

# Photolysis of fluometuron in the presence of natural water constituents

Sabrina Halladja, Amina Amine-Khodja, Alexandra Ter Halle, Abdelaziz Boulkamh, Claire Richard

## ▶ To cite this version:

Sabrina Halladja, Amina Amine-Khodja, Alexandra Ter Halle, Abdelaziz Boulkamh, Claire Richard. Photolysis of fluometuron in the presence of natural water constituents. Chemosphere, Elsevier, 2007, 69, pp.1647-1654. <10.1016/j.chemosphere.2007.05.035>. <halsde-00180209>

HAL Id: halsde-00180209 https://hal.archives-ouvertes.fr/halsde-00180209

Submitted on 18 Oct 2007

**HAL** is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers. L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

| 1  | Photolysis of fluometuron in the presence of natural water constituents  |
|----|--|
| 2  |  |
| 3  | Sabrina Halladja <sup>1,2</sup> , Amina Amine-Khodja <sup>1,2</sup> , Alexandra ter Halle <sup>1</sup> , Abdelaziz Boulkamh <sup>2</sup> and |
| 4  | Claire Richard <sup>1*</sup>   |
| 5  |  |
| 6  | <sup>1</sup> Laboratoire de Photochimie Moléculaire et Macromoléculaire, UMR 6505 CNRS-Université  |
| 7  | Blaise Pascal, Ensemble Universitaire des Cézeaux, F-63177 Aubière CEDEX, France   |
| 8  | <sup>2</sup> Laboratoire des Sciences et Technologies de l'Environnement, Université Mentouri, 25000   |
| 9  | Constantine, Algérie   |
| 10 |  |
| 11 | * corresponding author: Claire Richard   |
| 12 | E mail: claire.richard@univ-bpclermont.fr;   |
| 13 | tel +33 (0)4 73 40 71 42;  |
| 14 | fax: +33 (0)4 73 40 77 00  |
| 15 |  |
| 16 |  |

Abstract. Phototransformation of the herbicide fluometuron (1 μM) in natural sunlight was investigated in neutral Milli-Q water and in synthetic waters containing either fulvic acids, nitrate ions or both in order to mimic reactions taking place in aquatic environments. Fluometuron degradation followed a pseudo-first order kinetics. The reaction was faster in synthetic than in Milli-Q water. Fulvic acids (10 mg L<sup>-1</sup>) increased the rate of fluometuron photolysis by a factor 2.5 and nitrates (25 mg L<sup>-1</sup>) by a factor 15. Identification of major photoproducts was conducted under laboratory conditions using LC-ESI-MS. Numerous photoproducts were detected and tentatively characterized. In the presence of nitrates, hydroxylation of the aromatic ring with or without hydrolysis of CF<sub>3</sub> into CO<sub>2</sub>H and oxidation of the urea chain leading to demethylation were observed. In the presence of fulvic acids, hydroxylation of the aromatic ring was the major reaction route.

Key-words: phenylurea herbicide, photosensitization, nitrate ions, fulvic acids, hydroxyl radicals

#### Introduction.

Photochemistry is one of the main route for organic pollutants attenuation in surface waters. This abiotic degradation pathway has received increasing interest in the last thirty years (Zepp and Cline, 1977; Kotzias and Korte, 1981; Durand et al, 1991; Lemaire et al, 1991; Pirisi et al, 1996; Svenson and Hynning, 1997; Mansour et al, 1999). To predict the fate of pollutants in the natural environment and to assess the risk they may pose, it is necessary to improve our knowledge on their chemical reactions.

Several types of reactions may occur depending on the medium composition. Direct photolysis is possible if the considered pollutant absorbs solar light. In addition, photoinduced or photosensitized transformations mediated by components of the aquatic medium can also

take place. In particular, dissolved natural organic matter (DOM) which absorbs a large portion of photons is a potential photosensitizer. Singlet oxygen, superoxide ion/hydroperoxyl radicals, hydroxyl radicals, excited triplet states and alkylperoxyl radicals were proved or proposed to be generated in natural waters under the influence of sunlight (Zafiriou and True, 1979; Zepp et al, 1981; Hoigné et al, 1989; Haag and Hoigné, 1986; Cooper et al, 1989; Vaughan and Blough, 1998; Canonica et al, 1995). However, a part of these species are trapped by DOM itself. Nitrate ions that are present in surface waters at level varying from 0.2 to 25 mg L<sup>-1</sup> generate the highly oxidizing hydroxyl radicals under light excitation (Boule et al, 1999 for a review). Hydroxyl radicals that are unspecific oxidants are scavenged by a variety of aquatic components. The percentage of hydroxyl radicals trapped by pollutants is thus strongly dependent on the medium composition. Due to all these factors, studies can help to predict the fate of pollutants at the condition that they are undertaken under relevant experimental conditions.

In the present work, we focused on the phototransformation of the phenylurea herbicide fluometuron (FM). This compound is widely used for pre- and post-emergence control of weeds in fields of conventional cotton cultivars. It is persistent (Stoeckel et al, 1997) and may pose some risks to aquatic organisms (Muschal and Warne, 2003). However, literature data are very scarce, especially concerning its photochemical reactivity. FM poorly absorbs solar light (see Figure 1) but indeed its direct photolysis in simulated solar light ( $\lambda$  > 290 nm) was reported (Lam et al, 2005). The CF<sub>3</sub> group undergoes photohydrolysis into CO<sub>2</sub>H. FM was also included in a recent study (Lam et al, 2003) in which the contribution of indirect processes in surface waters under sunlight simulated irradiation conditions was evaluated. Nitrate ions at a level of  $8x10^{-4}$  M significantly enhanced FM degradation while humic acids gave unclear results.

The objective of the present work was to investigate the photolysis of FM in conditions approaching real ones. FM in the micromolar range was irradiated in natural solar light. The influence of DOM or/and nitrates on the phototransformation was assessed. In parallel, laboratories experiments were conducted to identify the main photoproducts on the basis of HPLC-ESI-MS analyses.

#### **Material and Methods**

- **Chemicals.** FM (97.7 % purity) was purchased from Riedel de Haën (Saint Quentin Fallavier,
- 73 France) and used as received. Elliott soil fulvic acids were purchased from International
- Humic Substances Society (IHSS). The salt used to add nitrate ions was NaNO<sub>3</sub> (Fluka, 99
- 75 %). All other reactants were of the highest grade available and used as received. Water was
- 76 purified using a Milli-Q device (Millipore).

**Preparation of solutions.** A stock solution of FM was prepared by dissolving 4 mg of substrate in 100 ml of Milli-Q purified water. Solubilization was achieved after a 4-day stirring. Stock solutions of fulvic acids and sodium nitrate were prepared in Milli-Q purified water. They were at a concentration of 40 mg L<sup>-1</sup> and 50 mg L<sup>-1</sup>, respectively. Stock solutions were then diluted in order to obtain the desired concentrations. Solutions were buffered in order to maintain constant the pH all along irradiations. A pH of 6.5 was chosen because natural aquatic medium is generally around neutrality. Phosphate buffers were used; their final concentration was equal to 10<sup>-3</sup> M. Deoxygenation of solutions was achieved by nitrogen bubbling for 20 minutes prior to irradiation and during irradiations. Solutions were

Photoreaction setup. For kinetic purpose, FM (1 μM) was irradiated (i) in MilliQ water, (ii) in water containing fulvic acids (10 mg L<sup>-1</sup>) used as a surrogate of DOM, (iii) in water

containing nitrate ions (25 mg L<sup>-1</sup> or 3x10<sup>-4</sup> M) and (iv) in water containing both fulvic acids (10 mg L<sup>-1</sup>) and nitrate ions (3x10<sup>-4</sup> M). Irradiation experiments in natural sunlight were performed at Clermont-Ferrand (46° N, 3° E) in June 2006. Cylindrical quartz glass reactors (14-mm internal diameter) were filled with 14 ml of solutions. A headspace was left above solutions. Reactors were closed by a septum, attached on a rack inclined by about 15° from horizontal and exposed to solar light. Samples received 13 h of sunshine per day. Aliquots of 0.5 mL were removed simultaneously from all the solutions at selected intervals. At each sampling, solutions were re-saturated with air. Samples were immediately analyzed by HPLC. Irradiations started at 9 am on the first day. The first samplings were made at 5 pm. Then, the next samplings were made at 9 am and 5 pm each day. Solutions were left outside overnights. Dark control experiments showed no loss of FM by adsorption on glass surfaces and no transformation, during the time required for the irradiations, irrespective of the aqueous media (variations less than 3 %).

Irradiations under laboratory conditions were also carried out for photoproducts identification. Irradiations were performed in a device equipped with six TLAD 15W/05 fluorescent tubes emitting within the wavelength range 300-450 nm with a maximum of emission at 365 nm and in a Pyrex glass reactor (14 mm i.d). The device was cylindrical and equipped with reflecting inner walls. A ventilator was used as a cooling system. The reactor was placed in the centre of the device and was surrounded by the six fluorescent tubes. Light intensity was measured using *p*-nitroanisole/pyridine as a chemical actinometer (Dulin and Mill, 1982). PNA (10<sup>-5</sup> M) and pyridine (10<sup>-4</sup> M) were irradiated in the same conditions as samples in the polychromatic device and in solar light. In the polychromatic device, PNA loss was 12 % after 1 hour and 63 % after 5 h. In solar light, 56 % of PNA had disappeared after 4 h of exposure between 10 am and 2 pm. It can be deduced that the average light intensity

delivered by the tubes of the polychromatic device was of the same magnitude order as that of solar light within the wavelength range 300-400 nm.

117

118

119

120

121

122

123

124

125

126

127

128

129

130

131

132

133

134

135

136

137

138

139

115

116

Analytical procedures. Loss of fluometuron and formation of photoproducts were monitored by HPLC-UV using a Waters apparatus equipped with two pumps (model 510), an autosampler, a photodiode array detector (model 996), a detector W2487 and a C<sub>18</sub> reversephase column (4.6 mm x 250 mm, Spherisorb S5 ODS2, 5 µm, Waters). Eluent was a mixture of water acidified with 0.1% of orthophosphoric acid and methanol (50 %/50 %) delivered at a constant flow of 1 mL min<sup>-1</sup>. For the HPLC mass spectrometry analyses, samples containing fulvic acids were prepared as follows. Portions of 0.6 ml of irradiated solutions were passed over an Oasis BHL column conditioned beforehand with 1 ml of methanol and 1 ml of water acidified with formic acid 0.1 %. Fulvic macromolecules were recovered by passing 1 ml of acidified water through the column while fluometuron and its photoproducts by passing 1 ml of methanol. The HPLC-UV-MS analyses were performed using a Waters/Micromass LC/QTOF (Micromass, Manchester, UK). For the HPLC conditions, a Waters Alliance 2695 HPLC equipped with a photodiode array detector (DAD) was used. A reversed-phase column (C18 Hypersil ODS, 5 µm, 100 mm x 2.1 mm; Interchim, Montluçon, France) was used at a flow rate of 0.3 mL min<sup>-1</sup>. The mobile phase was composed of acetonitrile (solvent A) and acidified water (formic acid, 0.4% v/v; pH 2.6) (solvent B). Gradient: 0-5 min, 5% A; 5-30 min, 5-95% A (linear); 30-34 min, 95% A; 34-35 min, 95-5% A; 35-40 min, 5% B (equilibrium period). The injection volume was 30 μL. The LC-ESI-MS worked both in positive and negative mode. The optimum voltages for the probe and ion source components (to produce maximum intensity) were 3 kV for the stainless-steel capillary, 35 V for the sample cone, and 2 V for the extractor cone. In the negative mode, voltages were switched, except the capillary voltage which was -2.1 kV for better sensitivity. UV spectra were

recorded on a Cary 3 (Varian) spectrophotometer. A 1-cm path quartz cell was used for all the experiments. The reference beam blank was always Milli-Q water.

142

143

144

145

146

147

148

149

150

151

152

153

154

155

156

157

158

159

160

161

162

163

164

authors.

140

141

#### **Results and Discussion**

## 1. Kinetics of phototransformation in solar light

We first compared the profiles of FM loss in various conditions. In the absence of fulvic acids (FA) and nitrate ions, the consumption of FM (1 µM) in solar light was very slow: less than 30 % had disappeared after 6.7 d. The addition of fulvic acids (10 mg L<sup>-1</sup>) significantly increased the rate of FM consumption: about 50 % of FM had disappeared after 6.7 d. A larger enhancement of the reaction rate was observed in the presence of nitrate ions (25 mg L<sup>-1</sup>): a complete FM loss was obtained after 3.3 d. In the presence of both fulvic acids (10 mg L<sup>-1</sup>) and nitrate ions (25 mg L<sup>-1</sup>), FM disappeared more slowly than in the presence of nitrates alone. To determine rate constants, we plotted  $lnc_0/c$  vs irradiation time, where  $c_0$  is the initial FM concentration and c the concentration at t, taking into account that samples were irradiated for 8 h between 9 am and 5 pm and for 5 h between 5 pm and 9 am (Figure 2). In all cases, FM consumption followed pseudo-first order kinetics. The rate coefficients, k, and R<sup>2</sup> values are reported in Table 1. These kinetic results bring insight into the photodegradability of FM in solar light. FM is hardly transformed in pure water due to the poor absorption of solar radiations. Using the rate coefficient given in Table 1 one computes a half-life of 175 h. This value is 2.6-fold higher than that of Lam et al. (2003). It shows that in the case of FM a one-day irradiation in June in our place is comparable to about 5 h of irradiation in the photosimulator used by these

The chromophoric constituents of water (nitrate and fulvic acids) promoted FM phototransformation. The effect of fulvic acids was quite moderate. In the presence of fulvic

acids (10 mg L<sup>-1</sup>), the rate coefficient was increased by a factor of 2.5 and the half-live reduced by the same factor. Nitrate ions (25 mg L<sup>-1</sup>) had a more pronounced influence increasing the rate coefficient and reducing the half-live by a factor of 15. In the presence of fulvic acids and nitrate ions, the rate of FM photodegradation was smaller by about 40 % than in the presence of nitrate ions alone. At the considered concentrations, the absorbance of nitrates is very small (around 0.002) while that of FA bigger (around 0.22) for a 1-cm path-length (see Figure 1). The inhibiting effect of FA on nitrate photolysis through screen effect is difficult to evaluate due to the cylindrical shape of reactors. Based on a mean path-length of 0.7 cm, one would compute a rate reduction between 15 and 20 %. It represents 50 % of the measured inhibition. The remaining 50 % of inhibition are likely to be due to the scavenging of hydroxyl radicals by FA (Brezonik and Fulkerson-Brekken, 1998; Schindelin and Frimmel, 2000, ter Halle and Richard, 2006).

In an attempt to delineate the role of hydroxyl radicals in the fulvic acids mediated phototransformation of FM, we studied the influence of 2-propanol added as a hydroxyl radical scavenger on the reaction (k = 1.9x10<sup>10</sup> M<sup>-1</sup> s<sup>-1</sup>, Buxton et al, 1988). These experiments were undertaken in laboratory conditions. In the absence of fulvic acids, the initial rate of FM phototransformation was equal to 2x10<sup>-8</sup> M h<sup>-1</sup>. In the presence of fulvic acids (5 mg L<sup>-1</sup>), it raised to 5.2x10<sup>-8</sup> M h<sup>-1</sup>. The addition of 2-propanol (0.015 M) reduced the latter rate by a factor comprised between 1.5 and 2, suggesting the involvement of hydroxyl radicals in the phototransformation reaction and showing the ability of FA to produce them under irradiation. We also studied the influence of oxygen on the reaction. After 5 h of irradiation, FM consumption was twice faster in air-saturated than in oxygen-saturated medium and 3-fold faster in nitrogen-saturated medium than in air-saturated medium. Thus, oxygen clearly inhibited the reaction.

## 2. Photoproducts identification

To achieve photoproducts characterization, we irradiated more highly concentrated FM solutions under laboratory conditions. The photolysis of FM ( $3x10^{-5}$  M) in Milli-Q water yielded the acidic compound resulting from the hydrolysis of CF<sub>3</sub> into CO<sub>2</sub>H as previously reported (Lam et al, 2005). Its UV absorption spectrum differs from that of FM: the far UV absorption band ( $\lambda_{max}$  = 224 nm) shows a shoulder and the second maximum of absorption is located at 297 nm instead of 275 nm in the case of FM.

The irradiation of FM ( $3x10^{-5}$  M) in the presence of fulvic acids ( $25 \text{ mg I}^{-1}$ ) yielded two HPLC-UV detectable photoproducts, I and II (Figure 3A). The UV spectrum of I exhibited the same shoulder as that of the acidic product and a second maximum red-shifted by 27 nm ( $\lambda_{max} = 297 \text{ nm}$ ). The UV spectrum of II resembled that of FM but the second maximum was red-shifted by 20 nm compared to it ( $\lambda_{max} = 295 \text{ nm}$ ). The HPLC-mass analysis of I gave a first molecular ion at m/z = 223 in ES mode and at m/z = 225 in ES mode corresponding to the loss of 9 amu that may be explained by the hydrolysis of CF<sub>3</sub> into CO<sub>2</sub>H and the addition of an oxygen atom. Fragments at m/z = 179 [M-44] and at m/z = 134 [M-89] were obtained in ES mode. These data are compatible with the structure proposed in Table 2. The first fragment would correspond to the loss of the carboxylic group and the second to the cleavage of the terminal amine group -N(CH<sub>3</sub>)<sub>2</sub> followed by loss of an H atom to yield an isocyanate as shown in Scheme 1.

Photoproduct II gave molecular ions at m/z = 247 and 249 in ES<sup>-</sup> and ES<sup>+</sup> modes, respectively, corresponding to the addition of an oxygen atom to FM. A fragment at m/z = 202 corresponding to M-45 was observed in ES<sup>-</sup> mode. The loss of 45 amu is likely to result from the departure of  $N(CH_3)_2$  by cleavage of the terminal urea C-N bond and the elimination of H to form the isocyanate. It is compatible with structure given in Table 2.

The irradiation of FM  $(3x10^{-5} \text{ M})$  in the presence of nitrate ions  $(25 \text{ mg I}^{-1})$  yielded seven photoproducts detectable by HPLC-UV (see Figure 3 B). Based on HPLC retention times, UV absorption spectra and mass data, we could conclude that photoproducts I and II were present among the seven photoproducts. Two other compounds III and IV with a molecular ion at m/z = 223 in ES<sup>-</sup> mode were also detected. As they showed distinct retention times and UV maxima, but similar fragmentation at m/z =179 and 134 with different percentages of fragments, they are likely to be isomers of I. Photoproducts I, III and IV should differ from each other in the position of the hydroxyl group on the ring. The photoproducts V and VI that were eluted just before FM showed similar absorption spectra. One of them, V, gave a molecular ion at m/z = 203 in ES<sup>-</sup> mode corresponding to the loss of 28 amu and a fragment at m/z = 160. The other one, VI, gave a molecular ion at m/z = 217 in ES<sup>-</sup> mode and the same fragment at m/z = 160 (see Scheme 2). These compounds could be assigned to demethylated products; VI would be the monodemethylated derivative and V the didemethylated derivative (see Table 2).

A molecular ion at m/z = 245 in ES mode was also detected at a retention time longer than that of FM. This mass corresponds to the addition of 14 amu. Three fragments were observed at m/z = 217 [M-28], 188 [M-28-29] and 160 [M-28-29-28]. This fragmentation indicates that the aromatic ring was not altered. It is in favour of photoproduct VII that bears a carbonyl group and for which successive losses of CO, NCH<sub>3</sub> and CO are again possible.

Finally, we used the selected peak method to detect some specific photoproducts formation of which was expected. In the case of nitrate ions, we could find nitro derivatives at m/z = 276 in ES<sup>-</sup> mode corresponding to the addition of 45 amu (+NO<sub>2</sub>-H). In the case of fulvic acids, careful examination of chromatograms revealed the presence of peaks at m/z = 217 and 245 in ES<sup>-</sup> mode and m/z = 219 in ES<sup>+</sup> mode, showing that fulvic acids

photosensitized transformation also yielded FM demethylation, but this pathway was very minor.

240

241

238

239

#### 3. Mechanisms of reactions

- FA and nitrates both accelerated the phototransformation of FM through different
- reaction mechanisms. Nitrate absorbs solar radiation in the actinic spectrum ( $\varepsilon_{304} = 7.4 \text{ M}^{-1}$
- 244 cm<sup>-1</sup>) and produces photoreactants through the two following primary processes (Boule et al,
- 245 1999):
- 246  $NO_3^- + hv \rightarrow NO_2^- + O$
- 247  $NO_3^- + h\nu \rightarrow NO_2^{\bullet} + O^{\bullet}$
- 248  $O^{-\bullet} + H^+ \leftrightarrow HO^{\bullet}$
- The formation of hydroxyl radicals is 9-fold more efficient than that of atomic oxygen.
- 250 As hydroxyl radicals are also much stronger oxidants than atomic oxygen, they are generally
- 251 considered to be the main species involved in the nitrate-photoinduced transformations of
- organic compounds, even though nitration reactions were also reported to occur (Machado
- and Boule, 1994; Vione et al., 2001). Hydroxylation of the aromatic ring and
- oxidation/elimination of the methyl groups of the urea aliphatic chain confirm that FM was
- oxidized by hydroxyl radicals (Scheme 3).
- The aromatic ring of FM is a site of attack for hydroxyl radicals. The OH-adduct radical is expected to yield the ring hydroxylated compound (photoproduct II). Photoproducts
- 258 I, III and IV are also ring hydroxylated products, but, contrary to II, they bear the CO<sub>2</sub>H
- substituent instead of the CF<sub>3</sub> substituent initially present in FM. The formation of these
- 260 products requires two steps. Either they are produced by photohydroxylation of the acid, or
- 261 they result from the direct photolysis of photoproduct II. Photoproduct II which shows a
- 262 maximum of absorption at 295 nm instead of 275 nm for FM absorbs solar light better than

FM, and thus could undergo faster photolysis. In both cases, several isomers corresponding to the possible sites of attack for hydroxyl radicals are expected. As three hydroxylated acids were detected, three hydroxylated derivatives of FM should be produced too. We only detected one peak at the mass m/z = 247 in ES<sup>-</sup>. Either, the three isomers show the same retention time and are eluted altogether, or we have failed to detect two of them.

The second site of attack of hydroxyl radicals is the urea aliphatic chain. Three photoproducts were detected: the carbonyl derivative VII, the monodemethylated product VI and the didemethylated product V. The first step should be the abstraction of a H atom from CH<sub>3</sub> to form the CH<sub>2</sub>• radical that can add oxygen and finally give V, VI and VII. On the HPLC chromatograms and by total ion count (TIC), the area of products V, VI and VII were higher than those of products I-IV. It seems therefore that the attack of the urea terminal chain is the main reaction pathway. However, without titration of photoproduct to measure chemical yields, this argument must be taken with caution because ring hydroxylation and hydrolysis of CF<sub>3</sub> significantly shift maxima of absorption to longer wavelengths and photoproducts I-IV might not accumulate due to photoreactivity. In the chromatograms of solar-light irradiated solutions of micromolar FM, only photoproducts VI and VII were detectable.

Fulvic acids are known to produce several photoreactants under light excitation (oxidant excited states, singlet oxygen, radicals). The quite small photosensitizing effect of fulvic acids on FM shows that FM is not easily oxidized by these species. A part of the reaction can be attributed to hydroxyl radicals as confirmed by the detection of product VI as a minor by-product and the inhibiting effect of 2-propanol on the reaction. The two main photoproducts were I and II. Contrary to what was observed in the nitrate-mediated phototransformation, only one hydroxylated acid derivative was produced in the presence of fulvic acids. This result suggests that the FA mediated ring hydroxylation does not involve hydroxyl radicals. As an alternative, the hydroxylation reaction could involve the triplet

excited state of FA ( ${}^{3}$ FA\*).  ${}^{3}$ FA\* reacts with FM either through energy transfer or through electron transfer. We found that the rate of FM phototransformation is accelerated upon deoxygenation of the medium by nitrogen bubbling as in the case of monuron (Richard et al, 1997). This result confirms the involvement of  ${}^{3}$ FA\* in FM photodegradation but does not clear up the mechanism. In the case where an energy transfer reaction would occur,  ${}^{3}$ FM\* would be produced. Oxidation photoproducts are formed. It implies that  ${}^{3}$ FM\* is oxidized. As direct photolysis of FM leads to acidification of CF3 substituent, one would conclude that singlet and triplet excited states of FM react differently. As an alternative, electron transfer from FM to  ${}^{3}$ FA\* could occur. It is the hypothesis proposed by Gerecke et al (2001). This hypothesis was also adopted in the case of fenuron, an analog of FM, because the FA mediated phototransformation of fenuron was found to be inhibited by removal of oxygen (Richard et al, 1997). Thus, the mechanism of reaction of FM is not clear. Further addition of dioxygen on the ring should finally produce the hydroxylated products, either photoproduct II in the case where the oxidation step involves FM or photoproduct I if it involves the acid, the photoproduct of direct photolysis. As proposed above, II might also produce I by photolysis.

# Conclusion

In conclusion, direct photolysis of aqueous FM is slow but photosensitizing processes may promote its transformation in surface waters. In the presence of hydroxyl radicals, both the lateral chain and the ring of FM are oxidized. Photoreactants deriving from fulvic acids privilege oxidation of the aromatic ring. Photoproducts of hydrolysis or hydroxylation of the aromatic ring absorb solar light better than FM itself. They are thus susceptible to undergo photolysis faster than FM. Photoproducts resulting from oxidation of lateral urea chain are more photostable. It would be now necessary to test the toxicity of each individual compound

312 or that of the irradiated mixtures toward reference organisms in order to access the risk that 313 may pose FM. 314 315 References 316 317 Boule, P., Bolte, M., Richard, C., 1999. Phototransformations induced in aquatic medium by 318 Fe<sup>III</sup>, NO<sub>3</sub>, NO<sub>2</sub> and humic substances. The Handbook of Environmental Chemistry. 319 Environmental Photochemistry, Editor P. Boule, Editor in Chief: O. Hutzinger, Bayreuth 320 (FRG), Part L, volume 2, 181-215. 321 322 Brezonik, P.L., Fulkerson-Brekken, J., 1998. Nitrate-induced photolysis in natural waters: 323 controls on concentrations of hydroxyl radical photo-intermediates by natural scavenging 324 agents. Environ. Sci. Technol. 32, 3004-3010. 325 Buxton, G.V., Greenstock, C.L., Helman, W.P., Ross A.B., 1988. Critical review of rate 326 327 constants for reactions of hydrated electrons, hydrogen atoms and hydroxyl radicals in aqueous solution. J. Phys. Chem. Ref. Data 17, 513-886. 328 329 330 Canonica, S., Jans, U., Stemmler, K., Hoigné, J., 1995. Transformation kinetics of phenols in 331 water: photosensitization by dissolved organic material and aromatic ketones. Environ. Sci. 332 Technol. 29, 1822-1831. 333 334 Cooper, W.J., Zika, R.G., Petasne, R.G., Fischer, A.M., 1989. Sunlight- induced 335 photochemistry of humic substances in natural waters: major reactive species. Adv. Chem. 336 Ser. 219, 333-362.

transformations of fluoxetine. Environ. Sci. Technol. 39, 513-522.

16 Lam, M.W., Tantuco, K., Mabury, S.A., 2003. PhotoFate: A new approach in accounting for 361 362 the contribution of indirect photolysis of pesticides and pharmaceuticals in surface waters. 363 Environ. Sci. Technol. 37, 899-907. 364 365 Lemaire, J., Boule, P., Sehili, T., Richard, C., 1991. Direct and indirect phototransformations 366 of chlorophenols and chlorobenzenes in laboratory conditions relevant to the aquatic 367 environment. In "Photochemical Conversion and Storage of Solar Energy" E. Pelizzetti, M. 368 Schiavello (eds), Kluwer Academic Publishers, Netherland, 477-495. 369 370 Machado, F., Boule, P., 1994. Phototransformation of phenolic derivatives in aqueous 371 solution induced by excitation of nitrate and nitrite ions. Trends Photochem. Photobiol. 3, 29-372 38. 373 374 Mansour, M., Feicht, E.A., Behechti, A., Schramm, K.W., Kettrup, A., 1999. Determination 375 photostability of selected agrochemicals in water and soil. Chemosphere 39, 575-585. 376 377 Muschal, M., Warne, M.S.J., 2003. Risk posed by pesticides to aquatic organisms in rivers of 378 northern inland New South Wales, Australia. Hum. Ecol. Risk Assess. 9, 1765-1787. 379 380 Pirisi, F.M., Cabras, P., Garau, V.L., Melis, M., Secchin E., 1996. Photodegradation of 381 pesticides. Photolysis rates and half-life of pirimicarb and its metabolites in reactions in water

and in solid phase. J. Agric. Food Chem. 44, 2417-2422.

Richard, C., Vialaton, D., Aguer, J.P., Andreux, F., 1997. Transformation of monuron

photosensitized by soil extracted humic substances: energy or hydrogen transfer mechanism?

J. Photochem. Photobiol. A 111, 265-271.

| 387 |  |
|-----|--|
| 388 | Schindelin, A.J., Frimmel, F.H., 2000. Nitrate and natural organic matter in aqueous solutions   |
| 389 | irradiated by simulated sunlight: Influence on the degradation of the pesticides dichlorprop     |
| 390 | and terbutylazine. Environ. Sci. Technol. 7, 205-210.  |
| 391 |  |
| 392 | Stoeckel, D.M., Mudd, E.C., Entry, J.A., 1997. Degradation of persistent herbicides in           |
| 393 | Riparian Wetlands. ACS Symp. Ser. 664, 114-132.  |
| 394 |  |
| 395 | Svenson, A., Hynning, P.A., 1997. Increased aquatic toxicity following photolytic conversion     |
| 396 | of an organochlorine pollutant. Chemosphere 34, 1685-1692.                                       |
| 397 |  |
| 398 | ter Halle, A., Richard, C., 2006. Simulated solar light irradiation of mesotrione in natural     |
| 399 | waters. Environ. Sci. Technol. 40, 3842-3847   |
| 400 |  |
| 401 | Vaughan, P., Blough, N., 1998. Photochemical formation of hydroxyl radical by constituents       |
| 402 | of natural waters. Environ. Sci. Technol. 32, 2947-2953.   |
| 403 |  |
| 404 | Vione, D., Maurino, V., Minero, C., Vincenti, M., Pelizzetti, E., 2001. Formation of             |
| 405 | nitrophenols upon UV irradiation of phenol and nitrate in aqueous solutions and in ${\rm TiO_2}$ |
| 406 | aqueous suspensions. Chemosphere 44, 237-248.  |
| 407 |  |
| 408 | Zafiriou, O.B., True, M.B. 1979. Nitrate photolysis in seawater by sunlight. Mar. Chem., 8       |
| 409 | 33-42.   |
| 410 |  |

| 411        | Zepp, R.G., Cline, D.M., 1977. Rates of direct photolysis in aquatic environment. Environ. |
|------------|--|
| 412        | Sci. Technol. 11, 359-366.   |
| 413        |  |
| 414        | Zepp, R.G., Baughman, G.L., Schlotzhauer, P.F., 1981. Comparison of photochemical          |
| 415        | behavior of various humic substances in water : Sunlight induced reactions of aquatic      |
| 416        | pollutants photosensitized by humic substances. Chemosphere 10, 109-117.                   |
| 417<br>418 |  |
| 410        |  |

| 419        | Captions for figures  |
|------------|---|
| 420        |   |
| 421        | Figure 1: Absorption spectrum of () FM at $3x10^{-4}$ M, (-) FA at $10 \text{ mg L}^{-1}$ , (***) nitrates a            |
| 422        | 0.01 M and solar light emission reaching the earth surface in summer from Zepp and Cline,                               |
| 423        | 1977.   |
| 424        |   |
| 425        | Figure 2 : Kinetics of FM (10 <sup>-6</sup> M) phototransformation in solar light in various aqueous                    |
| 426        | media. Plot of lnc <sub>0</sub> /c vs the irradiation time, where c and c <sub>0</sub> were the concentrations at t and |
| 427        | t=0, respectively.  |
| 428        | ■: in pH 6.5 MilliQ purified water  |
| 429        | •: in water containing FA (10 mg L <sup>-1</sup> )  |
| 430        | $\triangle$ : in water containing FA (10 mg L <sup>-1</sup> ) and nitrates (3.10 <sup>-4</sup> M)                       |
| 431        | $\nabla$ : in water containing nitrates (3.10 <sup>-4</sup> M)  |
| 432        |   |
| 433        | Figure 3: HPLC chromatogram of (A) a neutral solution of FM (3x10 <sup>-5</sup> M) containing fulvio                    |
| 434        | acids (25 mg L <sup>-1</sup> ) at a conversion extent of 20 % and (B) a neutral solution of FM (3x10 <sup>-5</sup> M)   |
| 435        | containing nitrate ions (25 mg L <sup>-1</sup> ) at a conversion extent of 46 %. Both solutions were                    |
| 436        | irradiated in simulated solar light. Photoproducts are numbered as indicated in Table 2.                                |
| 437<br>438 |   |
| 750        |   |

Table 1 : Phototransformation of FM (10<sup>-6</sup> M) in solar light. Rate coefficients deduced from logarithmic decay, correlation values and half-lives considering that samples were irradiated for 13 h per day.

|   | k                    | $R^2$ | t <sub>1/2</sub> |
|---|----------------------|-------|------------------|
| Conditions  | $(s^{-1})$           |       | (h)              |
|   |                      |       |                  |
| FM in MilliQ water, buffered at pH 6.5                  | 1.1×10 <sup>-6</sup> | 0.972 | 175              |
| FM in presence of FA (10 mg L <sup>-1</sup> )           | $2.7 \times 10^{-6}$ | 0.983 | <mark>71</mark>  |
|   |                      |       |                  |
| FM in presence of nitrate ions (25 mg L <sup>-1</sup> ) | $1.7 \times 10^{-5}$ | 0.991 | 11               |
|   |                      |       |                  |
| FM in presence of FA (10 mg L <sup>-1</sup> )           | 1.1×10 <sup>-5</sup> | 0.986 | 18               |
| and nitrate ions (25 mg L <sup>-1</sup> )               |                      |       |                  |

Table 2 : Proposed structure of detected photoproducts, retention time in HPLC-UV, maxima of absorption, molecular ions and main fragments detected by LC-ESI-MS.

| C 1   | 1      | 3.6.11    | C 1:/:                       |                    | / 1 .   |
|---|--------|-----------|------------------------------|--------------------|---|
| formula   | number | Molecular | Condition                    | $\lambda_{max}$    | m/z and main  |
|   |        | weight    | of                           | (nm)               | fragments   |
| Н   | EM     | 222       | obtention                    | 242/275            | (%)   |
| $\sim$ | FM     | 232       |                              | 242/275            | ES <sup>-</sup> : 231                                   |
| F <sub>3</sub> C  |        |           |                              |                    |   |
| HO <sub>2</sub> C CH <sub>3</sub>                                     | acid   | 208       | Pure<br>water                | 224/297            | ES : 207  |
| HOV H   | I      | 224       | FA or                        | 222/324            | ES <sup>-</sup> : 223                                   |
| HO <sub>2</sub> C H <sub>3</sub> CH <sub>3</sub> CH <sub>3</sub>      |        |           | nitrates                     |                    | (179, 134)<br>ES <sup>+</sup> : 225                     |
|   | TT     | 240       | T.A.                         | 242/205            | EC: 247   |
| HO H CO N CH <sub>3</sub>   | II     | 248       | FA or nitrates               | 242/295            | ES <sup>-</sup> : 247<br>(202)<br>ES <sup>+</sup> : 249 |
| F₃C´  |        |           |                              | 220/211            | 70  |
| HO HO CH <sub>3</sub>   | III    | 224       | Nitrates                     | 230/311            | ES <sup>-</sup> : 223 (179, 134)                        |
| HO <sub>2</sub> C isomer of I   |        |           |                              |                    |   |
| HO H   CH <sub>3</sub> CH <sub>3</sub> CH <sub>3</sub>                | IV     | 224       | Nitrates                     | 224/315            | ES <sup>-</sup> : 223 (179, 134)                        |
| HO <sub>2</sub> C isomer of I   |        |           |                              |                    |   |
| H   N—CO—NH <sub>2</sub>  | V      | 204       | Nitrates                     | 241/280            | ES <sup>-</sup> : 203 (160)                             |
| H H   | 371    | 210       | Nitratas                     | 242/270            | EC 217  |
| $F_3C$  | VI     | 218       | Nitrates<br>or FA<br>(minor) | 243/279            | ES <sup>-</sup> : 217 (160)                             |
| — Н   | VII    | 246       | Nitrates                     | 245 and            | ES <sup>-</sup> : 245                                   |
| N—CO—N—CH <sub>3</sub>  | V 11   | 240       | Milates                      | shoulder<br>at 280 | (217, 188, 160)   |
| F₃Ć   |        |           |                              |                    |   |

HO 
$$N = CO - N = CO$$

458 Scheme 1: Fragmentation of photoproduct I

$$H$$
 $CO-NH_2$ 
 $H$ 
 $CO-NH_2$ 
 $H$ 
 $CO-NH_2$ 
 $H$ 
 $CO-NH_2$ 
 $H$ 
 $F_3C$ 
 $F_3$ 

Scheme 2: Fragmentation of photoproducts V and VI

$$F_{3}C$$

$$V + VI + VII$$

$$OH$$

$$H$$

$$CO - N$$

$$CH_{3}$$

$$CH_{3}$$

$$F_{3}C$$

$$HO$$

$$CH_{3}$$

$$CH_{4}$$

$$CH_{4}$$

$$CH_{4}$$

481 Scheme 3: Proposed reaction scheme

484 Figure 1 

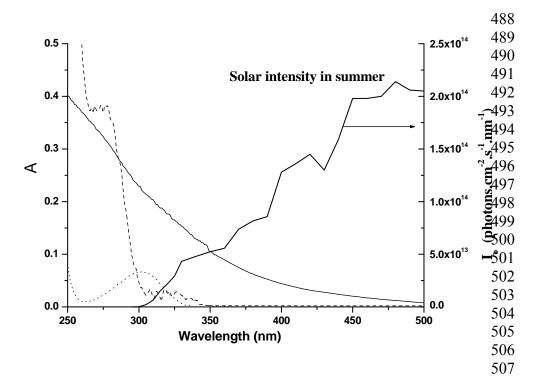
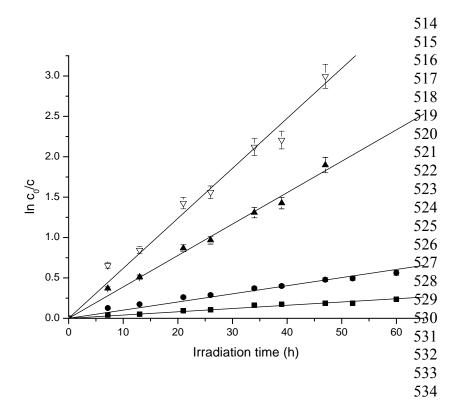


Figure 2



535 Figure 3 

