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# The cation exchange behavior of tylosin in loess-derived soil

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#### 22 ABSTRACT

23

24 metabolic inefficiencies, it enters the environment through manure applications. Ion exchange is 25 an important retention mechanism for Tyl, particularly for smectite clay. The objectives of this study are to characterize the exchange interactions of Tyl with common soil cations in subsoil 26 27 horizons that contain smectite and to investigate the interactions using in situ Fourier transform 28 infrared (FTIR) spectroscopy. Adsorbed Tyl in pH neutral, smectitic subsoil horizons is divided 29 into exchangeable and nonexchangeable forms. The percentage of adsorbed Tyl that is 30 exchangeable varies from 36% to 43% when Na<sup>+</sup> is the competing cation, and from 57% to 66%when Ca<sup>2+</sup> competes. In NaX-TylX binary exchange systems, neither Na<sup>+</sup> nor Tyl<sup>+</sup> is preferred 31 by the clay exchange phase, and the Vanselow selectivity coefficients ( $K_V$ ) for the NaX $\rightarrow$ TylX 32 33 exchange reaction range between 0.79 and 1.41. In the CaX<sub>2</sub>-TylX systems, Tyl<sup>+</sup> is preferred by 34 the clay exchange phase when the equivalent fraction of TylX ( $E_{TylX}$ ) is less than 0.4. The  $K_V$ 35 values for the CaX<sub>2</sub> $\rightarrow$ TylX exchange reaction are at a maximum at the lowest  $E_{\text{TylX}}$  values, with 17.6  $\leq K_V \leq$  58.1, then decrease with increasing  $E_{TylX}$  to 1.34  $\leq K_V \leq$  6.28. Adsorbed Tyl masks 36 37 the CEC of the soil clays; the effect is greatest in systems that are initially Tyl-saturated, and is 38 attributed to the steric effects of the large Tyl molecule. In situ FTIR indicates that Tyl interacts 39 with soil iron oxides through the dimethylamine moiety. 40 Keywords: Antibiotic, exchange selectivity, Vanselow, x-ray diffraction, ATR-FTIR 41 Abbreviations: ATR-FTIR, attenuated total reflectance Fourier transform infrared spectroscopy; 42 CEC, cation exchange capacity; DDI, distilled-deionized water; HATR, horizontal attenuated 43 total reflectance; HPLC, high pressure liquid chromatography, Tyl, tylosin; VM, Visual 44 **MINTEQ** 

Tylosin (Tyl) is a veterinary antibiotic commonly used in swine and poultry production. Due to

2

#### 45 **1.** Introduction

46 Tylosin (Tyl) is a broad-spectrum macrolide antibiotic first isolated from a strain of 47 Streptomycetes fradiae in soil from Thailand (Hamill et al., 1961). The antibiotic is commonly 48 used in swine and poultry production as a prophylactic and for the treatment of microbial 49 infections. It is also used as a dietary supplement to improve feed efficiency and to enhance 50 growth (Kumar et al., 2005). Veterinary-grade Tyl is sold as either a tartrate or phosphate salt 51 (Tylan®, Elanco, Greenfield, IN), which also contains minor amounts of other antimicrobial 52 agents, including desmycosin, macrocin, relomycin, and desmycinosyl tylosin (Fish and Carr, 53 1986). Tylosin consists of a 16-membered lactone ring with linked mycaminose and mycarose 54 sugars radiating from ring position 5, and the sugar mycinose attached at position 14 (Fig. 1). 55 The compound is relatively hydrophilic, with a water solubility of approximately 5.5 mmol  $L^{-1}$ 56 and an octanol-water partition coefficient (log Kow) of 1.15 to 2.50 (McFarland et al., 1997; 57 Wollenberger et al., 2000; Loke et al., 2002). The dimethylamine functional group is protonated 58 in neutral to acidic solutions ( $pK_a = 7.50 \pm 0.13$ ) (Qiang and Adams, 2004). 59 Animal manure application to agricultural lands is the primary mode of Tyl entry into the 60 environment (Joy et al., 2013; Soni et al., 2015; Luby et al., 2016; Sura et al., 2016; Dungan et 61 al., 2017). An important retention mechanism of Tyl by soil and soil minerals is cation exchange, 62 which occurs through the protonated dimethylamine functional group. Several studies have 63 illustrated a correlation between soil clay content and adsorption intensity and capacity, as well 64 as the influence of background electrolyte type and concentration on Tyl adsorption. Rabølle and 65 Spliid (2000) studied Tyl adsorption by soils that varied in clay content, cation exchange 66 capacity (CEC), and organic C and observed adsorption intensity to increase with increasing clay 67 content. They also noted that Tyl adsorption was not correlated to soil CEC or organic C. Similar 68 findings have been reported by others (Chander et al., 2005; Allaire et al., 2006; ter Laak et al., 69 2006). In addition, ter Laak et al. (2006) observed Tyl retention by clay loam and loamy sand 70 soil to decrease with increasing pH and with increasing ionic strength. Sassman et al. (2007) 71 noted that Tyl adsorption intensity was strongly correlated to clay content, CEC, and surface 72 area, but only weakly correlated to organic C content. They also observed the retention of Tyl to 73 be greater in smectite dominated soils relative to those dominated by kaolinite, and that NH<sub>4</sub> 74 acetate could provide the nearly complete removal of adsorbed Tyl when added to a methanol 75 extractant. In addition, Lee et al. (2014), Srinivasan et al. (2014), and Stromer et al. (2018) have 76 all illustrated the strong correlation of Tyl adsorption to the soil properties and experimental 77 conditions that influence ion exchange: greater retention in smectite dominated soils relative to 78 soils dominated by metal oxides and kaolinite; increased retention with decreasing pH and ionic 79 strength; and greater adsorption when the ionic strength is controlled by monovalent (e.g., 80 NaNO<sub>3</sub>) relative to divalent salts (e.g.,  $Ca(NO_3)_2$ ).

81 Bewick (1979) observed higher Tyl adsorption by Na- and Ca-saturated montmorillonites 82 relative to illite and kaolinite, and attributed this finding to the expansiveness of the 83 montmorillonites. It was also shown that the displacement of Tyl from the exchange complex by Ca<sup>2+</sup> was more difficult than the displacement of Na<sup>+</sup>. Essington et al. (2010) and Zhang et al. 84 85 (2013) examined the adsorption of Tyl by montmorillonite and kaolinite as a function pH, ionic 86 strength, and the dominant electrolyte cation. Tylosin adsorption was greater in montmorillonite 87 relative to kaolinite systems, decreased with increasing pH and ionic strength, and was greater in 88 NaNO<sub>3</sub> systems relative to Ca(NO<sub>3</sub>)<sub>2</sub> systems. Further, Tyl retention as a function of pH could be 89 modeled by considering only outer sphere surface complexation. Zhang et al. (2013) also 90 illustrated the intercalation of Tyl into montmorillonite using x-ray diffraction.

91 The hypothesis that ion exchange is an important retention mechanism for Tyl in soil was 92 directly tested by Call et al. (2019). They examined the exchange selectivity of Tyl relative to Ca<sup>2+</sup> and with Na<sup>+</sup> in reference vermiculite and montmorillonite systems using binary exchange 93 94 isotherms. They found Tyl to intercalate montmorillonite, but not vermiculite, and that >41% of 95 the Tyl adsorbed by montmorillonite was exchangeable in systems dominated by Na<sup>+</sup>, decreasing to >12% in systems dominated by Ca<sup>2+</sup>. The exchange isotherms also demonstrated that Tyl was 96 97 preferred by the smectite exchange phase relative to both Na<sup>+</sup> and Ca<sup>2+</sup>. The importance of cation 98 exchange as a retention mechanism for Tyl was confirmed using infrared spectroscopy. It was 99 also shown that polar and covalent surface interactions through carbonyl functional groups were 100 responsible for the nonexchangeable forms of adsorbed Tyl.

101 In this study, the selectivity of adsorbed Tyl relative to common soil cations addressed by Call et al. (2019) for reference smectite is investigated in soil systems that contain smectite clay. 102 103 The principal objective of the research was to determine if the preference of Tyl for the exchange 104 phase relative to both Na<sup>+</sup> and Ca<sup>2+</sup> is also present in soil smectite. Binary exchange isotherms 105 were developed to characterize Tyl exchange with Na<sup>+</sup> and Ca<sup>2+</sup>, to establish exchange 106 preference, and to determine the distribution between exchangeable and nonexchangeable forms 107 of the adsorbed Tyl. Adsorbed Tyl was characterized using in situ ATR-FTIR spectroscopy to 108 evaluate retention mechanisms. The IR spectra of Tyl adsorbed on hematite was also examined; 109 the hematite was used as a surrogate for soil iron oxides.

110 **2.** Materials and methods

111 2.1 Chemicals

112 Veterinary-grade Tyl tartrate was obtained from Elanco (Greenfield, IN) for use in the
113 exchange studies and consists of 95.5% Tyl, 3.1% desmycosin, and 1.4% macrocin. Analytical

114 grade Tyl tartrate (98.8% Tyl) was obtained from Sigma-Aldrich and used as a standard for the 115 chemical analysis of the equilibrium exchange systems. Other chemicals used include distilled-116 deionized (DDI) water (carbon dioxide free, >18  $\Omega$ ; Barnstead E-pure system) and analytical or 117 HPLC grade compounds (Fisher Scientific, Fair Lawn, NJ). Hematite was prepared using the 118 method described by Rakshit et al. (2017) and Sallman et al. (2018). Briefly, a 500-mL volume 119 of 2 M FeCl<sub>3</sub> was slowly added to a 500-mL volume of 5.4 M NaOH while continuously stirred. 120 The resulting precipitate was aged in a sealed Pyrex glass bottle for 8 days at 101°C. The 121 precipitate was centrifuge-washed with DDI until an electrical conductivity of the supernatant 122 liquid was  $<5 \ \mu S \ cm^{-1}$ . X-ray diffraction was used to confirm the precipitate was hematite 123 without detectable impurities. The suspension was freeze-dried, and the hematite stored dry at 124 room temperature.

125 2.2 Soils

126 Soil samples were collected from the University of Tennessee Milan Research and 127 Education Center in Milan, TN (35° 56' N, 88° 43' W, 125 m elevation). The samples were 128 obtained from the 15-30 cm and 30-46 cm depth increments of a Loring silt loam soil (Oxyaquic 129 Fragiudalfs) which represent the Bt1 and Bt2 horizons and contain approximately 23% to 27% clay (Lindbo et al., 1994) and < 1 g kg<sup>-1</sup> organic carbon (Jagadamma et al., 2019). The clay 130 131 mineralogy of the soil samples was determined through x-ray diffraction (described below). The 132 soil samples were air dried, disaggregated, passed through a 2 mm sieve, and then Na-saturated 133 by repeated centrifuge washings with 1 M NaCl. The  $<2 \mu m$  size separate was isolated using 134 Stoke's Law sedimentation with Na-saturated soil. The clay fraction was repeatedly centrifuge washed with DDI water to remove entrained salt, then freeze-dried. Clay subsamples were also 135 136 Ca- or Tyl-saturated with repeated centrifuge washes of Na-saturated samples with either 1.0 M

CaCl<sub>2</sub>, or 0.004 M or 0.01 M Tyl tartrate. Following the removal of entrained salts with DDI
water, the samples were freeze dried.

139 The cation exchange capacity (CEC) of the Bt1 and Bt2 clay separates was determined on 140 replicate samples using pH 7.2 Ca-acetate saturation, followed by NH<sub>4</sub>-acetate displacement 141 (modified from Jackson, 2005). The clay samples were centrifuge washed 3 times with pH 7.2 142 1.0 M Ca-acetate to remove native exchangeable cations and to saturate the exchange complex 143 with Ca. Entrained Ca-acetate was then removed by centrifuge washing 3 times with DDI. 144 Finally, the Ca-saturated samples were centrifuge washed 3 times with pH 7.2 1 M NH<sub>4</sub>-acetate, 145 the supernatant liquid collected into a 100 mL volumetric flask, brought to volume with NH<sub>4</sub>-146 acetate, and filtered through qualitative-grade filter paper. The NH<sub>4</sub>-acetate extracts were stored 147 under refrigeration until analyzed for Ca. The Ca concentration in the extract is a direct measure 148 of the CEC. The mean CEC ( $\pm$  sd) of the Bt1 clay separate was 43.9  $\pm$  1.9 cmol<sub>c</sub> kg<sup>-1</sup>, and that of 149 the Bt2 separate was  $36.1 \pm 1.9 \text{ cmol}_{c} \text{ kg}^{-1}$ .

#### 150 2.3 Binary exchange isotherms

151 Binary exchange isotherms were developed using 0.5 g of the Bt1 and Bt2 clay separates 152 in 50 mL polyethylene centrifuge tubes. The centrifuge tubes were wrapped in aluminum foil to 153 minimize Tyl photolysis. The exchange experiments were performed in triplicate and initiated 154 with clay samples that were initially Tyl- or Na-saturated for Tyl-Na exchange, or Tyl- or Casaturated for Tyl-Ca exchange. The soil clays were introduced to a 30 mL volume of solution 155 156 containing varied ratios of Tyl to Na, or Tyl to Ca, such that the total normality was 0.004 N. 157 Centrifuge tubes without solids were used as controls for each exchange experiment. These 158 control blanks were duplicated for each cation ratio in the experiment and used to determine Tyl 159 mass balance, total Cl and tartrate concentrations for ion speciation modeling, and to compute

160 total adsorbed concentration of Tyl. The suspensions and the blanks were equilibrated for 18 161 hours on an orbital shaker at ambient temperature (20-22°C). Preliminary kinetics studies 162 indicated that exchange equilibrium in a reference smectite system was achieved in less than 2 163 hours; the 18 hour reaction time was chosen for convenience. Following the exchange 164 equilibration period, the solution and solid phases were separated by centrifugation and filtration 165 through a 0.45 µm membrane filter and the solution pH was determined. For all experiments, the 166 pH varied between 7.13 and 7.57. The equilibrium solutions were stored under refrigeration until 167 analyzed. The remaining solids were centrifuge washed 3 times with DDI water to remove 168 entrained soluble salts, then washed 3 times with 1.0 M NH<sub>4</sub>-acetate (and the supernatant liquid 169 collected) to remove the exchangeable cations. The collected supernatant NH<sub>4</sub>-acetate solution 170 was brought to a volume of 100 mL and filtered through qualitative-grade filter paper. The NH<sub>4</sub>-171 acetate extracts were stored in the dark and under refrigeration until analyzed.

### 172 2.4 Clay mineralogy and x-ray diffraction

173 To determine clay mineralogy of the Loring clay separates (15-30 cm and 30-46 cm 174 depths), samples were sequentially treated to remove cementing agents with pH 5.0 ammonium 175 acetate (carbonate removal and Na-saturation), hydrogen peroxide (organic matter and 176 manganese oxide removal), and sodium citrate-bicarbonate-dithionite (iron oxide removal) as 177 described by Jackson (2005). The disaggregated clays were then centrifuge washed 4 times with 178 either 1 M KCl or 1 M MgCl<sub>2</sub> and then centrifuge washed 3 times with DDI water to remove 179 entrained soluble salt. Suspensions of K- and Mg-saturated samples were pipetted onto quartz 180 slides and dried overnight at ambient temperature. The slides were then placed in a desiccator 181 over a saturated MgCl<sub>2</sub> solution at a relative humidity of 33%. In addition to the ambient 182 temperature K-saturated clay slides, slides were selected for heat treatment in a muffle furnace at 183 330°C and 550°C for 2 hours; these slides were cooled and stored in a desiccator over anhydrite. 184 For the Mg-glycol saturation, an Mg-saturated clay suspension was combined with glycol and 185 the mixture placed on a quartz slide. The Mg-glycol slides were placed in a desiccator over a 186 beaker of glycol and allowed to dry. X-ray diffraction was then preformed to determine d-values 187 of the (001) spacings of the layer silicates for the five treatments: K-saturated ambient, K-188 saturated 330°C, K-saturated 550°C, Mg-saturated, and Mg-glycol saturated. X-ray 189 diffractograms were generated using a Bruker Model D8 with Ni-filtered, Cu Ka radiation. The 190 x-ray diffractometer operating parameters were set to 40 kV and 40 mA with a scan range of 2 to 191  $30^{\circ} 2\theta$ , a step of 0.02 °2 $\theta$ , and a count rate of 2 seconds per step.

192 2.5 Analytical

193 A Hewlett-Packard Series 1100 (Hewlett-Packard Palo Alto, CA) HPLC coupled with 194 UV detection was used to determine Tyl concentrations using a procedure described by 195 Essington et al. (2010) and Lee et al. (2014). An Ascentis C18 guard column (2 cm by 4.0 mm 196 and 5 µm) and an Ascentis C18 analytical column (15 cm by 4.6 mm and 5 µm) with an injection 197 volume of 100  $\mu$ L and a flow rate of 1 mL min<sup>-1</sup> was used. The mobile phase was an 198 acetonitrile-0.01 M KH<sub>2</sub>PO<sub>4</sub>, pH 7.0 gradient ranging from 20:80 to 60:40 in 10 minutes, 199 resulting in a Tyl retention time of 7.8 minutes. A UV detector wavelength of 280 nm results a method detecting limit of 0.008 µmol L<sup>-1</sup>. The Na and Ca concentrations were determined using 200 201 a Perkin Elmer AAnalyst 800 atomic absorption spectrophotometer (Wellseley, PA). Sodium 202 was analyzed with emission while Ca with absorbance. The samples and standards were spiked 203 with 12% LaCl<sub>3</sub> solution using 0.1 mL for every 10 mL of sample. Sodium and Ca standards 204 were made using atomic absorption standards from CPI International (Springfield, VA). Method 205 detection limits for both Na and Ca were  $0.01 \text{ mg L}^{-1}$ .

9

206 A chemical equilibrium modeling program, Visual MINTEQ (VM) (version 3.1) with the 207 built-in NIST 46.7 thermodynamic database (Gustafsson, 2018) was used to compute the free cation concentrations of Na<sup>+</sup>, Ca<sup>2+</sup>, and Tyl<sup>+</sup>, as well as their single-ion activities. The database 208 209 was augmented to include the  $pK_a$  value of 7.50 (Qiang and Adams, 2004) for the dissociation of 210 the Tyl dimethylamine functional group. Input data for the speciation model included the 211 equilibrium pH, and the total soluble concentrations of Na or Ca, Tyl, Cl, and tartrate. The 212 concentration of Cl was computed from the Na or Ca content of the blanks, and the concentration 213 of tartrate was computed from the Tyl content of the blanks.

214 2.6 Infrared spectroscopy

215 A flow-through, in-situ ATR-FTIR system was used to examine the adsorption of Tyl by 216 the Na- and Ca-saturated soil clays, and by hematite. This system, described in detail by Dolui et 217 al. (2018), consists of a Perkin Elmer Frontier Infrared Spectrometer, a liquid N<sub>2</sub> cooled MCT-A 218 (mercury cadmium telluride) detector, and an optics compartment purged with CO<sub>2</sub>- and H<sub>2</sub>O-219 free air. A reaction cell contains a horizontal 45° ZnSe ATR crystal (HATR) (Pike Technologies, 220 Madison, WI) attached to the ATR sample stage of the IR spectrometer. A peristaltic pump (Watson Marlow 400, Falmouth, UK) circulates an Ar-purged solution from a 500-mL solute 221 222 reservoir through the reaction cell at a rate of 2 mL min<sup>-1</sup>. The reservoir contains a pH-adjusted, 223 Tyl solution in a 0.01 M NaCl or CaCl<sub>2</sub> background electrolyte that is constantly stirred with a 224 magnetic stirring bar. A pH electrode monitors pH in the reservoir during Tyl adsorption. 225 A Na- or Ca-saturated clay or hematite sample (~ 2.5 mg) in 0.01M NaCl or CaCl<sub>2</sub> is 226 layered onto the HATR crystal before each ATR-FTIR experiment. The HATR crystal is 227 attached to the ATR sample stage after the mineral film is air-dried overnight. Background FTIR

spectra are collected while circulating a pH 7, 0.01 M NaCl or CaCl<sub>2</sub> solution over the mineral

film. All subsequent spectra are collected relative to the background. Tylosin is then added to the solute reservoir to yield a circulating solution that is 10  $\mu$ M Tyl. Spectra are accumulated until the intensity of the IR bands does not change with time. The concentration of Tyl in the reservoir is increased to 100  $\mu$ M and the accumulation of FTIR spectra is continued until the intensity of the IR bands does not change, which occurs at approximately 4.5 h reaction time. Band assignments were performed according to Amarasinghe et al. (2009) and Yang et al. (2016) (Table S1).

236 2.7 Exchange data analysis

The concentrations of cations in the exchange phase are directly determined by NH<sub>4</sub> extraction. An exchange isotherm for Tyl is a plot of the equivalent fraction of Tyl<sup>+</sup> on the exchange phase ( $E_{TylX}$ , y-axis) versus the equivalent fraction of Tyl<sup>+</sup> in the equilibrium solution ( $E_{Tyl^+}$ , x-axis). A detailed description of the development of exchange isotherms is provided by Essington (2015). For NaX-TylX exchange,  $E_{TylX}$  is

$$E_{\text{TylX}} = \frac{\{\text{TylX}\}}{\{\text{TylX}\}+\{\text{NaX}\}}$$
(1)

where {TylX} and {NaX} are the mol kg<sup>-1</sup> of Tyl<sup>+</sup> or Na<sup>+</sup> on the exchange phase, and X<sup>-</sup> represents a mole of exchange phase charge. Further, the equivalent fraction of Na<sup>+</sup> on the exchange phase is

246 
$$E_{\text{NaX}} = 1 - E_{\text{TylX}} = \frac{\{\text{NaX}\}}{\{\text{TylX}\}+\{\text{NaX}\}}$$
 (2)

247 The equivalent fraction of Tyl<sup>+</sup> in solution is

248 
$$E_{\text{Tyl}^+} = \frac{[\text{Tyl}^+]}{[\text{Tyl}^+] + [\text{Na}^+]}$$
 (3)

249 and

250 
$$E_{\text{Na}^+} = 1 - E_{\text{Tyl}^+} = \frac{[\text{Na}^+]}{[\text{Tyl}^+] + [\text{Na}^+]}$$
 (4)

251 where  $[Tyl^+]$  and  $[Na^+]$  are the mol  $L^{-1}$  of free cation in solution (computed using VM). A

similar set of expressions describe the CaX<sub>2</sub>-TylX exchange systems:

253 
$$E_{\text{TylX}} = \frac{\{\text{TylX}\}}{\{\text{TylX}\} + 2\{\text{CaX}_2\}}$$
 (5)

254 
$$E_{\text{CaX}_2} = 1 - E_{\text{TylX}} = \frac{2\{\text{CaX}_2\}}{\{\text{TylX}\} + 2\{\text{CaX}_2\}}$$
(6)

255 
$$E_{\text{Tyl}^+} = \frac{[\text{Tyl}^+]}{[\text{Tyl}^+] + 2[\text{Ca}^{2+}]}$$
 (7)

256 
$$E_{Ca^{2+}} = 1 - E_{Tyl^+} = \frac{2[Ca^{2+}]}{[Tyl^+] + 2[Ca^{2+}]}$$
(8)

Also plotted on the exchange isotherm is the non-preference isotherm, obtained from Essington (2015). For NaX-TylX exchange, the non-preference isotherm is  $E_{TylX} = E_{Tyl^+}$ . For CaX<sub>2</sub>-TylX exchange, the non-preference isotherm is

260 
$$E_{\text{TylX}} = \left\{ 1 + \frac{2}{\Gamma N_{\text{T}}} \left[ \frac{1}{E_{\text{Tyl}}^2} - \frac{1}{E_{\text{Tyl}}} \right] \right\}^{-0.5}$$
(9)

261 where  $N_{\rm T}$  is the total normality,  $\Gamma = \frac{\gamma_{\rm Tyl}^2}{\gamma_{\rm Ca}^2 +}$ , and the  $\gamma$ 's are single-ion activity coefficients. The 262 non-preference isotherm describes the condition where neither cation is preferred by the 263 exchange phase. The non-preference condition is met when the Vanselow selectivity coefficient 264 ( $K_{\rm V}$ ) for the exchange reaction is unity. For the NaX + Tyl<sup>+</sup> = TylX + Na<sup>+</sup> exchange reaction,  $K_{\rm V}$ 265 is

266 
$$K_{\rm V} = \frac{N_{\rm TylX}(\rm Na^{+})}{N_{\rm NaX}(\rm Tyl^{+})}$$
(10)

267 For the CaX<sub>2</sub> + 2Tyl<sup>+</sup> = 2TylX + Ca<sup>2+</sup> exchange reaction,  $K_V$  is

268 
$$K_{\rm V} = \frac{N_{\rm TylX}^2({\rm Ca}^{2+})}{N_{\rm CaX_2}({\rm Tyl}^+)^2}$$
(11)

269 In Eqs. (10) and (11) the parentheses represent activity (obtained from VM) and  $N_{TylX}$ ,  $N_{NaX}$ ,

270 and  $N_{CaX_2}$ , are the mole fractions:

271 
$$N_{\text{TylX}} = \frac{\{\text{TylX}\}}{\{\text{TylX}\} + \{\text{NaX}\}} ; N_{\text{NaX}} = 1 - N_{\text{TylX}} = \frac{\{\text{NaX}\}}{\{\text{TylX}\} + \{\text{NaX}\}}$$
(12)

272 
$$N_{\text{TylX}} = \frac{\{\text{TylX}\}}{\{\text{TylX}\} + \{\text{CaX}_2\}}; N_{\text{CaX}_2} = 1 - N_{\text{TylX}} = \frac{\{\text{CaX}_2\}}{\{\text{TylX}\} + \{\text{CaX}_2\}}$$
(13)

A plot of ln  $K_V$  [Eqs. (10) or (11)] as a function of  $E_{TylX}$  [Eqs. (1) or (5)] provides a mechanism to determine the true exchange equilibrium constant ( $K_{ex}$ ) (Essington, 2015).

275 A mass balance was performed to determine the nonexchangeable concentration of 276 adsorbed Tyl. For initially Na- or Ca-saturated samples, the mass of Tyl added to the suspensions is determined from the blanks. The mass balance expression for adsorbed Tyl is:  $^{Tyl}n_T = ^{Tyl}n_{ex} +$ 277  $^{Tyl}n_{nonex}$ , where  $^{Tyl}n_{T}$  represents the total mass of adsorbed Tyl determined by subtracting the 278 279 mass of Tyl in the equilibrium suspensions from that added;  $^{Tyl}n_{ex}$  is directly measured and represents the exchangeable mass of adsorbed Tyl; and  $^{Tyl}n_{nonex}$  is determined by difference and 280 281 represents the nonexchangeable forms of adsorbed Tyl. The adsorbed Tyl concentration of any 282 one of the adsorbed Tyl components is obtained by dividing *n* values by the mass of soil. For example, the total surface excess of adsorbed Tyl (in mmol kg<sup>-1</sup>) is  $q_T = {}^{Tyl}n_T/m_S$ , where  $m_S$  is the 283 mass of the adsorbant in kg. Adsorption isotherms were generated for total adsorbed and 284 285 exchangeable Tyl. An adsorption isotherm is a plot of the amount of Tyl adsorbed by the surface 286 (e.g.,  $q_T$ , y-axis) versus the total concentration of Tyl in the equilibrium solution ( $C_{eq}$  in mmol L<sup>-1</sup>, x-axis). The isotherms were described using the constant partition equation,  $q_T = K_P C_{eq}$ , 287 288 where  $K_{\rm P}$  is the partition constant obtained using least square linear regression analysis.

289 **3.** Results and discussion

290 3.1 Clay mineralogy

291 The adsorption capacity of soil for Tyl, and the distribution of adsorbed Tyl into 292 exchangeable and nonexchangeable forms, is dictated to a large degree by the clay mineralogy 293 (Bewick, 1979; Call et al., 2019). The adsorption of Tyl by non-expansive clay minerals (e.g., 294 the micas and kaolinite) and partially expansive vermiculites is restricted to the external surfaces, 295 while Tyl can intercalate the smectites, generating high adsorption capacities. The cation 296 saturations and heat treatments imposed on the soil size clay fraction of the Loring Bt horizons 297 provide a Group level characterization of the common soil phyllosilicates. The subsoil horizons 298 of the Loring soil display similar clay mineralogy (Fig. 2). The 1.47 nm diffraction in the Mg-299 saturated samples shifts to approximately 1.7 nm upon glycolation; this is diagnostic for the 300 occurrence of smectite. The 1.0 nm diffraction in the Mg- and Mg glycol-saturated samples is 301 indicative of the presence of mica, while the diffraction at > 2.0 nm indicates the occurrence of 302 an interstratified mica-smectite. The disappearance of the 0.72 nm diffraction upon heating to 303 550°C indicates kaolinite. The remaining diffractions are second order mica (0.5 nm) and 304 kaolinite (0.36 nm), and third order mica (0.33) or quartz. Thus, the clay mineral assemblage in 305 the Loring subsoil consists of mica, smectite, mica-smectite interstratification, and kaolinite.

306 *3.2* Binary exchange isotherms

For the NaX-TylX exchange systems, the exchange isotherms illustrate little preference for Tyl<sup>+</sup> or Na<sup>+</sup> by the exchange phase in the subsurface horizons (Fig. 3a). However, Na<sup>+</sup> is preferred relative to Tyl<sup>+</sup> in Bt1 systems when the clays are initially Tyl-saturated, and when  $E_{TylX} > 0.25$ . When initially Na-saturated,  $K_V$  values for NaX $\rightarrow$ TylX exchange [Eq. (10)] do not vary with exchange phase composition, averaging  $1.12 \pm 0.20$  for Bt1 and  $1.41 \pm 0.24$  for Bt2 within the  $E_{TylX}$  range studied (Fig. 4a). On average, there is no preference of Tyl<sup>+</sup> over Na<sup>+</sup> for the exchange phase when the clays are initially Na-saturated, as the  $K_V$  values are not 314 significantly different from unity. When initially Tyl-saturated, Tyl+ is preferred over Na<sup>+</sup> at low 315  $E_{\text{TylX}}$  values;  $K_{\text{V}}$  is 3.95 ± 0.51 when  $E_{\text{TylX}}$  = 0.08 in Bt1, and 3.91 ± 0.72 when  $E_{\text{TylX}}$  is 0.21 in 316 Bt2. However,  $K_V$  values decrease with increasing  $E_{TylX}$ , achieving constant values of 0.79 ± 317 0.29 for Bt1 and 1.18  $\pm$  0.32 when  $E_{TvlX} > 0.25$  (which are also not significantly different from 318 unity). The soil clay minerals are less selective for Tyl<sup>+</sup> over Na<sup>+</sup> than STx-1 reference 319 montmorillonite (Call et al., 2019). They observed a mean  $K_V$  value of 6.61 ± 2.62 when  $E_{TvIX}$ 320 ranged between 0.09 and 0.85. Tylosin is also less competitive with Na<sup>+</sup> than other monovalent 321 cations in smectites and smectitic soils. Vanselow selectivity coefficients evaluated when  $E_{\text{NaX}}$  = 322 0.5 range from 1.74 to 6.23 for NaX $\rightarrow$ KX, 3.54 to 20.6 for NaX $\rightarrow$ RbX, and 7.1 to 121 for 323 NaX→CsX exchange (Deist and Talibudeen, 1967; Gast, 1972; Jensen and Babcock, 1973; Xu 324 and Harsh, 1990; Tournassat et al., 2009).

325 In the CaX<sub>2</sub>-TylX exchange systems, CaX<sub>2</sub> dominates the exchange phase. However, Tyl<sup>+</sup> 326 is preferred because there is greater TylX then predicted by the nonpreference isotherm when 327  $E_{\text{TylX}} \le 0.2$  (Fig. 3b). The  $K_V$  values for the CaX<sub>2</sub> $\rightarrow$ TylX exchange reaction [Eq. (11)] vary as a 328 function of  $E_{\text{TvlX}}$  (Fig. 4b). When initially Ca-saturated,  $K_{\text{V}}$  decreases from 17.6 to 3.44 as  $E_{\text{TvlX}}$ 329 increases from 0.01 to 0.08 in Bt1, and from 58.1 to 6.28 as  $E_{TylX}$  increases from 0.02 to 0.14 in 330 Bt2. This trend is similar for initially Tyl-saturated clay, as K<sub>V</sub> decreases from 6.14 to 1.34 as 331  $E_{\text{TvIX}}$  increases from 0.09 to 0.40 in Bt1, and from 21.6 to 1.67 as  $E_{\text{TvIX}}$  increases from 0.14 to 0.36 in Bt2. In general,  $K_V$  values for CaX<sub>2</sub> $\rightarrow$ TylX exchange are greater in Bt2 relative to Bt1. 332 333 The preference of the soil clay for Tyl<sup>+</sup> over Ca<sup>2+</sup>, and the decreasing preference with increasing 334  $E_{\text{TylX}}$  is similar to that observed for CaX<sub>2</sub> $\rightarrow$ TylX exchange on STx-1 montmorillonite (Call et al., 335 2019). They observed  $K_V$  values to decrease from 388 to ~1 when  $E_{TvlX}$  increased from 0.02 to 336 0.40. For smectites and smectitic soils,  $K_V$  values for Na<sup>+</sup> and K<sup>+</sup> exchange with Ca<sup>2+</sup> also

decrease with increasing  $E_{\text{NaX}}$  and  $E_{\text{KX}}$ . When  $E_{\text{NaX}} = 0.5$ ,  $K_{\text{V}}$  values for CaX<sub>2</sub> $\rightarrow$ NaX exchange range from 0.15 to ~1 (Rytwo et al., 1996; Sposito et al., 1983; Tournasset et al., 2009). When  $E_{\text{KX}} = 0.5$ ,  $K_{\text{V}}$  values for CaX<sub>2</sub> $\rightarrow$ KX exchange range from 2.24 to 22.2 (Deist and Talibudeen, 1967; Jensen and Babcock, 1973; Evangelou, 1998; Essington, 2015; Ritter et al., 2017). These comparison indicate that Tyl<sup>+</sup> is less competitive than K<sup>+</sup> in exchange with CaX<sub>2</sub>, but that Tyl<sup>+</sup> and Na<sup>+</sup> have similar selectivity.

343 *3.3 Cation exchange capacity* 

344 The CEC values of the Bt1 and Bt2 clays measured by Ca-NH<sub>4</sub> acetate exchange were 345  $43.9 \pm 1.87$  and  $36.1 \pm 1.91$  cmol<sub>c</sub> kg<sup>-1</sup>. However, the *CEC* in the binary exchange systems, 346 determined by the summation of exchangeable cation charge, also varied as a function of the 347 initial saturating cation. For the Bt1 NaX-TylX systems that were initially Na-saturated, the CEC was  $10.3 \pm 1.60 \text{ cmol}_{c} \text{ kg}^{-1}$ . When initially Tyl-saturated, the CEC was  $3.45 \pm 0.33 \text{ cmol}_{c} \text{ kg}^{-1}$ . 348 349 The CEC was also depressed in the Bt2 NaX-TylX systems:  $15.9 \pm 1.80$  cmol<sub>c</sub> kg<sup>-1</sup> when initially Na-saturated, and  $4.81 \pm 0.43$  cmol<sub>c</sub> kg<sup>-1</sup> when Tyl-saturated. The CEC values in the 350 351 initially Ca-saturated CaX<sub>2</sub>-TylX exchange systems were similar to those determined by Ca-NH<sub>4</sub> acetate exchange;  $46.6 \pm 3.34$  cmol<sub>c</sub> kg<sup>-1</sup> for Bt1 and  $37.3 \pm 1.32$  cmol<sub>c</sub> kg<sup>-1</sup> for Bt2. However, 352 when initially Tyl-saturated, the CEC values were  $4.50 \pm 1.47$  cmol<sub>c</sub> kg<sup>-1</sup> for Bt1 and  $6.99 \pm 1.44$ 353 cmol<sub>c</sub> kg<sup>-1</sup> for Bt2. 354

The reduction in *CEC* associated with Tyl retention was also observed in montmorillonite systems (Call et al., 2019). These authors postulated that due to its size, intercalated Tyl was blocking the access of cations to the exchange locations. Ivanov (1998, 2002), using molecular

- 358 mechanics modeling techniques, determined that Tyl in solutions may exist in two
- 359 configurations: an unfolded linear structure was predicted to be the most stable, while a less

360 stable folded structure was also predicted. Call et al. (2019) reported the approximate solvent-361 accessible dimensions of the linear structure were 2.71 nm  $\times$  0.95 nm  $\times$  1.47 nm, while those of 362 the folded structure were 2.71 nm  $\times$  1.47 nm  $\times$  1.35 nm. For an example montmorillonite with a 363 unit cell layer charge of 1.12 and unit cell a and b dimensions of 0.52 nm and 0.9 nm, the surface 364 area per unit charge is  $0.84 \text{ nm}^2/z$ . If viewed as a box-like structure, the area of the Tyl molecule sides in the linear configuration are 4.0 nm<sup>2</sup>, 2.6 nm<sup>2</sup>, and 1.4 nm<sup>2</sup>; those for the folded 365 configuration are 3.1 nm<sup>2</sup>, 2.9 nm<sup>2</sup>, and 2.0 nm<sup>2</sup>. All of these exceed the area associated with unit 366 367 charge in a smectite interlayer, indicating that blocking the access of cations to exchange locations by Tyl reduces the apparent CEC of the adsorbent. 368

369 *3.4 Adsorption isotherms* 

370 The adsorption isotherms indicate that the total adsorption of Tyl is influenced by the horizon and the competing cation (Fig. 5). There is significantly greater adsorption in Bt2 371 372 relative to Bt1 in both the NaX-TylX and CaX<sub>2</sub>-TylX systems, as quantified by the linear 373 partition coefficients (Table 1). There is also greater Tyl retention in the NaX-TylX systems 374 relative to the CaX<sub>2</sub>-TylX systems, a finding that is consistent with that of prior research (Bewick, 1979; Essington et al., 2010; Lee et al., 2014). The fraction of the total adsorbed Tyl 375 376 that is exchangeable also varies with horizon and competing cation. Approximately 36% of 377 adsorbed Tyl is exchangeable in the Bt1, NaX-TylX systems, compared to 43% in Bt2. In the 378 CaX<sub>2</sub>-TylX systems, 57% of the total adsorbed Tyl is exchangeable in Bt1, 66% in Bt2. 379 There is also very little difference between exchangeable Tyl in the Na- and Ca-380 dominated systems in Bt1 (Fig. 5, Table 1), although there is significantly greater exchangeable Tyl in the Na-systems in Bt2. The CEC of the CaX<sub>2</sub>-TylX systems (46.6 cmol<sub>c</sub> kg<sup>-1</sup> for Bt1 and 381 382  $37.3 \text{ cmol}_{c} \text{ kg}^{-1}$  for Bt2) is significantly greater than that of the NaX-TylX systems (10.2 cmol\_{c})

383 kg<sup>-1</sup> for Bt1 and 15.9 cmol<sub>c</sub> kg<sup>-1</sup> for Bt2). However,  $E_{TylX}$  only ranges up to approximately 0.1

384 for CaX<sub>2</sub>-TylX exchange, but ranges up to 0.4 in NaX-TylX. Thus, the surface excess of

385 exchangeable Tyl are similar between the Ca- and Na systems. The nonexchangeable component

of adsorbed Tyl is greater in the Na-systems, with  $K_P$  values of 21.8 and 33.4 for Bt1 and Bt2,

387 relative to the Ca-systems (*K*<sub>P</sub> values of approximately 8 for both horizons) (Table 1).

388 3.5 Infrared spectroscopy

389 The sharp increase in IR band intensity at 1588 cm<sup>-1</sup> (Fig. 6a and b) indicates the possible 390 bonding of Tyl on mineral surfaces through the dimethylamine  $[-N(CH_3)_2]$  functional group. Bands at 1580-1590 cm<sup>-1</sup> are due to the N-H bending vibrations. Coupled with this is the strong 391 IR band at 1385 cm<sup>-1</sup> and the weaker bands at 1458 cm<sup>-1</sup> and 1320 cm<sup>-1</sup>, which represent the C-392 393 H bending vibrations of the dimethylamine (Amarasinghe et al., 2009). Conversely, the 1390-394 1310 cm<sup>-1</sup> region may also represent the O-H vibrations from phenols. The shoulder bands at 1712 and 1661 cm<sup>-1</sup> are most likely due to the interactions of the aromatic aldehyde (-CHO) and 395 396 carbonyl (-C=O) groups with the mineral surfaces. Interestingly, in the CaCl<sub>2</sub> medium, the IR band a near 1580-1590 cm<sup>-1</sup> appeared as a shoulder to the more pronounced 1627 cm<sup>-1</sup> band 397 398 (Fig. 6b). A broad IR band near 1627 cm<sup>-1</sup> and a shoulder at 1659 cm<sup>-1</sup> are features that differ 399 from the IR spectra collected in the NaCl medium. The salt effect on these IR bands may indicate 400 an outer-sphere adsorption mechanism through the electronegative carbonyl moieties and 401 positive surface sites.

402 The IR bands at 1077, 1029-1031, 1001-1011 cm<sup>-1</sup> (Figs. 6a and b) have been identified as 403  $\gamma$ (Si-O) vibration bands that represent the interlayer adsorption of highly polar molecules by 404 phyllosilicates (Amarasinghe et el., 2009; Yang et al., 2016). Call et al., (2019) observed these 405 bands to predominate the IR spectra of Tyl adsorbed by montmorillonite from both NaCl and 406 CaCl<sub>2</sub> media. Correspondingly, they also observed the direct intercalation of Tyl into the 407 montmorillonite interlayers. Other IR bands found in the montmorillonite systems (1590, 1462, 1383, 1317, and 1163 cm<sup>-1</sup>), which represent H-bonding, surface interactions through the 408 409 dimethylamine, and internal ring vibrations of the Tyl molecule were minor, indicating only the 410 minor covalent and external surface bonding of Tyl. By comparison, the intense IR bands in the 411 1590-1720 cm<sup>-1</sup> range indirectly support the hypothesis that the Tyl functional groups primarily 412 interact through covalent bonding with the external surfaces of the soil minerals. Further, only 413 approximately 36% to 43% of the adsorbed Tyl is exchangeable in NaCl systems, and 57% to 414 66% in CaCl<sub>2</sub>, supporting the significance of Tyl adsorption on the external clay surfaces, as well 415 as other soil minerals. To test this hypothesis, the IR spectra of Tyl adsorbed by hematite was 416 examined (Fig. 6c). The iron oxide content of the Loring soil 15-30 cm and 30-46 cm depth 417 increments is approximately 2 to 3% (unpublished data); thus, hematite is a surrogate for the soil iron oxides. The strong IR bands near 1586 cm<sup>-1</sup> and 1387 cm<sup>-1</sup>, with weaker bands in the 1334-418 419 1310 cm<sup>-1</sup> range, are similar to the band locations for the Tyl-soil interactions, suggesting similar 420 bonding mechanisms. Further, IR bands near 1139, 1084, and 1000 cm<sup>-1</sup>, which were identified 421 as  $\gamma$ (Si-O) vibration bands, also appear in the hematite spectra. It is possible some overlapping IR bands from H-bonding, ring-OH groups, and internal ring vibrations of the Tyl molecule also 422 423 appear in this region. Overall, the IR bands from Tyl-soil and Tyl-hematite interactions, coupled 424 with the significant distribution of adsorbed Tyl into non-exchangeable forms (Fig. 5), suggest 425 that adsorption on external phyllosilicate and iron oxide surfaces are important retention 426 mechanisms, possibly through carbonyl and dimethylamine moieties, with minor phyllosilicate 427 interlayer adsorption.

428 **4.** Conclusions

429	The adsorbed Tyl in pH neutral, smectitic subsoil horizons is divided into exchangeable
430	and nonexchangeable forms. The percentage of adsorbed Tyl that is exchangeable varies from
431	36% to 43% when Na <sup>+</sup> is the competing cation, and from 57% to 66% when $Ca^{2+}$ competes.
432	Thus, cation exchange is an important mechanism for the retention of Tyl by the soil clay
433	fraction. In NaX-TylX binary exchange systems, neither Na <sup>+</sup> nor Tyl <sup>+</sup> is preferred by the clay
434	exchange phase, and the Vanselow selectivity coefficients ( $K_V$ ) for the NaX $\rightarrow$ TylX exchange
435	reaction range between 0.79 and 1.41. In addition, $K_V$ is generally invariant with exchange phase
436	composition. In the CaX <sub>2</sub> -TylX systems, Tyl <sup>+</sup> is preferred by the clay exchange phase when the
437	equivalent fraction of TylX ( $E_{TylX}$ ) is less than 0.4. The $K_V$ values for the CaX <sub>2</sub> $\rightarrow$ TylX exchange
438	reaction are at a maximum at the lowest $E_{TylX}$ values, with 17.6 < $K_V$ < 58.1, then decrease with
439	increasing $E_{TylX}$ to 1.34 < $K_V$ < 6.28. Adsorbed Tyl masks the <i>CEC</i> of the soil clays; the effect
440	was greatest in systems that were initially Tyl-saturated. However, this effect was also apparent
441	to the lowest levels of TylX ( $E_{TylX} \le 0.01$ ), and was attributed to the steric effects of the large Tyl
442	molecule. The in-situ ATR-FTIR spectra indicate that surface interactions of Tyl with soil iron
443	oxides, through the dimethylamine functional group, may be an important retention mechanism.
444	The findings of this study directly illustrate the importance of ion exchange as a mechanism for
445	the retention of Tyl in soil, and provide a molecular-level characterization of the surface
446	complexation of Tyl in soil.

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- 450 **References**

- Allaire, S.E., Del Castillo, J., Juneau, V. 2006. Sorption kinetics of chlortetracycline and tylosin
  on sandy loam and heavy clay soils. J. Environ. Qual. 35, 969-972.
- 453 Amarasinghe, P.M., Katti, K.S., Katti, D.R. 2009. Nature of organic fluid-montmorillonite
- 454 interactions: an FTIR spectroscopic study. J. Colloid Interface Sci. 337, 97-105.
- Bewick, M.W.M. 1979. The adsorption and release of tylosin by clays and soils. Plant Soil 51,
  363-372.
- 457 Call, J.J., Rakshit, S., Essington, M.E. 2019. The adsorption of tylosin by montmorillonite and
- 458 vermiculite: exchange selectivity and intercalation. Soil Sci. Soc. Am. J.
- 459 doi:10.2136/sssaj2018.12.0475.
- Chander, Y., Kumar, K., Goyal, S.M., Gupta, S.C. 2005. Antibacterial activity of soil-bound
  antibiotics. J. Environ. Qual. 34, 1952-1957.
- 462 Deist, J., Talibudeen, O. 1967. Ion exchange in soils from the pairs K-Ca, K-Rb, and K-Na. J.
  463 Soil Sci. 18, 125-137.
- 464 Dolui, M., Rakshit, S., Essington, M.E., Lefèvre, G. 2018. Probing oxytetracycline sorption
- 465 mechanism on kaolinite in a single ion and binary mixtures with phosphate using in situ
- 466 ATR-FTIR spectroscopy. Soil Sci. Soc. Am. J. 82, 826-838.
- 467 Dungan, R.S., Snow, D.D., Bjorneberg, D.L. 2017. Occurrence of antibiotics in an agricultural
  468 watershed in south-central Idaho. J. Environ. Qual. 46, 1455-1461.
- Essington, M.E. 2015. Soil and Water Chemistry: An Integrative Approach, second ed. CRC
  press, Boca Raton, FL.
- 471 Essington, M. E., J. Lee, J., Seo, Y. 2010. Adsorption of antibiotics by montmorillonite and
- 472 kaolinite. Soil Sci. Soc. Am. J. 74, 1577-1588.

- 473 Evangelou, V.P. 1998. Environmental Soil and Water Chemistry: Principles and Applications, J.
  474 Wiley & Sons, New York.
- 475 Fish, B.J., Carr, G.P.R. 1986. Pharmacopoeial procedure for the determination of tylosin factors
- 476 by high-performance liquid chromatography. J. Chromatography 353, 39-50.
- 477 Gast, R.G. 1972. Alkali metal cation exchange on Chambers montmorillonite. Soil Sci. Soc. Am.
  478 J. 36, 14-19.
- 479 Gustafsson, J.P. 2018. Visual MINTEQ. Version 3.1. Stockholm: KTH Royal Institute of
  480 Technology. http://vminteq.lwr.kth.se/.
- 481 Hamill, R.L., Haney, M.E., Stamper, M., Wiley, P.F. 1961. Tylosin, a new antibiotic. II.
- 482 Isolation, properties, and preparation of desmycosin, a microbiological active degradation
  483 product. Antibiot. Chemother. 11, 328-334.
- 484 Ivanov, P.M. 1998. Molecular mechanics conformational analysis of tylosin. J. Molecular
  485 Structure 440, 121-130.
- 486 Ivanov, P.M. 2002. CONFLEX/MM3 search/minimization study of the conformations of the
- 487 macrolide tylosin. J. Molecular Structure 606, 217-229.
- Jackson, M.L. 2005. Soil Chemical Analysis-Advanced Course, second ed. University of
  Wisconsin-Madison Libraries, Madison, WI.
- 490 Jagadamma, S., Essington, M.E., Xu, S., Yin, X. 2019. Total and active soil organic carbon from
- 491 long-term agricultural management practices in West Tennessee. Agric. Environ. Letters.
- 492 doi: 10.2134/ael2018.11.0062.
- 493 Jensen, H.E., Babcock, K.L. 1973. Cation-exchange equilibria on a Yolo loam. Hilgardia 41,
- 494 475-488.

- 495 Joy, S.R., Bartelt-Hunt, S.L., Snow, D.D., Gilley, J.E., Woodbury, B.L., Parker, D.B., Marx,
- 496 D.B., Li, X. Fate and transport of antimicrobials and antimicrobial resistance genes in soil
  497 and runoff following land application of swine manure slurry. Environ. Sci. Technol. 47,
- 498 12081-12088.
- Kumar, K., Gupta, S.C., Chander, Y., Singh, A.K. 2005. Antibiotic use in agriculture and its
  impact on the terrestrial environment. Adv. Agron. 87.
- Lee, J., Seo, Y., Essington, M.E. 2014. Sorption and transport of veterinary pharmaceuticals in
  soil a laboratory study. Soil Sci. Soc. Am. J. 78, 1531-1543.
- 503 Lindbo, D.L., Rhoton, F.E., Bigham, J.M., Jones, F.S., Smeck, N.E., Hudnall, W.H., Tyler, D.D.
- 504 1994. Bulk density and fragipan identification in loess soils of the lower Mississippi River
  505 valley. Soil Sci. Soc. Am. J. 58, 884-891.
- 506 Loke, M.L., Tjørnelund, J., Halling-Sørensen, B. 2002. Determination of the distribution
- 507 coefficient (log *K*<sub>d</sub>) of oxytetracycline, tylosin A, olaquindox and metronidazole in manure.
  508 Chemosphere 48, 351-361.
- 509 Luby, E.M., Moorman, T.B., Soupir, M. 2016. Fate and transport of tylosin-resistant bacteria and
- 510 macrolide resistant genes in artificially drained agricultural fields receiving swine manure.
- 511 Sci. Total Environ. 550, 1126-1133.
- 512 McFarland, J.W., Berger, C.M., Froshauer, S.A., Hayashi, S.F., Hecker, S.J., Jaynes, B.H.,
- 513 Jefson, M.R., Kamicker, B.J., Lipinski, C.A., Lundy, K.M., Reese, C.P., Vu, C.B. 1997.
- 514 Quantitative structure-activity relationships among macrolide antibacterial agents: in vitro
- and in vivo potency against *Pasteurella multocida*. J. Med. Chem. 40, 1340-1346.
- 516 Qiang, Z., Adams, C. 2004. Potentiometric determination of acid dissociation constants (pKa) for
- 517 human and veterinary antibiotics. Water Res. 38, 2874-2890.

- 518 Rabølle, M., Spliid, N.H. 2000. Sorption and mobility of metronidazole, olaquindox,
- 519 oxytetracycline and tylosin in soil. Chemosphere 40, 715-722.
- 520 Rakshit, S., Sallman, B., Davantés, A., Lefèvre, G. 2017. Tungstate (VI) sorption on hematite: an
- 521 *in situ* ATR-FTIR probe on the mechanism. Chemosphere 168,685-691.
- 522 Ritter, S., DeSutter, T., O'Brien, P., Casey, F., Wick, A., Horsager, K., Khan, E. 2017. Binary
- 523 exchanges of calcium, magnesium, and potassium on thermally desorbed soil. Soil Sci.
  524 Soc. Am. J. 81, 1088-1095.
- 525 Rytwo, G., Banin, A., Nir, S. 1996. Exchange reactions in the Ca-Mg-Na-montmorillonite
  526 system. Clays Clay Miner. 44, 276-285.
- 527 Sallman, B., Rakshit, S., Lefèvre, G. 2018. Influence of phosphate on tungstate sorption on
  528 hematite: a macroscopic and spectroscopic evaluation of the mechanism. Chemosphere
  529 213,596-601.
- Sassman, S.A., Sarmah, A.K., Lee, L.S. 2007. Sorption of tylosin A, D, and A-aldol and
  degradation of tylosin A in soils. Environ. Toxicol. Chem. 26, 1629-1635.
- 532 Soni, B., Bartelt-Hunt, S.L., Snow, D.D., Gilley, J.E., Woodbury, B.L., Marx, D.B., Li, X. 2015.
- Narrow grass hedges reduce tylosin and associated antimicrobial resistance genes in
  agricultural runoff. J. Environ. Sci. 44, 895-902.
- 535 Sposito, G., Jouany, C., Holtzclaw, K.M., LeVesque C.S. 1983. Calcium-magnesium exchange
- on Wyoming bentonite in the presence of adsorbed sodium. Soil Sci. Soc. Am. J. 47, 10811085.
- 538 Srinivasan, P., Sarmah, A.K., Manley-Harris, M. 2014. Sorption of selected veterinary
- antibiotics onto dairy farming soils of contrasting nature. Sci. Total Environ. 472, 695-703.

540	Stromer, B.S., Woodbury, B., Williams, C.F. 2018. Tylosin sorption to diatomaceous earth
541	described by Langmuir isotherm and Freundlich isotherm models. Chemosphere 193, 912-
542	920.
543	Sura, S., Degenhardt, D., Cessna, A.J., Larney, F.J., Olson, A.F., McAllister, T.A. 2016.
544	Transport of three antimicrobials in runoff from windrows of composted beef cattle
545	manure. J. Environ. Qual. 45, 494-502.
546	ter Laak, T.L., Gebbink, W.A., Tolls, J. 2006. The effect of pH and ionic strength on the sorption
547	of sulfachloropyridazine, tylosin, and oxytetracycline to soil. Environ. Toxicol. Chem. 25,
548	904-911.
549	Tournassat, C., Gailanou, H., Crouzet, C., Braibant, G., Gautier, A., Gaucher, E.C. 2009. Cation
550	exchange selectivity coefficient values on smectite and mixed-layer illite/smectite minerals.
551	Soil Sci. Soc. Am. J. 73, 928-942.
552	Wollenberger, L., B. Halling-Sørensen, K.O. Kusk. 2000. Acute and chronic toxicity of
553	veterinary antibiotics to Daphnia magna. Chemosphere 40:723-730.
554	Xu, S., Harsh, J.B. 1990. Hard and soft acid-base model verified for monovalent cation
555	selectivity. Soil Sci. Soc. Am. J. 54, 1596-1601.
556	Yang, Y., Wang, S., Liu, J., Xu, Y., Zhou, X. 2016. Adsorption of lysine on Na-montmorillonite
557	and competition with Ca <sup>2+</sup> : a combined XRD and ATR-FTIR study. Langmuir 32,4746-
558	4754.
559	Zhang, Q., Yang, C., Huang, W., Dang, Z., Shu, X. 2013. Sorption of tylosin on clay minerals.
560	Chemosphere 93, 2180-2186.
561	

562 **List of Figures** 563 Figure 1. The structure of tylosin. Figure 2. X-ray diffractograms of (a) Bt1 and (b) Bt2 horizons of the Loring soil. A: Mg2+-564 saturated; B: Mg<sup>2+</sup>-glycol-saturated; C: K<sup>+</sup>-saturated ambient temperature; D: K<sup>+</sup>-saturated 565 566 330°C; E: K<sup>+</sup>-saturated 550°C. Diffractions are identified by their *d*-values in nm. 567 Figure 3. Exchange isotherms for the binary (a) NaX-TylX and (b) CaX<sub>2</sub>-TylX systems that 568 illustrate the equivalent fraction of tylosin on the exchange phase  $(E_{TylX})$  as a function of 569 the equivalent fraction of tylosin in solution  $(E_{Tvl})$ , initial cation saturation, and soil 570 horizon. The solid line represents the non-preference (NP) line. 571 Figure 4. The variation of the Vanselow selectivity coefficient [(as  $\ln K_V$ ), Eqs. (10) and (11)] in 572 (a) Bt1 and (b) Bt2 horizons as a function of the equivalent fraction of tylosin on the 573 exchange phase  $(E_{TvlX})$  and initial cation saturation. 574 Figure 5. Total and exchangeable tylosin adsorption isotherms in (a) Bt1 and (b) Bt2 horizons. 575 The lines represent the linear partition model (Table 1);  $q_{Tyl}$  and  $C_{eq}$  are the equilibrium 576 surface excess and the solution concentration of tylosin. 577 Figure 6. Growth of in-situ ATR-FTIR bands upon tylosin adsorption by Bt2 (15-30 cm) horizon 578 from (a) 0.01 M NaCl and (b) 0.01 M CaCl<sub>2</sub> solutions. The spectra in (c) represents tylosin 579 adsorption by hematite from 0.01 M NaCl. [color online]

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## Table 1

Linear partition coefficients ( $K_P$  in L kg<sup>-1</sup>) that describe the tylosin adsorption isotherms in Fig. 5 as a function of competing cation and soil horizon.

	Total			Exchangeable			Nonexchangeable			
Horizon	K <sub>P</sub>	SE <sup>a</sup>	CI <sup>b</sup>	K <sub>P</sub>	SE	CI	K <sub>P</sub>	SE	CI	
	NaX-TylX									
Bt1	34.0	0.91	1.9	12.2	0.32	0.7	21.8	0.76	1.6	
Bt2	58.2	1.83	3.9	24.8	0.84	1.8	33.4	1.38	2.9	
	CaX <sub>2</sub> -TylX									
Bt1	18.8	0.63	1.3	10.7	0.19	0.4	8.09	0.52	1.09*	
Bt2	24.5	0.98	2.1	16.2	0.33	0.7	8.27	0.76	1.61	

<sup>a</sup>Standard error of  $K_{\rm P}$ .

<sup>b</sup>The 95% confidence interval associated with  $K_{\rm P}$ .