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10-13-2020

## Effects of Low-Level Artificial Light at Night on Kentucky Bluegrass and Introduced Herbivore

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1 Word Count: 3,163  
2 Tables: 3  
3 Figures: 4

## 4 **Effects of low-level artificial light at night on Kentucky bluegrass and** 5 **introduced herbivore**

6 **Running Title:** Artificial light at night effects on bluegrass and herbivore  
7

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24 **Keywords:** photosynthesis, urban light, crickets, insects, growth rate.

### 25 **Abstract**

26 Increasing evidence suggests that artificial light at night (ALAN) can negatively impact  
27 organisms. However, most studies examine the impacts of ALAN on a single species or under high  
28 levels of artificial light that are infrequent or unrealistic in urban environments. We currently have  
29 little information on how low levels of artificial light emanating from urban skyglow affect plants  
30 and their interactions with herbivores. We examined how low levels of ALAN affect grass and  
31 insects, including growth rate, photosynthesis, and stomatal conductance in grass, and foraging  
32 behavior and survival in crickets. We compared growth and leaf-level gas exchange of Kentucky  
33 Bluegrass (*Poa pratensis*) under low-levels of ALAN (0.3 lux) and starlight conditions (night light at  
34 0.001 lux). Furthermore, each light treatment was divided into treatments with and without house  
35 crickets (*Acheta domesticus*). Without crickets present, bluegrass grown under artificial light at night  
36 for three weeks grew taller than plants grown under natural night light levels. Once crickets were  
37 introduced at the end of week three, grass height decreased resulting in no measurable effects of light  
38 treatment. There were no measurable differences in grass physiology among treatments. Our results  
39 indicate that low levels of light resulting from skyglow affect plant growth initially. However, with  
40 herbivory, ALAN effects on grass may be inconsequential. Gaining an understanding of how ALAN  
41 effects plant-insect interactions is critical to predicting ecological and evolutionary consequences of  
42 anthropogenic disturbance.

### 43 **1 Introduction**

44 Artificial light at night (ALAN) is an anthropogenic pollutant that is increasing spatially by a  
45 rate of 2.2% per year (Kyba et al., 2017). Direct ALAN sources, such as streetlights, can lead to  
46 skyglow: the atmospheric scattered light that can propagate up to several hundred kilometers into the  
47 environment (Aubé, 2015; Luginbuhl et al., 2009; Aubé, 2015). Skyglow results in light encroaching  
48 into natural areas where direct sources of light pollution are not present (Gaston et al., 2015; Garrett  
49 et al., 2020). The study of artificial light at night as an anthropogenic pollutant is a relatively young  
50 field (Longcore and Rich, 2004; Seymoure, 2018; Dominoni et al., 2020; Sanders et al., 2021), with  
51 most studies conducted at relatively high levels of nocturnal light pollution (e.g., 10-100 lux; (Gaston  
52 et al., 2013) but see (Alaasam et al., 2018; Sanders and Gaston, 2018). These high light levels are  
53 representative of organisms functioning under direct light pollution, such as directly beneath a  
54 streetlight, whereas most urban environments exist at lower light levels due to skyglow (e.g., 0.1 to 1  
55 lux), which can impact environments several hundred kilometers away from a direct light source  
56 (Gaston et al., 2013; Dominoni et al., 2014; Seymoure et al., 2019a). For reference, a full moon night  
57 could create ambient light levels of 0.3 lux on its brightest nights (Biberman et al., 1966; Kyba et al.,  
58 2017). Therefore, examining the impacts of light pollution at high intensities, although informative,  
59 is not representative of artificial light conditions in urban habitats at night. It remains an open  
60 question as to whether low levels of skyglow illumination (0.001 lux - 0.3 lux) affects communities  
61 to the same extent as direct illumination.

62 The intensity and spectral composition of light depends upon the phase of the moon, season,  
63 and weather, all of which create necessary cues for organisms (Kyba et al., 2015; Spitschan et al.,  
64 2016; Seymoure et al., 2019b). Plants use light as a cue for almost every physiological process  
65 including, but not limited to, seedling development, photosynthesis, growth, and budding (Takemiya  
66 et al., 2005; Bennie et al., 2016; Gaston et al., 2017; Singhal et al., 2018). Light influences plant  
67 growth, development, and photosynthetic efficiency (Briggs and Christie, 2002). In addition to  
68 powering the electron transport chain in thylakoid membranes, light intensity and direction increases  
69 photosynthetic efficiency through phototropism (i.e. the movement of the plant towards sunlight;  
70 (Celaya and Liscum, 2005), chloroplast movement (Wada et al., 2003), and light-induced stomatal  
71 opening to help optimize gas exchange efficiency (Dietrich et al., 2001). Periods of darkness are also  
72 important for plant metabolic processes, particularly stress recovery, which includes recovery from  
73 herbivory events (McNaughton, 1983; Singhal et al., 2018).

74 Increased levels of ALAN from urbanization are changing natural light regimes by increasing  
75 the intensity and duration of light available at night (Davies et al., 2013; Seymoure et al., 2019a;  
76 Buxton et al., 2020), potentially affecting plant photosynthesis, growth, and plant-herbivore  
77 interactions. For example, by masking natural night light levels, ALAN can mislead herbivores to be  
78 more active at night and disrupt plant-herbivore interactions and critical dark recovery periods for  
79 plants (Dominoni et al., 2020). Plants in light polluted environments experience changes in  
80 pollination, photoreceptor signaling, phenology and flowering (Ffrench-Constant et al., 2016; Singhal  
81 et al., 2018), which can have ecological consequences for food web dynamics (Polis et al., 2004).  
82 However, little is known about how constant illumination at the level of urban light alters plant-insect  
83 interactions. ALAN has led to declines in population sizes of a diversity of insect species through its  
84 interference with insect development, movement, foraging, and reproductive success, which can alter  
85 trophic systems (Owens and Lewis, 2018; Owens et al., 2020).

86 Here we test whether ALAN affects plant-insect interactions by modifying plant  
87 photobiology and growth rates. We exposed two common urban species—Kentucky bluegrass (*Poa*  
88 *pratensis*), a cool season common turfgrass (Weissman et al., 1977; Suplick-Ploense and Qian, 2005;  
89 Read et al., 1999; Weissman et al., 1977; Suplick-Ploense and Qian, 2005), and the house cricket

90 (*Acheta domesticus*), a nocturnal herbivore—to starlight (0.001 lux) and realistic urban night light  
91 levels (0.3 lux) (Dominoni et al., 2013; Alaasam et al., 2018; Seymoure et al., 2019a) in order to test  
92 the following hypotheses: 1) Low levels of ALAN affect plant physiology. We predicted that plants  
93 grown under urban light would have higher net photosynthesis and dark respiration, increased growth  
94 rates, and increased stomatal conductance compared to control plants grown under starlight  
95 conditions. 2) Herbivory interacts with ALAN to affect plant biomass. We predicted cricket  
96 herbivores would reduce the biomass and height of grass. However, as crickets are nocturnal  
97 foragers, we predicted they would consume less plant material under urban light than starlight  
98 conditions and have lower survival rates in urban light.

## 99 2 Materials and Methods

### 100 2.1 Light Treatments

101 We used a CMP6050 growth chamber (Version 4.06, Conviron, Winnipeg, Manitoba) set to a  
102 temperature of 22.2°C with light control to create artificial light environments (0.3 lux, hereafter  
103 “urban light”) and natural new moon light environments (0.001 lux, hereafter “starlight”)(Dominoni  
104 et al., 2013; Alaasam et al., 2018; Seymoure et al., 2019a; Jones et al., 2020). There were two  
105 different light types in the chamber - high pressure sodium and mercury vapor - placed in alternating  
106 positions on the ceiling of the chamber. To create urban light levels within the chamber, we used 4  
107 layers of filter gels over the light sources (Rosco E-Colour #211 .9 Neutral Density Filter, Stamford,  
108 CT) that attenuated 83% of light. To further attenuate light, 90% black shade cloth was placed over  
109 starlight treatments, and 22% white shade cloth was placed over urban light environments. These  
110 were constructed as square boxes and placed over the plant treatment groups using PVC pipe and  
111 shade cloth. We confirmed that light levels were approximately 0.3 lux and 0.001 lux using a highly  
112 sensitive spectroradiometer (StellarNet Silver Nova, Tampa Bay, FL) with a cosine corrected  
113 irradiance probe affixed to a 1000-micron optical fiber (StellarNet, Tampa Bay, FL). We checked  
114 irradiance measurements using SpectraWhiz software (StellarNet, Tampa Bay, FL); due to the low  
115 light levels, we set integration time to approximately 20 seconds for the 0.3 lux measurements and 8  
116 minutes for the 0.001 lux measurements. This confirmed that light levels throughout the enclosure  
117 were within one order of magnitude of the chosen light level for each treatment: 0.3 and 0.001 lux.

### 118 2.2 Experimental Design

119 On day 1, Kentucky bluegrass seeds were sown in 10 cm round pots (n=72) containing Scotts  
120 Miracle-Gro soil and placed in the growth chamber under experimental light conditions. On day 21,  
121 we measured the tallest blade of grass, then weeded down the pots randomly until there were 25  
122 shoots of grass remaining. After the initial 21-day growth period, one randomly selected juvenile  
123 cricket, male or female, was placed in each of 36 designated cricket pots. Herbivory and light  
124 environments were examined using a 2x2 factorial design in which light treatment was factorially  
125 crossed with cricket treatment in a 28-day experiment. The four treatments were arranged in a block  
126 test pattern, as shown in **Figure 1**. Treatment groups included: (1) plants without crickets in urban  
127 light, (2) plants without crickets in starlight, (3) plants with crickets in urban light, and (4) plants with  
128 crickets in starlight (n=18 per treatment). Nighttime lighting conditions were imposed in the middle  
129 of the day from start of the experiment to ensure nighttime measurements could be taken during  
130 regular working hours. Lighting conditions were altered twice daily; we placed filter paper and shade  
131 cloth structures over the plants at 08:00 and removed them at 18:00 to create a 14:10 light: dark cycle  
132 typical of summer in the northern hemisphere. Blocks were rotated daily one position clockwise to  
133 account for spatial variation in light levels within the chamber, and generously watered at this time.

134 Drierite (W.A. Hammond 23005, Xenia, OH) was placed in two trays on opposite sides of the  
 135 chamber to control humidity and prevent mold growth (Hammond, 1935).

136 Crickets were sourced as juveniles from a stock population from Premium Crickets (Winder,  
 137 Georgia) in December 2018 and May 2019 at the mean size of 1.9 centimeters, before adult morph.  
 138 From day 21 to 28, cricket survival was monitored daily (i.e., when light conditions were shifted) and  
 139 categorized as alive or dead. If a cricket was found dead, the cricket and its designated plant were  
 140 removed from the experiment. Upon removal, we measured the height of the tallest blade of grass  
 141 and recorded the length of time the plant/cricket spent in the chamber. We also cut and weighed  
 142 above ground biomass to determine wet and dry mass. On day 28, we removed all remaining plants  
 143 from the experiment and recorded the final height of the tallest blade of grass. We calculated the  
 144 average daily growth rate in week four (day 21 to day 28) to control for plants that were removed  
 145 prematurely due to cricket death.

### 146 **2.3 Gas Exchange Measurements**

147 To assess light treatment effects on bluegrass physiology independent of herbivory, we  
 148 measured leaf photosynthetic responses on day 19 before crickets were placed into pots. We  
 149 measured leaf gas exchange in each light treatment using a LI-6400XT infrared gas analyzer with a  
 150 leaf chamber fluorometer attached (Li-Cor Biosciences; Lincoln, NE) following previously published  
 151 methods with slight modifications (Lemoine et al., 2018). Plants were removed from the growth  
 152 chamber temporarily for gas exchange measurements. The environmental conditions inside the leaf  
 153 chamber were standardized across measurements; leaf temperature was maintained at 20°C, relative  
 154 humidity was maintained between 40-50%, sample chamber flow rate was set to 200  $\mu\text{mol s}^{-1}$ , and  
 155 reference chamber  $\text{CO}_2$  concentration was set to 400 ppm. Low flow settings are commonly used for  
 156 small leaved grasses with low photosynthetic rates (Taylor, 2014). Leaf level gas exchange was  
 157 measured under two light conditions: dark and low light (10  $\mu\text{mol m}^2 \text{s}^{-1}$  (740 lx) photosynthetically  
 158 active radiation; PAR). Gas exchange in the dark provides an estimation of leaf respiration. The low  
 159 light level was the minimum amount of light provided by the Li-6400 light source; thus, we were  
 160 unable to measure photosynthesis under the tested ALAN conditions imposed here (<10  $\mu\text{mol s}^{-1}$ , <740  
 161 lux), but instead measured whether treatments had an impact on plant photosynthetic responses to  
 162 low levels of light. Results are reported in regard to light treatment in the growth chamber (urban  
 163 light or starlight). A newly emerged and fully expanded leaf from each individual (n= 10 individuals  
 164 per treatment) was inserted into the leaf chamber. Prior to measurements, leaves were dark adapted  
 165 for 2 hours under a dark box that allowed no light to enter. Leaves were left in the chamber for 2-5  
 166 minutes to equilibrate to chamber conditions before gas exchange parameters (photosynthesis or  
 167 respiration, and stomatal conductance) were recorded (average of three logged values taken in rapid  
 168 succession). Steady-state fluorescence ( $F_s$ ) was measured continuously before exposing plants to a  
 169 saturating pulse of light (2750  $\mu\text{mol m}^{-2} \text{s}^{-1}$  of blue light or ~203,500 lux (Thimijan and Heins, 1983)  
 170 to measure maximum chlorophyll fluorescence. Light inside the chamber was then switched to the  
 171 low light level (10  $\mu\text{mol m}^2 \text{s}^{-1}$ ). Once gas exchange reached stability, net photosynthetic rate, and  
 172 stomatal conductance were recorded, and a saturating pulse was applied to estimate photosystem II  
 173 efficiency ( $\Phi\text{PSII}$ ):  $\Phi\text{PSII} = (F_m' - F_s)/F_m'$  where  $F_m'$  represents chlorophyll fluorescence under  
 174 low light. As grass blades rarely fill the entire chamber, the measured leaf area was estimated using  
 175 width and length, and photosynthetic parameters, which are based on the area of the chamber (6  $\text{cm}^2$ ),  
 176 were adjusted accordingly.

### 177 **2.4 Data Analysis**

178 All statistical analyses were performed in R version 3.4.3(R Development Core Team, 1999).  
179 We first confirmed the use of parametric tests to ensure our data was normally distributed. To test our  
180 first hypothesis that gas exchange increased under ALAN, we ran a MANOVA with net  
181 photosynthetic rate, stomatal conductance, dark respiration, and  $\Phi$ PSII as response variables and with  
182 light treatment and block as explanatory variables (**Figure 2**). For our second hypothesis that light  
183 and cricket treatments would affect plant height, we modeled daily percent change in height between  
184 day 21 and day 28 using a two-way ANOVA with light treatment, cricket treatment, and block as  
185 explanatory variables (**Figure 3**). We then analyzed the data using two-way ANOVA, again with  
186 light treatment, cricket treatment, and block as explanatory variables. We tested for an interaction  
187 between light treatment and cricket treatment, and we also analyzed cricket survival using Kaplan-  
188 Meier analysis with the “survival” package in R (**Figure 4**) (Therneau and Lumley, 2009).

### 189 **3 Results**

190 There was no difference in net photosynthesis, stomatal conductance, dark respiration, or  $\Phi$ PSII  
191 between grass grown in the two light treatments (**Table 1**). On day 21, bluegrass grown in urban light  
192 was taller (mean = 6.58cm, sd = 2.3) than bluegrass grown in starlight (mean = 7.10cm, sd = 2.67,  
193 **Table 2**). However, daily percent change in plant height from day 21 to day 28 was not significantly  
194 different (**Table 3**). The presence of crickets did affect plant height, whereby bluegrass with crickets  
195 present were shorter than bluegrass without crickets (**Table 3**).  
196 Crickets in the urban light treatment had a 25.0% probability of survival, whereas crickets in the  
197 starlight treatment had a survival probability of 32.1%, but this difference was not significant  
198 (Kaplan-Meier: n = 36, p = 0.37, see supplemental material). There was no difference in survival due  
199 to sex (Kaplan-Meier: n= 36, p= 0.80, see supplemental material).

### 200 **4 Discussion**

201 Our study explored how low levels of artificial light at night, which are widespread across  
202 ecosystems, may affect plants and plant-insect interactions. Contrary to our predictions, grass grown  
203 under urban light conditions after 19 days did not have higher net photosynthetic rates than those  
204 grown under starlight, nor did stomatal conductance, dark respiration, or  $\Phi$ PSII differ significantly  
205 between light treatments. However, plants under urban light conditions grew taller than plants grown  
206 under starlight conditions during the initial 21 days of growth before crickets were introduced.  
207 Additionally, we found no evidence that crickets under urban light consumed more plant matter than  
208 crickets in starlight treatments, and survival rates of crickets did not differ between treatments. The  
209 results from this study suggest that low levels of ALAN may not have significant effects on grass  
210 photobiology but may affect plant height.

211 Studies investigating grass responses to higher levels of illumination (e.g.,  $4 \pm 1 \mu\text{mol}\cdot\text{m}^{-2}\cdot\text{s}^{-1}$   
212 or 296 lux) found that plant photoreceptors were damaged causing changes to flowering phenology  
213 (Thimijan and Heins, 1983; Shin et al., 2010; Bennie et al., 2016). The lower levels of light tested  
214 here were likely not bright enough to induce these changes in bluegrass. Plants often use nighttime  
215 darkness to repair damage from UV rays, suggesting the low levels of ALAN in our treatments may  
216 be dark enough for plants to continue to repair damaged cells and photoreceptors (Singhal et al.,  
217 2018). Moreover, net photosynthesis is a dynamic measurement that can vary within samples due to  
218 time and day(Miller et al., 1996) and our single measurement at the end of week 3 may not have  
219 captured treatment differences occurring at other times.

220 We found no difference in stomatal conductance or respiration between plants grown in urban  
221 light and starlight. Other studies have noted differences in stomatal density and stomatal opening and  
222 closing in the presence of ALAN (Takemiya et al., 2005; Shimazaki et al., 2007). Another study  
223 found that yellow-poplar trees exposed to ALAN (high pressure sodium lighting ranging from 82 lx  
224 to 4100 lx) for three years resulted in reduced nighttime stomatal conductance (Kwak et al., 2018). It  
225 is possible that our light levels were too low, or grass was not subjected to our light levels for a long  
226 enough duration to induce such responses. Reduced chlorophyll and rubisco concentration has been  
227 observed in phytoplankton grown under low light levels (6.6 lux;(Poulin et al., 2014), and light as  
228 low as 3.5 lux has induced flowering in tree species across the United Kingdom (Ffrench-Constant et  
229 al., 2016). We also observed no treatment effects on photosystem II efficiency despite other studies  
230 noting adverse reactions in these physiological responses to light pollution (Zhang and Reisner, 2019;  
231 Meravi and Prajapati, 2020). Kentucky Bluegrass might be more adaptable to changing light regimes  
232 given that it is commonly used as a turf grass selected for its resilience to drought and heat stress  
233 (Wang and Huang, 2004). We observed a faster growth rate for grasses grown under urban light  
234 conditions compared to starlight conditions. Plant growth rate is determined by a variety of factors,  
235 including, but not limited to, photosynthetic rate, specific leaf area, leaf mass fraction, and nitrogen  
236 absorption rate(Poorter et al., 1991; Osone et al., 2008). Although we found no difference in net  
237 photosynthetic rate between treatments, growth rate differences could have been due to greater  
238 allocation to leaf area in urban light(Poorter and Remkes, 1990), although we did not measure such  
239 attributes.

240 ALAN is known to alter photoperiod detection in multiple organisms (Bennie et al., 2016)  
241 and these changes in photoperiod can impact plant growth and flowering (Cathey and Campbell,  
242 1975; Blanchard and Runkle, 2010; Basler and Körner, 2012; Craig and Runkle, 2016). Increased  
243 growth and biomass have been noted in *Poaceae* species when exposed to high levels of ALAN  
244 ranging from 0.349 - 1.145 $\mu\text{mol m}^{-2} \text{sec}^{-1}$  from metal halide bulbs (Flowers and Gibson, 2018),  
245 which is approximately 24.78 - 81.30 lux (Thimijan and Heins, 1983). Since we noted no change in  
246 Kentucky Bluegrass, photoperiod detection may not have been disrupted at our lower levels of  
247 ALAN, or it may have caused undetectable or non-measured physiological responses.

248 While animals rely on plants as a food source and shelter, we found no evidence that low-  
249 level light pollution would impact these typical interactions between plants and insects. Artificial  
250 light at the level of 0.3 lux was not significant enough to mask natural light cues in herbivores, nor  
251 mislead herbivores in foraging behaviors, but light pollution at higher levels could modify these  
252 interactions(Gaston et al., 2013; Macgregor et al., 2015; Bennie et al., 2016; Knop et al., 2017). High  
253 levels of ALAN could mask lunar cues, disrupting invertebrate behavior and feeding patterns and  
254 could attract invertebrates to artificially lit structures, deterring them from normal behavioral patterns  
255 (Longcore and Rich, 2004; Seymoure, 2018; Dominoni et al., 2020; Sanders et al., 2021).

256 Overall, our research detected few changes to plant physiology at low levels of urban light,  
257 suggesting that low levels of ALAN may not be as harmful to community interactions as predicted.  
258 Other studies conducted at high levels of ALAN suggest artificial light can induce large changes in  
259 physiology and community interactions(Longcore and Rich, 2004; Gaston et al., 2013; Seymoure et  
260 al., 2019a). There may be a threshold level at which artificial light becomes harmful, causing  
261 detrimental effects to individual and ecosystem function with additional increases in intensity and  
262 duration. Understanding and identifying this threshold would allow for more effective management  
263 of night skies and natural light conditions(Dominoni et al., 2020). With estimates suggesting two  
264 thirds of Key Biodiversity Areas experience ALAN(Seymoure et al., 2019a; Garrett et al., 2020), it is



265 important to identify the level at which artificial light becomes harmful and how natural night skies  
266 can be managed.

## 267 **5 Acknowledgements**

268 This work was supported through a Zoological Lighting Institute Grants-In-Aid of Research  
269 grant awarded to MC and CB. MC was awarded a SEEDS grant to present this research at the 2019  
270 meeting of the Ecological Society of America (ESA) where we received excellent feedback from the  
271 ESA community. Furthermore, this work was supported through the Colorado State University  
272 Honors Program. We are grateful for support from the Smith Lab, the Sound and Light Ecology  
273 Team at Colorado State University, and the Natural Sounds and Night Skies Division of the Natural  
274 Park Service. Jeremy White, Tammy Brenner, and Bob Meadows were foundational to the success of  
275 this study.

276 *The authors declare that the research was conducted in the absence of any commercial or financial*  
277 *relationships that could be construed as a potential conflict of interest.*

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439 **Table 1** MANOVA table of the gas exchange results evaluating differences in photosynthesis,  
 440 stomatal conductance in dark, stomatal conductance in light, fluorescence, and photosystem II  
 441 efficiency.

	<i>df</i>	<b>Pillai</b>	<i>f</i>	<i>p</i>
<b>Treatment</b>	<b>1</b>	<b>0.18</b>	<b>0.45</b>	<b>0.83</b>
<b>Block</b>	<b>3</b>	<b>0.95</b>	<b>1.09</b>	<b>0.40</b>
<b>Residuals</b>	<b>17</b>			

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443 **Table 2** ANOVA table comparing mean grass height at day 21 across light treatments and blocks. \*  
 444 indicates a significant response.

	<b>Sum of Squares</b>	<i>df</i>	<b>Mean Square</b>	<b>F</b>	<i>p</i>
<b>Light Treatment</b>	<b>3.50</b>	<b>1</b>	<b>3.50</b>	<b>5.63</b>	<b>0.021*</b>
<b>Block</b>	<b>7.87</b>	<b>6</b>	<b>1.31</b>	<b>2.11</b>	<b>0.064</b>
<b>Residuals</b>	<b>39.8</b>	<b>64</b>	<b>0.622</b>		

445

446 **Table 3** ANOVA table showing the effects of light treatment, cricket treatment, and block (plus  
 447 interactions between light and cricket treatment and cricket and block treatment) on daily percent  
 448 change in grass height between day 21 and the end of the experiment. \* indicates a significant  
 449 response.

	<b>Sum of Squares</b>	<i>df</i>	<b>Mean Square</b>	<b>F</b>	<i>p</i>
<b>Light Treatment</b>	<b>0.14</b>	<b>1</b>	<b>0.14</b>	<b>1.60</b>	<b>0.21</b>
<b>Cricket Treatment</b>	<b>2.82</b>	<b>1</b>	<b>2.82</b>	<b>32.04</b>	<b>5.3 x 10<sup>-7</sup>*</b>
<b>Block</b>	<b>0.85</b>	<b>6</b>	<b>0.14</b>	<b>1.62</b>	<b>0.16</b>
<b>Light: Cricket</b>	<b>0.002</b>	<b>1</b>	<b>0.002</b>	<b>0.023</b>	<b>0.88</b>
<b>Cricket: Block</b>	<b>0.90</b>	<b>6</b>	<b>0.15</b>	<b>1.70</b>	<b>0.14</b>
<b>Residuals</b>	<b>4.93</b>	<b>56</b>	<b>0.088</b>		

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453 **Figure 1:** Aerial view of treatment groups in the growth chamber after crickets were introduced (day  
454 21-28). The treatment groups were arranged in a block test pattern with 4 blocks of urban light  
455 treatments and 4 blocks of starlight treatments, totaling 8 groups (A-H). Within each block (A-H),  
456 nine plants (every other one) had a cricket.

457 **Figure 2:** (A) Net photosynthesis across light treatments, measured under low light conditions (10  
458  $\mu\text{mol m}^{-2} \text{s}^{-1}$  of light) and (B) stomatal conductance across light treatments. (C) Photosystem II  
459 efficiency is measured using a saturating pulse ( $\Phi\text{PSII}$ ):  $\Phi\text{PSII} = (\text{Fm}' - \text{F}_s)/\text{Fm}'$  where Fm is  
460 chlorophyll fluorescence under low light. (D) Dark respiration measured under low light level ( $<10$   
461  $\mu\text{mol m}^{-2} \text{s}^{-1}$  of light). There were no differences in net photosynthesis, stomatal conductance,  
462 Photosystem II efficiency, or dark respiration between light treatments.

463 **Figure 3:** (A) Bluegrass height at day 21 separated by light treatment when no crickets were present.  
464 Grass in urban light was taller than grass in starlight conditions. (B) Daily percent change in height of  
465 grass (change from day 21 to day 28 divided by the number of days in the chamber) separated by  
466 light treatment. There was no difference in daily percent change across light or cricket treatments.

467 **Figure 4:** Survival probability of crickets. (A) Survival probability of crickets under urban light and  
468 starlight treatments. (B) Survival probability of crickets under urban light and starlight treatments,  
469 split by sex in each treatment group. In all both comparisons (A-B), there were no differences in  
470 survival.