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1 Structures of benthic prokaryotic communities and their hydrolytic enzyme activities
2 resuspended from samples of intertidal mudflats: an experimental approach

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20

21 **Abstract**

22 Resuspended sediment can increase plankton biomass and the growth of bacteria, thus
23 influence the coastal planktonic microbial food web. But little is known about resuspension
24 itself: is it a single massive change or a whole series of events and how does it affect the

25 quantity and quality of resuspended prokaryotic cells? We simulated the sequential erosion of
26 mud cores to better understand the fate and role of benthic prokaryotes resuspended in the
27 water column. We analyzed the total, attached and free-living prokaryotic cells resuspended,
28 their structure and the activities of their hydrolytic enzymes in terms of the biotic and abiotic
29 factors that affect the composition of microphytobenthic biofilm.

30 Free living prokaryotes were resuspended during the fluff layer erosion phase (for shear
31 velocities below $5 \text{ cm}\cdot\text{s}^{-1}$) regardless of the bed sediment composition. At the higher shear
32 velocities, resuspended prokaryotes were attached to particulate matter. Free and attached
33 cells are thus unevenly distributed, scattered throughout the organic matter (OM) in the
34 uppermost mm of the sediment. Only 10-27% of the total cells initially resuspended were
35 living and most of the Bacteria were Cyanobacteria and Gamma-Proteobacteria; their
36 numbers increased to over 30% in parallel with the hydrolytic enzyme activity at highest
37 shear velocity. These conditions released prokaryotic cells having different functions that lie
38 deep in the sediment; the most important of them are Archaea. Finally, composition of
39 resuspended bacterial populations varied with resuspension intensity, and intense
40 resuspension events boosted the microbial dynamics and enzyme activities in the bottom
41 layers of sea water.

42 **Key words:** resuspension, mudflat biofilm, Bacteria, Archaea, community structure,
43 hydrolytic activities

44 **Introduction**

45 Terrestrial and marine influences converge at tidal flats to produce a complex patchwork of
46 habitats. Research on the productivity of coastal systems has established that intertidal flats
47 influence, both physically and trophically, the adjacent marine and land areas. The intense
48 microbial activity due to microalgae and prokaryotes in these areas results in great biological

49 productivity that is essential for aquaculture. Mudflats are therefore socially and economically
50 extremely important (Héral *et al.*, 1989). Most of the primary production in these areas is due
51 to benthic microalgae, mainly epipellic diatoms (Blanchard *et al.*, 1998; Leguerrier *et al.*,
52 2003) because the turbidity of the waters limits the presence of macrophytes and
53 phytoplankton (McLusky, 1989). The twice-daily emersion in water and the access to daylight
54 stimulates epipellic microalgae to migrate to the surface of the sediment, where they form a
55 continuous biofilm. This biofilm may contain over 20 mg of chlorophyll (Chl *a*) per m² and
56 its photosynthetic activity can increase the algal biomass, so doubling the size of the biofilm
57 (Blanchard *et al.*, 2002). This great productivity can lead to depletion of nitrogen or
58 phosphorus, conditions that stimulate microalgae to secrete exopolymeric substances (EPS).
59 These EPS are produced mainly by "overflow metabolism" or other processes like locomotion
60 (Brouwer and Stal, 2002; Orvain *et al.*, 2003). The EPS form a matrix around the microalgae
61 (Paterson & Black, 1999) that is one of the main resources leading to the rapid development
62 of prokaryotes (Goto *et al.*, 2001; Middelburg *et al.*, 2000; van Duyl *et al.*, 1999). Bacterial
63 production can be as high as or even higher than the production of the microphytobenthos
64 (Hamels *et al.*, 2001; Pascal *et al.* 2009; Van Duyl *et al.*, 1999). While the production of
65 bacteria fluctuates widely during the year this does not result in fluctuations in their
66 abundance. The poor correlation between biomass and bacterial production has been
67 classically explained by their "top-down" disappearance (Thingstad, 2000). Several
68 experiments designed to study grazing on mudflats (Brouage, France) (Pascal *et al.*, 2009)
69 throughout the year have shown that the bacteria consumed by larger animals never exceeded
70 6% of the bacterial production, so having only a limited impact on bacterial production and
71 stock. However, grazers may stimulate bacterial production by disturbing the biofilm which
72 allows the diffusion of oxygen and nutrients throughout the sediment (Alkemade *et al.*, 1992)
73 or by secreting nutrient-rich compounds such as mucus (Riemann and Helmke, 2002). The

74 activity of grazers can also increase the resuspension of sediment leading to the export of
75 benthic prokaryotes to the pelagic environment (Blanchard et al., 1997; Orvain et al., 2006).
76 The free-living or particle-associated prokaryotes transferred to the water column may then
77 find their way into coastal waters (Teal, 1962) where they may be consumed by filter-feeding
78 organisms or pelagic grazers (Guizien et al., 2014; Wainright, 1987).

79 Both tidal currents and wind-induced waves can cause the resuspension of sediment
80 (Blanchard et al., 2002). Sediment erosion generally increases with the rate of shear or friction
81 (u^* expressed in cm s^{-1}). Erosion rates depend on the balance between shear stress and the
82 critical threshold for bed erosion. This threshold depends on properties of the sediment, and
83 physical, geochemical and biological processes (Grabowski et al., 2011). The EPS matrix
84 secreted by the benthic microbial communities plays a key role in sediment
85 stabilization/destabilization, and hence in the export and retention of microphytobenthic and
86 prokaryotic biomasses (Gerbersdorf et al., 2009; Underwood and Paterson, 2003). The
87 sequential resuspension of microorganisms in a subtidal setting subjected to increasing
88 hydrodynamics energy has been attributed to differences in sediment erodibility and the
89 behaviour of microorganisms (Shimeta et al., 2003; Grabowski et al., 2011). This sequential
90 resuspension of microorganisms should be enhanced during the erosion of the biofilms that
91 form on the sediment surface in intertidal flats because the vertical structure of the biota is
92 well defined. Chronic erosion (type 0 erosion) generally occurs and can be defined as a fluff
93 layer erosion (i.e. simple detachment of loose aggregates from the sediment matrix, at low bed
94 shear stress). Orvain et al. (2006) demonstrated the relevance of bioturbation by macrofauna
95 in the creation of such a biogenic fluff layer. This chronic erosion can be followed by a
96 catastrophic erosion (namely mass erosion; type I and/or II according to the consolidation
97 status of the bed load), which corresponds to the general bed failure that can be achieved only
98 when wind-induced waves produce high values of bed shear stress that can overpass the

99 critical threshold for mass erosion (i.e. resistance force). This defined vertical distribution is
100 accompanied by a spatial structure due to top-down control by benthic herbivores (Weerman
101 et al., 2011). The drivers of sequential erosion and the critical erosion thresholds of the
102 microorganisms that make up a biofilm are analyzed in a companion paper (Dupuy et al., this
103 issue). Resuspension of the top few centimeters of sediment can increase the plankton
104 biomass and bacterial growth (Wainright, 1990). This affects the coastal planktonic microbial
105 food web due to the direct movement of cells from the sediment into the water column so
106 increasing the seston concentration or the mineralizing capacity (Wainright, 1987; 1990).
107 Dissolved nutrients and trophic interactions like bacterivory by nanoflagellates can also have
108 indirect effects (Blanchard et al., 1997; Garstecki et al., 2002). Resuspension has been shown
109 to increase the abundance and volume of bacteria in both field and experimental studies
110 (Ritzrau and Graf, 1992; Wainright, 1987), but most of these early studies considered
111 sediment resuspension to be a massive, homogeneous process. The present study resuspension
112 experiments were done to determine whether the sequential resuspension of prokaryotes under
113 varying environmental conditions could alter the direct and indirect effects on planktonic
114 food-web. We monitored the enrichment of resuspended prokaryotic cells and their structure
115 to assess if there was sequential resuspension under all the environmental conditions tested.
116 We also checked whether prokaryotic cells were alive or dead so as to better identify their
117 indirect effects on the plankton food-web. Lastly, we measured the activities of their
118 hydrolytic enzymes. These enzymes regulate the assimilation of nutrient monomers by
119 microbial cells and therefore play an important part in the cycling of organic matter and
120 remineralization.

121

122 **Material and Methods**

123 Field sampling

124 Muddy sediment samples were collected from Marennes–Oléron Bay (Atlantic Coast of
125 France) in July 2008 (Fig. 1) (45°54'50"N, 01°05'25"W) during spring tides (17 and 19 July),
126 maximum tides (21 July) and neap tides (22 and 24 July). At each sampling occasion, two 8
127 cm diameter sediment cores were taken at low tide (middle of the emerged period). Only
128 ridges were sampled since biofilms are not normally found on runnels. The sediment cores
129 were placed in a tank for transport back to the laboratory. The resuspension experiments were
130 done at the time of the immersion period in the field. Three additional 15 cm diameter cores
131 were taken to assess the biotic and abiotic parameters of the top 2 cm of sediment.

132 Resuspension experiments and instrumentation

133 The sediment samples were transferred from the cylindrical cores to the bottom of the flume
134 of an erosion device developed by IFREMER. This was modified to form a straight
135 recirculation flume, named “Erodimetre” (Guizien et al., 2012; Le Hir et al., 2006; Orvain et
136 al., 2007). The flume was filled with filtered artificial seawater (15L) to obtain a baseline
137 close to 0 for each type of microorganism. The discharge through the erosion device was
138 increased in twenty 5-minute steps to yield bed friction velocities ranging from 0 to 11cm.s⁻¹.
139 A total of 6 samples (1.5L each) were taken from the water column of the erodimeter as the
140 flow, and thus shear velocity, was increased. These were used to monitor the resuspension of
141 prokaryotes and the total particle matter (TPM) concentration. The induced bed shear stress
142 was calculated from measurements of the pressure head loss between the upstream and
143 downstream ends of the sample section (Guizien et al., 2012). Friction velocity was calculated
144 as the square root of the bed shear stress divided by the sea water density. Turbidity and
145 fluorescence were continuously recorded and used to measure chl *a* and suspended particular
146 matter. Erosion kinetics were analyzed to evaluate the relationship between bed erodibility

147 (erosion rates of suspended particulate matter and chl *a* biomass) and the sediment properties
148 (abiotic and biotic factors) and their temporal changes over 14 days (Orvain et al., this issue).
149 From this long set of experiments across a spring-neap tidal cycle, we selected 4 dates for a
150 detailed analysis of the fate of bacterial and archaeal benthic communities after resuspension.
151 The dates were chosen to evaluate the response of prokaryote resuspension for different
152 biofilm stages of development. The July 17 sample illustrated an increase in tidal range
153 (midtime between neap tides and spring tides, low tide occurs in the morning), those taken on
154 July 20 and 21 (spring tide) corresponded to a maximum tidal range (just before and just after
155 the peak, low tide occurred around noon), while the July 24 sample reflected the response to a
156 decrease in tidal range (midtime between spring tides and neap tides, low tide started after
157 noon).

158 Abiotic parameters

159 *Salinity and nutrient (NH₄, NO₃, NO₂, PO₄, silicates) concentrations in the upper layer of the*
160 *sediment*

161 Pore water was separated from 300 mL of fresh sediment by centrifugation (4 °C, 3000 g, 15
162 min). Several aliquots of interstitial water were obtained by passage through GF/C filters.
163 Salinity was determined with a refractometer.

164 Ammonia was assayed using the colorimetric Koroleff (1969) technique immediately after
165 filtration. The remainder of the filtered pore water was stored at -20°C. Nitrates, nitrites,
166 silicates and phosphates were assayed using a Skalar automated segmented flow analyzer with
167 classical spectrophotometric techniques (adapted from Treguer and Le Corre, 1975).

168

169 *Carbohydrates and proteins from EPS in the sediment*

170 EPS were extracted from 5 mL sediment as described in Takahashi et al. (2009). Bound EPS
171 were extracted from pellets by suspending each one with about 1 g of cation exchange resin
172 (Dowex Marathon C, Na⁺; Sigma-Aldrich) and 5 mL ASW and gently agitating the
173 suspension for 1 h at 4°C. The samples were then centrifuged (15°C, 3000 g, 10 min) and the
174 resulting supernatants were frozen (-20°C). Aliquots (1 mL) of each fraction were used to
175 measure the carbohydrate contents by the Dubois method (Dubois et al., 1956) with D-
176 glucose (Sigma-Aldrich) as a standard.

177

178 *Total particulate matter (TPM): particulate organic (POM) and mineral (PIM) matter*

179 TPM was measured according to Aminot and Chaussepied (1983). Erodimeter seawater
180 samples (100 - 500 ml) were filtered through a Whatman GF/C (47 mm in diameter) under
181 reduced pressure <10 mm Hg within 1 h after the end of the experiment. Each filter was
182 heated at 490°C for 2 h to eliminate any organic carbon matter and weighed. Filters that had
183 been used to treat samples were rinsed twice with distilled water to remove salt, dried at 60°C
184 for 12 h and weighed to measure the TPM. The proportions of particulate inorganic matter
185 (PIM) and particulate organic matter (POM) were determined by heating the filters at 490°C
186 for 2h and then weighing them.

187 Another 40 mL sample of sediment was stored frozen (-20 °C) and freeze-dried for estimation
188 of C and N using a nitrogen and carbon analyser 1500 (CARLO ERBA).

189

190 Biotic parameters

191 *The algal biomass* in sediment and water was assessed using chlorophyll *a* as a proxy, which
192 was measured fluorometrically (Lorenzen, 1966).

193 *Benthic prokaryotes* were extracted from the sediment as recommended by Danovaro et al.
194 (2001) and Manini and Danovaro (2006) and each sample was divided into two aliquots. One

195 was fixed with (1 %) glutaraldehyde and stored at 4°C. The other was stored at 4°C and used
196 to detect “live” and “dead” cells. Prokaryotes were counted in a FACSCalibur flow cytometer
197 (Becton Dickinson) equipped with an air-cooled laser providing 15 mW at 488 nm with the
198 standard filter set-up. Samples were stained with SYBR Green 1 (1:10X, 10,000 fold dilution
199 of commercial stock, Molecular Probes, Oregon, USA) (Brussaard, 2004; Duhamel and
200 Jacquet, 2006). Populations of prokaryotes differing in size and fluorescence intensity were
201 identified by plotting side scatter (SSC) against green fluorescence (530 nm wave- length,
202 fluorescence channel 1 of the instrument, FL1). We also identified free and particle-associated
203 prokaryotes (attached cells). The live and dead cells in unfixed samples were measured in
204 samples that had been diluted and stained with SYBR Green 1 (1:10X) and 10µg ml⁻¹
205 propidium iodide (Sigma Chemical Co.), and incubated for 15 min in the dark at room
206 temperature (Falcioni et al., 2008). A dot plot of red (670 nm fluorescence channel 3, FL3)
207 against green fluorescence (FL1) distinguished live cell clusters (cells with intact membranes
208 and DNA) from dead cells (with compromised membranes). FCM list modes were analysed
209 using CellQuest Pro software (BD Biosciences, version 4.0).

210 The efficiency with which cells were transferred to the erodimeter water during the erosion
211 was calculated from the enrichment factor (EF):

$$EF = \frac{\text{Water Cell concentration}}{TPM} / \text{Sediment cell content}$$

212

213 Where the water cell concentration is in cell.mL⁻¹, TPM is in mg.mL⁻¹, and the cell content of
214 the 1st cm of sediment core is in cell.mg⁻¹ dry sediment.

215

216 *Potential enzymatic activities*

217 The potential β -glucosidase and leucine aminopeptidase activities in sediment were measured
218 essentially as described by Mallet and Debroas (1999, 2001). The saturating concentrations of
219 methylumbelliferone (MUF)- β -D-glucoside used for the sediment and the water samples
220 were the same: 1mM for β -glucosidase activity and 100 μ M of L-leucine-7-amino-4-
221 β naphthylamide for leucine aminopeptidase activity. Controls and triplicate sediment samples
222 were incubated at *in situ* temperature for 45, 75 and 120 min to measure β -glucosidase
223 activity and for 10, 30 and 60 min to measure leucine aminopeptidase activity. Water samples
224 were incubated with substrate for 24h to measure β -glucosidase activity and for 6h to measure
225 leucine aminopeptidase activity. The fluorescence of the supernatants (14000g at 4°C for 2
226 min) were also measured (SAFAS FLXXenius spectrofluorimeter) using excitation at 365 nm
227 and emission at 460 nm for β -glucosidase activity and excitation at 340 nm and emission at
228 410 nm for aminopeptidase activity. Solutions of 4-methylumbelliferone (0 - 1 μ M) for β -
229 glucosidase activity and 2-naphthylamine (0 - 12.5 μ M) for aminopeptidase activity were
230 freshly prepared with sediment slurry and used as standards. Specific activities were also
231 calculated by dividing the total fluorescent by the total cell concentration ($\text{nmol}\cdot\text{cell}^{-1}\cdot\text{h}^{-1}$).

232 *Prokaryotic genetic structure*

233 Genomic DNA was extracted from fresh sediment (equivalent of 650 mg oven-dried
234 sediment) and purified using the Ultra Clean Soil DNA isolation kit (MoBio Laboratories)
235 according to the manufacturer's instructions. Genomic DNA was extracted from water onto
236 0.2 μ m filter (100 mL) and purified by alkaline lysis (Batisson et al., 2009). DNA quality was
237 checked by electrophoresis on 1% (w/v) agarose gels and quantified using NanoDrop.

238 The bacterial (V6-V8 regions) and archaeal 16S rDNA genes were amplified by PCR using
239 the primers GC-968f and 1401r for bacterial DNA and GC-934f and 1386r for archaeal DNA
240 (MWG-Biotech). The PCR mix (50 μ L) contained 1X PCR buffer, 2.5 mM MgCl_2 , 200 mM

241 of each dNTP, 20 pmol of each primer, 250 ng mL⁻¹ bovine serum albumin (BSA, Sigma), 1.5
242 units of HotStart Taq DNA polymerase (Qiagen), and 30 ng sediment DNA extract. The
243 samples were amplified in an iCycler thermocycler (Bio-Rad) using the following programs:
244 for bacteria: 15min at 95°C, 35 cycles of 1min at 97°C, 1min at 58°C and 1min at 72°C, and
245 finally 10min at 72°C; for archaea: 15min at 95°C for, 14 cycles of 45s at 95°C, for 1min at
246 65°C decreasing to 58.5°C in 0.5°C steps at each cycle, and 30s at 72°C then 20 cycles of 45s
247 at 95°C, 1min at 58°C and 30s at 72°C, and finally 7min at 72°C. PCR products were checked
248 by electrophoresis through 0.8% (w/v) agarose gels and quantified using the DNA
249 quantitation kit fluorescence assay (Sigma). PCR-16S rDNA gene products were analysed by
250 DGGE using a D-Code Universal Mutation Detection System (Bio-Rad). Aliquots of each
251 PCR product (500 ng) were loaded onto 8% polyacrylamide (w/v) denaturing gels with
252 linear gradients of 40-60% (100% denaturant contains 7 M urea and 40% formamide). Gels in
253 1X TAE buffer (pH 8) were subject to 70 V at 60°C for 16h. The DGGE gels were then
254 stained in 1X TAE buffer containing Gel Star diluted 1/20000 (Lonza, Rockland, ME-USA)
255 and the band patterns digitized using a BioSpectrumAC Imaging System (UVP). The digital
256 data were analysed using Gel ComparII software (Applied Maths, Kortrijk, Belgium). The
257 total band intensity of each lane was normalised among lanes and used to calculate the
258 relative abundance of each major OTUs (%). A DNA band was considered to significant if it
259 accounted for more than 0.5 % of the total lane intensity. The richness (*R*) was estimated as
260 the number of bands. The richest samples were cloned and sequenced. The genomic DNA in
261 these samples was amplified with 968f/1401r or 934f/1386r primers and cloned into a pGEM-
262 T-Easy vector (Promega) to construct clone libraries. These clones were reamplified and rerun
263 on DGGE gels to compare them with the parent bands. At least two cloned fragments that co-
264 migrated with the original bands were sequenced (MWG – Biotech). Sequences were
265 submitted to the National Center for Biotechnology Information (NCBI, BLASTn program

266 (Altschul et al., 1997) for species assignment and to the CHECK-CHIMERA program of the
267 Ribosomal Database Project (RDP (Maidak et al., 1999)) to detect potential chimeric artifacts
268 (Kopczynski et al., 1994).

269 Statistical analyses

270 Statistical analyses were carried out using PAST software (PAlaeontological STatistics,
271 Hammer et al., 2001). Principal component analysis (PCA) was used to group samples
272 according to environmental variables using the Pearson correlation coefficient. Prior to use
273 one-way analysis of variance (ANOVA), to test difference between sediment cores for a given
274 parameter, normality and homoskedasticity were checked (Shapiro-Wilk test and Levene's
275 test respectively). Differences between individual means were then compared using Tükey's
276 honest significant difference (HSD) post-hoc test. Significance was set at $p \leq 0.05$.
277 Correlations between abiotic parameters and chlorophyll *a* contents, enzymatic activities and
278 prokaryotic abundances in erodimeter water were checked with the Pearson correlation
279 coefficient (r) and its statistical significance. Pairwise similarity matrices were calculated for
280 the DGGE patterns using the Dice equation for presence/absence and the Bray–Curtis
281 equation for relative peak height data. Dendrograms were generated from the Dice matrix as
282 described by Ward (1963). The consistency of a cluster was described by the cophenetic
283 correlation, which calculates the correlation between the dendrogram-derived similarities and
284 the matrix similarities. Pairwise Analysis of Similarity (ANOSIM, Clarke and Warwick, 2001)
285 was used to test if the sediment samples were similar in term of chemical and biological
286 parameters. It tested the hypothesis that parameters within each sediment sample were more
287 similar to each other than to parameters in the other samples. The statistical significance of
288 the separation was given by a Bonferroni-corrected *p-value* ($R, p < 0.05$).

289

290 **Results**

291 *Sediment*

292 The biological and chemical parameters of the sediment cores taken at different dates were
293 significantly different (R: 0.55 p: 0.0001, Fig. 2). The July 17 cores were opposite to the other
294 samples (Fig. 2) with significant lowest salinity (Table 1), which implied the highest water
295 content. This date was also characterized by significant lower colloidal carbohydrate
296 concentrations than those of the July 21 and 22 (Table 1). The July 22 samples were also
297 significantly differed to the other sampling dates (Fig. 2) with silica concentrations
298 significantly lower than the other (Table 1). The profiles of the core samples taken on this date
299 seemed to be opposite to those of the July 21 and 24 samples, whose sediments had the
300 highest nitrogen concentrations (Fig. 2; Table 1). The highest prokaryotic abundances were
301 observed in the July 17 and 24 samples (Table 1). For the July 17 and 22 samples, highest
302 leucine aminopeptidase activities were observed with for the July 22 sample highest β -
303 glucosidase activity too (Table 1).

304 The structures of the bacterial communities in the four sediment samples differed significantly
305 (R: 0.9, p: 0.008). We found only 3 major phyla: Gamma-Proteobacteria (mainly
306 Pseudomonadales and Chromatiales and unidentified Gamma-Proteobacteria), Actinobacteria
307 and Cyanobacteria (Fig. 3A). The richness (as numbers of DGGE bands) varied from 22 on
308 July 24 to 33 on July 21 (Fig. 3A). Crenarchaeota formed the major group except in the July
309 17 samples (Fig. 3B). The July 22 cores were the least rich (14) and the July 17 cores were the
310 richest (33).

311 *Microorganism resuspensions in the erodimeter water*

312 We gradually increased the shear velocity (u^*) in the erosion flume to mimic low to highly
313 turbulent flow. There were two phases of total particulate matter (TPM) erosion. The shear
314 velocity in the first phase was low ($u^* < 5 \text{ cm.s}^{-1}$) and the concentration of particles was very
315 low. Only the biogenic fluff layer was resuspended during this phase. This layer is mostly
316 composed of particles detached from the sediment matrix like the fluff layer or track and/or
317 pseudofecal mounds created by faunal bioturbation). In the second phase mass erosion of the
318 sediment occurred and the number of particles increased (Orvain et al., this issue).

319 Enrichments of resuspended prokaryotes (total cell fraction) was more effective with
320 velocities below 5 cm.s^{-1} , before and during the precocious erosion phase, when the biogenic
321 fluff layer was resuspended, as this layer is created by surface fauna activity like tracks (Fig.
322 4A). The total cell abundance in the erosion flume fluctuated little, from $1.3 \cdot 10^5 \text{ cells.ml}^{-1}$ on
323 July 17 and $3.6 \cdot 10^5 \text{ cells.ml}^{-1}$ on July 22 under our controlled conditions. The enrichment in
324 free cells followed the same pattern as that for total prokaryotes (Fig. 4A) and represented
325 $75.1 \pm 16.5\%$ of total resuspended cells. Enrichment at the lowest friction was maximal on
326 July 17 and decreased as erosion increased. There was little enrichment in attached cells but it
327 increased during resuspension of the biogenic fluff layer (Fig. 4B). The concentration of
328 attached cells increased when there was mass erosion of the sediment and enrichment reached
329 25% on July 24 and 44% on July 21. Less than a third of the cells in the erodimeter water
330 were intact (potentially active). Their enrichments followed the same pattern as that of the
331 free cells but there were fewer of them (Fig. 4C). They accounted for only 10% of the total
332 cells on July 17 and 27% on July 21, but their abundance increased to more than 30% of the
333 total cells at the end of the experiment, during sediment mass erosion. The β -glucosidase
334 activity fluctuated between $2.4 \cdot 10^{-7}$ and $1.9 \cdot 10^{-6} \text{ nmol.cell}^{-1} \text{ h}^{-1}$ and that of leucine-
335 aminopeptidase between $6.5 \cdot 10^{-7}$ and $4.4 \cdot 10^{-6} \text{ nmol.cell}^{-1} \text{ h}^{-1}$. The activities had lowest values

336 when the shear velocity was less than $5 \text{ cm}\cdot\text{s}^{-1}$, except for the samples collected on July 21
337 and July 24 (Fig. 5A, B).

338 The PCA map indicated that the enrichments in total, free and live cells varied inversely as
339 enrichment in total particulate matter, chlorophyll *a* and the enzyme activities (Fig. 6). The
340 other axis was better characterized by the C:N ratio and the concentration of attached cells.
341 These two axes clearly separated the samples into two major clusters (Fig. 6). The samples
342 corresponding to mass erosion of sediment, rich in TPM lay on the first axis. Conversely, the
343 samples corresponding to erosion of the biogenic layer lay on the second.

344 *Structure of resuspended prokaryotes*

345 Attached cells accounted for more of the resuspended bacteria than did free cells for all dates
346 (Fig. 7). Bacterial populations were resuspended sequentially, especially for the July 17 and
347 July 24 samples, Gamma-proteobacteria and Pseudomonales appeared first, followed by
348 Actinobacteria and finally the Chromatiales (Fig. 7A) and the Cyanobacteria (Fig. 7E), which
349 were also resuspended at the beginning of the experiment. Nevertheless, enrichment varied
350 with the sampling date both qualitatively (OTUs) and quantitatively (relative abundance of
351 each OTU). Concerning the archaeal community, Archaea were only detected during the mass
352 erosion events. The dominant Archaea were crenarchaeotal populations (mainly unidentified
353 Crenarchaeota, MG I and thermoprotei) (Fig. 8). Archaeal cells were very effectively
354 resuspended, except from the July 17 sample, because the richnesses in the resuspended
355 fractions were greater than those of the sediment. The mass erosion of sediment collected on
356 July 17 produced the greatest richness and diversity and there was a sequential appearance of
357 populations, as for the Bacteria (Fig.8).

358

359 **Discussion**

360 It has been claimed that erosion of the sediment bed affects the pelagic trophic network by
361 inhibiting primary production because it increases turbidity. But it may also be affected
362 directly or indirectly by active prokaryotes released from the sediment. We have characterised
363 the sequential release of prokaryotes from bed sediment samples in terms of their structure
364 (free or attached cells and their composition), physiology (alive or dead) and enzyme activity
365 that were resuspended by increasing erosive forces from samples with and without biofilm.

366

367 *Resuspension of prokaryotic cells*

368 Subtidal sediments remain unconsolidated and are relatively easy to erode. But sediment in
369 intertidal areas exposed to the air tends to become consolidated after recurrent desiccation and
370 biofilm development (De Brouwer and Stal, 2001; Orvain et al., 2007; Porter et al. 2010;
371 Underwood and Paterson, 2003). The sediment samples we studied were very resistant to
372 erosion. Mass erosion was produced only by high shear velocities ($u^* > 5 \text{ cm.s}^{-1}$). These shear
373 velocities can only be generated by wind-induced waves in the field. This tendency held for
374 all sediments, despite differences in their chemical characteristics. Very low shear velocities
375 resulted in limited erosion of the TPM in all our experiments. The intensity of this 'fluff layer
376 erosion' varied with the composition of the bed sediment and the critical threshold shear
377 velocity (Dupuy et al., this issue; Orvain et al., this issue).

378 Prokaryotes were always resuspended during the fluff layer erosion phase when shear
379 velocities were below 5 cm.s^{-1} and before the great increase in particulate matter erosion at
380 higher shear velocities. Low friction velocities, when the concentrations of TPM were low,
381 produced the greatest enrichment in prokaryotic cells. Enrichment decreased at the end of the
382 fluff layer erosion when mass erosion began. The fluff layer in a mud site can be very
383 important for sediment transport (Orvain et al., 2007; Ziervogel et al., 2006), because this
384 material is especially enriched in OM and constitute a chronic erosion compared to the

385 general bed failure, which can be defined as ‘catastrophic’ (Mariotti and Fagherazzi, 2012).
386 The unconsolidated surface fluff layer, with a water content close to 85–98% of dry weight,
387 was resuspended using low hydrodynamic conditions. The cells were unevenly distributed;
388 they are scattered throughout the top mm of the sediment in the organic matter and this
389 distribution is probably due to the patchy distribution of microphytobenthic biofilms (Dupuy
390 et al., this issue). The surface layers of sediment always contained more bacteria than did the
391 subsurface suboxic sediment layers (Parkes et al., 2000). The decrease in microbial cell
392 counts down through the sediment profile and the shift in community structure probably
393 results from the decreased in organic carbon quality and availability in aged, deeply buried
394 sediments (Parkes et al., 2000). The volume of pore-water may also be a major determinant of
395 prokaryotic abundance in sediments (Schmidt et al., 1998). The force that determines the
396 decrease in prokaryotic abundance and changes in community structure down the vertical
397 profile of a sediment can be a complex mixture of biotic and abiotic factors, such as poor
398 mixing or bioturbation, sediment composition (grain size and distribution, sediment water
399 content), energy stress conditions (low food, electron acceptors, availability of dissolved
400 substrates and metabolites), and/or predatory pressures and competition.

401 Most of the cells in the water were free. Attached bacteria are less abundant in many pelagic
402 marine ecosystems than are free-living bacteria. They generally account for less than 5% of
403 the total bacteria, but this can rise to 10% (Lapoussière et al., 2011). However, particle-
404 attached bacteria can form a much greater fraction (up to 96%) of the total bacteria in
405 estuarine systems (Lapoussière et al., 2011). We found that they accounted for less than 20%
406 of the total suspended cells until the shear velocities increased above $5\text{cm}\cdot\text{s}^{-1}$, except in the
407 samples taken on July 21 (Fig 5A). Previous studies have demonstrated that the abundance
408 (or biomass) of particle-attached bacteria depended on the number and/or size of particles
409 available for attachment (Cho and Azam, 1988; Garneau et al., 2009; Lapoussière et al.,

410 2011). We see a negative relationship between the attached bacteria and the POM in the
411 samples taken on July 21 ($r = -0.97$, $p < 0.05$) and July 22 ($r = -0.88$, $p < 0.05$). The high density
412 of motile grazers caused the surface of the bed to be strongly pelletised, which significantly
413 increased the velocity at which eroded sediment settled. These results may also be influenced
414 by the fact that the abundance (or biomass) of particle-attached bacteria is linked to the
415 biogeochemical composition of the suspended material (Fandino et al., 2001; Wang and Yin,
416 2009). Particle-attached bacteria can play an important role in the transformation of freshly
417 produced POC despite their small contribution (often less than 30%) to total bacterial biomass
418 (Ghiglione et al., 2009). The positive correlation between leucine aminopeptidase activity and
419 attached cells ($r = 0.45$, $p < 0.05$) supports this idea. Extracellular enzymes may be important for
420 the dispersion or release of attached microbial cells.

421 Few of the resuspended cells were alive, in contrast to the percent of living cells in the
422 sediment (around 50%). Thus dead cells (those with damaged membranes) accounted for the
423 greatest fraction (70 - 74%) of resuspended bacterial assemblages in all samples analyzed.
424 These results suggest that the high mechanical stress produced by the shear velocity interferes
425 with the benthic cells, as there were many active cells in the sediment. But whether the cells
426 were attached or free did not seem to be important, as we found significant positive
427 relationships between the attached and active cells ($r = 0.98$, $p < 0.05$) and between free and
428 active cells ($r = 0.90$, $p < 0.05$). Nevertheless, our results corroborate Pusceddu et al. (2005) who
429 concluded that the metabolically active bacteria can respond differently to resuspensions of
430 sediment produced by low and high energies. They showed that disturbing a sediment with
431 low-energy force produced a slightly positive response from the active bacterial fraction only
432 during the initial resuspension event. But we found high shear velocities also enriched the
433 resuspended active cells (Fig.4D), unlike Pusceddu et al (2005), who observed that the
434 metabolically active bacteria decreased when turbulence was high. We conclude that

435 disturbing the sediment with high-energy forces has little effect on the fraction of
436 metabolically active bacteria resuspended.

437 Many studies (Chróst and Riemann, 1994; Cotner, 2000; Ritzrau and Graf, 1992; Ritzrau,
438 1996) have demonstrated that resuspending sediment enhances the microbial dynamics and
439 enzyme activities in the bottom layers of water. The extent to which sediment resuspension
440 affects the enzyme activity in the water column depends on the characteristics of the
441 suspended matter and on the depth of the water column (Ziervogel and Arnosti, 2009). We
442 found that the POM was positively correlated with both the β -glucosidase ($r=0.52$, $p<0.05$)
443 and leucine aminopeptidase ($r=0.66$, $p< 0.05$) activities, as was the PIM concentration
444 ($r=0.50$, $p<0.05$ for β -glucosidase and $r=0.63$, $p< 0.05$ for leucine aminopeptidase).
445 Microphytobenthic algae also channel a major fraction of their total primary production into
446 the synthesis of extracellular polymeric substances and are thus a significant source of oxygen
447 and organic matter for the growth of benthic bacteria (Boer et al., 2009; De Brouwer and Stal,
448 2001; Underwood and Kromkamp, 1999). Both the β -glucosidase ($r=0.72$, $p<0.05$) and
449 leucine aminopeptidase activities ($r=0.77$, $p< 0.05$) were positively correlated with the
450 chlorophyll *a* concentration. The association of specific enzymes with the EPS matrix may
451 prolong their activities and help them resist fluctuations in the environment. The activities of
452 enzymes and their patterns in the Delaware estuary were probably affected by the sediment-
453 associated microorganisms resuspended into the bottom water layers, increasing the number
454 of microbial cells and the hydrolytic activities of the extracellular enzymes (Ziervogel and
455 Arnosti, 2009).

456 *Erosion and enrichment of prokaryotic OTUs*

457 Proteobacteria are the most abundant bacteria in most surface marine sediments. They can
458 account for over 50% of the microbial biomass (Bowman and McCuaig, 2003; Kim et al.,

459 2008, Pachiadaki et al., 2011). Gamma-Proteobacteria seems to be the most significant clade
460 in most marine sediments (Inagaki et al., 2003; Feng et al., 2009; Polymenakou et al., 2005).
461 Gamma-Proteobacteria accounted for up to 10% of the total cells in the upper 2-cm layer of
462 sediment and for 20% of the prokaryotic rRNA in the Smeerenburgfjorden sediments
463 (Ravenschlag et al., 2001). The percentages are lower than those in our samples: Gamma-
464 Proteobacteria accounted for up to 50% of our total cells (Fig. 7). Analysis of the structures of
465 resuspended prokaryotic communities indicated a sequential enrichment in bacterial OTUs
466 with increasing shear velocity. Pseudomonadal OTUs predominated under low stress
467 condition. But the richness of erodimeter water was always lower than that of the sediment.
468 These increases were partly due to the emergence of Actinobacteria and Chromatiales OTUs.
469 These Bacteria are often found in sediments like those at Brouage mudflat that are exposed to
470 sunlight due to tides and shallow water (Borin et al., 2009). Here, some Cyanobacteria OTUs
471 appeared in the erodimeter water when shear were low, and reappeared when shear velocities
472 were high. The strong hydrodynamic forces encountered at the field site led to constant
473 vertical and horizontal mixing of the upper five centimeters of sediment (Hedtkamp, 2005).
474 This can impose strong selective pressure on the microbial community and restrict access to
475 this habitat to a relatively narrow range of Bacteria that can cope with occasional
476 resuspension, the physical abrasion generated by moving sediment particles, grazing, and
477 rapid fluctuations in the concentrations of oxygen and other nutrients (Boer et al., 2009).
478 Some authors have suggested that the bacteria attached to particles are phylogenetically
479 different from free-living bacteria (Crump et al., 1999; Gliglione et al., 2009). The present
480 study results do not support this idea: discrepancies may be due to differences in the
481 techniques used, or in the types of particles and trophic conditions. Attached bacteria may be
482 ubiquitous species and probably result from rapid exchanges with the more diverse reservoir
483 of free-living communities. The large number of similar OTUs we found in attached and free-

484 living bacteria leads us to believe that free-living and attached bacteria are not separate
485 entities but interacting assemblages, as suggested by Riemann and Winding (2001). Bacteria
486 often seem to develop concomitantly with benthic microalgae and they adapt quickly to
487 microalgal exudates by changing the profile of their enzymatic activities. The composition of
488 our bacterial communities differed markedly, depending on the presence (July 17) or absence
489 (July 21) of a microphytobenthic biofilm, in agreement with Lubarsky et al. (2010) (see also
490 Dupuy et al. this issue).

491 The resuspension produced by our technique provided a higher archaeal specific richness (30
492 OTUs on average) than Bacteria. However, this applied only to attached cells and samples
493 produced using the highest shear rates ($> 7 \text{ cm s}^{-1}$). The presence of large amounts of
494 Euryarchaeota and Crenarchaeota generally agrees with the clone library studies of Kim et al.
495 (2005; 2008). But they found that Euryarchaeota seemed to be more abundant and diverse
496 than we do. Most of the archaeal sequences in all our samples were of unidentified classes,
497 but we did find Thermoprotei, a group that includes anaerobes and sulfur reducers.
498 Methanomicrobia were the most abundant methanogens. Methanomicrobiales may have
499 occurred in the sulfate-free layers because there was less competition from sulfate reducers.
500 They use substrates released by decaying algae and other organisms, like methylamines and
501 dimethylsulfide. These substrates are not used by most other physiological groups and are
502 mostly found near the surface of the sediment (Wilms et al., 2006).

503 Conclusions

504 We have identified two distinct patterns of resuspended prokaryotic cells, produced by
505 applying different shear stresses to the sediment. Friction velocities of less than 5 cm s^{-1} ,
506 corresponding to a normal low choppy tide, lead to the resuspension of cells that are mostly
507 free-living, regardless of the physico-chemical characteristics of the sediment or the

508 development of a biofilm on its surface. These cells are mainly Cyanobacteria and Gamma-
509 Proteobacteria that secreted some extracellular enzymes and can become integrated into, and
510 thus affect, the coastal planktonic microbial food web (Garstecki et al., 2002). Friction
511 velocities greater than 5 cm s^{-1} , corresponding to rough tides with waves, favour the
512 resuspension of attached cells. Studies on the effects of fluid shear on microbial activity in the
513 water column have shown that particle-attached bacteria benefit nutritionally more than free-
514 living forms given that the host particle is large enough (Jumars et al., 1993; Lazier and
515 Mann, 1989). It is under these circumstances that the greatest proportion of bacterial and
516 archaeal OTUs with different functions are recruited from deep sediment. The shear velocity
517 was strongly correlated with hydrolytic enzyme activities and linked to the increase in
518 resuspended intact cells (potentially active). These attached prokaryotes could provide organic
519 matter to the pelagic zone and enhance the breakdown of benthic organic matter (Ståhlberg et
520 al., 2006). Resuspended sediments that are rich in organic and inorganic nutrients may
521 stimulate the large scale production of planktonic heterotrophs, even at low temperatures
522 (Fiordelmondo 2004).

523

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528

529 **References**

530 Alkemade, R., Wielemaker, A., de Jong, S.A., Sandee, A.J.J. 1992. Experimental evidence for
531 the role of bioturbation by the marine nematode *Diplolaimella brucei* in stimulating the
532 mineralization of *Spartina anglica* detritus. Marine Ecology Progress Series, 90, 149-155.

533 Altschul, S.F., Madden, T.L., Schaffer, A.A., Zhang, J., Zhang, Z., Miller, W., Lipman, D.J.
534 1997. Gapped BLAST and PSI-BLAST: a new generation of protein database search
535 programs. *Nucleic Acids Research*, 25, 3389–3402

536 Aminot, A., Chaussepied, M., 1983. Manuel des analyses chimiques en milieu marin.
537 CNEXO, Brest.

538 Batisson, I., Crouzet, O., Besse-Hoggan, P., Sancelme, M., Mangot, J. F., Mallet, C., Bohatier,
539 J. (2009). Isolation and characterization of mesotrione-degrading *Bacillus* sp. from soil.
540 *Environmental Pollution*, 157, 1195-1201.

541 Blanchard, G.F., Simon Bouhet B., Guarini, J.-M. (2002). Properties of the dynamics of
542 intertidal microphytobenthic biomass. *Journal of the Marine Biological Association of the*
543 *UK*, 82, 1027-1028

544 Blanchard, G.F., Guarini, J.-M., Bacher, C., Huet, V. 1998. Contrôle de la dynamique à court
545 terme du microphytobenthos intertidal par le cycle exondation-submersion. *Comptes*
546 *Rendus de l'Académie des Sciences*, 321,501-508

547 Blanchard, G.F., Sauriau, P-G., Cariou-Le Gall, V., Garet, M.J., Olivier, F. 1997. Kinetics of
548 tidal resuspension of microbiota: testing the effects of sediment cohesiveness and
549 bioturbation using flume experiments. *Marine Ecology Progress Series*, 151, 17-25.

550 Böer, S.I., Hedtkamp, S.I., van Beusekom, J.E., Fuhrman, J. A., Boetius, A., Ramette, A.
551 2009. Time-and sediment depth-related variations in bacterial diversity and community
552 structure in subtidal sands. *The ISME journal*, 3, 780-791.

553 Borin, S., Brusetti, L., Daffonchio, D., Delaney, E., Baldi, F. 2009. Biodiversity of prokaryotic
554 communities in sediments of different sub-basins of the Venice lagoon. *Research in*
555 *Microbiology*, 160, 307-314.

556 Bowman, J.P., McCuaig, R.D. 2003. Biodiversity, community structural shifts, and
557 biogeography of prokaryotes within Antarctic continental shelf sediment. *Applied and*
558 *Environmental Microbiology*, 69, 2463-2483.

559 Brouwer, J.F.C., de Stal, L.J 2002. Daily fluctuations of exopolymers in cultures of the
560 benthic diatoms *Cylindrotheca closterium* and *Nitzschia* sp. (Bacillariophyceae). *Journal of*
561 *Phycology*, 38, 464-472.

562 Brussaard, C.P. 2004. Optimization of procedures for counting viruses by flow cytometry.
563 *Applied and Environmental Microbiology*, 70, 1506-1513.

564 Cho, B.C., Azam, F. 1988. Major role of bacteria in biogeochemical fluxes in the ocean's
565 interior. *Nature*, 332, 441-443.

566 Chróst, R.J., Riemann, B. 1994. Storm-stimulated enzymatic decomposition of organic matter
567 in benthic/pelagic coastal mesocosms. *Marine Ecology Progress Series*, 108, 185-192.

568 Clarke K.R., Warwick R.W. 2001. *Change in Marine Communities: An Approach to Statistical*
569 *Analysis and Interpretation*, 2nd edn. Primer-E, Plymouth.

570 Cotner, J.B., Johengen, T.H. Biddanda, B.A. 2000. Intense winter heterotrophic production
571 stimulated by benthic resuspension. *Limnology and Oceanography*, 45, 1672–1676.

572 Crump, B.C, Armbrust, E.V., Baross, J.A. 1999. Phylogenetic analysis of particle-attached
573 and free-living bacterial communities in the Columbia River, its estuary, and the adjacent
574 coastal ocean. *Applied and Environmental Microbiology*, 65, 3192–3204.

575 Danovaro, R., Dell'Anno, A., Trucco, A., Serresi, M., Vanucci, S. 2001. Determination of
576 virus abundance in marine sediments. *Applied and Environmental Microbiology*, 67, 1384-
577 1387.

578 De Brouwer, J.F.C., Stal, L.J. 2001. Short-term dynamics in microphytobenthos distribution
579 and associated extracellular carbohydrates in surface sediments of an intertidal mudflat.
580 *Marine Ecology Progress Series*, 218, 33-44.

581 Dubois, M., Gilles, K.A, Hamilton, J.K., Rebers, P.A., Smith, F., 1956. Colorimetric method
582 for determination of sugars and related substances. *Analytical Chemistry*, 28, 350-356.

583 Duhamel, S., Jacquet, S. 2006. Flow cytometric analysis of bacteria-and virus-like particles in
584 lake sediments. *Journal of Microbiological Methods*, 64, 316-332.

585 Fandino, L.B., Riemann, L., Steward, G.F., Long, R.A., Azam, F. 2001. Variations in bacterial
586 community structure during a dinoflagellate bloom analyzed by DGGE and 16S rDNA
587 sequencing. *Aquatic Microbial Ecology*, 23, 119.

588 Feng, B.W., Li, X.R., Wang, J.H., Hu, Z.Y., Meng, H., Xiang, L.Y., Quan, Z.X. 2009.
589 Bacterial diversity of water and sediment in the Changjiang estuary and coastal area of the
590 East China Sea. *FEMS Microbiology Ecology*, 70, 236-248.

591 Fiordelmondo, C., Pusceddu, A. 2004. Short-term response of benthic bacteria and
592 nanoflagellates to sediment resuspension: an experimental study. *Chemistry and Ecology*,
593 20, 107-121.

594 Garneau, M.È., Vincent, W.F., Terrado, R., Lovejoy, C. 2009. Importance of particle-
595 associated bacterial heterotrophy in a coastal Arctic ecosystem. *Journal of Marine Systems*,
596 75, 185-197.

597 Garstecki, T., Wickham, S. A., Arndt, H. 2002. Effects of experimental sediment resuspension
598 on a coastal planktonic microbial food web. *Estuarine, Coastal and Shelf Science*, 55, 751-
599 762.

600 Gerbersdorf, S.U., Bittner, R., Lubarsky, H., Manz, W., Paterson, D.M. 2009. Microbial
601 assemblages as ecosystem engineers of sediment stability. *Journal of Soils and Sediments*,
602 9, 640-652.

603 Ghiglione, J.F., Conan, P., Pujo-Pay, M. 2009. Diversity of total and active free-living vs.
604 particle-attached bacteria in the euphotic zone of the NW Mediterranean Sea. *FEMS*
605 *Microbiology Letters*, 299, 9-21.

606 Goto, N., Mitamura O., Terai, H. 2001. Biodegradation of photosynthetically produced
607 extracellular organic carbon from intertidal benthic algae. *Journal of Experimental Marine*
608 *Biology and Ecology*, 257, 73-86.

609 Grabowski, R.C., Droppo, I.G., Wharton, G. 2011. Erodibility of cohesive sediment: the
610 importance of sediment properties. *Earth-Science Reviews*, 105, 101-120.

611 Guizien, K., Orvain F., Duchene J.C., Le Hir P. 2012. Accounting for rough bed friction
612 factors of mud beds due to biological activity in erosion experiments. *Journal of Hydraulic*
613 *Engineering*, 138, 979-984.

614 Guizien, K., Dupuy, C., Ory, P., Montanié, H., Hartmann, H., Chatelain, M., Karpytchev, M.
615 2014. Microorganism dynamics during a rising tide: Disentangling effects of resuspension
616 and mixing with offshore waters above an intertidal mudflat. *Journal of Marine Systems*,
617 129, 178-188.

618 Hamels, I., Muylaert, K., Casteleyn, G., Vyverman, W. 2001. Uncoupling of bacterial
619 production and flagellate grazing in aquatic sediments: a case study from an intertidal flat.
620 *Marine Ecology Progress Series*, 25, 31-42.

621 Hammer, Ø., Harper, D.A.T., Ryan, P.D. 2001. PAST: Paleontological Statistics Software
622 Package for Education and Data Analysis. *Palaeontologia Electronica* 4, 1-9.

623 Hedtkamp, S. 2005. Shallow subtidal sand: Permeability, nutrient dynamics
624 microphytobenthos and organic matter. Christian-Albrechts-Universität, Kiel.

625 Héral, M., Bacher, C., Deslous-Paoli, J.M. 1989. La Capacité trophique des bassins
626 ostréicoles. In: Troadec JP (ed) L'homme et les ressources halieutiques : essai sur l'usage
627 d'une ressource renouvelable Ifremer, Brest, pp 225-259.

628 Inagaki, F., Suzuki, M., Takai, K., Oida, H., Sakamoto, T., Aoki, K., ... & Horikoshi, K. 2003.
629 Microbial communities associated with geological horizons in coastal subseafloor
630 sediments from the Sea of Okhotsk. *Applied and Environmental Microbiology*, 69, 7224-
631 7235.

632 Jumars, P.A., Deming, J.W., Hill, P.S., Karp-Boss, L., Yager, P.L., Dade, W.B. 1993.
633 Constraints on marine osmotrophy in an optimal foraging context. *Marine Microbial Food*
634 *Webs*, 7, 121–159

635 Kim, B.S., Kim, B.K., Lee, J.H., Kim, M., Lim, Y.W., Chun, J. 2008. Rapid phylogenetic
636 dissection of prokaryotic community structure in tidal flat using pyrosequencing. *Journal*
637 *of Microbiology*, 46, 357-363.

638 Kim, B.S., H.M. Oh, H. Kang, J. Chun. 2005. Archaeal diversity in tidal flat sediment as
639 revealed by 16S rDNA analysis. *Journal of Microbiology*, 43, 144-151.

640 Kopczynski, E.D., Bateson, M.M., Ward, D.M. 1994. Recognition of chimeric small-subunit
641 ribosomal DNAs composed of genes from uncultivated microorganisms. *Applied and*
642 *Environmental Microbiology*, 60, 746–748

643 Lapoussière, A., Michel, C., Starr, M., Gosselin, M., Poulin, M. 2011. Role of free-living and
644 particle-attached bacteria in the recycling and export of organic material in the Hudson Bay
645 system. *Journal of Marine Systems*, 88, 434-445.

646 Lazier, J.R.N., Mann, K.H. 1989. Turbulence and the diffusive layers around small organisms.
647 Deep Sea Research Part A. Oceanographic Research Papers, 36, 1721-1733.

648 Le Hir, P., Cann, P., Jestin, H., Bassoulet, P., 2006. Instrumentation légère pour la mesure de
649 l'éroductibilité des sédiments vaseux ou sablovasseux." Proc. IXèmes Journées Nationales
650 Génie Côtier-Génie Civil, PARALIA, Nantes, France.

651 Leguerrier, D., Niquil, N., Boileau, N., Rzeznik, J., Sauriau, P.G, Le Moine, O., Bacher, C.
652 2003. Numerical analysis of the food web of an intertidal mudflat ecosystem on the
653 Atlantic coast of France. Marine Ecology Progress Series, 246, 17-37.

654 Lorenzen, C.J., 1966. A method for the continuous measurement of in vivo chlorophyll
655 concentration. Deep-Sea Research, 13, 223–227.

656 Lubarsky, H.V., Hubas, C., Chocholek, M., Larson, F., Manz, W., Paterson, D.M.,
657 Gerbersdorf, S.U. 2010. The stabilisation potential of individual and mixed assemblages of
658 natural bacteria and microalgae. PloS one, 5, e13794.

659 Maidak, B.L., Cole, J.R., Parker, Jr. C.T., Garrity, G.M., Larsen, N., Li B., Lilburn, T.G.,
660 McCaughey, M.J., Olsen, G.J., Overbeek, R., Pramanik, S., Schmidt, T.M., Tiedje, J.M.,
661 Woese, C.R. 1999. A new version of the RDP (ribosomal database project). Nucleic Acids
662 Research, 27, 171–173

663 Mallet, C., Debroas, D. 1999. Relations between organic matter and bacterial proteolytic
664 activity in sediment surface layers of a eutrophic lake (Lake Aydat, Puy de Dôme, France).
665 Archiv für Hydrobiologie, 145, 39-56.

666 Mallet, C., Debroas, D. 2001. Regulation of β - and α -glycolytic activities in the sediments of a
667 eutrophic lake. Microbial Ecology, 41, 106-113.

668 Manini, E., Danovaro, R. 2006. Synoptic determination of living/dead and active/dormant
669 bacterial fractions in marine sediments. *FEMS Microbiology Ecology*, 55, 416-423.

670 Mariotti, G., Fagherazzi, S. 2012. Modeling the effect of tides and waves on benthic biofilms.
671 *Journal of geophysical research*, 117, G04010, doi:10.1029/2012JG002064

672 McLusky, D.S. 1989. *The estuarine ecosystem*, Chapman & Hall, New York, 215 pp.

673 Middelburg, J.J., Baranguet, C., Boschker, H.T.S., Herman, P.M.J., Moens, T., Heip, C.H.R.
674 2000. The fate of intertidal microphytobenthos carbon: an in situ ¹³C-labeling study.
675 *Limnology and Oceanography*, 45, 1224-1234.

676 Moeseneder, M.M., Herndl, G.J. 1995. Influence of turbulence on bacterial production in the
677 sea. *Limnology and Oceanography*, 40, 1466-1473.

678 Orvain, F., Galois, R., Barnard, C., Sylvestre, A., Blanchard, G., Sauriau, P.-G. 2003.
679 Carbohydrate production in relation to microphytobenthic biofilm development: an
680 integrated approach in a tidal mesocosm. *Microbial Ecology*, 45, 231-257.

681 Orvain F., Sauriau P.G., Bacher C., Prineau M. 2006. The influence of sediment cohesiveness
682 on bioturbation effects due to *Hydrobia ulvae* on the initial erosion of intertidal sediments:
683 A study combining flume and model approaches. *Journal of Sea Research*, 55, 54-73.

684 Orvain, F., Sauriau, P.G., Le Hir, P., Guillou, G., Cann, P., Paillard, M. 2007. Spatio-temporal
685 variations in intertidal mudflat erodability: Marennes-Oléron Bay, western France.
686 *Continental Shelf Research*, 27, 1153-1173.

687 Pachiadaki, M.G., Kallionaki, A., Dähmann, A., De Lange, G.J., Kormas, K.A. 2011.
688 Diversity and spatial distribution of prokaryotic communities along a sediment vertical
689 profile of a deep-sea mud volcano. *Microbial Ecology*, 62, 655-668.

690 Parkes, R.J., Cragg, B.A., Wellsbury, P. 2000. Recent studies on bacterial populations and
691 processes in subseafloor sediments: a review. *Hydrogeology Journal*, 8, 11-28.

692 Pascal, P-Y, Dupuy, C., Richard, P., Mallet, C., Arminot du Châtelet, E., Niquil, N. 2009.
693 Seasonal variation in consumption of benthic bacteria by meio- and macrofauna in an
694 intertidal mudflat. *Limnology and Oceanography*, 54, 1048-1059.

695 Paterson, D.M., Black, K.S. 1999. Water flow, sediment dynamics and benthic biology. . In:
696 Nedwell D.B., Raffaelli D.G. (eds), *Estuaries. Advances in Ecological Research*, 29, 155-
697 193.

698 Polymenakou, P.N., Bertilsson, S., Tselepides, A., Stephanou, E.G. 2005. Bacterial community
699 composition in different sediments from the Eastern Mediterranean Sea: a comparison of
700 four 16S ribosomal DNA clone libraries. *Microbial Ecology*, 50, 447–462.

701 Porter, E.T., Mason R.P., Sanford L.P. 2010. Effect of tidal resuspension on benthic-pelagic
702 coupling in an experimental ecosystem study. *Marine Ecology Progress Series*, 413, 33-53.

703 Pusceddu, A., Fiordelmondo, C., Danovaro, R. 2005. Sediment resuspension effects on the
704 benthic microbial loop in experimental microcosms. *Microbial Ecology*, 50, 602-613.

705 Ravensschlag, K., Sahm, K., Amann, R. 2001. Quantitative molecular analysis of the microbial
706 community in marine Arctic sediments (Svalbard). *Applied and Environmental*
707 *Microbiology*, 67, 387-395.

708 Riemann, F., Helmke, E. 2002. Symbiotic relations of sediment-agglutinating nematodes and
709 bacteria in detrital habitats: the enzyme-sharing concept. *Marine Ecology*, 23, 93-113.

710 Riemann, L., Winding, A. 2001. Community dynamics of free-living and particle-associated
711 bacterial assemblages during a freshwater phytoplankton bloom. *Microbial Ecology*, 42,
712 274-285.

713 Ritzrau, W., Graf, G. 1992. Increase of microbial biomass in the benthic turbidity zone of Kiel
714 Bight after resuspension by a storm event. *Limnology and Oceanography*, 37, 1081-1086.

715 Ritzrau, W. 1996. Microbial activity in the benthic boundary layer: Small-scale distribution
716 and its relationship to the hydrodynamic regime. *Journal of Sea Research*, 36, 171-180.

717 Shapiro, S.S., Wilk, M.B. 1965. An analysis of variance test for normality (complete
718 samples). *Biometrika*, 52, 591-611.

719 Schmidt, J.L., Deming, J.W., Jumars, P.A., Keil, R.G. 1998. Constancy of bacterial abundance
720 in surficial marine sediments. *Limnology and Oceanography*, 43, 976-982

721 Shimeta, J., Amos, C.L., Beaulieu, S.E., Katz, S.L. 2003. Resuspension of benthic protists at
722 subtidal coastal sites with differing sediment composition. *Marine Ecology Progress
723 Series*, 259, 103-115.

724 Ståhlberg, C., Bastviken, D., Svensson, B.H., Rahm, L. 2006. Mineralisation of organic
725 matter in coastal sediments at different frequency and duration of resuspension. *Estuarine,
726 Coastal and Shelf Science*, 70, 317-325.

727 Takahashi, E., Ledauphin, J., Goux, D., Orvain, F. 2009. Optimising extraction of
728 extracellular polymeric substances (EPS) from benthic diatoms: comparison of the
729 efficiency of six EPS extraction methods. *Marine and Freshwater Research*, 60, 1201-1210.

730 Teal, J.M., 1962. Energy flow in the salt-marsh ecosystem of Georgia. *Ecology*, 43, 614-624.

- 731 Thingstad, F. 2000. Control of bacterial growth in idealized food webs. In: Kirchman DL (ed)
732 Microbial ecology of the Oceans. Wiley-Liss, New York, pp 229-260.
- 733 Underwood, G.J.C., Paterson, D.M. 2003. The importance of extracellular carbohydrate
734 production by marine epipellic diatoms. *Advances in Botanical Research*, 40, 183-240.
- 735 Underwood, G.J.C., Kromkamp, J. 1999. Primary production by phytoplankton and
736 microphytobenthos in estuaries. *Advances in Ecological Research*, 29, 93-153.
- 737 van Duyl, F.C., de Winder, B., Kop, A.J., Wollenzien, U. 1999. Tidal coupling between
738 carbohydrate concentrations and bacterial activities in diatom-inhabited intertidal mudflats.
739 *Marine Ecology Progress Series*, 191, 19-32.
- 740 Wainright, S.C. 1987. Stimulation of heterotrophic microplankton production by resuspended
741 marine sediment. *Science*, 238, 1710-1712.
- 742 Wainright, S.C. 1990. Sediment-to-water fluxes of particulate material and microbes by
743 resuspension and their contribution to the planktonic food web. *Marine Ecology Progress*
744 *Series*, 62, 271-281.
- 745 Ward, J.H. 1963. Hierarchical grouping to optimize an objective function. *Journal of the*
746 *American Statistical Association*, 58, 236-244.
- 747 Weerman, E.J., Herman, P.M.J., Van de Koppel, J. 2011. Top-down control inhibits spatial
748 self-organization of a patterned landscape. *Ecology*, 92, 487-495.
- 749 Wilms, R., Sass, H., Köpke, B., Köster, J., Cypionka, H., Engelen, B. 2006. Specific bacterial,
750 archaeal, and eukaryotic communities in tidal-flat sediments along a vertical profile of
751 several meters. *Applied and Environmental Microbiology*, 72, 2756-2764.

752 Ziervogel, K., Arnosti, C. 2009. Enzyme activities in the Delaware Estuary affected by
753 elevated suspended sediment load. *Estuarine, Coastal and Shelf Science*, 84, 253-258.

754 Ziervogel, K., Forster, S. 2006. Do benthic diatoms influence erosion thresholds of coastal
755 subtidal sediments? *Journal of Sea Research*, 55, 43-53.

756

757 **Figures captions**

758 Figure 1: Sampling site.

759 Figure 2: PCA of sediment environmental and biological parameters (showing the variables:
760 samples 1-3 were taken on July 17, 4-6 on July 21, 7-9 on July 22 and 10-12 on July 24).
761 (prot: protein; colloidal: colloidal carbohydrates from EPS; bound: bound carbohydrates from
762 EPS; chloro: chlorophyll *a*; muf: β -glucosidase activity; leu: leucine-aminopeptidase activity;
763 OM: organic matter; cells: prokaryotic abundance).

764 Figure 3: (A) Bacteria and (B) Archaea phyla found in the sediment samples (relative
765 abundance (%), see Materials and Methods for details) (numbers on the graph represent the
766 richness of taxons). (ungproteobacteria: unidentified Gamma-Proteobacteria;
767 uneuraryarchaeota: unidentified Euryarchaeota; uncrenarchaeota: unidentified Crenarchaeota;
768 unarchaea: unidentified Archaea).

769 Figure 4: (A) Total and free cells enrichments in the erodimeter water obtained by treating
770 samples collected on July 17, 21, 22 and 24, (B) attached cells enrichment and (C) active cells
771 enrichment, at each threshold friction velocity ($u^* \text{cm s}^{-1}$).

772

773 Figure 5: (A) β -glucosidase and (B) leucine aminopetidase specific activities in the
774 erodimeter water obtained with each threshold friction velocity ($u^* \text{cm s}^{-1}$) from samples
775 obtained on July 17, 21, 22 and 24.

776

777 Figure 6: PCA of erodimeter water environmental and biological parameters (showing the
778 variables; samples 1-6 from July 17, 7-10 from July 21, 11-15 from July 22 and 16-21 from
779 July 24). (attached: enrichment of attached prokaryotic cells; live: enrichment of live
780 prokaryotic cells; tot: enrichment of prokaryotic cells; free: enrichment of free prokaryotic
781 cells; muf: β -glucosidase activity; leu: leucine-aminopeptidase activity; MOP: particular
782 organic matter; MP: particulate matter; X: threshold friction velocity (u^*); chloro: chlorophyll
783 *a*).

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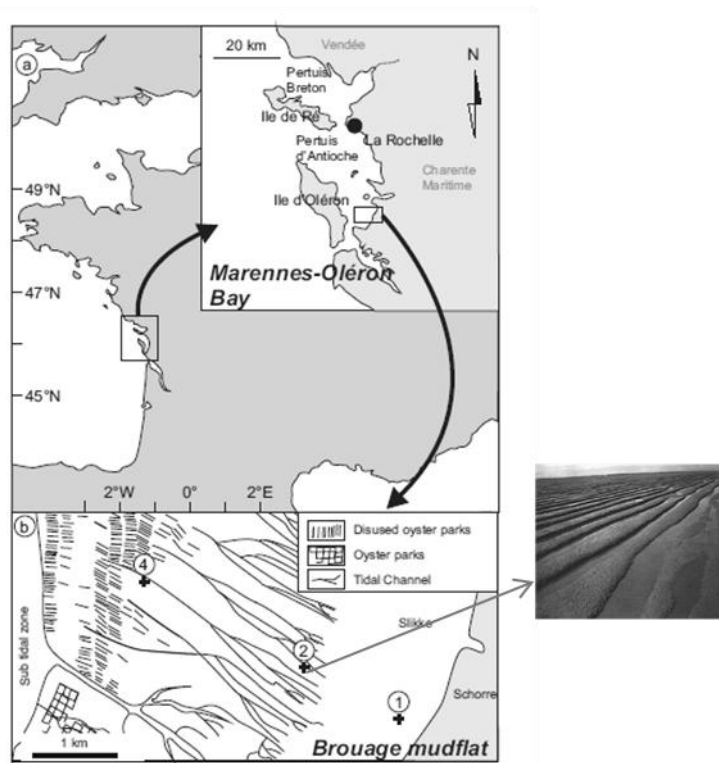
785 Figure 7: Bacterial phyla in the erodimeter water (relative abundance (%), see Materials and
786 Methods for details) after treating the July 17 sediment at each threshold friction velocity
787 ($u^* \text{cm s}^{-1}$): attached (A) and free cells (B) DNA ; (C) free cell DNA in the July 21 erodimeter
788 water; (D) attached cell DNA in the July 22 erodimeter water; (E) attached and (F) free cell

789 DNA in the July 24 erodimeter water (numbers on the graph represent the richness of taxons).
790 (ungproteobacteria: unidentified Gamma-Proteobacteria).

791 Figure 8: Archaeal phyla recovered (relative abundance (%), see Materials and Methods for
792 details) from the sediment sample at each threshold friction velocity ($u^* \text{cm s}^{-1}$). Attached cell
793 DNA in the erodimeter water for (A) July 17 of the, (B) July 21, (C) July 22 and (D) July 24.
794 (numbers on the graph represent the richness of taxons). (uneuryarchaeota: unidentified
795 Euryarchaeota; uncrenarchaeota: unidentified Crenarchaeota; unarchaea: unidentified
796 | Archaea).

797

798 Fig1

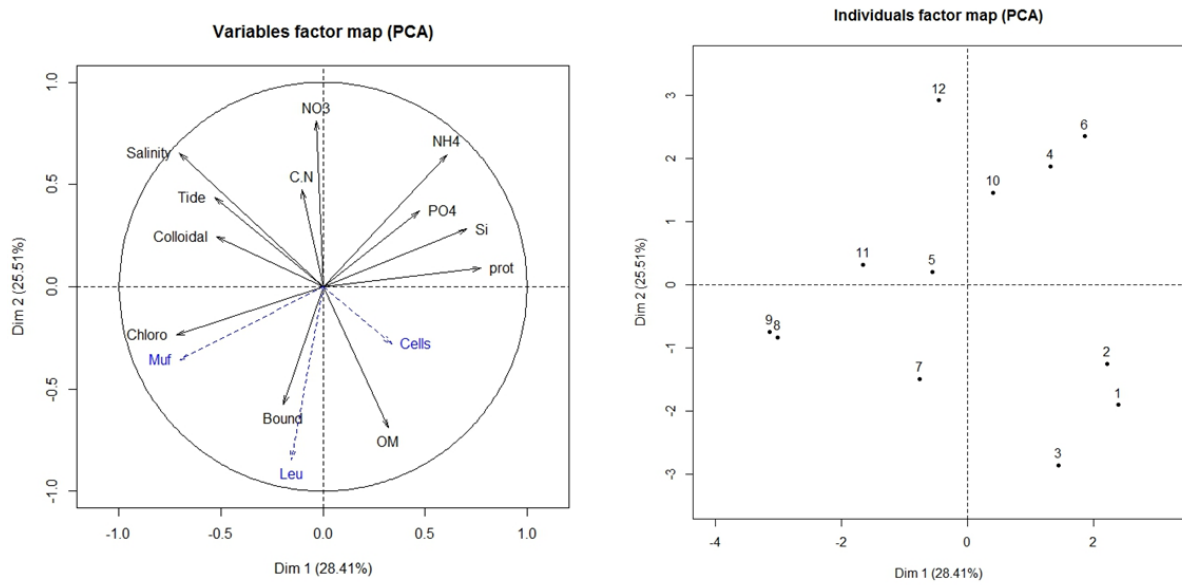


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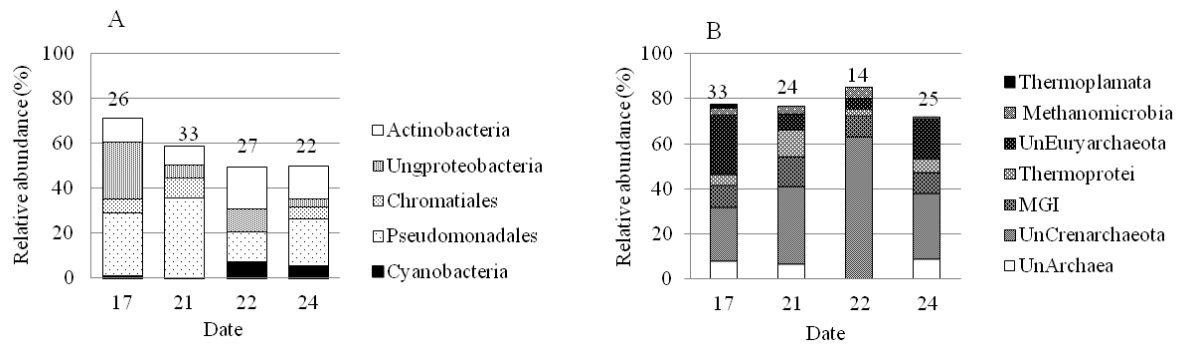
802 Fig2



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805 Fig3

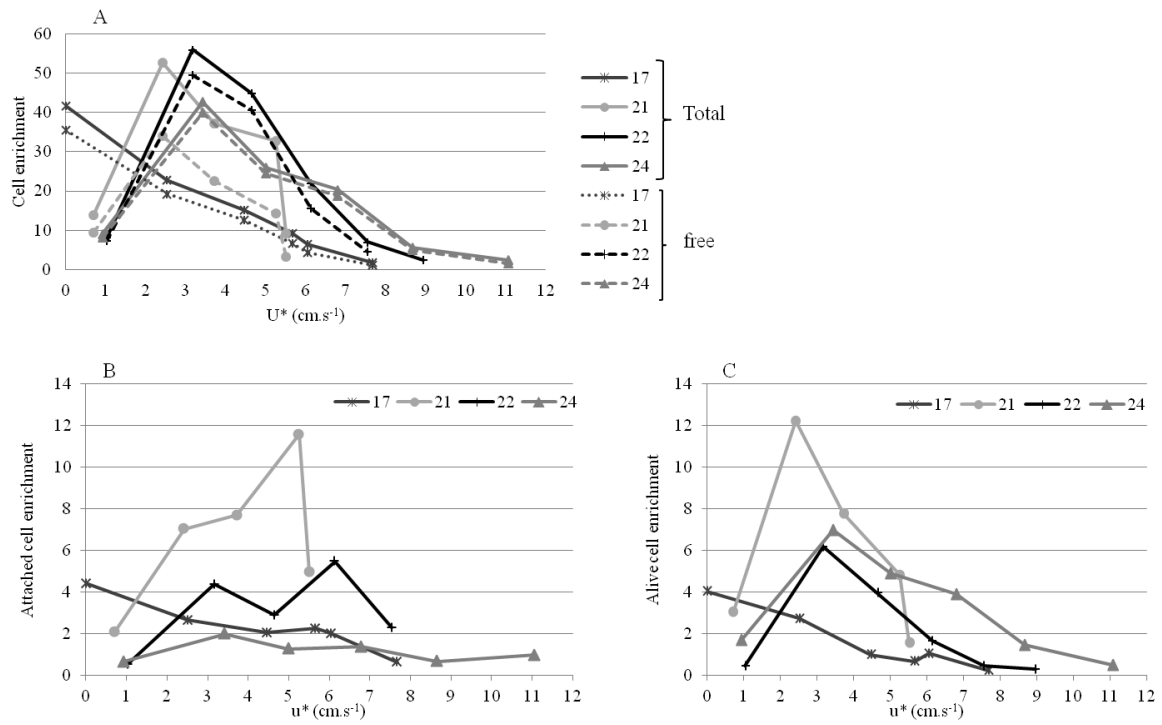


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809 Fig4

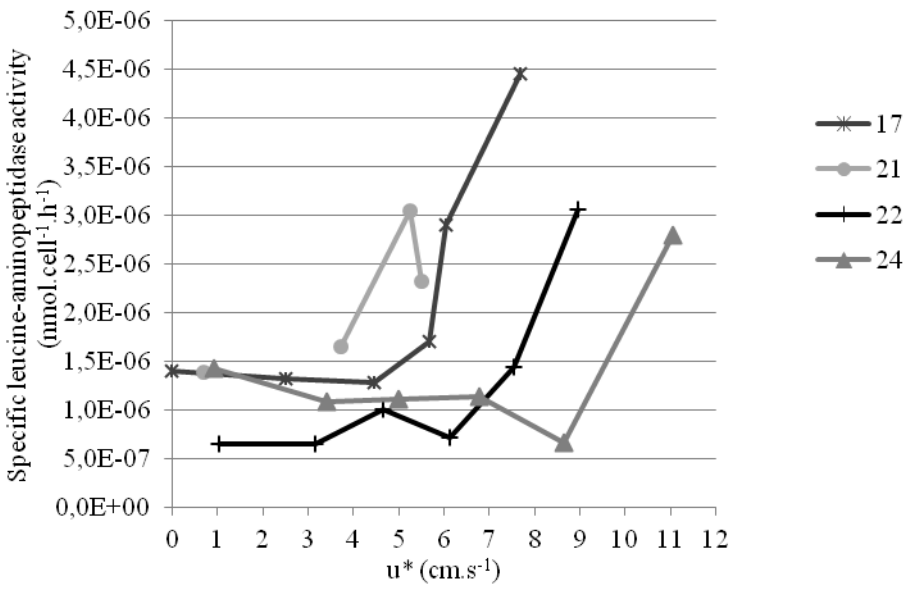
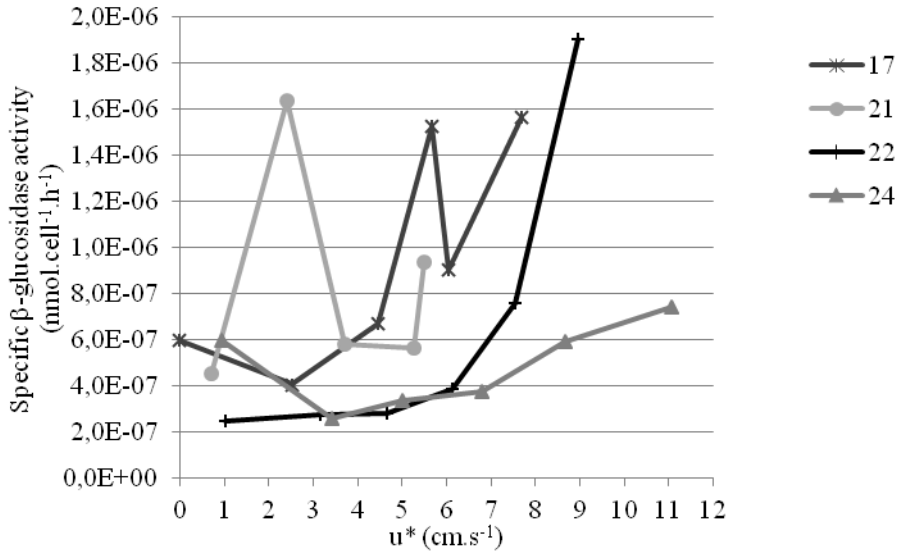


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813 Fig5

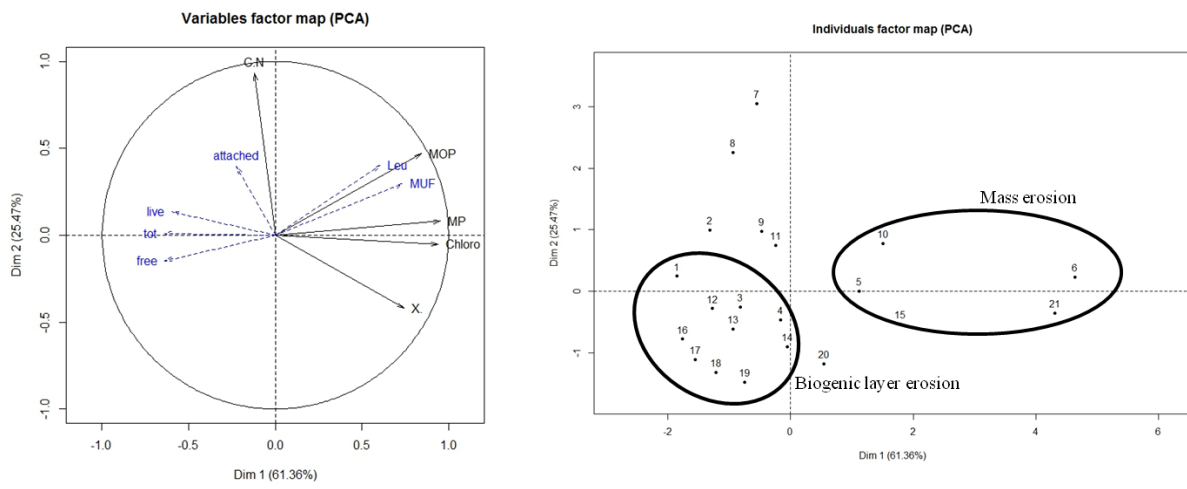


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817 Fig6

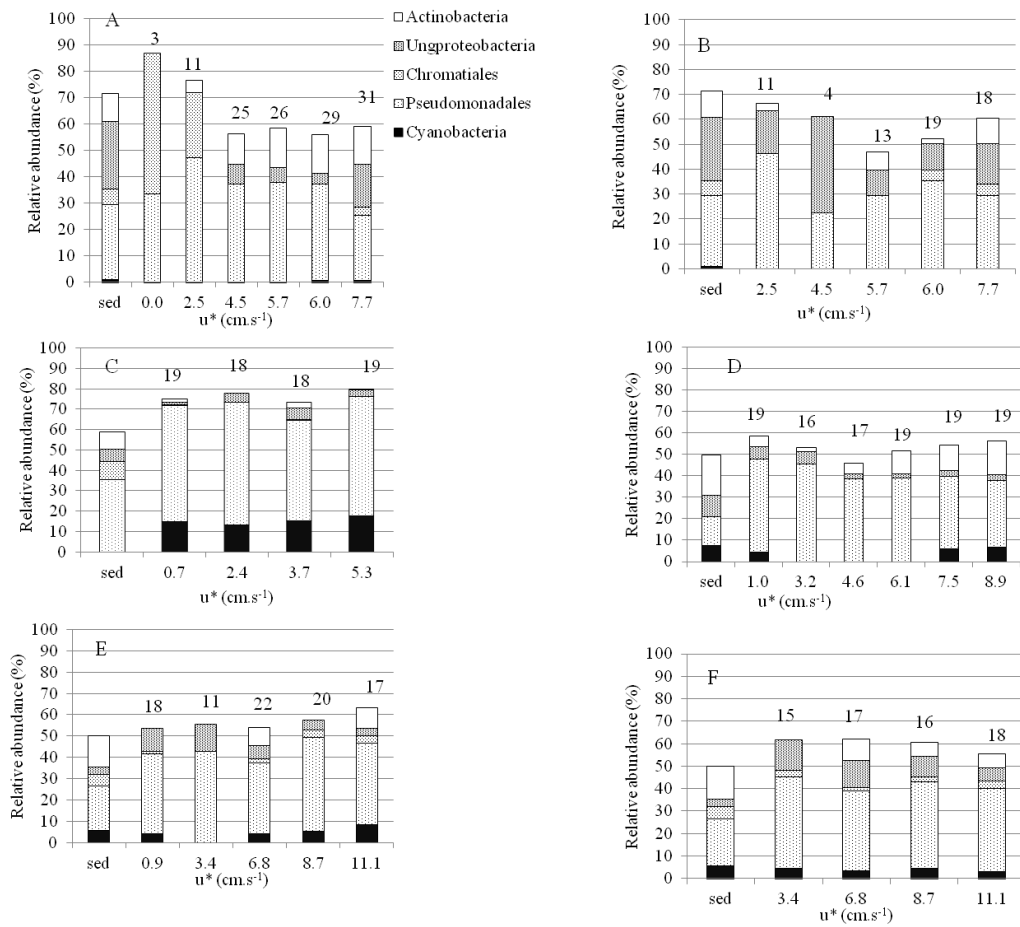


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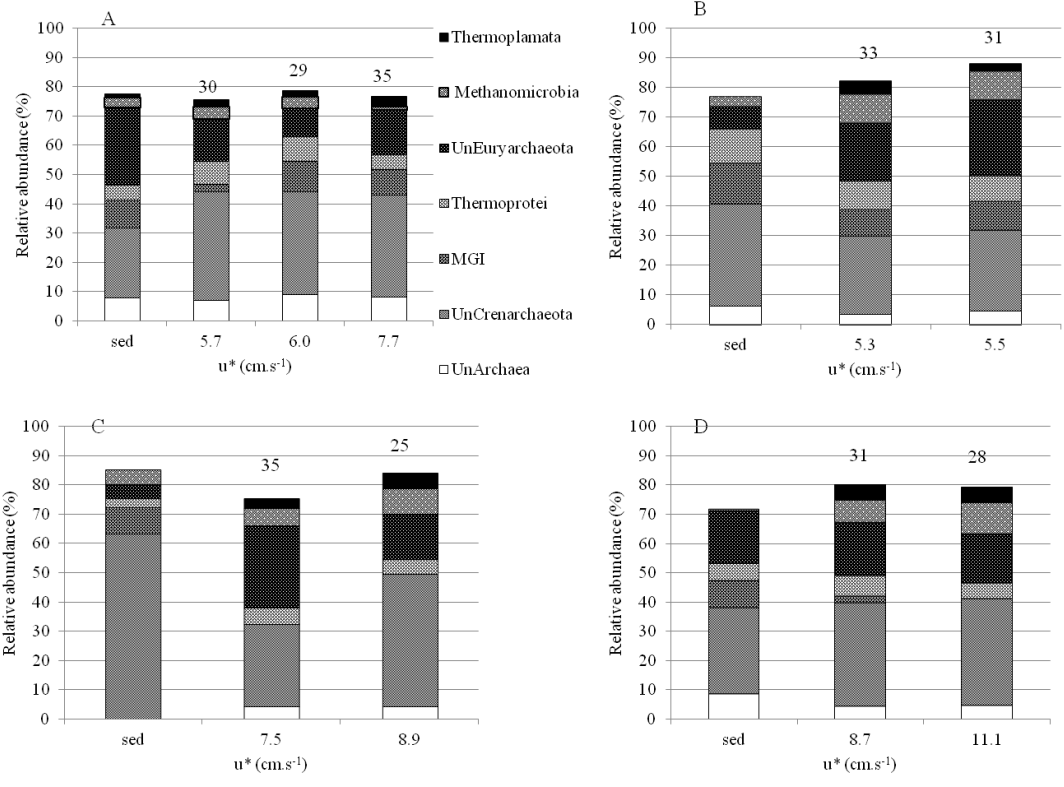
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821 Fig7



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826 Table 1 : Environmental and biological parameters of the sediment cores (means and standard deviation in brackets). (Leu: potential leucine aminopeptidase
827 activity. Muf: potential β -glucosidase activity). Different letters in the same column indicate differences between sampling date (according to an analysis of
828 variance followed by Tükey post hoc test, $p < 0.05$; ns: no significant difference)

829

Sampling dates	Tide coefficient	Salinity (o/oo)	NH4 $\mu\text{mol.L}^{-1}$	NO3 $\mu\text{mol.L}^{-1}$	PO4 $\mu\text{mol.L}^{-1}$	Si $\mu\text{mol.L}^{-1}$	Organic matter g.g^{-1}	Colloidal carbohydrates $\mu\text{g.g}^{-1}$	Bound carbohydrates $\mu\text{g.g}^{-1}$	Proteins $\mu\text{g.g}^{-1}$	C/N	Chloro <i>a</i> $\mu\text{g.g}^{-1}$	Prokaryotes 10^8 cell.g^{-1}	Leu nmol/g/h	Muf nmol/g/h
17	63	41.7 (1.3)a	21.2 (4.1)ab	1.2 (0.4)a	1.4 (0.3)	94.2 (3.8)a	0.2 (0.05)ns	5.1 (2.1)a	11.9 (0.8)ns	4.3 (0.2)ns	7.3 (0.3)ns	7.6 (1.5)ns	8.5 (0.1)a	127,8 (0.1)a	51,6 (0.2)a
21	82	46.4 (0.4)b	34.6 (6.7)a	4.7 (2.7)ab	1.7 (0.8)	96.0 (0.8)a	0.1 (0.03)ns	10.8 (1.7)b	11.1 (1.3)ns	4.4 (0.4)ns	7.8 (0.3)ns	7.2 (0.5)ns	5.6 (0.2)b	104,6 (0.2)b	51,5 (0.2)a
22	81	47.6 (1.1)b	14.9 (8.9)b	1.2 (0.7)a	1.2 (0.2)	75.8 (10.0)b	0.1 (0.07)ns	11.4 (1.6)b	12.5 (1.1)ns	3.8 (0.3)ns	7.3 (0.6)ns	9.2 (1.6)ns	6.3 (0.2)c	131,0 (0.2)a	105,3 (0.2)b
24	72	48.1 (1.9)b	26.1 (5.5)ab	10.3 (5.2)b	1.2 (0.3)	96.1 (5.8)a	0.1 (0.01)ns	8.4 (1.1)ab	10.6 (1.1)ns	3.8 (0.3)ns	9.5 (3.2)ns	8.9 (2.0)ns	8.2 (0.1)a	104,7 (0.2)b	49,1 (0.3)a

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