

Extensive sheep grazing is associated with trends in steppe birds in Spain: recommendations for the Common Agricultural Policy

Juan Traba^{1,2} and Cristian Pérez-Granados^{1,3}

¹ Ecology, Universidad Autónoma de Madrid, Madrid, Madrid, Spain

² Centro de Investigación en Biodiversidad y Cambio Global (CIBC-UAM), Universidad Autónoma de Madrid, Madrid, Madrid, Spain

³ Ecology, Universidad de Alicante, Alicante, Alicante, Spain

ABSTRACT

Iberian natural steppes have traditionally been used for extensive sheep grazing, which has been noted to be positively associated with steppe bird abundance and diversity. Sheep numbers in Spain, which harbors the largest European populations of many steppe bird species, decreased by 9.2 million (37.3%) between 1992 and 2020. Steppe birds in Spain have faced dramatic declines during the same period, but there is a lack of knowledge about the potential association between sheep and open-habitat bird declines. We used sheep data from the Spanish Ministry of Agriculture and bird data (1998–2018) from the Spanish Common Bird Monitoring Program to assess the association at the Spanish scale between sheep decline and the Farmland Bird Index (FBI) and the Natural Shrub-steppe Bird Index (SBI). We also used an independent dataset on population trends of the Dupont's Lark (*Chersophilus duponti*) to assess the relationship between sheep numbers and the decline of this threatened steppe specialist passerine in Spain, whose European population is restricted to Iberian natural steppes. To test for a spurious relationship between temporal series, variables were tested for cointegration. After confirming cointegration, we found a strong positive relationship between sheep abundance and the trends of the FBI and SBI indices during the period 1998–2018. The association between sheep abundance and trends of the Dupont's Lark (2004–2015) was positive although it was not statistically significant. Although the main causes of decline of farmland and steppe birds are mainly related to agricultural intensification and land use changes, the correlation found, using two independent cointegrated datasets, between the reduction in farmland and shrub-steppe birds and sheep numbers at the country scale suggests that the decline of steppe birds in Spain may be also associated with the decline in sheep numbers. This agrees with previous studies that found a positive relationship between intermediate levels of sheep grazing and steppe bird abundance in Iberian steppes. Further research (*e.g.* experimental studies) is needed to corroborate our study and identify the most appropriate level of grazing intensity for protecting the most farmland and shrub-steppe birds. Our results suggest that the promotion of extensive grazing should be considered as a key factor in future Common Agricultural Policy reforms and conservation programmes to protect steppe birds.

Submitted 12 October 2021
Accepted 11 January 2022
Published 28 February 2022

Corresponding author
Cristian Pérez-Granados,
cristian.perez@ua.es

Academic editor
David Stern

Additional Information and
Declarations can be found on
page 12

DOI [10.7717/peerj.12870](https://doi.org/10.7717/peerj.12870)

© Copyright
2022 Traba and Pérez-Granados

Distributed under
Creative Commons CC-BY 4.0

OPEN ACCESS

Subjects Agricultural Science, Biodiversity, Conservation Biology, Ecology, Zoology

Keywords *Chersophilus duponti*, Farmland bird index, Steppe birds, Extensive grazing, Sheep, Steppes

INTRODUCTION

European natural and semi-natural open habitats, such as steppes, uplands and moorlands (natural steppes hereinafter), are generally the result of altitudinal/climatic and edaphic tree-limiting conditions, and regular low-intensity human disturbances such as low-yield agriculture, fire or livestock (*Carboni et al., 2015*). All these factors have shaped the characteristic open, treeless landscape of such habitats (*Newton, 2004; Douglas et al., 2017*), which harbors high biodiversity values (*Carboni et al., 2015*). In these economically disadvantaged areas, sheep systems have been decisive in maintaining rural livelihoods, but also providing several ecosystem services, such as protecting biodiversity and maintaining natural resources (*Spiegel et al., 2019*). However, sheep farm income in the European Union is supported by public subsidies, such as the Common Agricultural Policy (*Bertolozzi, 2021*) (CAP hereinafter), which makes the sector strongly dependent on changes in the policy framework (*De Rancourt et al., 2006; Castillo et al., 2008; Reino et al., 2010*).

Several works have suggested the existence of a complex system of multiple interactions where extensive sheep grazing affects plant structure spatial heterogeneity and plant species composition (*Kurek et al., 2016; Adler, Raff & Lauenroth, 2001; Bugalho et al., 2011*), as well as arthropod abundance both directly through its depositions and indirectly, by increasing habitat heterogeneity (*Bugalho et al., 2011; Vickery et al., 2001; Dennis et al., 2008*). Heavy levels of grazing have proven to be disadvantageous for a large number of farmland and steppe birds, due to overconsumption of plants and high fertilization (*Fuller, 1996*). However, light or moderate grazing may have positive effects since grazing decreases vegetation height (*Vickery et al., 2001; Leal et al., 2019*), and facilitates the occurrence of coprophagous species (*De Rancourt et al., 2006; Jay-Robert et al., 2008; Perrin et al., 2019*). Moreover, the foraging efficiency of steppe birds increases in more open habitats (*Fuller, 1996; Haworth & Thompson, 1990; Buckingham & Peach, 2005; Zbyryt, Sparks & Tryjanowski, 2020*) and greater visibility in sparser vegetation minimizes the risk of predation for ground-standing birds (*Buckingham & Peach, 2005; Whittingham & Evans, 2004*). All of these factors may favor the abundance and space use of insectivorous birds (*Newton, 2004; Douglas et al., 2017; Gómez-Catasús et al., 2019; Smith et al., 2020*). As a result, numerous threatened bird species have their strongholds in these extensively sheep-grazed natural steppes (*Velado-Alonso et al., 2020*). Of particular conservation concern are the Iberian natural steppes (shrub steppes, *sensu* (*Traba, Sastre & Morales, 2013*)), which, as with many Mediterranean calcareous grasslands, are the consequence of both anthropogenic and natural processes, with an important role played by extensive grazing by sheep in the spring and autumn (*Suárez, 1994; Sainz Ollero, 2013*). In Spain, sheep are traditionally reared in extensive or semi-extensive systems due

to the hardiness of the autochthonous breeds and their good adaptation to adverse environmental conditions ([Castillo et al., 2008](#)).

Iberian natural steppes are among the European landscapes with the highest value for biodiversity conservation, since they harbor the main (or the entire) European population of several threatened steppe bird species, such as the Dupont's lark (*Chersophilus duponti*), the Greater short-toed lark (*Calandrella brachydactyla*), the Stone curlew (*Burhinus oedicnemus*), the Little bustard (*Tetrax tetrax*), both the Black-bellied (*Pterocles orientalis*) and Pin-tailed sandgrouses (*Pterocles alchata*), and the Tawny pipit (*Anthus campestris*), among many others ([Santos & Suárez, 2005](#)). Steppe and farmland birds are the most threatened group of birds in Europe ([European Environment Agency, 2015](#)) and overall 83% of the steppe bird species show an unfavorable conservation status in Europe ([Burfield, 2005](#)). Agricultural intensification, together with habitat loss and land-use changes have partly caused the large declines detected over the past several decades for open-habitat bird populations ([Carboni et al., 2015](#); [Santos & Suárez, 2005](#); [Fuller & Gough, 1999](#); [Traba & Morales, 2019](#)). In marginal regions such as those of Iberian natural steppes, where agricultural income has been traditionally low, agricultural intensification is also occurring at a field scale ([Reverter et al., 2021](#)). However, the negative trends for several steppe birds that only inhabit primary stages of vegetation succession are especially linked to shrub encroachment ([Robinson & Sutherland, 2002](#); [Herrando et al., 2014](#); [Holmes, Maestas & Naugle, 2017](#)). The spread of shrubs and dominant graminoids is one of the major consequences of grazing abandonment, altering local floristic and edaphic conditions ([Carboni et al., 2015](#); [Koch et al., 2015](#)) and ultimately affecting to the entire animal community ([Robinson & Sutherland, 2002](#)).

During the second half of the 20th century, land abandonment due to rural depopulation generated a reduction in extensive sheep farming in Spain, as well an increase in stocking rates ([Soto et al., 2016](#); [Martínez-Valderrama et al., 2021](#)). However, changes in CAP subsidies during the 21st century have also led to important structural changes in the sector ([De Rancourt et al., 2006](#)). An example is the decoupling of CAP payments in 2003, which provoked a decrease in sheep numbers in less-favoured areas, while the persistence of CAP subsidies coupled to cattle and beef-cattle production has provoked a shift from sheep-based to cattle-based systems in permanent pastures ([Ramos et al., 2021](#); [Faria & Morales, 2020](#); [Mújica et al., 2015](#)), usually with an increase in stocking rate. All of these changes have had a great impact on land use ([Soto et al., 2016](#); [Riedel, Casasús & Bernués, 2007](#)) and have led to an increase in shrub and tree cover in these areas ([Martínez-Valderrama et al., 2021](#)). While a number of studies have assessed the relationship between grazing pressure and bird abundance in farmlands and uplands in the United Kingdom ([Newton, 2004](#); [Douglas et al., 2017](#); [Fuller & Gough, 1999](#)), our current knowledge about the association between grazing and farmland and steppe birds in Mediterranean landscapes is still limited (see [Leal et al., 2019](#); [Riedel, Casasús & Bernués, 2007](#); [Reino et al., 2010](#)).

In this study, we aimed to assess the variation in steppe bird abundance in relation to changes in sheep abundance in Spain over the period from 1998–2018. We estimated the nation-wide relationship between trends in sheep abundance with an annual index in

population trends estimated for 35 common farmland bird species (Farmland Bird Index, [Table S1](#)) and for 20 common bird species typical of Iberian natural shrub-steppes (Steppe Bird Index, [Table S2](#)) to evaluate the potential impact of changes in sheep numbers on abundance of open-habitat bird species. Additionally, we also tested the relationship between trends in sheep numbers and population trends of the Dupont's lark (*Chersophilus duponti*). We selected this species because it is a threatened habitat-specialist whose European population is restricted to Iberian natural shrub-steppes ([Gómez-Catasús et al., 2018](#)), and for which accurate data have been collected during the last 15 years ([Gómez-Catasús et al., 2018](#)). Moreover, the promotion of extensive grazing has been traditionally described as one of the main conservation interventions to maintain optimal vegetation structure for the Dupont's lark (reviewed by [Pérez-Granados & López-Iborra \(2021\)](#)). Based on the assumed positive impact that low-to-moderate grazing pressure has on steppe bird abundance ([Vickery et al., 2001](#)), we expected to find a close association between the decline in sheep numbers and the decline in the population indices estimated for the community of farmland and steppe birds, as well as for the Dupont's lark. Our results might be useful in discussing the relationship between sheep numbers and farmland and steppe birds and to formulate recommendations for agri-environment schemes for these threatened groups of species.

MATERIAL AND METHODS

Study area

We focused the study in Spain, since this country is host to the largest proportion of steppe birds in Europe ([Santos & Suárez, 2005](#)) and is ranked second in terms of numbers of sheep (ca. 19% of the total ([Eurostat, 2020](#))).

Estimation of trends in sheep numbers

Annual data on total number of sheep in Spain were obtained from the General Directorate of Agricultural Productions and Markets (GDPME), of the Spanish Ministry of Agriculture, Fisheries and Food MAPA ([MAPA, 2020](#)), for the period 1992–2020. Data come from surveys collected from all of the country's agrarian districts, and later scaled up to the province level. This is the only exhaustive nation-wide information on livestock numbers and trend. We restricted our analyses to sheep livestock and did not include goat livestock, since goat numbers are low in Spain (ca. 17% in respect to sheep livestock for the year 2020) and mainly restricted to certain regions ([MAPA, 2020](#)). We estimated the trends in sheep numbers in Spain by calculating the trend (%) since: (a) 1992, for the whole time period, (b) 1998, for comparisons with Farmland and Shrub-steppe Bird Indices (see below), and (c) since 2004 for comparisons with Dupont's lark trends, using each mentioned year as the reference value (zero) in each analysis.

Estimation of trends in bird populations

For estimating farmland and steppe bird trends, we used bird data from the Spanish Common Breeding Bird Monitoring Program (SACRE), which is hosted by the Spanish representative of BirdLife (SEO/BirdLife). SACRE data is provided by volunteers who

perform bird censuses in a set of 10×10 km cells distributed across the country (see Fig. S1). These sites are sampled annually, during the breeding season, following a standardized methodology. The program currently involves over 1,000 sites (Gordo, 2018). Although SACRE started in 1996, we selected data for the period 1998–2018, since during the first 2 years of the program data were collected from a reduced number of sites.

Census data are analysed by SEO/BirdLife, which provides a bird population abundance index for each species and year, estimated using the Trend and Indices for Monitoring data (TRIM) software by fitting log-linear regression models to count data with Poisson error terms (Pannekoek & Strien, 2006). These indices were converted on an annual trend (%), using the year 1998 as baseline and following the same method described to estimate trends in sheep numbers. The trend of the TRIM population index can be used as an estimate of annual variations in the abundance of bird species (Gordo, 2018; Stjernman et al., 2013) (see (Traba & Morales, 2019) for a similar procedure). Indeed, the SACRE program provides the best information of population trends for common breeding bird species in Spain (SEO/BirdLife, 2016; Casas et al., 2019). The indices of a group of species can be summarized into a single estimate to analyze trends of related bird species (Bowler et al., 2019). Among these multi-species indices, the Farmland Bird Index (FBI) is notable. This is a summary population index that includes information of species classified as common farmland birds (e.g. Stjernman et al., 2013). The FBI is an official index, adopted by the European Union, of the quality of EU's agroecosystems for biodiversity and the effectiveness of agri-environmental interventions applied under European CAP (Eurostat, 2020). We used the official data provided by SEO/BirdLife for the FBI index in Spain, which comprised data for 35 common farmland bird species (see Table S1). Complementarily, we built a combined population index for a subset of 20 common species typical of Iberian natural shrub-steppes (Steppe Bird Index, SBI, see Table S2; for species selection see (Traba, Sastre & Morales, 2013; Santos & Suárez, 2005; Traba et al., 2007)), to further explore the relationship of birds typical of shrub-steppes with the variation in sheep numbers. The data about bird population index for each species and year was also provided by SEO/BirdLife.

Finally, we used the Dupont's Lark (*Chersophilus duponti*), as a threatened habitat-specialist of Mediterranean natural steppes, to evaluate the relationship between the abundance of this strong steppe-specialist lark and changes in sheep numbers. Due the difficulty of censusing this species using traditional surveys applied during the SACRE (Pérez-Granados & López-Iborra, 2017), and thus the uncertainty in SACRE's Dupont's lark data, we used an independent dataset collected using a standardized, species-specific, counting method. Trend of Dupont's Lark population was estimated by Gómez-Catasús et al. (2018) over the period 2004–2015 in Spain, using the year 2004 as baseline. This dataset comprised 42% of the Spanish populations of the species, which can be considered as representative of the population trends of the species in Spain (Gómez-Catasús et al., 2018). Annual trends were also estimated using the software TRIM (Gómez-Catasús et al., 2018), similar to the method described for the SACRE program.

Statistical analyses

The application of regression models to time series can generate spurious relationships between the considered variables, which although statistically related, may be purely coincidental (*Granger & Newbold, 1974*). Durbin Watson is a test statistic used to detect the presence of autocorrelation in the residuals of a regression analysis. If no causal relationship between the variables in a regression exists, the variance increases over time (non-stationarity) and affects the Durbin Watson statistic, whose low values may be due to this problem, and the coefficient of determination, which may reach high values close to 1 (*Granger & Newbold, 1974*). To deal with these problems, we carried out cointegration analyses between variables in the regressions described below. Two variables are cointegrated when they increase or decrease synchronously and maintain their relationship over time, suggesting a causal relationship (*Engle & Granger, 1987*). That is, when subtracting its expected value from the dependent variable, it is stationary (that is, mean value is stable along the time series) (*Engle & Granger, 1987*). We used the Johansen's method, which assesses the validity of the cointegrating relationship using a maximum likelihood estimates approach, to determine whether both variables are cointegrated. The null hypothesis for such test is that there is no cointegration between variables (*Johansen, 1988*). Once it is determined if two temporal series are cointegrated, traditional regression methods can be interpreted.

We fitted a linear regression to estimate the trend of sheep numbers over the period 1992–2020. We fitted independent linear regressions to assess the relationship between bird population trends (FBI, SBI) and sheep numbers, using trends of bird population as a variable response and the trend in sheep numbers as an independent variable over the period 1998–2018, using the year 1998 as a reference value (0). Finally, we assessed the relationship between the trends for Dupont's lark and sheep numbers through linear regression. We used the overall trends of Dupont's lark as a response variable and the trend in sheep numbers as an independent variable over the period 2004–2015, using the year 2004 as a reference value (0).

All statistics were performed with R 3.6.2 (*R Development Core Team, 2019*), using the package *lme4* (*Bates et al., 2015*) for linear regressions and packages *vars* and *urca* for cointegration analyses (*Bernhard, 2008*). The level of significance was $p < 0.05$.

RESULTS

Sheep numbers significantly declined in Spain during the period 1992–2020 (*Fig. 1*). In 2020, there were 37.3% fewer sheep than in 1992 (−9,175,782 sheep; linear regression, F-statistic = 128.2.10, df = 27, adjusted $R^2 = 0.82$; $p < 0.001$, see *Table S3* for full statistics, *Fig. 1*). The Johansen test showed significant cointegration between bird indices and sheep trends, which indicates that our results are unlikely to be due to spurious relationships: Spanish farmland bird trend (FBI) vs. Sheep trends Johansen test = 26.82; $p < 0.01$; Spanish steppe bird trend (SBI) vs. Sheep trend: Johansen test = 27.31; $p < 0.01$; Dupont's lark trend (FBI) vs. Sheep trend: Johansen test = 25.02; $p < 0.01$.

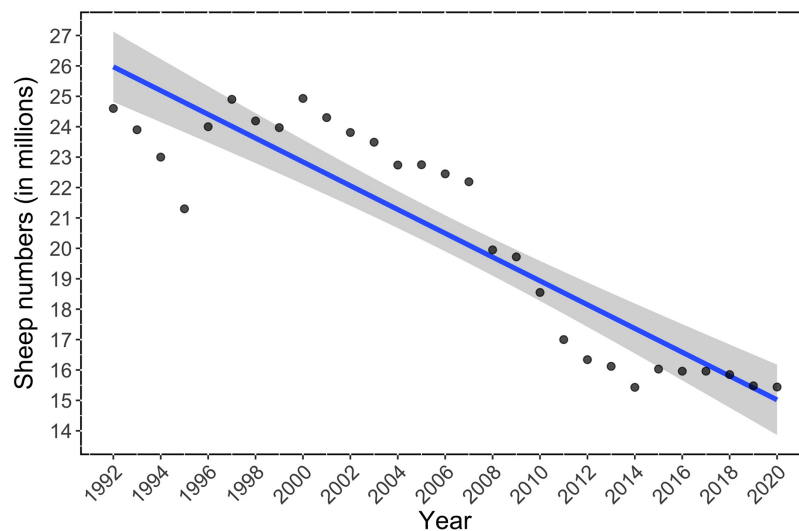


Figure 1 Temporal trend in the number of sheep in Spain during the period 1992–2020. The linear regression is shown in blue, and 95% Confidence Intervals in grey (linear regression: adjusted $R^2 = 0.82$; $p < 0.0001$). [Full-size !\[\]\(fd7fe780e8fd8eece60268c87d0c3e04_img.jpg\) DOI: 10.7717/peerj.12870/fig-1](https://doi.org/10.7717/peerj.12870/fig-1)

We found a strong, positive and significant relationship between the trend of Spanish farmland birds (FBI) during the period 1998–2018 and the sheep trend (linear regression, F-statistic = 39.85, $df = 19$, adjusted $R^2 = 0.66$, $p < 0.001$, Fig. 2A, Table S3). Similarly, a positive and significant association was found between the trend of natural steppe birds (SBI) and the sheep trend for the same period (linear regression, F-statistic = 27.4, $df = 19$, adjusted $R^2 = 0.57$, $p < 0.001$, Fig. 2B, Table S3). The linear relationship between Dupont’s lark trend and the sheep trend over the period 2004–2015 showed a positive relationship although it was not statistically significant (linear regression, F-statistic = 4.44, $df = 9$, adjusted $R^2 = 0.26$, $p = 0.06$, Fig. 3, Table S3).

DISCUSSION

Our results show a strong decline in sheep numbers in Spain during the last three decades, which is in agreement with the decreasing number of sheep livestock in many European countries: 15% for the period 2002–2018 (Eurostat, 2020; Pollock et al., 2013; Varga & Molnár, 2014; Varga et al., 2016; De Arriba & Barac, 2018). Nonetheless, the sheep decline detected in Spain has been especially steep since 2007, firstly, and even more so since 2009, when the number of sheep drastically declined (Fig. 1). This decrease in sheep numbers may be related to changes in CAP subsidies and largely explained by the uncoupling of sheep subsidies that started in 2006 and became permanent in 2010 (Mújica et al., 2015). After controlling for spurious correlations, we found a very strong and positive relationship between the reduction in sheep numbers and the decline in farmland and shrub-steppe birds at the national scale. The relationship between sheep numbers and the population trend of the Dupont’s Lark, a strong habitat specialist of Mediterranean natural steppes (Suárez, 2010), was just close to significant ($p = 0.06$). This result was

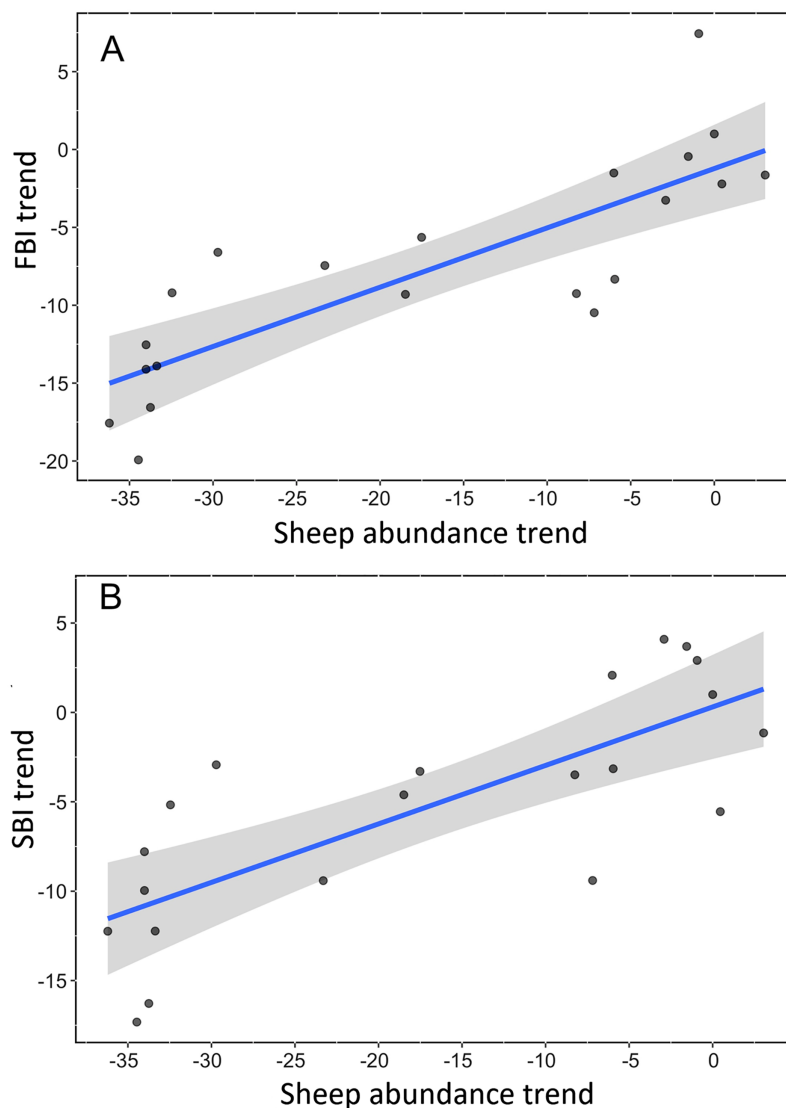


Figure 2 Linear relationship between trends of bird communities and trend of sheep abundance in Spain during the period 1998–2018. All variables are represented as the percent reduction using the year 1998 as the reference value (0). (A) Farmland Bird Index (adjusted $R^2 = 0.66$; $p < 0.0001$). (B) Steppe Bird Index (adjusted $R^2 = 0.56$; $p < 0.001$). [Full-size !\[\]\(5fd6ef84f97f42d7f8b34275f1b65312_img.jpg\) DOI: 10.7717/peerj.12870/fig-2](https://doi.org/10.7717/peerj.12870/fig-2)

obtained using an independent dataset (Gómez-Catasús *et al.*, 2018) and at a shorter temporal scale (11 vs. 20 years) than FBI and SBI analyses. The lack of a significant relationship for the Dupont's Lark case might be partly related to the more restricted distribution range of the species, which may make it difficult to detect nation-wide changes. Likewise, the decline in the Dupont's Lark has been associated with other land-use changes due to human activity, such as habitat loss due to ploughing (Tella *et al.*, 2005) and the development of wind farms (Gómez-Catasús, Garza & Traba, 2018), which may have muddled our results.

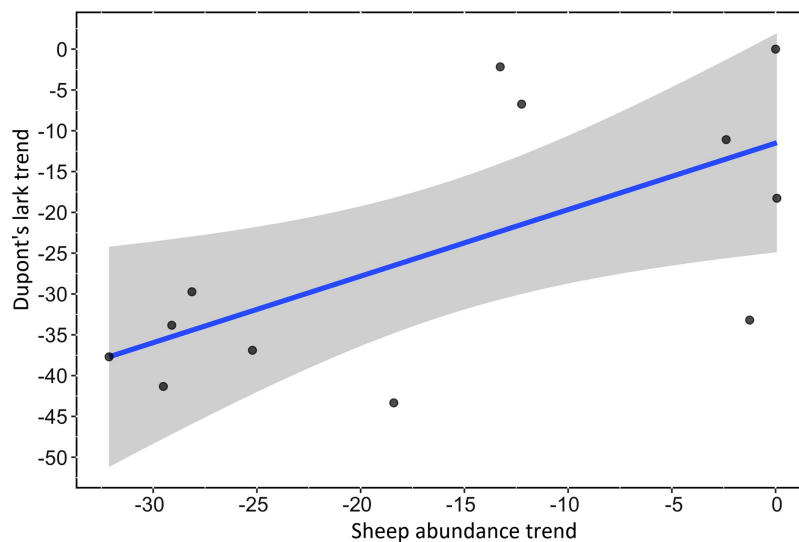


Figure 3 Linear relationship between Dupont's Lark trend and trend of sheep abundance in Spain during the period 2004–2015 (adjusted $R^2 = 0.25$; $p < 0.0649$). Both variables are represented as the percent reduction using the year 2004 as the reference value.

Full-size DOI: [10.7717/peerj.12870/fig-3](https://doi.org/10.7717/peerj.12870/fig-3)

We are aware that the decline in farmland and shrub-steppe birds in Spain is not exclusively explained by the decline in sheep numbers, but our results and previous studies strongly suggest that the significant relationships found are not simple correlations. It is well known that the negative trends in farmland and steppe birds in Europe are mainly related to agricultural intensification and land use changes, such as the decline in fallow lands (Traba & Morales, 2019; McMahon et al., 2010), which traditionally were an important grazing source (Ramos, Robles & González-Rebollar, 2010), as well as the increase in the use of pesticides and herbicides (Boatman et al., 2004; Chiron et al., 2014), higher mechanization (Santangeli, Di Minin & Arroyo, 2014), or the increase in trellis vineyards and irrigated woody crops, which are largely unsuitable habitats for the considered groups of birds (De Frutos, Olea & Mateo-Tomás, 2015; Casas et al., 2020), to cite a few. Velado-Alonso et al. (2020) recently found strong geographical associations between steppe bird richness and local sheep breed richness, which may be interpreted as an indicator of the intensity of extensive sheep grazing. Though these relationships are probably mediated by other environmental gradients, sheep grazing could have effects on increasing habitat heterogeneity, which could help to promote steppe bird richness and abundance (Velado-Alonso et al., 2020).

The associations found between the decline in both sheep numbers and farmland and shrub-steppe birds in Spain are in agreement with the well-recognized positive impact that extensive grazing has on the distribution and abundance of farmland and steppe birds (Douglas et al., 2017; Fuller, 1996; Haworth & Thompson, 1990; Traba, Sastre & Morales, 2013; Santangeli et al., 2019) (see Fuller, 1996; Evans et al., 2006) for a negative effect of both heavy or very low levels of grazing in steppe birds). For example, steppe birds are

adapted to open habitats, which are partly maintained by light or moderate grazing (Fuller, 1996). Grazing plays an important role by decreasing vegetation height and increasing spatial heterogeneity (Bugalho et al., 2011; Leal et al., 2019; Evans et al., 2006), which may increase the foraging efficiency of steppe birds by facilitating foraging in more open habitats (Fuller, 1996; Leal et al., 2019; Zbyryt, Sparks & Tryjanowski, 2020; Murray et al., 2016) and minimizing their predation risk while foraging on the ground (Buckingham & Peach, 2005; Whittingham & Evans, 2004). Sheep grazing has been described as a main driver of these habitats, as it may favor open, creeping plant phenotypes, which disperse their seeds through the livestock dung (Malo & Suárez, 1995; González-Rebollar & Ruiz-Mirazo, 2013). The amount of plants consumed by sheep in Mediterranean shrub steppes under extensive grazing has been estimated at around 1,500 kg/ha/year (dry weight), while dung production is around 600 kg/ha/year (Le Hourèou, 1991). Moreover, sheep dung is twice as attractive to Mediterranean dung beetles as cattle dung, three times more than red deer pellets, and four times more than horse dung (Dormont et al., 2007). Thus, extensive sheep grazing will reduce plant structure and density (Adler, Raff & Lauenroth, 2001) and facilitate the occurrence of dung-processing arthropods (Prather & Kaspari, 2019), an important food source for some farmland and steppe birds (Vickery et al., 2001; Dennis et al., 2008; Jay-Robert et al., 2008). Finally, these areas might be preferred by birds feeding on insects and especially on dung beetles and other coprophagous arthropods (Leal et al., 2019; Jay-Robert et al., 2008; Gómez-Catasús et al., 2019; Smith et al., 2020).

Goat abundance was around six times lower than sheep abundance in Spain in 2018, and especially linked to specific regions. Thus, in this work we only considered trends in sheep numbers. Nonetheless, the negative relationship between Iberian steppe birds and sheep numbers was maintained when the trends in goat numbers in Spain were included, which was around -10% during the period 2002–2018 according to official data (MAPA, 2020). Besides sheep number decline, during the last decades there has also been a great change in livestock husbandry towards higher intensification (Riedel, Casasús & Bernués, 2007; Bernhard, 2008) that may alter the role that sheep play in natural steppes and other plant communities. For example, in recent times sheep are more often kept indoors than free-ranging (Martínez-Valderrama et al., 2021) and the proportion of pregnant and lactating ewes, which are usually kept indoors, has increased (Martínez-Valderrama et al., 2021; Riedel, Casasús & Bernués, 2007; Naylor et al., 2005). These changes in sheep husbandry imply a substantial reduction in the workload and may maximize farmers' income, but reduce the intensity and duration of outdoor grazing (Soto et al., 2016; Mújica et al., 2015).

Under these circumstances, extensive sheep grazing in Spain has practically disappeared (Martínez-Valderrama et al., 2021), and the role of sheep as an ecosystem engineer species has disappeared (Naylor et al., 2005) or has even become negative due to pollution or overgrazing and soil degradation around intensive farms (O'Brien et al., 2016; Pulido et al., 2018). In Spain, shrub- and tree-encroachment has steeply increased in areas considered as

transitional from former extensive grazing to current land abandonment (*Martínez-Valderrama et al., 2021*), and the consequences of this functional change are yet to be fully determined.

Our results suggest that the alarming decline in Iberian farmland and shrub-steppe birds might be, at least partly, associated with the sharp decline in sheep livestock occurring in Iberia. Further research is needed to understand the role of sheep grazing in steppe and farmland bird declines. In the meantime, there is a need to reverse the negative effect that this trend in sheep numbers may have on habitat heterogeneity and quality (*i.e.* food resources) for steppe birds. Our findings suggest that extensive grazing should be considered as a key factor in future Common Agricultural Policy reforms and conservation programs to stabilize, and even better increase, the traditional sheep farming system. Such changes in CAP policies to promote sheep farming may partially tackle other concerns in the EU, such as the preservation of biodiversity, reduction of risks due to wildfire, valorization of environmentally friendly agricultural practices and prevention of desertification (*Mújica et al., 2015*). Besides, changes from sheep to cattle, which could be seen as an opportunity for farmers due to lower production and maintenance costs, has to be deeply evaluated, as may provoke relevant changes in ecosystem structure and functionality (*Ramos et al., 2021; Faria & Morales, 2020; Mújica et al., 2015*). As the same grazing pressure may be favorable for some species of conservation concern but detrimental to others (*Leal et al., 2019; Reino et al., 2010*), we encourage researchers to estimate the adequate grazing intensity for protecting most farmland and steppe birds, as well as to perform species-specific studies for proposing a concrete grazing intensity to protect the most threatened species. These studies would provide scientific evidence to managers and therefore increase the implementation of extensive grazing as a conservation intervention (*Gibbons, Wilson & Green, 2011*). Extensive sheep grazing should be promoted as a multirole and low impact practice, which may contribute to increasing habitat heterogeneity, reversing shrub encroachment and improving the situation for birds while avoiding the need to apply other resource-consuming, and potentially hazardous practices such as mowing, manual shrub-clearing and/or controlled burning (*Carboni et al., 2015*). In a climate change scenario, natural steppe habitats may need habitat management actions aimed at improving habitat quality for open-habitat bird species (*Leal et al., 2019; Gibbons, Wilson & Green, 2011; Menz, Brotons & Arlettaz, 2009; Pérez-Granados, Serrano-Davies & Nogueras, 2018; Hellicar & Kirschel, 2020*).

ACKNOWLEDGEMENTS

We wish to thank SEO/Birdlife and specifically Juan Carlos del Moral and Virginia Escandell for providing SACRE data. We are grateful to Piotr Tryjanowski, Francisco Moreira, Rita Ramos and one anonymous reviewer whose comments helped to improve the manuscript.

ADDITIONAL INFORMATION AND DECLARATIONS

Funding

This is a contribution to the Excellence Network Remedial 3CM (S2013/MAE2719), supported by Comunidad de Madrid and to the LIFE Ricotí (ES-LIFE15/NAT/ES/000802) and LIFE Connect Ricotí (ES-LIFE20/NAT/ES/000133) projects, supported by the European Commission. The funders had no role in study design, data collection and analysis, decision to publish, or preparation of the manuscript.

Grant Disclosures

The following grant information was disclosed by the authors:

Excellence Network Remedial 3CM: S2013/MAE2719.

LIFE Ricotí: ES-LIFE15/NAT/ES/000802.

LIFE Connect Ricotí: ES-LIFE20/NAT/ES/000133.

European Commission.

Competing Interests

The authors declare that they have no competing interests.

Author Contributions

- Juan Traba conceived and designed the experiments, performed the experiments, analyzed the data, prepared figures and/or tables, authored or reviewed drafts of the paper, and approved the final draft.
- Cristian Pérez-Granados performed the experiments, analyzed the data, prepared figures and/or tables, authored or reviewed drafts of the paper, and approved the final draft.

Data Availability

The following information was supplied regarding data availability:

The raw data is available in the [Supplemental File](#).

Supplemental Information

Supplemental information for this article can be found online at <http://dx.doi.org/10.7717/peerj.12870#supplemental-information>.

REFERENCES

- Adler P, Raff D, Lauenroth W. 2001.** The effect of grazing on the spatial heterogeneity of vegetation. *Oecologia* **128**(4):465–479 DOI [10.1007/s004420100737](https://doi.org/10.1007/s004420100737).
- Bates D, Mächler M, Bolker B, Walker S. 2015.** Fitting linear mixed-effects models using lme4. *Journal of Statistical Software* **67**(1):1–48 DOI [10.18637/jss.v067.i01](https://doi.org/10.18637/jss.v067.i01).
- Bernhard P. 2008.** Analysis of Integrated and Cointegrated Time Series with R. 2nd Edition. New York: Springer-Verlag, 188.
- Bertolozzi CD. 2021.** Resilience of extensive sheep farming systems in Spain: strategies and policy assessment. Doctoral dissertation, Agronomica. Universidad Politécnica de Madrid, Madrid, Spain.

- Boatman ND, Brickle NW, Hart JD, Milsom TP, Morris AJ, Murray AW, Murray KA, Robertson PA. 2004. Evidence for the indirect effects of pesticides on farmland birds. *Ibis* 146(Suppl. 2):131–143 DOI 10.1111/j.1474-919X.2004.00347.x.
- Bowler DE, Heldbjerg H, Fox AD, de Jong M, Böhning-Gaese K. 2019. Long-term declines of European insectivorous bird populations and potential causes. *Conservation Biology* 33:1120–1130 DOI 10.1111/cobi.13307.
- Buckingham DL, Peach WJ. 2005. The influence of livestock management on habitat quality for farmland birds. *Animal Science* 81:199–203 DOI 10.1079/ASC50700199.
- Bugalho MN, Lecomte X, Gonçalves M, Caldeira MC, Branco M. 2011. Establishing grazing and grazing-excluded patches increases plant and invertebrate diversity in a Mediterranean oak woodland. *Forest Ecology and Management* 261(11):2133–2139 DOI 10.1016/j.foreco.2011.03.009.
- Burfield I. 2005. The conservation status of steppic birds in Europe. In: Bota G, Morales MB, Mañosa S, Camprodon J, eds. *Ecology and Conservation of Steppe-land Birds*. Barcelona: Lynx Edicions & Centre Tecnològic Forestal de Catalunya, 119–139.
- Carboni M, Dengler J, MantiLla-Contreras J, Venn S, Török P. 2015. Conservation value, management and restoration of Europe's semi-natural open landscapes. *Hacquetia* 14(1):5–17 DOI 10.1515/hacq-2015-0017.
- Casas F, Gurarieb E, Faganb WF, Mainalib K, Santiagod R, Hervás I, Palacínf C, Morenoa E, Viñuelac J. 2020. Are trellis vineyards avoided? Examining how vineyard types affect the distribution of great bustards. *Agriculture, Ecosystems and Environment* 289:106734 DOI 10.1016/j.agee.2019.106734.
- Casas F, Mougeot F, Arroyo B, Morales MB, Hervás I, de la Morena ELG, Fagan WF, Viñuela J. 2019. Opposing population trajectories in two Bustard species: a long-term study in a protected area in Central Spain. *Bird Conservation International* 29:308–320 DOI 10.1017/S0959270918000254.
- Castillo LP, Rubio MTM, Puyalto EF, Sepúlveda WS. 2008. The diversity of sheep production systems in Aragón (Spain): characterisation and typification of meat sheep farms. *Spanish Journal of Agricultural Research* 4:497–507 DOI 10.5424/sjar/2008064-344.
- Chiron F, Chargé R, Julliard R, Jiguet F, Muratet A. 2014. Pesticide doses, landscape structure and their relative effects on farmland birds. *Agriculture, Ecosystems & Environment* 185:153–160 DOI 10.1016/j.agee.2013.12.013.
- De Arriba R, Barac M. 2018. Homo economicus and the shepherd: the traditional sheep farmer facing the modernization (or intensification) of European livestock. *Eastern European Countryside* 24:171–187 DOI 10.2478/eec-2018-0009.
- De Frutos A, Olea PP, Mateo-Tomás P. 2015. Responses of medium-and large-sized bird diversity to irrigation in dry cereal agroecosystems across spatial scales. *Agriculture, Ecosystems & Environment* 207:141–152 DOI 10.1016/j.agee.2015.04.009.
- De Rancourt M, Fois N, Lavín MP, Tchakerian E, Vallerand F. 2006. Mediterranean sheep and goats production: an uncertain future. *Small Ruminant Research* 62(3):167–179 DOI 10.1016/j.smallrumres.2005.08.012.
- Dennis P, Skartveit J, McCracken DI, Pakeman RJ, Beaton K, Kunaver A, Evans DM. 2008. The effects of livestock grazing on foliar arthropods associated with bird diet in upland grasslands of Scotland. *Journal of Applied Ecology* 45:279–287 DOI 10.1111/j.1365-2664.2007.01378.x.
- Dormont L, Rapior S, McKey DB, Lumaret JP. 2007. Influence of dung volatiles on the process of resource selection by coprophagous beetles. *Chemoecology* 17(1):23–30 DOI 10.1007/s00049-006-0355-7.

- Douglas DJT, Beresford A, Selvidge J, Garnett S, Buchanan GM, Gullett P, Grant MC. 2017.** Changes in upland bird abundances show associations with moorland management. *Bird Study* 64:242–254 DOI 10.1080/00063657.2017.1317326.
- Engle R, Granger CWJ. 1987.** Co-integration and error correction: representation, estimation and testing. *Econometrica* 35(2):251–276 DOI 10.2307/1913236.
- European Environment Agency. 2015.** State of nature in the EU. Results from reporting under the nature directives 2007–2012. European Environment Agency, Technical Report No. 2/2015.
- Eurostat. 2020.** Statistics explained. Available at http://ec.europa.eu/eurostat/statistics-explained/index.php/Main_Page (accessed 1 March 2020).
- Evans DM, Redpath SM, Evans SA, Elston DA, Gardner CJ, Dennis P, Pakeman RJ. 2006.** Low intensity, mixed livestock grazing improves the breeding abundance of a common insectivorous passerine. *Biology Letters* 2:636–638 DOI 10.1098/rsbl.2006.0543.
- Faria N, Morales MB. 2020.** Livestock species and grazing rotational patterns modulate grassland bird assemblages in Mediterranean drylands. *Agriculture, Ecosystems & Environment* 295:106893 DOI 10.1016/j.agee.2020.106893.
- Fuller RJ. 1996.** Relationships between grazing and birds with particular reference to sheep in the British uplands. BTO Research Report No. 164. Norfolk, United Kingdom.
- Fuller RJ, Gough SJ. 1999.** Changes in sheep numbers in Britain: implications for bird populations. *Biological Conservation* 91(1):73–89 DOI 10.1016/S0006-3207(99)00039-7.
- Gibbons DW, Wilson JD, Green RE. 2011.** Using conservation science to solve conservation problems. *Journal of Applied Ecology* 48:505–508 DOI 10.1111/j.1365-2664.2011.01997.x.
- González-Rebollar JL, Ruiz-Mirazo J. 2013.** El papel del ganado doméstico en la naturalización del monte mediterráneo. *Pastos* 43:7–12.
- Gordo O. 2018.** Are two days enough? Checking the accuracy of the survey protocols used in common bird monitoring schemes. *Ardeola* 65:41–52 DOI 10.13157/arla.65.1.2018.ra3.
- Granger CWJ, Newbold P. 1974.** Spurious regressions in econometrics. *Journal of Econometrics* 2(2):111–120 DOI 10.1016/0304-4076(74)90034-7.
- Gómez-Catasús J, Garza V, Morales MB, Traba J. 2019.** Hierarchical habitat-use by an endangered steppe bird in fragmented landscapes is associated with large connected patches and high food availability. *Scientific Report* 9:19010 DOI 10.1038/s41598-019-55467-2.
- Gómez-Catasús J, Garza V, Traba J. 2018.** Wind farms affect the occurrence, abundance and population trends of small passerine birds: the case of the Dupont's lark. *Journal of Applied Ecology* 55(4):2033–2042 DOI 10.1111/1365-2664.13107.
- Gómez-Catasús J, Pérez-Granados C, Barrero A, Bota G, Giralt D, López-Iborra GM, Serrano D, Traba J. 2018.** European population trends and current conservation status of an endangered steppe-bird species: the Dupont's lark *Chersophilus duponti*. *PeerJ* 6:e5627 DOI 10.7717/peerj.5627.
- Haworth PF, Thompson DBA. 1990.** Factors associated with the breeding distribution of upland birds in the South Pennines, England. *Journal of Applied Ecology* 27:562–577.
- Hellicar MA, Kirschel AN. 2020.** Grazing pressure and the interaction dynamics of the endemic Cyprus Warbler *Sylvia melanothorax* and its recently colonising congener the Sardinian Warbler *S. melanocephala*. *Bird Conservation International* 31(2):239–254 DOI 10.1017/S0959270920000180.
- Herrando S, Anton M, Sardà-Palomera F, Bota G, Gregory RD, Brotons L. 2014.** Indicators of the impact of land use changes using large-scale bird surveys: land abandonment in a Mediterranean region. *Ecological Indicators* 45(1):235–244 DOI 10.1016/j.ecolind.2014.04.011.

- Holmes AL, Maestas JD, Naugle DE. 2017. Bird responses to removal of Western Juniper in sagebrush-steppe. *Rangeland Ecology & Management* 70(1):87–94 DOI 10.1016/j.rama.2016.10.006.
- Jay-Robert P, Niogret J, Errouissi F, Labarussias M, Paoletti É, Luis MV, Lumaret J-P. 2008. Relative efficiency of extensive grazing vs. wild ungulates management for dung beetle conservation in a heterogeneous landscape from Southern Europe (Scarabaeinae, Aphodiinae, Geotrupinae). *Biological Conservation* 141:2879–2887 DOI 10.1016/j.biocon.2008.09.001.
- Johansen S. 1988. Statistical analysis of cointegration vectors. *Journal of Economic Dynamics and Control* 112(2–3):231–254 DOI 10.1016/0165-1889(88)90041-3.
- Koch B, Edwards PJ, Blanckenhorn WU, Walter T, Hofer G. 2015. Shrub encroachment affects the diversity of plants, butterflies, and grasshoppers on two Swiss subalpine pastures. *Arctic, Antarctic, and Alpine Research* 47(2):345–357 DOI 10.1657/AAAR0013-093.
- Kurek P, Steppa R, Grzywaczewski G, Tryjanowski P. 2016. The silence of the lambs? Plant diversity in abandoned sheep pens. *Plant, Soil and Environment* 62(1):1–8 DOI 10.17221/327/2015-PSE.
- Le Hourèou HN. 1991. Impact of man and his animals on Mediterranean vegetation. In: Di Castri F, Goodall DW, Spetch RL, eds. *Ecosystems of the World*. Vol. 11. New York: Elsevier, 643.
- Leal AI, Acácio M, Meyer CF, Rainho A, Palmeirim JM. 2019. Grazing improves habitat suitability for many ground foraging birds in Mediterranean wooded grasslands. *Agriculture, Ecosystems & Environment* 270:1–8 DOI 10.1016/j.agee.2018.10.012.
- Malo JE, Suárez F. 1995. Herbivorous mammals as seed dispersers in a Mediterranean dehesa. *Oecologia* 104(2):246–255 DOI 10.1007/BF00328589.
- MAPA. 2020. Statistics. Available at <https://www.mapa.gob.es/es/estadistica/temas/estadisticas-agrarias/ganaderia/encuestas-ganaderas/> (accessed 20 September 2021).
- Martínez-Valderrama J, Sanjuán ME, del Barrio G, Guirado E, Ruiz A, Maestre FT. 2021. Mediterranean landscape re-greening at the expense of South American agricultural expansion. *Land* 10:204 DOI 10.3390/land10020204.
- McMahon BJ, Giralt D, Raurell M, Brotons L, Bota G. 2010. Identifying set-aside features for bird conservation and management in northeast Iberian pseudo-steppes. *Bird Study* 57(3):289–300 DOI 10.1080/00063651003749680.
- Menz MH, Brotons L, Arlettaz R. 2009. Habitat selection by Ortolan Buntings *Emberiza hortulana* in post-fire succession in Catalonia: implications for the conservation of farmland populations. *Ibis* 151:752–761 DOI 10.1111/j.1474-919X.2009.00961.x.
- Murray C, Minderman J, Allison J, Calladine J. 2016. Vegetation structure influences foraging decisions in a declining grassland bird: the importance of fine-scale habitat and grazing regime. *Bird Study* 63(2):223–232 DOI 10.1080/00063657.2016.1180342.
- Mújica PMT, Aguilar C, Vera RR, Capote CB, Rivas J, Martínez AG. 2015. Changes in the pastoral sheep systems of semi-arid Mediterranean areas: association with common agricultural policy reform and implications for sustainability. *Spanish Journal of Agricultural Research* 13(2):19 DOI 10.5424/sjar/2015132-6984.
- Naylor R, Steinfeld H, Falcon W, Galloway J, Smil V, Bradford E, Alder J, Mooney H. 2005. Losing the links between livestock and land. *Science* 310:1621–1622 DOI 10.1126/science.1117856.
- Newton I. 2004. The recent declines of farmland bird populations in Britain: an appraisal of causal factors and conservation actions. *Ibis* 146(4):579–600 DOI 10.1111/j.1474-919X.2004.00375.x.

- O'Brien D, Bohan A, McHugh N, Shalloo L. 2016. A life cycle assessment of the effect of intensification on the environmental impacts and resource use of grass-based sheep farming. *Agricultural Systems* 148:95–104 DOI 10.1016/j.agsy.2016.07.004.
- Pannekoek J, Strien A. 2006. *TRIM version 3.54 (trends and indices for monitoring data)*. Voorburg: Statistics Netherlands.
- Perrin W, Jay-Robert P, Buatois B, Tatin L. 2019. A comparative analysis of dung beetle assemblages (Coleoptera: Scarabaeidae: Scarabaeinae, Aphodiinae) attracted to sheep and little bustard excrement in Southern France. *The Coleopterists Bulletin* 73(1):185–192 DOI 10.1649/0010-065X-73.1.185.
- Pollock ML, Holland JP, Morgan-Davies C, Morgan-Davies J, Waterhouse A. 2013. Reduced sheep grazing and biodiversity: a novel approach to selecting and measuring biodiversity indicators. *Rangeland Ecology & Management* 66:387–400 DOI 10.2111/REM-D-11-00123.1.
- Prather RM, Kaspari M. 2019. Plants regulate grassland arthropod communities through biomass, quality, and habitat heterogeneity. *Ecosphere* 10(10):e02909 DOI 10.1002/ecs2.2909.
- Pulido M, Schnabel S, Lavado Contador JF, Lozano-Parra J, González F. 2018. The impact of heavy grazing on soil quality and pasture production in rangelands of SW Spain. *Land Degradation & Development* 29:219–230 DOI 10.1002/ldr.2501.
- Pérez-Granados C, López-Iborra GM. 2017. Assessment of counting methods for estimating the endangered Dupont's lark *Chersophilus duponti*. *Ardeola* 64:75–84 DOI 10.13157/arla.64.1.2017.sc2.
- Pérez-Granados C, López-Iborra GM. 2021. Research-conservation gap: an endangered bird case study. *Oryx* 10:1–8 DOI 10.1017/S003060532000023X.
- Pérez-Granados C, Serrano-Davies E, Noguerales V. 2018. Returning home after fire: how fire may help us manage the persistence of scrub-steppe specialist bird populations. *Biodiversity Conservation* 27:3087–3102 DOI 10.1007/s10531-018-1586-y.
- R Development Core Team. 2019. *R: a language and environment for statistical computing*. Vienna: The R Foundation for Statistical Computing. Available at <http://www.R-project.org>.
- Ramos RF, Diogo JA, Santana J, Silva JP, Reino L, Schindler S, Beja P, Lomba A, Moreira F. 2021. Impacts of sheep versus cattle livestock systems on birds of Mediterranean grasslands. *Scientific Reports* 11(1):10827 DOI 10.1038/s41598-021-89975-x.
- Ramos ME, Robles AB, González-Rebollar JL. 2010. Ley-farming and seed dispersal by sheep: two methods for improving fallow pastures in semiarid Mediterranean environments? *Agriculture, Ecosystems & Environment* 137(1–2):124–132 DOI 10.1016/j.agee.2010.01.012.
- Reino L, Porto M, Morgado R, Moreira F, Fabião A, Santana J, Delgado A, Gordinho L, Cal J, Beja P. 2010. Effects of changed grazing regimes and habitat fragmentation on Mediterranean grassland birds. *Agriculture, Ecosystems & Environment* 138(1–2):27–34 DOI 10.1016/j.agee.2010.03.013.
- Reverter M, Gómez-Catasús J, Barrero A, Traba J. 2021. Crops modify habitat quality beyond their limits. *Agriculture, Ecosystems & Environment* 319:107542 DOI 10.1016/j.agee.2021.107542.
- Riedel JL, Casasús I, Bernués A. 2007. Sheep farming intensification and utilization of natural resources in a Mediterranean pastoral agro-ecosystem. *Livestock Science* 111:153–163 DOI 10.1016/j.livsci.2006.12.013.
- Robinson RA, Sutherland WJ. 2002. Post-war changes in arable farming and biodiversity in Great Britain. *Journal of Applied Ecology* 39(1):157–176 DOI 10.1046/j.1365-2664.2002.00695.x.

- Sainz Ollero H. 2013.** Steppes across the world: an overview with emphasis on the Iberian Peninsula. In: Morales MB, Traba J, eds. *Steppe Ecosystems: Biological Diversity, Management and Restoration*. New York: Nova Science Publishers, 1–26.
- Santangeli A, Di Minin E, Arroyo B. 2014.** Bridging the research implementation gap. Identifying cost-effective protection measures for Montagu's harrier nests in Spanish farmlands. *Biological Conservation* **177**:126–133 DOI [10.1016/j.biocon.2014.06.022](https://doi.org/10.1016/j.biocon.2014.06.022).
- Santangeli A, Lehikoinen A, Lindholm T, Herzon I. 2019.** Organic animal farms increase farmland bird abundance in the boreal region. *PLOS ONE* **14**:e0216009 DOI [10.1371/journal.pone.0216009](https://doi.org/10.1371/journal.pone.0216009).
- Santos T, Suárez F. 2005.** Biogeography and population trends of Iberian steppe birds. In: Bota G, Morales MB, Mañosa S, Camprodon J, eds. *Ecology and Conservation of Steppe-land Birds*. Barcelona: Lynx Edicions & Centre Tecnològic Forestal de Catalunya, 69–102.
- SEO/BirdLife. 2016.** *Programas de seguimiento y grupos de trabajo de*. Madrid: SEO/BirdLife.
- Smith BM, Aebischer NJ, Ewald J, Moreby S, Potter C, Holland JM. 2020.** The potential of arable weeds to reverse invertebrate declines and associated ecosystem services in cereal crops. *Frontiers in Sustainable Food Systems* **3**:118 DOI [10.3389/fsufs.2019.00118](https://doi.org/10.3389/fsufs.2019.00118).
- Soto D, Infante-Amate J, Guzmán GI, Cid A, Aguilera E, García R, de Molina MG. 2016.** The social metabolism of biomass in Spain, 1900–2008: from food to feed-oriented changes in the agro-ecosystems. *Ecological Economics* **128**:130–138 DOI [10.1016/j.ecolecon.2016.04.017](https://doi.org/10.1016/j.ecolecon.2016.04.017).
- Spiegel A, Slijper T, De Mey Y, Poortvliet M, Rommel J, Hansson H, Vigani M, Soriano B, Wauters E, Appel F, Antonioli F, Harizanova H, Gavrilescu C, Gradziuk P, Neumeister D, Meuwissen M. 2019.** Report on farmers' perceptions of risk and resilience capacities—a comparison across EU farmers. SURE-Farm Deliverable 2.1 (H2020, No.727520).
- Stjernman M, Green M, Lindström Å, Olsson O, Ottvall R, Smith HG. 2013.** Habitat-specific bird trends and their effect on the Farmland Bird Index. *Ecological Indicators* **24**:382–391 DOI [10.1016/j.ecolind.2012.07.016](https://doi.org/10.1016/j.ecolind.2012.07.016).
- Suárez F. 1994.** *Mediterranean steppe conservation: a background for the development of a future strategy*. Brussels: European Commission.
- Suárez F. 2010.** *La alondra ricotí Chersophilus duponti*. Madrid: Dirección General para la Biodiversidad. Ministerio de Medio Ambiente y Medios Rural y Marino.
- Tella JL, Vögeli M, Serrano D, Carrete M. 2005.** Current status of the threatened Dupont's lark *Chersophilus duponti* in Spain: overestimation, decline, and extinction of local populations. *Oryx* **39**:90–94 DOI [10.1017/S0030605305000165](https://doi.org/10.1017/S0030605305000165).
- Traba J, de la Morena ELG, Morales MB, Suárez F. 2007.** Determining high value areas for steppe birds in Spain: hot spots, complementarity and the efficiency of protected areas. *Biodiversity Conservation* **16**:3255–3275 DOI [10.1007/978-1-4020-6865-2_2](https://doi.org/10.1007/978-1-4020-6865-2_2).
- Traba J, Morales MB. 2019.** The decline of farmland birds in Spain is strongly associated to the loss of fallowland. *Scientific Reports* **9**(1):9473 DOI [10.1038/s41598-019-45854-0](https://doi.org/10.1038/s41598-019-45854-0).
- Traba J, Sastre P, Morales MB. 2013.** Factors determining species richness and community composition in steppe birds of peninsular Spain. In: Morales MB, Traba J, eds. *Steppe Ecosystems: Biological Diversity, Management and Restoration*. New York: NOVA Science Publishers, 29–46.
- Varga A, Molnár Z. 2014.** The role of traditional ecological knowledge in managing wood-pastures. In: Hartel T, Plieninger T, eds. *European Wood-Pastures in Transition*. London, New York: Routledge, 187–202.
- Vargaa A, Molnára Z, Biróa M, Demeterb L, Gellényc K, Miókovicsd E, Molnáre Á, Molnárd K, Ujházyf N, Ulicsnic V, Babaig D. 2016.** Changing year-round habitat use of extensively grazing

cattle, sheep and pigs in East-Central Europe between 1940 and 2014: consequences for conservation and policy. *Agriculture, Ecosystems and Environment* **234**:142–153
[DOI 10.1016/j.agee.2016.05.018](https://doi.org/10.1016/j.agee.2016.05.018).

Velado-Alonso E, Morales-Castilla I, Rebollo S, Gómez-Sal A. 2020. Relationships between the distribution of wildlife and livestock diversity. *Diversity and Distributions* **26(10)**:1264–1275
[DOI 10.1111/ddi.13133](https://doi.org/10.1111/ddi.13133).

Vickery JA, Tallowin JR, Feber RE, Asteraki EJ, Atkinson PW, Fuller RJ, Brown VK. 2001. The management of lowland neutral grasslands in Britain: effects of agricultural practices on birds and their food resources. *Journal of Applied Ecology* **38**:647–664
[DOI 10.1046/j.1365-2664.2001.00626.x](https://doi.org/10.1046/j.1365-2664.2001.00626.x).

Whittingham MJ, Evans KL. 2004. The effects of habitat structure on predation risk of birds in agricultural landscapes. *Ibis* **146**:210–220 [DOI 10.1111/j.1474-919X.2004.00370.x](https://doi.org/10.1111/j.1474-919X.2004.00370.x).

Zbyryt A, Sparks TH, Tryjanowski P. 2020. Foraging efficiency of white stork *Ciconia ciconia* significantly increases in pastures containing cows. *Acta Oecologica* **104**:103544
[DOI 10.1016/j.actao.2020.103544](https://doi.org/10.1016/j.actao.2020.103544).