



Climate-driven interannual variability of water scarcity in food production potential: a global analysis

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Abstract. Interannual climatic and hydrologic variability has been substantial during the past decades in many regions. While climate variability and its impacts on precipitation and soil moisture have been studied intensively, less is known on subsequent implications for global food production. In this paper we quantify effects of hydroclimatic variability on global “green” and “blue” water availability and demand in global agriculture, and thus complement former studies that have focused merely on long-term averages. Moreover, we assess some options to overcome chronic or sporadic water scarcity. The analysis is based on historical climate forcing data sets over the period 1977–2006, while demography, diet composition and land use are fixed to reference conditions (year 2000). In doing so, we isolate the effect of interannual hydroclimatic variability from other factors that drive food production. We analyse the potential of food production units (FPUs) to produce a reference diet for their inhabitants (3000 kcal cap⁻¹ day⁻¹, with 80 % vegetal food and 20 % animal products). We applied the LPJmL vegetation and hydrology model to calculate the variation in green-blue water availability and the water requirements to produce that very diet. An FPU was considered water scarce if its water availability was not sufficient to produce the diet (i.e. assuming food self-sufficiency to estimate dependency on trade from elsewhere). We found that 24 % of the world’s population lives in chronically water-scarce FPUs (i.e. water is scarce every year), while an additional 19 % live under occasional water scarcity (water is scarce in some years). Among these 2.6 billion people altogether, 55 % would have to rely on international trade to reach the reference diet, while for 24 % domestic trade would be enough. For the remaining 21 % of

the population exposed to some degree of water scarcity, local food storage and/or intermittent trade would be enough to secure the reference diet over the occasional dry years.

1 Introduction

Climatic and hydrologic conditions vary considerably around the globe, both spatially and temporally (Zachos et al., 2001; Trenberth et al., 2007). Interannual hydroclimatic variability is important for many ecologic (e.g. Notaro, 2008) and societal functions (Brown and Lall, 2006; Brown et al., 2010). Although the global interannual variabilities of precipitation (e.g. Fatichi et al., 2012; Sun et al., 2012), temperature (Sakai et al., 2009) and surface wetness (Ma and Fu, 2007) are rather well understood, less is known on variability of runoff or river discharge and soil moisture at the global scale, and on the subsequent effects on availability of “blue” (i.e. freshwater in rivers and aquifers) and “green” (i.e. naturally infiltrated rain, attached to soil particles and accessible by roots) water resources for ecosystems and human societies. Not least of all, this is due to constraints in data coverage (Dettinger and Diaz, 2000; Ward et al., 2010). Recently, global hydrological models have enabled the assessment of average conditions, variabilities and trends in global runoff and discharge with greater spatial coverage (Hirabayashi et al., 2005; Piao et al., 2007; Gerten et al., 2008; Haddeland et al., 2011; Ward et al., 2014), though interannual variability has not been the main focus, except in Ward et al. (2014) who assessed the sensitivities of annual flood peaks to El Niño–Southern Oscillation (ENSO). Furthermore, meteorological

and hydrological droughts have been assessed globally, yet basically constrained to a purely hydroclimatological perspective (Sheffield et al., 2009; Dai, 2011).

Sufficiency of blue water to meet certain demands can be measured with simple scarcity indices (Falkenmark et al., 1989; Falkenmark, 1997; Vörösmarty et al., 2000; Alcamo et al., 2003; Arnell, 2004; Oki and Kanae, 2006; Kummu et al., 2010). Although blue water and irrigation are crucial for global food production, still around 60% of food is produced solely with green water resources on rainfed land (Rockström et al., 2009). Accordingly, green water contributes about 90% to agricultural water consumption (Rost et al., 2008; Liu et al., 2009; Hoekstra and Mekonnen, 2012), such that a blue-water-based analysis only captures scarcities related to irrigation (and domestic and industrial uses). The importance of green water in food production has led to a quest for integrated green-blue water (GBW) scarcity indicators. Rockström et al. (2009) found a global GBW shortage by referring to a threshold of available GBW resources of $1300 \text{ m}^3 \text{ cap}^{-1} \text{ yr}^{-1}$, which is the amount of water needed to produce a “standard diet” (Falkenmark and Rockström, 2004). Gerten et al. (2011) developed a locally specific GBW scarcity indicator by taking explicitly into account the water productivity, i.e. the amount of GBW needed to produce a benchmark for hunger alleviation ($3000 \text{ kcal cap}^{-1} \text{ day}^{-1}$, assumed to consist of 80% vegetal food and 20% animal products).

All these studies have focused on water scarcity under long-term average climate conditions. Besides, recent global studies are available that have focused on average seasonal (i.e. intra-annual) blue water scarcity (Hanasaki et al., 2008; Wada et al., 2011b; Hoekstra et al., 2012), while interannual variability (i.e. the extent to which individual years diverge from the average condition) of blue water scarcity due to climatic variation has been analysed by Wada et al. (2011a). The latter study analysed the role of climate variability in the historical evolution (1960–2001) of blue water stress. It found that increased water demand was the main factor for increased water stress, while climatic variation was often the main cause of extreme events, e.g. when prolonged dry period notably decreased the water availability. Wada et al. (2011a) fixed the population and water use for the year 1960, but still no assessment exists on how interannual hydroclimatic variability affects more present water scarcity. We expect that this effect is notable; especially since the importance of interannual climate variability for food production and underlying water resources has been highlighted recently for many regions (e.g. Haile, 2005; Tubiello et al., 2007; Zhang et al., 2010).

We argue that it is imperative to improve the understanding of the variability and frequency of water scarcity in food production, as areas exposed to occasional scarcity require essentially different response measures to overcome the food deficit than areas exposed to chronic water scarcity. Thus, quantitative knowledge of average water scarcity,

assessed for example over 30 yr (Gerten et al., 2011) or 10 yr (Rockström et al., 2009), might not reveal the areas that suffer water scarcity occasionally during dry years. On the other hand, these studies might classify areas to be water scarce although the stress is not present every year. Moreover, recently the climate variability, and related extreme climate events, have been reported to be increased in various parts of the globe (Coumou and Rahmstorf, 2012; IPCC, 2012) and their role in food production has been widely reported, e.g. in India (World Bank, 2012), Australia (Qureshi et al., 2013), and Eastern Africa (Moore et al., 2012).

In this paper we quantify the impact of interannual climatic variability on global GBW availability and requirements for food production potential, using the GBW scarcity index introduced by Gerten et al. (2011). Our analysis is based on climate forcing data for the past 30 yr (1977–2006) while diet composition, population and land cover settings are fixed to specific reference conditions, in order to assess the *isolated* effect of climate variability on GBW scarcity. We thus aim to assess how the hydroclimatic variability impacts on food producing units’ (FPUs) potential to produce a given diet for their inhabitants. Moreover, we quantify whether the variability has changed over time by comparing two climatic periods (1947–1976 and 1977–2006), and use these results to assess whether water availability, requirements and scarcity would have been significantly different given the climatic variability of the former period. All calculations are performed with the LPJmL vegetation and hydrology model (Bondeau et al., 2007; Rost et al., 2008).

2 Data and methods

In this study we assess how present water demand for food production potential is influenced by hydroclimatic variability. To cover this variability (and its changes as observed in the past), we used climatic records for the past decades. As the purpose is to assess the isolated effect of climate variability on green-blue water scarcity, we keep other variables (e.g. population, land use, diet and agronomic practices) constant at their year 2000 values. Introduced below are the methodologies and data used to conduct the study.

2.1 LPJmL model and data

The process-based, dynamic global vegetation and water balance model LPJmL (Bondeau et al., 2007; Rost et al., 2008) simulates – among other processes – water requirements, water productivities and crop yields, as well as green and blue water availabilities at a daily time step and on a global $0.5^\circ \times 0.5^\circ$ spatial grid. Specifically, it computes the growth, production and phenology of nine natural vegetation types, of grazing land, and of crops as classified into 12 “crop functional types” (CFTs). The fractional coverage of grid cells with CFTs was prescribed here using data sets

of the reference (around year 2000) cropland distribution (Ramankutty et al., 2008) and maximum monthly irrigated and rainfed harvested areas (Portmann et al., 2010). Seasonal phenology of CFTs is simulated based on meteorological conditions. Agronomic practices are calibrated for the period around 2000 by adjusting three model parameters determining the vegetation density and the maximum achievable leaf area index (LAI). This ensures that simulated yields best match those reported by FAOSTAT (see Fader et al., 2010 for details).

In LPJmL carbon fluxes and pools as well as water fluxes are modelled in direct coupling with vegetation dynamics. Possible effects of rising atmospheric CO₂ concentration on plant growth and water use efficiency could have been included, but the concentration was held constant here at 370 ppm, corresponding approximately to the year 2000 level to isolate the impact of climate (variability) on plant growth. Water requirements, water consumption (i.e. evapotranspiration, distinguished into transpiration, evaporation and interception loss) and crop water productivity (water consumption per unit of total biomass produced) are calculated for both irrigated and rainfed systems. On rainfed areas, all consumed water is green water per definition, whereas on areas equipped for irrigation, we distinguish the fractions of green and blue water. The latter is assumed to be withdrawn from rivers, reservoirs, lakes and shallow aquifers to the extent required by crops and unfulfilled by green water, considering country-specific irrigation efficiencies. We assume that the irrigation water requirements of each CFT can always be met, with implicit contributions from fossil groundwater, river diversions or other large-scale water transports (details in Rost et al., 2008; Konzmann et al., 2013). It should be noted that this assumption may lead to overestimations in some parts of the world. River flow directions are determined as in Haddeland et al. (2011), and reservoir distribution and management as in Biemans et al. (2011).

The areal distribution of CFTs and grazing land, the calibration parameters and the irrigation efficiencies are held constant at the year 2000 level throughout the simulation period. By doing so, we exclude any effects of agronomic practices and cropland expansion and thus allow for the separation of climate (variability) effects on GBW scarcity. Monthly values of temperature and fraction of cloud cover are taken from CRU TS 3.1 (Mitchell and Jones, 2005) and linearly interpolated to daily values. Monthly precipitation totals are taken from GPCC v5 (Rudolf et al., 2010; extended to cover the full CRU grid) and the number of monthly precipitation days is derived using the method from Heinke et al. (2013). Daily precipitation values are calculated from these two parameters by a statistical weather generator (Gerten et al., 2004), hence short-term droughts and their effects on yields are not necessarily captured.

2.2 Analysis and reporting scales

The model results, with resolution of 0.5°, are aggregated primarily to the FPU scale. FPUs divide the world into 281 sub-areas, being a hybrid between river basin and economic regions (Cai and Rosegrant, 2002; Rosegrant et al., 2002; De Fraiture, 2007) modified here as in Kummu et al. (2010). The final FPU set used includes 309 units with an average size of $467 \times 10^3 \text{ km}^2$ and an average population of 19.6 million people. Although the appropriate scale depends on market access and varies from region to region, we decided to focus on FPUs, as they are at a hydro-political scale within which the demand for water and food can be assumed to be managed (Kummu et al., 2010).

We introduce another three reporting scales to further analyse and present our results: countries, administrative regions, and hydrobelts. The administrative regions divide the world into 12 regions (see Fig. S1A in the Supplement) based on country borders. For this we use a regional data set originating from the UN (2000), which was further modified by Kummu et al. (2010). The hydrobelts divide the world into eight zones determined by specific hydrological characteristics and formed based on river basin boundaries (Meybeck et al., 2013). The country and administrative region scales are used for a multi-scale analysis that reveals the spatial scale of need for food storage and/or trade as an option to reach the reference diet (see Sect. 2.5).

2.3 Calculation of green-blue water availability, demand, and scarcity

The procedures to calculate water availability, requirements and scarcity are described in detail in Gerten et al. (2011) and are only briefly summarised here. Water availability is given by the sum of green water and blue water resources. The former is defined as the evapotranspiration from cropland and partly from grazing land during the growing season. Hence, green water availability depends not only on hydroclimatic conditions but also on the spatial extent of cropland and grazing land. Blue water availability is defined to be 40 % of the sum of runoff and water storage in lakes and aquifers. This 40 % threshold represents the maximum amount of water that should be withdrawn to avoid water stress (Vörösmarty et al., 2000; Oki and Kanae, 2006; Falkenmark et al., 2007). We acknowledge that a global value of 40 % does not take into account the spatial heterogeneity of, for example, the environmental flow requirements, but work is in progress to get a better, spatially explicit, understanding of this limit (Gerten et al., 2013). Moreover, while in green water availability the seasonality is taken into account by calculating crop phenology (Gerten et al., 2011), the modelled growing period may not always be realistic. A seasonal discrepancy might happen in areas where blue water is not available when the water is required and reservoir storage capacity is not adequate to buffer this (Barnett et al., 2005). Therefore, in some

occasions our approach might overestimate the blue water availability.

The total availability of GBW (i.e. green and blue water), aggregated to FPUs, is then compared with the amounts of GBW needed to produce the raw products for the reference diet of 3000 kcal cap⁻¹ day⁻¹ (with 80 % vegetal and 20 % animal-based share) for all inhabitants of the FPU, following the method developed by Gerten et al. (2011). The reference diet is based on the WHO and FAO recommended production level required to eradicate hunger (WHO, 2003; FAO, 2013c) and it includes the average food losses and waste (in terms of calories 24 %; Kummu et al., 2012). Subtracting this loss and waste from the total production gives a food consumption of around 2280 kcal cap⁻¹ day⁻¹, being almost exactly the global average dietary energy requirement (country data averaged over 2007–2012) of 2245 kcal cap⁻¹ day⁻¹ defined by FAO (2013c). We use the same reference diet, in terms of food supply, for all FPUs to follow the recommendations for hunger alleviation. It should be noted, however, that the actual calorie level and content of the diet vary from place to place. Moreover it is good to mention, that should the animal-based share be higher (lower) from the used 20 %, the pressure on water resources would also be higher (lower) due to the higher water consumption of animal-origin foodstuff (e.g. Hoekstra and Mekonnen, 2012).

Since we constrain our analysis to the effects of local hydroclimatic variability on the production of a given diet, we omit the decoupling of food production potential and consumption areas. Although this does not reflect the current patterns, which are governed by international trade of agricultural commodities and associated virtual water trade, this analysis provides relevant information due to the following reasons: (i) security of domestic food supply for a possible emergency situation remains important for many countries; (ii) many low- and medium-income countries do not have sufficiently strong economies to enter global food markets and are thus mainly dependent on domestic food production, and (iii) globally more than 80 % of the food (energy-wise) is still consumed in the country it is produced in (based on Food balance sheets; FAO, 2013b). Thus, we argue, that knowing the potential of reaching food self-sufficiency by FPUs and by countries is highly important information in assessing global and regional food security. Furthermore, by disregarding current trade patterns in the analysis, we are able to identify the dependency of FPUs on this trade.

The GBW requirements result from the crop water productivity (determined at grid cell level and influenced by climate and agronomic practices), the transpirational demand given by meteorological conditions, and the soil moisture. They are computed from both the water requirements to produce the vegetal calorie share on present cropland (represented by the CFTs) and from a provisional livestock sector. Contributions from the latter come from both grazing land and cropland (i.e. the shares used for feed production, assigned according to the scheme used in Gerten et al., 2011). The water

requirements from grazing land are computed slightly differently as compared to Gerten et al. (2011): here we weigh the green water available on each country's and FPU's grazing land according to its water productivity, while Gerten et al. (2011) uses a water productivity relative to the global average for grass.

GBW scarcity is given by the ratio between the GBW availability and the GBW requirements for producing the reference diet. A region (FPU) is considered to be GBW scarce if its domestic GBW availability falls below the GBW requirements. It is acknowledged that in the case of GBW scarcity the gap between the availability and requirement would be small for some regions while large for others, when using average climate conditions. We argue that when taking into account the interannual variability in climatic conditions, this sharp threshold is less problematic, as then the areas close to the threshold (on either side of it) are classified to be under occasional water scarcity (see more in Sect. 2.4).

2.4 Methods for analysing the variability and change in variability

As GBW scarcity is assessed on the basis of GBW requirements and GBW availability, it is important to understand the impact of climatic variability on both of them. To quantify this variability we use the coefficient of variation (CV; i.e. standard deviation divided by mean) that is comparable between different areas, and also between the three variables (GBW availability, requirements, scarcity). Further, we measure the frequency of years when an FPU in question is under GBW scarcity (i.e. when it does not have enough water to produce the reference diet). This frequency analysis allows for the classification of the FPUs into three main groups, of which occasional GBW scarcity is further divided into four sub-groups:

1. no GBW water scarcity (enough GBW to produce the reference diet in all years);
2. occasional GBW water scarcity:
 - a. sporadic GBW scarcity (GBW scarcity in 1–25 % of the years);
 - b. medium frequent GBW scarcity (GBW scarcity in 25–50 % of the years);
 - c. highly frequent GBW scarcity (GBW scarcity in 50–75 % of the years);
 - d. recurrent GBW scarcity (GBW scarcity in 75–99 % of the years);
3. chronic GBW scarcity (GBW scarcity in all years).

Moreover, analogous to former studies (see Introduction), we analyse the average GBW scarcity for each FPU using the average values of GBW requirements and availability over the past 30 yr. This reveals whether an area in question has

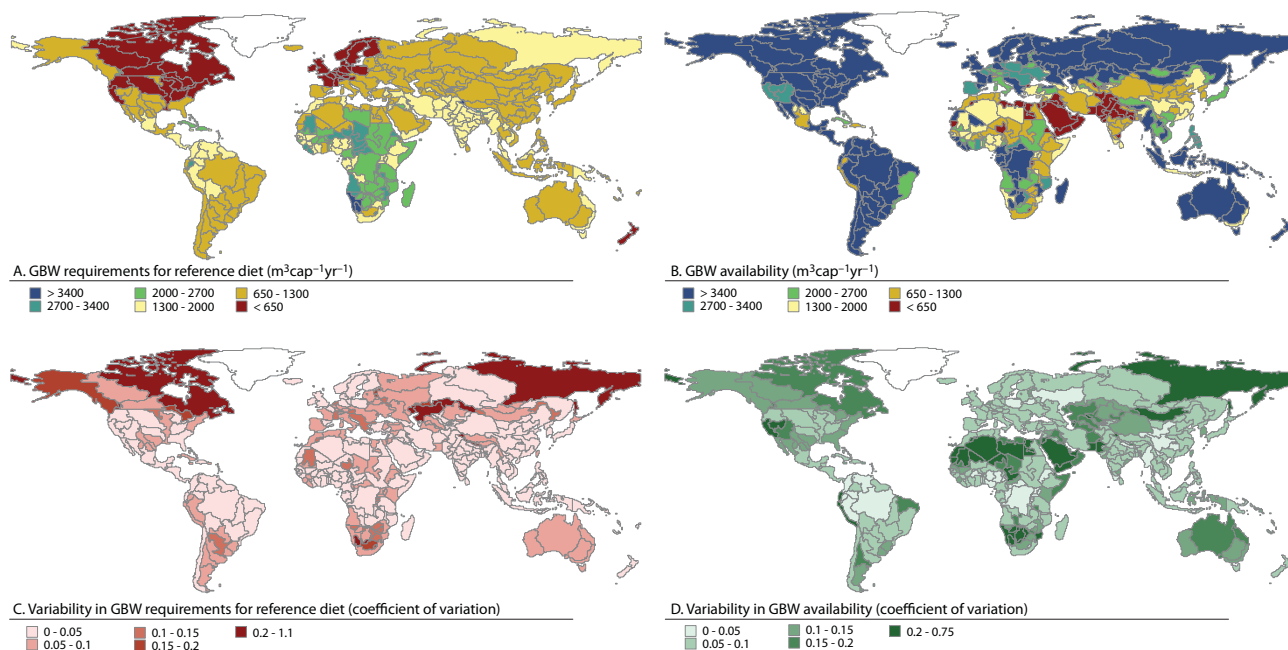


Fig. 1. Green-blue water (GBW) requirements for reference diet and GBW availability. (A) Average GBW requirements over 30 yr climatic period; (B) average GBW availability over 30 yr climatic period; (C) variability in GBW requirements measured with coefficient of variation (CV); and (D) variability in GBW availability measured with CV.

the potential to produce enough food with available water resources under long-term average climate conditions. We also investigate whether there have been changes over time in GBW scarcity, as affected by changes in both the hydro-climatic variability and average hydroclimatic conditions. In so doing, we first test with a one-way ANOVA whether the mean values of these two parameters are equal in two 30 yr climatic periods (1947–1976 and 1977–2006). Moreover, we analyse the changes in variability of GBW requirements and GBW availability within both 30 yr climatic periods. For that we use the Brown–Forsythe Levene’s test for equality of variances (Brown and Forsythe, 1974), which assesses whether there is a difference in group variances between these two periods. All statistical tests are performed with SPSS v20. We also perform the frequency analysis of GBW scarcity for both climatic periods and compare those.

2.5 Response options and stress drivers: multi-scale analysis and GBW matrix

We further conduct a multi-scale scenario analysis to scrutinise possible response measures on how each FPU under GBW scarcity could theoretically reach the reference diet. The results from GBW scarcity analyses are used at FPU, country and regional scale to identify the possible response measures as follows (see also Table 2 in Sect. 3.3):

- *Food self-sufficiency*: an FPU that does not suffer from any degree of GBW scarcity, hence without a need of measures to reach food self-sufficiency;

- *Need for local food storage and/or intermittent trade*: if an FPU is subject to occasional scarcity but is self-sufficient under average climate conditions, it would need to store food in surplus years to overcome the deficit years and/or import (export) food in deficit (surplus) years;
- *Need for domestic trade*: if an FPU is not self-sufficient under average climate conditions, but the country in which it is located is self-sufficient, it would need domestic trade to overcome the deficit years;
- *Regional and inter-regional trade*: if a country, in which an FPU is located, is not self-sufficient under average climate conditions, an FPU is classified to need either regional trade (if the region of FPU in question is self-sufficient) or inter-regional trade (if it is not).

3 Results

3.1 Interannual variability of GBW scarcity

The GBW requirements for the reference diet, calculated over the 1977–2006 period, show a distinct spatial pattern (Fig. 1a). This is due primarily to differences in climatic conditions and agronomic practices (irrigated vs. rain-fed, and other agronomic practices implicitly considered in

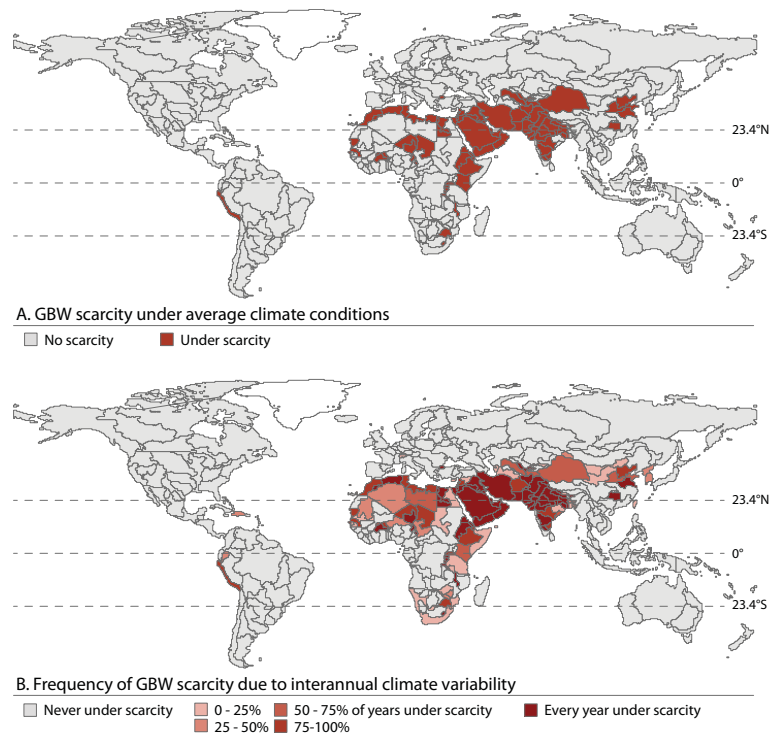


Fig. 2. Green-blue water (GBW) scarcity mapped by FPU. (A) GBW scarcity under average climate conditions over 30 yr climatic period; and (B) frequency of GBW scarcity due to interannual climate variability. The marked latitudes represent the Tropic of Cancer (23.4° N), Equator (0°), and Tropic of Capricorn (23.4° S). Note that the calculations were made for constant reference population and agronomical practices (year 2000 situation) but varying climate over the period of 1977–2006, in order to isolate the impact of climate variability on GBW scarcity.

the calibration, see Sect. 2.1). The requirements are lowest ($< 650 \text{ m}^3 \text{ cap}^{-1} \text{ yr}^{-1}$) in western Europe and large parts of North America, moderate ($650\text{--}1300 \text{ m}^3 \text{ cap}^{-1} \text{ yr}^{-1}$) in southern parts of North America, South America and large part of Asia, and highest ($> 1300 \text{ m}^3 \text{ cap}^{-1} \text{ yr}^{-1}$) in northern parts of Latin America, Africa (except the northernmost part) and South Asia (Fig. 1a). GBW availability per capita is lowest in very dry areas (e.g. North Africa, Middle East) and in highly populated places, such as in South Asia and China (Fig. 1b). High requirements often co-occur with low availability (Fig. 1a and b), resulting in water-scarce conditions (Fig. 2).

The variability in GBW requirements is mostly rather low ($\text{CV} < 0.1$) except for some areas (e.g. Canada, Siberia) where CV exceeds 0.2 (Fig. 1c). Variability of GBW availability is much higher and spatially more heterogeneous, being particularly high in dry areas, such as in North Africa, the Middle East, Australia, Central Asia and western China (Fig. 1d). The variability is, on the other hand, rather low or very low in large parts of South America, Europe, Sub-Saharan Africa, South Asia and eastern Asia (Fig. 1d).

We find that 34 % of the global population (year 2000 level) live in FPUs affected by GBW scarcity under long-term average climate conditions, i.e. the average GBW

requirements for the reference diet were larger than the average available GBW resources (Table 1; Fig 2a). When analysing the frequency in GBW scarcity due to interannual climate variability, we find that 44 % of the world's population live in FPUs under some degree of scarcity (Table 1; Fig. 2b). For about half of these people water is chronically scarce (equalling 24% of global population), while for the other half the scarcity is occasional (Table 1).

The GBW-scarce areas form a belt-like pattern from the westernmost tip of North Africa towards eastern China (Fig. 2a and b). The majority of the water-scarce FPUs in the Southern Hemisphere are located between the Equator and Tropic of Capricorn (23.4° S) while in the Northern Hemisphere they are located mostly around the Tropic of Cancer (23.4° N) (Fig. 2). The highest GBW scarcity frequencies are found in the region from the Middle East to South Asia (Fig. 2b), where the vast majority of people live under some degree of GBW scarcity (Fig. 3a; Table S2 in the Supplement; see region division in Fig. S1A in the Supplement) and 75 % in scarcity under long-term climatic conditions. In North Africa around 84 % of the population live under GBW scarcity, but only half of them in scarcity under long-term climatic conditions. According to our analysis, North America and Australia & Oceania are the only regions without GBW

Table 1. Global population (in millions) under green-blue water (GBW) scarcity. Frequency of scarcity over the climatic period of 1977–2006 and over the 30 yr climatic period before that (1947–1976). Bottom row (“average scarcity”) represents the GBW scarcity under average climate conditions within those climatic periods. See also Figs. 2 and 4e. Note that the calculations were made for constant reference population and agronomical practices (year 2000 situation) but varying climate within the indicated time periods.

Frequency	Population under GBW scarcity (in millions)			
	1977–2006	% of total	1947–1976	% of total
0 %	3471	57.4 %	3524	58.3 %
0–25 %	332	5.5 %	247	4.1 %
25–50 %	197	3.3 %	240	4.0 %
50–75 %	212	3.5 %	198	3.3 %
75–100 %	375	6.2 %	370	6.1 %
100 %	1456	24.1 %	1463	24.2 %
Total	6042		6042	
Average scarcity	2027	33.6 %	1885	31.2 %

scarcity, while FPU in Central America and Western Europe do not suffer long-term scarcity, but some are under occasional GBW scarcity (Figs. 2b and 3a).

When the results for frequency of GBW scarcity are aggregated by hydrobelts (see division in Fig. S1B in the Supplement), they reveal that in Northern and Southern Dry belts over half of the reference population suffer some degree of GBW scarcity, while in other belts less than half of the population is exposed to water scarcity (Fig. 3b). In contrast, in the boreal and equatorial belts, 0 and 3 % of population, respectively, live under occasional scarcity. The belts in the Northern Hemisphere have relatively higher GBW scarcity than their southern analogues (Fig. 3b), mostly due to their much higher population densities (Kummu and Varis, 2011; Meybeck et al., 2013).

3.2 Change in GBW scarcity over time

When comparing the 30 yr climatic period before 1977 to the climatic period of 1977–2006, we find that the average GBW requirements for the reference diet decreased statistically significantly ($p < 0.05$) in 98 FPUs, while they increased in only 13 FPUs (Fig. 4a). It should be noted, however, that because the variability in GBW requirements is relatively small (Fig. 1c), these changes mostly result from rather small absolute changes (on average ~ 5 % of the mean). The areas with significant changes are concentrated in East Asia, Africa and along the west coast of Latin and North America. Changes in the variability of GBW requirements are less pronounced and significant only for a few FPUs (Fig. 4c). The most distinct changes in average GBW availability – namely decreases ($p < 0.05$) – occurred in West Africa and Southeast Asia (Fig. 4b). In large parts of Europe and in the southern part of South America, by contrast, the second period

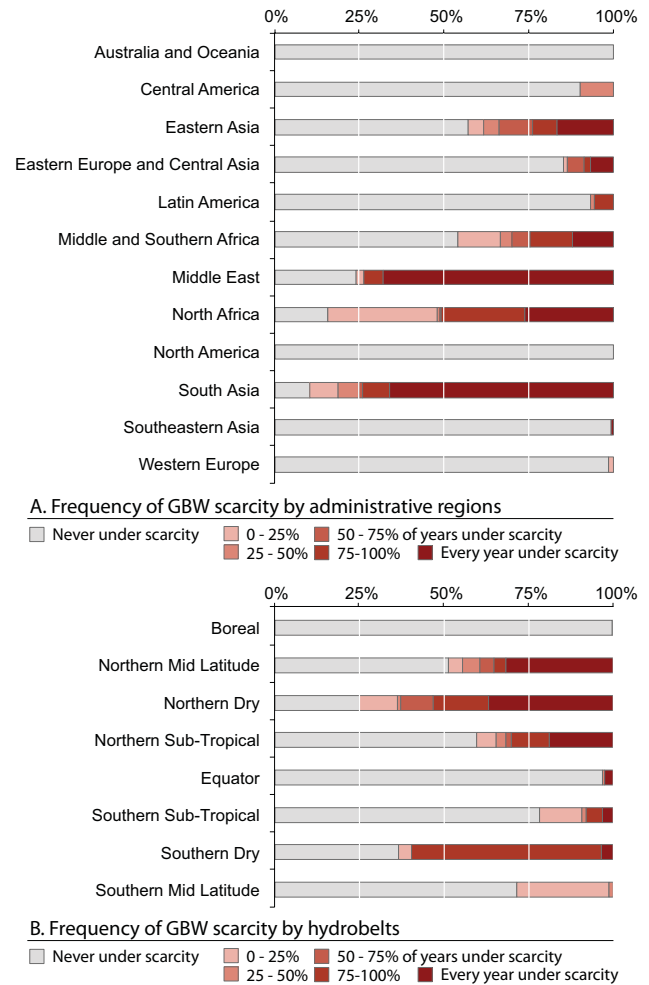


Fig. 3. Frequency of green-blue water (GBW) scarcity aggregated by (A) administrative regions, and (B) hydrobelts. See maps of the administrative regions and hydrobelts in Supplementary (Fig. S1A and S1B in the Supplement, respectively).

was wetter compared to the first one. The spatial pattern of changes, and the extent of those, in GBW variability (Fig. 4d) is less distinct than in the case of average GBW availability (Fig. 4b).

These changes in GBW requirements and GBW availability in response to changing climatic conditions are mirrored in the frequency of GBW scarcity (Fig. 4e). The frequency increases in large parts of Northern Africa and in Eastern China, while it decreases in Central Asia, in a few FPUs in Central America, and in large part of East Africa. Globally, the earlier climatic period resulted in a slightly lower population exposed to GBW scarcity (Table 1).

3.3 FPU dependency on trade and food storage

Over half of the global reference population would have the potential to reach food self-sufficiency (when measured with the reference diet) with the respective FPU’s GBW resources,

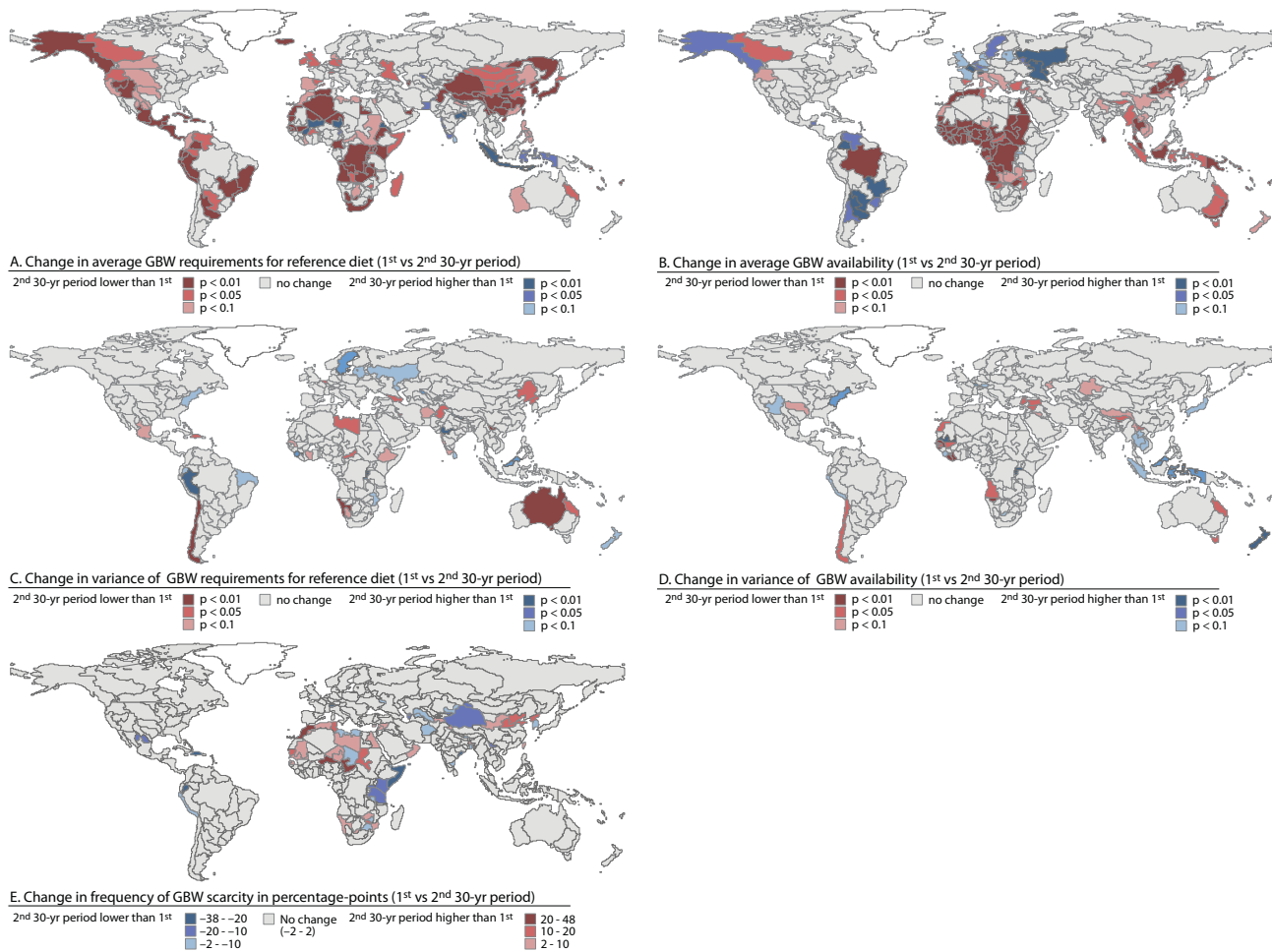


Fig. 4. Comparison of two 30 yr climatic periods (i.e. model forced by hydrometeorological data of 1947–1976 vs. 1977–2006). (A) Change in average green-blue water (GBW) requirements for the reference diet; (B) change in average GBW availability; (C) change in variance of GBW requirements; (D) change in variance of GBW availability; and (E) change in frequency of GBW scarcity. The change assessment in average values was conducted with One-way ANOVA and the change assessment in variability was conducted with the Brown–Forsythe Levene’s test. Note that the calculations were made for constant reference population and agronomical practices (year 2000 situation) but varying climate within the indicated time periods.

while for another 9 % the local storage and/or intermittent trade (import in deficit years and export in surplus years) would be enough (Table 2). The multi-scale analysis thus reveals that FPU’s inhabited by about two-thirds of the world’s population would have the potential to be independent of either domestic or international annual trade of food products if they chose to produce the reference diet for all their inhabitants on their own (and based on the products grown on their territory). Of those living in FPU’s dependent on trade (34 % of global population), 70 % depend on international or interregional trade while for the rest, 30 % domestic trade (FPU’s with 48 million people supported with national storage) would be enough to secure the reference food supply (Table 2). FPU’s depending on interregional trade are located in South Asia and the Middle East, being the only regions not having the potential to produce intra-regionally the reference

diet for all inhabitants with available GBW resources on present cropland and grazing land (Fig. 5). The FPU’s dependent on intra-regional trade are mainly located in North and East Africa and Central Asia, while the FPU’s dependent only on domestic trade (i.e. trade within a nation) are mainly located in China. In Europe, Central America and Southeast Asia, the FPU’s under occasional water scarcity would be able to reach food self-sufficiency with local storage and/or intermittent trade (Fig. 5).

4 Discussion and conclusions

Interannual climate variability has not been previously included in green-blue water scarcity studies, despite its crucial role for food production potential. By including this variability, we thus provide a notable extension to the existing

Table 2. Response options for FPU, depending on the ability of to reach reference diet of 3000 kcal cap⁻¹ yr⁻¹ (see map in Fig. 5 and FPU in GBW matrix in Fig. 6). Scarcity frequency refers to the frequency of GBW scarcity (see Fig. 2b) and average scarcity on GBW scarcity under average climate conditions (see Fig. 2a) over our study period of 1977–2006.

	FPU			Country			Region		Population	
	0 %	> 0 %	> 0 %	0 %	> 0 %	> 0 %	no	yes	10 ⁶	%
Scarcity frequency										
Average scarcity	no	no	yes	no	no	yes				
No response needed	X								3471	57.4 %
Local food storage and/or intermittent trade		X							544	9.0 %
Domestic trade			X	X					563	9.3 %
National storage and domestic trade			X		X				48	0.8 %
Regional trade			X			X	X		266	4.4 %
Inter-regional trade			X			X		X	1163	19.0 %

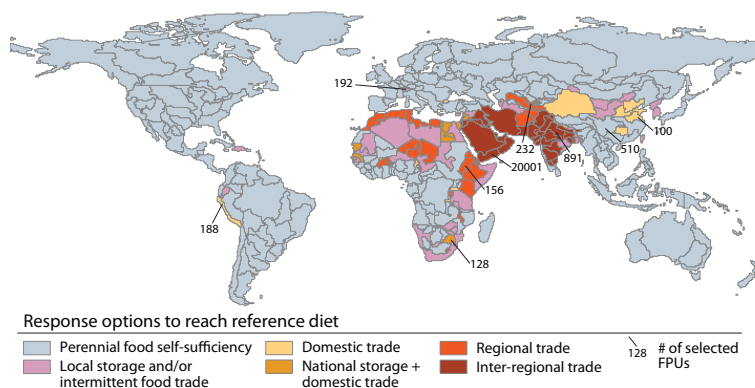


Fig. 5. FPU response options that would be needed to reach the local reference diet of 3000 kcal cap⁻¹ day⁻¹ (see Table 2 for more information). Selected FPUs are identified with their code; see Fig. 6 for their location in GBW matrix.

methodology, concept and knowledge on GBW scarcity and its impact on food production potential across the globe, as further discussed below. Our calculations indicate that 43 % of the planet's population (relative to the year 2000 reference population) dwell in areas characterised by at least some level of GBW scarcity, i.e. the GBW availability is insufficient for producing the reference diet in the respective FPUs.

4.1 Uncertainties and evaluation of results

In this study we analysed the interannual variability in global water scarcity, as imposed by climatic variability, over several decades. We used the GBW scarcity indicator developed by Gerten et al. (2011), measuring the ability to produce a reference diet of 3000 kcal cap⁻¹ yr⁻¹ within a given spatial unit, e.g. an FPU, based on current agricultural areas, agronomic practices and population levels (fixed for a reference period to separate the climate-only effect) with available GBW resources. Prior to our study, the impact of interannual climate variability on global-scale water scarcity was assessed by Wada et al. (2011a), who used the year 1960 as a reference year to isolate the climatic impact from the anthropogenic impact of the trend in blue water stress during

the period 1960–2001. In our study we used the year 2000 as a reference year and thus, we assessed the impact of interannual climate variability on considerably more recent conditions compared to Wada et al. (2011a). Furthermore, we considered not only blue water but also green water for computing water requirements and availabilities for food production, resulting in GBW scarcity estimations. Moreover, local crop water productivities were accounted for in calculating the GBW demand for the reference diet. Our analysis further extends the current knowledge about water scarcity by assessing the interannual frequency of GBW scarcity (i.e. whether an area suffers from occasional or chronic water scarcity) instead of capturing only the average scarcity as resulting from longer-term average conditions, as was done in earlier studies (Vörösmarty et al., 2000; Alcamo et al., 2003; Arnell, 2004; Oki and Kanae, 2006; Falkenmark et al., 2009; Rockström et al., 2009; Kummu et al., 2010; Gerten et al., 2011).

The LPJmL model used here has been comprehensively validated in terms of biogeochemical, agricultural and hydrological simulations (Gerten et al., 2004; Bondeau et al., 2007; Rost et al., 2008; Biemans et al., 2009; Fader et al., 2010). To evaluate our present results, we compared the computed GBW availability (Fig. 1d) to other studies. Our findings

accord with the few interannual variability analysis using surface observations of streamflow (Dettinger and Diaz, 2000) and precipitation (Fatichi et al., 2012). Fatichi et al. (2012) further used three gridded data sets (NCEP-NCAR, ERA-40, and GPCC Full Reanalysis) to assess the interannual variability in precipitation; the patterns of the discharge variability found here agree rather well with those, except for eastern Siberia where we suggest a higher variability in GBW availability (Fig. 1d) as compared to rather low precipitation variability. It should be noted, however, that although precipitation is the key driver in GBW availability, other factors are relevant as well (e.g. Gerten et al., 2008), such as the modelled duration of crop growing periods.

We found that the GBW availability has changed over time due to changes in climate. The patterns of changes in the variability (Fig. 4d) of GBW availability were rather similar to the trend in the variability of precipitation as compiled from seven databases for the years 1940–2009 (Sun et al., 2012). Further, the changes in mean GBW availability (Fig. 4b) were similar to observed trends in precipitation (IPCC, 2013) and modelled trends in blue water availability, i.e. river discharge (Gerten et al., 2008). It should be noted, however, that our timescale was somewhat different to these studies and we only mapped statistically significant changes.

The GBW requirements for the reference diet were also found to change over time and are significantly lower in many areas during climatic period of 1977–2006 than in 1947–1976 (Fig. 4a), although the absolute change between these periods was not very large. Spatially explicit identification of the underlying mechanisms requires further analyses. However, the variability in GBW requirements did not change significantly between these two periods (Fig. 4c). The climate-driven changes in GBW availability and GBW requirements did not drastically alter the global number of people under GBW scarcity (Table 1), although local changes occurred (Fig. 4e).

Our analysis suggests that North America and Australia & Oceania were the only regions without GBW scarcity (Figs. 2b and 3a). This might be counter-intuitive at first glance, given the drought proneness of both areas, particularly Australia (Qureshi et al., 2013). This is probably due to two factors: (i) both areas are very large food exporters (FAO, 2013a) and it seems that even during the driest years the FPU would have enough GBW resources to produce the reference diet for local population; and (ii) the FPU scale might, in some cases, be too large to identify local scarcity. Indeed, our method was not able to pick up very local, but still notable, GBW scarcity due to the used analysis scale, namely FPU. We do believe, however, that FPU is an appropriate scale to analyse demand for water and food. Thus for more local-scale analysis, a finer analysis scale should be used in future studies.

In summary, it should be noted that our findings are subject to the model assumptions and parameterisations, especially regarding the calculation of yields, which, however,

were calibrated here to minimise such biases. The forcing data used also exert strong influences on yield and water resources assessments (Biemans et al., 2009). Given these restrictions, we recommend interpreting our results only at regional and global scales. To increase the robustness of our results, multiple forcing data, and even multiple models, could be used. It should be further noted that we performed the calculations for a single year's reference population, cropland extent, and agronomic practices to reveal the impact of climate variability on GBW scarcity. Hence, we recommend that the diverse anthropogenic effects on GBW scarcity, and the historical trajectory of it, be investigated separately.

4.2 Response options beyond storage and trade

To facilitate the discussion of possible options to reach potential food self-sufficiency at FPU level, we designed the concept of a GBW matrix (see Fig. S2 in the Supplement), inspired by the blue water scarcity matrix developed by Falkenmark (1997). In the GBW matrix the GBW requirements for the reference diet (y axis) are plotted against the GBW availability ratio (x axis; i.e. GBW availability vs. GBW requirements). The matrix thus also reveals the “distance” of each FPU to the scarcity threshold (Fig. 6). When plotting all FPUs grouped by response options (Table 2) in the GBW matrix, we see that the majority (90%) of the FPUs that depend on international trade have GBW requirements larger than $1300 \text{ m}^3 \text{ cap}^{-1} \text{ yr}^{-1}$ (Fig. 6). Thus, the possible response options for these FPUs, and for others with high GBW requirements, would involve improvements in agronomic practices. This would result in higher crop yield per used GBW resources. Another response option would be cropland expansion.

Indeed, Fader et al. (2013) found that many countries would possess the potential to produce the required food on their own – even under increasing population in the future – if agronomic practices would continuously improve at current rates. Yet, in some countries the local food self-sufficiency would require the expansion of their cropland. Cropland expansion, however, introduces notable challenges to environmental sustainability (Wirsenius et al., 2010; Tilman et al., 2011) and the potential has been known to be quite limited in most parts of the world already over several decades (Kendall and Pimentel, 1994; Pfister et al., 2011). Foley et al. (2011) conclude that when better agronomic practices (closing yield gaps and increasing cropping efficiency) are combined with shifting diets and reducing waste, global food security could be increased considerably. Political priorities related to food self-sufficiency should be brought into the picture as well, we argue; otherwise such discussions remain somewhat theoretical.

The FPUs under GBW scarcity with moderate GBW requirements (i.e. $< 1300 \text{ m}^3 \text{ cap}^{-1} \text{ yr}^{-1}$) (Fig. 6) have fewer response options, as their level of agronomic practices is already rather high. For these areas the option to ensure

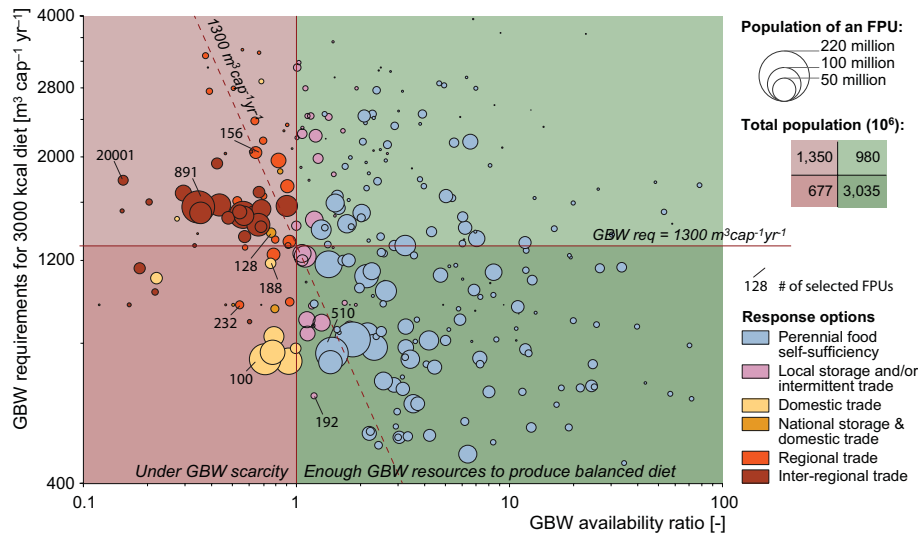


Fig. 6. FPUs mapped in the GBW matrix, grouped by response options. The location of an FPU is based on green-blue water (GBW) requirements for the reference diet (y axis) and GBW availability ratio (x axis; i.e. GBW availability vs GBW requirements) under average climate conditions. Thus, the FPUs with response option of “local storage and/or intermittent trade” are under occasional scarcity although not under average GBW scarcity. The dashed red line represents the threshold for GBW scarcity in Rockström et al. (2009), i.e. GBW availability of $1300 \text{ m}^3 \text{ cap}^{-1} \text{ yr}^{-1}$, and thus provides a comparison to their method. See Table 2 for more information of these options, and Fig. S2 in Supplement for the concept of the GBW matrix. The selected FPUs are identified with their code; see Fig. 5 for their location in a map. Note: log10 scale in both axes.

production of the reference diet (from a GBW resources view) would be either expanding the croplands (to get access to the green water on these areas) or transferring blue water from elsewhere. Long-distance water transfer is already happening in many parts of the world, including China where water is being diverted from the water-rich south to the water-scarce north (Liu and Zheng, 2002). Northwest China and North China Plain are the areas where most of the country’s FPUs under GBW scarcity are located (Fig. 2b). Indeed, in these areas cropland expansion is no longer feasible (Liu et al., 2005; Pfister et al., 2011).

To verify some of our response option findings we reflected our results to an assessment of countries’ food availability and their trade dependency to meet the food energy consumption over the period of 1965–2005, conducted by Porkka et al. (2013). Their findings reveal, for example, that the food availability has increased notably in many of the areas facing GBW scarcity in our analysis (e.g. North Africa and Middle East) since 1965. Many of the regions’ countries have been able to raise their food availability from critically low ($< 2000 \text{ kcal cap}^{-1} \text{ day}^{-1}$) to adequate level ($> 2500 \text{ kcal cap}^{-1} \text{ day}^{-1}$) – mainly with the help of high food imports. At the same time, some other countries suffering from GBW scarcity (e.g. in Eastern Africa, South Asia – i.e. countries with generally lower gross domestic production (GDP) than in the above-mentioned areas) have not managed to lift the food availability to an adequate level. Porkka et al. (2013, p. 7) indeed concluded that “average per cap GDP in countries that achieve sufficient food supply

by imports was approximately tenfold compared to countries with insufficient food supply and production”. Therefore, it is very much an economical issue whether a country can use the trade as a response option to overcome insufficient food production. When we compared our GBW scarcity results (Fig. 2b) with the food trade results of Porkka et al. (2013, p. 6), we found that all the countries in which the GBW-scarce FPUs are located were net food importers in year 2005.

4.3 Further research directions

In this study we kept the land cover, population, diet composition and agronomic practices constant in order to trace the sole impact of climate variability on GBW scarcity. While our study revealed new and interesting findings, follow-up studies should examine each year’s actual water limitations, accounting for historic land use, and the dynamics in population and variability in diets. Such comprehensive historical assessments would provide important insight on how cropland extension (and thus increased green water availability) has been linked with population growth and migrations over time. Moreover, such studies could analyse in more detail the historical record on the impact of climate variability – and climatic extremes, namely droughts – on water scarcity in individual years and seasons. This would increase the understanding of the extent and character of the required responses and their evolution over time, such as the presently soaring trend in virtual water trade (Carr et al., 2013) and increasing

dependency of many countries on agricultural trade to secure their food supply (Porkka et al., 2013). Moreover, the impact of producing non-food crops, such as fibres, should be accordingly analysed, since their role in many water-scarce FPU is substantial.

We used FPU as an analysis scale, as they are a hydro-political scale within which the demand for water and food can, in general, be assumed to be managed. In future studies the impact of resolution on the results should be studied. To find a single suitable resolution for a global scale might be difficult however, as the scale depends on food markets and infrastructure. While in some parts of the globe the markets are still very local, increasingly the food is coming from further away. This is particularly the case with large cities, and in their case the use of a very rough scale (e.g. 0.5°) could result in misrepresented conclusions.

Many FPU that are under GBW scarcity today, or are approaching that, are facing rapid population growth and thus, the situation can be expected to become even more challenging in the near future (Gerten et al., 2011; Fader et al., 2013; Suweis et al., 2013). It might actually already be so, since 14 yr have passed since the conditions of our population reference year. The projected climate change, as well as population growth, thus add another stress dimension. This is particularly the case in regions such as the Middle East, Northern and Southern Africa, and parts of Australia. Therefore, it would be important to assess how the projected increase in hydroclimatic variability in the future (e.g. Boer, 2009; Wetherald, 2010) might impact on the frequency of GBW scarcity. As the future climate would concur with population projections, showing rapid growth in many areas under water scarcity (UN, 2011), much more FPU would turn from water abundant to water scarce (as suggested by Arnell, 2004; Arnell et al., 2011 for the river-basin scale).

Our analysis reveals the theoretical potential of FPU to reach food self-sufficiency, or if that cannot be reached, the dependence level on either domestic or international food trade. We thus encourage the linking of our approach to investigations on national and regional food policies across the globe in order to bridge the calculations of theoretical potentials to actual policy level priorities for meeting the demand for food.

Supplementary material related to this article is available online at <http://www.hydrol-earth-syst-sci.net/18/447/2014/hess-18-447-2014-supplement.pdf>.

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