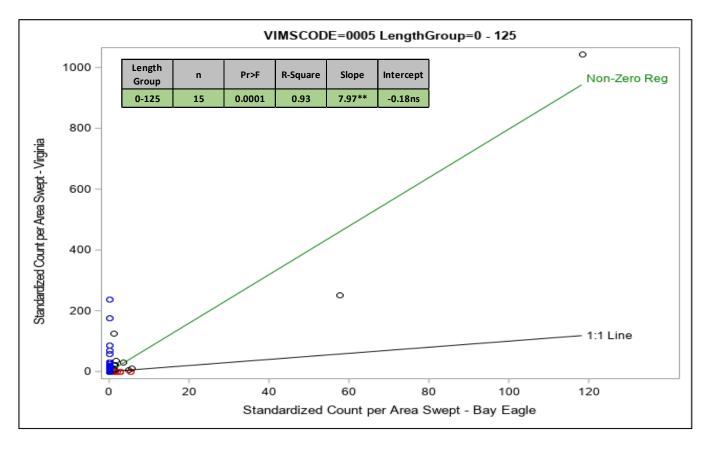
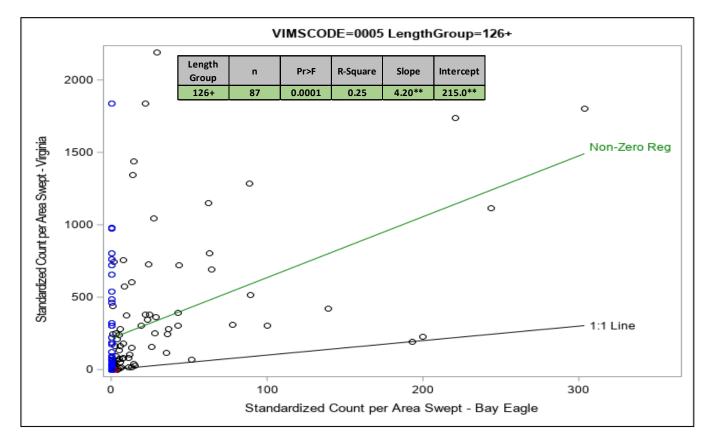


Figure 71. Preliminary comparison of catch rates during calibration tows for Atlantic Croaker, 0 – 125mm.

Figure 72. Preliminary comparison of catch rates during calibration tows for Atlantic Croaker, 126 – 180mm.



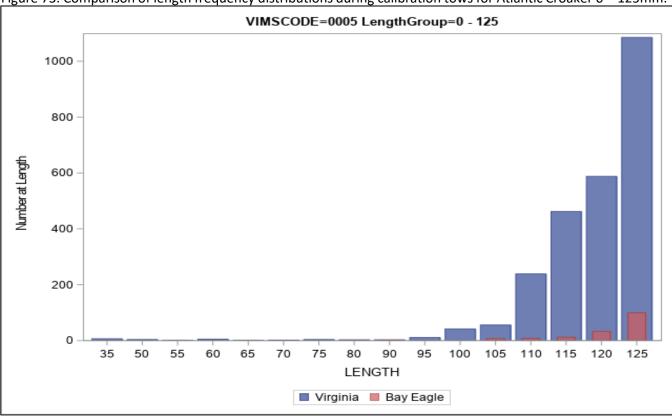
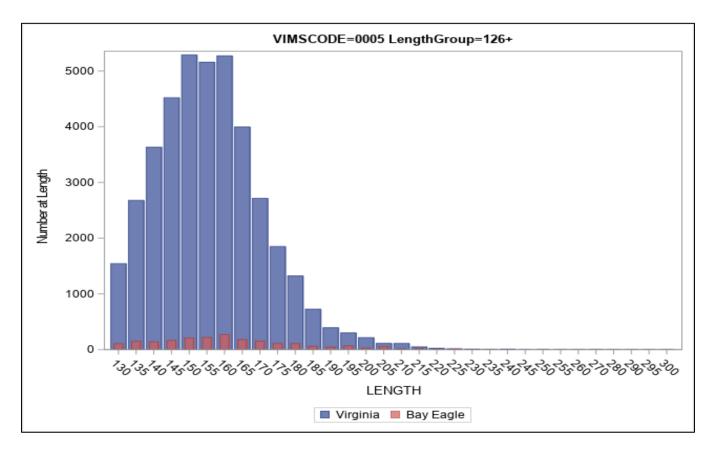


Figure 73. Comparison of length frequency distributions during calibration tows for Atlantic Croaker 0 – 125mm.

Figure 74. Comparison of length frequency distributions during calibration tows for Atlantic Croaker 126+mm.



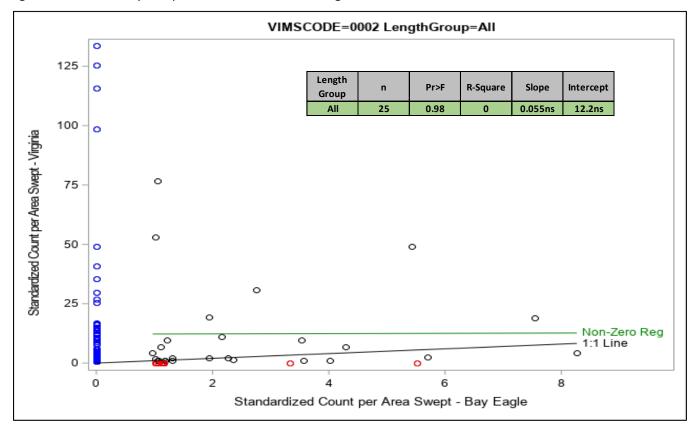
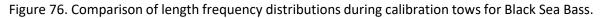
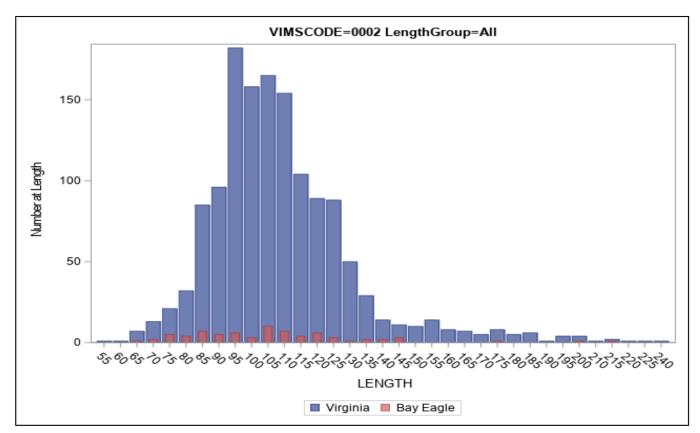
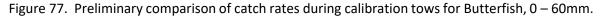


Figure 75. Preliminary comparison of catch rates during calibration tows for Black Sea Bass, all sizes.







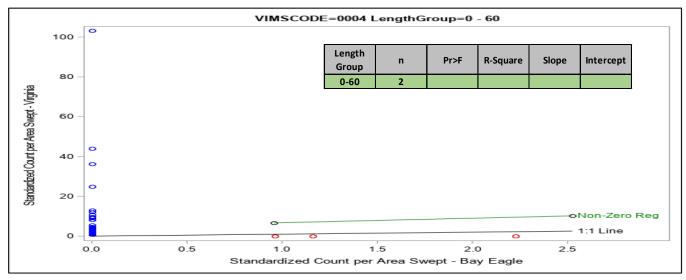


Figure 78. Preliminary comparison of catch rates during calibration tows for Butterfish, 61+mm.

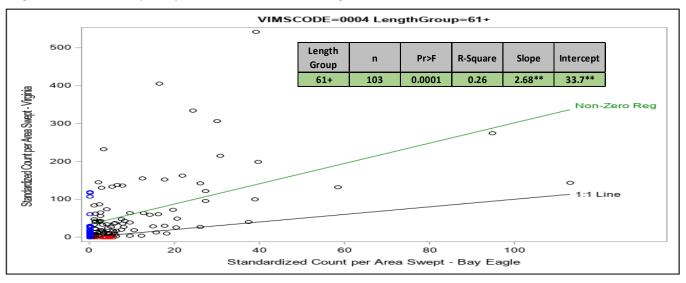
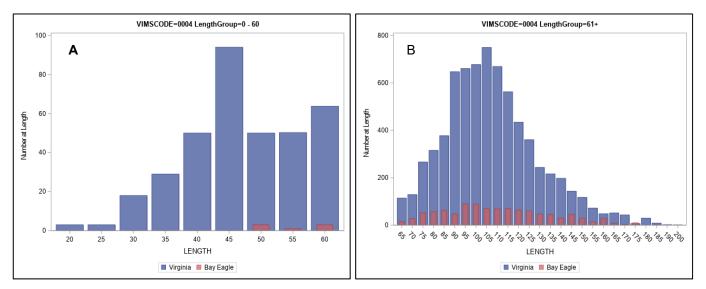
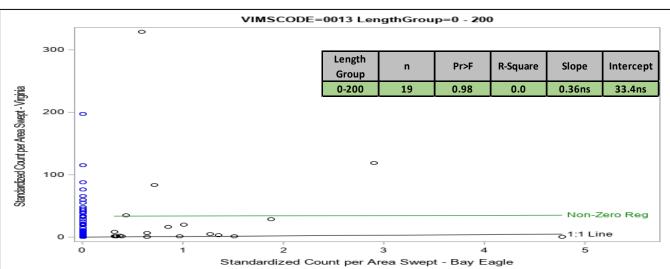


Figure 79. Comparison of length frequency distributions during calibration tows for Butterfish, 0 – 60mm (A) and 61+mm (B).





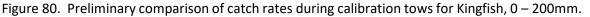


Figure 81. Preliminary comparison of catch rates during calibration tows for Kingfish, 201+mm.

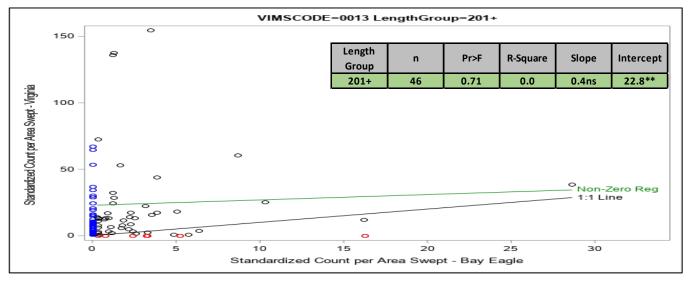
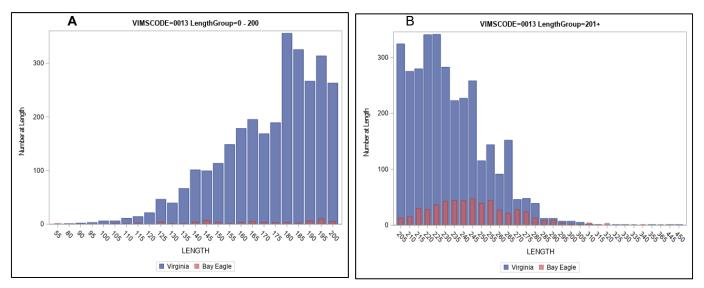


Figure 82. Comparison of length frequency distributions during calibration tows for Kingfish, 0 – 200mm (A) and 200+mm (B).



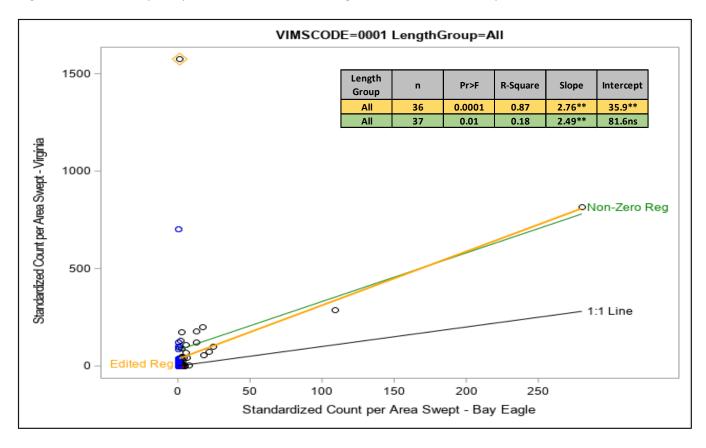
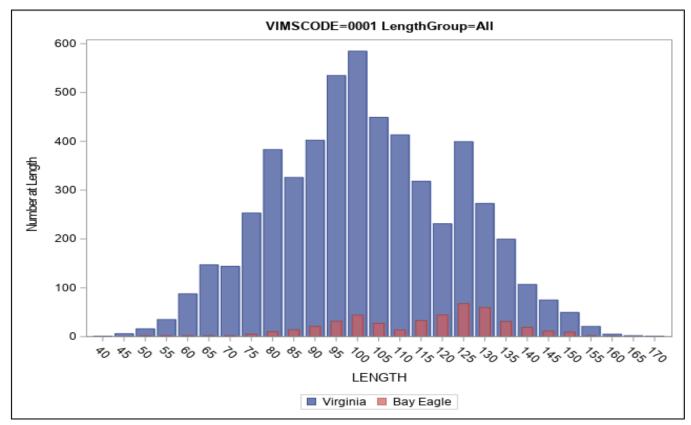


Figure 83. Preliminary comparison of catch rates during calibration tows for Scup, all sizes.

Figure 84. Comparison of length frequency distributions during calibration tows for Scup.





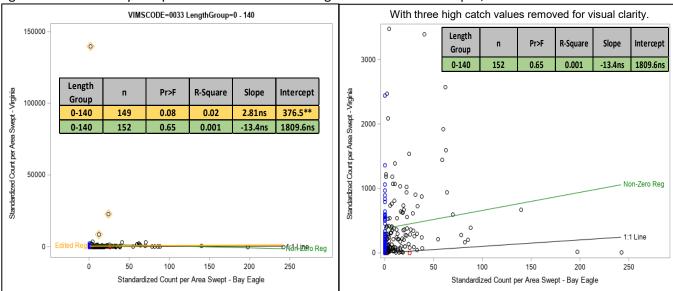


Figure 86. Preliminary comparison of catch rates during calibration tows for Spot, 141+mm.

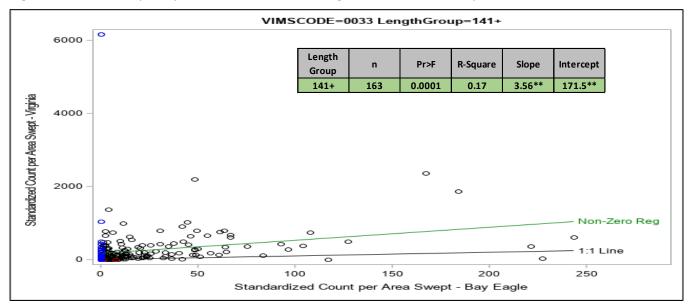
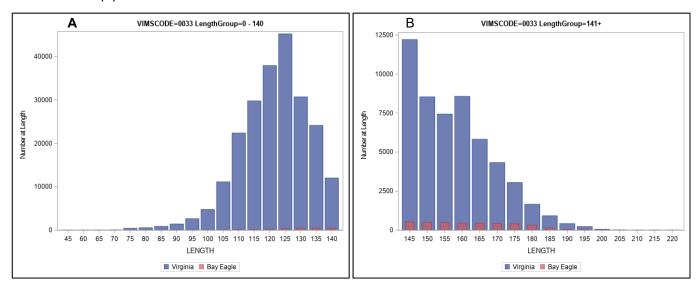


Figure 87. Comparison of length frequency distributions during calibration tows for Spot, 0 – 140mm (A) and 140+mm (B).



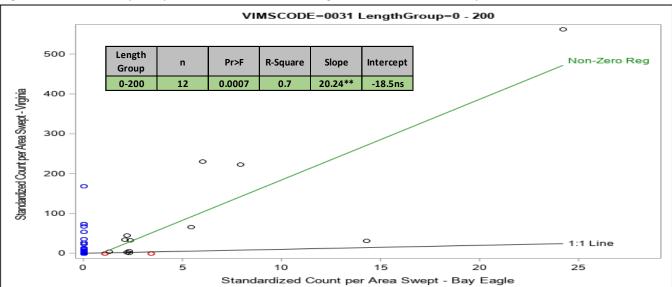


Figure 88. Preliminary comparison of catch rates during calibration tows for Striped Bass, 0- 200mm.

Figure 89. Preliminary comparison of catch rates during calibration tows for Striped Bass, 201-300mm.

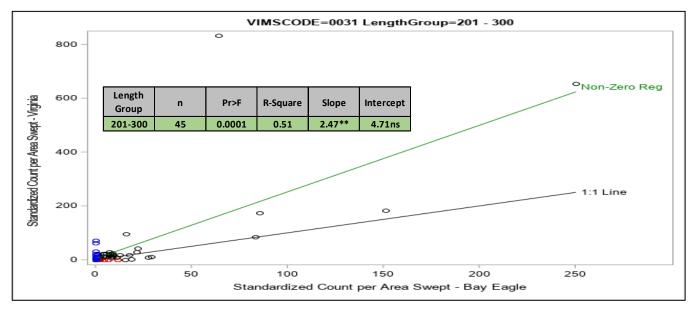
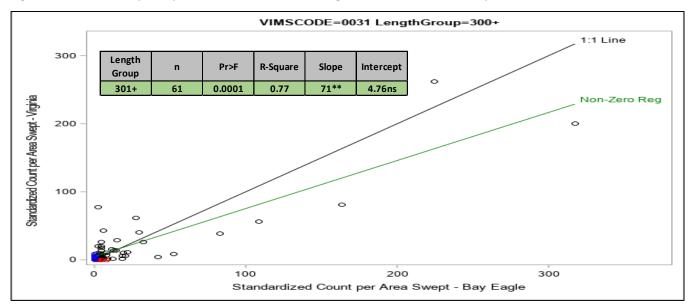


Figure 90. Preliminary comparison of catch rates during calibration tows for Striped Bass, 301+mm.



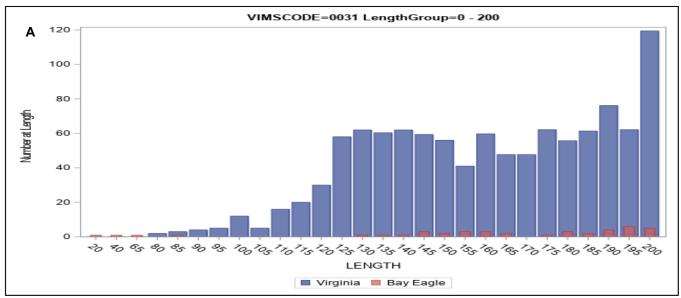
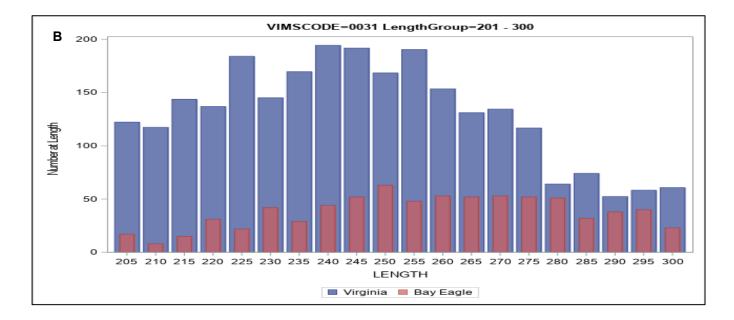
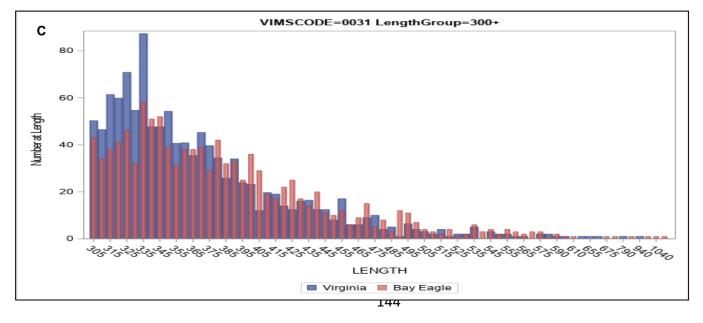


Figure 91. Comparison of length frequency distributions during calibration tows for Striped Bass 0-200mm (A), 201-30mm (B) and 300+mm (C).





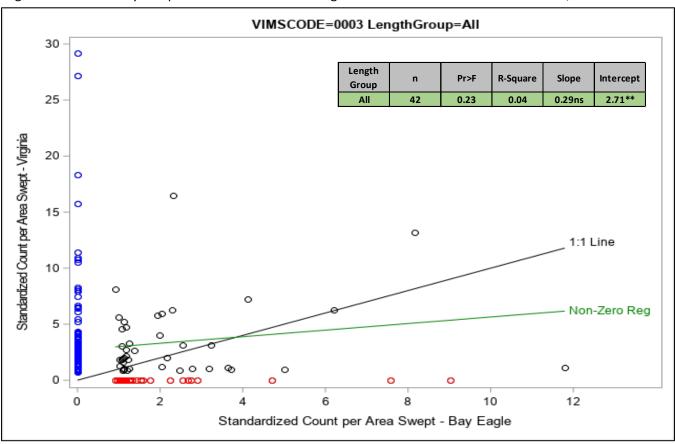
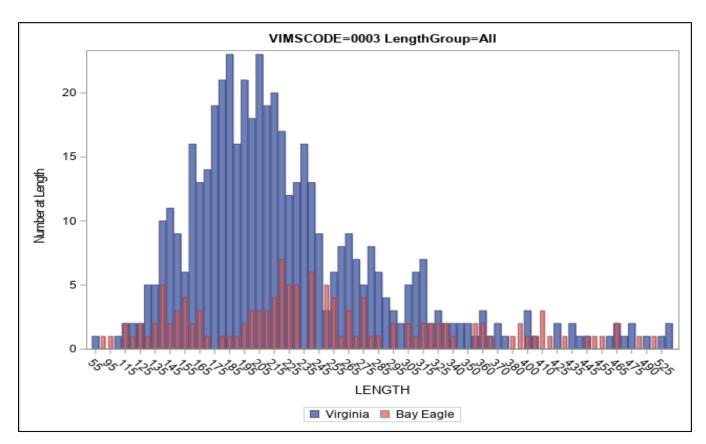


Figure 92. Preliminary comparison of catch rates during calibration tows for Summer Flounder, all sizes.

Figure 93. Comparison of length frequency distributions during calibration tows for Summer Flounder.



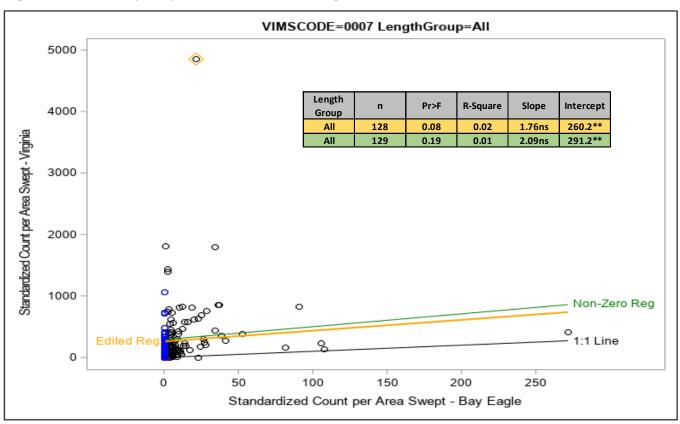


Figure 94. Preliminary comparison of catch rates during calibration tows for Weakfish, all sizes.

Figure 95. Comparison of length frequency distributions during calibration tows for Weakfish.

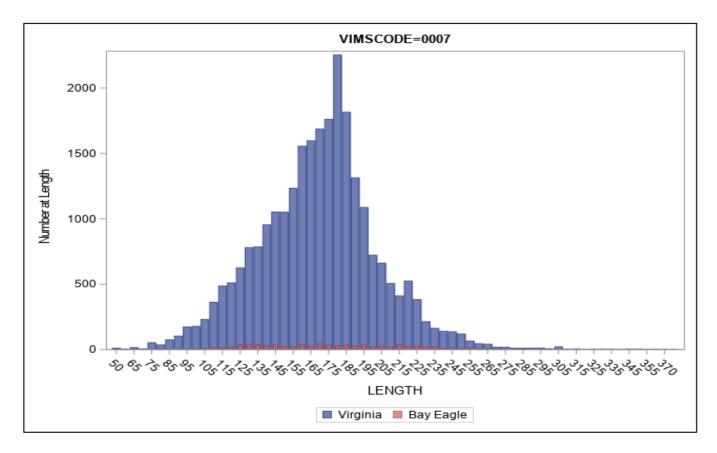


Figure 96. Preliminary comparison of catch rates during calibration tows for White Perch, 0 – 190mm.

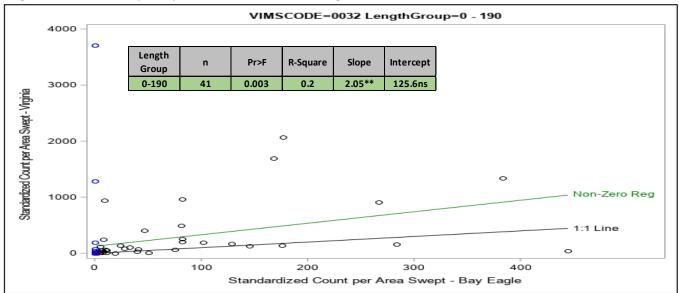


Figure 97. Preliminary comparison of catch rates during calibration tows for White Perch, 0 – 190mm.

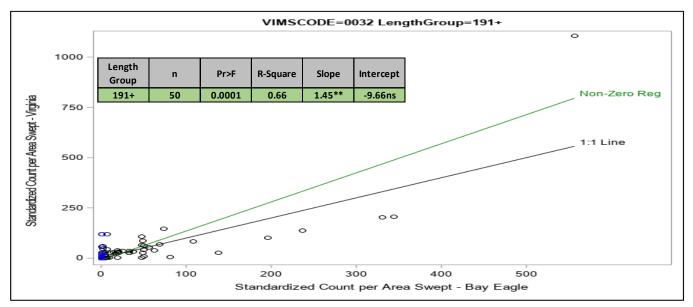


Figure 98. Comparison of length frequency distributions during calibration tows for White Perch, 0 – 190mm (A) and 191+mm (B).

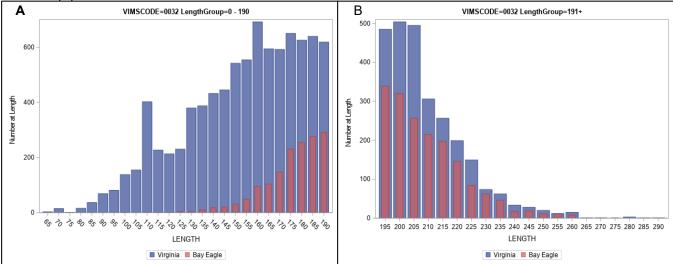
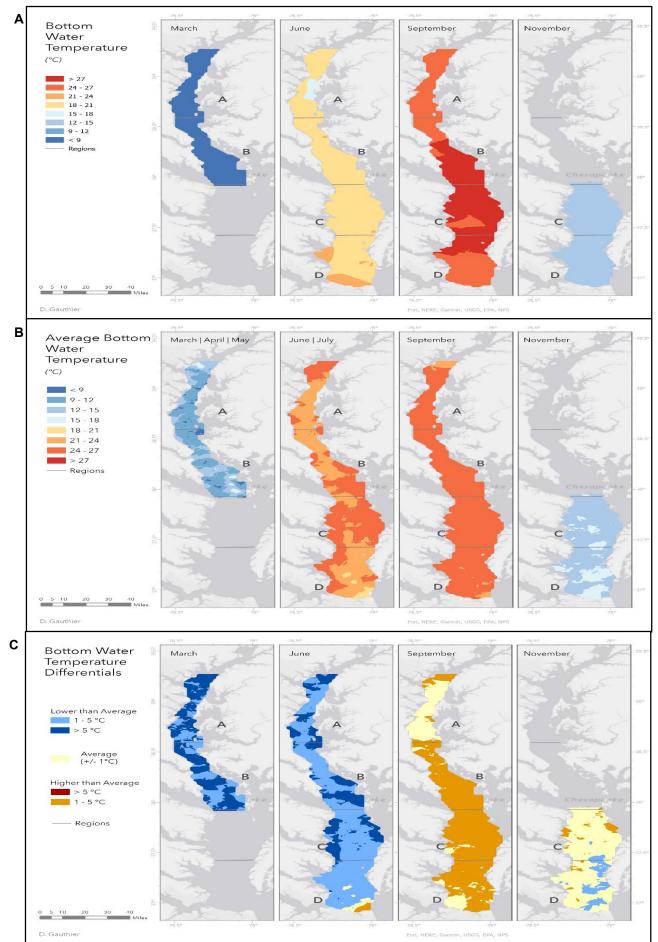
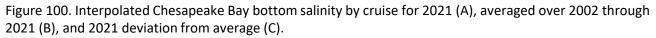


Figure 99. Interpolated Chesapeake Bay bottom water temperature by cruise for 2021 (A), averaged over 2002 through 2021 (B), and 2021 deviation from average (C).





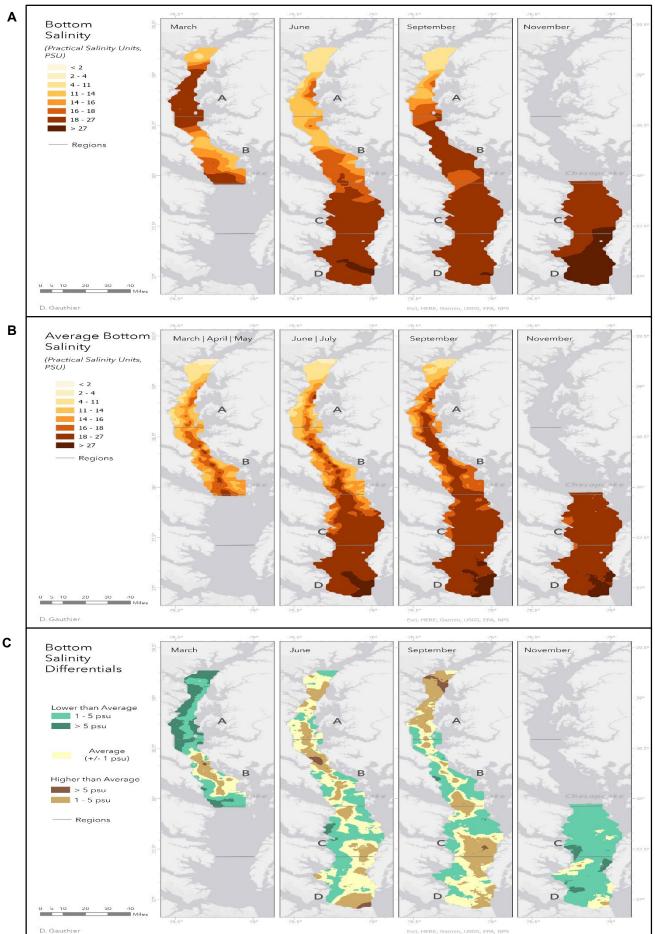
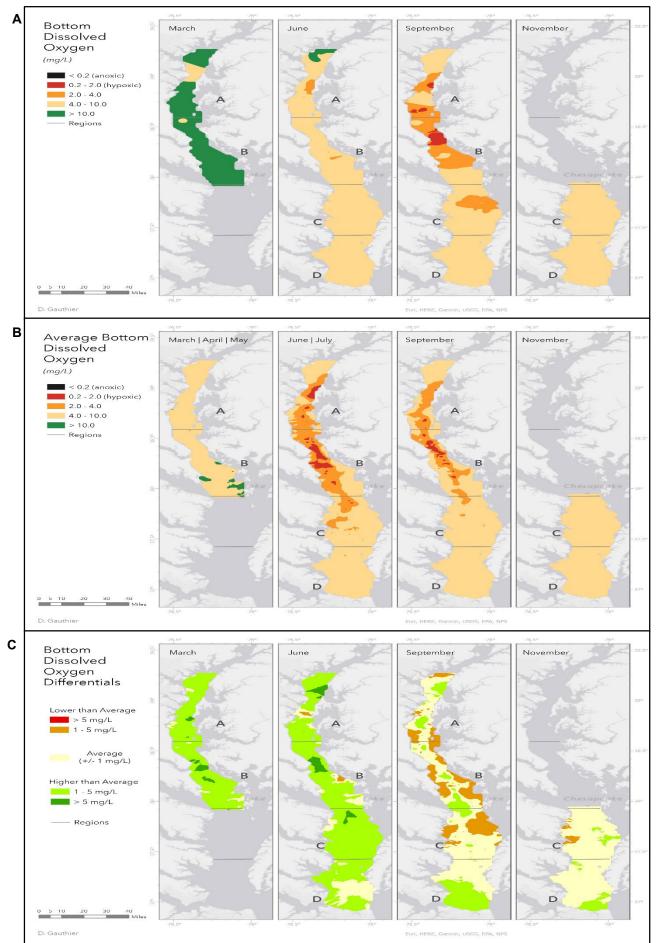


Figure 101. Interpolated Chesapeake Bay bottom dissolved oxygen by cruise for 2021 (A), averaged over 2002 through 2021 (B), and 2021 deviation from average (C).



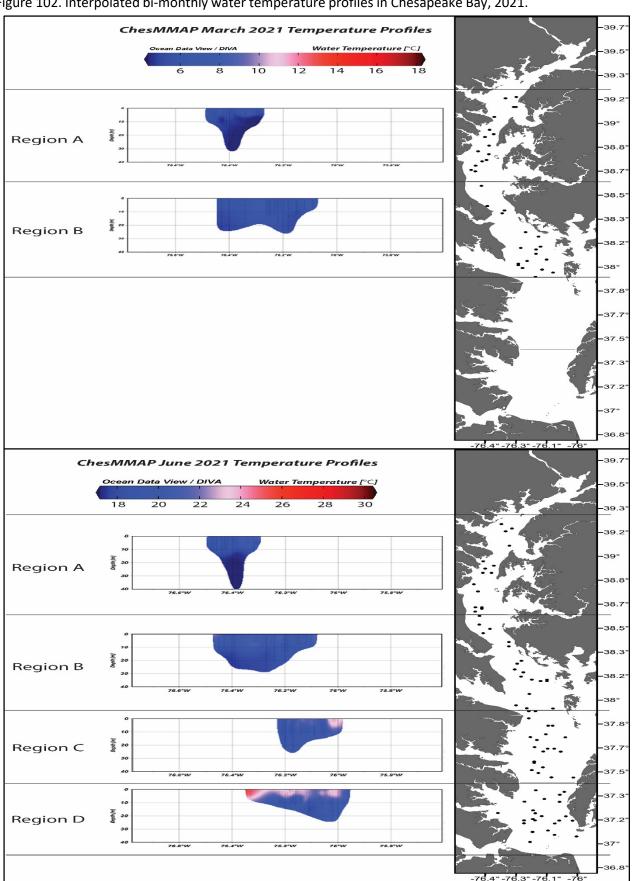
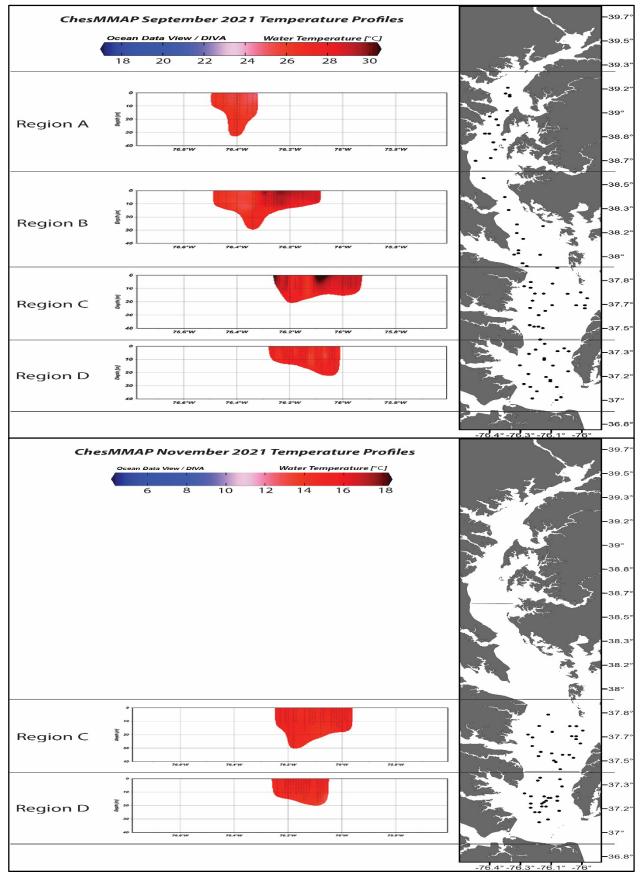
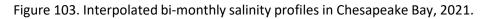


Figure 102. Interpolated bi-monthly water temperature profiles in Chesapeake Bay, 2021.





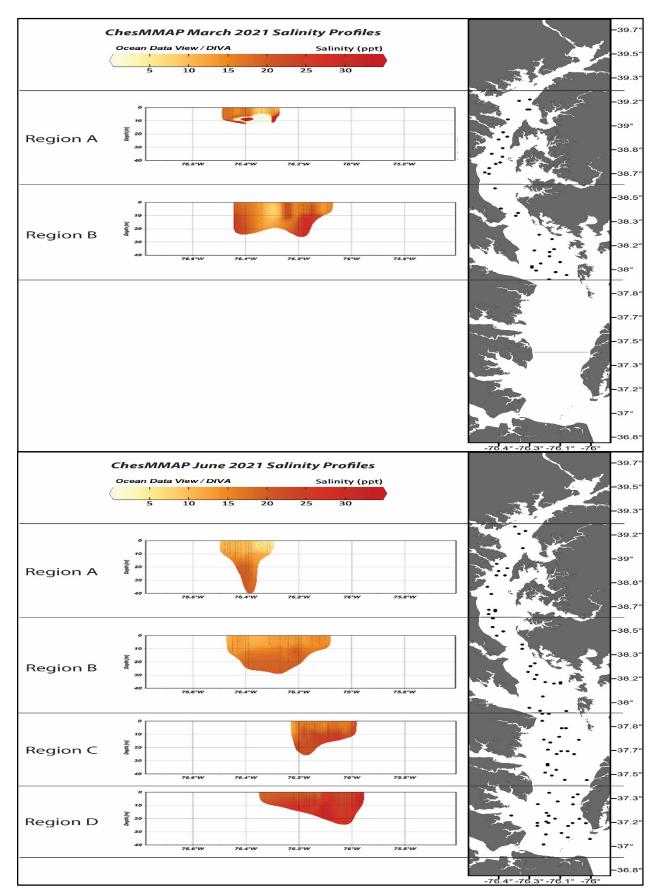
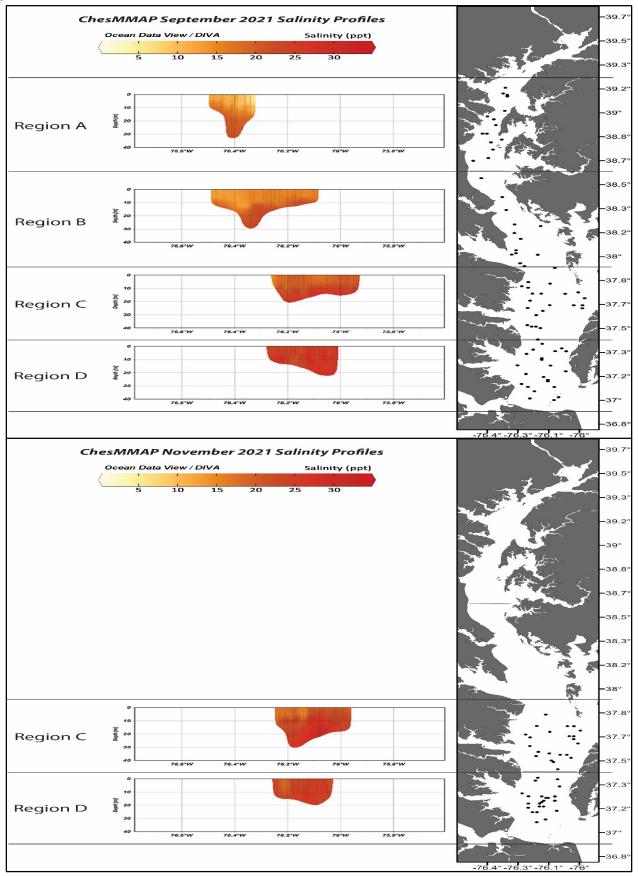


Figure 103. cont.



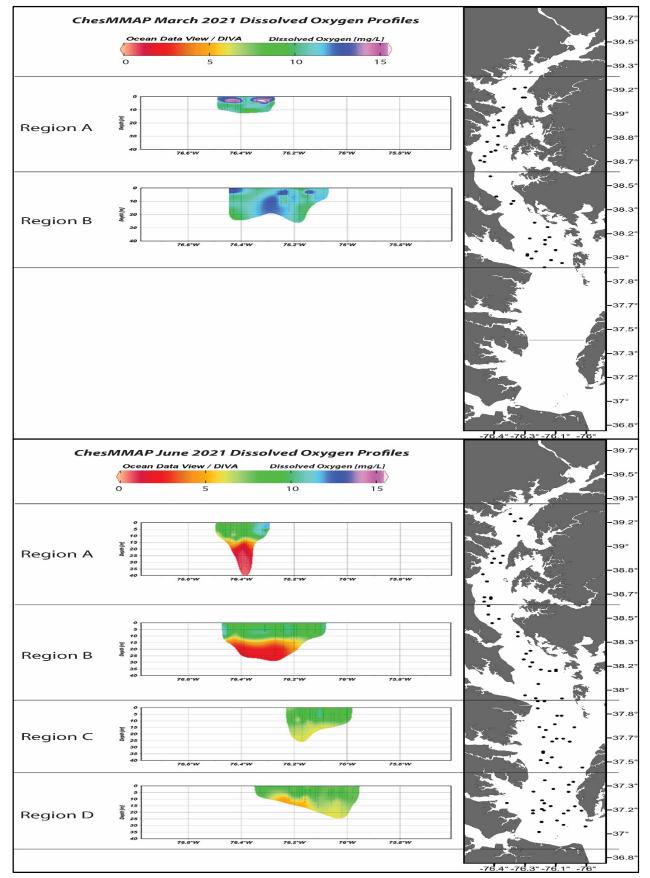
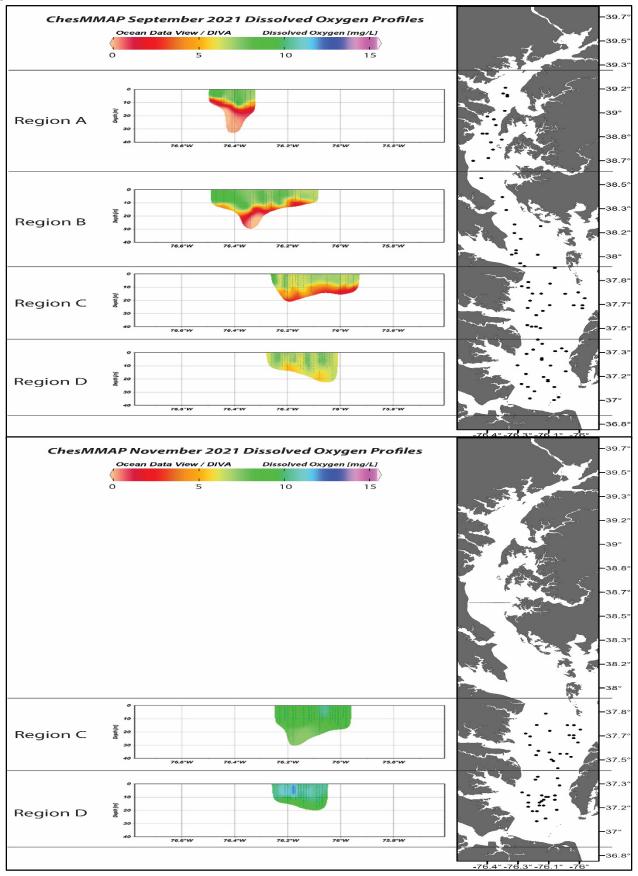


Figure 104. Interpolated bi-monthly dissolved oxygen profiles in Chesapeake Bay, 2021.

Figure 104. cont.



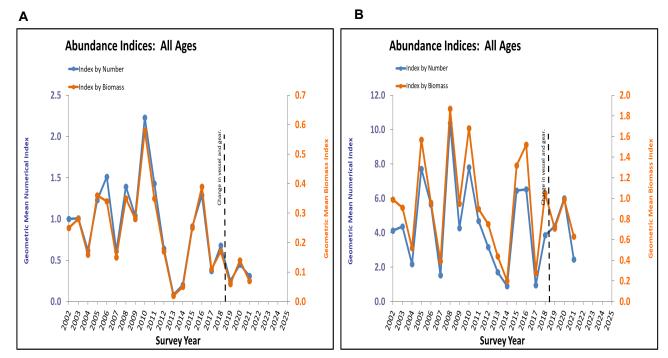
Appendix

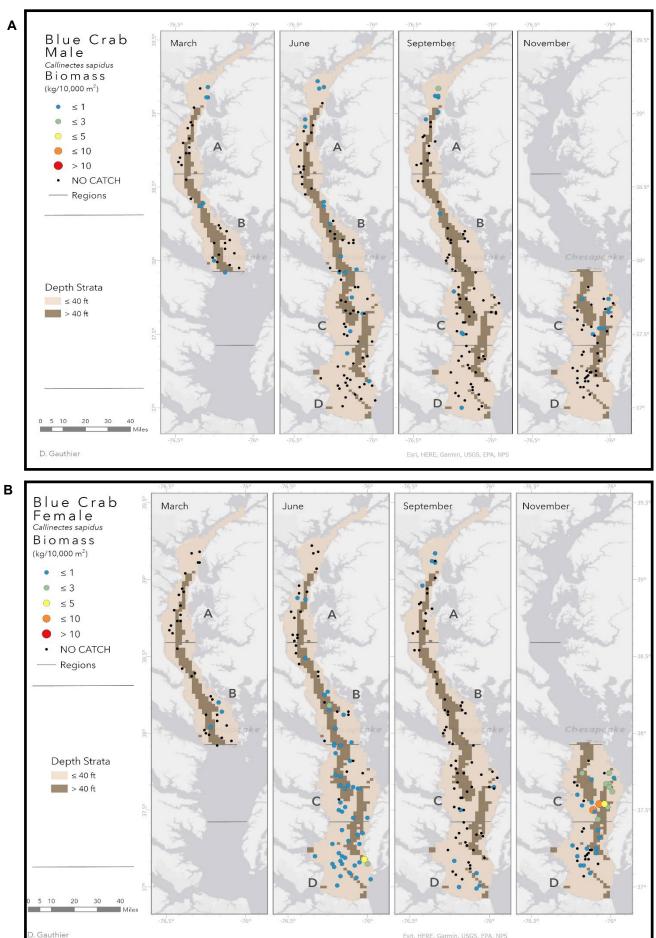
Blue Crab and Clearnose Skate Abundance

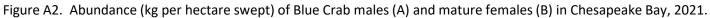
Table A1. Blue Crab male (A) and mature female (B) geometric mean indices of abundance, by number and biomass, overall.

| Α | | | | | | | | | В | | | | | | | | |
|------|-----|-----|-----------------|-------|-------|---------------|-------|-------|------|-----|----|-----------------|-------|-------|---------------|------------|------|
| Year | Age | n | Numerical Index | | | Biomass Index | | | Year | Age | n | Numerical Index | | | Biomass Index | | |
| | | | LCI | Index | UCI | LCI | Index | UCI | | | | LCI | Index | UCI | LCI | Index | UCI |
| 2002 | All | 116 | 0.740 | 1.000 | 1.290 | 0.180 | 0.250 | 0.320 | 2002 | All | 40 | 2.24 | 4.1 | 7.14 | 0.50 | 1.0 | 1.64 |
| 2003 | | 110 | 0.720 | 1.010 | 1.360 | 0.190 | 0.280 | 0.370 | 2003 | | 60 | 2.95 | 4.4 | 6.31 | 0.62 | 0.9 | 1.25 |
| 2004 | | 117 | 0.430 | 0.620 | 0.830 | 0.100 | 0.160 | 0.220 | 2004 | | 46 | 1.33 | 2.2 | 3.33 | 0.31 | 0.5 | 0.77 |
| 2005 | | 115 | 0.830 | 1.230 | 1.710 | 0.220 | 0.360 | 0.510 | 2005 | | 45 | 4.91 | 7.7 | 11.83 | 1.04 | 1.6 | 2.22 |
| 2006 | | 108 | 1.150 | 1.510 | 1.920 | 0.250 | 0.340 | 0.440 | 2006 | | 41 | 3.80 | 5.6 | 8.17 | 0.67 | 1.0 | 1.29 |
| 2007 | | 56 | 0.340 | 0.610 | 0.950 | 0.080 | 0.150 | 0.230 | 2007 | | 44 | 0.83 | 1.5 | 2.54 | 0.21 | 0.4 | 0.60 |
| 2008 | | 118 | 1.000 | 1.390 | 1.850 | 0.210 | 0.350 | 0.490 | 2008 | | 45 | 7.40 | 10.4 | 14.44 | 1.35 | 1.9 | 2.51 |
| 2009 | | 116 | 0.730 | 1.040 | 1.420 | 0.190 | 0.280 | 0.380 | 2009 | | 45 | 2.54 | 4.3 | 6.85 | 0.57 | 1.0 | 1.43 |
| 2010 | | 117 | 1.690 | 2.230 | 2.870 | 0.430 | 0.580 | 0.740 | 2010 | | 45 | 4.95 | 7.8 | 12.04 | 1.14 | 1.7 | 2.36 |
| 2011 | | 116 | 0.990 | 1.430 | 1.970 | 0.240 | 0.350 | 0.470 | 2011 | | 44 | 3.16 | 4.7 | 6.76 | 0.64 | 0.9 | 1.21 |
| 2012 | | 112 | 0.400 | 0.640 | 0.930 | 0.100 | 0.170 | 0.240 | 2012 | | 39 | 1.66 | 3.2 | 5.55 | 0.39 | 0.8 | 1.20 |
| 2013 | | 118 | 0.030 | 0.080 | 0.150 | 0.000 | 0.020 | 0.030 | 2013 | | 44 | 0.92 | 1.7 | 2.83 | 0.22 | 0.4 | 0.70 |
| 2014 | | 118 | 0.090 | 0.200 | 0.310 | 0.020 | 0.050 | 0.090 | 2014 | | 45 | 0.50 | 0.9 | 1.43 | 0.11 | 0.2 | 0.30 |
| 2015 | | 118 | 0.640 | 0.910 | 1.230 | 0.170 | 0.250 | 0.330 | 2015 | | 45 | 4.03 | 6.5 | 10.06 | 0.86 | 1.3 | 1.91 |
| 2016 | | 118 | 0.870 | 1.290 | 1.820 | 0.260 | 0.390 | 0.520 | 2016 | | 45 | 3.97 | 6.5 | 10.37 | 0.95 | 1.5 | 2.27 |
| 2017 | | 118 | 0.220 | 0.370 | 0.550 | 0.060 | 0.110 | 0.160 | 2017 | | 45 | 0.46 | 1.0 | 1.62 | 0.13 | 0.3 | 0.44 |
| 2018 | | 95 | 0.430 | 0.680 | 0.960 | 0.110 | 0.170 | 0.240 | 2018 | | 45 | 2.07 | 3.9 | 6.66 | 0.53 | 1.1 | 1.73 |
| 2019 | | 78 | 0.140 | 0.250 | 0.380 | 0.030 | 0.060 | 0.090 | 2019 | | 45 | 3.14 | 4.4 | 6.02 | 0.52 | 0.7 | 0.93 |
| 2020 | | 79 | 0.180 | 0.450 | 0.780 | 0.020 | 0.140 | 0.280 | 2020 | | 45 | 4.49 | 6.0 | 7.93 | 0.74 | 1.0 | 1.27 |
| 2021 | | 79 | 0.140 | 0.310 | 0.510 | 0.020 | 0.070 | 0.120 | 2021 | | 45 | 1.50 | 2.5 | 3.78 | 0.38 | 0.6 | 0.92 |
| 2022 | | | | | | | | | 2022 | | | | | | | | |
| 2023 | | | | | | | | | 2023 | | | | | | | | |
| 2024 | | | | | | | | | 2024 | | | | | | | | |
| 2025 | | | | | | | | | 2025 | | | | | | | | |

Figure A1. Blue Crab male (A) and mature female (B) geometric mean indices of abundance, by number and biomass.







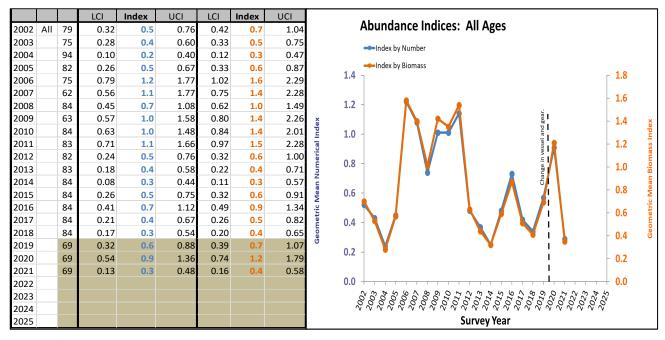


Figure A3. Clearnose Skate geometric mean indices of abundance, by number and biomass.

Figure A4. Abundance (kg per hectare swept) of Clearnose Skate in Chesapeake Bay, 2021.

